

1 Short Title: Shifts in temperature and *Batrachochytrium*

2 **Shifts in temperature influence how *Batrachochytrium dendrobatidis* infects**
3 **amphibian larvae**

4 Paul W. Bradley¹, Michael D. Brawner², Thomas R. Raffel³, Jason R. Rohr⁴, Deanna H.
5 Olson⁵, and Andrew R. Blaustein²

6 ¹ Department of Biology, University of San Diego, 5998 Alcala Park, San Diego CA,
7 93110, USA.

8 ² Department of Integrative Biology, 3029 Cordley Hall, Oregon State University,
9 Corvallis, OR, 97331, USA.

10 ³ Department of Biology, 375 Dodge Hall, Oakland University, Rochester, MI, 48309,
11 USA.

12 ⁴ Department of Integrative Biology, University of South Florida, 4202 East Fowler
13 Avenue, Tampa, FL, 33620, USA.

14 ⁵ USDA Forest Service, Pacific Northwest Research Station, 3200 SW Jefferson Way,
15 Corvallis, OR, 97331, USA.

16

17 Corresponding author: paulwilliambradley@gmail.com

18

19 **Author contribution statement**

20 ARB, JRR, and TRR originally formulated the idea. PWB designed the experiment and developed the
21 methodology. PWB performed the experiment. PWB and MDB performed the molecular analysis. PWB
22 and TRR performed the statistical analyses. ARB, JRR, TRR, and DHO obtained funding. PWB wrote the
23 manuscript and other authors provided editorial advice.

24

Abstract:

Many climate change models predict increases in mean temperature, and increases in frequency and magnitude of temperature fluctuations. These potential shifts may impact ectotherms in several ways, including how they are affected by disease. Shifts in temperature may especially affect amphibians, a group with populations that have been challenged by several pathogens. Because amphibian hosts invest more in immunity at warmer than cooler temperatures and parasites may acclimate to temperature shifts faster than hosts (creating lags in optimal host immunity), researchers have hypothesized that a temperature shift from cold-to-warm might result in increased amphibian sensitivity to pathogens, whereas a shift from warm-to-cold might result in decreased sensitivity. Support for components of this climate-variability based hypothesis have been provided by prior studies of the fungus *Batrachochytrium dendrobatidis* (Bd) that causes the disease chytridiomycosis in amphibians. We experimentally tested whether temperature shifts before Bd exposure alter susceptibility to Bd in the larval stage of two amphibian species – western toads (*Anaxyrus boreas*) and northern red legged frogs (*Rana aurora*). Both host species harbored elevated Bd infection intensities under constant cold (15° C) temperature in comparison to constant warm (20° C) temperature. Additionally, both species experienced an increase in Bd infection abundance when shifted to 20° C from 15° C, compared to a constant 20° C but they experienced a decrease in Bd when shifted to 15° C from 20° C, compared to a constant 15° C. These results are in contrast to prior studies of adult amphibians that found increased susceptibility to Bd infection after a temperature shift in either direction, highlighting the potential for species and stage differences in the temperature-dependence of chytridiomycosis.

Keywords: amphibian declines, *Batrachochytrium dendrobatidis*, chytridiomycosis, climate variability hypothesis, infectious disease, temperature, *Rana aurora*, *Anaxyrus boreas*

Introduction

Climate change represents one of the greatest challenges to biodiversity and conservation because it might compromise ecosystem functions worldwide. Most studies of climate-change induced effects on ecological communities emphasize the role of predicted changes to annual or seasonal mean temperatures or precipitation (Paaijmans, Read & Thomas 2009, Paaijmans *et al.* 2010). However, many climate change models predict increases in the frequency and magnitude of extreme weather events such as heat waves and droughts (Schar *et al.* 2004, Horton *et al.* 2016) leading to increases in temperature variability at shorter timescales (Easterling *et al.* 2000, Rummukainen 2012). These predicted climate-change induced increases in short-term temperature fluctuations can affect species interactions (Hoover, Knapp & Smith 2014, Teskey *et al.* 2015). Yet few studies have investigated how increases in temperature variability on these shorter timescales affect disease dynamics despite the likelihood that such variability might differentially affect hosts and pathogens (Ben-Horin, Lenihan & Lafferty 2012, Bannerman & Roitberg 2014, Raffel *et al.* 2015).

The aquatic chytrid fungal pathogen *Batrachochytrium dendrobatidis* (Bd) causes chytridiomycosis, an emerging infectious disease of amphibians (Longcore, Pessier & Nichols 1999). Bd is widespread globally (Liu, Rohr & Li 2013, Olson *et al.* 2013) and is

associated with worldwide amphibian population declines (Stuart *et al.* 2004, Skerratt *et al.* 2007). Moreover, models based on IPCC climate projections predict that the range of Bd will increase, thus potentially making more amphibian species at risk (Xie, Olson & Blaustein 2016).

In larvae, Bd infection can cause host mortality in some species (Blaustein *et al.* 2005, Garner *et al.* 2009). However the infection is localized to keratinized larval mouthparts, (Marantelli *et al.* 2004, McMahon & Rohr 2015) often resulting in sublethal effects (Han, Bradley & Blaustein 2008, Buck *et al.* 2012, Gervasi *et al.* 2013).

