

1 Combining molecular gut content analysis and functional response models shows 2 how body size affects prey choice in soil predators

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16

17 **Summary**

18 1. Predator-prey interactions are a core concept of animal ecology and functional response
19 models provide a powerful tool to predict the strength of trophic links and assess motives for
20 prey choice. However, due to their reductionist set-up, these models may not display field
21 conditions, possibly leading to skewed results.

22 2. We tested the validity of functional response models for multiple prey by comparing them
23 with empirical data from DNA-based molecular gut content analysis of two abundant and
24 widespread macrofauna soil predators, lithobiid and geophilomorph centipedes.

25 3. We collected soil and litter dwelling centipedes, screened their gut contents for DNA of nine
26 abundant decomposer and intraguild prey using specific primers and tested for different prey
27 and predator traits explaining prey choice. In order to calculate the functional response of
28 same predators, we used natural prey abundances and functional response parameters from
29 published experiments and compared both approaches.

30 4. Molecular gut content results showed that prey choice of centipedes is driven by predator
31 body size and prey identity. Results of functional response models significantly correlated
32 with results from molecular gut content analysis for the majority of prey species.

33 5. Overall, the results suggest that functional response models are a powerful tool to predict
34 trophic interactions in soil, however, species-specific traits have to be taken into account to
35 improve predictions.

36

37 **Keywords**

38 allometric scaling, food webs, generalist predator, molecular prey detection, predator-prey
39 interaction

40

41 ***Introduction***

42 Analysis of consumer-resource interactions is key to understand the structure and dynamics of food
43 webs, eventually explaining composition, stability and development of communities and ecological
44 processes coupled with them. Depending on the specific problem and scale of feeding interactions,
45 21st century ecologists are in the comfortable position to select from a broad spectrum of methods,
46 from field observations to molecular tracking of nutrients and DNA in the consumer's body.
47 Measuring the functional response, i.e. the intake rate of a consumer (hereafter referred to as
48 *predator*) as a function of food resource (hereafter referred to as *prey*) density has been
49 demonstrated to be a powerful method not only to track feeding interactions but also to assess the
50 interaction strength (Holling 1959). Based on a small set of parameters including densities and body
51 sizes of prey and predator, functional response models allow predicting general patterns and
52 mechanisms of trophic interactions in very different systems, spanning from *Daphnia* water fleas
53 feeding on phytoplankton to wolf packs preying on moose (Sarnelle & Wilson 2008, Messier 1994).
54 The approach allows investigating feeding interactions on a large scale, and can be modified to
55 include changes in body size (Hansen *et al.* 1997, Pawar *et al.* 2012, Rall *et al.* 2012), ambient
56 temperature (Hansen *et al.* 1997, Englund *et al.* 2011, Rall *et al.* 2012) as well as habitat structure
57 (Hauzy *et al.* 2010, Kalinkat *et al.* 2013a, Kalinkat & Rall 2015).

58 The simplicity of functional responses, however, may come at the cost of accuracy. Functional
59 response curves, in particular those of invertebrate species, are typically based on single-prey-
60 predator laboratory feeding trials, which lack many characteristics of natural settings. Among these
61 potentially important characteristics are habitat structure, presence of competitors and alternative
62 prey as well as different physiological states of prey and predator (e.g. sick prey). Thus, functional
63 response models based on idealized laboratory settings may be of limited use to predict feeding

64 interactions in the field. Here, other methods apply, which allow us to analyse the character and
65 intensity of predator-prey interactions under natural settings directly in the field.

66 DNA-based molecular gut content analysis offers a state-of-the-art technique (Pompanon *et al.* 2012;
67 Traugott *et al.* 2013) to identify trophic links under challenging conditions, from sea shores (Peters
68 *et al.* 2014), over arctic tundra (Wirta *et al.* 2015) to arable soils (Wallinger *et al.* 2014). Using
69 specifically designed PCR assays targeting prey DNA in a predator's gut, species-specific trophic
70 interactions can be tracked even several days after the feeding event, allowing unravelling of
71 trophic links in unprecedented detail (Eitzinger *et al.* 2013). Hence, molecular gut content analysis
72 allows to empirically assess complex trophic interactions in the field and provides the opportunity
73 to evaluate functional response models under natural conditions.

74 We adopted this approach for the first time using for a soil predator-prey system in European
75 deciduous forests. Here, we analysed the predation frequency on extra- and intraguild prey of
76 centipedes (Chilopoda, Myriapoda), widespread generalist predators in the litter and soil layers of
77 temperate forests (Lewis 1981; Poser 1988) using predictive models from functional response
78 experiments and compare these with empirically quantified trophic links using molecular gut
79 content analysis. By the combined use of both approaches we aimed at achieving an integrated view
80 of food web interactions in complex systems and evaluate the suitability and effectiveness of the
81 approaches for analysing trophic interactions.

82 Centipedes, in particular lithobiid (Lithobiidae) and geophilomorph (Geophilomorpha) species prey
83 on a variety of prey taxa including Collembola, Diptera larvae and Lumbricidae (Günther *et al.*
84 2014). Lithobiids predominantly colonize the litter layer and perform a sit-and-wait strategy of prey
85 capture, whereas geophilomorph centipedes are active hunters in crevices of the mineral soil (Lewis
86 1981; Poser 1988; Eitzinger *et al.* 2013). Prey capture of centipedes specifically depends on body
87 size indicating an allometric relationship between predator and prey size (Schneider *et al.* 2012,
88 Günther *et al.* 2014). Typically, small predators have narrow diets while large predators feed on a
89 wider range of prey including higher trophic level taxa, i.e. intraguild prey (Woodward & Hildrew

90 2002; Riede *et al.* 2011). Body-size dependent prey-switching, coupled with feeding on intraguild
91 prey may be a key factor reducing dietary niche overlap (Woodward & Hildrew 2002). Moreover,
92 this might explain coexistence of different centipede species and other predators in forest soils.
93 Studies employing functional response models suggested that the body size acts as a supertrait,
94 explaining most of the variance in predator-prey interactions in soil systems (Vucic-Pestic *et al.*
95 2010, Kalinkat *et al.* 2013b). Hence, allometry-based functional response models may be applied to
96 many different predator-prey-interactions.

