Plant root pathogens over 120,000 years of temperate rainforest ecosystem development

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developed the methods for, and AMW carried out, molecular characterisation. SJR collected

plant community data and additional plant trait data. IAD conducted data analysis with input

from AMW and JMT. IAD produced the figures and wrote the first draft with AMW, all

other authors made editorial input.

Abstract

The role of pathogens, including oomycetes, in long-term ecosystem development has

remained largely unknown, despite hypotheses that pathogens drive primary succession,

determine mature ecosystem plant diversity, or dominate in retrogressive, nutrient-limited

ecosystems. Using DNA sequencing from roots, we investigated the frequency and host

relationships of oomycete communities along a 120 000 year glacial chronosequence.

Oomycetes were frequent in early successional sites (5 - 70 yrs), occurring in 38 – 65% of

plant roots, but rare (average 3%) in all older ecosystems (280 yrs and older). Oomycetes

were highly host specific, and more frequent on plant species that declined most strongly in

abundance between ecosystem ages. In contrast, oomycetes were not correlated with plant

abundance or plant root traits associated with retrogression. The results support the

importance of root pathogens in early succession, but not thereafter, suggesting root

pathogen-driven dynamics may be important in driving succession but not long-term

diversity maintenance.

Introduction

As ecosystems develop, biotic communities undergo succession both in composition and in the nature of mutualistic and antagonistic interactions. Many of these changes have been well studied, including changes in plant communities and traits, competition intensity, and mutualisms (Lambers *et al.* 2008). Nonetheless, long-term successional changes in composition and community interactions of many other organisms, such as pathogens, remain poorly understood. Indeed, plant pathogens are often out-of-sight out-of-mind, until large-scale forest diebacks (Scott & Williams 2014) or major crop losses (Martin & Loper 1999) occur. This lack of information is particularly apparent in forest disease outbreaks, where it is often unclear whether the pathogen is native, from where it may have originated, or what background levels of pathogens are typical in healthy ecosystems (Podger & Newhook 1971). Nonetheless, pathogens occur as part of healthy ecosystems and play an important role in ecosystem processes (Castello *et al.* 1995; Gomez-Aparicio *et al.* 2012; Bever *et al.* 2015). For example, *Pythium* and other oomycetes (Oomycota = Peronosporomycota) can drive negative density dependence in plants (Packer & Clay 2000) and hence are believed to contribute to the maintenance of forest plant diversity (Mangan *et al.* 2010a). Pathogenic organisms, including oomycetes, are also an important component of soil biodiversity, as pathogens of plant roots and soil invertebrates, and as saprotrophs (Arcate *et al.* 2006). However, the ecological literature is highly divided in predicting when pathogens might be important in ecosystem succession and development.

One view of soil pathogens in relation to ecosystem development is that they play a particularly critical role in driving early succession, which we define as the period of large changes in vegetation composition with time (van der Putten *et al.* 1993; van der Putten

2003). This view is based, in part, on the assumption that newly establishing plant species

will encounter only low levels of pathogens in the absence of conspecific host plants during

succession. Over time, these established plants accumulate pathogens, which prevent

establishment of conspecific seedlings but have less negative effects on other plant species.

Part of the role of pathogens in succession is driven by plant age, as pathogens that kill

seedlings are often tolerated by established plants (Martin & Loper 1999; van der Putten

2003). This implies that early-successional (*r*-selected) plants may have limited selection for

pathogen resistance. Instead, these species may accumulate and tolerate pathogens as mature

plants, and rely on seed dispersal into new habitats to escape pathogens during the vulnerable

seedling stage (van der Putten 2003). Support for the role of pathogens in early succession

comes from studies of invertebrates in grasslands, where dominant early-successional species

are more suppressed than subordinate or late-successional species (de Deyn *et al.* 2003) and

from studies showing negative plant-soil-feedback early in succession, with mid- and late-

successional ecosystems showing neutral or positive feedback, respectively (Kardol *et al.*

2006).

A second view is that pathogens are of little importance early in succession, but increase in

importance in mature ecosystems, where vegetation composition is more stable (Reynolds *et*

al. 2003; Bardgett *et al.* 2005; Peltzer *et al.* 2010). Soil biotic communities as a whole

increase in biomass and species diversity over periods of tens to hundreds of years (Bardgett

et al. 2005), and pathogen communities may follow similar patterns of increasing biomass

and diversity. Further, Gilbert (2002) suggests that increasing nutrient competition in mature

ecosystems may stress plants and increase susceptibility to pathogens. Pathogen host-

specificity in these systems leads to density-dependent mortality (Augspurger 1984). Density

dependence may occur within a cohort, driven by a high density of co-occurring conspecific

seedlings near seed sources, or cross-cohort, with mature plants supporting high levels of

pathogens on their roots and hence reducing establishment of conspecific seedlings (Connell

1970), but most studies do not distinguish these two possibilities. Evidence for negative

density dependence has come from both tropical and temperate forests (Packer & Clay 2003;

Mangan *et al.* 2010b), although other studies have found positive density dependence or no

effect (Reinhart *et al.* 2012).

One of the signatures of density dependence should be a correlation between plant abundance

and plant-soil feedback strength, where soil feedbacks are defined as the growth of seedlings

of a species in a soil community developing under con-specific plants relative to other species

in the same soil (Bever *et al.* 2010). Several studies suggest that correlation between plant

abundance and feedback strength should be negative, with more abundant species showing

stronger negative feedback (Queenborough *et al.* 2007). However, positive correlations of

plant abundance and the strength of negative soil feedbacks have also been taken as evidence

supporting pathogen driven density-dependence as a mechanism for supporting mature forest

plant diversity (Xu *et al.* 2015). A logical argument can be made for either direction of

correlation, depending on whether you assume that observed abundances are the outcome of

prior negative density dependence (Xu *et al.* 2015) as opposed to drivers of current negative

density dependence. Direct measurement of pathogen frequency is not typically part of plant-

soil feedback studies, but strong negative feedback is often assumed to indicate high

pathogen loads. Nonetheless, other soil organisms, including mycorrhizas, can drive similar

outcomes (Bever *et al.* 2010).

Finally, a recently proposed third view of the role of pathogens in ecosystem development is

that soil pathogens may be particularly critical in retrogressive ecosystems (Laliberté *et al.*

2015), where declining soil nutrient availability results in declining vegetation stature and

diversity (Richardson *et al.* 2004; Porder *et al.* 2007; Peltzer *et al.* 2010). Laliberté and

colleagues (2015) suggest that the highly weathered, P-limited soils of retrogressive

ecosystems favour species with ephemeral roots to maximize nutrient uptake, but that this

imposes a trade-off with increased susceptibility to pathogens. Ecosystem retrogression has

been linked to increased specific root length (length per unit mass), thinner roots, and

increased root branching (Holdaway *et al.* 2011), all of which may increase susceptibility to

root pathogens (Laliberté *et al.* 2015). This hypothesis makes no explicit prediction about

how plant abundance will be correlated with pathogens.

We formalized each of these views from the literature into three testable, non-mutually

exclusive, sets of hypotheses. Each set makes specific predictions about the occurrence of

pathogens at different ecosystem stages (H1.1, H2.1, H3.1) and the occurrence of pathogens

on different hosts based on plant strategies and root traits (H1.2, H2.2, H3.2).

H1. Early-successional ecosystem hypotheses

H1.1 Pathogen frequency and diversity are high in early-successional ecosystems relative

to later stages.

H1.2 Pathogen-driven succession will be evident as plant species that decline in

abundance having higher pathogen loads than plant species that increase or maintain their

abundance.

H2. Mature ecosystem hypotheses

H2.1 Pathogen frequency and diversity are high in mature ecosystems relative to other

stages.

H2.2 Negative density-dependent mortality in mature ecosystems will be evident as a

correlation between plant abundance with pathogen frequency, either positively

(Queenborough *et al.* 2007; Bever *et al.* 2015), or negatively (Xu *et al.* 2015).