Temperature is considered one of the most important environmental factors driving chytridiomycosis (Forrest & Schlaepfer 2011, Voyles *et al.* 2017). Because physiologies of both the host and pathogen are strongly influenced by environmental temperature, climate change has been used to explain several major Bd outbreaks and amphibian population declines, (reviewed in Li, Cohen & Rohr 2013, Rohr *et al.* 2013). Yet, the host and pathogen are not expected to share a uniform response to a given temperature (Brown *et al.* 2004, Rohr *et al.* 2013) and thermal responses measured in constant-temperature artificial environments might not reflect organism responses in more realistic variable-temperature environments. Providing evidence of the lack of a uniform response between Bd and amphibians to temperature shifts, Rohr and Raffel (2010) found a strong correlation between elevated month-to-month temperature variability and Bd-associated amphibian population declines of *Atelopus* spp. across Central and South America. Further support of the relationship between chytridiomycosis and temperature variability has been provided by laboratory studies (Raffel *et al.* 2013, Raffel *et al.* 2015).

The potential for temperature variability to increase disease severity in amphibians was first postulated by Raffel *et al.* (2006) and has subsequently been referred to as the “climate variability hypothesis” (Rohr & Raffel 2010). This hypothesis posits that parasites acclimate to the new temperature more rapidly than their hosts, leading to lags in host acclimation following a temperature shift that could make hosts more susceptible to infection (Raffel *et al.* 2013). However, Raffel *et al.* (2006) also pointed out potential complexities in acclimation of the ectotherm immune system that might lead to alternative predictions. According to the “lag effect” hypothesis (Raffel *et al.* 2006), changes in levels of temperature-dependent immune parameters might simply lag behind environmental temperature shifts (Fig. 1) because it takes time to produce necessary, or remove unnecessary, immune cells from the host. Thus, the “lag effect” hypothesis predicts the opposite effect from the “climate variability hypothesis” following a temperature decrease, at least on a short timescale. These mechanistic hypotheses are not mutually exclusive, and it is unclear which effects might be more important for a given host-parasite combination.

We tested the general prediction that an amphibian shifted to a new temperature before Bd exposure would respond to infection differently than a host already acclimated to the exposure temperature. We postulated that the direction of the effect would depend upon the direction of the temperature shift, in accordance with the “lag effect” hypothesis of Raffel *et al.* (2006). Given the differences in size between the host and the pathogen, and associated physiological process rate differences, we assumed Bd would physiologically respond to the temperature shift faster than the host, such that an idealized host-immune response to Bd exposure would temporarily lag behind the

temperature shift. Thus, we predicted that a temperature shift from cold-to-warm would result in an *increase* in susceptibility to Bd exposure, whereas a temperature shift from warm-to-cold would result in a *decrease* in susceptibility to Bd exposure.

Materials and Methods

In a laboratory study under controlled conditions necessary for increased precision and to allow for better replication, we examined how temperature shifts may alter larval amphibian infection dynamics. We selected two species of amphibian hosts, the northern red legged frog (*Rana aurora*) and the western toad (*Anaxyrus boreas*) as adults of both species have been observed in the field with Bd infections (Pearl *et al.* 2007, Piovia-Scott *et al.* 2011) and both species are susceptible to chytridiomycosis (Han, Bradley & Blaustein 2008, Gervasi *et al.* 2013). Amphibians were collected as eggs from natural oviposition sites where Bd is not known to be endemic. Red legged frog eggs were collected from a permanent pond located near Florence, Oregon, USA (Lincoln County, elevation 12 m; latitude/longitude: 44.088/-124.123) in the Oregon Coast Range on 11-Feb-2012. Western toad eggs were collected from Little Three Creeks Lake (Deschutes County, elevation 2,000 m; latitude/longitude: 44.009/-121.643) in the Cascade Range on 9-Jul-2011. Immediately after collection, eggs were transported to a laboratory at Oregon State University where they were maintained at 14° C, under a 12-12 photoperiod in 40-liter aquaria filled with dechlorinated water. Upon hatching, larvae were maintained at a density of approximately 200 individuals per aquarium and fed *ad libitum* a mixture of Tetramin fish food and ground alfalfa pellets (1:3 ratio by volume). Water was changed every seven days. The 40-day trials for each species were not run

concurrently, but identical protocols were used for both species and both trials consisted of individuals of identical larval stage (Gosner stage 26).

Acclimation Period

Independent trials for each host species began with a 20-day acclimation period with 80 (Gosner stage 26) larvae randomly selected, and individually placed into 80 plastic 500-mL containers where they were housed for the duration of the acclimation period and experiment. Each container was filled with 14° C dechlorinated water and covered with a lid to help maintain water temperature and limit evaporation. Each container had 2-mm diameter holes drilled between the water line and the lid to allow air circulation into the container. Pairs of containers were then placed within 40 individual temperature-controlled chambers that were set at 15° C to avoid cold-shocking the larvae. Each temperature-controlled chamber was independently controlled via its own thermostat and the interior measured approximately 37 cm deep x 21 cm wide x 13 cm in height. Half of the 40 temperature-controlled chambers were then randomly selected to begin the acclimation period at 20° C (warm treatment) and the other half were kept at 15° C (cold treatment). The placement of temperature chambers within the laboratory was randomized, as was the placement of 500-mL containers within each temperature chamber.