97 Based on the generalised allometric functional response model by Kalinkat *et al.* (2013b), we
98 calculated body-size dependent trophic interaction strength of centipede predators as a function of
99 natural abundances of different prey groups present in soil of unmanaged beech forests in central
100 Germany. We then analysed the gut content of field-collected centipedes from the same forests
101 using nine group- and five species-specific primers for DNA of extra- and intraguild prey taxa. We
102 hypothesized that (i) feeding interactions of centipedes are driven by predator-prey body-size ratios
103 rather than by taxonomy, and that (ii) functional response models can correctly predict actual
104 feeding interactions in a complex system such as soil.

105

106 ***Materials and Methods***

107 *Sampling*

108 Invertebrate predators were collected in four unmanaged beech forests (> 120 years old) within the
109 national park Hainich near Mülverstedt (Thuringia, Germany). Each study plot spanned 100 × 100
110 m and formed part of the Biodiversity Exploratories, an integrated biodiversity project (Fischer *et al.*
111 2010). To investigate trophic links during periods of maximum invertebrate activity, we sampled
112 animals in autumn and spring/early summer, each represented by four sampling dates (October 8,
113 20 and 28 and November 3, 2009; June 15, 24 and 29 and July 8, 2010). Centipedes were collected
114 by sieving litter, transferred individually to 1.5 mL microcentrifuge tubes and placed immediately at
115 -20 °C.

116 To record the species spectrum and abundance of prey organisms, two large (20 cm diameter, 10 cm
117 deep) and two small (5 cm diameter, 10 cm deep) soil cores per plot were taken in May of 2008 and
118 2011 (Klarner *et al.* 2014). Animals were extracted using a high gradient extractor (Kempson *et al.*
119 1963), stored in 75% ethanol and identified to the species level (except dipteran larvae).
120 Additionally, lumbricids were collected by hand after application of mustard solution (Eisenhauer *et*
121 *al.* 2008). Average densities between the two sampling dates were taken to represent prey density at
122 the sampling dates of centipedes. We assume this to be justified as soil arthropod composition and
123 density changes little between years (Bengtsson 1994).

124 A total of 532 field-caught *Lithobius* spp. and 65 geophilomorph centipedes were identified to
125 species level using the keys of Eason (1964) and Latzel (1880). Further, we determined
126 developmental stages and body length of each individual. Body mass of lithobiid centipedes was
127 calculated using the following equation:

128

$$\log_{10}M = 2.32784 * \log_{10}L - 1.24015 \quad (1)$$

129

130 where M is the fresh body mass and L the body length of individuals. The equation is based on body
131 length - body mass relationship of 560 lithobiid individuals used in laboratory studies (Eitzinger *et*
132 *al.* 2014). Based on body size of collected specimens from the study site the body mass of
133 geophilomorph centipedes and all prey taxa was calculated using formulas given in Gowing and
134 Recher (1984) and Mercer (2001). Body mass (for predator and prey) and prey abundance were
135 \log_{10} -transformed prior to statistical analyses.

136

137 *DNA extraction*

138 Centipedes were subjected to CTAB-based DNA-extraction protocol (Juen & Traugott 2005) with
139 modifications given in Eitzinger *et al.* (2013). DNA extracts were purified using GeneClean Kit (MP
140 Biomedicals, Solon, OH, USA). To test for DNA carry-over contamination a blank control was

141 included within a batch of 47 individuals. None was found when testing all extracts for false
142 negatives and false positives, using the universal invertebrate primer pair LCO1490/HCO2198
143 (Folmer *et al.* 1994) amplifying a *c.* 700 bp fragment of the cytochrome *c* oxidase subunit I gene
144 (*COI*). Each 10 μ L PCR contained 5 μ L PCR SuperHot Mastermix (2 \times), 1.25 mM MgCl₂ (both
145 Geneaxxon, Ulm, Germany), 0.5 μ L bovine serum albumin (BSA, 3%; Roth, Karlsruhe, Germany),
146 0.5 μ M of each primer and 3 μ L of DNA extract. PCR cycling conditions were 95 °C for 10 min
147 followed by 35 cycles at 95 °C for 30 s, 48 °C for 30 s, 72 °C for 90 s and a final elongation at 72
148 °C for 10 min. PCR products were separated in 1% ethidium bromide-stained agarose gels and
149 visualized under UV-light.

150

151 *Screening predators for prey DNA*

152 DNA extracts were screened for five extraguild and three intraguild prey (i.e. other predators) taxa
153 using group-specific primers. PCR mixes and thermocycling conditions were the same as above
154 only differing in applied primers, an elongation step at 72 °C for 45 s and the primer pair-specific
155 annealing temperature. Geophilomorph centipedes additionally were tested for consumption of
156 *Lithobius* spp. intraguild prey. All predator samples scoring positive for Collembola were
157 subsequently tested for abundant Collembola species *Ceratophysella denticulata*,
158 *Folsomia quadrioculata*, *Lepidocyrtus lanuginosus*, *Protaphorura armata* and *Pogonognathellus*
159 *longicornis* (for primers and annealing temperature see Table S1, Supporting Information).