H3. Retrogression hypotheses:

H3.1 Pathogen frequency is high in retrogressive ecosystems relative to earlier stages.

H3.2 Trade-offs between nutrient acquisition and pathogen resistance will be evident as

plant species with high specific root length, thinner roots, and abundant root hairs having

higher pathogen loads than species without these traits.

In addition, we tested the hypothesis (H4) that oomycete communities are host-specific in

their plant associations, as host specificity is requisite for pathogens to either drive early-

successional vegetation change (H1) or to contribute to maintaining plant diversity in mature

ecosystems (H2).

We tested our hypotheses using an extensively studied soil and ecosystem development

chronosequence created by the Franz Josef glacier in New Zealand, where glacial advances

and retreats have created a series of soils of varying age in close proximity (Walker & Syers

1976; Richardson *et al.* 2004). We focused on oomycetes as widespread pathogens that have

been directly linked to plant succession and negative density-dependence in forest

ecosystems, as well as forest die-back (Packer & Clay 2000). Using direct DNA sequencing

from roots, we provide the first study of pathogen dynamics throughout ecosystem

development from early-succession to retrogression, and find strong support for the role of

oomycetes in early-succession but not in later stages of ecosystem development.

Methods

Study site and sampling

The Franz Josef chronosequence includes early successional (5, 15, 70 years of development), mature (280, 500, 1000, 5000, 12 000 years), and retrogressive (60 000, 120 000 years) sites. As soils age, there are dramatic changes in nutrient availability (declining P, increasing and then decreasing N), pH, and physical properties (Walker & Syers 1976). These changes are linked to changes in plant communities, with plant biomass increasing through succession to mature stages, and then entering retrogression where declining soil nutrients result in a concomitant decline in plant biomass, stature, and diversity (Richardson *et al.* 2004) and shifts in root traits (Holdaway *et al.* 2011). Vegetation shifts from a rock field with sparse herbaceous plants and sub-shrubs (5 years) to shrub land (15 years) to angiosperm forest (70 years), followed by an increasing dominance of large gymnosperm trees (Podocarpaceae) through mature stages, with an eventual decline in plant biomass, canopy height, and canopy closure in retrogression. Rainfall is high along the entire chronosequence, ranging from 3.5 to 6.5 m, and all sites are below 365 m elevation (Richardson *et al.* 2004).

In a previous study we collected 510 individual plant roots from ten sites along the chronosequence and characterized both plant identity and arbuscular mycorrhizal fungal communities in those roots (Martinez-Garcia *et al.* 2015). Here we use these same samples but used taxon-specific primers to amplify DNA from oomycetes. The collection of samples is described in full in Martinez-Garcia *et al.* (2015). In brief, we collected 51 individual root fragments, taking a single root fragment (approx. 15 mg dry weight) every 2 m along two parallel transects with equal sampling from three depths. For the 5-year-old site, which comprised sparse vegetation in rocks, we collected the nearest plant to the sample point and

sampled one root from that plant. This was a necessary given that most of the soil at the 5-

year-old site was root free. Roots from all sites were thoroughly rinsed in water and dried

before DNA extraction. DNA was extracted using a MoBio soil DNA kit, and plant species

identified by PCR amplification and DNA sequencing of the tRNL gene region, except for

the 5-year-old site, where the plant was already known due to collection method. The one

difference in the sampling between this study and the prior study (Martinez-Garcia *et al.*

2015) was that in the earlier paper a spare ("B") sample was used if the first ("A") sample

failed to yield both plant and arbuscular-mycorrhizal fungal PCR products. In the present

study we did not use the "B" samples, as it would have made quantification difficult.

Oomycete PCR and identification

We primarily based our oomycete detection, operational taxonomic unit (OTU) clustering,

and identification on a nested PCR of large-subunit (28S) DNA (supplementary methods). A

negative (omitting DNA template) and a positive control (including genomic DNA from a

Pythium species) were included in every PCR. PCR products were purified using DNA Clean

and Concentrator Kit (Zymo Research Corporation) prior to performing Sanger sequencing

(Canterbury Sequencing and Genotyping, University of Canterbury, New Zealand).

Some ecosystem ages had a high proportion of samples that failed to amplify a product of the

expected size in the first-round PCR. In order to ascertain whether these were true absences

of oomycetes or due to PCR inhibition, nine samples from each of these sites with the highest

failure rate (500, 100, 1200, 60 000 and 120 000 yrs) were tested for PCR inhibitors by

repeating PCR reactions in duplicate for each sample, with one of the duplicate reactions

spiked with 20 ng positive control DNA. Only one out of the 45 samples spiked with positive

control DNA failed to amplify, suggesting PCR inhibitors were not likely to be causing low

detection rates.

To confirm identities of sequenced oomycetes, samples that produced positive large subunit

PCR products were sequenced for the internal transcribed spacer (ITS) 1 region

(supplementary material). As DNA sequences were obtained from environmental samples

(root fragments) potentially containing multiple oomycetes, the ITS sequence may or may not

represent the same species as the 28S sequence. ITS sequences were therefore used to help

inform identification, but all analyses were based on the 28S results.

Sequence bioinformatics

All DNA sequences were matched against GenBank using BLAST to find the closest

matching sequence, and, where the closest match had no reliable taxonomic identity, the

closest matching sequence associated with a taxonomic identity. Sequences that matched

non-oomycete specimens were discarded. Three sequences had a closest match to a

Cercozoan sequence, but only at 78% identity. The phylogeny of Cercozoans remains

unclear, but some classifications place this group within the same kingdom (Chromista) as

oomycetes (Cavalier-Smith 2010). We therefore considered these three sequences to be

potentially basal oomycetes and retained them in the analysis. Out of 510 root samples, 122

had positive PCR products with oomycete primers. After filtering for sequence quality and

matching to an oomycete sequence, the final dataset contained 91 DNA sequences of

oomycetes with a mean sequence length of 310 bp. Plant IDs were obtained for 458 samples

overall and 86 of the 91 samples with oomycete products through DNA sequencing of the

trnL gene (Martinez-Garcia *et al.* 2015). We clustered sequences into OTUs using BLASTn

to merge any sequences with > 97% similarity over at least 95% of the shorter sequence

length based on single-linkage clustering. Finally, we assembled all of our environmental

DNA sequences and a selection of named sequences from GenBank to examine how our

OTUs were reflected in phylogenetic clustering. Phylogenetic clustering was determined

using Clustal Omega for alignment and ClustalW2 phylogeny based on both UPGMA

clustering (ebi.ac.uk). The single-linkage BLAST clustering and phylogenetic clustering were

broadly consistent.

Root traits

Root trait data (root diameter, root hairs per mm, and specific root length (SRL)) had been

previously obtained for most of our plant species in Holdaway *et al.* (2011). Intraspecific root

trait variability across sites is minimal (Holdaway *et al.* 2011). We collected new trait data

for two species that were abundant in our data but missing from Holdaway *et al.* (2011),

Raoulia hookeri and *Epilobium glabellum*. Additional trait data were obtained for *Nertera*

depressa root diameter from Johnson (1976). Roots that could not be identified to species

(e.g., *Coprosma* spp.) were excluded from root trait analysis due to high variability of traits

within genus (Holdaway *et al.* 2011).

Vegetation cover

The percent cover of all vascular plants within sites was measured following standard

protocols (Hurst & Allen 2007) in a 20 x 20 m area located within 50 m of the root sample

transects. For testing H1.2 and 2.2 we used the mid-point of vegetation cover classes summed

across height tiers as a measure of the abundance of each plant species. Vegetation data are

archived in the New Zealand National Vegetation Survey Databank

(https://nvs.landcareresearch.co.nz; last accessed 8 Feb 2016).