Temperature Shifts

On day 20 of the experiment, half of the temperature chambers in each of the two acclimation temperatures (15° C and 20° C) were randomly selected to undergo a temperature shift, either from 20° to 15° C or from 15° C to 20° C. The other half of the temperature chambers underwent no shift in temperature. Thus, each of the temperature

chambers was subjected to one of four temperature treatments: a constant 15° C (cold) throughout the experiment; a constant 20° C (warm) throughout the experiment; a temperature shift from 15° C to 20° C (cold-to-warm); or a temperature shift from 20° C to 15°C (warm-to-cold).

Bd exposure

On day 24, one of the two 500-mL containers within each temperature-controlled chamber was randomly selected to undergo a *Bd*-exposure treatment and the other was selected as a control. Larvae in the *Bd*-exposure treatment were exposed to a single inoculate of *Bd* strain JEL 274, which was grown in pure culture on 1% tryptone agar in 10-cm diameter Petri dishes for 10 days. To harvest the zoospores, 10 plates were flushed with 15 mL of 15° C dechlorinated water and remained undisturbed for 10 minutes. The inoculum from each plate was then pooled in a beaker, quantified, and then diluted to 10,000 zoospores/mL. Individuals in the *Bd*-exposed treatments were exposed to 10 mL of inoculum transferred into the 500-mL container housing an individual larva. Control individuals were exposed to 10 mL of sham inoculum lacking the *Bd* culture (made from 1% tryptone sterile agar plates following the same methods). Thus, the individual larva underwent their exposure treatment on day 24, four days after the water temperature shift for chambers in the two temperature shift treatments.

During the 40-d trial, survival and metamorphic status were checked daily. Water for each 500-mL container within the temperature chambers was changed every 12 days and consisted of dechlorinated water of the same temperature (15° C and 20° C). Individuals that survived until the end of the trial (i.e., day 40) were euthanized in a 2% solution of MS-222, and then preserved in 95% ethanol. Individuals that reached

metamorphosis (Gosner stage 42: emergence of forelimbs) were euthanized, measured, and preserved as previously described.

Determining infection status

We used quantitative polymerase chain reaction (qPCR) to determine infection status and quantify Bd-infection intensity of all individuals in the Bd-exposure treatments. Additionally, we investigated Bd-infection status in eight randomly selected control individuals per species. To sample the individuals for Bd, we extracted whole mouthparts of the larvae using sterile dissection scissors. We conducted qPCR using an ABI PRISM 7500 sequencer (Applied Biosystems) according to the methods of Boyle *et al.* (2004) except that we used 60 μ L of Prepman Ultra (Applied Biosystems, Carlsbad, California, USA), instead of the 40 μ L in the DNA extraction.

Statistical Analyses

Each temperature-controlled chamber was subjected to one of four temperature regimes consisting of a Bd-exposure temperature combined with a temperature shift status (constant cold, constant warm, shifted to cold, and shifted to warm). Further, the pairs of containers within each temperature-controlled chamber were subjected to one of two exposure treatments (Bd exposed and Bd unexposed).

Survival was compared between temperature treatments for western toad larvae with a Cox proportional hazards model using TIBCO Spotfire S+ version 8.1. The model consisted of the main effects of the temperature treatment, temperature shift status (constant versus shifted), and an interaction between the two variables. Due to losses of western toad larvae prior to the application of the exposure treatment, we lacked the power to statistically compare survival in western toad larvae between the Bd exposure

treatments

Bd infection abundance (Bd genomic equivalents) among temperature treatments and between host species was analyzed using R version 3.1.1. We used a zero-inflated negative-binomial generalized linear model (function ‘zeroinf’ in package ‘pscl’) as described by Raffel *et al.* (2010). Our full model investigated the effects of all of the explanatory variables including host species, exposure temperature, temperature shift status, and all two- and three-way interactions on Bd abundance. Interpretation of this analysis required further reduced models to investigate the effect of exposure temperature and temperature shift for each species (species model) and the effect of temperature shift for each Bd-exposure temperature and host species combination (Bd-exposure temperature model).

Results

Survival differences were not detected between exposure temperatures (Cox, $Z = -1.099$, $p = 0.27$) or temperature shift status (Cox, $Z = -0.277$, $p = 0.78$) in Bd-exposed western toad larvae. We were unable to detect survival differences in red legged frog larvae, as only one individual larva experienced mortality after application of the exposure treatment (Table S1).

Infection Abundance

We detected a host species by temperature shift interaction ($\chi^2_1 = 3.83$, $p = 0.050$; Table S2) and a Bd-exposure temperature by temperature shift interaction ($\chi^2_1 = 7.50$, $p =$

0.006; Table S2). We investigated these interactions with reduced models to investigate effects on Bd abundance at the levels of species and exposure temperature.

