

1 **Pleistocene climate change and the formation of regional species pools**

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11

12 **Abstract**

13 Despite the description of bioregions dates back from the origin of biogeography, the
14 processes originating their associated species pools have been seldom studied. Ancient
15 historical events are thought to play a fundamental role in configuring bioregions, but
16 the effects of more recent events on these regional biotas are largely unknown. We
17 use a network approach to identify regional and sub-regional faunas of European
18 *Carabus* beetles, and analyse the effects of dispersal barriers, niche similarities and
19 phylogenetic history on their configuration. We identify a transition zone matching the
20 limit of the ice sheets at Last Glacial Maximum. While southern species pools are
21 mostly separated by dispersal barriers, in the north species are mainly sorted by their
22 environmental niches. Strikingly, most phylogenetic structuration of *Carabus* faunas
23 occurred since the beginning of the Pleistocene. Our results show how extreme recent
24 historical events –such as Pleistocene climate cooling, rather than just deep-time
25 evolutionary processes, can profoundly modify the composition and structure of
26 geographic species pools.

27

28 **Introduction**

29 Naturalists have long been captivated by the geographic distribution of world biotas.
30 Rooted in the seminal ideas by Alexandre von Humbolt, this fascination has promoted
31 a long-term research agenda aiming to delineate biogeographic regions according to

32 their integrating faunas and floras (e.g. Wallace 1876, Holt et al. 2013, Rueda et al.
33 2013). Regional biotas are known to determine ecological and evolutionary dynamics
34 taking place at finer scales (Ricklefs 2008, 2015). For instance, regional species pools
35 can modulate local diversity patterns (Ricklefs 2011, Medina et al. 2014, Ricklefs and
36 He 2016), the structure and functioning of ecosystems (Naeslund and Norberg 2006),
37 or co-evolutionary processes (Calatayud et al. 2016a). However, despite their
38 fundamental importance, the processes that have configured regional biotas have
39 been seldom studied (and particularly the historical ones), and most explanations on
40 their origin and dynamics remain largely narrative (Crisp et al. 2011).

41 Perhaps the earliest speculations about the formation of regional species pools took
42 place during the flourishing of bioregionalizations in the mid-19th century (reviewed
43 by Ebach 2015). During that time, and beyond referring to physical factors (climate,
44 soils, and physical barriers), some authors already started to emphasize historical
45 influences as key elements determining the configuration of plant and animal regions.
46 For instance, when Wallace (1876) proposed his famous zoogeographic regions, he
47 argued that while the distribution of ancient lineages such as genera and families would
48 more likely reflect major geological and climatic changes spanning the early and mid-
49 Cenozoic, that of species would be more influenced by recent events such as
50 Pleistocene glaciations (see Rueda et al. 2013). Indeed, these recent events could have
51 fostered many additions and subtractions of species to regional faunas through
52 dispersal and diversification processes. Increasing evidence suggests that Pleistocene
53 glacial-interglacial dynamics may have driven population extinction (e.g. Barnes et al.
54 2002), allopatric speciation in glacial refugia (e.g. Johnson et al. 2004) and post-glacial
55 recolonization events (e.g. Hewitt 1999; Theissinger et al. 2013). As a consequence,
56 Pleistocene glaciations are known to influence current diversity patterns for many
57 taxa, particularly in the Holarctic (e.g. Svenning and Skov 2007; Hortal et al. 2011;
58 Calatayud et al. 2016b). However, the role of Pleistocene glaciations in shaping
59 regional species pools remains largely unknown.

60

61 Besides historical contingencies, niche-based processes may also determine the
62 composition of regional species pools (Mittelbach and Schemske 2015), mainly
63 throughout their effects on species distribution ranges (Soberon 2007, Hortal et al.

64 2010, Hortal et al. 2012). These processes integrate responses to abiotic conditions
65 along geographical gradients and to local and regional biotic environment (Colwell et
66 al. 2009). Furthermore, they involve a complex interplay between evolutionary,
67 ecological and biogeographical factors. For instance, species with similar climatic
68 tolerances can coexist in regions of similar climate, but their arrival (and *in situ*
69 evolution) may be constrained by geographical barriers, which may also lead to
70 climatic-tolerance divergent species pools (Fig.1a). Further, if species' resemblance in
71 climatic tolerance is phylogenetically constrained, climate-driven regional species
72 pools will be also composed of evolutionary related species (i.e. phylogenetically
73 clustered species pools), although this effect is again filtered by biogeographical
74 processes (Gouveia et al. 2014). Indeed, diversification of lineages within regions
75 separated by strong dispersal barriers (e.g. mountain ranges) may also lead to
76 phylogenetically clustered species pools (Warren et al. 2014; Fig.1a). Over the mid-
77 term, historical contingencies may contribute to erase the signature of such
78 geographically-linked diversification. For example, differential diversification rates may
79 be the predominant driver of regional species pools in climatically stable regions
80 (Cardillo 2011), yet regions where Pleistocene glaciations exerted a greater influence
81 may harbour pools of species shaped by the joint effects of current climate and post-
82 glacial dispersion (Svenning et al. 2015).

83

84 In this study we aim to disentangle the relative importance of the processes that may
85 contribute to the formation of regional species pools, using European *Carabus*
86 (Coleoptera: Carabidae) as a model lineage. *Carabus* is a species-rich ground beetle
87 genus of great popularity due to the beautiful jewel-like appearance of some species
88 (Turin et al. 2003). In general, *Carabus* species are flightless nocturnal predators of
89 snails, earthworms and caterpillars. They hold hydrophilic adaptations and are typically
90 associated to deciduous forests (Deuve et al. 2012). Previous evidence suggests that
91 the richness of species from this genus in Europe is determined to a large extent by
92 both current environmental conditions (i.e. climate and habitat) and glacial-interglacial
93 dynamics (Calatayud et al. 2016b). This makes European *Carabus* an ideal case study to
94 evaluate the joint effects of evolutionary, ecological and historical contingency
95 processes as drivers of regional species pools.

