Stochastic processes dominate community assembly in cichlid communities in Lake

Tanganyika

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ABSTRACT

The African Great Lakes are characterized by an extraordinary diversity of endemic cichlid

fish species. The cause of this diversity is still largely unknown. Most studies have tried to

solve this question by focusing on macro-evolutionary processes, such as speciation.

However, the ecological processes determining local cichlid diversity have so far been

understudied, even though knowledge on these might be crucial for understanding larger scale

biodiversity patterns.

Using trait, environmental and abundance data of cichlid fishes along 36 transects, we have

studied how differences in local environmental conditions influence cichlid community

assembly in the littoral of Lake Tanganyika, Zambia. We investigated changes in average trait

values and in trait-based community assembly processes along three key environmental

gradients.

Species diversity and local abundance decreased with increasing sand cover and diet-

associated traits changed with depth. Analyses on within-community trait diversity patterns

indicated that cichlid community assembly was mainly driven by stochastic processes, to a

smaller extent by processes that limit the similarity among co-existing species and least by

filtering processes that limit the range of species traits occurring in an environment. Despite

the low impact of habitat filtering processes, we find community dissimilarity to increase with

increasing environmental difference.

Our results suggest that local environmental conditions determine cichlid abundance, while

the predominance of stochastic community assembly across all environments explains why

the communities with the highest abundances contain most species.

    INTRODUCTION

    The stunning diversity of cichlid fish in the African Rift lakes has fascinated scientists during

    the past 50 years (Brooks 1950; Fryer & Iles 1972; Coulter 1991; Kocher 2004; Wagner *et al.*

    2012). In contrast to the multitude of research focusing on the evolutionary explanations of

    cichlid diversity (Sturmbauer *et al.* 2001; Salzburger *et al.* 2005; Muschick *et al.* 2012;

    Brawand *et al.* 2014), there are fewer studies aiming at understanding ecological mechanisms

    underlying local coexistence and diversity of cichlid species. Empirical studies on local scale

    diversity have focused either on temporal trends (Hori *et al.* 1993; Takeuchi *et al.* 2010), the

    impact of human disturbance (Alin *et al.* 1999), opportunities for relieving fishing efforts

    (Duponchelle *et al.* 2003; Weyl *et al.* 2005), the impact of protected areas on cichlid

    communities (Sweke *et al.* 2013) or have been restricted to descriptions only (Hori *et al.*

    1983; Kuwamura 1987; Van Steenberge *et al.* 2011). While these studies have shown

    interesting patterns of local cichlid communities, no studies so far have attempted to

    understand the processes underlying local community composition patterns.

    Community assembly is generally viewed from two alternative perspectives, either from a

    niche-based perspective, or a neutral perspective. The niche-based hypothesis postulates that

    species are adapted to their local environment and occupy a species' specific niche: a set of

    conditions in which the species thrives, and outcompetes other species (Hutchinson 1959;

    Tilman 1982; Chesson 2000). The traits of a species reflect the adaptation of a species to its

    own niche, and studying patterns in trait-distributions can inform us about any underlying

    processes driving species-coexistence and community composition. In benign environments

    that do not pose strong requirements on traits, the niche-based hypothesis predicts that species

    are abundant, and that the presence or absence of species' traits is mainly shaped through

    interactions with each other, rather than interactions with the abiotic environment. Due to

    competitive exclusion of species with overlapping niches or due to exclusion of species with

shared specialist predators, niche-based assembly is expected to lead to an *increase* in trait

diversity of co-occurring species in benign environments (Macarthur & Levins 1967;

Mayfield & Levine 2010). In harsh environments, species with traits that make them

intolerant to stress, herbivory and/or predation pressures might be excluded from a local

community, leading to a *reduced* observed trait diversity of co-occurring species (Weiher &

Keddy 1995; Cornwell *et al.* 2006).

In contrast, the neutral hypothesis, which treats all individuals from all species equally,

explains community composition from a stochastic point of view, where local abundance of a

species is the outcome of stochastic birth, death and migration over time (Hubbell 2001;

Rosindell *et al.* 2011, 2012). Local community composition is assumed to be a dynamic

equilibrium between random immigration from the species pool, and local ecological drift.

The neutral hypothesis acknowledges that there might be benign and harsh environments, but

that these environments affect all individuals equally. As a consequence, benign environments

have more individuals than stressful environments, but both benign and stressful areas contain

individuals that form a (dispersal-limited) random subset of the species pool. The null

expectation is then that areas with high abundances also have high diversity, as a result of

random sampling.

