Can functional traits explain phylogenetic signal in the composition of a plant community?

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Abstract:

Phylogeny-based and functional trait-based analyses are two principle ways to study community

assembly and underlying ecological processes. In principle, knowing all information about

species traits would make phylogenetic information redundant, at least that component of

phylogenetic signal in the distribution of species among communities that is caused by

phylogenetically related species sharing similar traits. In reality, phylogenies may contain more

information than a set of singular, discretely measured traits because we cannot measure all

species traits and may misjudge which are most important. The extent to which functional trait

information makes phylogenetic information redundant, however, has not been explicitly studied

with empirical data in community ecology. Here, we use phylogenetic linear mixed models to

analyze community assembly of 55 understory plant species in 30 forest sites in central

Wisconsin. These communities show strong phylogenetic attraction, yet variation among sites in

20 environmental variables could not account for this pattern. Most of the 15 functional traits we

measured had strong phylogenetic signal, but only three varied strongly among sites in ways that

affected species' abundances. These three traits explained only 19% of variation in phylogenetic

patterns of species co-occurrence. Thus, phylogenies appear to provide considerably more

information about community assembly than the functional traits measured in this study,

demonstrating the value of phylogeny in studying of community assembly processes even with

abundant functional traits.

Introduction

Functional traits, arising as innovations through evolution, can capture essential aspects of species' morphology, ecophysiology, and life-history strategy (McGill *et al.* 2006; Violle *et al.* 2007). Although closely related species can differ greatly in some functional traits due to rapid evolution or convergence (Losos, 2008, 2011), most functional traits show strong phylogenetic signal (Moles *et al.* 2005, Donoghue 2008, Freckleton *et al.* 2002; Webb *et al.* 2002). Functional traits, with or without phylogenetic signal, are known to influence the species composition of communities, thereby providing mechanistic links between fundamental ecological processes and community structure (McGill *et al.* 2006; Violle *et al.* 2007; Adler *et al.* 2013). Functional traits also provide a common currency that facilitates comparisons among species and across regions, allowing us to assess the generality of patterns and predictions in community ecology (McGill *et al.* 2006). This has lead to a proliferation of studies using functional traits to understand community structure and composition. Functional trait-based approaches, however, are limited by the fact that it is impossible to measure all potentially important functional traits affecting the distribution of species across communities.

Even in the absence of functional trait information, it is still possible to infer the effects of (unmeasured) functional traits on community composition by investigating phylogenetic patterns in community composition. Phylogenies play an important role in community ecology by giving information about evolutionary relationships among species (Graves & Gotelli, 1993; Losos 1996; Baum & Smith, 2012). Because phylogenetically related species often share similar functional trait values, we expect phylogenetically related species to co-occur more often in the same communities reflecting their shared environmental tolerances. Conversely, if

phylogenetically related species have similar traits that cause them to compete more, then closely

related species may be less likely to co-occur. These and other processes relating functional traits

to community composition likely lead to phylogenetic signatures in how species are distributed

among communities (Webb *et al.* 2002). However, in principle, if we have information for all

relevant functional traits, then we expect phylogeny to provide little additional information

relevant for community composition. That is, when all of the functional traits affecting

community composition are known, we do not expect the unexplained residual variation in the

occurrence of species to have phylogenetic signal (Ives & Helmus, 2011).

In practice, we cannot obtain information about all relevant functional traits. In addition,

phylogenetic signals in community composition may result from factors beyond functional traits

that generate phylogenetic signal, such as the biogeographical patterns generated as species

disperse across a landscape (Ricklefs *et al.* 1993; Moen *et al.* 2009). Therefore, even after

accounting for those functional traits whose measurements are available, we should expect

phylogenies to contain additional information about community composition (Vane-Wright *et al.*

1991; Cadotte *et al.* 2009). As far as we know, however, no study has explicitly assessed how

much information about community assembly provided by traits and phylogeny is overlapping.

Here, we ask how much of the phylogenetic signal in the composition of a plant community

assemblage can be explained by functional traits (Fig. 1).

To address this question, we take advantage of our data set of 30 understory communities

containing a total of 55 species from Wisconsin pine barrens for which we built a highly resolved

phylogeny, amassed a large database on 15 functional traits, and measured 20 environmental

variables at each community. We first investigate whether there is phylogenetic pattern in

community composition, using a phylogenetic community mixed model that tests for both

"phylogenetic attraction" (phylogenetically related species more likely to occur in the same

communities) and "phylogenetic repulsion." If there is phylogenetic pattern, then it could be

produced by measured functional traits that themselves have phylogenetic signal (Fig. 1, arrows

2, 4, and 7), unmeasured functional traits with phylogenetic signal (Fig. 1, arrows 2, 5, and 8), or

phylogenetic processes unrelated to functional traits (Fig. 1, arrow 6). We then develop a

phylogenetic community mixed model incorporating the functional traits we measured to ask

whether there is phylogenetic signal in the residual variation in community composition after the

effects of these functional traits are removed. This analysis tests the hypothesis that we can

explain all the phylogenetic pattern in community composition using measured functional traits.

Finally, we use a phylogenetic community mixed model to investigate whether phylogenetically

related species respond similarly to environmental gradients across the communities. The

motivation for this analysis is to indirectly identify possible unmeasured functional traits that

might play a role in community composition. Cases where phylogenetically related species

respond similarly to an environmental gradient suggest that they share traits that confer similar

tolerances to, or preferences for, specific environmental conditions; this analysis therefore may

suggest what additional functional traits to measure in order to explain patterns in community

composition.

Methods

Data

Community composition. – We sampled 30 pine barrens forest sites in the central Wisconsin sand

plains in 2012 using 50 1-m^2 quadrats placed along five transects at each site. Within each

quadrat, we recorded the presence/absence of all vascular plant species (see Li & Waller 2015

for details). Across all sites, we recorded 152 species. For the analyses other than the initial

exploration of phylogenetic patterns in community composition, we focused on the 55 species

that occurred in three or more communities. This was done for two reasons. First, we did not

have values for all functional traits for many of the rare species. Second, we wanted to limit the

number of zeros in the data set, especially in analyses of responses of species to environmental

variables in which one or two observations of a species give little information.