96
97 Specifically, we use data on the distribution and evolutionary relationships of *Carabus*
98 species, and network and phylogenetic analyses, to evaluate six hypotheses: First,
99 given the presumed low dispersal capacity of the species from this genus, we
100 hypothesize that (H1) European *Carabus* species pools are mainly shaped by the main
101 orographic barriers of the continent, but also, that (H2) glacial-interglacial dynamics
102 have led to strong differentiation between northern and southern regional species
103 pools. If this differentiation is true, (H3) northern European *Carabus* faunas will be
104 composed of species adapted to cold climatic conditions that colonized newly vacant
105 habitats after the withdrawal of the ice sheet, and hence their regional distribution will
106 be mostly determined by current climate. In contrast, (H4) southern faunas will be
107 mainly shaped by the joint influence of diversification events and dispersal limitations,
108 due to the combined effect of higher climatic stability towards southern latitudes (e.g.
109 climatic refugia) and a more complex orography (Alps, Pyrenees, Carpathians).
110 Therefore, (H5) species forming northern regional pools will exhibit comparatively
111 lower levels of regional endemism, whereas those forming southern regional pools
112 will show comparatively higher levels of regional fidelity. Finally, according to Wallace
113 (1876), the advance and retreat of the ice sheets during the Pleistocene should have
114 determined the spatial distribution of lineages, eroding the effects of the former
115 configuration of the distribution of the main *Carabus* lineages. Therefore, (H6) we
116 expect a temporal signal coincident with the Pleistocene in the phylogenetic structure
117 of *Carabus* faunas, and no effect of deep-time events on the current geographical
118 distribution these lineages.

119

120

121 **Material and methods**

122

123 **Data origin**

124 Data on the European distribution of *Carabus* species comes from a recent analysis on
125 the determinants of diversity of this genus in the continent (Calatayud et al. 2016b).
126 Briefly, expert-based range maps of all *Carabus* species inhabiting Europe (n = 131)
127 delineated by Turing et al. (2003) were digitized and overlaid into a 100-km equal-area

128 grid using ARCGIS (grid available at <http://dataservice.eea.europa.eu>). The resulting
129 gridded distribution maps were reviewed by several experts, eventually correcting cells
130 wrongly classified as either presences or absences (see Calatayud et al. 2016b for
131 further details). The resulting presence-absence matrix was used to represent the
132 distribution of *Carabus* species in all analyses.

133

134 We used environmental and physical geography GIS data from several commonly used
135 public domain digital repositories. Bioclimatic variables were extracted from Worldclim
136 (v1.4 Hijmans et al. 2005; available at <http://www.worldclim.org/>). Altitudinal data
137 were derived from the 30-arcsecond digital elevation model GTOPO30 provided by the
138 U. S. Geological Survey (available at <https://lta.cr.usgs.gov/GTOPO30>), and the location
139 of water bodies was extracted from vector information coming from Natural Earth
140 (available at <http://www.naturalearthdata.com/>). Finally, data for the geographical
141 distribution of forest habitats come from MODIS Land Cover (Channan et al. 2014,
142 available at <http://glcf.umd.edu/data/lc/>).

143

144 **Statistical analyses**

145 **Rationale and structure of the analyses**

146 Exploring the determinants of regional faunas requires analysing ecological,
147 evolutionary and historical factors jointly. We did so through three consecutive steps
148 (Fig.1b). First, we identified distinct regional species pools within Europe by using a
149 network community detection algorithm. From this analysis we derive a species
150 pairwise similarity matrix of occurrence into different modules that represent different
151 regions. Second, we assessed the relative importance of the environmental, spatial
152 and evolutionary determinants of such similarity. We constructed four pairwise
153 matrices to describe the relationships among species; namely, *i*) a matrix of climatic
154 niche similarity, *ii*) a matrix of habitat similarity, *iii*) a matrix of spatial connectivity
155 among their ranges, and *iv*) a phylogenetic distance matrix. Then, we used generalized
156 partial matrix regression to model the similarity in species occurrences as a function of
157 these four matrices (Fig.1b). We used this workflow to explore the factors involved in
158 the configuration of *Carabus* faunas either at the European scale (i.e. co-occurrence
159 patterns of European *Carabus* species across the delimited regions, hereafter “regional

160 scale”) and independently (i.e. within regions co-occurrence patterns of *Carabus*
161 species within each delimited region, “hereafter sub-regional scale”). Finally, we also
162 applied ancestral range estimation analysis in order to identify the time period from
163 which ancestral areas are estimated with less uncertainty. By doing so, we aimed to
164 detect important historical periods contributing to the regional organization of
165 *Carabus* lineages.

166
167 The interpretation of the joint and independent effects of explanatory matrices can
168 shed light on the different processes configuring regional faunas (see Fig.1a). In that
169 way, if niche similarities (i.e. represented by the climatic and habitat similarity
170 matrices) and phylogenetic distances altogether explained the regional co-occurrence
171 of species, then this could be interpreted as indicative of constrained niche evolution
172 (or a tendency to resemble ancestral niches) in shaping regional faunas (Fig.1a.i).
173 However, if spatial connectivity also accounted for part of this co-occurrence, this
174 would indicate that this niche conservatism pattern can be caused by geographical
175 constraints (Fig.1a.ii). Further, the independent effects of niche similarities together
176 with spatial connectivity can be more likely the consequence of a convergence of
177 climatic niches due to geographical isolation (Fig.1a.iii). Whereas the effects of
178 connectivity and phylogeny would be indicative of a primacy of intra-regional
179 speciation driven by geographical barriers. Niche similarities alone would point to an
180 unconstrained niche evolution shaping regional faunas, while phylogeny alone would
181 indicate a primacy of geographically unconstrained intra-regional speciation events.
182 Finally, either a *cul-de-sac* effect (i.e. the accumulation of species in past climatic
183 refugia) or a primacy of vicariant speciation events could lead to the existence of
184 independent effects of connectivity and regional co-occurrence (Fig.1a.iv).

185

186 **Identification of regional species pools**

187 We took advantage of community detection analysis —borrowed from network
188 theory— to identify *Carabus* regional species pools in Europe. To do so, we derived a
189 bipartite network from the presence-absence matrix, where species and grid cells
190 constitute two disjoint sets of nodes that are mutually connected according to the
191 presence of species in grid cells (e.g. Calatayud et al. 2016a). Then, we conducted a

192 modularity analysis using the index proposed by Barber (2007) and the Louvain
193 algorithm (Blondel et al. 2008, as implemented in the Matlab function “Gen Louvain”,
194 available at <http://netwiki.amath.unc.edu>; Mucha et al. 2010) to identify the optimal
195 modular structure of the bipartite network. That is, the optimal groups of grid cells
196 that show similar *Carabus* species composition together with their associated species
197 (i.e. network modules). We performed a heuristic search for an optimal solution,
198 where this analysis was repeated iteratively to obtain 100 different modular solutions,
199 retaining the one that showed the highest modularity value. This optimal solution was
200 used to conduct all subsequent analyses, although all the solutions were qualitatively
201 similar. We evaluated the statistical significance of the modules by comparing its
202 associated modularity value to a null distribution of values ($n = 100$) where the original
203 presence absence matrix was randomized using the independent swap algorithm (a
204 fixed-fixed null model implemented in the R package “picante”, Kembel et al. 2010). In
205 order to detect potential sub-modules nested within modules (i.e. sub-regional species
206 pools within regional species pools), we derived a new bipartite network from each of
207 the modules previously identified in the optimal solution, and applied the procedure
208 described above in each case.

