

1 Building realistic assemblages with a Joint Species Dis- 2 tribution Model

3 David J. Harris (davharris@ucdavis.edu)

4 **Running title:** Building realistic species assemblages

5 **Keywords:** Species distribution model, birds, breeding bird survey, species assemblages,
6 neural network, probabilistic graphical model

7 6712 words (including references etc.)

8 David J. Harris

9 Center for Population Biology

10 1 Shields Avenue

11 Davis, CA 95616 (USA)

12 Abstract

- 13 1. Species distribution models (SDMs) can be used to predict how individual species—and
14 whole assemblages of species—will respond to a changing environment. Until now, these
15 models have either assumed (1) that species' occurrence probabilities are uncorrelated,
16 or (2) that species respond linearly to preselected environmental variables. These two
17 assumptions currently prevent ecologists from modeling assemblages with realistic co-
18 occurrence and species richness properties.
- 19 2. This paper introduces a stochastic feedforward neural network, called mistnet, which makes
20 neither assumption. Thus, unlike most SDMs, mistnet can account for non-independent
21 co-occurrence patterns driven by unobserved environmental heterogeneity. And unlike
22 recently proposed Joint SDMs, mistnet can also learn nonlinear functions relating species'
23 occurrence probabilities to environmental predictors.
- 24 3. Mistnet makes more accurate predictions about the North American bird communities
25 found along Breeding Bird Survey transects than several alternative methods tested. In
26 particular, typical assemblages held out of sample for validation were nearly 50,000 times
27 more likely under the mistnet model than under independent combinations of single-species
28 models.
- 29 4. Apart from improved accuracy, mistnet shows two other important benefits for ecological
30 research and management. First: by analyzing co-occurrence data, mistnet can identify
31 unmeasured—and perhaps unanticipated—environmental variables that drive species
32 turnover. For example, mistnet identified a strong grassland/forest gradient, even though
33 only temperature and precipitation were given as model inputs. Second: mistnet is able

34 to take advantage of incomplete data sets to guide its predictions towards more realistic
35 assemblages. For example, mistnet automatically adjusts its expectations to include more
36 forest-associated species in response to a stray observation of a forest-dwelling warbler.

37 Introduction

38 Programs for managing and understanding biodiversity each require information about where
39 species occur and where they could occur. Statistical approaches to these questions, such as
40 species distribution models (SDMs), are important because they can help us anticipate how
41 beneficial species might fare—or how harmful species might spread—in scenarios that we
42 cannot observe directly (Elith & Leathwick 2009). Modern SDMs need not assume that species
43 respond to environmental variation in a pre-specified way (e.g. linearly or quadratically);
44 relaxing this assumption has substantially improved our ability to make predictions about
45 where species can occur (Elith *et al.* 2006).

46 Unfortunately, existing nonlinear approaches do not always answer the most pressing questions
47 for ecologists. Ecologists are not only interested in individual species; we are also interested
48 in learning about higher-level patterns, such as community structure, species richness, species
49 turnover, and alternative stable states (Chase 2003). While SDMs are often combined
50 (“stacked”) to generate assemblage-level predictions (Pellissier *et al.* 2013), doing so requires
51 assuming that species’ occurrence probabilities are uncorrelated (Clark *et al.* 2013; Calabrese
52 *et al.* 2014). As shown in more detail below, ignoring these correlations leads stacked
53 models to predict incoherent jumbles of species rather than realistic assemblages (Clark *et al.*
54 2013). A major source of non-independence among species—which stacked SDMs ignore—is
55 shared dependence on unobserved environmental factors (McInerny & Purves 2011; Figure
56 1; Calabrese *et al.* 2014). Given that most models only use climate variables as predictors
57 (Austin & Van Niel 2011), the set of unobserved factors will usually include *all of ecology*
58 apart from climatic influences. SDMs’ failure to model other ecological processes is thus
59 widely considered to be a major omission from statistical ecology’s toolbox (Austin & Van

60 Niel 2011; Guisan & Rahbek 2011; Kissling *et al.* 2012; Wisz *et al.* 2013; Clark *et al.* 2013).

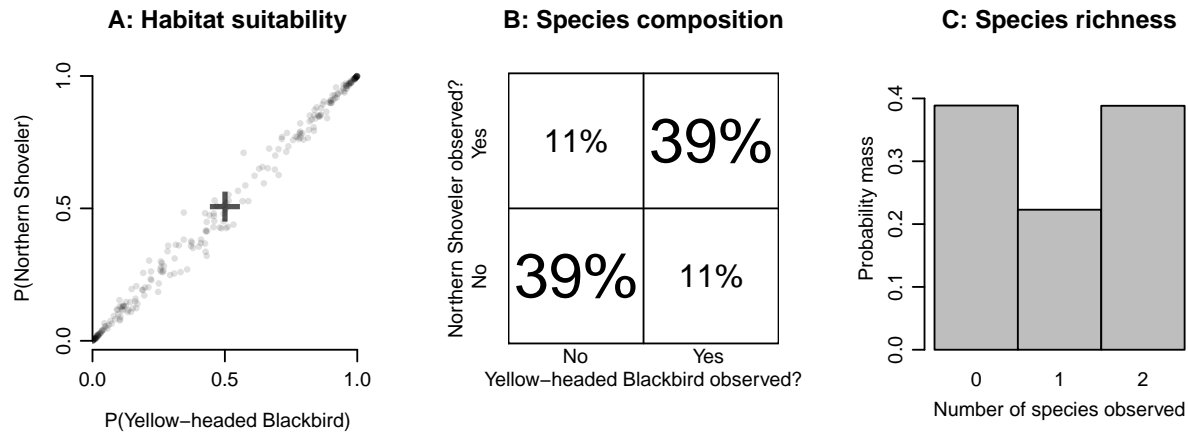


Figure 1: Unobserved environmental heterogeneity can induce correlations between species; ignoring this heterogeneity can produce misleading results. **A**: Based on climate predictors, a pair of single-species models might predict 50% occurrence probabilities for each of two wetland species (black cross). Climate predictors are not sufficient in this case, however: a site's suitability for these species cannot really be determined without information about the availability of wetland habitat. Real habitats will tend to be suitable for both species (dense cloud of points in upper-right corner) or neither (lower-left corner), depending on this unmeasured variable. **B** This correlation among species substantially alters the set of assemblages one would expect to observe. (Under independence, all four possibilities would be equally probable.) **C** Positive correlations among species can even induce a strongly bimodal distribution of species richness values.

61 In the last few years, several mixed models have been proposed to help explain the co-
62 occurrence patterns that stacked SDMs ignore (Latimer *et al.* 2009; Ovaskainen, Hottola
63 & Siitonen 2010; Golding 2013; Clark *et al.* 2013; Pollock *et al.* 2014). These *joint* species
64 distribution models (JSDMs) can produce mixtures of possible species assemblages (points
65 in Figure 1a), rather than relying on a small number of environmental measurements to
66 fully describe each species' probability of occurrence (which would collapse the distribution
67 in Figure 1a to a single point; Pollock *et al.* 2014). In JSDMs (as in nature), a given set
68 of temperature and precipitation measurements could be consistent with a number of very
69 different possible sets of co-occurring species, depending on factors that ecologists have not

70 necessarily measured or even identified as important. JSDMs represent these unobserved
71 (latent) factors as random variables whose true values are unknown but whose existence
72 would still help explain discrepancies between the data and the stacked SDMs' predictions
73 (Figures 1b and 1c). While JSDMs represent a major advance in community-level modeling
74 (Clark *et al.* 2013; Pollock *et al.* 2014), existing implementations have all assumed that
75 species' responses to the environment are linear (in the sense of a generalized linear model).
76 Thus, these JSDMs sacrifice the flexibility of modern single-species models, reducing their
77 accuracy and limiting their utility.