Statistics

Changes in the frequency of oomycetes as a function of site age were tested by

treating site age as a factor, and as a function of successional stage were tested by treating age

as a three-level factor. Sites were allocated to successional stage levels on the basis of soil

nutrient concentrations, vegetation height, species richness and biomass, and plant traits

(Walker & Syers 1976; Richardson *et al.* 2004). Early-successional sites (5, 15, 70 years) had

abundant N-fixing trees; vascular plant traits of high foliar nutrient concentrations, low leaf

mass per unit area; very high soil P, rapidly accumulating biomass and species richness of

vascular plants, and increasing nitrate-N. Mature ecosystems (280 to 12 000 years) had slow

rates of change in species richness, height and biomass among sites. These sites also had the

highest biomass and plant diversity. Retrogressive sites (60 000, 120 000 years) had a decline

in vegetation height, richness and biomass and exceptionally proficient phosphorus resorption

(Richardson *et al.* 2005; Peltzer *et al.* 2010; Vitousek *et al.* 2010).

The significance of changes in the frequency of oomycetes as a function of site age

(H1.1, 1.2 and 1.3) was tested using a chi-square test, treating site age as a factor. We tested

changes in the frequency of oomycetes as a function of successional stage using a binomial

glm, fitted with quasi-likelihood to account for over-dispersion. To test whether declining

plant species had higher than random oomycete levels (H1.2), we calculated the change in

cover of each plant species from one time point (t_0) to the next (t_1) as a log ratio (log((cover t_0

+1) / (cover t_1 +1))) and then used a binomial mixed effects model to test whether the

presence / absence of oomycetes in root fragments at t_0 could be predicted by change in cover

from t_0 to t_1 across the nine intervals between site ages. This test was carried out using lmer in

the R package lme4, with site age and plant species (to account for species occuring across

multiple sites) as random effects and the canonical logit link function. For trait correlations,

only observations with more than three root samples were included in order to have some

degree of confidence in oomycete frequency estimation. We similarly tested the correlation

between oomycete frequency and vegetation abundance as cover scores (H2.2) and root traits

(H3.2), again using mixed effects models and including random terms for plant species and

site. For all binary tests, over dispersion was tested using the Pearson residuals divided by

degrees of freedom, tested against a chi-square distribution (from

http://glmm.wikidot.com/start; accessed 10 July 2015). Host-specificity was tested within

ecosystem age using chi-square tests.

Results

Oomycete frequency as a function of stage of ecosystem development (H1.1, 2.1 and 3.1)

In early-successional ecosystems (5, 15 and 70 years), 38 to 65% of root samples had

oomycete DNA detected (Figure 1). In contrast, in mature and retrogressive ecosystems (280

to 120 000 years) an average of only 2% and never more than 8% of roots had oomycete

DNA found (Figure 1; χ^2 = 194, df = 9, P < 2.2 x 10^{-16}). Treating the sites as representing

three stages following our hypotheses, the early successional stages had a significantly higher

frequency of oomycetes than mature stages (quasibinomial family glm; t = -6.6, P = 0.00050)

or retrogressive stages (t = -4.6, P = 0.0025), and the difference between mature and

retrogressive ecosystems was not significant (t = -1.1, P = 0.32).

Oomycete frequency as a function of plant strategies and traits (H1.2, 2.2, 3.2)

The frequency of oomycetes was significantly correlated with the decline of vegetation cover

between ecosystem ages (Figure 2a, z = -2.4, P = 0.018), supporting hypothesis 1.2. There

was no correlation between oomycete frequency and plant abundance (z = 0.84, P = 0.40). Of

the root traits tested, root hair abundance was significantly correlated with oomycete

frequency (Figure 2b, z = 3.0, P = 0.003), providing partial support for hypothesis 3.2, while

other root traits were not significant (SRL, z = 0.013, P = 0.99; root diameter, z = 0.82, P =

0.42).

Oomycete diversity, host specificity and ecosystem age (H4)

A total of 37 different OTUs were found (Figure 3). Five OTUs were found more than 5

times. The most abundant OTU occurred 17 times, and had DNA-sequence affinities to

Lagena radicicola and an uncultured oomycete from New York, USA agricultural soils (17

occurrences). Other abundant OTUs included an OTU matching *Pythium dissotocum* and

Pythium aquatile (13 occurrences); *Pythium tracheiphilum* (6 occurrences); *Pythium volutum*

(6 occurrences); and *Phytophthora hedraiandra* (28S) / *cactorum* (ITS) (5 occurrences).

Within ecosystem age there was a significantly non-random distribution of oomycete OTU

identity across plant species at 5 (χ^2 = 42.96, df = 12, p-value = 0.000022) and 15 years (χ^2 =

225.87, df = 182, p-value = 0.015), but only marginally significant at 70 years (χ^2 = 109.85,

df = 90, p-value = 0.076; Figure 4). There were too few observations at the ecosystem stages

after 70 years to meaningfully test for plant by OTU host-specificity at later ecosystem

stages.

Discussion

Our results strongly support the importance of pathogens in early succession, with high frequencies observed in the three earliest ecosystem stages (5, 15, and 70 years) but oomycetes being present only at very low levels thereafter. Although the importance of pathogens is often discussed in the ecological literature (Bagchi *et al.* 2014), relatively few studies have actually quantified pathogen abundance in natural ecosystems (Gomez-Aparicio *et al.* 2012) and we believe our results are the first study of pathogenic organisms throughout a complete ecosystem development sequence. The community of oomycetes was diverse, including multiple *Pythium* and *Phytophthora* species and, surprisingly, a very abundant sequence matching *Lagena,* a pathogen previously only reported on grass, on *Raolia hookeri* (Asteraceae) and *Epilobium glabellum* (Onograceae).

Pathogens as drivers in early successional ecosystems

The early succession hypothesis postulates that pathogens drive early-successional change in plant communities (Kardol *et al.* 2006). The high frequency of oomycetes in early successional ecosystems strongly supports this hypothesis. In addition, those plant species that had the highest frequency of oomycetes at a given ecosystem stage declined in percent cover most strongly before the next stage. The higher frequency of oomycetes on plant roots in early-successional ecosystems compared to mature and retrogressive ecosystems was not due to PCR inhibition in older sites, as samples spiked with positive control oomycete DNA showed no evidence of inhibition.

Oomycetes, particularly *Pythium*, are often tolerated by established plants, but prevent seedling establishment (Martin & Loper 1999; van der Putten 2003). The correlation of oomycete frequency with decline in cover may not, therefore, reflect a direct negative effect

of the pathogens on established plants. Instead, we suggest that these pathogens are primarily

preventing re-establishment of plants and hence contributing to vegetation change across

cohorts (van der Putten 2003).

Our definition of "early succession" comprises the major transition from rock field (5

years) to shrub land (15 years) to forest (70 years). Therefore our finding of high pathogen

levels is potentially consistent with Packer & Clay (2000), who found strong negative

feedback in a shade-intolerant, seral species (*Prunus serotina*) in a forest described as "at

least 70 years old" (Packer & Clay 2000).

Changes in oomycete frequency may, in part, reflect a substantial site-age effect and

direct effects of changing soil environments on oomycete populations, independent of

changing plant communities (akin to the "habitat" hypothesis for mycorrhizal fungi of Zobel

& Opik (2014). Meaningful testing of soil variables as predictors of oomycetes was not

possible, as the three early-successional sites had both uniformly high oomycete frequency

and uniformly high (pH, total P) or low (total C) soil measures. Thus, while we can conclude

that oomycetes were particularly abundant in early-successional ecosystems and on early-

successional plants, our data are insufficient to determine whether this was due to the

presence of particular plant hosts, or to soil abiotic conditions driving both host plants and

oomycete frequency (Martin & Loper 1999).