160 Specificity of the PCR assays was warranted by testing against a set of up to 119 non-target
161 organisms (Eitzinger *et al.* 2013). PCR products were separated using the capillary electrophoresis
162 system QIAxcel (Qiagen, Hilden, Germany); fragments of the expected size and a relative
163 fluorescent value ≥ 0.1 RFU were scored as positive. PCR products showing no result were re-
164 tested once.

165

166 *Statistical analysis*

167 To compare prey DNA detection rates between predator taxa at the $P < 0.05$ level, 95% tilting
168 confidence intervals (CI; Hesterberg *et al.* 2003) were calculated by 9999 bootstrap resamples using
169 s-plus 8.0 (Insightful Corporations, Seattle, WA, USA).

170 Relationships between prey detection rates and predator identity, predator body mass, square of
171 predator body mass, predator development stage (immature or adult), prey identity, prey body mass
172 and prey abundance were analysed by generalized linear models (GLM) in R 2.12.2 (R
173 Development Core Team 2011) using the function `glm {stats}`. Based on Akaike information
174 criterion (AIC) we selected the most parsimonious model (Burnham and Anderson 2004). Prey
175 DNA detection data was coded as binary (prey DNA present or absent).

176 A multi-prey functional response model was used to calculate feeding rates F of centipede predator
177 i and prey j when alternative prey organisms k are present (note that k includes j ; Kalinkat *et al.*
178 2011):

179

$$F_{ij} = \frac{b_{ij} N_j^{1+q_{ij}}}{1 + \sum_{k=1}^{k=n} b_{ik} h_{ik} N_k^{1+q_{ik}}} \quad (2)$$

180

181 with N the prey density (individuals/m²), n the number of alternative prey items, h [s] the handling
182 time (time for killing, ingesting and digesting prey), b the capture coefficient and q the scaling
183 exponent that converts hyperbolic type-II ($q = 0$) into sigmoid type-III ($q > 0$) functional responses
184 (Kalinkat *et al.* 2013b). We used prey-specific body masses [g] and values for generalised
185 allometric functional response (Kalinkat *et al.* 2013b) to calculate b , h and q for each of the eight
186 most important prey groups and added plot-specific prey density data (see above). The relative
187 proportion of each of the eight prey-specific feeding rates per plot and for all plots combined was
188 measured, resulting in prey-specific feeding ratios, F_{rel} :

$$F_{rel_{ij}} = \frac{F_{ij}}{\sum_{k=1}^{k=n} F_{ik}} \quad (3)$$

189

190 Additionally, we related both prey detection and feeding ratios to body size of predators.

191 For each prey group, we then compared the relative proportion of prey in the predator's diet with

192 the proportion of prey-DNA-positive predators using Pearson's correlation coefficient in R 2.12.2.

193

194 **Results**

195 *Centipede community*

196 Among the 597 centipedes collected during the sampling periods, nine species of lithobiid

197 (*Lithobius aulacopus*, *L. crassipes*, *L. curtipes*, *L. dentatus*, *L. melanops*, *L. muticus*, *L. mutabilis*,

198 *L. nodulipes* and *L. piceus*) and three species of geophilomorph centipedes (*Geophilus* sp.,

199 *Schendyla nemorensis*, *Strigamia acuminata*) of both sexes and different developmental stages were

200 identified. Body sizes / body masses ranged between 2-18 mm / 0.28 - 48.07 mg in lithobiids and 8-

201 47 mm / 1.58 - 16.70 mg in geophilomorph centipedes.

202

203 *Prey DNA screening*

204 A total of 532 *Lithobius* spp. and 65 geophilomorph centipedes collected at the eight sampling dates

205 were tested for DNA of five and four extra- and intraguild prey taxa, respectively. Per sampling

206 date 41-91 *Lithobius* spp. and 4-12 geophilomorph centipedes were investigated.

207 DNA of each of the prey organisms tested could be detected in at least one predator individual.

208 Lithobiid predators were significantly more often tested positive for Collembola than for any other

209 prey group (Fig. 1A). Detection rates of Diptera and Lumbricidae were significantly higher than

210 those of other extraguild prey, such as Isopoda and Oribatida. Intraguild prey formed only a minor

211 fraction of lithobiid prey: detection frequencies of Mesostigmata were followed by Staphylinidae

212 and Araneida. In 69 predators two or three prey taxa were detected in one individual. The lithobiids

213 which tested positive with the general Collembola primers ($n=141$) consumed significantly more

214 *Folsomia quadrioculata* than any other of the four tested Collembola species (Fig. 1B).

215 In geophilomorph centipedes extraguild prey, such as Collembola and Diptera, were most often
216 detected followed by Lumbricidae, Isopoda and Oribatida (Fig. 1C). Detection rates for intraguild
217 prey were highest for Staphylinidae, followed by Araneida and Mesostigmata. None of the five
218 Collembola species could be detected in geophilomorph centipedes tested positive for Collembola.
219 In 14 geophilomorph centipedes two or three prey taxa were detected simultaneously.

220

221 *Factors influencing prey consumption*

222 We selected the most parsimonious model based on AIC comparison, thereby rejecting models
223 containing factors centipede identity and development stage. Overall, lithobiid feeding was
224 significantly affected by prey identity and predator body mass (Table 1), with preferences of
225 predators for certain prey sizes. For Collembola and Lumbricidae prey, the probability of prey
226 detection in relation to predator body mass followed a unimodal curve, peaking at body masses of
227 6.3 mg and 4.9 mg, respectively (Fig. 2). In contrast, detection probability of Diptera prey increased
228 exponentially with predator body mass, indicating that Diptera are increasingly fed on by larger
229 lithobiids while being rejected by smaller ones. Prey detection probabilities for Oribatida,
230 Mesostigmata, Staphylinidae and Isopoda, despite being generally low, also increased with predator
231 body mass, with the curve flattening at 25, 60, 62 and 69 mg predator body mass, respectively.
232 Feeding on another intraguild prey, Araneida, however, showed a steady decrease with body mass.