Here we investigate how niche-based and neutral-based community assembly shapes

biodiversity in cichlid fishes across three environmental gradients: depth, sand cover and

topographical complexity. If community assembly is highly niche-driven, we firstly expect

that decreasing algae densities with depth provide less scope for algae-feeding cichlids and

that hence, communities in deep areas are predominantly assembled by filtering processes.

Secondly, because many cichlids are highly dependent on rocks for either food (as substrate

for algae growth) or for shelter (Konings 2005), we expect that an increase in sand cover,

which brings a reduction of available rock surface, will increase habitat filtering and decrease

diversity. However, males of several cichlid species use sand to construct bowers to attract

females and are thus highly dependent on sand, which could counteract filtering processes and

possibly drive limiting similarity processes (Sefc 2011). Thirdly, we expect that areas with

high topographical complexity induce higher levels of limiting similarity and higher diversity,

because a more complex habitat provides more shelter and a larger surface area for algae

growth. Furthermore, complex habitats are often associated with a reduction in territoriality

(Danley 2011), a larger number of niches (Willis *et al.* 2005) and higher diversity (Ding *et al.*

2014).

Alternatively, if neutral-based processes dominate, we expect resource-rich areas, such as

shallow, rocky and complex habitats that allow for abundant algae growth, to show highest

diversity, as these habitats can sustain the highest number of individuals. Furthermore, we

expect these habitats to have the highest species diversity, and we expect that the trait

compositions of such communities to reflect a random subset of the larger species pool.

METHODS

Abundance data

Abundance and composition data of cichlids was collected in Lake Tanganyika, near

Kalambo Lodge (8°37'22.29"S, 31°12'1.89"E), Zambia, Africa (Figure 1) using SCUBA

diving. In total, 12 different sites were sampled, with a total of 36 transects, which were

placed parallel to the shore (Figure 1, Table 2). Individuals were recorded along 20m x 4m

transects by two divers in two steps: first, all individuals within 2 meters on one side of the

transect were sampled. After an interval of 10 minutes, all individuals within 2 meters on the

other side of the transect were sampled.

Figure 1. Sketch of the sampling positions in front of the Kalambo Lodge, located in the south
of Lake Tanganyika. Relative position of the transects at every site is indicated in the left
panel. Numbers in the figure refer to site numbers in Tables S1, S2 and S3.

Environmental data

At each transect, we measured three different ecological variables. Firstly, we recorded the

depth at the beginning and the end of each transect. Secondly, we took 40 photos of 50 x 50

cm quadrats per transect, to estimate the percentage of sand cover (20 photographs on the left

side of the transect line, and 20 photographs on the right side). Percentage of sand cover per

quadrat was calculated using an image analysis script in MATLAB (supplementary

information). The average proportion of sand across all photographs taken along the transect

was used to quantify sand cover of the transect. Thirdly, we recorded the topographical

complexity of the substrate using a variation of the 'chain method' (Risk 1972). Along the

transect, a link-brass chain was laid over the substrate, such that the chain would closely

follow the contour of the substrate (Shumway *et al.* 2007). Topographical complexity was

calculated as the ratio between the distance following the contour (measured with the chain)

and the horizontal linear distance. High values indicate high relief/complexity, caused by an

alternation between large rocks and small rocks or sand, while low values indicate low

relief/complexity.

Species traits

Data on mean trait values of each cichlid species as reported in Muschick et al. (2012, 2014)

were collated for 11 traits: standard length, total length, weight, stable isotope ratios of carbon

and nitrogen, lower pharyngeal jaw height, lower pharyngeal jaw width, gut length, lower

pharyngeal jaw shape and body shape. In order to obtain a set of traits that does not strongly

correlate with each other, to avoid the overemphasis of the importance of some traits over

others, we only used a subset of these traits. Firstly we checked the Pearson Rank correlation

between traits and removed one of the traits that correlated strongly. Total length and weight

were found to correlate strongly with standard length (standard length vs total length: $R^2 =$

0.98, standard length vs weight: $R^2 = 0.75$), and we decided to only use standard length.