Functional traits. – For the 55 focal species, we measured 11 continuous and four categorical

functional traits on at least 12 individuals from at least three populations using standard protocols

(Pérez-Harguindeguy *et al.* 2013). Continuous traits include seed mass (*g*/seed), plant height

(*cm*), specific leaf area (SLA, m^2/kg), leaf dry matter content (LDMC, %), leaf circularity

(dimensionless), leaf length (*cm*), leaf width (*cm*), leaf thickness (*mm*), leaf carbon concentration

(%), leaf nitrogen concentration (%), and stem dry matter content (SDMC, %). We aggregated

categories of each categorical trait into two levels: growth form (woody vs. non-woody), life

cycle (annual vs. non-annual), and pollination mode (biotic vs. abiotic). We divided seed

dispersal mode into three binary variables (wind dispersed vs. not, animal dispersed vs. not, and

unassisted vs. assisted dispersal). Collectively, these functional traits, covering the leaf-height-

seed (LHS) plant ecology strategy (Westoby, 1998), represent multidimensional functions of

plants associated with resource use, competitive ability, dispersal ability, etc. For analyses, trait

values were Z-transformed to have means of zero and standard deviations of one, allowing

coefficients in the mixed models to be interpreted as effects sizes.

Phylogeny. – The phylogeny used in this study is a subset of a phylogeny for all vascular plants

in Wisconsin (Cameron *et al. unpublished manuscript*[1]). Briefly, Cameron *et al.* used two plastid

DNA barcode loci *rbcL* and *matK* to build the phylogeny using maximum likelihood (ML) in the

program R$_{AXML}$ (Stamatakis, 2014). The phylogeny was then time-calibrated using the branch

length adjuster (*bladj*) available in the program phylocom (Webb *et al.* 2008).

Environmental data. – At each site, we collected six soil samples and then pooled the samples

together to measure the soil properties listed in Table 4. We also took six vertical fish-eye

photographic images at each site to measure canopy cover. To characterize climatic conditions,

we extracted daily precipitation and minimum temperature for each site from interpolated values

estimated by Kucharik *et al*. (2010) from 2002 to 2006 (data after 2006 were not available). All

environmental variables were also scaled to have means of zero and standard deviations of one.

Phylogenetic community composition

141   We first tested for phylogenetic community structure without including environmental or

142   functional trait information. We used traditional metrics and randomization tests (i.e., null

143   models) to identify whether there was phylogenetic pattern (phylogenetic attraction or repulsion)

144   in the composition of our 30 communities. Specifically, we measured the phylogenetic structure

145   of species abundances at each site using phylogenetic species evenness (PSE, Helmus *et al.*

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[1] Cameron, K., R. Kriebel, M. Pace, D. Spalink, P. Li, and K. Sytsma. *In prep*. A complete molecular community phylogeny for the flora of Wisconsin based on the universal plant DNA barcode.

2007) and mean phylogenetic distance (MPD, Webb, 2000). For a site with m individuals from n

species, if we let $\mathbf{M}$ to be a $n \times 1$ column vector containing abundance of all species and let

variance-covariance matrix $\mathbf{C}$ describing phylogenetic relationship of these species, we can

calculate PSE as $\frac{m\,\mathrm{diag}(\boldsymbol{C})'\boldsymbol{M} - \boldsymbol{M}'\boldsymbol{CM}}{m^2(1-1/n)}$ (the prime denotes transpose) and calculate MPD as

$\frac{\sum_i^n \sum_j^n C_{[i,j]}\boldsymbol{M}_i\boldsymbol{M}_j}{\sum_i^n \sum_j^n \boldsymbol{M}_i\boldsymbol{M}_j}, i \neq j$. PSE is scaled from zero to one, with one occurring when $\mathbf{C}$ is the identity

matrix and all species have equal abundance. We then calculated the mean of these 30 PSE

($\overline{\mathrm{PSE}}_{\mathrm{obs}}$) and 30 MPD values ($\overline{\mathrm{MPD}}_{\mathrm{obs}}$). To test for phylogenetic pattern using these metrics, we

permuted species randomly among sites (SIM2 in Gotelli, 2000) 4999 times and then calculated

metrics base on each permutation data set. This null permutation model retains the prevalence of

each species across sites, but allows sites to change in species richness. If $\overline{\mathrm{PSE}}_{\mathrm{obs}}$ or $\overline{\mathrm{MPD}}_{\mathrm{obs}}$

falls above (or below) 97.5% of the permutation values, then we infer a statistically significant

phylogenetic repulsion (attraction). Using this null model where sites can vary in species

richness is justified because under the null hypothesis of no phylogenetic signal, the values of

PSE and MPD are independent of species richness at the sites.

In addition to testing the significance of the observed PSE and MPD values via null models, we

fit a phylogenetic linear mixed model (PLMM) to test for phylogenetic community patterns in

species abundances. A PLMM establishes a flexible statistical base to subsequently incorporate

functional trait and environmental variables. Furthermore, PLMMs tend to have greater

statistical power than other metrics examined using permutation tests (Ives & Helmus, 2011). To

build the PLMM, let n be the number of species distributed among m sites. Letting Y be the $mn \times$

1 vector containing the abundance of species j ($j = 1, \ldots, n$) at site s ($s = 1, \ldots, m$), the PLMM is

167            $\log(Y + 1) = \alpha + a_{\mathrm{spp}[i]} + b_{\mathrm{spp}[i]} + c_i + d_{\mathrm{site}[i]} + e_i$

$a \sim \text{Gaussian}(\mathbf{0}, \sigma_a^2 \mathbf{I}_n)$

$b \sim \text{Gaussian}(\mathbf{0}, \sigma_b^2 \mathbf{S}_{\text{spp}})$

$c \sim \text{Gaussian}(\mathbf{0}, \text{kron}(\mathbf{I}_m, \sigma_c^2 \mathbf{S}_{\text{nested}}))$

$d \sim \text{Gaussian}(\mathbf{0}, \sigma_d^2 \mathbf{I}_m)$

$e \sim \text{Gaussian}(\mathbf{0}, \sigma_e^2 \mathbf{I}_{mn})$ (1)

We use the convention of multilevel models here (Gelman & Hill, 2007), with fixed and random

effects given by Greek and Latin letters, respectively. The function spp[i] maps the observation i

in vector Y to the identity of the species (Gelman & Hill, 2007, p251-252), so i takes value from

1 to mn. The intercept α estimates the overall average log abundance of species across all sites.

The following three random variables $a_{\text{spp}[i]}$, $b_{\text{spp}[i]}$ and c_i incorporate variation in abundance

among plant species. Specifically, the n values of $a_{\text{spp}[i]}$ give differences among species in mean

log abundance across all sites and are assumed to be drawn independently from a Gaussian

distribution with mean 0 and variance σ_a^2. The n values of $b_{\text{spp}[i]}$ also give differences in mean log

abundance across sites but are assumed to be drawn from a multivariate Gaussian distribution

with covariance matrix $\sigma_b^2 \mathbf{S}_{\text{spp}}$, where the $n \times n$ matrix $\mathbf{S}_{\text{spp}}$ is derived from the phylogeny (see

next paragraph below), and the scalar σ_b^2 dictates the overall strength of the phylogenetic signal.