Oomycetes in mature ecosystems and forest diversity

The mature ecosystem hypothesis suggests that pathogens are important drivers of

diversity in mature forest ecosystems, with evidence from tropical and temperate forests

(Bagchi *et al.* 2010; Bagchi *et al.* 2014), although other studies have failed to support a role

of pathogens in mature ecosystems (Reinhart *et al.* 2012). Our finding of only low levels of

oomycetes on roots in mature ecosystems suggest that any role of pathogens in maintaining

mature forest diversity is unlikely to be driven by oomycete populations on the roots of

established plants. Further, we found no significant correlation between the abundance of a

particular tree species and its pathogen abundance, whereas the mature ecosystems

hypothesis would predict either a positive (Queenborough *et al.* 2007) or negative correlation

(Xu *et al.* 2015) depending on interpretation.

Our findings do not rule out negative density dependent mechanisms linking pathogen

infection to forest diversity, but suggest an important restriction of potential mechanisms. In

the original formulation, negative density dependent theories distinguished between distance-

dependent and density-dependent agents of seed and seedling mortality (Janzen 1970).

Distance-dependent agents are primarily driven by the presence of the parent tree itself, while

density-dependent agents are attracted by the density of seeds and seedlings around that tree.

While it was previously known that established plant individuals are less susceptible to

oomycetes than seedlings, we had limited prior knowledge of whether roots in mature

ecosystems tolerated oomycetes, hence supporting populations that could cause distance-

dependent mortality, or resisted oomycete infection altogether (c.f., Spear *et al.* 2015). Our

results suggest that roots of established plants in mature ecosystems are not supporting high

oomycete populations, implying that distance-dependent oomycete infection is not an

important driver of diversity, at least in this temperate rainforest. This is consistent with a

number of studies showing positive, rather than negative, effects of established plants on

conspecific seedlings (Simard & Durall 2004). On the other hand, our results do not rule out

density-dependent mechanisms driven by oomycetes on seedlings. A high density of

conspecific seedlings may still lead to increased risk of oomycete driven mortality, as has

been found in other studies (Bagchi *et al.* 2010).

While Packer and Clay (2000) clearly demonstrated oomycetes driving seedling

dynamics in a relatively young temperate forest, a more recent study by Bagchi and

colleagues (2014) in a tropical seasonal forest found that eliminating fungi alone had a

stronger effect on negative density-dependence than eliminating both fungi and oomycetes.

Further research is needed, but this may suggest that fungal pathogens are more important

drivers of forest diversity in mature ecosystems than oomycetes.

Oomycetes in retrogression

Laliberté and colleagues (2015) suggested that root traits adapted to highly P-limited

retrogressive ecosystems would increase susceptibility to root pathogens. Of the root traits we

tested, only root hair abundance was correlated with a high frequency of oomycetes in roots.

While this supports the general concept that root traits can determine pathogen susceptibility,

species with root hair abundance shows no increase with ecosystem age at this site

(Holdaway et al. 2011) and hence this correlation did not drive an increase in oomycete

frequency during retrogression. Studies of oomycete populations in other retrogressive

ecosystems will be needed to confirm our findings, but at present the suggestion of Laliberté

and colleagues (2015) of increasing pathogen loads in retrogression is not supported by our

findings, despite part of the mechanistic basis of that hypothesis (root traits determining

pathogen loads) was supported at least in the case of root hairs.

In comparing our results to other studies, it may be important to distinguish between

ecosystem age, as in the pedogenic sequence studied here, and response of pathogens to

disturbance or secondary succession (e.g., Reinhart *et al.* 2010). None of our older sites

showed any sign of stand-replacing disturbance in the past hundreds to thousand years, but

smaller-scale disturbance is a normal part of ecosystem development (Peltzer *et al.* 2010).

The retrogressive sites, in particular, have much younger vegetation (the oldest trees being no

more than 1000 years old) than underlying soils, and were unlikely to have supported tall

forests during the Pleistocene when these sites were exposed by glacial retreat.

Novel organisms or native?

Having baseline data on oomycete ecology may be important for understanding and

managing future pathogen outbreaks. For example, *Phytophthora agathidicida* (=

Phytophthora taxon *Agathis*) has been implicated as a cause of forest dieback of the iconic

New Zealand-endemic *Agathis australis*, but, like other oomycete disease outbreaks, the

origin of the species remains uncertain (Weir *et al.* 2015). The single-direction sequencing

we performed directly from environmental samples was not designed for accurate

phylogenetic classification, even with two gene regions sequenced per sample, but it does

suggest considerable diversity of *Phytophthora*. We found sequences with affinities to a

number of species already known to be present in New Zealand, including *P. cinnamomi*, *P.*

cactorum, *P. infestans* and *P. kernovieae* (Scott & Williams 2014). Given that our samples

came from relatively pristine ecosystems, we believe these OTUs are likely to be from native

Phytophthora.

More surprising was that the most common OTU in our data had affinities to *Lagena*

radicicola in both 28S and many of the ITS sequence results. *Lagena* is a monotypic genus

widespread as a pathogen of grasses in North America (Blackwell 2011), but Barr and

Désaulniers (1990) suggest it may be under-reported due to being morphologically very

similar to *Pythium*, and not being easily culturable. Spores described as "resembling

Lagenocystis [syn. *Lagena*] spp." were noted by Skipp and Christensen (1989) in New

Zealand *Lolium perenne* pastures, but our finding is the first report of a putative *Lagena*

species outside of grasses. We believe this most likely reflects a lack of prior knowledge, as

very few prior studies have used molecular methods to detect oomycetes from healthy forest

ecosystems, and none of those have taken place in temperate southern hemisphere rainforest.

Our sampling of washed plant roots and choice of primers that preferentially amplify

the Peronosporales focused our analysis on putative plant pathogens within the Oomycota

occurring in soils. This is reflected in the affinity of most of our sequences to known plant

pathogens. Nonetheless, we note that the Oomycota also includes animal pathogens and free-

living saprotrophs as well as foliar pathogens; potential pools of species that were not likely

to be included in our sampling.

Conclusions

Much of the Earth's surface is covered in young, early successional ecosystems

comparable to our early successional sites of 5, 15, and 70 years (Vitousek *et al.* 1997;

Haddad *et al.* 2015). Our results support the concept that oomycete pathogens are a diverse

part of these ecosystems. Further, the strong plant-host specificity and correlation of plant

decline and oomycete frequency is consistent with the suggestion that these pathogens

contribute to vegetation succession. While we do not find abundant oomycetes in older

ecosystems, oomycete DNA was detected in all but one site. These low levels of oomycetes

in roots, along with oospores in soil, may still be sufficient to prevent establishment of

susceptible plant species or to drive pathogen outbreaks following external climatic stressors,

but seem incompatible with density dependent mortality being driven by oomycetes on

established plant roots.

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Figure 1. Frequency of oomycetes on plant roots (%) as a function of ecosystem age. The 5,

15, and 70-year-old sites represent a successional sere from rock field through shrubland to

angiosperm forest. Sites 280 to 5000 years of age are mature forest with high biomass and

plant diversity, while after 12 000 to 60 000 years plant biomass, stature, and diversity

decline due to retrogressive P-limitation.

Figure 2. The two significant predictors of oomycete frequency on roots: vegetation cover (a)

and the abundance of root hairs (b), with circles representing each plant species between site

transitions (a) or within site (b). Lines indicate random effects to account for site in a binary

mixed effects model with random intercepts for site and plant species. Color coding in (a)

reflects the site age at t_0 for each transition, hence no points for 120 000 years are shown.