Furthermore, lower pharyngeal jaw (LPJ) width and lower pharyngeal jaw height were found

to correlate with each other ($R^2 = 0.67$) and to correlate with standard length (lpj height vs

standard length: $R^2 = 0.44$, lpj width vs standard length: $R^2 = 0.37$). Partly to avoid

overemphasis of pharyngeal jaw traits, and partly because of the high correlation coefficients

we decided to omit both these traits. Lower pharyngeal jaw shape and body shape were

assessed using landmark-based geometric morphometric methods. *xy* coordinates of the

landmarks were combined into PCA components for both body shape and lower pharyngeal

jaw shape. From these PCA components we included the first PCA component for both body

and lower pharyngeal jaw shape, which contained 37% and 49% of the variation for body and

pharyngeal jaw shape respectively. The inclusion of more PCA components would allow us

to cover a larger part of variance in shapes, but would overemphasize the importance of shape

relative to the other traits.

The final trait set thus consisted of 6 traits: standard length, stable isotope ratios of carbon and

nitrogen, gut length, the first PCA component of LPJ shape and the first PCA component of

body shape. Missing trait values were imputed using the MICE package for R (Buuren &

Groothuis-Oudshoorn 2011). The MICE imputation process uses a Gibbs sampler technique

to impute the missing data, assuming a multivariate distribution for the missing data.

Trait-based community assembly

In order to infer the relative contribution of limiting similarity, habitat filtering and dispersal

assembly we used the STEPCAM approach (Van der Plas *et al.* 2015). The STEPCAM model

is a STEPwise Community Assembly Model, which implements community assembly by

applying three types of processes that mediate species selection from the species pool to the

local community: 'filtering processes', 'limiting similarity processes' and 'stochastic

processes'. Starting with all observed species in the dataset, species are removed in a stepwise

fashion until the number of species observed in the local community is reached. Removal of

species occurs either 1) because their traits are too dissimilar from the observed mean trait

distribution in the community, which is assumed to be the habitat optimum ('filtering'), 2)

because their traits are too similar to the other remaining species ('limiting similarity'), or 3)

due to a stochastic event, which results in a random removal step, where the probability of

removal is proportional to the number of local communities in the dataset where the species is

observed, which is used as a proxy for the species pool. The STEPCAM model was fitted

using Approximate Bayesian Computation, where using the model, data is simulated and

compared with the observed data. Comparison between simulated and observed data occurred

through comparing four summary statistics, Functional Richness, Functional evenness,

Functional divergence and community trait means (Villéger et al. 2008, van der Plas et al.

2015). We applied a Sequential Monte Carlo algorithm (ABC-SMC) using the function

STEPCAM_ABC from the package STEPCAM (Janzen & van der Plas 2014). We used 1000

particles and a final acceptance rate of 1 in 20,000. The reported estimates for dispersal

assembly, filtering and limiting similarity are mean estimates over 3 replicate STEPCAM

runs, with the random number generator seeded with different seeds for each replicate.

The effect of the environment on community trait means

In order to understand changes in trait values that might be linked to changes in the

environment, we analyzed the relationship between the different traits and environmental

variables. We calculated, per transect, community-level weighted trait means (CWM), where

trait values are weighted by the relative frequency of the species in the transect (Lavorel *et al.*

2008). CWM values were calculated using the function 'dbFD' from the R package FD

(Laliberté *et al.* 2014). We used Linear Mixed Models to test how CWM values changed over

environmental gradients. We constructed full models, where CWM values were treated as

dependent variable, the three habitat components as fixed effect, and the site as a random

effect. We then used stepwise removal to remove non-significant effects. Conditional R^2

values were calculated following Nakagawa and Schielzeth (2013).

We repeated this approach using unweighted trait means, thus treating every species equally

and disregarding differences in abundance. Using unweighted trait means has two benefits

over using weighted means: firstly it does not overemphasize species that occur in a clustered

fashion, such as the breeding complexes of *Neolamprologus pulcher*. Secondly it does not

underestimate species that, although occurring in low abundance, influence species around

them strongly, for instance due to territoriality or aggressive behavior, such as *Plecodus*

straeleni (Boileau *et al.* 2015).

Community dissimilarity

To assess the simultaneous effect of all three environmental gradients on community

composition, we compared community dissimilarity between all communities. We

constructed an environmental distance score by calculating the distance between all transects

for each gradient (depth, sand and complexity). To ensure that all environmental factors had a

similar weight on environmental heterogeneity, we normalized the distance scores by the

maximum distance, such that all individual distance scores where between -1 and 1. We then

obtained the total normalized environmental distance by calculating the Euclidian distance by

taking the square root of the sum of squared distance scores.