78 Here, I present a new R package for assemblage-level modeling—called *mistnet*—that does not
79 rely on independence (as stacks of single-species models do) or linearity (as previous JSDMs
80 do). *Mistnet* is a stochastic feed-forward neural network (Neal 1992; Tang & Salakhutdinov
81 2013) that combines the nonlinear flexibility of modern single-species models with the latent
82 variables found in previous JSDMs (cf Hutchinson, Liu & Dietterich 2011). In order to
83 demonstrate the value of this approach, I compared *mistnet*'s predictive likelihood with
84 that of several existing models, using observational data from thousands of North American
85 Breeding Bird Survey transects (BBS; Sauer *et al.* 2011). A high predictive likelihood
86 indicates that the model expects to see assemblages like those found along transects held
87 out-of-sample, while a very low likelihood means that the model has effectively ruled those
88 assemblages out due to overfitting or underfitting.

89 An accurate JSDM would up new possibilities for research and effective management. For
90 example, although most models only have access to climate data (Austin & Van Niel 2011),
91 a successful model of community structure should also be able to identify the major axes of
92 non-climate variation that drive species turnover based on the species' observed co-occurrence

93 patterns. Moreover, a successful assemblage-level model would be able to take advantage of
94 partially-completed samples or other kinds of prior information about a few species to inform
95 its predictions about the rest of the assemblage. Since data collection efforts are frequently
96 asymmetrical or incomplete, the ability to transfer information from well-documented taxa to
97 more cryptic or rare species would prove valuable for community ecologists and conservationists
98 alike. While a model’s ability to infer, for example, that “waterbirds like water” would not
99 provide any novel biological insights, it would demonstrate that a modeling framework is
100 ready to tackle more difficult problems where the biology is not already known.

101 **Materials and Methods Methods**

102 Methods are presented in four main sections: (1) an introduction to the data sets used in
103 this analysis, (2) a description of mistnet, (3) a summary of the existing methods used for
104 model comparison, and (4) criteria for model evaluation.

105 **Data**

106 Field survey data was obtained from the 2011 Breeding Bird Survey (BBS; Sauer *et al.* 2011).
107 The BBS data consists of thousands of transects (“routes”), which I used as the main unit
108 for my analysis. Each route includes 50 stops, about 0.8 km apart. At each stop, all the
109 birds observed in a 3-minute period are recorded, using a standardized procedure. Following
110 BBS recommendations, I omitted nonstandard routes and data collected on days with bad
111 weather.

112 In order to evaluate SDMs’ capacities for predicting species composition, I split the routes

113 into a “training” data set consisting of 1559 routes and a “test” data set consisting of 280
114 routes (Figure 2; Appendix A). The two data sets were separated by a 150-km buffer to
115 ensure that models could not rely on spatial autocorrelation to make accurate predictions
116 about the test set (Bahn & McGill 2007) (Appendix A). Each model was fit to the same
117 training set, and then its performance was evaluated out-of-sample on the test set.

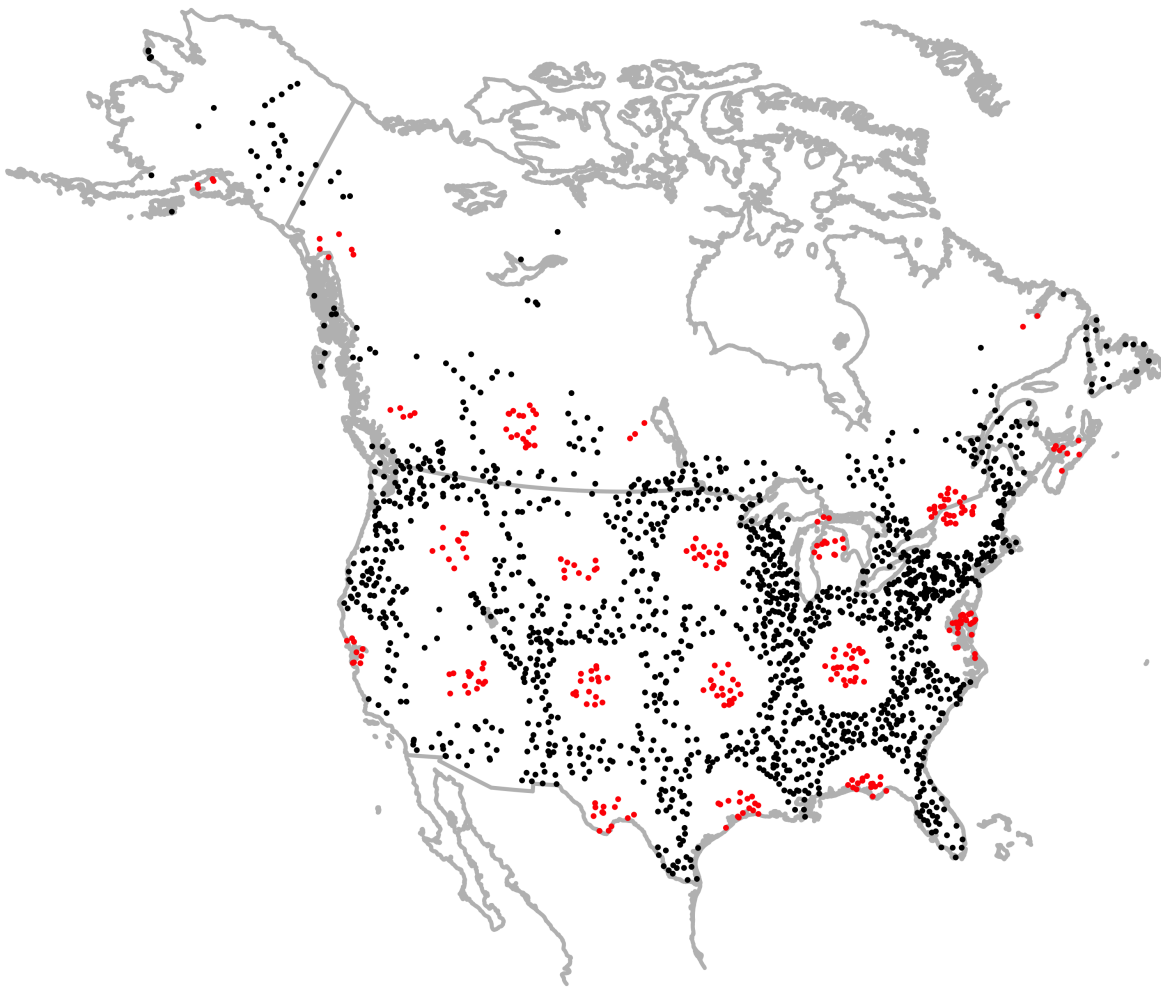


Figure 2: Map of the BBS routes used in this analysis. Black points are training routes; red ones are test routes. The training and test routes are separated by a 150-km buffer in order to minimize spatial autocorrelation across the two partitions.

118 Observational data for each species was reduced to “presence” or “absence” at the route level,

119 ignoring the possibility of observation error for the reasons outlined in (Welsh, Lindenmayer
120 & Donnelly 2013). It would be possible to incorporate the possibility of such errors in the
121 model-fitting procedure if appropriate data were available, as was done in (Hutchinson *et al.*
122 2011). 368 species were chosen for analysis according to a procedure described in Appendix
123 A.