233 Feeding of geophilomorph centipedes varied with prey identity, predator body mass (including
234 square of predator body mass) and prey abundance (Table S2, Supporting Information). In contrast
235 to lithobiids, detection rates followed a unimodal curve for each of the prey taxa (Fig. S3,
236 Supporting Information).

237

238 *Prey proportions according to functional response models*

239 According to the functional response models, Collembola, Oribatida and Mesostigmata accounted

240 for most of the diet of lithobiid and geophilomorph centipedes, showing a bimodal relationship with
241 predator body mass (Fig. 3; Fig. S4, Supplementary Information). Diptera and Isopoda prey
242 portions increased slightly at highest body masses, while other prey did not form part of the diet of
243 the centipede predators.

244

245 *Comparison of functional response models with molecular gut content analysis*

246 The relative proportion of a specific prey in the centipedes' diet, as calculated by functional
247 response models and the proportion of prey-DNA-positive centipedes, as calculated from the
248 molecular gut content analysis significantly correlated for each of the prey group (Pearson
249 correlation coefficient, $P < 0.001$; Fig. 4). While we found a positive correlation for the five prey
250 groups Collembola, Diptera, Isopoda, Oribatida and Staphylinidae the other three prey groups had a
251 negative relationship. In geophilomorph centipedes, only correlations with Lumbricidae,
252 Staphylinidae and Collembola were significantly positive ($P < 0.05$), while Mesostigmata showed a
253 significant negative correlation ($P < 0.001$). The other prey groups did not show any significant
254 correlation.

255

256 **Discussion**

257 The present study provides the first strong evidence that generalised allometric functional response
258 models are an appropriate method to assess predator-prey interactions in complex systems, which
259 include high levels of habitat structure, competitors and alternative prey. We tested if these models
260 correctly predict relative feeding strength of generalist predators in a species- rich soil system by
261 comparing with empirically quantified prey proportions in the diet of predators as indicated by
262 molecular gut content analysis. Model and empirical data positively correlated in five of eight tested
263 prey species, suggesting high explanatory power of the functional response models. Corroborating
264 previous studies employing functional response models (Vucic-Pestic *et al.* 2010, Rall *et al.* 2011),
265 we also empirically showed that 'predator body size' and 'prey identity' are two major drivers of

266 prey capture in soil-dwelling predators.

267 The functional response models predicted high feeding rates of both lithobiid and geophilomorph
268 centipedes on mesofaunal prey including Collembola, oribatid and mesostigmatid mites. A
269 combination of high prey abundance, facilitating high encounter rates, and an optimal predator-prey
270 body mass relationship allows the predator to forage on a maximum of prey individuals with a
271 minimum of handling time, thereby reducing energetic costs (Aljetlawi *et al.* 2004, Brose *et al.*
272 2008, Vucic-Pestic *et al.* 2010). Results of the model used in this study allowing to track shifts from
273 a hyperbolic (type-II) to a sigmoid (type-III) functional response suggest that with increasing
274 predator body mass relative feeding rates follow a roller-coaster-pattern, peaking at the respective
275 optimal body-mass ratios.

276 Feeding rates on other than mesofauna prey, however, were consistently low, only increasing
277 slightly in large lithobiids and geophilomorph centipedes. As metabolism increases with body size,
278 consumers require a higher energy uptake which is covered by the ingestion of more prey biomass,
279 i.e. more small prey or larger prey individuals (Kalinkat *et al.* 2011). This is in line with earlier
280 studies (Woodward & Hildrew 2002, Kalinkat *et al.* 2011) showing that with the increase in
281 predator body mass prey preference shifts towards bigger prey while at the same time still being
282 able to exploit small prey.

283 Results from the molecular gut content analysis corroborate the body-size dependent change in prey
284 capture in the mathematical model. Centipedes exhibit unimodal feeding responses for 75% of the
285 studied prey taxa, with large predator individuals more frequently feeding on more prey taxa than
286 small predators. Analogous to the model, mesofauna taxa constitute the most important prey except
287 for oribatid mites, which were detected in only 0.94% and 4.62% of the tested lithobiid and
288 geophilomorph centipedes, respectively. While their high abundances and optimal body size
289 suggest them to be ideal prey in the model, other traits, particularly their hard exoskeleton and toxic
290 secretions seem to be effective defence traits, explaining why they were only rarely consumed

291 (Peschel *et al.* 2006, Heethoff *et al.* 2011).

292 Collembola-DNA was detected in most centipedes, particularly medium-sized individuals.

293 Collembola are abundant in virtually any terrestrial ecosystem and of high nutritional value thereby

294 functioning as major prey for a wide range of predators in soil throughout the globe (Marcussen *et*

295 *al.* 1999, Bilde *et al.* 2000, Oelbermann *et al.* 2008). Using a taxonomic-allometric model, Rall *et al.*

296 (2011) calculated an optimal body mass ratio of 649 between the lithobiid centipede species

297 *L. forficatus* and the Collembola species *Heteromurus nitidus*. In our study a similar ratio applied to

298 *L. lanuginosus* and *P. armata*, the second and third most often detected Collembola prey species of

299 lithobiid centipedes, respectively.