Red legged frog larvae had higher Bd abundance when they were exposed to infection at 15° C when compared to 20° C ($\chi^2_1 = 3.88$, $p = 0.049$; Fig. 2). The main effect of temperature shift was marginally significant in the reduced species model analysis ($\chi^2_1 = 3.50$, $p = 0.061$), but there was a significant effect of temperature shift for individuals exposed at 20° C in the reduced Bd-exposure model ($\chi^2_1 = 5.7$, $p = 0.017$), with individuals shifted from 15° C to 20° C having higher Bd abundance than red legged frog larvae experiencing constant 20° C (Fig. 2). In contrast, there was no evidence that a temperature shift influenced Bd infection when red legged frog larvae were exposed to Bd at 15° C ($\chi^2_1 = 0.6$, $p = 0.4$; Fig. 2). There was no statistically significant interaction between exposure temperature and temperature shift for red legged frog larvae ($\chi^2_1 = 2.4$, $p = 0.13$).

We detected an interactive effect of exposure temperature and temperature shift on Bd abundance in western toad larvae ($\chi^2_1 = 5.2$, $p = 0.023$). This was driven by elevated Bd abundance in individuals under the constant 15° C temperature when compared to individuals that experienced a temperature shift from 20° to 15° C, but no evidence of an effect of shifting temperature from 15° C to 20° C (Fig. 2). There were no main effects of exposure temperature ($\chi^2_1 = 0.50$, $p = 0.5$) or temperature shift ($\chi^2_1 < 0.01$, $p = 0.9$) on Bd abundance in western toad larvae. Further, when investigating the exposure temperatures individually in the reduced Bd-exposure model, there was no evidence that a temperature shift influenced Bd infection in western toad larvae after exposure to Bd at 15° C ($\chi^2_1 = 3.4$, $p = 0.066$) or 20° C ($\chi^2_1 = 2.5$, $p = 0.11$).

We failed to find evidence that the two host species differed in response to exposure to the pathogen, leading us to conclude that general patterns for both species were similar (Fig. 2).

A number of western toad individuals died or metamorphosed before the end of the experiment (Table S2) but the model for Bd abundance on western toads was not significantly improved by adding either a variable coding whether individuals were near metamorphosis when sampled ($\chi^2_1 = 4.00$, $p = 0.150$) or a covariate indicating the sampling date ($\chi^2_1 = 3.33$, $p = 0.068$). Furthermore, neither variable qualitatively changed the contribution of exposure temperature or temperature shift status to the model. Therefore, we omitted both covariates from the final model for western toads.

Discussion

Numerous climate change models predict increases in annual or seasonal mean temperatures in many locations (IPCC 2007). These models often also predict elevated chances of extreme weather events occurring at, and over, much shorter timescales (Rummukainen 2012, Horton *et al.* 2016). Temperature shifts that may be associated with the onsets and conclusions of these weather events have the potential to affect alter species interactions – including host-pathogen interactions (Rohr & Raffel 2010, Ben-Horin, Lenihan & Lafferty 2012, Bannerman & Roitberg 2014).

Our results suggest that Bd infection dynamics in larval amphibians can be affected by a shift in water temperature before host exposure to the pathogen, and that the direction of temperature shift determines the outcome of Bd exposure. Importantly, we detected the effects of temperature shifts despite the host having a four-day head start on

acclimating to the Bd exposure temperature relative to the pathogen. This suggests that we are likely underestimating the strength of these effects and that their magnitudes might have been larger if the host and pathogen experienced the shifts concurrently, as which would most likely be expected in the field.

Amphibian species do not all respond similarly to a given Bd exposure (Searle *et al.* 2011, Gervasi *et al.* 2017). For some susceptible host species, temperature-shift induced changes in Bd abundance might alter the outcome of infection by either pushing Bd abundance over or under a tolerance threshold. Such changes in relation to pathogen abundance and pathogen tolerance may result in altering the strength of negative effects of Bd infection.

Our results were consistent with predictions of the “lag effect” hypothesis (Raffel *et al.* 2006, Rohr & Raffel 2010), and were generally consistent with previous studies showing that a shift in temperature influences Bd infection in postmetamorphic amphibians (Raffel *et al.* 2013, Raffel *et al.* 2015). In particular, our finding of decreased resistance to infection following a temperature increase (relative to warm-acclimated individuals) mirrored a laboratory study of post-metamorphic red-spotted newts (*Notophthalmus viridescens*), where juvenile newts exhibited decreased Bd resistance following a shift from 15° C to 25° C (Raffel *et al.* 2015). These findings of fluctuating temperature effects on Bd infection across amphibian taxonomic groups and life-stages suggest that effects of temperature shifts and Bd-related chytridiomycosis susceptibility might be widespread within amphibians. However, our finding of increased resistance to Bd infection following a temperature decrease (relative to cold-acclimated individuals) was opposite the pattern observed in red-spotted newts and Cuban treefrogs (Raffel *et al.*

2013, Raffel *et al.* 2015) These contrasting results suggests that there are important among-taxa or among-stage differences in the underlying mechanisms driving the effects of temperature fluctuation on Bd infection; whereas our results in pre-metamorphic life-stage of western toads and red legged frogs are consistent with the “lag effect” hypothesis, results of similar studies investigating post-metamorphic red-spotted newts and Cuban treefrogs support the “climate variability hypothesis.”