124 To obtain environmental predictors for the model, I extracted the 18 Bioclim climate variables
125 for each route from Worldclim (version 1.4; Hijmans *et al.* 2005). I omitted variables that
126 were nearly collinear with one another (i.e. $|r| > 0.8$) using the `findCorrelation` function in
127 the `caret` package (Wing *et al.* 2013), leaving eight climate-based predictors (Appendix A).
128 Since most SDMs do not use land cover data (Austin & Van Niel 2011) and one of mistnet's
129 goals is to make inferences about unobserved environmental variation, no other variables
130 were included in this analysis.

131 Finally, I obtained habitat classifications for each species from the Cornell Lab of Ornithology's
132 All About Birds website (www.allaboutbirds.org) using an R script written by K. E. Dybala.

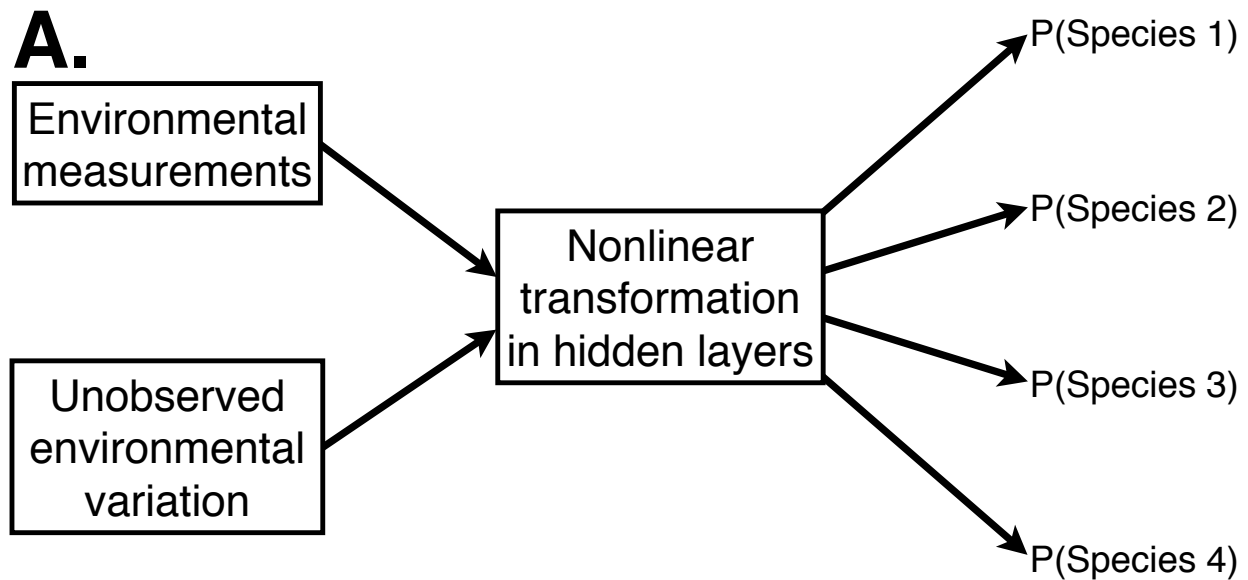
133 **Introduction to stochastic neural networks**

134 Neural networks describe nonlinear mappings from input variables to predictions about one
135 or more output variables. In general, ecologists have not had much success using neural
136 networks for SDM, compared with other methods (e.g. Dormann *et al.* 2008). However,
137 modern neural networks have recently outperformed other machine learning techniques in a
138 wide range of applied contexts (Bengio 2013) and are thus worth a second look.

139 Mistnet models are *stochastic* neural networks, meaning that they include latent random
140 variables (Neal 1992; Tang & Salakhutdinov 2013). In such a model, species' occurrence

141 probabilities are not fully specified the variables ecologists happen to measure, but can also
142 depend on factors that have not been observed. In the absence of any information about these
143 variables, mistnet (like other JSDMs) represents them using standard normal distributions.
144 Depending on which values are sampled from these normal distributions and fed through the
145 neural network, the model will expect to see different kinds of species assemblages (Figure 3).
146 While the model's main function is to make predictions about the species found in a given
147 environment, inference can also proceed backward through the network, so that the presence
148 (or absence) of a particular species can provide indications about the local environment—and
149 thus about the likely configuration of the rest of the assemblage. This kind of inference could
150 be useful in a variety of important contexts. For example, data is often more plentiful about
151 waterfowl than about other wetland species, due to interest from hunters and conservation
152 groups. If waterfowl are known to be present along a route, then a JSDM should recognize
153 that suitable habitat was available, automatically increasing the estimated probability of
154 occurrence for other species known to have similar habitat requirements. Notably, none of
155 this extra inferential power requires that the mistnet user understand *which* environmental
156 factors are driving the correlations between species, since these correlations are automatically
157 inferred from species' co-occurrence patterns.

158 The neural network used here (illustrated in Figure 3b) is trained to find a way of representing
159 different environmental conditions such that each species' response to the environment can
160 be described using a small number of coefficients (e.g. 15 in this analysis; Appendix B). The
161 small number of coefficients and the uniformity of their functions makes mistnet models highly
162 interpretable: the coefficients linking the second hidden layer to a given species' probability of
163 occurrence essentially describe that species' responses to a few leading principal components



B.

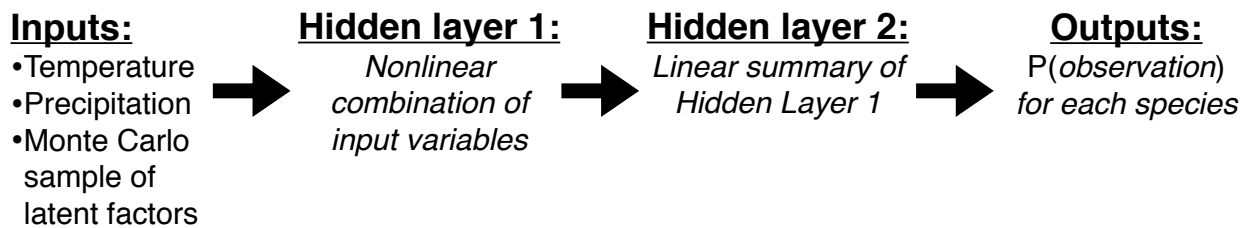


Figure 3: **A** A generalized diagram for stochastic feed-forward neural networks that transform environmental variables into occurrence probabilities multiple species. The network’s hidden layers perform a nonlinear transformation of the observed and unobserved (“latent”) environmental variables; each species’ occurrence probability then depends on the state of the final hidden layer. **B** The specific network used in this paper, with two hidden layers. The inputs include Worldclim variables involving temperature and precipitation, as well as random draws from each of the latent environmental factors. These inputs are multiplied by a coefficient matrix and then nonlinearly transformed in the first hidden layer. The second hidden layer uses a different coefficient matrix to linearly transform its inputs down to a smaller number of variables (like Principal Components Analysis of the previous layer’s activations). A third matrix of coefficients links each species’ occurrence probability to each of the variables in this linear summary (like one instance of logistic regression for each species). The coefficients are all learned using a variant of the backpropagation algorithm.

164 of environmental variation (cf Vincent et al. (2010)). For comparison, the boosted regression
165 tree SDMs used below (Elith, Leathwick & Hastie 2008) have tens of thousands of coefficients
166 per species, with entirely new interpretations for each new species' coefficients.