300 Lumbricidae, on the other hand, were a far more important prey than expected from the functional

301 response model. Lumbricidae for long have been regarded as major prey of centipedes, in particular

302 geophilomorph species (Lewis 1981), however, their low abundances and big size (as compared to

303 mesofauna taxa) make them an unlikely prey in our allometric model. Using their poison claws,

304 however, centipedes kill prey far below the optimal body-mass ratio (Eason 1964), and this resulted

305 in underestimation of the importance of earthworms as prey of centipedes.

306 Interestingly, we found a strong increase in feeding on Diptera larvae with lithobiid body size, even

307 stronger than predicted by the model. In combination with reduced feeding on other important prey,

308 Collembola and Lumbricidae, this suggests prey switching towards this abundant prey of high

309 nutritional value (Oelbermann & Scheu 2002). Prey switching has been reported in many studies

310 (Hohberg & Traunspurger 2005, Petchey *et al.* 2008) and its frequency is increasing if predators

311 become larger, presumably due to a combination of effects of habitat structure and optimal foraging

312 processes (Murdoch & Oaten 1975, Kalinkat *et al.* 2013a) as described as follows:

313 Habitat structure modifies lithobiid feeding by allowing small prey such as Collembola but also

314 small Lumbricidae, to take refuge from predation, forcing particularly large predator individuals to

315 focus on more accessible prey dwelling in the upper litter layer (Günther *et al.* 2014).

316 Simultaneously, larger predators have higher energetic demands forcing them to hunt for larger prey,
317 i.e. bigger individuals of species already feeding upon or a new, larger species. Higher energetic
318 costs of killing, ingesting and digesting (i.e. 'handling time') prey, such as tipulid fly larvae or large
319 earthworms are more easily balanced by the prey's high nutritional value. However, the results
320 suggest that to meet their nutritional and energetic demands, large lithobiid centipedes cannot be too
321 selective in their prey choice: their spectrum still includes mesofauna prey and also encompasses
322 intraguild prey, such as spiders and staphylinid beetles. These results confirm earlier studies
323 showing that the prey spectrum of predators broadens with predator body size, suggesting that large
324 predators exploit prey communities more efficiently (Cohen *et al.* 1993; Woodward & Hildrew
325 2002). On the other hand our findings argue against suggestions that at high density of extraguild
326 prey intraguild predation is negligible (Halaj & Wise 2002, Eitzinger & Traugott 2011). Further, the
327 results contradict findings that the role of intraguild predation is reduced in well-structured habitats
328 providing refuge for intraguild prey (Finke & Denno 2002, Janssen *et al.* 2007).

329

330 *Conclusions*

331 The present study, for the first time, investigated the impact of predator body size and prey
332 abundance on predator consumption using two different approaches, functional response models
333 and molecular gut content analysis. Both methods proved to be useful to study trophic interactions,
334 the first one to analyse feeding strengths based on body size ratios and abundances, the latter to
335 examine predator-prey interactions of individual predators on small scale. While these methods
336 measure different parameters, i.e. feeding rate and prey DNA detection frequency, respectively,
337 results of the present study suggest that they complement each other allowing to prove and extend
338 theoretical predictions under natural settings. Therefore, combining these two techniques may
339 ultimately allow uncovering the structure of food webs in particular those in opaque habitats
340 colonized by minute animal species.

341 Combining functional responses with molecular gut content analyses and including predator-prey
342 body size ratios we are able to explain the majority of feeding interactions in belowground systems.
343 This emphasizes that allometric constraints override taxonomic constraints in structuring soil food
344 webs. Further, in contrast to food webs in simply structured habitats, such as aquatic systems, prey
345 abundance did not affect prey ingestion rates in this soil system, pointing to the importance of prey
346 identity effects as driving factors. Therefore, for improving the effectiveness of allometric
347 functional response models in predicting food web interactions in the field, additional traits of prey
348 species, such as defence characteristics, have to be included.

349

350

351 *Author's contributions*

352 B.E. and B.C.R. conceived the ideas and designed methodology with contributions from M.T. and
353 S.S.; B.E. collected the data, and B.E. and B.C.R. analysed the data; B.E. drafted the manuscript.
354 All authors contributed to later drafts and gave final approval for publication.

355

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369

370 ***Data accessibility***

371 If the manuscript gets accepted, the authors will make data available on the Dryad Digital
372 Repository (www.datadryad.org).

373

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503 **Legends to figures**

504

505 **Fig 1.** Prey detection rates of lithobiid (**A**; n= 532) and geophilomorph centipedes (**C**; n=65) sampled in
506 autumn 2009 and spring 2010. Specimens tested positive for Collembola prey (**B**; n=141) further were tested
507 for Collembola prey species. Error bars indicated 95% confidence intervals and letters denote significant
508 differences in DNA detection rates at $P < 0.05$.

509

510 **Fig 2.** Body-size-dependent probability of positive prey-DNA detection of eight taxa in lithobiid centipedes
511 (n= 532) sampled in autumn 2009 and spring 2010. Rugs on top and bottom of each diagram display single
512 data points with values 1 or 0.

513

514 **Fig 3.** Body-size-dependent proportion of eight prey taxa in the diet of centipede predators as based on the
515 functional response model using abundance and body-size data of invertebrates sampled in autumn 2009 and
516 spring 2010. Upper and lower limit indicate highest and lowest diet proportion in the four forest sites.