209

210 It is important to note that despite species and grid cells were assigned to just one
211 module, they could participate with different degrees in other modules. For example,
212 despite most species in a grid cell will belong to the same module the cell does, this
213 cell could also hold species of other modules. Similarly, although a species will mostly
214 be present in cells assigned to its module, it may also occur in cells from other
215 modules. Thus, we calculated the degree of module affinity for each node of the
216 bipartite network as the proportion of links shown by a given node within its module
217 divided by the total number of its links (note that this index is related to the inter-
218 modular participation index of Guimera and Amaral 2005). Species with low module
219 affinity will tend to be widespread throughout Europe, belonging to different regional
220 species pools. While species with high module affinity will be mainly distributed within
221 their corresponding modules (highly endemic species). On the other hand, grid cells
222 with low module affinity may represent transition zones between regions.

223

224

225 **Assessing the determinants of regional species pools**

226 In order to disentangle the determinants of the current configuration of *Carabus*
227 faunas in Europe, we first generated a species-per-module matrix, where each entry
228 of the matrix represents the percentage of the distributional range of a certain species
229 that lies in a given module. Then we derived a co-occurrence pairwise similarity matrix
230 from the former matrix using the Schoener index (Schoener 1970). This index describes
231 the degree of overlap between species pairs according to their distributions
232 throughout the modules (see Krasnov et al. 2012) for a previous application to similar
233 purposes). It ranges from 0 (i.e. complete lack of overlap) to 1 (i.e. identical pattern of
234 module distribution). Note that this similarity matrix reflects the co-occurrence
235 similarities at regional scale, ignoring lower-scaled distributional patterns (i.e. two
236 species may have identical regional distribution but differ in the grid cell they are
237 present). The resultant co-occurrence pairwise similarity matrix was used as
238 dependent variable. We generated four different pairwise similarity matrices to be
239 used as explanatory variables. Two of them to account for environmental factors,
240 namely (i) a climatic-niche similarity matrix, and (ii) a habitat similarity matrix; and the
241 other two for geographical and evolutionary factors, respectively: (iii) a spatial-
242 connectivity matrix, and (iv) a phylogenetic distance matrix.

243

244 *i) Climatic-niche similarity matrix.* We characterized the climatic niche of each *Carabus*
245 species in the dataset following the approach proposed by Broennimann et al. (2012).
246 We selected six bioclimatic variables to account for the main water and energy aspects
247 of climate –namely mean annual temperature, temperature of the warmest quarter,
248 temperature of the driest quarter, total annual precipitation, total precipitation of the
249 warmest quarter and total precipitation of the driest quarter– and altitudinal range to
250 account for the effects of mesoclimatic gradients within each grid cell. These variables
251 may be among the main determinants of the distribution of *Carabus* species diversity
252 within Europe (see Calatayud et al. 2016b). We conducted a principal components
253 analysis (PCA) on these variables to obtain a bidimensional climatic space defined by
254 the two main axes (81.4% of the variance, Fig.S2). Finally, we divided this climatic
255 space into 100 grid cells, and applied kernel smoothers to the species occurrence

256 densities in the gridded climatic space to calculate niche overlap between species
257 using the Schoener index (see above). Note that the kernel density function requires at
258 least five occurrences of the species, hence species occurring in less than five grid cells
259 were excluded from subsequent analysis ($n = 12$).

260

261 *ii) Habitat similarity matrix.* The distribution of *Carabus* species may also be shaped by
262 their forest preferences (Turin et al. 2003). Thus, we made use of information relative
263 to the type of vegetation where each species tends to occur in order to characterize
264 their environmental niche. We calculated the fraction of each species' distributional
265 range laying within each vegetation category, according to 10 vegetation categories
266 derived from MODIS dataset (Evergreen broadleaf forest, deciduous needle leaf forest,
267 deciduous broadleaf forest, mixed forest, closed shrub lands, open shrub lands, woody
268 savannas, savannas and grasslands). Then, we computed pairwise similarities in the
269 preference to different vegetation types using the Schoener index (see above).

270

271 *iii) Spatial connectivity matrix.* To evaluate the potential influence of dispersal barriers
272 on the current distribution of *Carabus* species we first created a dispersal-cost surface
273 by weighting each pixel in the study area according to both its topography (in this case,
274 slope) and the presence of water bodies. Slope values ranging from 0 to 100% at each
275 pixel and were determined from GTOPO30 altitudinal data using the GRASS tool
276 `r.slope` (GRASS Development Team 2017). Water bodies from Nature Earth were assigned
277 arbitrary values of friction to the dispersal of *Carabus* species, namely 30% for pixels
278 containing rivers and lakes and 99% for pixels that lay on sea water masses (note that
279 *Carabus* species show hydrophilic adaptations). Then, the connectivity between all
280 pairs of cells was calculated as least-cost path over the dispersal-cost surface that
281 connect both cells, using the “`gdistance`” R package (van Etten 2015). Finally, the spatial
282 connectivity between two species' distributional ranges was estimated as the average
283 distance among all grid cells within the range of each species. Average distances were
284 preferred over absolute least-cost distances to avoid disproportionate differences in
285 spatial connectivity between overlapping and non-overlapping distributional ranges.

286

287 *iv) Phylogenetic distance matrix.* To unravel the evolutionary history of *Carabus* lineage
288 and the importance of evolutionary processes in determining the formation of *Carabus*
289 species pool, we reconstructed a species-level time-calibrated molecular phylogeny
290 including 89 species. We used a maximum-likelihood inference based on ten DNA
291 markers (see Appendix S3 for full details). We used taxonomic information and
292 phylogenetic uncertainty methods (Rangel et al. 2015) to place species lacking
293 molecular information into the phylogeny (see Appendix S3). Thus, we derived 100
294 different phylogenetic hypotheses from the maximum-likelihood phylogeny, by
295 randomly inserting missing species within their most derived consensus clade based on
296 taxonomic knowledge.