167 How do we train the model to make good predictions? As with most neural networks,
168 mistnet's coefficients are initialized randomly, and then the model climbs the log-likelihood
169 surface by iteratively adjusting the coefficients toward better values. In mistnet models, the
170 adjustments are calculated with a variant of the backpropagation algorithm (Rumelhart,
171 Hinton & Williams 1986; Murphy 2012) suggested by Tang & Salakhutdinov (2013) for
172 stochastic neural networks. The fitting procedure alternates between inferring the states of
173 the latent variables (via importance sampling) and updating the model's coefficients (via
174 backpropagation). Both phases of model fitting are described in more detail in Appendix
175 B. Despite importance sampling's imprecision, this generalized expectation maximization
176 procedure will converge to a local optimum on the likelihood surface with probability one
177 (Neal & Hinton 1998; Tang & Salakhutdinov 2013), ensuring that the expected likelihood
178 is high after averaging over the possible random samples. Following best practices (Orr &
179 Müller 1998; Murphy 2012), mistnet constrains the coefficients using L_2 regularization to
180 prevent overfitting; the strength of this "weight decay" term was chosen by cross-validation,
181 as described in the Appendix.

182 The mistnet source code can be viewed and downloaded from <https://github.com/davharris/mistnet>.

183 While the user interface and most of the algorithms are written in R, a small portion of
184 the code is written in C++, using Rcpp (Eddelbuettel & Francois 2011) to manage the
185 interface between languages and RcppArmadillo (Eddelbuettel & Sanderson 2014) to access
186 the Armadillo linear algebra library for faster matrix manipulations (Sanderson 2010).

187 Existing models used for comparison

188 I compared mistnet's predictive performance with two machine learning techniques and with
189 a linear JSDM called BayesComm (Golding 2013; Golding & Harris 2014). Each of these
190 techniques is described briefly below; implementational details and settings for each method
191 can be found in the Appendix.

192 The first machine learning method I used for comparison, boosted regression trees (BRT;
193 Elith *et al.* 2008), is among the most powerful techniques available for single-species SDM
194 (Elith *et al.* 2006; Elith *et al.* 2008). I trained one BRT model for each species using R's `gbm`
195 package (Ridgeway 2013) and stacked them following the recommendations in (Calabrese *et*
196 *al.* 2014).

197 I also used a neural network model with no stochastic latent variables as a baseline against
198 which to compare mistnet. Such neural networks do share some information among species
199 (i.e. all species' log-odds of occurrence are linear combinations of the same hidden layer), but
200 like most other multi-species SDMs (De'ath 2002; Leathwick *et al.* 2005; Ferrier *et al.* 2007)
201 they are not JSDMs and do not explicitly model co-occurrence (Clark *et al.* 2013). The
202 neural net baseline was trained using the `nnet` package (Venables & Ripley 2002).

203 Finally, I trained a BayesComm model (Golding 2013; Golding & Harris 2014) to evaluate
204 the importance of mistnet's nonlinearities compared to a linear alternative that also models
205 co-occurrence explicitly.

206 To ensure a level playing field, each modeling approach was given about 15 hours on the same
207 computer for cross-validation and to make its predictions, as described in the Appendix.

208 **Evaluating model predictions along test routes**

209 I evaluated mistnet's predictions both qualitatively and quantitatively. Qualitative assess-
210 ments involved looking for patterns in the model's predictions and comparing them with
211 ornithological knowledge (e.g. the habitat classifications provided by the Cornell Lab of
212 Ornithology).

213 Each model was evaluated quantitatively on the test routes (red points in Figure 2) to
214 assess its predictive accuracy out-of-sample. Models were scored according to their predictive
215 likelihoods, i.e. the probabilities they assigned to various scenarios observed in the test
216 data. Models with high likelihoods expect realistic co-occurrence patterns, and should yield
217 more biologically relevant insights about the processes underlying those patterns. Models
218 that overfit or underfit will have lower out-of-sample likelihoods, and should be trusted less
219 to provide these kinds of insights. I tested each model's ability to make several kinds of
220 predictions, ranging from estimates of the probability of observing particular species at a given
221 location, to predictions about the species richness and composition of entire assemblages.

222 To quantify the difficulties each model faced as it made predictions about increasingly large
223 assemblages, I estimated their route-level predictive likelihoods for randomly-chosen groups
224 of species, ranging in size from individual species pairs to the full set of 368 species in
225 the data set. Models that assumed species were uncorrelated should see an exponential
226 decay in their likelihoods as the number of species increases (since the probability of making
227 correct predictions for a set of uncorrelated species equals the product of their individual
228 probabilities), while BayesComm and mistnet should be able to take advantage of correlations
229 to simplify problem of making predictions for the larger assemblages.

230 Finally, each model predicted a range of possible species richness values for each test route;

231 I calculated quantiles for each model's predictions using the Poisson-binomial distribution
232 (Hong 2013), as recommended in Calabrese et al. (2014).

233 **Results and Discussion**

234 **Mistnet's view of North American bird assemblages**

235 I began by decomposing the variance in the mistnet's species-level predictions among-routes
236 (which varied in their climate values) and residual variation within routes. On average,
237 the residuals accounted for 29% of the variance in mistnet's predictions, indicating that
238 non-climate factors play a substantial role in habitat filtering at continental scales.

239 If the non-climate factors mistnet identified were biologically meaningful, then there should be
240 a strong correspondence between the 15 coefficients assigned to each species by mistnet and
241 the habitat classifications assigned by the Cornell lab of Ornithology. A linear discriminant
242 analysis (LDA; Venables & Ripley 2002) demonstrated such a correspondence (Figure 4). The
243 two-dimensional subspace in Figure 4 explains 19% of the total variance in species' coefficients
244 (representing an even greater portion of the non-climate variance). Mistnet's coefficients
245 cleanly distinguished several groups of species by habitat association (e.g. "Grassland" species
246 versus "Forest" species), though the model largely failed to distinguish "Marsh" species from
247 "Lake/Pond" species and "Scrub" species from "Open Woodland" species. These results
248 indicate that the model has identified the broad differences among communities, but that it
249 lacks some fine-scale resolution for distinguishing among types of wetlands and among types
250 of partially-wooded areas. Alternatively, perhaps these finer distinctions are not as salient at
251 the scale of a 40-km transect.

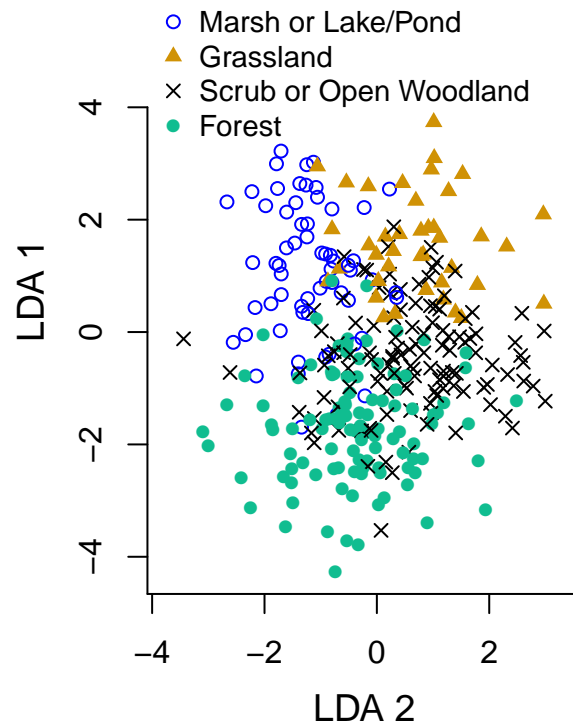


Figure 4: Each species' mistnet coefficients have been projected into a two-dimensional space by linear discriminant analysis (LDA) in order to maximize the spread between the six habitat types assigned to species by the Cornell Lab of Ornithology's All About Birds website. The figure shows that mistnet cleanly separates "Grassland" species from "Forest" species, with "Scrub" and "Open Woodland" species representing intermediates along this axis of variation. "Marsh" and "Lake/Pond" species cluster together in the upper-left. The other habitat classes were included in the LDA, but are not shown here.