RESULTS

Species compositions

We recorded on average 22 species per transect, and an accumulated total number of 53

species (Table S3). *Telmatochromis temporalis* was the most common species, contributing

12% of 4926 recorded individuals. The seven most common species combined accounted for

50% of all observed individuals (*Telmatochromis temporalis, Variabilichromis moorii,*

Tropheus moorii, Neolamprologus pulcher, Interochromis loocki, Telmatochromis vittatus

and *Xenotilapia boulengeri*), whilst the 24 most common species accounted for 90% of all

individuals.

Transects with a higher sand cover had a lower number of individuals (R^2 = 0.41, p = 0.021),

and a lower species richness (R^2 = 0.47, p = 0.001) (Table 1, Figure 2). Neither depth nor

habitat complexity had significant effects on abundance or species richness (Table 1, Figure

2).

Figure 2. Changes in abundance and in species richness across the three habitat gradients.
Points depict the different transects. Significant correlations (table 1) are plotted as a line.

Figure 3. Changes in Bray-Curtis dissimilarity between sampled transects versus the
normalized environmental distance between sampled transects. Normalized environmental
distance is the Euclidian distance between two plots, where the environmental distances are
normalized by the maximum value recorded across all plots.

We compared the Bray-Curtis dissimilarity between transects with the normalized Euclidian

distance between the environmental conditions at those transects. Bray-Curtis dissimilarity

was found to correlate significantly with the normalized environmental distance score (figure

3, R^2 = 0.59, Mantel test, mantel-r statistic = 0.766, p = 0.00001, 100.000 permutations).

Relative contribution of community assembly processes

Fitting STEPCAM to the trait distributions of the 36 different transects yielded an average

contribution of dispersal assembly steps of 72%, an average contribution of habitat filtering

steps of 9% of the steps and 19% of removal steps due to limiting similarity (Figure 4). We

found no significant correlations between relative contribution of either of the three processes

and any of the three habitat gradients (Figure 5).

Figure 4. Relative contribution of stochasticity, habitat filtering and limiting similarity steps
across all 36 sites.

Figure 5. Relative contribution of stochasticity, habitat filtering and limiting similarity steps
as estimated using STEPCAM, plotted against the three measured habitat components: depth,
sand cover and complexity. None of these relationships was significant.

Relationships between traits and habitat components

We found that the three habitat gradients explained only a low proportion of variation (R^2 around 0.5) of community-weighted means (CWM) of cichlid traits (Table 1). We observed high R^2 values ($R^2 > 0.7$) for δ15N content, δ13C content, lower pharyngeal jaw width, the first axis of the PCA of the Lower Pharyngeal Jaw shape, and the third axis of the PCA of the body shape (Table 1). Using community unweighted trait means we find similar effects, with similar R^2 scores (Table 1), suggesting that weighting the trait values by abundance does not influence our findings.

Table 1, Significant components of linear mixed-effects models, where species traits were used as response variable, the habitat components as predictor variables, and the sampling site as a random effect. Only those components that were significant after stepwise removal of all non-significant components are reported. The conditional R^2 of the final model is reported in the last column. Used abbreviations: LPJ = Lower Pharyngeal Jaw, PCA = Principal Component Axis.

Response variable	Depth	Sand	Complexity	R^2
Species richness		-0.1143		0.47
[10]log abundance		-0.0033		0.67
Weighted Trait Means				
Standard Length		0.1386		0.54
Total Length		0.1652		0.57
Weight		0.0505	5.1724	0.40
δ15N	0.0553		-0.3885	0.78
δ13C	-0.1754			0.73
LPJ height		0.0074		0.69
LPJ width	0.0002	0.0001		0.72
Gut Length			171.79	0.52
LPJ PCA axis 1	-0.2727	-0.0185		0.84
LPJ PCA axis 2	-0.0611		2.3922	0.54
LPJ PCA axis 3		-0.0100	0.8428	0.56
Body PCA axis 1		0.0404		0.67
Body PCA axis 2	0.0770	-0.0092		0.67
Body PCA axis 3	-0.0989	0.0141	-1.6877	0.86
Unweighted Trait Means				
Standard Length	-0.2700	0.0413	7.7057	0.54
Total Length			7.3057	0.48
Body Weight	-0.3291		5.3659	0.58
δ15N	0.0366			0.76
δ13C	-0.0899	-0.0066		0.78
LPJ height	-0.0109			0.32
LPJ width	-0.0002			0.29
Gut Length	-8.8453		130.8656	0.58
LPJ PCA 1	-0.0840	-0.0109	-1.3614	0.78
LPJ PCA 2	-0.0880		0.8963	0.54
LPJ PCA 3		-0.0039	0.7303	0.51
Body PCA 1	0.0668	0.0126		0.65
Body PCA 2	0.0667	-0.0041		0.67
Body PCA 3	-1.5261			0.60