517

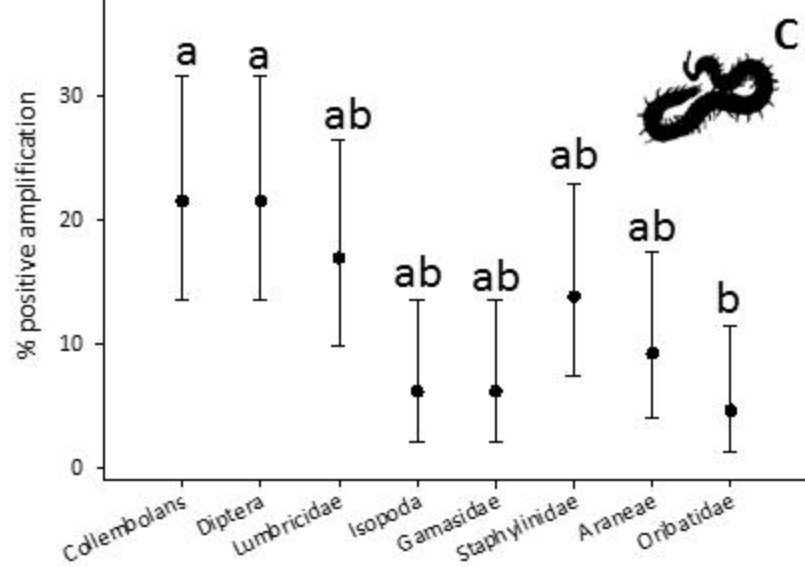
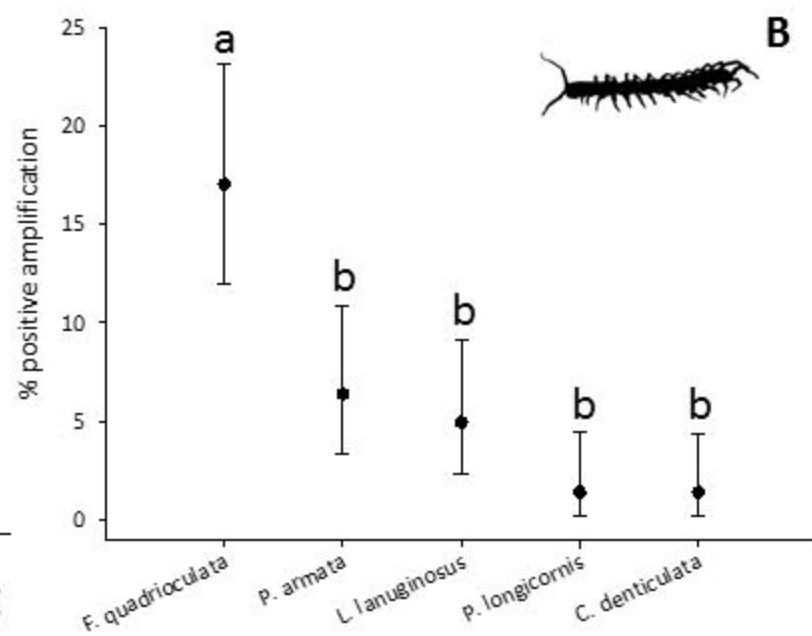
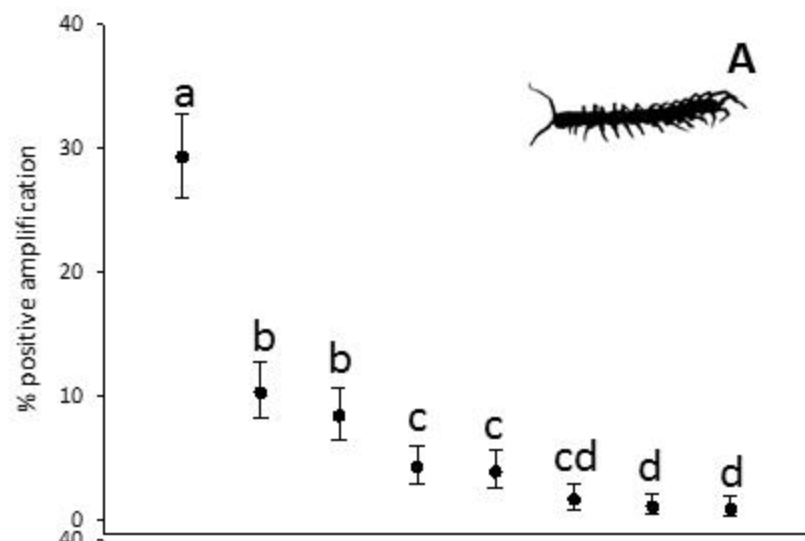
518 **Fig 4.** Pearson correlation coefficient between the relative proportion of prey in the centipede's diet (as
519 calculated by functional response models) and the proportion of prey-DNA-positive tested centipede
520 *Lithobius* sp. (based on molecular gut content data) for each of the eight main prey groups.