252 Figure 5A shows how the forest/grassland gradient identified by mistnet affects the model's
253 predictions for a pair of species with opposite responses to forest cover. The model cannot
254 tell *which* of these two species will be observed (since it was only provided with climate data),
255 but the model has learned enough about these two species to tell that the probability of
256 observing *both* along the same 40-km transect is much lower than would be expected if the
257 species were uncorrelated.

258 Figure 5A reflects a great deal of uncertainty, which is appropriate considering that the model
259 has no information about a crucial environmental variable (forest cover). Often, however,
260 additional information is available that could help resolve this uncertainty, and the mistnet
261 package includes a built-in way to do so, as indicated in Figures 5B and 5C. These panels
262 show how the model is able to use an observation of a forest-associated Nashville Warbler
263 (*Oreothlypis ruficapilla*) to indicate that a whole suite of other forest-dwelling species are
264 likely to occur nearby, and that a variety of species that prefer open fields and wetlands
265 should be absent. Similarly, Figure 5D shows how the presence of a Redhead duck (*Aythya*
266 *americana*) can inform the model that a route is suitable habitat for a variety of other ducks,
267 as well as for other wetland-associated species such as marsh-breeding blackbirds, sandpipers,
268 and rails (along with a few other species that do not fit this theme as nicely). None of these
269 inferences would be possible from a stack of disconnected single-species SDMs.

270 **Model comparison: species richness**

271 Environmental heterogeneity plays an especially important role in determining species richness,
272 which is often overdispersed relative to models' expectations (O'Hara 2005). Figure 6 shows
273 that mistnet's predictions respect the heterogeneity one might find in nature: areas with

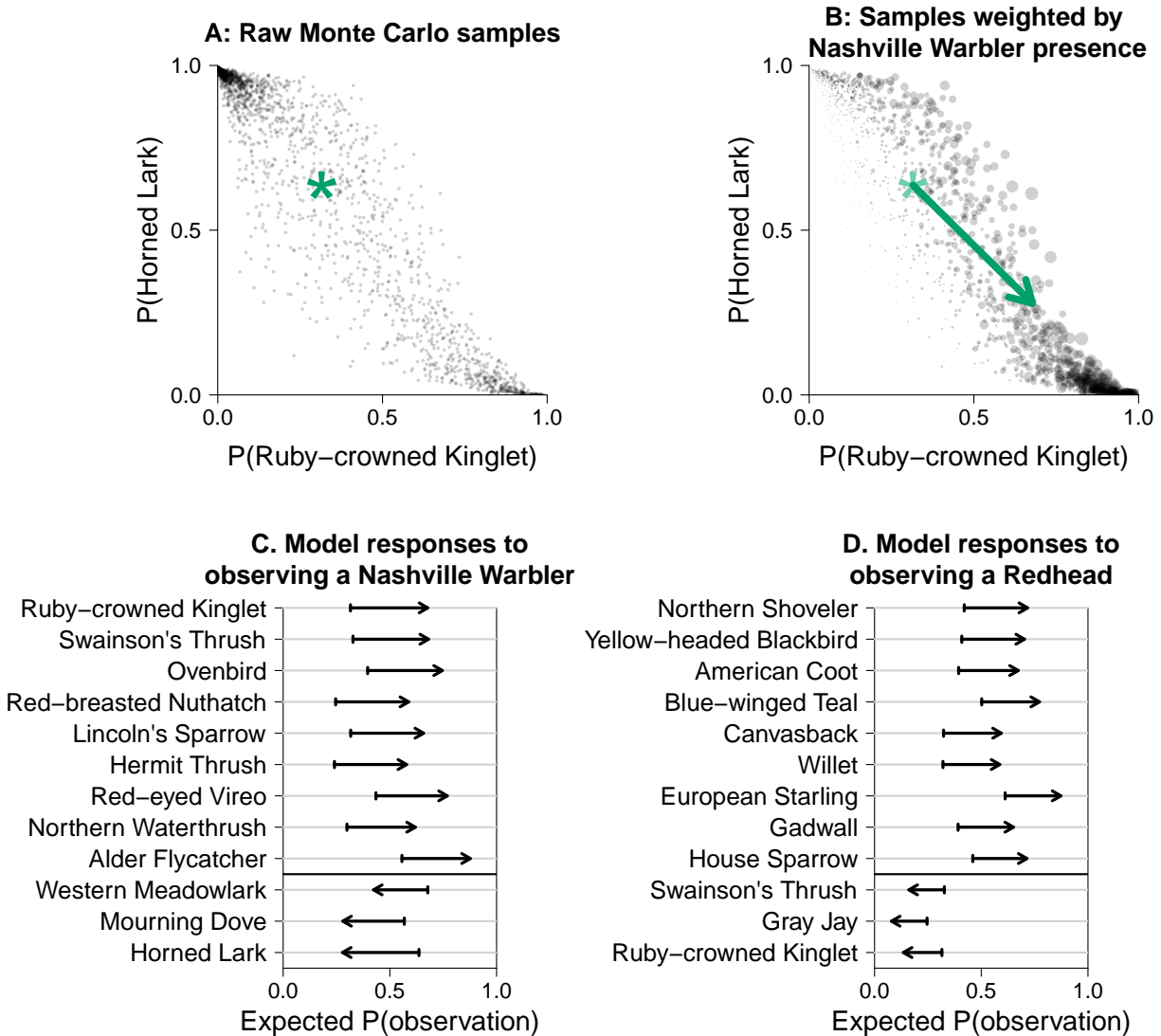


Figure 5: **A.** The mistnet model has learned that Ruby-crowned Kinglets (*Regulus calendula*) and Horned Larks (*Eremophila alpestris*) have opposite responses to some environmental factor whose true value is unknown. Based on these two species' biology, an ornithologist could infer that this unobserved variable is related to forest cover, with the Kinglet favoring more forested areas and the Lark favoring more open areas. **B.** The presence of a forest-dwelling Nashville Warbler (*Oreothlypis ruficapilla*) provides the model with a strong indication that the area is forested, increasing the weight assigned to Monte Carlo samples that are suitable for the Kinglet and decreasing the weight assigned to samples that are suitable for the lark. **C.** The Nashville Warbler's presence similarly suggests increased occurrence probabilities for a variety of other forest species, as well as decreased probabilities for species associated with wetlands and grasslands. **D.** If a Redhead (*Aythya americana*) has been observed along a route, the model correctly expects to see more ducks, rails and sandpipers in the same area.

274 a given climate could be largely unsuitable for waterfowl (Anatid richness < 2 species) or
275 marshy and open (Anatid richness > 10 species). Under the independence assumption used
276 for stacking SDMs, however, both of these very plausible scenarios would be ruled out (Figure
277 6A).