Thus, $a_{\text{spp}[i]}$ and $b_{\text{spp}[i]}$ together capture variation in mean species log abundances that is either

unrelated to phylogeny or has phylogenetic signal. The random variable c_i accounts for

covariance in the log abundances of plant species nested within sites (using the Kronecker

product, ⊗). Specifically, c_i assesses whether phylogenetically related plant species are more or

less likely to co-occur at the same sites. Hence, c_i is used to measure either phylogenetic

attraction or phylogenetic repulsion; because σ_c^2 dictates the overall strength of these

phylogenetic patterns, it is the key term we are interested in. Random effect $d_{\text{site}[i]}$ is assumed to

contain m values, one for each site, that are distributed by a Gaussian distribution with variance

σ_d^2 to account for differences in the average log abundances of species from site to site. Finally, e_i

captures residual variance σ_e^2.

We base the phylogenetic covariance matrix $\mathbf{S}_{\text{spp}}$, on the assumption of Brownian motion

evolution. If a continuous-valued trait evolves up a phylogenetic tree with a constant probability

of increasing or decreasing slightly, the covariance in trait values between two species will be

proportional to the length of shared evolution, given by the distance on the phylogenetic tree

between the root and the species' most recent common ancestor (Martins & Hanson 1997). This

gives a direct way to convert the phylogeny into a hypothesis about the covariance matrix. For

the assessment of phylogenetic attraction within sites, c_i, we use $\mathbf{S}_{\text{nested}} = \mathbf{S}_{\text{spp}}$. For phylogenetic

repulsion, we use the matrix inverse of $\mathbf{S}_{\text{spp}}$, $\mathbf{S}_{\text{nested}} = (\mathbf{S}_{\text{spp}})^{-1}$. Theoretical justification for $\mathbf{S}_{\text{nested}} =$

$(\mathbf{S}_{\text{spp}})^{-1}$ comes from a model of competition among community members (Ives & Helmus 2011,

Appendix A). Briefly, if the strength of competition between species is given by $\mathbf{S}_{\text{spp}}$, as might be

the case if closely related species are more likely to share common resources, then the relative

abundances of species will have covariance matrix $(\mathbf{S}_{\text{spp}})^{-1}$.

Equation 1 is the same as model I in Ives & Helmus (2011), except model I includes variation

among species in mean log abundance across sites as fixed effects rather than two random

effects, $a_{\text{spp}[i]}$ and $b_{\text{spp}[i]}$. This change allows us to align equation 1 with equation 3 below that

includes variation in the relationship between trait values and log abundance within sites as

random effects. In our analyses, treating variation among species in mean log abundance as fixed

effects (results not presented) led to almost identical estimates of phylogenetic signal (estimates

of σ_c^2), and therefore our treatment of $a_{spp[i]}$ and $b_{spp[i]}$ as random effects does not change the

conclusions.

We fit the PLMM with maximum likelihood using function communityPGLMM in the pez

library of R (R Core Team, 2015). Statistical significance of the variance estimates σ^2 was

determined using a likelihood ratio test. Because the null hypothesis $\sigma^2 = 0$ is on the boundary of

the parameter space (because σ^2 cannot be negative), we used the $0.5\chi_0^2 + 0.5\chi_1^2$ mixture

distribution of Self & Liang (1987) for significance tests. The distribution of χ_0^2 represents a

distribution with a point mass at 0, and the *p*-values given by the constrained likelihood ratio test

are one-half the values that would be calculated from a standard likelihood ratio test using χ_1^2.

Simulations suggest that *p*-values calculated in this way are more conservative than those from a

parametric bootstrap (Appendix Text S1).

Our data set contained many zeros, raising the question of the validity of applying a linear model

to transformed data. Nonetheless, transforming data and applying a linear analysis is robust when

assessing the significance of regression parameters (Ives, 2015). To check this robustness for our

results, we repeated all analyses using phylogenetic generalized linear mixed models (PGLMM)

after converting our abundance data to presence/absence data. PGLMMs yielded qualitatively

similar results in all of the analyses we present in the main text; these PGLMM analyses are

presented in Appendix. We present the PLMM results in the main text, because including

abundance data in phylogenetic community structure analyses can provide more information

about community assembly (Freilich & Connolly, in press).

Can functional traits explain phylogenetic community composition?

To quantify how much of the variation in phylogenetic patterns can be explained by measured functional traits, we estimated PLMMs with and without functional traits, and then compared the strength of phylogenetic signal in the residual variation: if functional traits alone serve to explain phylogenetic community composition, then as functional traits are included, the strength of the phylogenetic signal in the residuals should decrease. We selected functional traits one by one based on the two conditions necessary for them to generate phylogenetic signal in community composition. First, a functional trait must show phylogenetic signal among species, because in the absence of phylogenetic signal among species, a trait could not produce phylogenetic signal in species' abundances. Second, there must be variation among sites in which species are selected according to their functional traits; if a trait has phylogenetic signal but there is no variation of relationships between plant functional trait values and abundances among sites, then species with similar trait values will also have similar overall abundance. This will contribute to the overall phylogenetic signal of species abundance and will be captured by $b_{spp[i]}$ in equation 1. Therefore, we only investigate traits that exhibit both strong phylogenetic signal *and* variation among sites in the apparent advantages the traits give to species.

Phylogenetic signal of functional trait was tested with model-based methods. Each continuous trait was tested with Pagel's λ (Pagel, 1999) using `phylolm` (Ho & Ané, 2014). For the binary traits, we applied phylogenetic logistic regression (Ives & Garland, 2010) as implemented by `phyloglm` (Ho & Ané, 2014). We also tested phylogenetic signal of functional traits via Blomberg's K (Blomberg *et al.* 2003) with `picante` (Kembel *et al.* 2010).

We tested variation of relationships between trait values and log abundances with the LMM

$\quad$ $\log(Y + 1) = \alpha + a_{\mathrm{spp}[i]} + (\beta + b_{\mathrm{site}[i]})t_{\mathrm{spp}[i]} + e_i$

$\quad$ $a \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_a^2 \mathbf{I}_n)$

$\quad$ $b \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_b^2 \mathbf{I}_m)$

$\quad$ $e \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_e^2 \mathbf{I}_{mn})$ $\hspace{6cm}$ (2)

where $t_{\mathrm{spp}[i]}$ is the focal functional trait value of the species corresponding to observation i, and σ_b^2

gives the variation among sites of the relationship between species trait values and log

abundances. If $\sigma_b^2 > 0$, we conclude that different sites select species differently based on the

tested trait. Significance of σ_b^2 was tested with a likelihood ratio test.