297

298 We used multiple regression on distance matrices and variance partitioning to
299 disentangle the relative importance of climatic niche, habitat preferences, dispersal
300 barriers and evolutionary history in determining *Carabus* species pools in Europe. First,
301 we conducted single regressions between the co-occurrence pairwise similarity matrix
302 and each of the four explanatory matrices described above to seek for significant
303 associations between the variables. Given that the distribution of co-occurrence
304 pairwise values was rather bimodal with modes at 0 and 1, respectively, we set a
305 binomial family for error distribution and “logit” as the link function (see Ferrier et al.
306 2007 and Calatayud et al. 2016a for a similar approach). To assess for significance, we
307 randomized the original species per module matrix using the independent swap
308 algorithm (see above) to derive 999 null occurrence pairwise similarity matrices. Then,
309 we fitted single regressions between the null occurrence pairwise similarity matrices
310 and each of the explanatory similarity matrices, and considered a variable to have a
311 significant effect when the regression based on observed data explained a higher
312 deviance than 99% of the null regressions. In the case of phylogenetic pairwise
313 distances we repeated this procedure for each phylogenetic hypothesis to take into
314 account phylogenetic uncertainties, applying the same criterion for significance.
315 Finally, we retained those variables that showed significant associations with the co-
316 occurrence pairwise similarity matrix, and conducted partial multiple regression
317 (Legendre and Legendre 2012) on distance matrices to explore patterns of covariation
318 among niche similarities (i.e. climatic and habitat similarity matrices) dispersal barriers

319 and phylogenetic history. We repeated this whole procedure for each module
320 independently, to further explore the determinants of *Carabus* species co-occurrence
321 into submodules.

322

323 **Ancestral range estimation**

324 Probabilistic models of geographic range evolution have proven to be of great utility to
325 deal with historical biogeographical questions (see Ronquist and Sanmartín 2011 and
326 references therein). Here we used these models to try to unravel whether, as
327 predicted by Wallace (1876), deep historical signals were eroded by Pleistocene
328 glaciations in the configuration of European *Carabus* faunas. We used a maximum
329 likelihood approach for inferring geographical range evolution implemented in the R
330 package BioGeoBears (Matzke 2014), using independently each one of the 100
331 phylogenetic hypotheses created before. BioGeoBears fits a variety of models that
332 differ in their rates and/or types of dispersal, extinction, sympatric speciation,
333 vicariance and founder event speciation. We fitted the six available models (DEC,
334 DEC+J, DIVALIKE, DIVALIKE+J, BAYAREALIKE, BAYAREALIKE+J, see Matzke 2013)
335 comparing their informative capacity through the Akaike information criterion (AIC).
336 Species ranges were here coded as present/absent in each region detected in the
337 former network analyses.

338

339 The estimation of ancestral ranges usually tends to be more ambiguous in deeper
340 nodes of the phylogeny, as the potential lability of geographical ranges would tend to
341 blur deep time signals (Losos and Glor 2003). Yet, if the Pleistocene glacial periods had
342 important effects in the distribution of species it could be expected that ancestral
343 range estimations will begin to be more accurate around the Pleistocene (that is, pre-
344 Pleistocene signals on distributional range evolution will be eroded). To explore this we
345 attended to potential changes in the relationship between node age and the marginal
346 probability of the single most-probable ancestral state at each internal node. We
347 obtained these probabilities from the best model among the six geographical range
348 evolution models in BioGeoBears (see above) and in each of the phylogenetic
349 hypotheses used. Then, general additive mixed models (GAMMs) were fitted to the
350 node marginal probability as a function of node age, including the phylogenetic

351 hypothesis as a random factor. Further, in case we found different best models in
352 different phylogenetic hypotheses we also included a fixed variable with the model
353 type. Finally, we also carried out generalized linear mixed models (GLMMs) using a
354 piecewise regression procedure to assess for a shift in the relationship between
355 marginal probability and node age. The breakpoint in the piecewise regression was
356 assessed by including the breakpoint as a new parameter in the GLMM and then
357 minimizing the deviance of the fitted model using the function “optimize” in the R
358 package lme4 (Bates et al. 2014). Finally, we used AICc to compare the models with
359 and without breakpoint. In both cases we used a binomial family and a loglink
360 function, since the marginal probabilities of the nodes vary between zero and one. All
361 analyses were carried out in R (R core team 2015),: GAMM with the package gamm4
362 (Wood and Scheipl 2014), GLMM with lme4 (Bates et al. 2014) and AICc with
363 AICcmodavg (Mazerolle 2011).

364

365 **Results**

366 **Identification of regional faunas**

367 The *Carabus* occurrence network was significantly modular ($M=0.385$, $p=0.01$), dividing
368 Europe in seven modules that group zoogeographically distinct regions with their
369 associated faunas; that is, different regional species pools (Figs. 2 and S1). Further, all
370 modules but module 2 showed a significant sub-modular structure, presenting a
371 decrease in modularity with latitude (mean $M=0.316$, ranging from 0.154 to 0.468; all
372 p -values < 0.05 , see Table S1). Module 1 holds 21 species mainly living in Southwestern
373 Palearctic (Iberian Peninsula, North of Africa, Balearic Islands, Corsica, Sardinia and the
374 western half of Sicilia). This module was subdivided into four submodules. Module 2
375 only groups two species which are endemic of Crete. Module 3 identifies an East
376 Mediterranean region including the Italic Peninsula, part of Greece and Turkey. This
377 module holds 18 species and was subdivided into five submodules. Module 4 depicts a
378 Central European region embracing the Alps and Carpathian Mountains, as well as
379 Central European plains. This module is the richest one, with forty-nine *Carabus*
380 species, and was split into 4 submodules. Module 5 and module 6 correspond with
381 northern regions and are the poorest ones, holding ten species each. The former
382 expands from Iceland and the British Isles almost to the Ural Mountains. The latter

383 includes The Ural Mountains and expands to the easternmost zone of the study area.
384 Both modules were divided into 3 submodules. Finally, module 7 includes 21 species
385 and embraces a south-eastern central European region expanding from the Carpathian
386 Mountains to the south Ural Mountains. This module was split into 3 submodules.

387

388 The transition zones between regions appear to be associated with geographical
389 barriers such as the Pyrenees, Alps, Carpathian and Ural Mountains, as well as the
390 Bosphorus strait (Fig. 2c). Further, we also found a transition zone running through
391 Central Europe in a west-to-east belt, separating southern and northern regions.
392 Interestingly, this transition zone closely follows the southern limits of the ice sheet at
393 the Last Glacial Maximum (LGM), suggesting a link between the configuration of
394 *Carabus* regional faunas and Pleistocene glacial conditions.