Higher Bd abundances were observed for both host species under the constant cold temperature treatment compared to the constant warm temperature treatment. These results are consistent with previous experiments that showed increased Bd abundance (Raffel *et al.* 2015) and Bd-induced mortality (Kilpatrick, Briggs & Daszak 2010, Raffel *et al.* 2015) associated with lower temperatures. This is despite Bd growing best in culture at about 23° C, which is much closer to the warm than cold temperature treatments in this experiment (Piotrowski, Annis & Longcore 2004, Woodhams *et al.* 2008). This might be because the larval immune response to Bd infection increases with increasing temperatures at a faster rate than the infectivity or growth rate of Bd (Raffel *et al.* 2013), or alternatively because of the differences between the growth rate of Bd in culture compared to the growth rate on host tissue (Venesky *et al.* 2013).

Alternatively, differences in Bd abundance between the two constant temperature treatments might be due to temperature effects on the pathogen rather than the host (Woodhams *et al.* 2008, Voyles *et al.* 2012). The Bd was cultured at 15° C; it is possible that the temperature shift experienced by the pathogen in the warm exposure treatment caused the depressed Bd abundances observed in both host species compared to the elevated Bd abundance in the cold exposure temperature treatment. A decrease in

temperature may cause an increase in the number of Bd zoospores released from zoosporangia (Woodhams *et al.* 2008), however the effect of a similar increase in temperature on Bd physiology is unclear.

In conclusion, our results provide additional evidence for climate variability affecting Bd infection in amphibians but suggest important among-taxa differences in the directionality of these effects. Our study highlights the complexity that temperature plays in determining the outcome of Bd-amphibian interactions and the role that a fluctuating temperature might play in altering these interactions. Furthermore, this study increases the diversity of amphibian species and stages that have been shown to exhibit thermal acclimation effects on disease, and the broad generality of this pattern across four disparate taxa suggests that fluctuating-temperature effects on amphibian infection may be widespread.

Acknowledgments

All applicable institutional and national guidelines for the care and use of animals were followed; this research was conducted under Oregon State University IACUC animal care and use permit 3917. Collection of amphibian eggs was approved by the Oregon Department of Fish and Wildlife (Oregon Scientific Taking Permit #006-12 issued to ARB). We thank S. Bauer, E. Davis, E. Hunt, A. Koosman, B. Meyers, M. Ouspenskaya, E. Peseke, V. Raffeale, and C. Rains for their help performing the experiment, K. Boersma for her help with the experimental design, and E. Boersley for her support and assistance. Additionally we thank J. Spatafora, V. Weis, and the Center for Genome Research and Biocomputing at Oregon State University for providing

laboratory space for qPCR. This research was supported by grants from the National Science Foundation (EF-1241889), National Institutes of Health (R01GM109499, R01TW010286), U.S. Department of Agriculture (NRI 2006-01370, 2009-35102-0543), and U.S. Environmental Protection Agency (CAREER 83518801) to JRR and NSF grant IOS 1121529 to TRR. Support was provided by the U.S. Forest Service Pacific Northwest Research Station, Corvallis, Oregon to DHO.

Conflict of Interest: The authors declare that they have no conflict of interest.

Supporting Information

Additional supporting information may be found in electronic supplementary material for this article.

357 **Literature Cited**

- 358 Bannerman, J. A. & B. D. Roitberg (2014) Impact of extreme and fluctuating
359 temperatures on aphid-parasitoid dynamics. *Oikos*, **123**, 89-98.
- 360 Ben-Horin, T., H. S. Lenihan & K. D. Lafferty (2012) Variable intertidal temperature
361 explains why disease endangers black abalone. *Ecology*, **94**, 161-168.
- 362 Blaustein, A. R., S. S. Gervasi, P. T. J. Johnson, J. T. Hoverman, L. K. Belden, P. W.
363 Bradley & G. Y. Xie (2012) Ecophysiology meets conservation:
364 understanding the role of disease in amphibian population declines. *Philos.*
365 *Trans. R. Soc., B*, **367**, 1688-1707.
- 366 Blaustein, A. R., J. M. Romansic, E. A. Scheessele, B. A. Han, A. P. Pessier & J. E.
367 Longcore (2005) Interspecific variation in susceptibility of frog tadpoles to the
368 pathogenic fungus *Batrachochytrium dendrobatidis*. *Conserv. Biol.*, **19**, 1460-
369 1468.
- 370 Boyle, D. G., D. B. Boyle, V. Olsen, J. A. T. Morgan & A. D. Hyatt (2004) Rapid
371 quantitative detection of chytridiomycosis (*Batrachochytrium dendrobatidis*)
372 in amphibian samples using real-time Taqman PCR assay. *Dis. Aquat. Org.*,
373 **60**, 141-148.
- 374 Brown, J. H., J. F. Gillooly, A. P. Allen, V. M. Savage & G. B. West (2004) Toward
375 a metabolic theory of ecology. *Ecology*, **85**, 1771-1789.