We quantified the contribution of a trait to the observed phylogenetic pattern using the model

$\quad$ $\log(Y + 1) = \alpha + a_{\mathrm{spp}[i]} + b_{\mathrm{spp}[i]} + c_i + d_{\mathrm{site}[i]} + (\beta + f_{\mathrm{site}[i]})t_{\mathrm{spp}[i]} + e_i$

$\quad$ $a \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_a^2 \mathbf{I}_n)$

$\quad$ $b \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_b^2 \mathbf{S}_{\mathrm{spp}})$

$\quad$ $c \sim \mathrm{Gaussian}(\mathbf{0}, \mathrm{kron}(\mathbf{I}_m, \sigma_c^2 \mathbf{S}_{\mathrm{nested}}))$

$\quad$ $d \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_d^2 \mathbf{I}_m)$

$\quad$ $f \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_f^2 \mathbf{I}_m)$

$\quad$ $e \sim \mathrm{Gaussian}(\mathbf{0}, \sigma_e^2 \mathbf{I}_{mn})$ $\hspace{6cm}$ (3)

This model is the same as equation 1 used to assess phylogenetic patterns in community

composition, except that it includes functional trait values $t_{\mathrm{spp}[i]}$. Random variables $a_{\mathrm{spp}[i]}$, $b_{\mathrm{spp}[i]}$,

c_i, and $d_{\text{site}[i]}$ are defined as in equation 1. The proportion of phylogenetic signal in species

composition (measured by σ_c^2) that this trait can explain is then assessed by comparing σ_c^2

between models with and without this trait. Finally, to evaluate the overall contribution of

functional traits to the observed phylogenetic patterns, we built a multivariate version of equation

3 which included all traits that have both phylogenetic signal *and* strong variation among sites.

Does any environmental variable drive phylogenetic pattern?

If phylogenetic patterns in community composition are observed, yet no functional traits can

explain the patterns, how could we identify additional functional traits that might be responsible?

Phylogenetically related species usually are assumed to be ecologically similar due to niche

conservatism (Wiens *et al.* 2010). Therefore, related species will tend to have similar responses

to environmental variable. If these environmental variables are strong enough to drive

phylogenetic patterns in community composition, then functional traits that are associated with

tolerance or sensitivity to these environmental variables will likely be important in explaining

community composition. Thus, we investigated phylogenetic patterns in the responses of species

to environmental variables to suggest additional, unmeasured functional traits that might be

important to explain phylogenetic patterns in community composition.

Given the large number of environmental variables we have, we first selected variables for which

there is significant variation among species' responses without accounting for their phylogenetic

relationships, using linear mixed model (LMM) structured as:

$$\log(Y + 1) = \alpha + a_{\text{spp}[i]} + (\beta + b_{\text{spp}[i]})x_{\text{site}[i]} + e_i$$

$$a \sim \text{Gaussian}(\mathbf{0}, \sigma^2_a \mathbf{I}_n)$$

$b \sim \text{Gaussian}(\mathbf{0}, \sigma^2_b \mathbf{I}_n)$

$e \sim \text{Gaussian}(\mathbf{0}, \sigma^2_e \mathbf{I}_{mn})$ (4)

As in model 1, spp[i] gives the species identity of datum i while site[i] gives the site identity of

datum i. This makes $x_{\text{site}[i]}$ the value of the focal environmental variable at the site corresponding

to observation i in the data. Variation among species in their response to the environmental

variable x is given by the σ^2_b term here. If there is significant variation in species' responses to

this variable ($\sigma^2_b > 0$), we then further tested whether phylogenetically related species respond to

environmental variables in a similar way using the PLMM model

$\log(Y + 1) = \alpha + a_{\text{spp}[i]} + b_{\text{spp}[i]} + (\beta + c_{\text{spp}[i]} + d_{\text{spp}[i]})x_{\text{site}[i]} + e_i$

$a \sim \text{Gaussian}(\mathbf{0}, \sigma^2_a \mathbf{I}_n)$

$b \sim \text{Gaussian}(\mathbf{0}, \sigma^2_b \mathbf{S}_{\text{spp}})$

$c \sim \text{Gaussian}(\mathbf{0}, \sigma^2_c \mathbf{I}_n)$

$d \sim \text{Gaussian}(\mathbf{0}, \sigma^2_d \mathbf{S}_{\text{spp}})$

$e \sim \text{Gaussian}(\mathbf{0}, \sigma^2_e \mathbf{I}_{mn})$ (5)

Here, $c_{\text{spp}[i]}$ and $d_{\text{spp}[i]}$ represent non-phylogenetic and phylogenetic variation among species in

their response to environmental variable x (see model II in Ives & Helmus, 2011). The key

parameter of interest is σ^2_d, which we tested using a likelihood ratio test. If $\sigma^2_d > 0$,

phylogenetically related species respond to environmental variable x in similar ways, suggesting

the existence of an unmeasured phylogenetically inherited trait that is associated with species

tolerances or sensitivities to x.

Results

Phylogenetic community composition

Phylogenetically related species co-occur more often than expected by chance in pine barrens communities in central Wisconsin (Fig. 2). Permutation tests including all 152 species show that closely related species are likely to co-occur within communities, as judged by either phylogenetic species evenness ($\overline{PSE}_{obs}$ = 0.32, p = 0.03) or mean phylogenetic distance ($\overline{MPD}_{obs}$ = 338, p = 0.01). Confining the analyses to the 55 focal species that occurred in at least three communities, the permutation tests did not statistically significant phylogenetic patterns ($\overline{PSE}_{obs}$ = 0.27, p = 0.29; $\overline{MPD}_{obs}$ = 286, p = 0.17), although the PLMM gave statistically significant phylogenetic patterns (p = 0.008; Table 1). Converting the data to presence/absence among communities, the PGLMM also confirmed phylogenetic attraction pattern in these communities (Appendix Table S1). The stronger statistical results for the PLMM and PGLMM are not unexpected, given that model-based approaches are generally statistically more powerful that metric-based approaches (Ives & Helmus 2011).

Can functional traits explain phylogenetic community composition?

Most functional traits show strong phylogenetic signal (Table 2). Three traits – SLA, leaf circularity, and leaf thickness – significantly affected plant species' abundances among sites ($\sigma_b^2 >$ 0, equation 2, Table 2). In other words, different sites selected different species based on these three functional traits. Individually, SLA, leaf circularity, and leaf thickness reduced the phylogenetic variance by 6%, 2%, and 6% measured by the reduction in σ_c^2 with the inclusion of these functional traits (equation 3). Including all three traits in the final model, the phylogenetic

variation σ_c^2 decreased 19%. In other words, the many functional traits we measured in this study

can only explain 19% of the phylogenetic signal in species composition among these

communities. Converting the data to presence/absence and using the PGLMM, σ_c^2 was reduced

by 53% (Appendix, Table S2). Thus, functional traits explained more of the phylogenetic

patterns in community composition in the presence/absence of species from communities than in

their log abundance, although this reduction still only explains about one-half of the phylogenetic

pattern in community composition.