395

396

397 **Correlates of regional co-occurrence**

398 Matrix regressions showed that both environmental niche similarity (i.e. climate and
399 habitat) and spatial connectivity were significantly related to species co-occurrence in
400 regions and sub-regions ($p < 0.01$). In contrast, evolutionary relatedness was not related
401 with regional co-occurrence, as phylogenetic distances did not show a significant
402 relationship at $p > 0.01$ for any single phylogenetic hypothesis. Environmental niche
403 similarity was the best predictor of co-occurrence in both regions and sub-regions,
404 although its effects were stronger when predicting the co-occurrence into sub-regions
405 (Fig. 3). Spatial connectivity had stronger effects when predicting the co-occurrence
406 into regions relative to sub-regions, where its effect mostly overlapped with that of
407 environmental niche. The relative importance of these two sets of factors showed a
408 clear latitudinal trend. The effects of connectivity were stronger in southern regions
409 (i.e. modules 1 and 3), which is not surprising given the more complex orography of
410 these regions. But importantly, although niche similarity was the best predictor of co-
411 occurrence in all cases, its effects became more evident towards northern regions.
412 Indeed, niche similarities and spatial connectivity were significant predictors of the co-
413 occurrence into sub-regions in all regional pools except the two northern ones (Table
414 S2). Only niche similarities (mainly climate) showed significant effects in these two

415 modules (i.e. 5 and 6), as expected if the species in these regional pools colonized the
416 northern regions only recently, and were geographically sorted according to current
417 climate.

418

419 **Ancestral range estimation**

420 BAYAREALIKE+J model was the best supported by data in 59% of the phylogenetic
421 hypotheses, followed by DEC+J and DEC, which were the best in 23 and 18% of the
422 hypotheses, respectively (Table 1). Interestingly, the BAYAREALIKE+J model does not
423 account for vicariant cladogenetic events. Rather, it considers founder effect
424 speciation, which could be seen as a signal of Pleistocene glaciations promoting
425 speciation in glacial refugia (i.e. acting as islands) and eroding the ancestors' ranges.
426 Indeed, according to this model the range contraction parameter had the highest
427 estimate (Table 1). This parameter was however of little importance in the other two
428 models, which accounted for vicariant events in both cases. Regardless the
429 biogeographical model used, GAMM results showed that node marginal probability of
430 the most probable state increased as expected towards younger nodes ($P < 0.01$, R^2
431 $= 0.34$). However, this increase became abrupt coinciding with the beginning of the
432 Pleistocene – which started around 2.59 Mya.– (Fig.4). Indeed, we found that the
433 GLMM including the breakpoint was better in terms of AICc (AICc weight for the model
434 including the breakpoint = 1), and that the relationship between state probability and
435 node age changed at 2.47 Mya. (confidence interval at 95% = 2.23, 2.73). This points to
436 the signal left by the Pleistocene on the configuration of European *Carabus* faunas.
437 These results were confirmed using different approaches (see Appendix S4).

438

439

440 **Discussion**

441 More than 140 years ago, Wallace (1876) foresaw that the influence of Pleistocene
442 glaciations on the distribution of diversity had been strong enough so as to erode the
443 imprint of previous events. Our results confirm Wallace's thoughts, showing a
444 remarkable coincidence between the distribution of the ice sheet at the Last Glacial
445 Maximum and the current configuration and evolutionary structure of European
446 *Carabus* Faunas. There is a close spatial relationship between the southern limits of the

447 area glaciated at LGM, and the transition zone separating the southern and northern
448 regions identified by our analyses. Indeed, this border also coincides with the line
449 identified by Calatayud et al. (2016b), where the relationship between *Carabus* species
450 richness and current climate changes in shape and strength (Fig. 2). Thus, it seems that
451 the climate changes underwent during the Pleistocene not only shape local diversity
452 patterns (e.g. Svenning and Skov 2007, Araujo et al. 2008, Hortal et al. 2011), but they
453 have also left a strong imprint on the geographical structure of species composition at
454 a regional scale. Accordingly, the species from the northernmost region (module 5)
455 show the lowest level of endemism (Fig. S2, as expected for regional faunas composed
456 of species that have recently colonized the north of Europe from southern glacial
457 refugia; Araujo et al. 2008, Calatayud et al. 2016b). In fact, although these species are
458 widely distributed across southern Europe, all of their ranges only overlap near to the
459 northern Carpathian Mountains, an area that has been shown to be a glacial refugia
460 for some *Carabus* species (Homburg et al. 2013, Fig. S3). Additionally, the decrease in
461 modularity values with latitude also points to a lesser geographical structure of
462 northern assemblages, which can be interpreted as the result of a post-glacial
463 colonization.

464

465 Besides the Pleistocene effects in the definition and geographical structure of regional
466 pools of species, we also found evidence of the imprint of this geological period on the
467 processes configuring the distribution of *Carabus* faunas. The generally strong
468 relationship between regional patterns of co-occurrence and both niche similarities
469 and spatial connectivity shows that co-occurring species tend to have similar realized
470 environmental niches and that also tend to be geographically constrained by the same
471 dispersal barriers. This latter result was expected, given the low dispersal capacity of
472 *Carabus* species (see Turin et al. 2003), which is likely to be behind the spatial
473 coincidence of the zones of transition between regional faunas and geographical
474 barriers. Perhaps more unexpected is the relationship of regional co-occurrence and
475 niche similarities without any significant phylogenetic effect. This implies that the
476 geographical configuration of barriers to dispersal has restricted species within regions
477 of similar climate, rather than climatic-niche conservatism constraining their co-
478 occurrences. These results point to that *Carabus* niche evolution is, to some extent,

479 evolutionary unconstrained, which is congruent with the high adaptation capacity of
480 insects in general (e.g. Overgaard and Sørensen 2008). Going further, this questions to
481 what extent the observed species' regional occurrence is the consequence of their
482 environmental niche and not the other way around (see Hortal et al. 2012, Wüest et al.
483 2015).

484
485 Whatever the origin of the relationship between species occurrence and
486 environmental conditions, which is certainly true is that its strength changes among
487 regions. These changes follow a latitudinal gradient in the importance of
488 environmental niche similarities. In northern regions, the similarity in the realized
489 niche seems to be stronger related to the occurrence into sub-regions than in the
490 south. This might be a direct consequence of the processes that determine regional
491 pools being dependent on the particular history of each region (Ricklefs 2015). These
492 findings are also congruent with the effects of post-glacial colonization, where
493 formerly glaciated areas show a clear sorting of species northwards due to low time for
494 dispersal (Svenning and Skov 2007). This contrasts with southern regions, where
495 climate has been milder and more stable and the species have had more time to
496 diversify. This suggests that species' geographical arrangement in southern latitudes is
497 more likely the result of historical contingency rather than environmental preferences
498 (Hortal et al. 2011, Calatayud et al. 2016b).