Figure 3. Mapping of DNA sequences into OTUs and against known species. Phylogenetic

clustering (at left) based on 28S gene of samples and a selection of named sequences from

GenBank, using Clustal Omega 1.2.1 for alignment and phylogeny based on UPGMA

(Unweighted Pair Group Method with Arithmetic Mean) clustering. OTUs were based on

pairwise BLASTn of 28S sequences at 97% similarity; resulting clusters shown as colored

squares, with singletons shown as black squares. Bold indicates representative sequence from

each non-singleton OTU accessioned into GenBank (numbers KU863588-KU863606). For

each collection, the best matching 28S sequence name in GenBank is shown, followed by the

sequence similarity (%) / percent of query matched and the actual length of the query

sequence (in bp). ITS1 sequences were obtained from most samples, best match in GenBank

shown, again with % identity over % of query and query length. Where the ITS1 sequence is

likely to be a different species sequenced from the same sample, this is indicated by

parentheses.

Figure 4. Plant by site age visualisation of oomycete frequency within sampled roots. Each

square represents a single root sample, with coloured squares indicating the presence of

oomycete pathogens with black indicating an OTU found only once and other colours

indicating the OTU groupings identified in figure 3. Grey dots indicate species not found in

root samples for that site.

Supplemental files S1: Detailed molecular methods and discussion.

References

Arcate, J.M., Karp, M.A. & Nelson, E.B. (2006) Diversity of peronosporomycete (oomycete)

communities associated with the rhizosphere of different plant species. *Microb Ecol*, **51**, 36-

50.

Augspurger, C.K. (1984) Seedling survival of tropical tree species: interactions of dispersal

distance, light-gaps, and pathogens. *Ecology*, **65**, 1705-1712.

Bagchi, R., Gallery, R.E., Gripenberg, S., Gurr, S.J., Narayan, L., Addis, C.E. *et al.* (2014)

Pathogens and insect herbivores drive rainforest plant diversity and composition. *Nature*,

506, 85-88.

Bagchi, R., Swinfield, T., Gallery, R.E., Lewis, O.T., Gripenberg, S., Narayan, L. *et al.*

(2010) Testing the Janzen-Connell mechanism: pathogens cause overcompensating density

dependence in a tropical tree. *Ecol Lett*, **13**, 1262-1269.

Bardgett, R.D., Bowman, W.D., Kaufmann, R. & Schmidt, S.K. (2005) A temporal approach

to linking aboveground and belowground ecology. *Trends in Ecology & Evolution*, **20**, 634-

641.

Barr, D.J.S. & Désaulniers, N.L. (1990) The life cycle of *Lagena radicicola*, an oomycetous

parasite of wheat roots. *Canadian journal of botany*, **68**, 813-824.

Bever, J.D., Mangan, S.A. & Alexander, H.M. (2015) Maintenance of Plant Species Diversity

by Pathogens. *Annu. Rev. Ecol. Evol. Syst.*, **46**, 305-325.

Bever, J.D., Dickie, I.A., Facelli, E., Facelli, J.M., Klironomos, J., Moora, M. *et al.* (2010)

Rooting theories of plant community ecology in microbial interactions. *Trends Ecol Evol*, **25**,

468-478.

Blackwell, W.H. (2011) The genus Lagena (Stramenopila: Oomycota), taxonomic history and

nomenclature. *Phytologia*, **93**, 157-167.

Castello, J.D., Leopold, D.J. & Smallidge, P.J. (1995) Pathogens, patterns, and processes in

forest ecosystems. *Bioscience*, 16-24.

Cavalier-Smith, T. (2010) Kingdoms Protozoa and Chromista and the eozoan root of the

eukaryotic tree. *Biol Lett*, **6**, 342-345.

Connell, J.H. (1970) Diversity in tropical rain forests and coral reefs. *Science*, **199**, 1302-

1309.

de Deyn, G.B., Raaijmakers, C.E., Zoomer, H.R., Berg, M.P., de Ruiter, P.C., Verhoef, H.A.

et al. (2003) Soil invertebrate fauna enhances grassland succession and diversity. *NATURE*,

422, 711-713.

Gilbert, G.S. (2002) Evolutionary ecology of plant diseases in natural ecosystems. *Annu Rev*

Phytopathol, **40**, 13-43.

Gomez-Aparicio, L., Ibanez, B., Serrano, M.S., De Vita, P., Avila, J.M., Perez-Ramos, I.M.

et al. (2012) Spatial patterns of soil pathogens in declining Mediterranean forests:

implications for tree species regeneration. *New Phytol*, **194**, 1014-1024.

Haddad, N.M., Brudvig, L.A., Clobert, J., Davies, K.F., Gonzalez, A., Holt, R.D. *et al.* (2015)

Habitat fragmentation and its lasting impact on Earth's ecosystems. *Science Advances*, **1**,

e1500052.

Holdaway, R.J., Richardson, S.J., Dickie, I.A., Peltzer, D.A. & Coomes, D.A. (2011)

Species- and community-level patterns in fine root traits along a 120 000-year soil

chronosequence in temperate rain forest. *Journal of Ecology*, **99**, 954-963.

Hurst, J.M. & Allen, R.B. (2007) *The recce method for describing New Zealad vegetation:*

field protocols. Manaaki Whenua - Landcare Research, Lincoln, N. Z.

Janzen, D.H. (1970) Herbivores and the number of tree species in tropical forests. *The*

American Naturalist, **104**, 501-528.

Johnson, P.N. (1976) Effects of soil phosphate level and shade on plant growth and

mycorrhizas. *New Zealand Journal of Botany*, **14**, 333-340.

Kardol, P., Bezemer, T.M. & van der Putten, W.H. (2006) Temporal variation in plant-soil

feedback controls succession. *Ecol Lett*, **9**, 1080-1088.

Laliberté, E., Lambers, H., Burgess, T.I. & Wright, S.J. (2015) Phosphorus limitation, soil-

borne pathogens and the coexistence of plant species in hyperdiverse forests and shrublands.

New Phytol, **206**, 507-521.

Lambers, H., Raven, J.A., Shaver, G.R. & Smith, S.E. (2008) Plant nutrient-acquisition

strategies change with soil age. *Trends Ecol Evol*, **23**, 95-103.

Mangan, S.A., Herre, E.A. & Bever, J.D. (2010a) Specificity between Neotropical tree

seedlings and their fungal mutualists leads to plant-soil feedback. *Ecology*, **91**, 2594-2603.

Mangan, S.A., Schnitzer, S.A., Herre, E.A., Mack, K.M.L., Valencia, M.C., Sanchez, E.I. *et*

al. (2010b) Negative plant–soil feedback predicts tree-species relative abundance in a tropical

forest. *Nature*, 752-755.

Martin, F.N. & Loper, J.E. (1999) Soilborne Plant Diseases Caused by *Pythium* spp.:

Ecology, Epidemiology, and Prospects for Biological Control. *Critical Reviews in Plant*

Sciences, **18**, 111-181.

Martinez-Garcia, L.B., Richardson, S.J., Tylianakis, J.M., Peltzer, D.A. & Dickie, I.A. (2015)

Host identity is a dominant driver of mycorrhizal fungal community composition during

ecosystem development. *New Phytol*, **205**, 1565-1576.

Packer, A. & Clay, K. (2000) Soil pathogens and spatial patterns of seedling mortality in a

temperate tree. *Nature*, **404**, 278-281.

Packer, A. & Clay, K. (2003) Soil pathogens and *Prunus serotina* seedling and sapling

growth near conspecific trees. *Ecology*, **84**, 108-119.

Peltzer, D.A., Wardle, D.A., Allison, V.J., Baisden, W.T., Bardgett, R.D., Chadwick, O.A. *et*

al. (2010) Understanding ecosystem retrogression. *Ecological Monographs*, **80**, 509-529.

Podger, F.D. & Newhook, F.J. (1971) *Phytophthora cinnamomi* in indigenous plant

communities in New Zealand. *New Zealand Journal of Botany*, **9**, 625-638.

Porder, S., Vitousek, P.M., Chadwick, O.A., Chamberlain, C.P. & Hilley, G.E. (2007) Uplift,

Erosion, and Phosphorus Limitation in Terrestrial Ecosystems. *Ecosystems*, **10**, 159-171.