Does any environmental variable drive phylogenetic pattern?

There is significant variation among species in their responses to most environmental variables,

including most soil properties we measured, canopy shade, precipitation and minimum

temperature (Table 4). However, none of these variables show phylogenetic signal (last column

in Table 4). Therefore, no environmental variable we measured can explain the observed

phylogenetic pattern in community composition. When using PGLMM with presence/absence

data, we found that minimum temperature and soil pH, Ca, and Mn concentration show

phylogenetic patterns, with related species showing similar patterns in presence/absence in

response to these environmental variables (Appendix Table S3).

Discussion

We found phylogenetic patterns in community composition, in which phylogenetically related

species were more likely to occur in the same communities, yet we could not explain this pattern

completely using information among species' functional traits. When functional traits that

themselves showed phylogenetic signal were included in the phylogenetic linear mixed model

(PLMM) for log abundances of species in communities, that component of the residual variance

having phylogenetic covariances decreased by only 19%. The decrease in the phylogenetic

component of residual variation decreased 53% in the analyses of presence/absence data, yet

even this leaves considerable phylogenetic pattern in the unexplained variation in the

presence/absence of species among communities. Thus, even though we measured 15 functional

traits, including most of the "standard" functional traits used for studies on community structure,

we could not explain the phylogenetic patterns in community composition. This suggests that

there are either important functional traits that we have not measured, or that there are

phylogenetic processes unrelated to functional traits that we have not identified.

Phylogenetic community composition

The permutation tests using PSE and MPD, and the PLMM and PGLMM, found phylogenetic

attraction in community composition for the plant communities of central Wisconsin when all

152 species were included. However, only the PLMM and PGLMM found statistically

significant phylogenetic patterns when using the subset of 55 species that occurred in three or

more communities for which we had complete information on functional traits. Ives & Helmus

(2011) found that phylogenetic mixed models have greater power than metrics such as PSE and

MPD used with permutation tests to detect phylogenetic community structure. Simulations

tailored for our plant community data (Appendix Text S1) showed that the PLMM analyses

tended to have, if anything, incorrectly low Type I error rates, implying that our PLMM results

were unlikely to be the result of false positives. Therefore, we can conclude that closely related

species are more likely to co-occur than expected by chance in these sand plain communities.

Functional traits and phylogenetic patterns

We used our extensive database of functional traits to answer a key question in trait-based and

phylogeny-based community ecology: Can information about functional traits explain

phylogenetic patterns in community composition? Incorporating our measured functional traits

into a PLMM for log species' abundance only reduced by 19% the phylogenetic component of

residual variation in species composition (Table 3). For presence/absence data and a PGLMM,

this increased to 53%, although there still remained considerable phylogenetic covariances. A

possible explanation for this residual phylogenetic variation is some unknown historical process

(Fig. 1B, IV). However, all of our sites are located within 100 km with each other (Li & Waller,

2015), and therefore historical biogeographical processes are unlikely to affect plant community

composition.

We think the main source of phylogenetic patterns that were not explained by our measured

functional traits is additional functional traits that we did not measure. Further analyses of

presence/absence data with PGLMMs suggested that soil pH, Ca and Mn concentrations, and

minimum temperature are the potential driving variables for the residual phylogenetic patterns

(Appendix Table S3). However, none of the functional traits we measured are likely to explain

how plants respond to these environmental variables. Measurements of additional functional

traits that control plant performance related to these environmental variables (e.g., root structure,

micorrhizal associations, frost tolerance, etc.) might be able to explain more of the phylogenetic

pattern in community composition.

Implications for other studies

Our study has several implications for community ecology. First, it is clear that studying

community composition should incorporate analyses of both phylogenetic structure and variation

in functional traits. These data clearly complement each other in allowing sophisticated analyses

that can partition the amount of phylogenetic signal in community assembly that is associated

with functional trait variation (Fig. 1). Our results provide empirical support for the argument

that phylogeny can provide more information than a set of singular, discretely measured traits

(Vane-Wright *et al.* 1991; Cadotte *et al.* 2009). Although functional traits are necessary in order

to accurately infer processes from phylogenetic patterns (Kraft *et al.* 2007; Cavender-Bares *et al.*

2009), functional traits alone might not give a complete picture of community assembly. This

implies that both phylogenetic and trait-based data are needed if we are to understand how

measured and unmeasured traits, along with biogeography, influence community composition.

Second, our finding also provides an explanation for different (sometimes even opposite)

conclusions about community assembly processes based on comparisons between dispersion

patterns of phylogeny and traits (e.g. Swenson & Enquist, 2009; Kraft & Ackerly, 2010; Graham

et al. 2012). As functional traits and phylogeny provide overlapped and complementary

information about species and communities, different conclusions can be possible depend on the

functional traits used in the analyses. If the functional traits used provide similar information

with phylogeny, then conclusion about community assembly processes may be the same and vice

versa. Instead of compare and contrast dispersion patterns of traits and phylogeny, we argue that

integrating both into a model-based method such as PLMM used here. Model-based methods are

being increasingly applied in ecology because they are more interpretable, flexible, and powerful

than either null models or conventional algorithmic multivariate analyses (Warton *et al.* 2014).

With phylogenetic linear mixed models (PLMM), we not only detected phylogenetic patterns in

community composition, but also assessed whether these could be explained by functional traits.

Thus, both phylogenies and functional traits could be incorporated into the same statistical

model, with PLMMs (and PGLMMs) providing an integrated and quantitative framework for

analyzing ecological communities and predicting abundance of one taxon from others.

Finally, we can use phylogenetic analyses to suggest possible unmeasured functional traits that

underlie patterns in community composition and that therefore should be measured. If species

respond differently to an environmental variable, and if these differences are phylogenetic (i.e.,

related species respond to the environmental variable in similar ways), then there is likely to be a

functional trait or traits that underlie the response of species to this environmental variable. In

our study, the phylogenetic patterns in species responses to edaphic conditions like soil

chemistry highlighted our lack of data on the specific functional traits related to roots or

water/nutrient uptake. While this reveals that our study is incomplete, it also provides a valuable

lesson and demonstrates the power of the integrated PLMM approach.

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434

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Tables:

542   Table 1   Estimated variance of random effects for the PLMM (equation 1) used to detect phylogenetic

543   patterns in community composition.