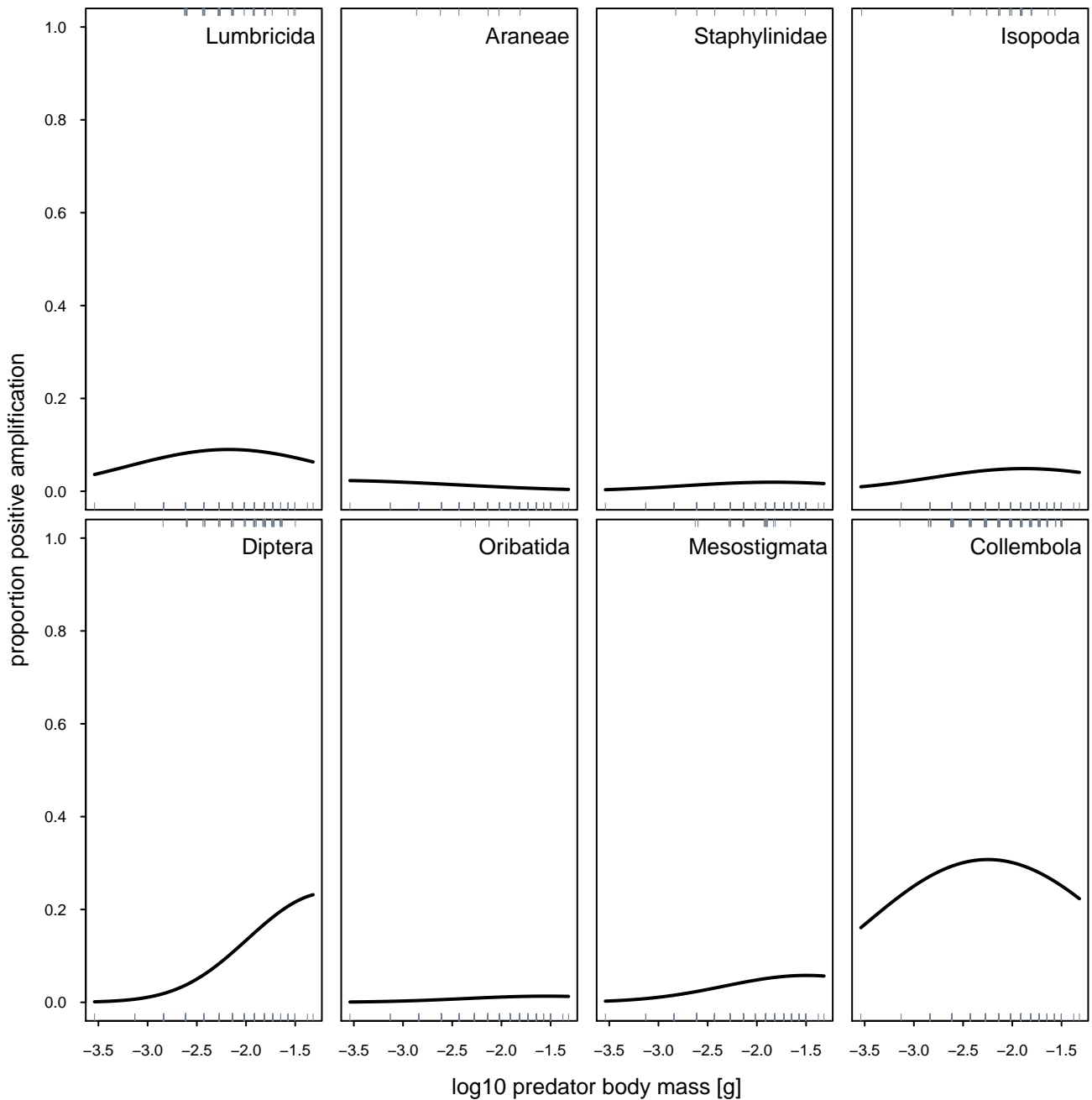
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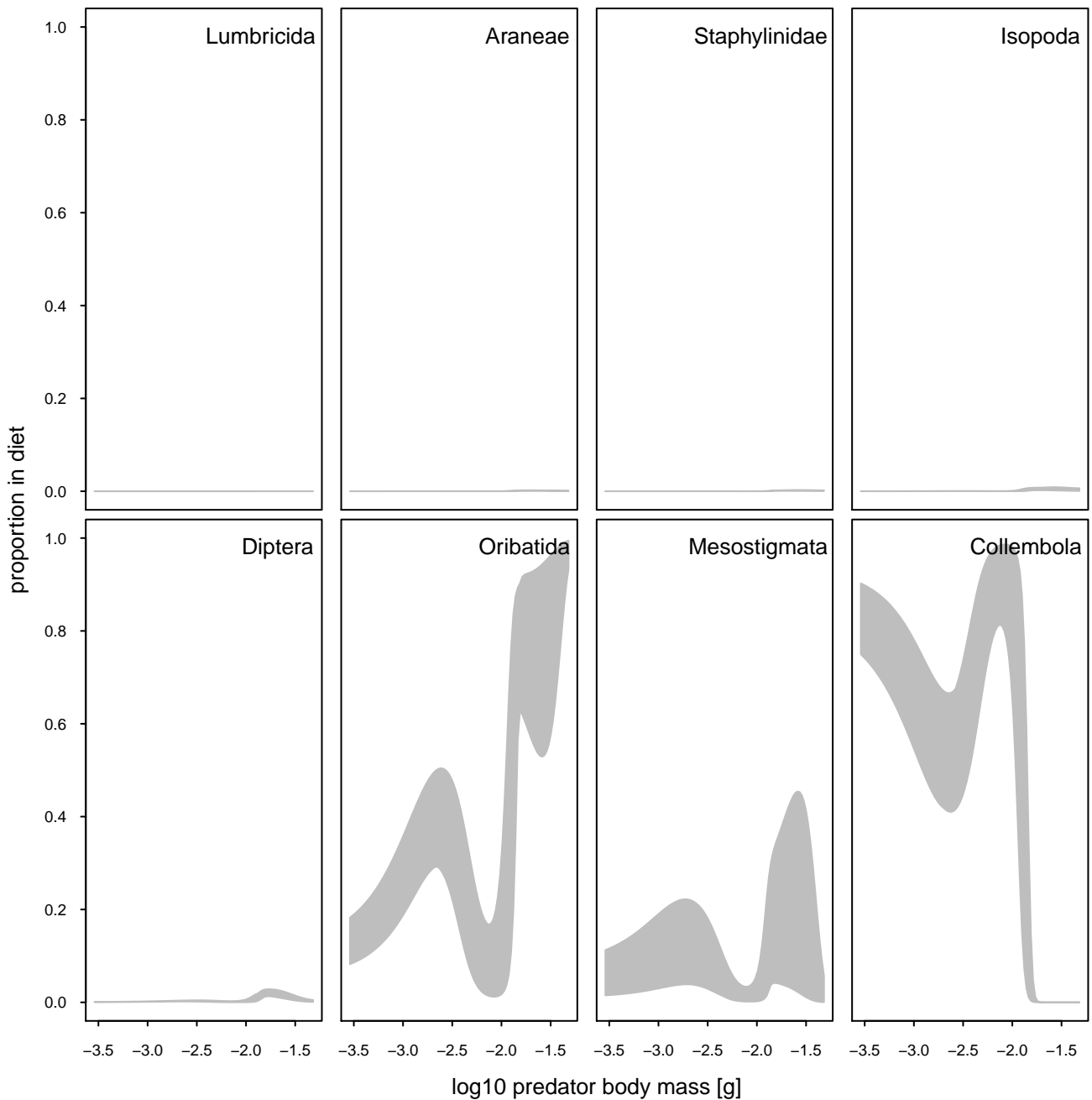
522