- 376 Buck, J. C., E. A. Scheessele, R. A. Relyea & A. R. Blaustein (2012) The effects of
377 multiple stressors on wetland communities: pesticides, pathogens and
378 competing amphibians. *Freshwat. Biol.*, **57**, 61-73.
- 379 Easterling, D. R., G. A. Meehl, C. Parmesan, S. A. Changnon, T. R. Karl & L. O.
380 Mearns (2000) Climate extremes: observations, modeling, and impacts.
381 *Science*, **289**, 2068-2074.
- 382 Forrest, M. J. & M. A. Schlaepfer (2011) Nothing a hot bath won't cure: Infection
383 rates of amphibian chytrid fungus correlate negatively with water temperature
384 under natural field settings. *PLoS ONE*, **6**, e28444.
- 385 Garner, T. W. J., S. Walker, J. Bosch, S. Leech, J. M. Rowcliffe, A. A. Cunningham
386 & M. C. Fisher (2009) Life history tradeoffs influence mortality associated
387 with the amphibian pathogen *Batrachochytrium dendrobatidis*. *Oikos*, **118**,
388 783-791.
- 389 Gervasi, S., C. Gondhalekar, D. H. Olson & A. R. Blaustein (2013) Host identity
390 matters in the amphibian-*Batrachochytrium dendrobatidis* system: Fine-scale
391 patterns of variation in responses to a multi-host pathogen. *PLoS ONE*, **8**,
392 e54490.
- 393 Gervasi, S. S., P. R. Stephens, J. Hua, C. L. Searle, G. Y. Xie, J. Urbina, D. H. Olson,
394 B. A. Bancroft, V. Weis, J. I. Hammond, R. A. Relyea & A. R. Blaustein
395 (2017) Linking ecology and epidemiology to understand predictors of multi-

- 396 host responses to an emerging pathogen, the amphibian chytrid fungus. *PLoS*
397 *ONE*, **12**, e0167882.
- 398 Han, B. A., P. W. Bradley & A. R. Blaustein (2008) Ancient behaviors of larval
399 amphibians in response to an emerging fungal pathogen, *Batrachochytrium*
400 *dendrobatidis*. *Behav. Ecol. Sociobiol.*, **63**, 241-250.
- 401 Hoover, D. L., A. K. Knapp & M. D. Smith (2014) Resistance and resilience of a
402 grassland ecosystem to climate extremes. *Ecology*, **95**, 2646-2656.
- 403 Horton, R. M., J. S. Mankin, C. Lesk, E. Coffel & C. Raymond (2016) A Review of
404 Recent Advances in Research on Extreme Heat Events. *Current Climate*
405 *Change Reports*, **2**, 242-259.
- 406 IPCC, (2007) Climate Change 2007: Synthesis Report. 104. R. K. Pachauri & A.
407 Reisinger (Eds.). IPCC, Geneva.
- 408 Kilpatrick, A. M., C. J. Briggs & P. Daszak (2010) The ecology and impact of
409 chytridiomycosis: an emerging disease of amphibians. *Trends Ecol. Evol.*, **25**,
410 109-118.
- 411 Li, Y., J. M. Cohen & J. R. Rohr (2013) Review and synthesis of the effects of
412 climate change on amphibians. *Integr Zool*, **8**, 145-161.
- 413 Liu, X., J. R. Rohr & Y. Li (2013) Climate, vegetation, introduced hosts and trade
414 shape a global wildlife pandemic. *Proceedings of the Royal Society B:*
415 *Biological Sciences*, **280**, 20122506.

- 416 Longcore, J., A. Pessier & D. Nichols (1999) *Batrachochytrium dendrobatidis* gen. et
417 sp. nov., a chytrid pathogenic to amphibians. *Mycologia*, **91**, 219-227.
- 418 Marantelli, G., L. Berger, R. Speare & L. Keegan (2004) Distribution of the
419 amphibian chytrid *Batrachochytrium dendrobatidis* and keratin during tadpole
420 development. *Pac. Conserv. Biol.*, **10**, 173-179.
- 421 McMahon, T. A. & J. R. Rohr (2015) Transition of chytrid fungus infection from
422 mouthparts to hind limbs during amphibian metamorphosis. *EcoHealth*, **12**,
423 188-193.
- 424 Olson, D. H., D. M. Aanensen, K. L. Ronnenberg, C. I. Powell, S. F. Walker, J.
425 Bielby, T. W. J. Garner, G. Weaver, M. C. Fisher & T. B. M. Group (2013)
426 Mapping the global emergence of *Batrachochytrium dendrobatidis*, the
427 amphibian chytrid fungus. *PLoS ONE*, **8**, e56802.
- 428 Paaijmans, K. P., S. Blanford, A. S. Bell, J. I. Blanford, A. F. Read & M. B. Thomas
429 (2010) Influence of climate on malaria transmission depends on daily
430 temperature variation. *Proc. Natl. Acad. Sci. U. S. A.*, **107**, 15135-15139.
- 431 Paaijmans, K. P., A. F. Read & M. B. Thomas (2009) Understanding the link between
432 malaria risk and climate. *Proc. Natl. Acad. Sci. U. S. A.*, **106**, 13844-13849.
- 433 Pearl, C. A., E. L. Bull, D. E. Green, J. Bowerman, M. J. Adams, A. Hyatt & W. H.
434 Wente (2007) Occurrence of the amphibian pathogen *Batrachochytrium*
435 *dendrobatidis* in the Pacific Northwest. *J. Herpetol.*, **41**, 145-149.