PLMM	σ_a^2	σ_b^2	σ_c^2	σ_d^2	σ_e^2	$p(\sigma_c^2 = 0)$	AIC
Phylogenetic attraction: $c \sim \text{Gaussian}(\mathbf{0}, \text{kron}(\mathbf{I}_m, \sigma_c^2\mathbf{S}_{spp}))$	0.9789	1.265×10^{-7}	6.503×10^{-3}	7.712×10^{-10}	0.5154	0.008	3900
Phylogenetic repulsion: $c \sim \text{Gaussian}(\mathbf{0}, \text{kron}(\mathbf{I}_m, \sigma_c^2\,(\mathbf{S}_{spp})^{-1})$	0.9785	1.715×10^{-7}	1.227×10^{-9}	2.282×10^{-2}	0.5308	0.496	3906
Non-nested model: c removed	0.9838	2.287×10^{-2}	-	8.828×10^{-7}	0.5306	-	3904

560 Table 2 Phylogenetic signal and site variation for each functional trait. P-values for the null hypothesis

561 $\sigma_b^2 = 0$ (equation 2) implying no difference among sites in the effects of trait values on log abundance are

562 given in the column labeled $p(\sigma_b^2 = 0)$. Functional traits with strong phylogenetic signal and $p(\sigma_b^2 = 0) <$

563 0.05 are considered to be important in explaining phylogenetic patterns.

Trait	Pagel's λ	K	$p(\sigma_b^2 = 0)$
Leaf specific area (SLA, *m2/kg*)	0.70**	0.26**	**0.002**
Leaf circularity (Dimensionless)	1.00***	0.71***	**0.001**
Leaf thickness (*mm*)	0.96***	1.80***	**0.032**
Life cycle (Annual or non-annual)	0.00	0.30	0.479
Growth habit (woody or non-woody)	1.08***	0.24**	0.500
Pollination mode (Biotic or abiotic)	0.00	0.08	0.500
Seed mass (*g/seed*)	1.00***	0.46	0.077
Leaf dry mass content (LDMC, %)	0.52	0.16	0.500
Stem dry mass content (SDMC, %)	0.00	0.14	0.500
Plant height (*cm*)	0.00	0.16	0.500
Leaf length (*cm*)	0.98***	0.66**	0.500
Leaf width (*cm*)	1.00***	0.57***	0.206
Leaf carbon content (%)	0.65***	0.26**	0.500
Leaf nitrogen content (%)	0.34	0.09	0.334
Wind dispersal (Yes or no)	1.15***	0.45***	0.196
Animal dispersal (Yes or no)	0.64***	0.28**	0.072
Unassisted dispersal (Yes or no)	0.00	0.15	0.500

564 * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$

570   Table 3   Reduction of the phylogenetic variance in community composition caused by the inclusion of

571   functional traits (equation 3).

Trait	σ_c^2 with traits	σ_c^2 without traits	$100 \times \sigma_{c\ (\text{with traits})}^2 / \sigma_{c\ (\text{without traits})}^2$
SLA	0.006054	0.006461	6.30
Leaf circularity	0.006304	0.006415	1.73
Leaf thickness	0.006057	0.006459	6.22
SLA + circularity + thickness	0.005208	0.006437	19.09

590   Table 4   Variation in the response of species abundances to environmental variables (equation 4).

591   Although 13/20 environmental variables generated variation in species composition among communities,

592   none of these showed phylogenetic signal in which related species responded more similarly to the

593   environmental variable.

Environmental variables	P-values σ_c^2 (no phylogenetic signal)	P-values for σ_d^2 (phylogenetic signal)
Minimum temperature	**< 0.001**	0.500
Precipitation	**< 0.001**	0.500
Canopy shade	**0.002**	0.500
Total exchange capacity	**0.002**	0.500
Organic matter	**0.001**	0.500
pH	**< 0.001**	0.500
N	**< 0.001**	0.500
P	**0.039**	0.500
Mg	**0.030**	0.500
K	**0.007**	0.500
Na	**< 0.001**	0.500
Mn	**< 0.001**	0.354
Ca	**< 0.001**	0.122
Clay	0.110	-
Silt	0.070	-
Sand	0.117	-
Fe	0.500	-
S	0.458	-
Zn	0.500	-
Al	0.500	-

Figures:

Figure 1: Schematic diagram of the conceptual framework of the study. (A) Evolution is the ultimate source of all traits, although only some traits have phylogenetic signal that reflects phylogenetic history (arrows 2, 4 and 5). Other traits do not (arrows 1 and 3), possibly because these traits evolve rapidly or experience convergent evolution. Community composition is determined by unmeasured and measured traits, and also by additional processes that could generate phylogenetic signal, such as biogeographical patterns in the distribution of species. Phylogenetic patterns in community composition can be generated from measured and unmeasured traits with phylogenetic signal (arrows 7 and 8), and by other phylogenetic processes (arrow 6). The question we address is how much of the phylogenetic signal in community composition can be explained by measured functional traits, and whether after accounting for these traits there is residual phylogenetic signal that could have been generated by unmeasured traits or other phylogenetic processes. (B) Traits and phylogeny contain overlapping and complementary information about how communities are assembled. Here, we focus on estimating the proportion of this overlapping information that the phylogeny contains (i.e., $\frac{i}{i+ii+iii}$). Note that we do not try to explain the proportion of overlapping information that functional traits contain (i.e., $\frac{I}{I+II+III}$) due to our inability to estimate the amount of information provided by unmeasured traits and hence estimate ($I + II + III$). Note: panel B is just a heuristic diagram.

Figure 2: Phylogeny and relative abundance of common plant species found in the pine barrens of central Wisconsin in 2012. The area of dots is proportional to abundances within each site.

620   Figure 1

Sites

Aster macrophyllus
Euthamia graminifolia
Krigia biflora
Hieracium aurantiacum
Diervilla lonicera
Viburnum acerifolium
Aralia nudicaulis
Melampyrum lineare
Galium triflorum
Mitchella repens
Apocynum androsaemifolium
Ilex verticillata
Cornus canadensis
Vaccinium angustifolium
Gaylussacia baccata
Gaultheria procumbens
Epigaea repens
Pyrola rotundifolia
Lysimachia quadrifolia
Trientalis borealis
Aronia melanocarpa
Prunus virginiana
Prunus serotina
Spiraea alba
Fragaria virginiana
Potentilla simplex
Rosa spp
Rubus spp
Rhamnus frangula
Corylus americana
Betula papyrifera
Comptonia peregrina
Quercus ellipsoidalis
Quercus alba
Populus tremuloides
Populus grandidentata
Viola spp
Toxicodendron rydbergii
Acer rubrum
Anemone quinquefolia
Coptis trifolia
Phalaris arundinacea
Poa spp
Carex spp
Maianthemum canadense
Maianthemum racemosum
Cypripedium acaule
Smilax tamnoides
Uvularia sessilifolia
Pinus resinosa
Pinus banksiana
Pinus strobus
Osmunda claytoniana
Osmunda cinnamomea
Lycopodium clavatum

622   Figure 2

Appendix

Table S1　Estimated variance of random effects within phylogenetic generalized linear mixed models used to detect phylogenetic patterns (phylogenetic attraction and phylogenetic repulsion, estimated by σ_c^2).