499
500 The lack of relationship between the phylogenetic distances among *Carabus* species
501 and their regional co-occurrence can be the outcome of either a generalized
502 preponderance of vicariant events and/or a “cul-de-sac effect” (O'Regan 2008). The
503 former will more likely imply that the generation of the dispersal barriers that shaped
504 the regions will also promote the geographical split of many lineages and subsequent
505 allopatric speciation events (Weeks et al. 2016). Yet, the formation of the geographical
506 accidents associated to the delimitation of *Carabus* regions largely predates the origin
507 of the genus (see Beccaluva et al. 1998, Deuve et al. 2012). On the other hand, a
508 generalized dispersion into climatic refugia, together with a subsequent stagnancy
509 within them (i.e. a “cul-de-sac” effect) may also produce the observed mixing of
510 unrelated lineages into regions. Although it is difficult to distinguish between both

511 processes, we believe it is more plausible to think that southern regions have
512 accumulated unrelated species while acting as glacial refugia, whereas northern
513 regions were recolonized by unrelated species with similar environmental niches ---or
514 simply with higher dispersal capacity (see above).

515

516 In agreement with this idea, the model of ancestral range estimation that is best
517 supported by the data does not include vicariant events. On the contrary, this model
518 takes into account founder speciation events, a process typical from islands (Provine
519 1989). This could be interpreted on the light of southern regions acting as islands
520 during the different Pleistocene glacial maxima. Should this be true, a temporal signal
521 of this period on the spatial organization of *Carabus* lineages should be evident in their
522 phylogeny. We found such signal on the relationship between node maximum
523 probability state and node age. Indeed, the striking coincidence between the
524 breakpoint where this relationship becomes steeper and the beginning of the
525 Pleistocene argues in favour of the imprint of this epoch on the configuration of
526 *Carabus* regional faunas. We obtained similar results using several other approaches
527 (Appendix S4), supporting the notion that the current regional organization of *Carabus*
528 species and lineages has its very roots at the beginning of the Pleistocene. This contrasts
529 with ancestral range estimations for clades inhabiting areas that were never glaciated,
530 where more ancient signals were found in the spatial sort of lineages (Condamine et al.
531 2015, Economo et al. 2015, Tänzler et al. 2016, Toussaint and Balke 2016). It thus
532 seems that the repeated advances and retreats of ice sheets during this geological
533 period produced the repeated cycles of retreat to southern regions and advance
534 towards northern regions of *Carabus* species, a hustle-and-bustle process that
535 ultimately led to the observed mixing of unrelated lineages.

536

537 In sum, our results provide solid arguments in favour of the importance of Pleistocene
538 glaciations along with geographical barriers in structuring the regional faunas of this
539 group. On the one hand, European *Carabus* faunas are primarily delimited by the
540 location of the southern limit of the ice sheet at LGM, which separates two large
541 regions that differ not only in species composition, but also in the processes underlying
542 the spatial organization of these species. On the other hand, the phylogenetic

543 structure of these faunas coincides with the beginning of the Pleistocene. This not only
544 implies that the geographical distribution of species and lineages is deeply shaped by
545 past climates, but also that the ecological processes (Naeslund and Norberg 2006,
546 Madrigal et al. 2016) and evolutionary mechanisms (Wüest et al. 2015, Calatayud et
547 al. 2016a) that are dependent on regional species pools may be profoundly affected by
548 the history of Earth climates. It is therefore essential to take into account past
549 historical events while trying to understand, not only current diversity patterns and the
550 processes behind them, but also processes and patterns occurring at local scales.

551

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- 741

742 **Table 1.** Results of the best ancestral range estimation models. d= dispersion, e=
743 extinction; j= founder speciation. Prop.best = proportion of phylogenetic hypotheses
744 where a model was the best in terms of AICc. Mean values among different
745 phylogenetic hypotheses (incl. standard deviation within brackets), are shown for each
746 parameter.

Model	N° param.	d	e	j	Prop. Best
BAYAREALIKE+J	3	0.015(0.001)	0.063(0.017)	0.015(0.005)	0.59
DEC+J	3	0.025(0.001)	<0.001(<0.001)	0.008(0.003)	0.23
DEC	2	0.026(0.001)	0.003(0.001)	0	0.18
BAYAREALIKE	2	0.019(0.001)	0.111(0.006)	0	0
DIVALIKE	2	0.031(0.001)	0.002(0.001)	0	0
DIVALIKE+J	3	0.03(0.001)	0(0)	0.006(0.003)	0

747

748

749 **Figure 1.** Hypothetical examples of the factors configuring regional faunas and work
750 flow. a) Figure showing four hypothetical processes configuring regional faunas.
751 Dotted lines depict different regions while colours correspond with different climates.
752 In each case, the tips of the phylogeny point to regional distribution of the species. b)
753 Workflow and potential results: 1) Hypothetical results of modularity analysis over the
754 occurrence network; 2) similarity matrix of occurrence into modules; 3) pairwise
755 matrix of environmental niche similarities; phylogenetic distances and topographical
756 connectivity; and 4) hypothetical results and interpretations of a partial matrix
757 regression on species occurrence similarities as a function of niche similarities,
758 phylogenetic distances and connectivity.

759

760 **Figure 2.** European *Carabus* regions found by the network community detection
761 analysis. a) Geographical location of modules (i.e. regions) and submodules. b) Module
762 simplification of the occurrence network. Circles represent a module, being their size
763 proportional to the species group within them. Links depict the species shared among
764 regions, being its width proportional to the number of species. c) Values of module
765 affinity per grid cell; green colours (i.e. cells with low affinity) identify transition zones.
766 The dotted black line corresponds with the southern limit of the ice sheet at LGM
767 (extracted from Ehlers and Gibbard 2004). The blue line depicts the breakpoint where
768 the temperature--*Carabus* richness relationship changes, as found in Calatayud et al.
769 (2016b).

770

771 **Figure 3.** Results of the partial generalized matrix regression of similarity in regional
772 co-occurrence, as a function of environmental niche similarity (climate and habitat),
773 topographical connectivity and phylogenetic distances. The first and second bars
774 correspond with the models including all regions and subregions, respectively. The
775 remaining bars correspond with the models where the similarities in subregional
776 occurrence were analysed independently for the species of each region. Con.= spatial
777 connectivity; Envir. = environmental niche similarities; Overl.= Overlap.