- 436 Piotrowski, J. S., S. L. Annis & J. E. Longcore (2004) Physiology of
437 *Batrachochytrium dendrobatidis*, a chytrid pathogen of amphibians.
438 *Mycologia*, **96**, 9-15.
- 439 Piovia-Scott, J., K. L. Pope, S. P. Lawler, E. M. Cole & J. E. Foley (2011) Factors
440 related to the distribution and prevalence of the fungal pathogen
441 *Batrachochytrium dendrobatidis* in *Rana cascadae* and other amphibians in
442 the Klamath Mountains. *Biol. Conserv.*, **144**, 2913–2921.
- 443 Raffel, T. R., N. T. Halstead, T. A. McMahon, A. K. Davis & J. R. Rohr (2015)
444 Temperature variability and moisture synergistically interact to exacerbate an
445 epizootic disease. *Proceedings of the Royal Society B: Biological Sciences*,
446 **282**, 20142039.
- 447 Raffel, T. R., P. J. Michel, E. W. Sites & J. R. Rohr (2010) What drives chytrid
448 infections in newt populations? Associations with substrate, temperature, and
449 shade. *EcoHealth*, **7**, 526-536.
- 450 Raffel, T. R., J. R. Rohr, J. M. Kiesecker & P. J. Hudson (2006) Negative effects of
451 changing temperature on amphibian immunity under field conditions. *Funct.*
452 *Ecol.*, **20**, 819-828.
- 453 Raffel, T. R., J. M. Romansic, N. T. Halstead, T. A. McMahon, M. D. Venesky & J.
454 R. Rohr (2013) Disease and thermal acclimation in a more variable and
455 unpredictable climate. *Nat. Clim. Change*, **3**, 146–151.

- 456 Rohr, J. R. & T. R. Raffel (2010) Linking global climate and temperature variability
457 to widespread amphibian declines putatively caused by disease. *Proc. Natl.*
458 *Acad. Sci. U. S. A.*, **107**, 8269-8274.
- 459 Rohr, J. R., T. R. Raffel, A. R. Blaustein, P. T. J. Johnson, S. H. Paull & S. Young
460 (2013) Using physiology to understand climate-driven changes in disease and
461 their implications for conservation. *Conserv. Physiol.*, **1**, cot022.
- 462 Rummukainen, M. (2012) Changes in climate and weather extremes in the 21st
463 century. *Wiley Interdiscip. Rev.: Clim. Change*, **3**, 115-129.
- 464 Schar, C., P. L. Vidale, D. Luthi, C. Frei, C. Haberli, M. A. Liniger & C. Appenzeller
465 (2004) The role of increasing temperature variability in European summer
466 heatwaves. *Nature*, **427**, 332-336.
- 467 Searle, C. L., S. S. Gervasi, J. Hua, J. I. Hammond, R. A. Relyea, D. H. Olson & A.
468 R. Blaustein (2011) Differential host susceptibility to *Batrachochytrium*
469 *dendrobatidis*, an emerging amphibian pathogen. *Conserv. Biol.*, **25**, 965-974.
- 470 Skerratt, L., L. Berger, R. Speare, S. Cashins, K. McDonald, A. Phillott, H. Hines &
471 N. Kenyon (2007) Spread of chytridiomycosis has caused the rapid global
472 decline and extinction of frogs. *EcoHealth*, **4**, 125-134.
- 473 Stuart, S. N., J. S. Chanson, N. A. Cox, B. E. Young, A. S. L. Rodrigues, D. L.
474 Fischman & R. W. Waller (2004) Status and trends of amphibian declines and
475 extinctions worldwide. *Science*, **306**, 1783-1786.