Queenborough, S.A., Burslem, D.F.R.P., Garwood, N.C. & Valencia, R. (2007)

Neighborhood and community interactions determine the spatial pattern of tropical tree

seedling survival. *Ecology*, **88**, 2248-2258.

Reinhart, K.O., Johnson, D. & Clay, K. (2012) Conspecific plant-soil feedbacks of temperate

tree species in the southern Appalachians, USA. *PLoS One*, **7**, e40680.

Reinhart, K.O., Royo, A.A., Kageyama, S.A. & Clay, K. (2010) Canopy gaps decrease

microbial densities and disease risk for a shade-intolerant tree species. *Oecologiia*, **36**, 530-

536.

Reynolds, H.L., Packer, A., Bever, J.D. & Clay, K. (2003) Grassroots ecology: Plant-

microbe-soil interactions as drivers of plant community structure and dynamics. *ECOLOGY*,

84, 2281-2291.

Richardson, S.J., Peltzer, D.A., Allen, R.B. & McGlone, M.S. (2005) Resorption proficiency

along a chronosequence: responses among communities and within species. *Ecology*, **86**, 20-

25.

Richardson, S.J., Peltzer, D.A., Allen, R.B., McGlone, M.S. & Parfitt, R.L. (2004) Rapid

development of phosphorus limitation in temperate rainforest along the Franz Josef soil

chronosequence. *Oecologia*, **139**, 267-276.

Scott, P. & Williams, N. (2014) *Phytophthora* diseases in New Zealand forests. *NZ Journal*

of Forestry, **59**, 15.

Simard, S.W. & Durall, D.M. (2004) Mycorrhizal networks: a review of their extent,

function, and importance. *Canadian Journal of Botany*, **82**, 1140-1165.

Skipp, R.A. & Christensen, M.J. (1989) Fungi invading roots of perennial ryegrass (L.) in

pasture. *New Zealand Journal of Agricultural Research*, **32**, 423-431.

Spear, E.R., Coley, P.D. & Kursar, T.A. (2015) Do pathogens limit the distributions of

tropical trees across a rainfall gradient? *J Ecol*, **103**, 165-174.

Van der Putten, W.H., Dijk, C.V. & Peters, B.A.M. (1993) Plant-specific soil-borne diseases

contribute to succession in foredune vegetation. *Nature*, **362**, 53-56.

van der Putten, W.H. (2003) Plant defense belowground and spatiotemporal processes in

natural vegetation. *Ecology*, **84**, 2269-2280.

Vitousek, P.M., Porder, S., Houlton, B.Z. & Chadwick, O.A. (2010) Terrestrial phosphorus

limitation: mechanisms, implications, and nitrogen-phosphorus interactions. *Ecological*

Applications, **20**, 5-15.

Vitousek, P.M., Mooney, H.A., Lubchenco, J. & Melillo, J.M. (1997) Human domination of

Earth's ecosystems. *Science*, **277**, 494-499.

Walker, T.W. & Syers, J.K. (1976) The fate of phosphorus during pedogenesis. *Geoderma*,

15, 1-19.

Weir, B.S., Paderes, E.P., Anand, N., Uchida, J.Y., Pennycook, S.R., Bellgard, S.E. *et al.*

(2015) A taxonomic revision of *Phytophthora* Clade 5 including two new species,

Phytophthora agathidicida and *P. cocois*. *Phytotaxa*, **205**, 21.

Xu, M., Wang, Y. & Yu, S. (2015) Conspecific negative density dependence decreases with

increasing species abundance. *Ecosphere*, **6**, art257.

Zobel, M. & Opik, M. (2014) Plant and arbuscular mycorrhizal fungal (AMF) communities -

which drives which? *J Veg Sci*, **25**, 1133-1140.

Phylogenetic tree figure. Column headers:

OTU | **Best Match 28S** | **Best Match ITS**

Tree tip labels (top to bottom):

KJ150290.1 Spongospora sp.
DA13581
DA13562
DA13723
DQ007471.1 Albugo candida
DA13796
DA13866
JX436350.1 Apodachlya brachynema
AF119590.1 Apodachlya brachynema
AF235937.1 Apodachlya minima
AF119597.1 Leptomitus lacteus
DA13734
DA13775
DA13757
DA13862
DA13794
DA13893
DA13798
DA13799
DA13872
DA13748
DA13786
DA13834
DA13772
DA13781
AY598692.2 Pythium grandisporangium
AB468704.1 Pythium sp.
DA13882
HM441674.1 Uncultured oomycete
KJ716869.1 Lagena radicicola
DA13774
DA13891
DA13879
DA13877
DA13894
DA13831
DA13812
DA13830
DA13819
DA13829
DA13835
DA13885
DA13880
HQ395652.1 Lagenidium sp.
DA13761
DA13758
DA13795
DA13864
DA13782
DA13715
DA13718
AY598627.2 Pythium aristosporum
AY598686.2 Pythium volutum
AY598681.2 Pythium dissimile
DA13717
DA13713
DA13762
DA13764
DA13763
HQ665207.1 Pythium tracheiphilum
DA13803
DA13865
DA13793
DA13719
DA13739
DA13806
DA13595
DA13721
DA13800
AY598634.2 Pythium dissotocum
AB468701.1 Pythium aquatile
DA13711
DA13873
DA13760
DA13801
DA13809
DA13871
DA13789
DA13805
DA13745
DA13702
DA13776
HQ643697.2 Pythium monospermum
DA13766
DA13753
HQ395673.1 Pythium adhaerens
DA13792
DA13867
HQ395660.1 Lagenidium sp.
HQ643956.2 Pythium viniferum
AY598717.2 Pythium violae
DA13770
DA13773
DA13720
AY598648.2 Pythium iwayamai
DA13826
DA13889
DA13836
DA13820
DA13817
DA13810
DA13755
DA13376
DA13526
DA13394
AB688459.1 Phytophthora cinnamomi
ICMP 17027 Pythophthora agathidicida
AB688479.1 Phytophthora cryptogea
EU199118.1 Phytophthora inundata
HQ665148.1 Phytophthora humicola
HQ665201.1 Phytophthora megasperma
DA13863
DA13722
AB700469.1 Phytophthora hedraiandra
DA13714
DA13783
DA13777
DA13641
GU594862.1 Phytophthora polonica
DA13744
DA13765
DA13396

Best Match 28S column:

Spongospora sp. 76%/78% 299 bp
Spongospora sp. 76%/79% 297 bp
Spongospora sp. 75%/79% 306 bp
Albugo candida 99%/98% 382 bp
Pythium viniferum 81%/97% 378 bp
Uncultured oomycete 83%/98% 243 bp
Uncultured oomycete 89%/71% 342 bp
Sapromyces elongatus 77%/98% 381 bp
Uncultured oomycete 87%/91% 156 bp
Pythium dissotocum 80%/92% 380 bp
Uncultured rhizosphere 85%/98% 200 bp
Pythium monospermum 88%/98% 338 bp
Pythium tracheiphilum 84%/99% 373 bp
Pythium tracheiphilum 86%/97% 376 bp
Pythium tracheiphilum 94%/93% 180 bp
Uncultured oomycete 91%/100% 206 bp
Pythium dissimile 91%/95% 143 bp
Pythium viniferum 98%/95% 97 bp
Uncultured oomycete 98%/100% 369 bp
Uncultured oomycete 98%/99% 162 bp
Uncultured oomycete 99%/98% 210 bp
Uncultured oomycete 97%/94% 181 bp
Uncultured oomycete 98%/98% 259 bp
Uncultured oomycete 100%/97% 150 bp
Uncultured oomycete 99%/98% 208 bp
Uncultured oomycete 99%/96% 121 bp
Uncultured oomycete 98%/98% 407 bp
Uncultured oomycete 99%/98% 298 bp
Uncultured oomycete 99%/100% 368 bp
Uncultured oomycete 99%/98% 241 bp
Uncultured oomycete 97%/97% 194 bp
Uncultured oomycete 98%/99% 374 bp
Uncultured oomycete 98%/100% 369 bp
Lagena radicicola 97%/80% 136 bp
Uncultured oomycete 91%/100% 197 bp
Uncultured oomycete 93%/98% 363 bp
Pythium aquatile 92%/98% 372 bp
Uncultured oomycete 97%/98% 210 bp
Uncultured rhizosphere 97%/100% 221 bp
Uncultured oomycete 96%/98% 373 bp
Uncultured oomycete 99%/98% 369 bp
Uncultured oomycete 98%/100% 380 bp
Pythium tracheiphilum 99%/98% 372 bp
Pythium tracheiphilum 99%/100% 374 bp
Pythium tracheiphilum 99%/98% 373 bp
Pythium tracheiphilum 95%/100% 385 bp
Pythium tracheiphilum 94%/100% 366 bp
Pythium tracheiphilum 98%/99% 370 bp
Uncultured rhizosphere 95%/96% 249 bp
Uncultured oomycete 99%/97% 197 bp
Pythium dissotocum 94%/81% 367 bp
Pythium aquatile 99%/98% 367 bp
Pythium aquatile 99%/100% 374 bp
Pythium dissotocum 99%/98% 373 bp
Pythium aquatile 99%/98% 383 bp
Pythium dissotocum 99%/100% 374 bp
Pythium dissotocum 99%/100% 368 bp
Pythium dissotocum 99%/98% 370 bp
Pythium dissotocum 99%/98% 376 bp
Pythium aquatile 99%/98% 375 bp
Pythium dissotocum 98%/98% 370 bp
Pythium dissotocum 97%/100% 379 bp
Uncultured oomycete 95%/100% 206 bp
Uncultured oomycete 93%/98% 254 bp
Uncultured oomycete 98%/97% 162 bp
Uncultured oomycete 98%/98% 377 bp
Uncultured oomycete 98%/98% 238 bp
Uncultured oomycete 98%/98% 378 bp
Uncultured oomycete 94%/98% 374 bp
Pythium violae 98%/100% 368 bp
Pythium violae 94%/95% 239 bp
Pythium violae 89%/100% 383 bp
Pythium iwayamai 98%/100% 261 bp
Uncultured oomycete 98%/95% 307 bp
Uncultured oomycete 97%/97% 197 bp
Phytophthora megasperma 93%/100% 381 bp
Phytophthora megasperma 91%/100% 383 bp
Phytophthora cryptogea 94%/100% 384 bp
Phytophthora cryptogea 91%/100% 382 bp
Phytophthora cinnamomi 99%/98% 254 bp
Phytophthora cinnamomi 100%/99% 383 bp
Phytophthora cinnamomi 99%/100% 385 bp
Phytophthora hedraiandra 99%/98% 385 bp
Phytophthora hedraiandra 99%/98% 384 bp
Phytophthora hedraiandra 98%/98% 392 bp
Phytophthora hedraiandra 99%/97% 218 bp
Phytophthora hedraiandra 98%/98% 359 bp
Peronospora grisea 99%/98% 381 bp
Phytophthora polonica 99%/99% 384 bp
Phytophthora polonica 98%/98% 388 bp
Phytophthora polonica 99%/100% 129 bp

Best Match ITS column:

(Pythium intermedium 94%/49% 304 bp)*
(Phytophthora lateralis 89%/99% 222 bp)*
(Phytophthora infestans 95%/27% 199 bp)*
(Phytophthora uliginosa 99%/99% 257 bp)*
(Pythium sp. 81%/42% 310 bp)*
Pythium intermedium 94%/38% 298 bp
Pythium intermedium 86%/99% 305 bp
Pythium rostratum 76%/96% 311 bp
Pythium intermedium 84%/34% 304 bp
Pythium emineosum 96%/33% 299 bp
Phytophthora lateralis 88%/100% 305 bp
Pythium sp. 98%/99% 286 bp
Pythium sp. 88%/97% 292 bp
Pythium sp. 84%/99% 282 bp
Pythium sp. 99%/100% 287 bp
Pythium helicandrum 76%/53% 232 bp
Lagena radicicola 96%/99% 205 bp
Lagena radicicola 92%/99% 209 bp
(Phytophthora infestans 94%/54% 444 bp)*
Lagena radicicola 95%/97% 202 bp
Lagena radicicola 95%/99% 202 bp
Lagena radicicola 90%/71% 447 bp
Lagena radicicola 93%/99% 205 bp
Lagena radicicola 93%/72% 445 bp
Lagena radicicola 80%/69% 201 bp
Lagena radicicola 96%/99% 205 bp
Lagena radicicola 96%/100% 206 bp
Lagena radicicola 95%/99% 205 bp
Lagena radicicola 96%/99% 207 bp
(Pythium pachycaule 92%/94% 310 bp)*
(Phytophthora lateralis 89%/96% 228 bp)*
(Pythium intermedium 96%/99% 309 bp)*
(Phytophthora lateralis 89%/99% 308 bp)*
Pythium sp. 97%/97% 291 bp
Pythium sp. 90%/99% 202 bp
Pythium sp. 85%/97% 202 bp
Pythium sp. 91%/99% 200 bp
Pythium sp. 92%/100% 203 bp
Pythium intermedium 85%/100% 302 bp
Pythium tracheiphilum 91%/96% 306 bp
Pythium intermedium 98%/34% 300 bp
Pythium tracheiphilum 91%/57% 305 bp
Pythium intermedium 97%/42% 289 bp
Pythium tracheiphilum 89%/98% 209 bp
Pythium sp. 91%/100% 201 bp
Pythium sp. 98%/95% 307 bp
(Phytophthora alni 79%/44% 276 bp)*
(Phytophthora lateralis 89%/99% 309 bp)*
(Phytophthora lateralis 87%/99% 216 bp)*
Pythium sp. 98%/95% 308 bp
(Phytophthora lateralis 89%/99% 222 bp)*
Pythium emineosum 97%/34% 298 bp
Pythium irregulare 94%/35% 308 bp
Pythium irregulare 98%/45% 300 bp
Pythium intermedium 90%/47% 302 bp
Pythium heterothallicum 86%/80% 210 bp
Pythium sp. 100%/20% 257 bp
Pythium attrantheridium 93%/38% 308 bp
Pythium intermedium 96%/99% 310 bp
(Saprolegnia litoralis 81%/30% 231 bp)*
Pythium irregulare 100%/28% 307 bp
Phytophthora lateralis 87%/99% 223 bp
Pythium sp. 99%/99% 286 bp
Pythium sp. 98%/99% 286 bp
Pythium sp. 85%/100% 283 bp
Pythium canariense 88%/67% 277 bp
Phytophthora infestans 93%/54% 444 bp
Phytophthora infestans 96%/53% 449 bp
Phytophthora infestans 96%/53% 446 bp
Phytophthora infestans 95%/53% 447 bp
Phytophthora lateralis 89%/99% 221 bp
Phytophthora lateralis 89%/99% 220 bp
Phytophthora cinnamomi 99%/100% 254 bp
Phytophthora cinnamomi 99%/100% 254 bp
Phytophthora cinnamomi 99%/100% 254 bp
Phytophthora cactorum 98%/99% 230 bp
Phytophthora cactorum 100%/99% 233 bp
Phytophthora cactorum 99%/99% 238 bp
Phytophthora cactorum 99%/99% 245 bp
Phytophthora cactorum 98%/70% 340 bp
Peronospora agrestis 91%/95% 222 bp
Phytophthora kernoviae 96%/99% 214 bp
(Pythium intermedium 91%/99% 301 bp)*

*Probable second species from 28S

Figure showing a matrix of oomycete associations across plant species (rows) and ecosystem age (columns).