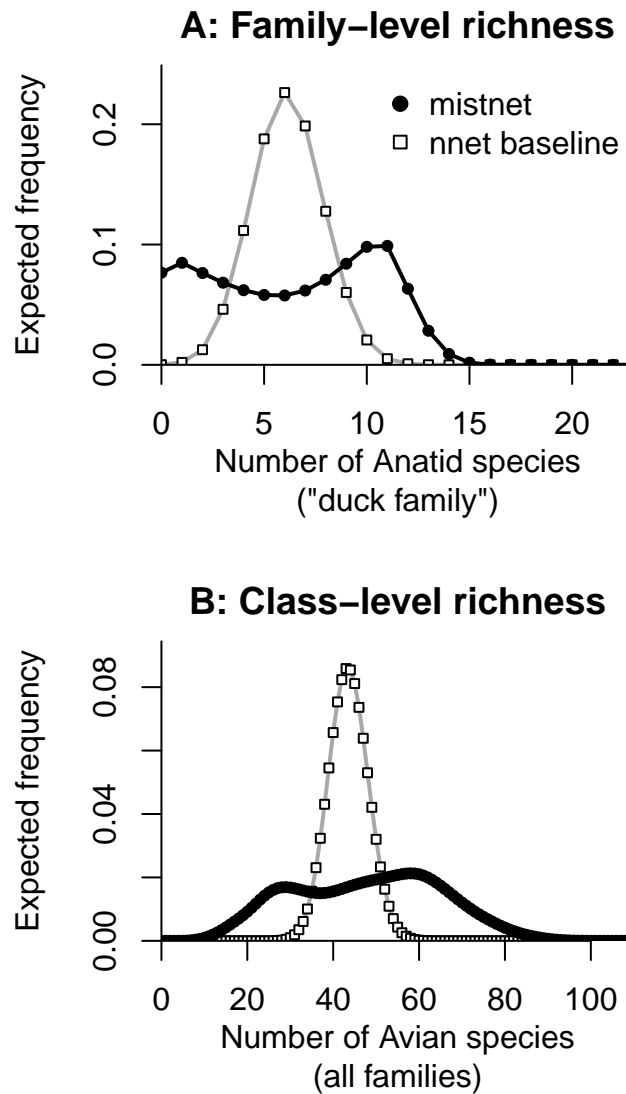


Figure 6: The predicted distribution of species richness one would expect to find based on predictions from mistnet and the baseline neural network. **A.** Anatid species (waterfowl). **B.** All bird species. BRT's predictions (not shown) are similar to the baseline network, since neither one accounts for the effects of unmeasured environmental heterogeneity.

278 Unfortunately, stacking leads to even larger errors when predicting richness for larger groups,

279 such as the complete set of birds studied here. Models that stacked independent predictions
280 underestimated the range of biologically possible outcomes (Figure 6B), frequently putting
281 million-to-one or even billion-to-one odds against species richness values that were actually
282 observed. In more concrete terms, half of the observed species richness values fell outside
283 these models' 95% confidence intervals. The overconfidence associated with stacked models
284 could have serious consequences in both management and research contexts if we fail to
285 prepare for species richness values outside such an unreasonably narrow range.

286 Mistnet, on the other hand, was able to explore the range of possible non-climate environments
287 to avoid these missteps: 90% of the test routes fell within mistnet's 95% confidence intervals,
288 and the log-likelihood ratio decisively favored it over stacked alternatives.

289 **Model comparison: single species**

290 The two neural network models had the best performance at the level of individual species
291 (Table 1). The neural networks' advantage over BRT was largest for low-prevalence species
292 (linear regression of log-likelihood ratio versus log-prevalence; $p = 0.004$). This is consistent
293 with previous observations that multi-species models can outperform single-species approaches
294 for rare species (Leathwick, Elith & Hastie 2006), which will often be of the greatest
295 conservation concern. BayesComm's predictions were substantially worse than any of the
296 machine learning methods, which I attribute to its inability to learn nonlinear responses to
297 the environment.

method	expected.log.likelihood	likelihood.ratio
nnet	-48.7	21.3
mistnet	-48.7	21
BRT	-51.7	1
BayesComm	-56.6	0.00771

Table 1: Expected species-level log-likelihood for each method, summed over all test routes and averaged across all species. The likelihood ratio compares each model to BRT, representing single-species SDMs. Sharing information among species with either of the neural net models improves the predictive likelihood more than twenty-fold for a typical species compared to BRT. Note also that BayesComm averages less than 1% of the machine learning methods' likelihoods because of its linearity assumption.

298 **Model comparison: community composition**

299 While making predictions about individual species observations is fairly straightforward
300 with this data set (since most species have relatively narrow breeding ranges), community
301 ecology is more concerned with co-occurrence and related patterns involving community
302 composition (Chase 2003). As expected, models that combined their single-species predictions
303 independently (including the neural network baseline) showed exponential decay in their
304 likelihoods as the number of species per prediction increased. The JSDMs (mistnet and
305 BayesComm) showed sub-exponential declines, since correlations reduce the number of
306 independent bits of information needed to make an accurate prediction. As a result, mistnet
307 became increasingly advantageous over independent combinations of single-species predictions
308 as the assemblage size increased (Figure 7). Mistnet's log-likelihood averaged 10.8 units higher
309 than BRT's for full assemblages of 368 species, corresponding to a 47000-fold improvement
310 in likelihood for a typical transect in the test set. Mistnet's ability to focus its predictions
311 on plausible combinations of species indicates that it has captured a great deal more of
312 the underlying ecological processes than existing SDM approaches. While some of this
313 improvement can be attributed to mistnet's overall tendency to make better predictions about

314 individual species (Table 1), the difference is mainly due to mistnet’s ability to keep ahead of
315 the combinatorial explosion of possible assemblages by exploiting correlations among species.

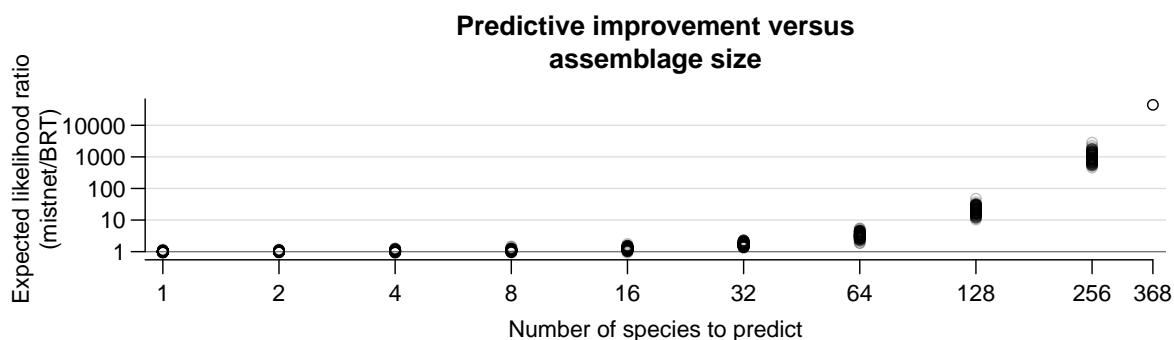


Figure 7: The likelihood ratio favoring mistnet over BRT grows super-exponentially with assemblage size. Each circle corresponds to a randomly-generated set of N species, where the value of N is indicated along the horizontal axis. Note the log scale on both axes.

316 **Comparison with BayesComm**

317 BayesComm’s ability to make out-of-sample predictions was severely limited by its assumption
318 that species respond linearly to climate variables, highlighting the the need for nonlinear
319 methods that can learn the functional forms of species’ responses to the environment. Adding
320 quadratic and interaction terms would have led to severe overfitting for many rare species,
321 and may still not have provided enough flexibility to compete with nonlinear techniques.

322 Even without the added complexity of nonlinear terms, the BayesComm model required
323 70,000 parameters, most of which served to to identify a distinct correlation coefficient
324 between a single pair of species. Tracing this many parameters through hundreds of Markov
325 chain iterations routinely caused BayesComm to exceed my machine’s 8 gigabytes of memory
326 and crash, even after the code was modified to reduce its memory footprint. Storing long
327 Markov chains over a dense, full-rank covariance matrix (as has apparently been done in all

328 other JSDMs to date) thus appears not to be a feasible strategy with large assemblages.

329 **Conclusion**

330 These results show conclusively that both linearity and independence are unwarranted
331 assumptions; either assumption can substantially impair our ability to model and understand
332 large assemblages. Linear JSDMs are not flexible enough, and models without latent random
333 variables cannot match the properties of real assemblages.