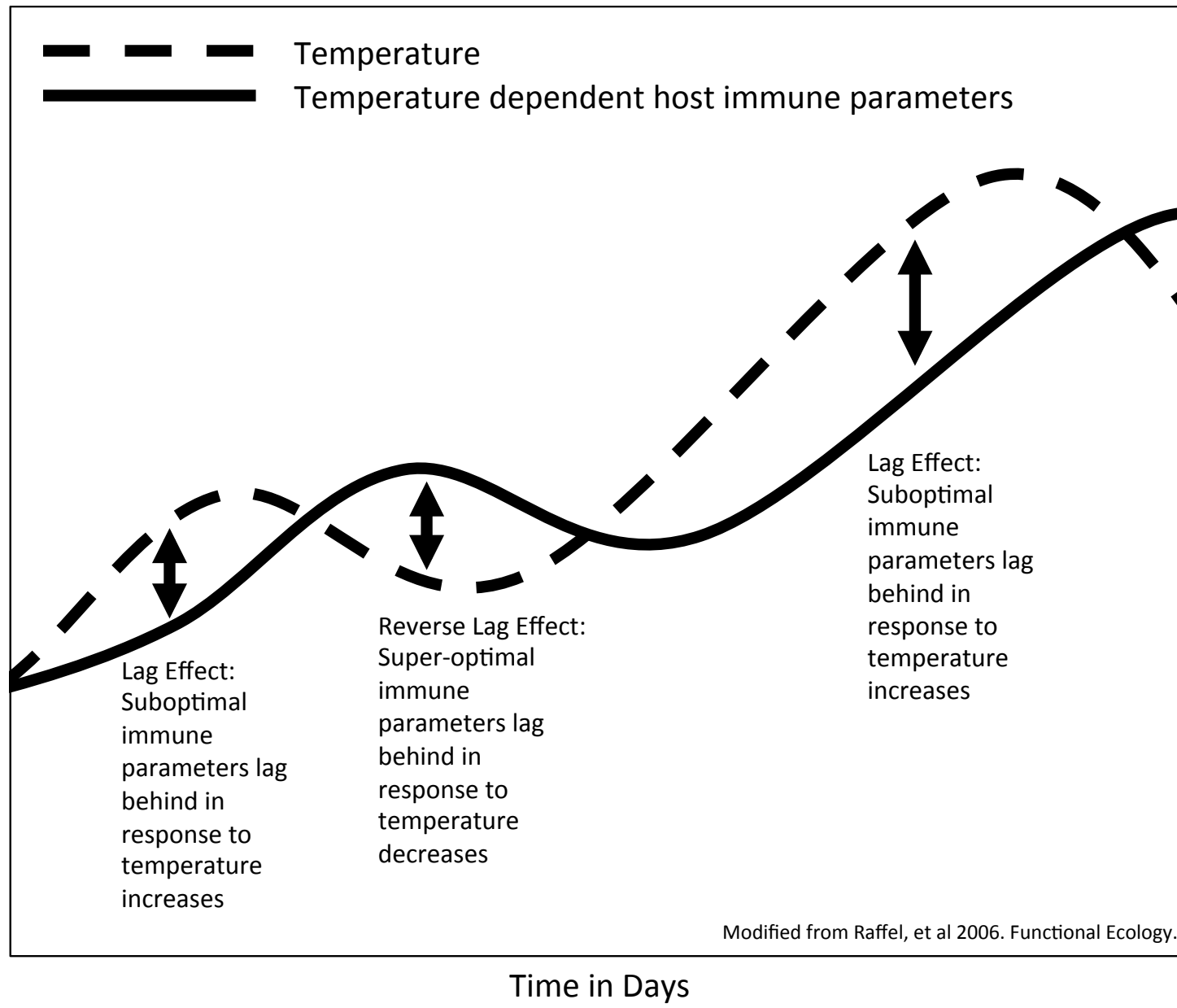
- 476 Teskey, R., T. Wertin, I. Bauweraerts, M. Ameye, M. A. McGuire & K. Steppe
 477 (2015) Responses of tree species to heat waves and extreme heat events. *Plant*
 478 *Cell Environ*, **38**, 1699-1712.
- 479 Venesky, M. D., T. R. Raffel, T. A. McMahon & J. R. Rohr (2013) Confronting
 480 inconsistencies in the amphibian-chytridiomycosis system: implications for
 481 disease management. *Biol. Rev. Camb. Philos. Soc.*, **89**, 477-483.
- 482 Voyles, J., L. R. Johnson, C. J. Briggs, S. D. Cashins, R. A. Alford, L. Berger, L. F.
 483 Skerratt, R. Speare & E. B. Rosenblum (2012) Temperature alters
 484 reproductive life history patterns in *Batrachochytrium dendrobatidis*, a lethal
 485 pathogen associated with the global loss of amphibians. *Ecology and*
 486 *Evolution*, **2**, 2241-2249.
- 487 Voyles, J., L. R. Johnson, J. Rohr, R. Kelly, C. Barron, D. Miller, J. Minster & E. B.
 488 Rosenblum (2017) Diversity in growth patterns among strains of the lethal
 489 fungal pathogen *Batrachochytrium dendrobatidis* across extended thermal
 490 optima. *Oecologia*, 1-11.
- 491 Woodhams, D. C., R. A. Alford, C. J. Briggs, M. Johnson & L. A. Rollins-Smith
 492 (2008) Life-history trade-offs influence disease in changing climates:
 493 strategies of an amphibian pathogen. *Ecology*, **89**, 1627-1639.
- 494 Xie, G. Y., D. H. Olson & A. R. Blaustein (2016) Projecting the global distribution of
 495 the emerging amphibian fungal pathogen, *Batrachochytrium dendrobatidis*,
 496 based on IPCC climate futures. *PLoS ONE*, **11**, e0160746.

497

498

499 **Fig 1.** Hypothesized lag effect showing the relationship between fluctuating
 500 temperatures (over days to weeks) and the optimal levels of a hypothetical
 501 temperature-dependent host immune parameter. The immune parameter follows and
 502 lags behind temperature changes – resulting in periods of a compromised immune
 503 status after a temperature increase, and resulting in an over-active (or unnecessarily
 504 costly) immune status after a temperature decrease. Modified from Raffel *et al.*
 505 (2006).

506
 507 **Fig 2.** Mean *Batrachochytrium dendrobatidis* (Bd) infection abundance (\pm SE)
 508 measured at death, or at euthanasia 16-days after Bd exposure, in both western toad
 509 (*Anaxyrus boreas*) larvae and red legged frog (*Rana aurora*) larvae from Oregon,
 510 USA, and between the two temperatures at the time of Bd-exposure (cold [15° C]
 511 versus warm [20° C]) and between larvae having experienced either a constant or
 512 shifted temperature. Bd infection abundance is quantified as the log (1 + Bd genomic
 513 equivalents) per excised larval mouthparts of all individuals exposed to the pathogen.
 514



Bd Infection Abundance (Log (1 + genomic equivalents))