DISCUSSION

We have investigated whether community assembly in cichlid communities at the littoral of

Lake Tanganyika, Zambia is mostly driven by niche-based processes, or by neutral-based

processes. In line with the neutral hypothesis, we found that while abundance and richness of

cichlids became higher in benign conditions, within-community traits distributions were

highly similar to random samples from the species pool distribution. Despite some modest

relationships between average trait values and habitat components, the within-community trait

distribution reflects a dominance of stochastic processes in community assembly. Niche-

based processes only contributed minimally to community assembly and instead, observed

community patterns pointed to an important role for neutral processes in community

assembly.

Although habitat filtering only contributed to community assembly to a limited extent,

environmental conditions were not without influence on local community patterns. We found

that an increase in sand cover was correlated with a lower total abundance in the local

community. This could be an effect of a reduction in food resources, as there is less substrate

for algae growth (a common food source for many cichlids). However, individuals of most

species showed similar trends, which explains why a) both in areas with high and low sand

cover local communities resembled a random subset of the species pool, and b) diversity is

highest in low sand areas: this is mostly the result of random sampling from the species pool.

Sand cover thus drives local community patterns through neutral based processes, rather than

niche-based processes.

Although we found no changes in the relative contribution of the different community

assembly processes across the different environmental gradients, we did find that species

compositions shifted over these gradients. With an increasing difference between the local

conditions between two sites, their species composition became more dissimilar. This seems

to suggest that habitat filtering does impose an effect on community composition. One

explanation of our lack of finding habitat filtering effects in the STEPCAM analysis could lie

in our choice of traits. In our trait-based analyses we focused on traits associated with shifts in

diet and survivability (total length, gut length and pharyngeal jaw shape). We expected these

traits to differ most across different habitats, as they are thought to be related to different

trade-offs regarding resource retrieval and survival (Muschick *et al.* 2012; McGee *et al.*

2015). The inclusion of traits specifically linked to sexual selection, such as male coloration,

sexual dichromatism or sexual dimorphism, could change our estimates and possibly reveal

novel relationships between habitat components and stochasticity, habitat filtering or limiting

similarity.

Even though observed community patterns were in line with a dominant role for stochastic

processes in community assembly, trait-based processes were not completely absent. This was

shown by our stepwise community assembly models, and confirmed by patterns in

community weighted trait means over environmental gradients. We found significant trait-

environment relationships for stable nitrogen and carbon isotopic signatures, lower

pharyngeal jaw shape and shape of the body. The first three of these traits are associated with

food uptake and diet and mainly change with depth (associated with light availability and

hence algal density), thus confirming our expectation that there is a diet shift along our

environmental gradients.

Here we have presented a first analysis of cichlid diversity from a community assembly perspective. We have focused on disentangling niche-based and neutral processes and found that the majority of community assembly in cichlid communities at the littoral of Lake Tanganyika, Zambia is driven by stochastic effects, rather than trait-based effects. Interestingly, the dominance of stochastic processes in shaping local diversity is in line with previous findings focusing on factors influencing cichlid diversity from a more macro-evolutionary (rather than local ecological) perspective. Using advanced regression techniques, Wagner and colleagues (2012, 2014) have shown that the success, and the diversity, of a cichlid radiation depends mainly on the depth, the size and the total received solar input of a lake. These three factors are the main drivers behind the maximum number of individuals that a lake can sustain, which in turn determines the number of species, following neutral theory (Rosindell & Cornell 2009).

Furthermore we suggest that our findings are in line with the previously hypothesized stages of adaptive radiation in cichlid fish (Kocher 2004). Kocher hypothesized that cichlids have first adapted to the local habitat (habitat filtering), after which diversification proceeded through diversification of the feeding apparatus (limiting similarity), and lastly diversification proceeded through diversification of the color pattern due to sexual selection. Our findings resonate this order, as we find that adaptations to the local habitat (habitat filtering) are currently least important in shaping diversity, followed secondly by trophic adaptations driven by competition for food (limiting similarity). Although the current contribution of community assembly processes is not necessarily a strong predictor of the evolutionary path that cichlids have followed, the resemblance is striking at least.

Here we studied cichlid biodiversity from a micro-ecological perspective, rather than from a

(macro-) evolutionary perspective. Ultimately, a combination of these alternative approaches

might be most effective in uncovering why the diversity of cichlids in the East African Rift

lakes is so extremely high.

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