334 SDMs' failure to sufficiently consider correlations among has kept these models from explaining
335 and anticipating the full range of complex assemblages found in nature (Austin & Van Niel
336 2011). Mistnet's predictions are much more compatible with these sorts of complexities. In
337 particular, the model's predictions need not be unimodal, allowing the model to express
338 conditional predictions, such as that "the probability of observing a Redhead duck will be very
339 high if other wetland species are present, but very low otherwise." Such conditional predictions
340 are important because the available data will not always contain enough information to
341 narrow the possibilities down to a single assemblage type or a single group of species. In
342 such situations, stacked models will provide a false sense of security out-of-sample, leading
343 to bad decisionmaking and biased estimates of nature's variability. Mistnet provides better
344 confidence intervals that are much more likely to actually contain the observed values when
345 we look out-of-sample.

346 Mistnet can also identify some of the same similarities among species that a skilled biologist
347 would expect to find, which will be important for studying taxa that are more diverse and
348 harder to observe (such as microbes). For taxa on the frontier of our knowledge, a model

349 like mistnet could help guide the biologists to ask the best questions and organize their
350 understanding by suggesting which species have similar habitat requirements—even when
351 the factor controlling their occurrence are still unknown.

352 Unlike with stacked methods, one can read this straight out of mistnet’s coefficient tables
353 with no more difficulty than interpreting a Principal Components Analysis.

354 Mistnet’s ability to use asymmetrical or low-quality data sources to improve its predictions
355 should increase the value of low-effort data collection procedures such as short transects—
356 especially since these improvements can be incorporated without need for fitting a new model.

357 Future research should look for ways to use other forms of ecological knowledge about species
358 to impose some structure on models coefficients and nudge the models toward more biologically
359 reasonable predictions (Kearney & Porter 2009; Kissling *et al.* 2012). Such a research program
360 could also be useful in other areas of predictive ecology [[@pearse_predicting_2013](#)].

361 Finally, it should be noted that, while one *could* describe direct interactions among species
362 using latent variables (Ovaskainen *et al.* 2010; Golding 2013), existing JSDMs are not
363 particularly well-suited for learning about species interactions. Other models, such as Markov
364 random fields (Azaele *et al.* 2010), or ensembles of classifier chains (Yu *et al.* 2011) would
365 be much more appropriate for inferring coefficients related to species interactions, as they
366 include direct dependencies among species. Latent variable-based JSDMs, including mistnet,
367 are more appropriate for studies like this one at large spatial scales where direct species
368 interactions will tend to be weaker and most of the variation is driven by environmental
369 filtering and species’ range limits.

370 In conclusion, mistnet’s accuracy, as well as its flexibility to work with opportunistic samples
371 should make it useful for a variety of basic and applied contexts. Assemblage-level models,

372 such as mistnet, also have the potential to yield new biological insights. With charismatic and
373 well-studied species like North American birds, most models will mainly be telling information
374 that we already know. Still, mistnet's ability to capture useful information about axes of
375 variation among birds and to match preconceptions about which species co-occur due to
376 habitat variables may indicate that the model can teach us new things about taxa that are
377 harder to study.

378 **Acknowledgements**

379 This work benefitted greatly from discussions with A. Sih and his lab meeting group, M. L.
380 Baskett, R. J. Hijmans, R. McElreath, J. H. Thorne, M. W. Schwartz, B. M. Bolker, R. E.
381 Snyder, A. C. Perry, and C. S. Tysor. It was funded by a Graduate Research Fellowship
382 from the National Science Foundation, the UC Davis Center for Population Biology, and the
383 California Department of Water Resources. I gratefully acknowledge the field biologists that
384 collected the BBS data, as well as the US Geological Survey, Worldclim, and Cornell Lab of
385 Ornithology for their efforts and for making their data sets publicly available.

386 **Data Accessibility:**

- 387 • All data sets used here are freely downloadable from their original sources.
- 388 • The mistnet source code can be downloaded from <https://github.com/davharris/mistnet/>.
389 The easiest way to install the package is with the `devtools` package's `install_github`
390 command (e.g. `devtools::install_github("mistnet", "davharris")`).

391 • Some code has been improved since the analyses were run; however, the web site includes
392 a complete version history. The analyses in this paper had essentially all been run by the
393 commit at <https://github.com/davharris/mistnet/tree/1e2eaaeabf9b4b4360f19b00c0d06508578d7f15>.

394 References

395 Austin, M.P. & Van Niel, K.P. (2011) Improving species distribution models for climate
396 change studies: variable selection and scale. *Journal of Biogeography*, **38**, 1–8.

397 Azaele, S., Muneeppeerakul, R., Rinaldo, A. & Rodriguez-Iturbe, I. (2010) Inferring plant
398 ecosystem organization from species occurrences. *Journal of Theoretical Biology*, **262**, 323–
399 329.

400 Bahn, V. & McGill, B.J. (2007) Can niche-based distribution models outperform spatial
401 interpolation? *Global Ecology and Biogeography*, **16**, 733–742.

402 Bengio, Y. (2013) Deep Learning of Representations: Looking Forward. *Statistical Language
403 and Speech Processing* (eds & trans A.-H. Dediu), C. Martín-Vide), R. Mitkov), & B. Truthe),
404 pp. 1–37. Springer Berlin Heidelberg.

405 Calabrese, J.M., Certain, G., Kraan, C. & Dormann, C.F. (2014) Stacking species distribution
406 models and adjusting bias by linking them to macroecological models. *Global Ecology and
407 Biogeography*, **23**, 99–112.

408 Chase, J.M. (2003) Community assembly: when should history matter? *Oecologia*, **136**,
409 489–498.

410 Clark, J.S., Gelfand, A.E., Woodall, C.W. & Zhu, K. (2013) MORE THAN THE SUM OF
411 THE PARTS: FOREST CLIMATE RESPONSE FROM JOINT SPECIES DISTRIBUTION

412 MODELS. *Ecological Applications*.

413 De'ath, G. (2002) MULTIVARIATE REGRESSION TREES: A NEW TECHNIQUE FOR
414 MODELING SPECIES–ENVIRONMENT RELATIONSHIPS. *Ecology*, **83**, 1105–1117.

415 Dormann, C.F., Purschke, O., Márquez, J.R.G., Lautenbach, S. & Schröder, B. (2008)
416 Components of uncertainty in species distribution analysis: a case study of the great grey
417 shrike. *Ecology*, **89**, 3371–3386.

418 Eddelbuettel, D. & Francois, R. (2011) Rcpp: Seamless R and C++ Integration. , **40**, 1–18.

419 Eddelbuettel, D. & Sanderson, C. (2014) RcppArmadillo: Accelerating R with high-
420 performance C++ linear algebra. *Computational Statistics and Data Analysis*, **71**,
421 1054–1063.

422 Elith, J. & Leathwick, J.R. (2009) Species Distribution Models: Ecological Explanation and
423 Prediction Across Space and Time. *Annual Review of Ecology, Evolution, and Systematics*,
424 **40**, 677–697.

425 Elith, J., H. Graham*, C., P. Anderson, R., Dudík, M., Ferrier, S., Guisan, A., J. Hijmans,
426 R., Huettmann, F., R. Leathwick, J., Lehmann, A., Li, J., G. Lohmann, L., A. Loiselle, B.,
427 Manion, G., Moritz, C., Nakamura, M., Nakazawa, Y., McC. M. Overton, J., Townsend
428 Peterson, A., J. Phillips, S., Richardson, K., Scachetti-Pereira, R., E. Schapire, R., Soberón,
429 J., Williams, S., S. Wisz, M. & E. Zimmermann, N. (2006) Novel methods improve prediction
430 of species' distributions from occurrence data. *Ecography*, **29**, 129–151.