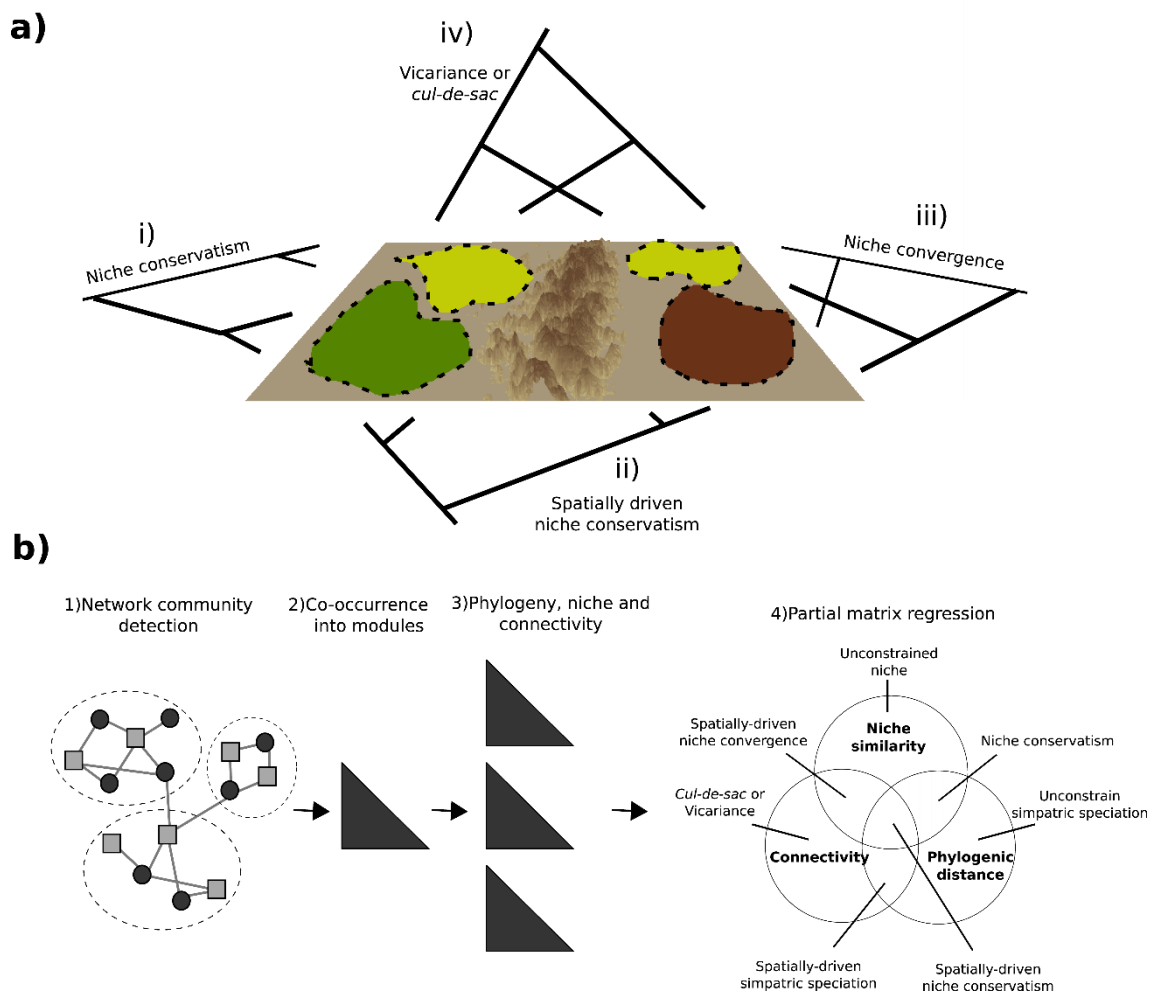
778

779 **Figure 4.** Results of the ancestral range estimation. a) Marginal probability of the most
780 probable state at each internal node for a phylogenetic hypothesis where the
781 BAREALIKE+J model was the best supported by the data. The size of the circles depicts
782 the value of probability, while the colours represent the combinations of regions
783 forming the state. The phylogeny is labelled at subgenera level. b) GAMM predictions
784 of the marginal probability as a function of node age for the BAREALIKE+J model. The
785 shaded area corresponds with the interval confidence at 95%. The dotted black line
786 represent the breakpoint found by piecewise GLMM regression. The interval confidence
787 at 95% for the breakpoint is also provided. See Fig. S5 for the predictions of the other
788 models.
789

790 **Figure 1.**

791

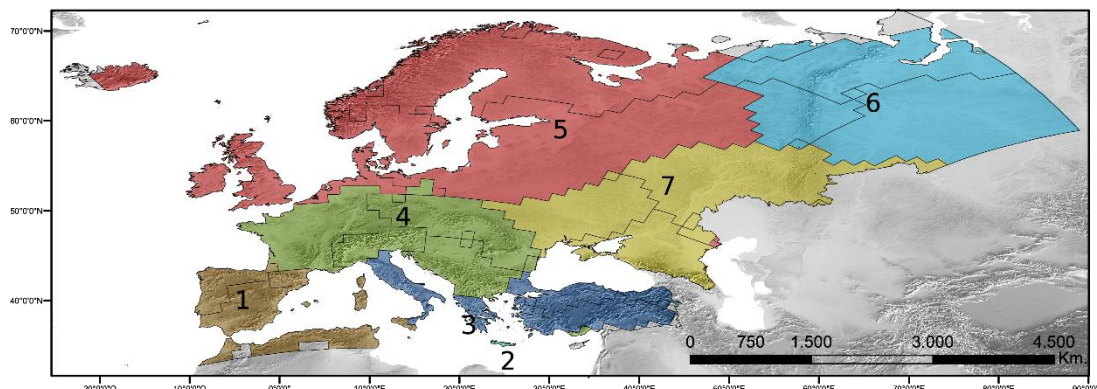
792



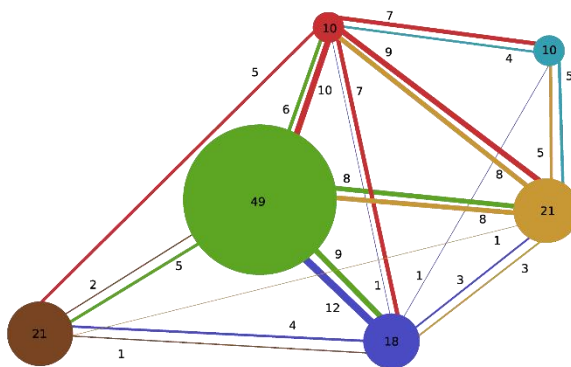
793

794 **Figure 2.**

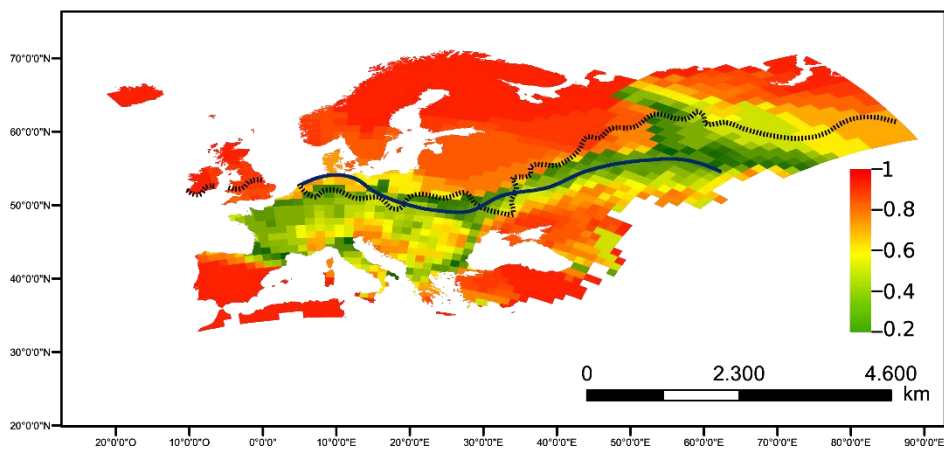
a)



b)



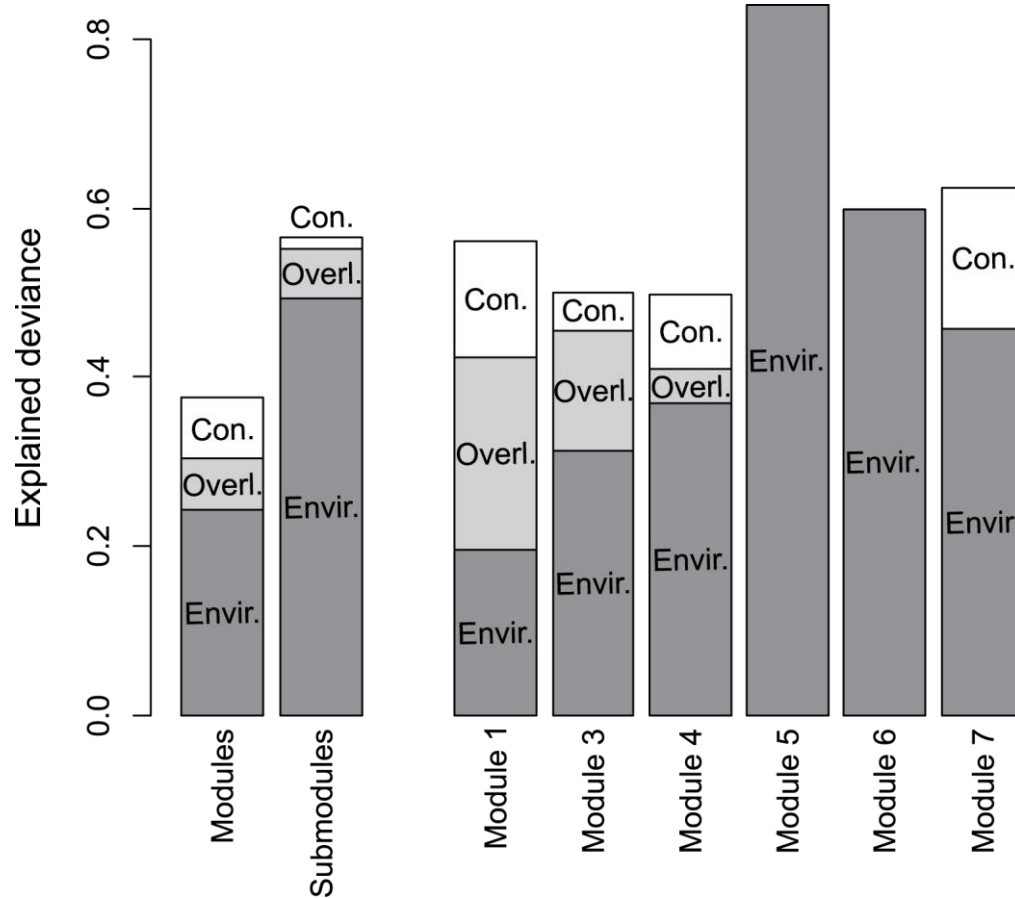
c)



795

796

797 **Figure 3**

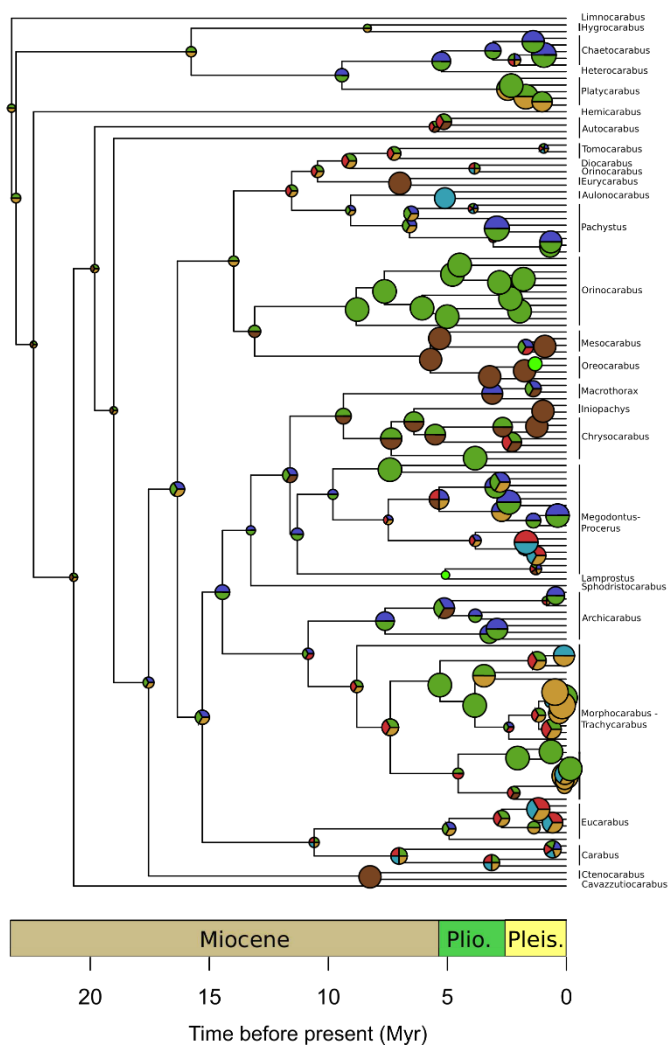


798

799

800 **Figure 4.**

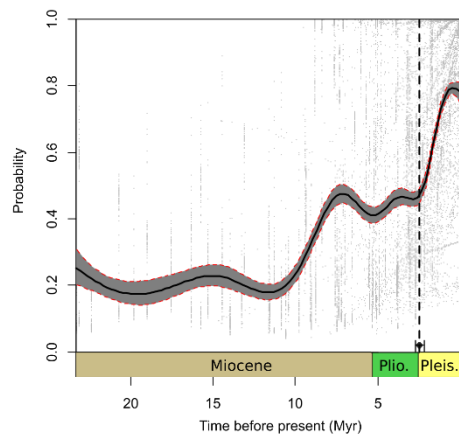
a)



State probability



b)



801