523 **Table 1.** Results of Generalized linear model (GLM) on the effect of predator body mass, square of predator
524 body mass, prey identity and the two-way interactions on the detection of prey DNA in *Lithobius* predators.
525 Significant effects are highlighted in bold. Df: degrees of freedom

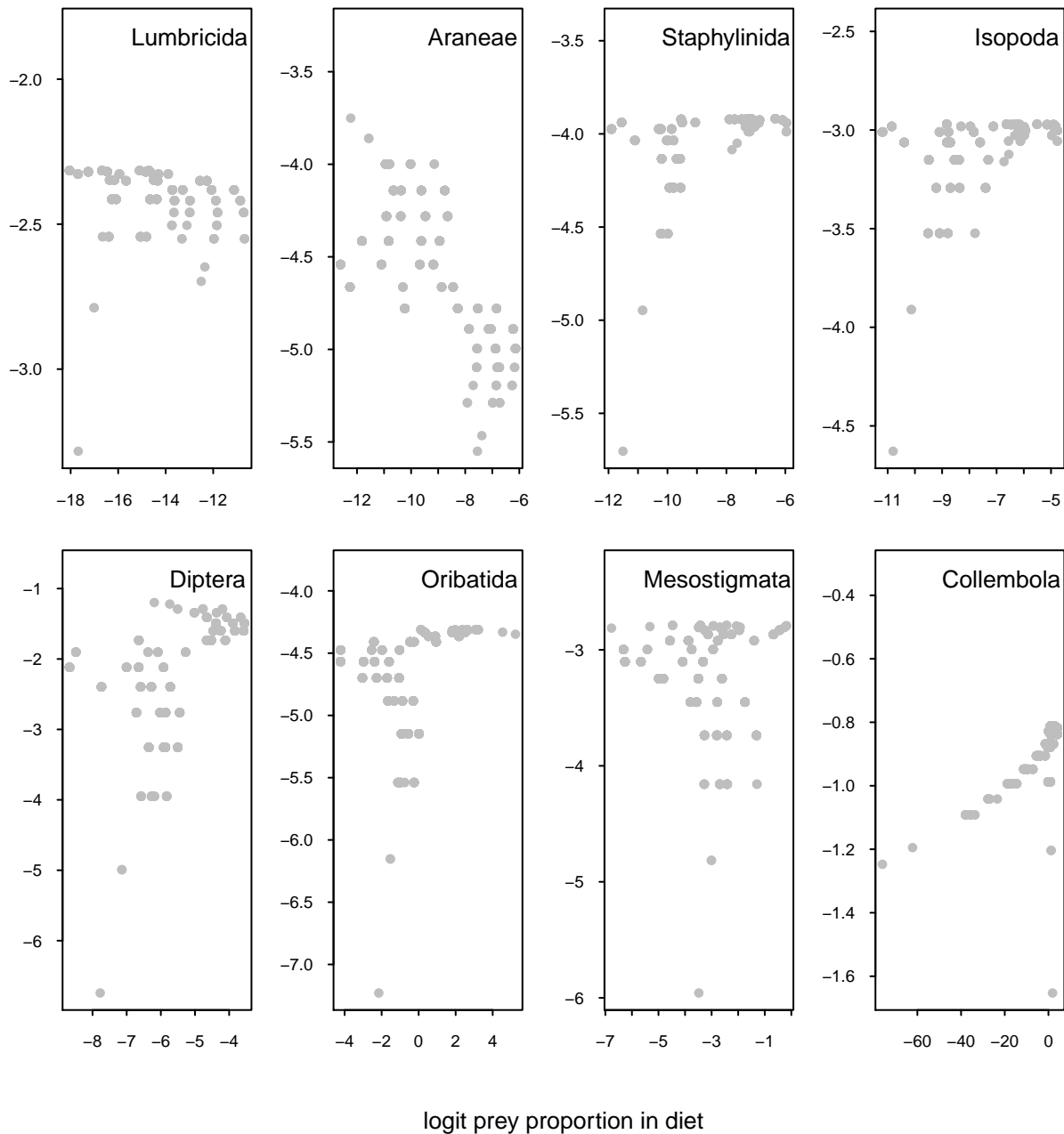
Variable	Df	Deviance	Resid. Df	Resid. Dev	P(> Chi)
NULL			4247	2270.2	
Log ₁₀ predator body mass	1	5.38	4246	2264.8	0.0204
Prey identity	7	386.35	4239	1878.5	<0.001
Prey identity× Log ₁₀ predator body mass ²	8	19.05	4231	1859.5	0.0146







logit ratio prey positive predators



logit prey proportion in diet