431 Elith, J., Leathwick, J.R. & Hastie, T. (2008) A working guide to boosted regression trees.
432 *Journal of Animal Ecology*, **77**, 802–813.

433 Ferrier, S., Manion, G., Elith, J. & Richardson, K. (2007) Using generalized dissimilarity

- 434 modelling to analyse and predict patterns of beta diversity in regional biodiversity assessment.
435 *Diversity and Distributions*, **13**, 252–264.
- 436 Golding, N. (2013) PhD thesis: Mapping and understanding the distributions of potential
437 vector mosquitoes in the UK: New methods and applications.
- 438 Golding, N. & Harris, D.J. (2014) *BayesComm: Bayesian Community Ecology Analysis*.
- 439 Guisan, A. & Rahbek, C. (2011) SESAM – a new framework integrating macroecological and
440 species distribution models for predicting spatio-temporal patterns of species assemblages.
441 *Journal of Biogeography*, **38**, 1433–1444.
- 442 Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G. & Jarvis, A. (2005) Very high resolution
443 interpolated climate surfaces for global land areas. *International Journal of Climatology*, **25**,
444 1965–1978.
- 445 Hong, Y. (2013) *Poibin: The Poisson Binomial Distribution*.
- 446 Hutchinson, R.A., Liu, L.-P. & Dietterich, T.G. (2011) Incorporating boosted regression
447 trees into ecological latent variable models. *Twenty-Fifth AAAI Conference on Artificial*
448 *Intelligence* pp. 1343–1348.
- 449 Kearney, M. & Porter, W. (2009) Mechanistic niche modelling: combining physiological and
450 spatial data to predict species' ranges. *Ecology letters*, **12**, 334–350.
- 451 Kissling, W.D., Dormann, C.F., Groeneveld, J., Hickler, T., Kühn, I., McNerny, G.J.,
452 Montoya, J.M., Römermann, C., Schiffers, K., Schurr, F.M., Singer, A., Svenning, J.-C.,
453 Zimmermann, N.E. & O'Hara, R.B. (2012) Towards novel approaches to modelling biotic
454 interactions in multispecies assemblages at large spatial extents. *Journal of Biogeography*, **39**,
455 2163–2178.

- 456 Latimer, A.M., Banerjee, S., Sang Jr, H., Mosher, E.S. & Silander Jr, J.A. (2009) Hierarchical
457 models facilitate spatial analysis of large data sets: a case study on invasive plant species in
458 the northeastern United States. *Ecology Letters*, **12**, 144–154.
- 459 Leathwick, J.R., Elith, J. & Hastie, T. (2006) Comparative performance of generalized
460 additive models and multivariate adaptive regression splines for statistical modelling of
461 species distributions. *Ecological modelling*, **199**, 188–196.
- 462 Leathwick, J.R., Rowe, D., Richardson, J., Elith, J. & Hastie, T. (2005) Using multivariate
463 adaptive regression splines to predict the distributions of New Zealand’s freshwater diadromous
464 fish. *Freshwater Biology*, **50**, 2034–2052.
- 465 McNerny, G.J. & Purves, D.W. (2011) Fine-scale environmental variation in species distri-
466 bution modelling: regression dilution, latent variables and neighbourly advice. *Methods in*
467 *Ecology and Evolution*, **2**, 248–257.
- 468 Murphy, K.P. (2012) *Machine Learning: A Probabilistic Perspective*. The MIT Press.
- 469 Neal, R.M. (1992) Connectionist learning of belief networks. *Artificial Intelligence*, **56**,
470 71–113.
- 471 Neal, R.M. & Hinton, G.E. (1998) A view of the EM algorithm that justifies incremental,
472 sparse, and other variants. *Learning in graphical models* pp. 355–368. Springer.
- 473 Orr, G.B. & Müller, K.-R. (1998) *Neural Networks: Tricks of the Trade*. Springer-Verlag.
- 474 Ovaskainen, O., Hottola, J. & Siitonen, J. (2010) Modeling species co-occurrence by mul-
475 tivariate logistic regression generates new hypotheses on fungal interactions. *Ecology*, **91**,
476 2514–2521.
- 477 O’Hara, R.B. (2005) Species richness estimators: how many species can dance on the head of

478 a pin? *Journal of Animal Ecology*, **74**, 375–386.

479 Pellissier, L., Espíndola, A., Pradervand, J.-N., Dubuis, A., Pottier, J., Ferrier, S. & Guisan,
480 A. (2013) A probabilistic approach to niche-based community models for spatial forecasts of
481 assemblage properties and their uncertainties. *Journal of Biogeography*, **40**, 1939–1946.

482 Pollock, L.J., Tingley, R., Morris, W.K., Golding, N., O’Hara, R.B., Parris, K.M., Vesk, P.A.
483 & McCarthy, M.A. (2014) Understanding co-occurrence by modelling species simultaneously
484 with a Joint Species Distribution Model (JSDM). *Methods in Ecology and Evolution*.

485 Ridgeway, G. (2013) *Gbm: Generalized Boosted Regression Models*.

486 Rumelhart, D.E., Hinton, G.E. & Williams, R.J. (1986) Learning representations by back-
487 propagating errors. *Nature*, **323**, 533–536.

488 Sanderson, C. (2010) Armadillo: An Open Source C++ Linear Algebra Library for Fast
489 Prototyping and Computationally Intensive Experiments.

490 Sauer, J.R., Hines, J.E., Fallon, J., Pardieck, K.L., Ziolkowski Jr, D.J. & Link, W.A. (2011)
491 The North American breeding bird survey, results and analysis 1966-2011. *Version 2011.0*.

492 Tang, Y. & Salakhutdinov, R. (2013) Learning Stochastic Feedforward Neural Networks.
493 *Advances in Neural Information Processing Systems 26* (eds & trans C.J.C. Burges), L.
494 Bottou), M. Welling), Z. Ghahramani), & K.Q. Weinberger), pp. 530–538.

495 Venables, W.N. & Ripley, B.D. (2002) *Modern Applied Statistics With S*. Springer, New York.

496 Vincent, P., Larochelle, H., Lajoie, I., Bengio, Y. & Manzagol, P.-A. (2010) Stacked denoising
497 autoencoders: Learning useful representations in a deep network with a local denoising
498 criterion. *The Journal of Machine Learning Research*, **9999**, 3371–3408.

499 Welsh, A.H., Lindenmayer, D.B. & Donnelly, C.F. (2013) Fitting and Interpreting Occupancy

500 Models. *PLoS ONE*, **8**, 52015.

501 Wing, M.K.C. from J., Weston, S., Williams, A., Keefer, C., Engelhardt, A. & Cooper, T.
502 (2013) *Caret: Classification and Regression Training*.

503 Wisz, M.S., Pottier, J., Kissling, W.D., Pellissier, L., Lenoir, J., Damgaard, C.F., Dormann,
504 C.F., Forchhammer, M.C., Grytnes, J.-A., Guisan, A., Heikkinen, R.K., Høye, T.T., Kühn,
505 I., Luoto, M., Maiorano, L., Nilsson, M.-C., Normand, S., Öckinger, E., Schmidt, N.M.,
506 Termansen, M., Timmermann, A., Wardle, D.A., Aastrup, P. & Svenning, J.-C. (2013)
507 The role of biotic interactions in shaping distributions and realised assemblages of species:
508 implications for species distribution modelling. *Biological Reviews*, **88**, 15–30.

509 Yu, J., Wong, W.-K., Dietterich, T., Jones, J., Betts, M., Frey, S., Shirley, S., Miller,
510 J. & White, M. (2011) Multi-label Classification for Multi-Species Distribution Modeling.
511 *Proceedings of the 28th International Conference on Machine Learning*.