Phylogenetic linear mixed models	σ_a^2	σ_b^2	σ_c^2	σ_d^2	$p(\sigma_c^2 = 0)$
Phylogenetic attraction: $\Pr(Y_i = 1) = \text{logit}^{-1}(\alpha + a_{\text{spp}[i]} + b_{\text{spp}[i]} + c_i + d_{\text{site}[i]})$ $c \sim \text{Gaussian}(\mathbf{0}, \text{kron}(\mathbf{I}_m, \sigma_c^2 \mathbf{S}_{\text{spp}}))$	2.8416	6.738×10^{-4}	0.0452	0.0110	<0.001
Phylogenetic repulsion: $\Pr(Y_i = 1) = \text{logit}^{-1}(\alpha + a_{\text{spp}[i]} + b_{\text{spp}[i]} + c_i + d_{\text{site}[i]})$ $c \sim \text{Gaussian}(\mathbf{0}, \text{kron}(\mathbf{I}_m, \sigma_c^2 (\mathbf{S}_{\text{spp}})^{-1}))$	3.1011	1.026×10^{-4}	0.0011	0.1936	0.5
$\Pr(Y_i = 1) = \text{logit}^{-1}(\alpha + a_{\text{spp}[i]} + b_{\text{spp}[i]} + d_{\text{site}[i]})$	2.8323	1.454×10^{-5}	-	0.1796	-

642   Table S2   Proportion of phylogenetic signal of species composition in communities explained by

643   individual functional trait and multiple functional traits. With selected multiple functional traits, only

644   about 19% percent of phylogenetic variation was explained, suggesting that phylogenies can provide

645   additional information about community assembly beyond measured functional traits. See equation 3 in

646   the Methods section for details about models.

Trait	σ_c^2 with trait	σ_c^2 without trait	$100 \times \sigma_{c\,(\text{with trait})}^2 / \sigma_{c\,(\text{without trait})}^2$
SLA	0.036213	0.042060	13.90
Leaf circularity	0.034731	0.041094	15.48
Leaf thickness	0.024457	0.042004	41.77
SLA + circularity + thickness	0.019557	0.044395	53.17

662   Table S3   There are strong variations in species' relationships between their abundance and most

663   environmental variables (*p* value of each environmental variable was presented in the *P*-values for

664   variation column). However, none of these variations show phylogenetic signal. For environmental

665   variable that has no strong variation in species' responses, no further test for phylogenetic signal of

666   variation was conducted (thus "-" in the third column). P-values that are less than 0.05 are in bold.

Environmental variables	P-values for variation	P-values for phylogenetic signal of variation
Minimum temperature	**<0.001**	**0.002**
Precipitation	**<0.001**	0.500
Canopy shade	**0.001**	0.500
Total exchange capacity	0.149	-
Organic matter	0.161	-
pH	**0.005**	**<0.001**
N	0.052	-
P	0.343	-
Mg	0.500	-
K	0.206	-
Na	**0.004**	0.500
Mn	**<0.001**	**<0.001**
Ca	**0.012**	**<0.001**
Clay	0.431	-
Silt	0.494	-
Sand	0.500	-
Fe	0.379	-
S	0.500	-
Zn	0.500	-
Al	0.500	-

669　Text S1 Codes to compare p-values of null hypothesis $\sigma^2 = 0$ calculated from the $0.5\chi_0^2 + 0.5\chi_1^2$

670　mixture distribution and paramatric bootstrap. The p-values based on the mixture Chi-squre