Ecosystem age (columns): 5, 15, 70, 280, 500, 1000, 5000, 12000, 60000, 120000

Plant species (rows): *Raoulia hookeri*, *Epilobium glabellum*, *Chionochloa conspicua*, *Poa spp.*, *Epilobium pernitens*, *Parahebe lyallii*, *Aristotelia serrata*, *Coriaria arborea*, *Nertera depressa*, *Acaena sp.*, *Melicytus ramiflorus*, *Pseudopanax crassifolius*, *Festuca arundinacea*, *Olearia arborescens*, *Ranunculus repens*, *Uncinia sp.*, *Schefflera digitata*, *Astelia fragrans*, *Blechnum novae zelandiae*, *Holcus lanatus*, *Leycesteria formosa*, *Olearia avicenniifolia*, *Rytidosperma species*, *Weinmannia racemosa*, *Cyathea smithii*, *Raukaua simplex*, *Asplenium gracillimum*, *Neomyrtus pedunculata*, *Metrosideros umbellata*, *Dicksonia squarrosa*, *Rubus sp.*, *Coprosma foetidissima*, *Coprosma spp.*, *Pseudopanax colensoi*, *Hedycarya arborea*, *Carpodetus serratus*, *Ripogonum scandens*, *Luzula species*, *Dacrydium cupressinum*, *Prumnopitys ferruginea*, *Leptopteris superba*, *Podocarpus hallii*, *Griselinia littoralis*, *Metrosideros diffusa*, *Quintinia acutifolia*, *Hymenophyllum ferrugineum*, *Blechnum discolor*, *Elaeocarpus dentatus*, *Leptospermum scoparium*, *Gleichenia dicarpa*, *Lagarostrobos colensoi*, *Lepidothamnus intermedius*, *Phyllocladus alpinus*

Legend:
- No root sampled
- Root without oomycete
- Lagena sp.
- Pythium aff. aquatile/dissotocum
- Phytophthora aff. cinnamomi / cactorum
- Pythium sp.
- Pythium aff. tracheiphilum
- Pythium aff. adhaerens / monospermum
- Phytophthora aff. megasperma / infestans
- Phytophthora aff. polonica / kernoviae
- Chromista aff. Spongospora sp.
- Pythium violae
- Singleton

Detailed molecular methods and discussion

Molecular methods

We primarily based our oomycete detection, operational taxonomic unit clustering, and identification on a nested PCR of large-subunit (28S) DNA. The primer pair LR0R: ACCCGCTGAACTTAAGC (LoBuglio *et al.* 1991) and LR5: TCCTGAGGGAAACTTCG (Vilgalys & Hester 1990) was used for first-round PCR amplification, followed by using 1μL of a 10-fold dilution from this PCR product as the template for the second-round PCR amplification with Oom1F: GTGCGAGACCGATAGCGAACA and Oom1R: TCAAAGTCCCGAACAGCAACAA (Arcate *et al.* 2006) primers, fluorochrome-labelled with VIC and 6FAM respectively. PCR reactions for LR0R-LR5 were carried out in a 20 μl volume containing 2 μl 10X PCR Buffer, 1.5 mM $MgCl_2$, 0.2 mM dNTPs, 0.2 μM of each primer, 0.2 mg ml^{-1} BSA, 0.8 U Roche Faststart Taq DNA Polymerase, and 1 μl DNA extract; with thermocycling of 94°C for 1 min, then 35 cycles at 94 °C for 30s, 47 °C for 30s and 72 °C for 60s, final extension at 72 °C for 7 min. PCR reactions for the Oom1F-Oom1R primer pair were the same, with the exception that $MgCl_2$ was 2.5 mM, annealing temperature of 57 °C, and extension time was 45 s with final extension of 10 min. A negative (omitting DNA template) and a positive control (including genomic DNA from a *Pythium* species) were included in every PCR. Our intention in using fluorochrome-labelled primers was to provide the option of performing tRFLP analysis. However, initial results suggested that most samples contained a single tRFLP type. We therefore moved to a direct-sequencing technique. PCR products were purified using DNA Clean and Concentrator Kit (Zymo Research Corporation) prior to performing Sanger sequencing (Canterbury Sequencing and Genotyping, University of Canterbury, New Zealand).

To confirm identities of sequenced oomycetes, samples that produced positive large subunit

PCR products were sequenced for the ITS region. The oomycete specific primer pair

OOMUP18Sc: TGCGGAAGGATCATTACCACAC (Lievens *et al.* 2004)

and ITS2-OOM: GCAGCGTTCTTCATCGATGT (Lievens *et al.* 2006)

was used to amplify the ITS1 gene region (~ 240 bp). PCR reactions were carried out in a 25

µl volume containing 2.5 µl 10X Ex Taq Buffer, 2 µl Ex Taq dNTP Mixture (2.5 mM each),

0.5 µl of 10 µM of each primer, 5 µl of 10 mg ml-1 RSA, 1 U TaKaRa Ex Taq DNA

Polymerase, and 1 µl DNA extract, and sterilized distilled water up to 25 µl; with

thermocycling of 94 °C for 2 min, then 35 cycles at 94 °C for 30s, 60 °C for 30s and 72 °C

for 60s, final extension at 72 °C for 7 min.

Additional root trait data

Two species were present on the 5 year site that had no previously measured root traits:

Epilobium glabellum and *Raoulia hookeri*. Root traits were measured using the same

methods as for the earlier data set (Holdaway *et al.* 2011). Diameters = 0.22 ± 0.015 mm and

0.28 ± 0.018 mm; specific root lengths = 1437 ± 147 and 2354 ± 387 cm g^{-1}; and root hairs =

9.3 ± 4.8 and 106 ± 3.4 hairs cm^{-1} for *Epilobium glabellum* and *Raoulia hookeri,*

respectively.

Supplemental discussion of methods

Using direct Sanger sequencing from environmental samples results in somewhat short and

lower quality sequences than might be achievable with other methods, such as clone libraries.

Nonetheless, direct sequencing is a cost-effective and efficient way of detecting and

identifying the oomycetes in samples with a single dominant DNA sequence. This greater cost-effecitveness allows greater investment in replication; a critical factor in ecological studies of cryptic organisms (Prosser 2010).

Most samples that produced a positive PCR product in large subunit sequencing also produced a positive PCR product with ITS primers. In most cases, the two sequences appear to represent the same species (Figure 3). The exceptions tended to be samples at the base of the phylogeny, matching *Spongospora, Albugo*, and *Sapromyces* in 28S and either failing to produce ITS PCR products at all, or producing ITS sequences that were distantly related. This suggests that the ITS primers were more specific to a subset of the Oomycota, and may omit some basal groups; while the 28S primers better amplify basal Oomycota but potentially amplify other Chromalveolata as well. Neither gene region was sufficient to identify all OTUs to species, largely reflecting the relative paucity of oomycota sequences in GenBank compared to better studied groups, such as fungi. Large subunit sequences also lacked resolution within *Phytophythora* compared to ITS sequences.

References

Arcate, J.M., Karp, M.A. & Nelson, E.B. (2006) Diversity of peronosporomycete (oomycete) communities associated with the rhizosphere of different plant species. *Microb Ecol*, **51**, 36-50.

Holdaway, R.J., Richardson, S.J., Dickie, I.A., Peltzer, D.A. & Coomes, D.A. (2011) Species- and community-level patterns in fine root traits along a 120 000-year soil chronosequence in temperate rain forest. *Journal of Ecology*, **99**, 954-963.

Lievens, B., Claes, L. & Vanachter…, A.C.R.C. (2006) Detecting single nucleotide polymorphisms using DNA arrays for plant pathogen diagnosis. *FEMS …*,

Lievens, B., Hanssen, I.R.M. & Vanachter…, A. (2004) Root and foot rot on tomato caused by Phytophthora infestans detected in Belgium. *Plant …*,

LoBuglio, K.F., