

Coral *Symbiodinium* community composition across the Belize Mesoamerican Barrier Reef

System is driven by host species and environmental variability

Baumann JH^{1*}, Davies SW^{1,2}, Aichelman HE^{1,3}, Castillo KD¹

¹*Department of Marine Sciences, University of North Carolina at Chapel Hill, Chapel Hill, North Carolina, 27599-3300 United States of America*

²*Boston University Department of Biology, 5 Cummington Mall, Boston MA 02215*

³*Department of Biological Sciences, Old Dominion University, 302 Miles Godwin building, Norfolk VA, 23529*

*Corresponding author: baumannj@live.unc.edu, 513-306-1516, ORCID ID: 0000-0003-0113-0491

Keywords: coral, *Symbiodinium*, symbiosis, marine science, environmental variability

Abstract

Reef-building corals maintain a symbiotic relationship with dinoflagellate algae of the genus *Symbiodinium* and this symbiosis is vital for the survival of the coral holobiont. *Symbiodinium* community composition within the coral host has been shown to influence a coral's ability to resist and recover from stress. A multitude of stressors including ocean warming, ocean acidification, and eutrophication have been linked to global scale coral decline in coral health and cover in recent decades. Three distinct thermal regimes (high_{TP}, mod_{TP}, and low_{TP}) following an inshore-offshore gradient of declining average temperatures and thermal variation were identified on the Belize Mesoamerican Barrier Reef System (MBRS). Quantitative metabarcoding of the ITS-2 locus was employed to investigate differences and similarities in *Symbiodinium* genetic diversity of the Caribbean corals *Siderastrea siderea*, *S. radians*, and *Pseudodiploria strigosa* between the three thermal regimes. *Siderastrea siderea* associated with distinct *Symbiodinium* communities when compared to their congener *S. radians* as well as *P. strigosa*, demonstrating host-specificity of *Symbiodinium* along the MBRS. *Symbiodinium* community differences were only detected across thermal regimes for *S. siderea*; however, thermal parameters influenced *Symbiodinium* communities in all coral species investigated. Interestingly, *Symbiodinium trenchi*, a symbiont known to confer thermal tolerance, was dominant only in *S. siderea* at one sampled offshore site and was rare inshore, suggesting that coral thermal tolerance in more thermally variable inshore habitats is achieved through alternative mechanisms. Overall, thermal parameters alone were not the primary drivers of *Symbiodinium* community composition, suggesting that environmental variables unrelated to temperature (i.e., light availability, or nutrients) may play key roles in structuring coral-algal communities in Belize.

Introduction

Obligate symbioses, relationships in which two or more organisms depend on one another for nutrition and survival, occur globally. Such symbioses are ubiquitous in plants and Mycorrhiza [1], ants and bacteria [2], and lichens [3]. The effects of climate change are expected to disrupt proper functioning of many symbioses, including that of reef-building corals [4-6]. The success of coral reefs worldwide depends on the symbiosis between the coral host and photosynthetic algae of the genus *Symbiodinium* [7-9]. Under stressful conditions this coral-*Symbiodinium* relationship breaks down, resulting in the loss of endosymbiont cells and/or photosynthetic pigments from the coral tissue in a process known as ‘coral bleaching’ [10]. Coral bleaching is most commonly associated with thermal stress [11-15] and is predicted to increase in frequency and severity as the world’s climate continues to change [5, 16-21]. Increased thermal stress resulting from climate change combined with other local stressors such as eutrophication, habitat destruction, and overfishing has created an uncertain future for coral reefs [6, 13, 22]. In the Caribbean Sea, warming rates are higher than in any other tropical basin [23] and coral cover has declined by as much as 80% in recent decades [24]. It has been predicted that Caribbean coral reefs may suffer biannual bleaching events within the next 20-30 years [17] and annual bleaching by 2040 [25].

In the face of a changing climate and widespread reef declines, corals will need to rapidly increase their thermal tolerance in order to persist in their current form [18, 26]. Coral thermal tolerance has been shown to be influenced by a coral’s thermal history, which among other factors includes average environmental temperature and extent of thermal variability [27, 28]. On average, corals previously exposed to warmer temperatures show decreased mortality during bleaching events [29] and more stable growth patterns [30] compared with corals exposed to

cooler temperatures, which exhibit greater mortality during heat stress and declining growth rates with increased temperatures [29, 30]. Exposure to greater daily thermal variation has also been shown to increase coral thermal tolerance [31] and has been associated with higher coral cover and slower mortality rates when compared to reefs exposed to less thermal variation [32]. Coral thermal tolerance is also heritable with larvae from parent colonies on lower-latitude (warmer) reefs showing a 10-fold increase in survival under heat stress when compared to larvae from cooler reefs locations [33]. A growing body of evidence suggests that the coral host plays a significant role in thermal tolerance [34-37], however, plasticity or specificity of coral-associated *Symbiodinium* communities also plays a significant role in overall thermal tolerance [38-41].

The clades, lineages, or species of *Symbiodinium* hosted by a coral are critical to its survival and resilience to stress. The genus *Symbiodinium* is genetically diverse and comprises at least nine divergent clades [clades A-I; 42]. These clades can be further broken down into lineages, corresponding approximately to species level diversity [43], with some species conferring variable benefits [38, 42, 44]. In particular, some *Symbiodinium* are more thermally tolerant than others [9, 38, 45], specifically *Symbiodinium* clade D [46]. In contrast, clade C is more thermally sensitive [47-49], yet it includes *Symbiodinium thermophilum*, a thermally tolerant species within clade C endemic to the Red Sea [50]. This example illustrates that making clade level generalizations is problematic due to the physiological diversity within a single *Symbiodinium* clade [51]. Specific lineages within clades can also confer various advantages. For example, C1 enhances growth rate [52], *S. thermophilum* confers heat tolerance [50], and B2 confers cold tolerance [53]. Additionally, species D1a (*Symbiodinium trenchi*) has been shown to be both heat tolerant [54, 55], and cold tolerant [45]. However, the increased thermal tolerance of a coral which predominantly hosts clade D *Symbiodinium* appears to come at a cost of lower

lipid stores, reproductive potential, growth, and carbon fixation rates compared with corals that host other clades [56-59]. Due to the high levels of variation in coral host-*Symbiodinium* interactions, it is essential to identify which lineages are present in order to help predict how a coral may respond to environmental stressors.

The majority of coral species host one dominant *Symbiodinium* lineage [42, 60, 61] along with several non-dominant lineages [62], each proliferating primarily by asexual cloning [51]. However, other corals can host multiple dominant lineages or clades [38, 51]. Recent advances in genetic techniques, especially next-generation sequencing (NGS), have allowed researchers to identify cryptic and low-abundance symbionts comprising 0.1% or more of the total *Symbiodinium* community within a host [36, 63]. It is important to understand these low-abundance *Symbiodinium*, as they have the potential to play important roles in coral-algal holobiont physiology under ambient and stressful conditions [64-66, but see also 67]. Identifying trends in *Symbiodinium* community variation (including cryptic or low abundance lineages) within and between species across a coral reef may allow for a better understanding of the role of *Symbiodinium* communities in modulating coral response to environmental variation.

Symbiodinium communities have been shown to vary regionally [between reef systems; 59, 68, 69], locally [within a reef system; 68], temporally [across time on the same reef; 70], and within a colony [69]. Studies of this variation have revealed geographically endemic lineages of *Symbiodinium* which may play a significant role in local and regional scale coral survival and stress tolerance [38, 69, 71]. While temperature stress may play a role in structuring *Symbiodinium* communities [72], variations in other environmental factors have also been shown to drive *Symbiodinium* community composition. For example, physical processes and total suspended solids (a proxy for nutrients and flow) drive *Symbiodinium* associations within the

Orbicella annularis species complex in Belize and Panama [68]; however, on a regional scale (e.g., the entire Caribbean Sea), *O. annularis* *Symbiodinium* communities differed based on patterns of chronic thermal stress [73]. Additionally, the presence of several subclades of *Symbiodinium* correlated with other environmental parameters, such as cooler summers, nutrient loading, and turbidity [73]. Taken together, these studies demonstrate that variation in *Symbiodinium* communities can be driven by a variety of environmental parameters and may be specific to each coral species in each specific environment.

The majority of Caribbean *Symbiodinium* biogeography studies have focused on the *Orbicella* species complex [68, 69, 73] as *Orbicella* spp. has experienced significant declines over the last two decades [74] and is now listed as ‘threatened’ under the Endangered Species Act. However, the variation in *Symbiodinium* communities of other more stress tolerant corals, such as *Siderastrea siderea* and *S. radians* [75-80], remain relatively understudied. Here, we assess *Symbiodinium* community composition in three species of ubiquitous Caribbean corals (*Siderastrea siderea*, *S. radians*, and *Pseudodiploria strigosa*) across three distinct thermal regimes along the Belize Mesoamerican Barrier Reef System (MBRS) previously shown to influence coral community composition [81]. Coral-associated *Symbiodinium* communities were examined across an inshore-offshore thermal gradient and a latitudinal gradient to elucidate the role that coral species, local habitat, and a suite of thermal parameters play in structuring *Symbiodinium* communities in the western Caribbean Sea.

Methods:

Site selection and characteristics

Ten sites along the Belize MBRS, that were previously characterized into three thermally distinct regimes (low_{TP}, mod_{TP}, high_{TP}) and exhibited variations in coral species diversity and

richness [81], were selected. High_{TP} sites (inshore) were characterized by larger annual temperature variation, higher annual maximum temperatures, and are exposed to temperatures above the regional bleaching threshold of 29.7°C (Aronson et al., 2002) more often than mod_{TP} sites (mid-channel reefs) and low_{TP} sites (offshore) [81]. High_{TP} sites were dominated by stress tolerant and weedy coral species while corals representing all four coral life histories [stress tolerant, weedy, competitive, and generalist; 80] were present in low_{TP} and mod_{TP} sites [81].

Sample Collection

In November 2014, five to ten (quantity depended on local availability) coral tissue microsamples (approx. 2 mm diameter) were collected at 3 to 5 m depth from three coral species (*Siderastrea siderea*, *S. radians*, and *Pseudodiploria strigosa*) at nine sites across four latitudes along the Belize MBRS (Fig 1; Table 1). Each latitudinal transect contained a low_{TP}, mod_{TP}, and high_{TP} site. The transects from north to south were: Belize City, Dangriga, Placencia, and Punta Gorda (Fig 1). All three sites within the Punta Gorda and Placencia transects were sampled, but only the low_{TP} and high_{TP} sites were sampled along the Belize City and Dangriga transects due to time constraints. Samples collected at the Belize City high_{TP} site were collected in October 2015, as no corals were located in the area in 2014, but patch reefs were located in 2015. Coral microsamples were collected using a hammer and chisel and sampled colonies were separated by at least 1m to randomize micro-environmental and host genetic effects in order to attain more site-specific representative samples. Microsamples were collected from colony edges to avoid unnecessary damage to the larger colony and to limit effects of *Symbiodinium* zonation within an individual [69]. Tissue microsamples were placed on ice immediately following collection for transport to mainland Belize. Microsamples were then preserved in 96% ethanol and stored on

ice at -20°C . Preserved microsamples were transported on ice to the coral ecophysiology lab at the University of North Carolina at Chapel Hill and stored at -20°C until DNA isolation.

Sea Surface Temperature

Daily 1-km horizontal resolution sea surface temperature (SST) estimates were acquired from the NASA Jet Propulsion Laboratory's Multi-Scale High Resolution SST (JPL MUR SST) product via NOAA Environmental Research Division's Data Access Program (ERDDAP-<https://coastwatch.pfeg.noaa.gov/erddap/index.html>) [82] and analyzed following Baumann et al [81]. Briefly, SST data were binned by latitude and longitude for each site and annual values for maximum temperature, temperature range, number of days above the regional bleaching threshold (29.7°C , Aronson et al. [83]), and longest streak of consecutive days above the bleaching threshold were calculated for each site and for the entire MBRS reef area. Using standard deviations from the mean for all four parameters, sites were classified as low_{TP}, mod_{TP}, and high_{TP} [81]. Several additional temperature parameters were taken into account for this study, including: annual degree heating days (similar to degree heating weeks, as per Gleeson and Strong [84]), annual minimum temperature, annual average temperature, annual winter average temperature, and annual summer average temperature. Values for these parameters within the three thermal regimes are reported in Table S1.

DNA Extraction

Coral holobiont (coral, algae, and microbiome) DNA was isolated from each sample following a modified phenol-chloroform [83, 85, 86] method described in detail by Davies et al (2013). Briefly, DNA was isolated by immersing the tissue in digest buffer (100 mM NaCl,

10mM Tris-Cl pH 8.0, 25 mM EDTA pH 9.0, 0.5% SDS, 0.1 mgml⁻¹ Proteinase K, and 1 µgml⁻¹ RNaseA) for 1 h at 42°C followed by a standard phenol-chloroform extraction. Extracted DNA was confirmed on an agarose gel and quantified using a Nanodrop 2000 Spectrophotometer (Thermo Scientific).

PCR amplification and metabarcoding

The ITS-2 region (350 bp) was targeted and amplified in each sample using custom primers that incorporated *Symbiodinium* specific ITS-2-dino-forward and its2rev2-reverse regions [63, 71, 87]. Each primer was constructed with a universal linker, which allowed for the downstream incorporation of Illumina specific adapters and barcodes during the second PCR as well as four degenerative bases whose function was to increase the complexity of library composition. The forward primer was 5'-GTCTCGTCGGCTCGG + *AGATGTGTATAAGAGACAG* + NNNN + **CCTCCGCTTACTTATATGCTT**-3' where the underlined bases are the 5'- universal linker, italicized bases indicate spacer sequences, N's denote degenerative bases and the bold bases are the ITS-2-dino. The reverse primer was 5'- TCGTCGGCAGCGTCA + *AGATGTGTATAAGAGACAG* + NNNN + **GTGAATTGCAGAACTCGTG**-3'.

Each 20uL PCR reaction contained 5-100 ng DNA template, 12.4 µL MilliQ H₂O, 0.2 µM dNTPs, 1µM forward and 1µM reverse primers, 1X *Extaq* buffer, and 0.5 U (units) *Extaq* polymerase (Takara Biotechnology). PCR cycles were run for all samples using the following PCR profile: 95°C for 5 min, 95°C for 40 s, 59°C for 2 min, 72°C for 1 min per cycle and a final elongation step of 72°C for 7 min. The optimal number of PCR cycles for each sample was determined from visualization of a faint band on a 2% agarose gel (usually between 22 and 28

cycles) as per Quigley et al. (2014). PCR products were cleaned using GeneJET PCR purification kits (Fermentas Life Sciences) and then a second PCR reaction was performed to incorporate custom barcode-primer sequences [63] modified for Illumina Miseq as in Klepac et al. [88]. Custom barcode primer sequences included 5'-*Illumina adaptor* + 6 bp **barcode sequence** + one of two universal linkers-3' (e.g.: 5'- *CAAGCAGAAGACGGCATACGAGAT* + **GTATAG** + GTCTCGTGGGCTCGG-3', or 5'- *AATGATACGGCGACCACCGAGATCTACAC* + **AGTCAA** + TCGTCGGCAGCGTC-3'). These universal linking barcoded adapters can be used to target any loci and therefore significantly reduce costs associated with Miseq sequencing of multiple loci. Following barcoding, PCR samples were visualized on a 2% agarose gel and pooled based on band intensity (to ensure equal contributions of each sample in the pool). The resulting pool was run on a 1% SYBR Green (Invitrogen) stained gel for 60 minutes at 90 volts and 120 mAmps. The target band was excised, soaked in 30 uL of milli-Q water overnight at 4°C, and the supernatant was submitted for sequencing to the University of North Carolina at Chapel Hill High Throughput Sequencing Facility across two lanes of Illumina MiSeq (one 2x250, one 2x300). The two lanes produced similar mapping efficiencies (73% and 73%, respectively; Table S3).

Bioinformatic Pipeline

The bioinformatic pipeline used here builds upon previous work by Quigley et al. [63] and Green et al. [71]. Raw sequences were renamed to retain sample information and then all forward (R1) and reverse (R2) sequences were concatenated into two files, which were processed using CD-HIT-OTU[89]. CD-HIT-OTU clusters concatenated reads into identical groups at 100% similarity for identification of operational taxonomic units (OTUs). Each sample was then

mapped back to the resulting reference OTUs and a counts table for each sample across all OTUs was produced. A BLASTn search of each reference OTU was then run against the GenBank (NCBI) nucleotide reference collection using the representative sequence from each OTU to identify which *Symbiodinium* lineage was represented by each OTU (Table S2).

The phylogeny of representative sequences of each distinct *Symbiodinium* OUT was constructed using the PhyML tool [90, 91] within Geneious version 10.0.5 (<http://geneious.com>) [92]. PhyML was run using the GTR+I model (chosen based on delta AIC values produced from jModelTest [90, 93]) to determine the maximum likelihood tree. The TreeDyn tool in Phylogeny.fr was used to view the tree (Fig 2) [94-96]. The reference sequences included in the phylogeny were accessed from GenBank (Table S6).

Statistical Analysis

OTU count analysis used the R [R Core97] package *MCMC.OTU* and followed methods described in Green et al. [71]. First, outlier samples with low sequence coverage (total log counts ≥ 2.5 standard deviations below the mean of all samples) were identified and removed, which removed 3 samples. Next, rare OTUs ($< 0.1\%$ of the global sum of counts [as per 63]) were identified and discarded leaving 56 of the original 5,132 OTUs. Many remaining OTUs were identified as having the same *Symbiodinium* lineage (i.e., C1 or D1a) and these OTUs were regressed against one another. Positive correlations between OTUs within a lineage may indicate paralogous loci from the same genome [36, 71]. As a result, reads from OTUs within the same lineage that showed a positive R^2 and significant p -value following linear regression were pooled in order to control for possible overestimation of biodiversity [98]. Pooling resulted in a final OTU table containing ten OTUs (Table S2). Raw reads, trimmed reads, mapped reads, and

percentage of reads mapped per species were calculated and reported in Table 2. Final pooled OTUs were run through the MCMC.OTU package in R and fit to a model that included fixed effect for host species, collection site, and thermal regime (Table S4). Differences between fixed effects were calculated based on their sampled posterior distributions and statistical significance was calculated as per Matz et al. [99]. OTU count data were converted to relative abundances (%), which were used to generate Fig 3 (Table S5).

To visualize differences in symbiont communities between temperature regimes, latitude, and species, principal component analyses (PCA) were performed on all OTUs that passed filtering using the *vegan* package in R [100]. Count data were transformed using Bray-Curtis similarity and were used as input for PCA. PERMANOVA was carried out on each PCA using the *adonis* function of the *vegan* package in R [100]. Canonical Correlation Analysis (CCA), which is widely used for ecological applications [101], was undertaken using the *cca* function of the *vegan* package in R. CCA was used to measure associations between *Symbiodinium* communities within a species and temperature variables (Table S1) at each temperature regime.

Results

Symbiodinium diversity and abundance across the Belize MBRS

Our analysis produced 118,834 unique sequences of which 89,211 mapped to 10 OTUs (Table 1). The dominant OTU (hereafter referred to as lineage) in *S. siderea* was C1.I (74.39%), while B1.I dominated *S. radians* (70.31%) and *P. strigosa* (51.74%) samples (Table S5, Fig 3). Nine out of ten *Symbiodinium* lineages were present in *S. siderea* and *P. strigosa* while all ten were present in *S. radians* (Table S5). The four most abundant lineages in *S. siderea* were C1.I, C1.III, D1a, and B1.I (74.39%, 12.94%, 9.29%, and 2.94%, respectively; Table S5, Fig 3A) and

date of collection did impact the dominate *Symbiodinium* lineages (all samples collected in 2014 except for Belize City high_{TP} which were collected in 2015; Fig 3). *Symbiodinium* D1a (*S. trenchi*) was most abundant in *S. siderea* at low_{TP} sites, particularly the low_{TP} site along the most southern Punta Gorda transect (Table S5, Fig 3A) while lineage C1.II is more abundant in central and northern Belize (Belize City and Dangriga transects; Figs 1, 2).

The four most abundant lineages in *S. radians* were B1.I, C1.I, B1.II, and C1.II (70.31%, 13.41%, 6.54%, and 2.19% respectively; Table S5, Fig 3B). B1.I was the dominant symbiont across all thermal regimes and all latitudes, but C1.I and C1.II were the most abundant *Symbiodinium* lineages in several samples from the central Placencia transect (Table S5, Fig 3B). Lineage C1.II was only present in proportions above 1% in 2 samples, both from the mod_{TP} site along the Placencia transect (Table S5, Fig 3B). D1a (*S. trenchi*) was only present in low abundance in *S. radians* (Table S5, Fig 3B).

The four most abundant lineages in *P. strigosa* were B1.I, C1.I, C1.II, and C1.III (51.74%, 21.87%, 16.92%, and 6.24%, respectively). C1.II was the most abundant lineage at the low_{TP} site in the Placencia transect, but B1.I was most abundant at all other sites (Table S5, Fig 3). C1.I was the second most abundant lineage in mod_{TP} and high_{TP} sites and C1.II was the second most abundant lineage in the low_{TP} site (Table S5, Fig 3C). D1a (*S. trenchi*) was only present in low abundance in *P. strigosa* (Table S5, Fig 3C).

Host species specificity in Symbiodinium community composition

Symbiodinium communities differed significantly between *S. siderea* and the other two coral host species (Table S4, Fig 4A, p -value=0.001). This difference appears to be driven by higher relative abundances of C1.I and D1a (*S. trenchi*) in *S. siderea* compared to *P. strigosa* and

S. radians (Fig 3A). Within *S. siderea*, *Symbiodinium* communities varied by thermal regime and site, but not by latitude (Table S4, Fig 4B). *Symbiodinium* communities in *S. radians* and *P. strigosa* did not differ significantly by thermal regime, site, or latitude (Table S4). As *Symbiodinium* communities did not differ significantly by latitude in any of the three coral species (Table S4), there does not appear to be a significant effect of Illumina lane (northern and southern latitudes were run on separate lanes; Table 1) on dominant *Symbiodinium* lineages.

Correlation of Symbiodinium community structure with temperature parameters

Canonical correlation analysis (CCA) revealed that thermal parameters measured in this study correlated with 29.6% of the variance in *Symbiodinium* communities within *S. siderea*, 11.5% of the variance in *S. radians*, and 28.4% of the variance in *P. strigosa* (Fig 5). 17.2% of total variation in *S. siderea* symbiont community is explained by CCA axis 1 and 8.9% is explained by CCA axis 2 (Fig 5A). CCA axis 1 explained 7.6% of the variance in *S. radians* and 19.3% of the variance in *P. strigosa*; while CCA axis 2 explained 2.4% and 8.2% of the variation in *S. radians* and *P. strigosa*, respectively (Figs 5B, C). Average annual temperature, average annual minimum temperature, annual temperature range, degree heating days, and summer average temperature all appear to influence variation in *Symbiodinium* communities in *S. siderea* (Fig 5A); while summer average temperature, annual average temperature, and annual minimum temperatures appeared to play principal roles in *S. radians* (Fig 5B). Average annual temperature, annual days above the bleaching threshold, and the annual longest streak of days above the bleaching threshold best explained variation within *P. strigosa* (Fig 5C).

Discussion

Host-specificity drives Symbiodinium community composition

This study indicates that *Siderastrea siderea* hosts significantly different *Symbiodinium* communities than *S. radians* and *P. strigosa* on the Belize MBRS (Table S5, Fig 3), providing evidence to support previous findings of host-specific *Symbiodinium* associations [51, 102]. The three coral species studied here were found to be dominated by the two most abundant *Symbiodinium* clades in the Caribbean [103]: B1 in *S. radians* and *P. strigosa* colonies and C1 in *S. siderea* (Table S5, Fig 3). These associations are consistent with previous studies that identified the same dominant *Symbiodinium* in these species on the Belize MBRS [102]; but contrast with findings of other studies on the same species elsewhere in the Caribbean, that identified other dominant *Symbiodinium* lineages in these species [102, 104, 105], supporting previous evidence for regional endemism within the Caribbean Sea and specificity of the coral-algal symbiosis [102, 106]. Differences in *Symbiodinium* communities between coral host species appear to be driven by the relative abundance of B1 and C1 as well as the presence or absence of D1a (Fig 4A). Interestingly, *Symbiodinium* communities appear more similar between *S. radians* and *P. strigosa* than between *S. radians* and *S. siderea*, indicating that members of the same coral genus do not necessarily share a common dominant *Symbiodinium* partner. Presence of multiple lineages of C1 and B1 (Table S2, Table S5) support previous evidence of phylogenetic partitioning, or highly specific lineages, in clades B and C [69, 102, 107, 108]. Differences in *Symbiodinium* communities between *S. siderea* and *S. radians*/ *P. strigosa* is suggestive that corals species are differentially affected by the environmental gradients sampled here.

Temperature parameters shape Symbiodinium community composition in S. siderea, but not other species

Symbiodinium communities varied significantly across thermal regimes in *S. siderea* (Table S4, Fig 4B), supporting previous evidence that habitat type [109] and temperature [110] are correlated with differences in *Symbiodinium* associations. *Symbiodinium* communities did not differ significantly across thermal regimes in *S. radians* or *P. strigosa*, possibly due to low sample size at each sampling site for these two coral species (Table 1; Fig 3). While *Symbiodinium* communities did not differ between thermal regimes in *S. radians* or *P. strigosa* (Table S4), temperature parameters accounted for a proportion of the variation in these species (approx. 10% and 28%, respectively; Fig 5B, C). Thermal parameters also explained up to 30% of the variation in *Symbiodinium* communities in *S. siderea*, with temperature range and degree heating days playing the largest roles (Fig 5A). This finding supports evidence from previous studies that temperature, and more specifically chronic thermal stress (degree heating days and days above bleaching threshold; Fig 5A, C), can be important drivers of *Symbiodinium* associations in some, but not all coral species [73, 110]. In this study, the role of temperature parameters in influencing *Symbiodinium* associations varied based on coral host species and was most significant for *S. siderea*. However, temperature parameters did not account for all of the variance in *Symbiodinium* communities for any coral host species investigated in the current study (Fig 5), indicating that other local factors, such as nutrients, light availability, and/or sedimentation may play a role [46, 111-115].

Role of local impacts on Symbiodinium communities

It has previously been shown that prevalence of specific *Symbiodinium* types within a coral host species can differ based on local scale environmental parameters such as nutrient loading and turbidity [73]. While these variables were not quantified in this study, chlorophyll-*a* (*chl-a*), a proxy for nutrient input, has previously been shown to be positively correlated with thermal regime in Belize. Specifically, high_{TP} sites had higher *chl-a* than low_{TP} sites across the Belize MBRS [81]. Therefore, a PERMANOVA that shows significant differences in *Symbiodinium* communities between thermal regimes includes a confounding effect of nutrient input (Table S4). Since significant differences in *Symbiodinium* communities occurred between thermal regimes in *S. siderea* only, it is possible that nutrient loading or turbidity played a role in *Symbiodinium* variation within *S. siderea*, but may not have significantly influenced *Symbiodinium* communities in *S. radians* or *P. strigosa*. However, the magnitude of this influence cannot be teased apart from the effect of thermal regime without extensive quantification of nutrient concentrations across the Belize MBRS.

Coral host plays a significant role in thermal tolerance

In this study, the relative abundance of thermally tolerant *Symbiodinium* D1a (*S. trenchi*) was not associated with inshore reefs as in Toller et al. [116], marginal reefs as in Hennige et al. [117] and LaJeunesse et al. [103], sites exposed to the highest temperatures as in Baker et al. [46], or sites exposed to the widest range of thermal fluctuations as in Abrego et al. [118], Fabricius et al. [119], and LaJeunesse et al. [39, 120]. Instead, *S. trenchi* was most prevalent at the southern Punta Gorda low_{TP} and mod_{TP} sites (Table S1, S5, Fig 3). Since *S. trenchi* is often associated with recently bleached and/or recovering corals [46, 121], but can be replaced or outcompeted following recovery [105], it could be possible that a recent bleaching event may

have occurred at these sites, however these data are not available. In summer 2014, temperatures at all sites in this study exceeded the published local bleaching threshold of 29.7°C [83] (Fig S1), yet *S. trenchi* was only the dominant symbiotic partner in eight *S. siderea* samples, all of which were from the same two sites (Punta Gorda low_{TP} and mod_{TP}; Fig 3). The presence of *S. trenchi* in several *P. strigosa* corals taken from the Punta Gorda mod_{TP} site provides additional evidence of temperature stress at these sites (Punta Gorda low_{TP} and mod_{TP}). However, no evidence of mass bleaching was observed at any of the study sites during collection in Nov 2014 or Oct 2015 so corals at these sites had either bleached recently or retained *S. trenchi* as a dominant symbiont following bleaching, possibly as a way to increase thermal tolerance. Lower thermal tolerance has been proposed previously at these sites (Punta Gorda low_{TP} and mod_{TP}) and may be due to nutrients and sediments exported from Guatemala and Honduras by currents that wash over this area of the Belize MBRS [122-124]. Low abundances of *S. trenchi* at other low_{TP} and mod_{TP} sites corroborates this hypothesis, as estimated thermal stress occurred at all latitudes at roughly the same magnitude (Fig S1). Overall, lack of *S. trenchi* in high_{TP} sites indicates that regardless of warmer and more variable conditions, these three coral species do not associate with this thermally tolerant symbiont. Therefore, presumed increased thermal tolerance at high TP sites may be due to local adaptation of the coral host [36, 125] or strains of *Symbiodinium* [126, 127]. Further research into coral host and symbiont local adaptation would be needed to confirm this hypothesis.

Conclusion

This study demonstrates that *Symbiodinium* communities associated with corals in Belize are dependent on both host species as well as environmental variables. *S. siderea* *Symbiodinium*

communities were divergent from *S. radians* and *P. strigosa* (Fig 3; Fig 4A). Temperature parameters played a role in driving *Symbiodinium* community composition in all three coral host species, but overall significant differences across thermal regimes were only detected in *S. siderea*. Temperature parameters did not account for all of the variation in *Symbiodinium* communities within any of the three coral host species, suggesting that local impacts such as nutrients, sediment, or light availability may influence *Symbiodinium* communities on the Belize MBRS. Additionally, low abundance of *S. trenchi* in inshore high_{TP} sites indicates thermal tolerance at these sites must be conferred through alternative mechanisms, such as local adaptation.

Acknowledgements

We thank J. Watkins, L. Speare, and A. Knowlton for laboratory assistance and C. Berger for assistance with coding. We also thank NASA JPL and NOAA ERDAAP for access to MUR SST data used in this paper, Belize Fisheries Department for issuing research and collection permits, and Garbutt's Marine for providing local expert guides and boats for field research. This work was supported by the Rufford Foundation (<http://www.rufford.org>) Small Grant to JHB (15802-1); National Science Foundation (Oceanography) ([nsf.gov](http://www.nsf.gov)) to KDC (OCE 1459522); Department of Defense NDSEG fellowship to JHB. The authors declare that no conflict of interests exists.

References

1. Meyer FH (1966) Mycorrhiza and other plant symbioses. *Symbiosis* 1: 171-255.
2. Degnan PH, Lazarus AB, Brock CD, Wernegreen JJ (2004) Host-symbiont stability and fast evolutionary rates in an ant-bacterium association: Cospeciation of *Camponotus* species and their endosymbionts, *Candidatus Blochmannia*. *Systematic Biology* 53: 95-110.
3. Honegger R (1991) Functional aspects of the lichen symbiosis. *Annual review of plant biology* 42: 553-578.
4. Brosi GB, McCulley RL, Bush LP, Nelson JA, Classen AT, Norby RJ (2011) Effects of multiple climate change factors on the tall fescue-fungal endophyte symbiosis: infection frequency and tissue chemistry. *New Phytologist* 189: 797-805.

5. Coles SL, Brown BE (2003) Coral Bleaching-Capacity for Acclimatization and Adaptation. *Advances in Marine Biology* 46: 183-213.
6. Hoegh-Guldberg O, Mumby PJ, Hooten AJ, Steneck RS, Greenfield P, Gomez E, Harvell CD, Sale PF, Edwards AJ, Caldeira K, Knowlton N, Eakin CM, Iglesias-Prieto R, Muthiga N, Bradbury RH, Dubi A, Hatzitolos ME (2007) Coral Reefs Under Rapid Climate Change and Ocean Acidification. *Science* 318: 1737-1742.
7. Muscatine L (1990) The role of symbiotic algae in carbon and energy flux in reef corals. In: Dubinsky, Z (ed.) *Ecosystems of the World 25: Coral Reefs*. Elsevier, New York, pp. 75-87
8. DeSalvo MK, Sunagawa S, Fisher PL, Voolstra CR, IGLESIAS-PRIETO R, Medina M (2010) Coral host transcriptomic states are correlated with Symbiodinium genotypes. *Molecular Ecology* 19: 1174-1186.
9. Warner ME, Fitt WK, Schmidt GW (1996) The effects of elevated temperature on the photosynthetic efficiency of zooxanthellae in hospite from four different species of reef coral: a novel approach. *Plant, Cell and Environment* 19: 291-299.
10. Glynn PW (1993) Coral reef bleaching: ecological perspectives. *Coral Reefs* 12: 1-17.
11. Hughes TP, Kerry JT, Álvarez-Noriega M, Álvarez-Romero JG, Anderson KD, Baird AH, Babcock RC, Beger M, Bellwood DR, Berkelmans R (2017) Global warming and recurrent mass bleaching of corals. *Nature* 543: 373-377.
12. Heron SF, Maynard JA, Ruben van Hooidonk C (2016) Warming Trends and Bleaching Stress of the World's Coral Reefs 1985–2012. *Scientific reports* 6.
13. Hughes TP, Baird AH, Bellwood DR, Card M, Connolly SR, Folke C, Grosberg R, Hoegh-Guldberg O, Jackson JBC, Kleypas J, Lough JM, Marshall P, Nystrom M, Palumbi SR, Pandolfi JM, Rosen B, Roughgarden J (2003) Climate change, human impacts, and the resilience of coral reefs. *Science* 301: 929-933.
14. Hoegh-Guldberg O (1999) Climate change, coral bleaching and the future of the world's coral reefs. *Marine Freshwater Research* 50: 839-866.
15. Wild C, Hoegh-Guldberg O, Naumann MS, Colombo-Pallotta MF, Ateweberhan M, Fitt WK, Iglesias-Prieto R, Palmer C, Bythell JC, Ortiz J-C, Loya Y, van Woesik R (2011) Climate change impedes scleractinian corals as primary reef ecosystem engineers. *Marine and Freshwater Research* 62: 205-215. doi: <http://dx.doi.org/10.1071/MF10254>
16. Wooldridge S, Done T, Berkelmans R, Jones R, Marshall P (2005) Precursors for resilience in coral communities in a warming climate: a belief network approach. *Marine Ecology Progress Series* 295: 157-169.
17. Donner SD, Knutson TR, Oppenheimer M (2007) Model-based assessment of the role of human-induced climate change in the 2005 Caribbean coral bleaching event. *Proceedings of the National Academy of Sciences* 104: 5483-5488.
18. Donner SD, Skirving WJ, Little CM, Oppenheimer M, Hoegh-Guldberg O (2005) Global assessment of coral bleaching and required rates of adaptation under climate change. *Global Change Biology* 11: 2251-2265.
19. Jokiel PL, Coles SL (1990) Response of Hawaiian and other Indo-Pacific reef corals to elevated temperature. *Coral Reefs* 8: 155-162.
20. D'Croz L, Mate JL, Oke JE (2001) Responses to elevated seawater temperature and UV radiation in the coral *Porites lobata* from upwelling and non-upwelling environments on the Pacific coast of Panama. *Bulletin of Marine Science* 69: 203-214.
21. McWilliams JP, Cote IM, Gill JA, Sutherland WJ, Watkinson AR (2005) Accelerating impacts of temperature-induced coral bleaching in the Caribbean. *Ecology* 86: 2055-2060.

- 477 22. Frieler K, Meinshausen M, Golly A, Mengel M, Lebek K, Donner S, Hoegh-Guldberg O (2013)
478 Limiting global warming to 2 [thinsp][deg] C is unlikely to save most coral reefs. *Nature Climate*
479 *Change* 3: 165-170.
- 480 23. Chollett I, Müller-Karger FE, Heron SF, Skirving W, Mumby PJ (2012) Seasonal and spatial
481 heterogeneity of recent sea surface temperature trends in the Caribbean Sea and southeast Gulf
482 of Mexico. *Marine Pollution Bulletin* 64: 956-965.
- 483 24. Gardner TA, Cote IM, Gill JA, Grant A, Watkinson AR (2003) Long-term region-wide declines in
484 Caribbean corals. *Science* 301: 958-960.
- 485 25. Van Hooidonk R, Maynard JA, Liu Y, Lee SK (2015) Downscaled projections of Caribbean coral
486 bleaching that can inform conservation planning. *Global Change Biology*.
- 487 26. Barshis DJ (2015) Genomic Potential for Coral Survival of Climate ChangeCoral Reefs in the
488 Anthropocene. Springer, pp. 133-146
- 489 27. Middlebrook R, Hoegh-Guldberg O, Leggat W (2008) The effect of thermal history on the
490 susceptibility of reef-building corals to thermal stress. *Journal of Experimental Biology* 211:
491 1050-1056. doi: 10.1242/jeb.013284
- 492 28. Castillo KD, Helmuth BST (2005) Influence of thermal history on the response of *Montastraea*
493 *annularis* to short-term temperature exposure. *Marine Biology* 148: 261 - 270.
- 494 29. Pineda J, Starczak V, Tarrant A, Blythe J, Davis K, Farrar T, Berumen M, da Silva JC (2013) Two
495 spatial scales in a bleaching event: Corals from the mildest and the most extreme thermal
496 environments escape mortality. *Limnology and Oceanography* 58: 1531-1545.
- 497 30. Castillo KD, Ries JB, Weiss JM, Lima FP (2012) Decline of forereef corals in response to recent
498 warming linked to history of thermal exposure. *Nature Climate Change* 2: 756-760.
- 499 31. Oliver T, Palumbi S (2011) Do fluctuating temperature environments elevate coral thermal
500 tolerance? *Coral Reefs* 30: 429-440.
- 501 32. Soto IM, Muller Karger FE, Hallock P, Hu C (2011) Sea Surface Temperature Variability in the
502 Florida Keys and Its Relationship to Coral Cover. *Journal of Marine Biology* 2011: 10. doi:
503 10.1155/2011/981723
- 504 33. Dixon GB, Davies SW, Aglyamova GV, Meyer E, Bay LK, Matz MV (2015) Genomic determinants
505 of coral heat tolerance across latitudes. *Science* 348: 1460-1462.
- 506 34. Baird AH, Bhagooli R, Ralph PJ, Takahashi S (2009) Coral bleaching: the role of the host. *Trends*
507 *in Ecology & Evolution* 24: 16-20.
- 508 35. Kenkel C, Meyer E, Matz M (2013) Gene expression under chronic heat stress in populations of
509 the mustard hill coral (*Porites astreoides*) from different thermal environments. *Molecular*
510 *Ecology* 22: 4322-4334.
- 511 36. Kenkel C, Goodbody-Gringley G, Caillaud D, Davies S, Bartels E, Matz M (2013) Evidence for a
512 host role in thermotolerance divergence between populations of the mustard hill coral (*Porites*
513 *astreoides*) from different reef environments. *Molecular Ecology* 22: 4335-4348.
- 514 37. Barshis DJ, Ladner JT, Oliver TA, Seneca FO, Traylor-Knowles N, Palumbi SR (2013) Genomic basis
515 for coral resilience to climate change. *Proceedings of the National Academy of Sciences* 110:
516 1387-1392.
- 517 38. Rowan R, Knowlton N, Baker A, Jara J (1997) Landscape ecology of algal symbionts creates
518 variation in episodes of coral bleaching. *Nature* 388: 265-269.
- 519 39. Lajeunesse TC, Smith, R., Walther, M., Pinzon, J., Pettay, D.T., McGinley, M., Aschaffenburg, M.,
520 Medina-Rosas, P., Cupul-Magana, A.L., Perez, A.L., Reyes-Bonilla, H., and M.E. Warner (2010)
521 Host-symbiont recombination vs. natural selection in response of coral-dinoflagellate symbioses
522 to environmental disturbance. . *Proceedings of the Royal Society B*.

40. Baker AC (2003) Flexibility and Specificity in Coral-Algal Symbiosis: Diversity, Ecology, and Biogeography of *Symbiodinium*. Annual Review of Ecology, Evolution, and Systematics 34: 661-689.
41. Silverstein RN, Cunning R, Baker AC (2015) Change in algal symbiont communities after bleaching, not prior heat exposure, increases heat tolerance of reef corals. Global Change Biology 21: 236-249.
42. Coffroth MA, Santos SR (2005) Genetic diversity of symbiotic dinoflagellates in the genus *Symbiodinium*. Protist 156: 19-34.
43. Lajeunesse TC (2001) Investigating biodiversity, ecology, and phylogeny of endosymbiotic dinoflagellates in the genus *Symbiodinium* using the ITS region: In search of a 'species' level marker. Journal of Phycology 37: 866-880.
44. Lajeunesse T, Trench R (2000) Biogeography of two species of *Symbiodinium* (Freudenthal) inhabiting the intertidal sea anemone *Anthopleura elegantissima* (Brandt). The Biological Bulletin 199: 126-134.
45. Silverstein RN, Cunning R, Baker AC (2017) Tenacious D: *Symbiodinium* in clade D remain in reef corals at both high and low temperature extremes despite impairment. The Journal of Experimental Biology. doi: 10.1242/jeb.148239
46. Baker AC, Starger CJ, McClanahan TR, Glynn PW (2004) Corals' adaptive response to climate change. Nature 430: 741.
47. Rowan R (2004) Thermal adaptation in reef coral symbionts. Nature 430: 742.
48. Tchernov D, Gorbunov MY, de Vargas C, Yadav SN, Milligan AJ, Haggblom M, Falkowski PG (2004) Membrane lipids of symbiotic algae are diagnostic of sensitivity to thermal bleaching in corals. Proceedings of the National Academy of Sciences of the United States of America 101: 13531-13535.
49. Berkelmans R, Van Oppen MJ (2006) The role of zooxanthellae in the thermal tolerance of corals: a 'nugget of hope' for coral reefs in an era of climate change. Proceedings of the Royal Society of London B: Biological Sciences 273: 2305-2312.
50. Hume BC, D'Angelo C, Smith EG, Stevens JR, Burt J, Wiedenmann J (2015) *Symbiodinium thermophilum* sp. nov., a thermotolerant symbiotic alga prevalent in corals of the world's hottest sea, the Persian/Arabian Gulf. Scientific reports 5: 8562.
51. Thornhill D, Howells E, Wham D, Steury T, Santos S (2017) Population genetics of reef coral endosymbionts (*Symbiodinium*, Dinophyceae). Molecular Ecology.
52. Little AF, Van Oppen MJ, Willis BL (2004) Flexibility in algal endosymbioses shapes growth in reef corals. Science 304: 1492-1494.
53. Thornhill DJ, Kemp DW, Bruns BU, Fitt WK, Schmidt GW (2008) Correspondence between Cold Tolerance and Temperate Biogeography in a Western Atlantic *Symbiodinium* (Dinophyta) Lineage1. Journal of Phycology 44: 1126-1135.
54. Jones AM, Berkelmans R, van Oppen MJ, Mieog JC, Sinclair W (2008) A community change in the algal endosymbionts of a scleractinian coral following a natural bleaching event: field evidence of acclimatization. Proceedings of the Royal Society of London B: Biological Sciences 275: 1359-1365.
55. Lajeunesse TC, Smith RT, Finney J, Oxenford H (2009) Outbreak and persistence of opportunistic symbiotic dinoflagellates during the 2005 Caribbean mass coral 'bleaching' event. Proceedings of the Royal Society of London B: Biological Sciences 276: 4139-4148.
56. Jones AM, Berkelmans R (2011) Tradeoffs to thermal acclimation: energetics and reproduction of a reef coral with heat tolerant *Symbiodinium* type-D. Journal of Marine Biology 2011.

57. Cuning R, Gillette P, Capo T, Galvez K, Baker AC (2015) Growth tradeoffs associated with thermotolerant symbionts in the coral *Pocillopora damicornis* are lost in warmer oceans. *Coral Reefs* 34: 155-160. doi: 10.1007/s00338-014-1216-4
58. Cantin NE, van Oppen MJ, Willis BL, Mieog JC, Negri AP (2009) Juvenile corals can acquire more carbon from high-performance algal symbionts. *Coral Reefs* 28: 405.
59. Kennedy EV, Foster NL, Mumby PJ, Stevens JR (2015) Widespread prevalence of cryptic *Symbiodinium* D in the key Caribbean reef builder, *Orbicella annularis*. *Coral Reefs* 34: 519-531.
60. Diekmann O, Bak R, Tonk L, Stam W, Olsen J (2002) No habitat correlation of zooxanthellae in the coral genus *Madraca* on a Curacao reef. *Marine Ecology Progress Series* 227: 221-232.
61. Baker AC, Rowan R, Knowlton N (1997) Symbiosis ecology of two Caribbean Acroporid corals. In: Lessios, HA, Macintyre, IG (eds.) *Proceedings of the 8th International Coral Reef Symposium*, vol. 2, Panama, pp. 1296-1300.
62. Silverstein RN, Correa AMS, Baker AC (2012) Specificity is rarely absolute in coral-algal symbiosis: implications for coral response to climate change. *Proceedings of the Royal Society B: Biological Sciences* 279: 2609-2618. doi: 10.1098/rspb.2012.0055
63. Quigley KM, Davies SW, Kenkel CD, Willis BL, Matz MV, Bay LK (2014) Deep-sequencing method for quantifying background abundances of *Symbiodinium* types: exploring the rare *Symbiodinium* biosphere in reef-building corals. *Plos ONE* 9: e94297.
64. Davy SK, Allemand D, Weis VM (2012) Cell biology of cnidarian-dinoflagellate symbiosis. *Microbiology and Molecular Biology Reviews* 76: 229-261.
65. Jones A, Berkelmans R (2010) Potential Costs of Acclimatization to a Warmer Climate: Growth of a Reef Coral with Heat Tolerant vs. Sensitive Symbiont Types. *Plos ONE* 5: e10437. doi: 10.1371/journal.pone.0010437
66. Correa AMS, McDonald MD, Baker AC (2009) Development of clade-specific *Symbiodinium* primers for quantitative PCR (qPCR) and their application to detecting clade D symbionts in Caribbean corals. *Marine Biology* 156: 2403-2411. doi: 10.1007/s00227-009-1263-5
67. Lee MJ, Jeong HJ, Jang SH, Lee SY, Kang NS, Lee KH, Kim HS, Wham DC, LaJeunesse TC (2016) Most Low-Abundance "Background" *Symbiodinium* spp. Are Transitory and Have Minimal Functional Significance for Symbiotic Corals. *Microbial Ecology*: 1-13.
68. Garren M, Walsh SM, Caccone A, Knowlton N (2006) Patterns of association between *Symbiodinium* and members of the *Montastraea annularis* species complex on spatial scales ranging from within colonies to between geographic regions. *Coral Reefs* 25: 503-512.
69. Kemp DW, Thornhill DJ, Rotjan RD, Iglesias-Prieto R, Fitt WK, Schmidt GW (2015) Spatially distinct and regionally endemic *Symbiodinium* assemblages in the threatened Caribbean reef-building coral *Orbicella faveolata*. *Coral Reefs* 34: 535-547.
70. Warner ME, LaJeunesse TC, Robison JD, Thur RM (2006) The ecological distribution and comparative photobiology of symbiotic dinoflagellates from reef corals in Belize: Potential implications for coral bleaching. *Limnology and Oceanography* 51: 1887-1897.
71. Green EA, Davies SW, Matz MV, Medina M (2014) Quantifying cryptic *Symbiodinium* diversity within *Orbicella faveolata* and *Orbicella franksi* at the Flower Garden Banks, Gulf of Mexico. *PeerJ* 2: e386.
72. Pettay DT, Wham DC, Smith RT, Iglesias-Prieto R, LaJeunesse TC (2015) Microbial invasion of the Caribbean by an Indo-Pacific coral zooxanthella. *Proceedings of the National Academy of Sciences* 112: 7513-7518.
73. Kennedy EV, Tonk L, Foster NL, Chollett I, Ortiz J-C, Dove S, Hoegh-Guldberg O, Mumby PJ, Stevens JR (2016) *Symbiodinium* biogeography tracks environmental patterns rather than host genetics in a key Caribbean reef-builder, *Orbicella annularis*. *Proc R Soc B*, vol. 283. The Royal Society, pp. 20161938.

74. Miller J, Muller E, Rogers C, Waara R, Atkinson A, Whelan K, Patterson M, Witcher B (2009) Coral disease following massive bleaching in 2005 causes 60% decline in coral cover on reefs in the US Virgin Islands. *Coral Reefs* 28: 925.
75. Lirman D, Manzello D (2009) Patterns of resistance and resilience of the stress-tolerant coral *Siderastrea radians* (Pallas) to sub-optimal salinity and sediment burial. *Journal of Experimental Marine Biology and Ecology* 369: 72-77.
76. Lirman D, Fong P (2007) Is proximity to land-based sources of coral stressors an appropriate measure of risk to coral reefs? An example from the Florida Reef Tract. *Marine Pollution Bulletin* 54: 779-791.
77. Lirman D, Manzello D, Maciá S (2002) Back from the dead: the resilience of *Siderastrea radians* to severe stress. *Coral Reefs* 21: 291-292.
78. Castillo KD, Ries JB, Weiss JM (2011) Declining coral skeletal extension for forereef colonies of *Siderastrea siderea* on the Mesoamerican Barrier Reef System, Southern Belize. *Plos ONE* 6: e14615.
79. Guzman HM, Tudhope AW (1998) Seasonal variation in skeletal extension rate and stable isotopic ($^{13}\text{C}/^{12}\text{C}$ and $^{18}\text{O}/^{16}\text{O}$) composition in response to several environmental variables in the Caribbean reef coral *Siderastrea siderea*. *Marine Ecology Progress Series* 166: 109-118.
80. Darling ES, Alvarez-Filip L, Oliver TA, McClanahan TR, Côté IM (2012) Evaluating life-history strategies of reef corals from species traits. *Ecology Letters* 15: 1378-1386.
81. Baumann JH, Townsend JE, Courtney TA, Aichelmann HE, Davies SW, Lima FP, Castillo KD (2016) Temperature Regimes Impact Coral Assemblages along Environmental Gradients on Lagoonal Reefs in Belize. *Plos ONE* 11: e0162098. doi: 10.1371/journal.pone.0162098
82. Simons R (2011) ERDDAP- The Environmental Research Division's Data Access Program. .
83. Aronson RB, Precht WF, Toscano MA, Koltes KH (2002) The 1998 bleaching event and its aftermath on a coral reef in Belize. *Marine Biology* xxx.
84. Gleeson M, Strong A (1995) Applying MCSST to coral reef bleaching. *Advances in Space Research* 16: 151-154.
85. Davies SW, Rahman M, Meyer E, Green EA, Buschiazzi E, Medina M, Matz MV (2013) Novel polymorphic microsatellite markers for population genetics of the endangered Caribbean star coral, *Montastraea faveolata*. *Marine Biodiversity* 43: 167-172.
86. Chomczynski P, Sacchi N (2006) The single-step method of RNA isolation by acid guanidinium thiocyanate-phenol-chloroform extraction: twenty-something years on. *Nature protocols* 1: 581-585.
87. Stat M, Loh WKH, Hoegh-Guldberg O, Carter DA (2009) Stability of coral-endosymbiont associations during and after a thermal stress event in the southern Great Barrier Reef. *Coral Reefs* 28: 709-713.
88. Klepac CN, Beal J, Kenkel CD, Sproles A, Polinski JM, Williams MA, Matz MV, Voss JD (2015) Seasonal stability of coral-Symbiodinium associations in the subtropical coral habitat of St. Lucie Reef, Florida. *Marine Ecology Progress Series* 532: 137-151.
89. Li W, Fu L, Niu B, Wu S, Wooley J (2012) Ultrafast clustering algorithms for metagenomic sequence analysis. *Briefings in bioinformatics*: bbs035.
90. Guindon S, Gascuel O (2003) A simple, fast, and accurate algorithm to estimate large phylogenies by maximum likelihood. *Systematic Biology* 52: 696-704.
91. Guindon S, Dufayard J-F, Lefort V, Anisimova M, Hordijk W, Gascuel O (2010) New algorithms and methods to estimate maximum-likelihood phylogenies: assessing the performance of PhyML 3.0. *Systematic Biology* 59: 307-321.

92. Kears M, Moir R, Wilson A, Stones-Havas S, Cheung M, Sturrock S, Buxton S, Cooper A, Markowitz S, Duran C (2012) Geneious Basic: an integrated and extendable desktop software platform for the organization and analysis of sequence data. *Bioinformatics* 28: 1647-1649.
93. Darriba D, Taboada GL, Doallo R, Posada D (2012) jModelTest 2: more models, new heuristics and parallel computing. *Nature methods* 9: 772-772.
94. Dereeper A, Guignon V, Blanc G, Audic S, Buffet S, Chevenet F, Dufayard J-F, Guindon S, Lefort V, Lescot M (2008) Phylogeny. fr: robust phylogenetic analysis for the non-specialist. *Nucleic acids research* 36: W465-W469.
95. Dereeper A, Audic S, Claverie J-M, Blanc G (2010) BLAST-EXPLORER helps you building datasets for phylogenetic analysis. *BMC evolutionary biology* 10: 8.
96. Chevenet F, Brun C, Bañuls A-L, Jacq B, Christen R (2006) TreeDyn: towards dynamic graphics and annotations for analyses of trees. *BMC bioinformatics* 7: 439.
97. Team RC (2017) R: A language and environment for statistical computing. R Foundation for Statistical Computing Vienna, Austria.
98. Thornhill DJ, LaJeunesse TC, Santos SR (2007) Measuring rDNA diversity in eukaryotic microbial systems: how intragenomic variation, pseudogenes, and PCR artifacts confound biodiversity estimates. *Molecular Ecology* 16: 5326-5340.
99. Matz MV, Wright RM, Scott JG (2013) No control genes required: Bayesian analysis of qRT-PCR data. *Plos ONE* 8: e71448.
100. Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara R, Simpson GL, Solymos P, Stevens M, Wagner H (2013) Package 'vegan'. *R Packag ver* 254: 20-28.
101. Ter Braak CJ (1986) Canonical correspondence analysis: a new eigenvector technique for multivariate direct gradient analysis. *Ecology* 67: 1167-1179.
102. Finney JC, Pettay, D.T., Sampayo, E.M., Warner, M.E., Oxenford, H.A. and T.C. LaJeunesse (2010) On the relative significance of host-habitat, irradiance, and dispersal in the ecological distribution and speciation of coral endosymbionts. *Microbial Ecology* In Press.
103. LaJeunesse TC, Loh WK, Van Woesik R, Hoegh-Guldberg O, Schmidt GW, Fitt WK (2003) Low symbiont diversity in southern Great Barrier Reef corals, relative to those of the Caribbean. *Limnology and Oceanography* 48: 2046-2054.
104. Correa AMS, Brandt ME, Smith TB, Thornhill DJ, Baker AC (2009) Symbiodinium associations with diseased and healthy scleractinian corals. *Coral Reefs* 28: 437-448. doi: 10.1007/s00338-008-0464-6
105. Thornhill DJ, LaJeunesse TC, Kemp DW, Fitt WK, Schmidt GW (2006) Multi-year, seasonal genotypic surveys of coral-algal symbioses reveal prevalent stability or post-bleaching reversion *Marine Biology* 148: 711-722.
106. Thornhill DJ, Xiang Y, Fitt WK, Santos SR (2009) Reef endemism, host specificity and temporal stability in populations of symbiotic dinoflagellates from two ecologically dominant Caribbean corals. *Plos ONE* 4: e6262.
107. LaJeunesse TC, Lee S, Bush S, Bruno JF (2005) Persistence of non-Caribbean algal symbionts in Indo-Pacific mushroom corals released to Jamaica 35 years ago. *Coral Reefs* 24: 157-159.
108. Santos S, Shearer T, Hannes A, Coffroth M (2004) Fine-scale diversity and specificity in the most prevalent lineage of symbiotic dinoflagellates (Symbiodinium, Dinophyceae) of the Caribbean. *Molecular Ecology* 13: 459-469.
109. Bongaerts P, Riginos C, Ridgway T, Sampayo EM, van Oppen MJ, Englebert N, Vermeulen F, Hoegh-Guldberg O (2010) Genetic divergence across habitats in the widespread coral *Seriatopora hystrix* and its associated Symbiodinium. *Plos ONE* 5: e10871.
110. Tong H, Cai L, Zhou G, Yuan T, Zhang W, Tian R, Huang H, Qian P-Y (2017) Temperature shapes coral-algal symbiosis in the South China Sea. *Scientific reports* 7: 40118. doi: 10.1038/srep40118

- <http://www.nature.com/articles/srep40118#supplementary-information>
111. Buddemeier RW, Fautin DG (1993) Coral bleaching as an adaptive mechanism. *BioScience* 43: 320-326.
 112. Glynn PW, Mate JL, Baker AC, Calderon MO (2001) Coral bleaching and mortality in Panama and Ecuador during the 1997-1998 El Nino Southern Oscillation event: spatial/temporal patterns and comparisons with the 1982-1983 event. *Bulletin of Marine Science* 69: 79-109.
 113. Frade P, Englebert N, Faria J, Visser P, Bak R (2008) Distribution and photobiology of Symbiodinium types in different light environments for three colour morphs of the coral *Madracis pharensis*: is there more to it than total irradiance? *Coral Reefs* 27: 913-925.
 114. Ulstrup KE, Ralph PJ, Larkum AWD, Kuhl M (2006) Intra-colonial variability in light acclimation of zooxanthellae in coral tissues of *Pocillopora damicornis*. *Marine Biology* 149: 1325 - 1335.
 115. Ulstrup KE, Van Oppen M (2003) Geographic and habitat partitioning of genetically distinct zooxanthellae (Symbiodinium) in *Acropora* corals on the Great Barrier Reef. *Molecular Ecology* 12: 3477-3484.
 116. Toller WW, Rowan R, Knowlton N (2001) Repopulation of zooxanthellae in the Caribbean corals *Montastraea annularis* and *M. faveolata* following experimental and disease-associated bleaching. *Biological Bulletin* 201: 360-373.
 117. Hennige SJ, Smith DJ, Walsh S-J, McGinley MP, Warner ME, Suggett DJ (2010) Acclimation and adaptation of scleractinian coral communities along environmental gradients within an Indonesian reef system. *Journal of Experimental Marine Biology and Ecology* 391: 143-152. doi: <http://dx.doi.org/10.1016/j.jembe.2010.06.019>
 118. Abrego D, Van Oppen MJ, Willis BL (2009) Highly infectious symbiont dominates initial uptake in coral juveniles. *Molecular Ecology* 18: 3518-3531.
 119. Fabricius K, Mieog J, Colin P, Idip D, H VAN OPPEN M (2004) Identity and diversity of coral endosymbionts (zooxanthellae) from three Palauan reefs with contrasting bleaching, temperature and shading histories. *Molecular Ecology* 13: 2445-2458.
 120. Lajeunesse TC, Pettay DT, Sampayo EM, Phongsuwan N, Brown B, Obura DO, Hoegh-Guldberg O, Fitt WK (2010) Long-standing environmental conditions, geographic isolation and host-symbiont specificity influence the relative ecological dominance and genetic diversification of coral endosymbionts in the genus Symbiodinium. *Journal of Biogeography* 37: 785-800.
 121. Baker AC (2001) Reef corals bleach to survive change. *Nature* 411: 765-766.
 122. Paris CB, Cherubin LM (2008) River-reef connectivity in the Meso-American Region. *Coral Reefs* 27: 773-781.
 123. Carilli JE, Norris RD, Black BA, Walsh SM, McField M (2009) Local stressors reduce coral resilience to bleaching. *Plos ONE* 4: e6324.
 124. Carilli JE, Prouty NG, Huguen KA, Norris RD (2009) Century-scale records of land-based activities recorded in Mesoamerican coral cores. *Marine Pollution Bulletin* 58: 1835-1842.
 125. Howells EJ, Berkelmans R, van Oppen MJ, Willis BL, Bay LK (2013) Historical thermal regimes define limits to coral acclimatization. *Ecology* 94: 1078-1088.
 126. Howells E, Beltran V, Larsen N, Bay L, Willis B, Van Oppen M (2012) Coral thermal tolerance shaped by local adaptation of photosymbionts. *Nature Climate Change* 2: 116-120.
 127. Hume BC, Voolstra CR, Arif C, D'Angelo C, Burt JA, Eyal G, Loya Y, Wiedenmann J (2016) Ancestral genetic diversity associated with the rapid spread of stress-tolerant coral symbionts in response to Holocene climate change. *Proceedings of the National Academy of Sciences* 113: 4416-4421.

Tables and Figures

Table 1: Sampling locations and microsamples information for *S. siderea* (SSID), *S. radians* (SRAD), and *P. strigosa* (PSTR). Locations are listed in order of descending latitude (Northernmost to Southernmost).

Transect	Thermal regime	Collection Date	Illumina Lane	Lat (°N)	Long (°W)	SSID	SRAD	PSTR
Belize City	Low	Nov 2014	2	17.64363	88.0264	n=10	n=0	n=0
Belize City	High	Oct 2015	2	17.48685	88.1207	n=10	n=0	n=0
Dangriga	Low	Nov 2014	2	17.078	88.01285	n=9	n=0	n=0
Dangriga	High	Nov 2014	2	16.79491	88.27699	n=10	n=0	n=0
Placencia	Low	Nov 2014	1	16.45816	88.01295	n=7	n=7	n=5
Placencia	Mod	Nov 2014	1	16.49995	88.16527	n=6	n=7	n=6
Placencia	High	Nov 2014	1	16.4654	88.31315	n=9	n=9	n=5
Sapodilla	Low	Nov 2014	1	16.15729	88.25073	n=8	n=0	n=0
Sapodilla	Mod	Nov 2014	1	16.13013	88.33234	n=6	n=0	n=6
Sapodilla	High	Nov 2014	1	16.2245	88.62943	n=8	n=6	n=0

Table 2: Average number of raw reads, trimmed reads, and mapped reads including mapping efficiency (% of trimmed reads that mapped) for each species.

Species	Raw reads	Trimmed reads	Mapped reads	Mapping efficiency
<i>S. siderea</i>	46161	28453	22048	73%
<i>S. radians</i>	51081	46812	35290	75%
<i>P. strigosa</i>	88888	43928	31873	69%
Total	186130	118834	89211	75%

Figure Legends

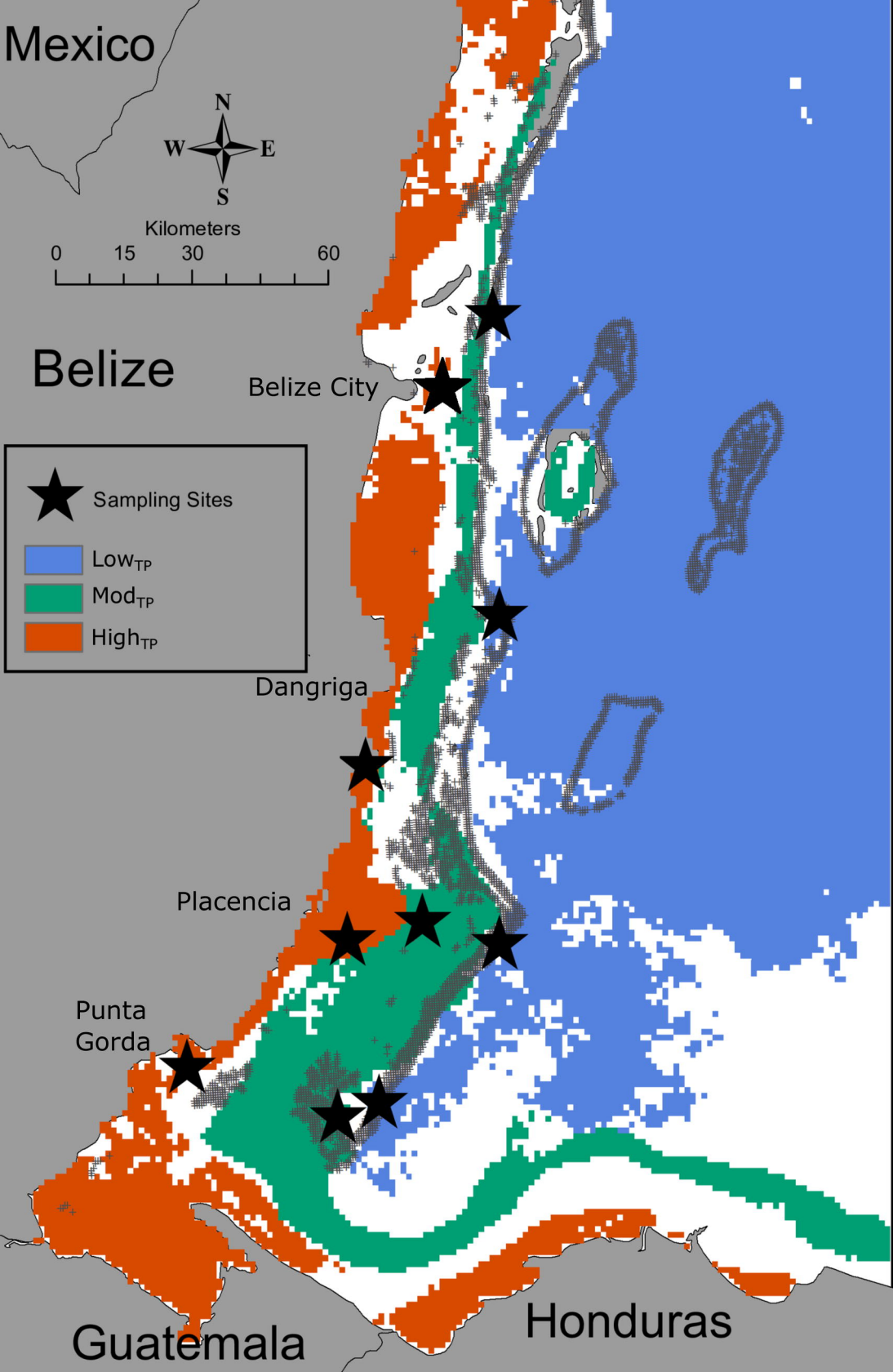
Fig 1: Thermal regime designations for sampling sites on the Belize MBRS [81]. Stars indicate sites where coral tissue samples were collected for *Symbiodinium* community analysis. Low_{TP}, mod_{TP}, and high_{TP} are defined based on combined averages of annual maximum temperature, annual temperature range, annual days above the bleaching threshold, and annual longest streak of consecutive days above the bleaching threshold. Low_{TP} sites exhibit the lowest values for all parameters measured and high_{TP} sites exhibit the highest. A more detailed description of classification of these thermal regimes can be found in Baumann et al. [81].

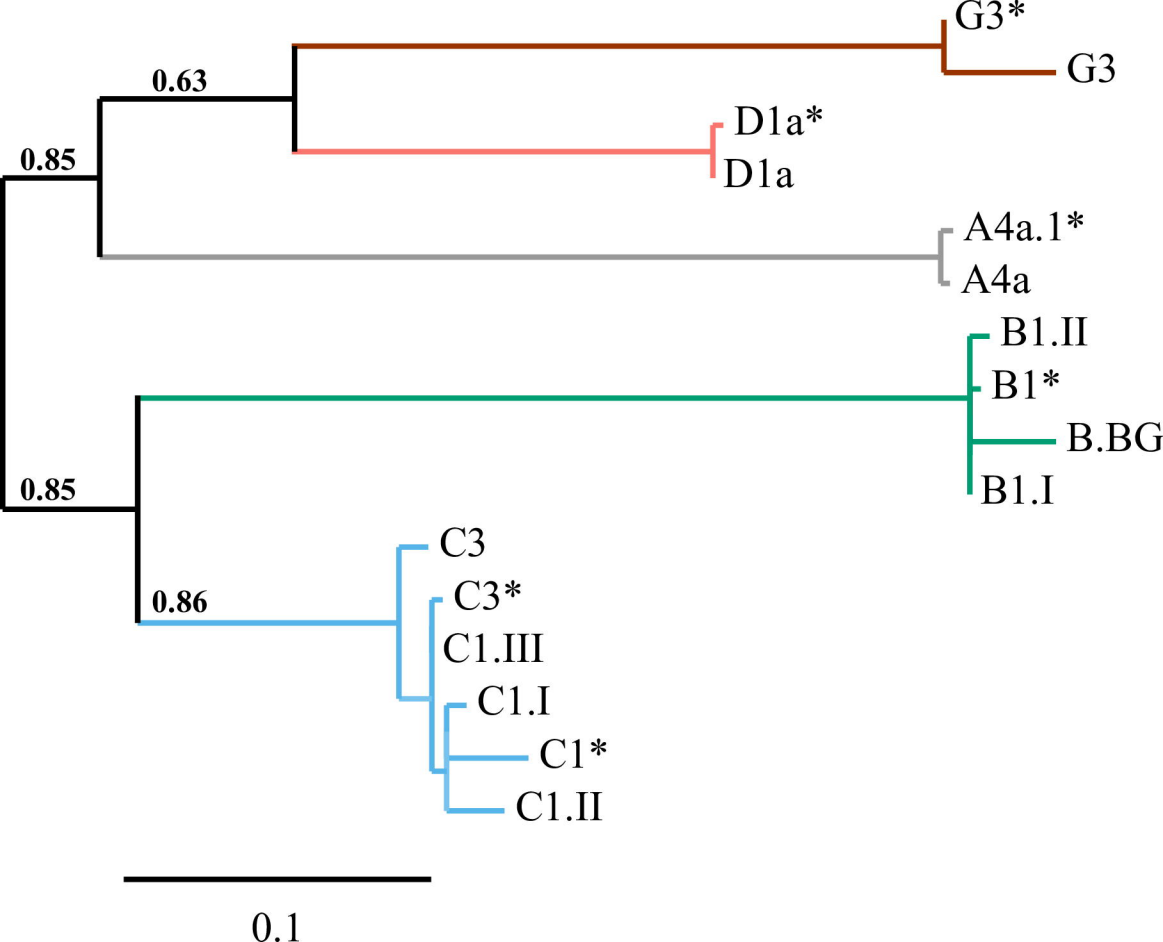
Fig 2: Phylogenetic analysis of ITS-2 sequences of representative OTUs from this study in addition to reference sequences for each clade (indicated by *). Branch support values are shown on the branches at divisions between distinct clades. The scale bar represents replacements per nucleotide site.

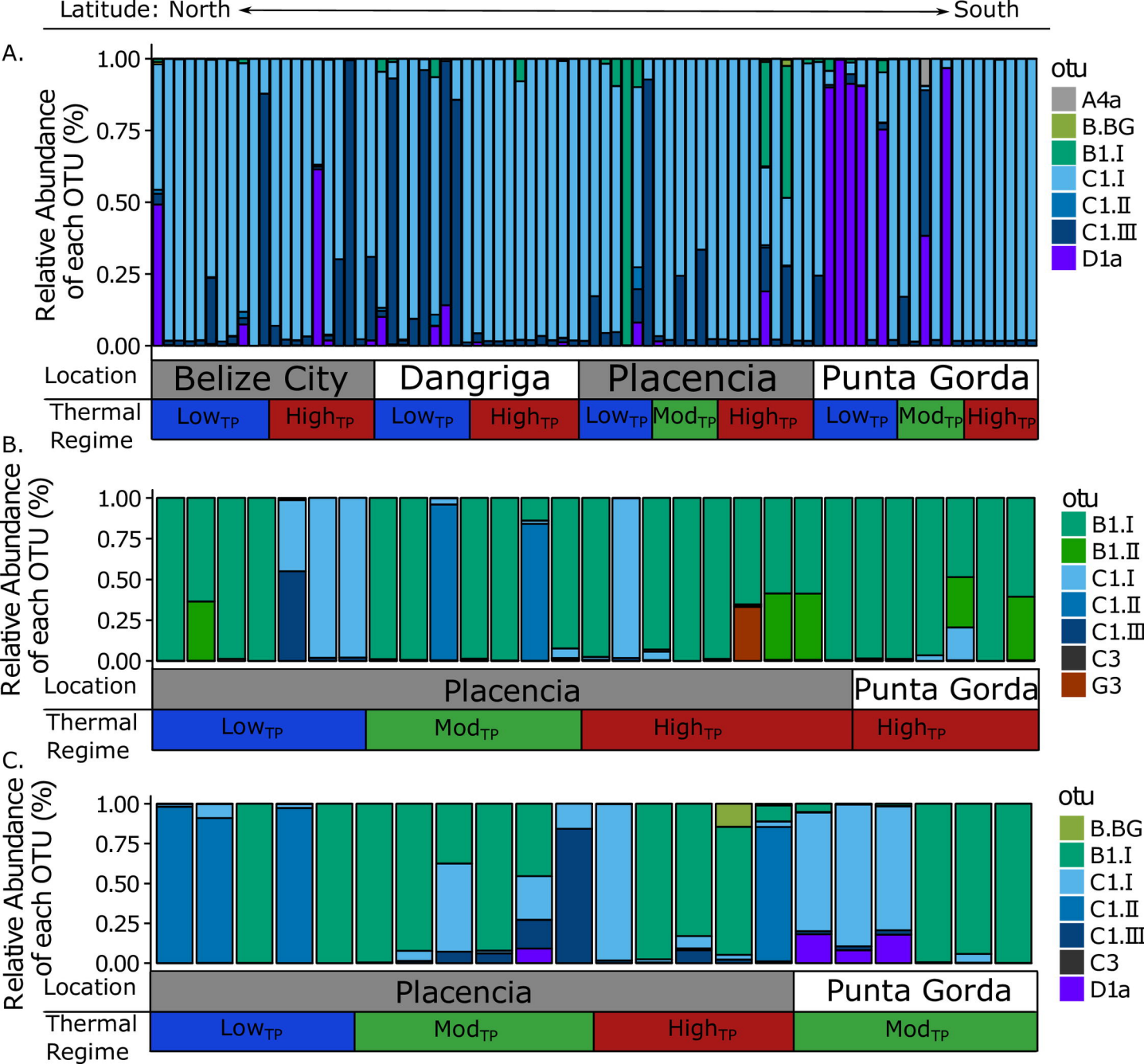
Fig 3. Relative abundance (%) of each OTU (lineage) in *S. siderea* (A), *S. radians* (B), and *P. strigosa* (C). Each column represents an individual sample. Columns are arranged by latitudinal transect (as indicated by site names in alternating gray and white boxes) and then by thermal regime (blue boxes indicate low_{TP} sites, green boxes indicate mod_{TP} sites, and red boxes indicates high_{TP} sites).

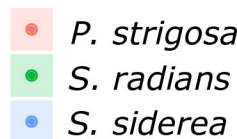
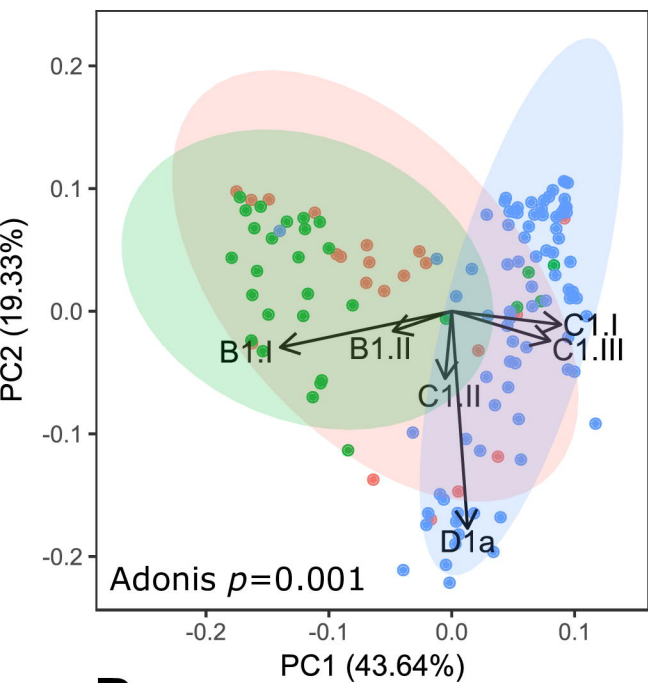
Fig 4. Principal component analysis (PCA) plots of *Symbiodinium* communities by species (A) and by thermal regime for *S. siderea* (B). Percentages on each axis indicate the amount of variation explained by each axis. Adonis *p-values* indicate significant results of PERMANOVA tests. See Table S4 for additional PERMANOVA results. Black arrows indicate loadings showing the magnitude and direction of the effect of each OTU on the total variance. Colored ellipses indicate 95% confidence intervals.

Fig 5. Canonical correlation analysis (CCA) showing relationship between thermal parameters (Table S1), *Symbiodinium* lineages, and *Symbiodinium* communities within *S. siderea* (A), *S. radians* (B), and *P. strigosa* (C). CCA scores for each sample are represented by a filled circle (colored by thermal regime) and scores for each lineage are denoted by the name of each lineage. Percentages on each axis indicate the total variation explained by that axis.







A.**B.**