671　districution are conservative.

```
672  # packages used
673  library(ape)  # for phylogeny reading
674  library(plyr)
675  library(MASS)
676  library(dplyr, quietly = TRUE)
677  library(pez)  # for commuityPGLMM function
678  library(parallel)  # for multiple cores parallel computation, not available
679  # for Windows operation system

680  # data: vegetation data, phylogeny
681  load("d_li_data.RData")
682  # select 20 sites and 20 species
683  test = veg.aggr.wide.1958[1:20, 1:20]
684  test1 = filter(veg.aggr.long.1958, sp %in% names(test), site %in% rownames(te
685  st))

687  # this function calculates log likelihood of the fitted model on observed
688  # data, then simulates data based on the fitted model, and fits model on
689  # simulated data and calculates the log likelihood of the fitted model; then
690  # calculates the p-value of the log likelihood of the fitted model on
691  # observed data based all simulated ones (i.e. parametric bootstrap); so we
692  # can compare the p-value get in this way (parametric bootstrap) with the
693  # one from the mixture Chi-square distribution.
694  q1_obs_sim = function(veg.long, phylo = pb.phylo, date = 1958, trans = NULL,
695      nsim = 100, ncores = 5) {
696      # transformation of freq
697      if (!is.null(trans)) {
698          if (trans == "log") {
699              veg.long$Y <- log(veg.long$freq + 1)
700          }

702          if (trans == "asin") {
703              veg.long <- group_by(veg.long, site) %>% mutate(Y = asin(sqrt((fr
704  eq + 1)/ifelse(date == 1958, 20 + 2, 50 + 2)))) %>% ungroup() %>%
705                  as.data.frame()
706          }
707      }

709      veg.long$sp = as.factor(veg.long$sp)
710      veg.long$site = as.factor(veg.long$site)
711      nspp <- nlevels(veg.long$sp)
712      nsite <- nlevels(veg.long$site)

714      # Var-cov matrix for phylogeny
715      phy <- drop.tip(phylo, tip = phylo$tip.label[!phylo$tip.label %in% levels
716  (veg.long$sp)])
```

```
717        Vphy <- vcv(phy)
718        Vphy <- Vphy[order(phy$tip.label), order(phy$tip.label)]
719        Vphy <- Vphy/max(Vphy)
720        Vphy <- Vphy/det(Vphy)^(1/nspp)
721        Vphy.inv = solve(Vphy)
722
723        show(c(nlevels(veg.long$sp), Ntip(phy)))  # should be equal
724
725        # random effect for site
726        re.site <- list(1, site = veg.long$site, covar = diag(nsite))
727        re.sp <- list(1, sp = veg.long$sp, covar = diag(nspp))
728        re.sp.phy <- list(1, sp = veg.long$sp, covar = Vphy)
729        # sp is nested within site, to test phylo attraction or repulsion
730        re.nested.phy <- list(1, sp = veg.long$sp, covar = Vphy, site = veg.long$
731    site)
732        re.nested.rep <- list(1, sp = veg.long$sp, covar = Vphy.inv, site = veg.l
733    ong$site)
734
735        z <- communityPGLMM(Y ~ 1, data = veg.long, family = "gaussian", sp = veg
736    .long$sp, site = veg.long$site, random.effects = list(re.sp, re.sp.phy, re.si
737    te, re.nested.phy), REML = F, verbose = F, s2.init = 0.1)
738        show(z$ss)
739        z0 <- communityPGLMM(Y ~ 1, data = veg.long, family = "gaussian", sp = ve
740    g.long$sp, site = veg.long$site, random.effects = list(re.sp, re.sp.phy, re.s
741    ite),
742            REML = F, verbose = F, s2.init = 0.1)
743        z.rep <- communityPGLMM(Y ~ 1, data = veg.long, family = "gaussian", sp =
744    veg.long$sp, site = veg.long$site, random.effects = list(re.sp, re.sp.phy, re
745    .site, re.nested.rep), REML = F, verbose = F, s2.init = 0.1)
746        show(z.rep$ss)
747
748        # observed ouput, p-values are get from Chisq approx.
749        output_obs = data.frame(LRT_attract = (z$logLik - z0$logLik), p_attract =
750    pchisq(2 * (z$logLik - z0$logLik), df = 1, lower.tail = F)/2, LRT_repulse = (
751    z.rep$logLik - z0$logLik), p_repulse = pchisq(2 * (z.rep$logLik - z0$logLik),
752    df = 1, lower.tail = F)/2, obs_sim = "obs")
753
754        # the fitting model z0: log(y_i + 1) = alpha + a_spp[i] +
755        # b_spp.phy[i] + c_site[i] + err[i]
756        alpha = z0$B  # intercept, overall mean of all sp
757        alpha.se = z0$B.se  # SE
758        LRT_sim = mclapply(1:nsim, function(x) {
759            # multi-cores
760            set.seed(x)
761            # z0$ss: random effects' SD for the cov matrix \sigma^2 * V, in order
762    : [1]
763            # sp with no phylo; [2] sp with Vphy; [3] site random effect
764            a_spp = rnorm(nspp, 0, z0$ss[1])  # simulate a_spp
```

```
765        # simulate b_spp.phy
766        b_spp.phy = MASS::mvrnorm(1, mu = rep(0, nspp), Sigma = z0$ss[2] * Vp
767   hy)
768        mu_spp = alpha + a_spp + b_spp.phy  # mean freq of sp
769        c_site = rnorm(nsite, 0, z0$ss[3])  # site random
770        mu_spp_site = rep(mu_spp, nsite) + rep(c_site, each = nspp)  # each s
771   p at each site
772        y_i = rnorm(nspp * nsite, mean = mu_spp_site, sd = alpha.se)  # inclu
773   de SE of intercept
774        y_i_count = ceiling(exp(y_i) - 1)  # exp transf and round to positive
775   interge
776        test1_sim = data.frame(sp = names(mu_spp_site), site = rep(1:nsite,
777          each = nspp), Y = y_i, freq = y_i_count)
778
779        test1_sim$sp = as.factor(test1_sim$sp)
780        test1_sim$site = as.factor(test1_sim$site)
781
782        # refit models on simulated data random effect for site
783        re.site.sim <- list(1, site = test1_sim$site, covar = diag(nsite))
784        re.sp.sim <- list(1, sp = test1_sim$sp, covar = diag(nspp))
785        re.sp.phy.sim <- list(1, sp = test1_sim$sp, covar = Vphy)
786        # sp is nested within site
787        re.nested.phy.sim <- list(1, sp = test1_sim$sp, covar = Vphy, site =
788   test1_sim$site)
789        re.nested.rep.sim <- list(1, sp = test1_sim$sp, covar = Vphy.inv, sit
790   e = test1_sim$site)
791
792        z_sim <- communityPGLMM(Y ~ 1, data = test1_sim, family = "gaussian",
793          sp = test1_sim$sp, site = test1_sim$site, random.effects = list(r
794   e.sp.sim, re.sp.phy.sim, re.site.sim, re.nested.phy.sim), REML = F, verbose =
795   F, s2.init = 0.1)
796        # show(z_sim$ss)
797        z0_sim <- communityPGLMM(Y ~ 1, data = test1_sim, family = "gaussian"
798   , sp = test1_sim$sp, site = test1_sim$site, random.effects = list(re.sp.sim,
799          re.sp.phy.sim, re.site.sim), REML = F, verbose = F, s2.init =
800   0.1)
801        # show(z0_sim$ss)
802        z.rep_sim <- communityPGLMM(Y ~ 1, data = test1_sim, family = "gaussi
803   an", sp = test1_sim$sp, site = test1_sim$site, random.effects = list(re.sp.si
804   m, re.sp.phy.sim, re.site.sim, re.nested.rep.sim), REML = F, verbose = F, s2.
805   init = 0.1)
806        # show(z.rep_sim$ss)
807
808        # log lik of refitted models on simulated data
809        data.frame(LRT_attract = (z_sim$logLik - z0_sim$logLik), LRT_repulse
810   = (z.rep_sim$logLik - z0_sim$logLik))
811     }, mc.cores = ncores)
812
813
```

```
        # output results
    list(output_obs, ldply(LRT_sim))
}

qqq = q1_obs_sim(test1, trans = "log", nsim = 1000, ncores = 6)
saveRDS(qqq, "qqq.rds")

qqq = readRDS("qqq.rds")
qqq[[1]]

##    LRT_attract p_attract  LRT_repulse p_repulse obs_sim
## 1   0.3006013 0.2190598 -1.82719e-05       0.5     obs

head(qqq[[2]])

##    LRT_attract  LRT_repulse
## 1  -0.9611412 -15.68606765
## 2   0.1303866  -0.14712624
## 3  -2.9661583   0.06437319
## 4  -0.2152182  -1.91538503
## 5  -0.4204626   0.04073069
## 6  -1.3125998  -0.40523844

qqq[[2]]$obs_sim = "sim"
q1_sim = rbind(select(qqq[[1]], -p_attract, -p_repulse), qqq[[2]])
1 - (rank(q1_sim$LRT_attract)[1] + 1)/1001   # 0.12088 vs 0.219 from Chisq

## [1] 0.1208791

1 - (rank(q1_sim$LRT_repulse)[1] + 1)/1001   # 0.40959 vs 0.5 from Chisq

## [1] 0.4095904
```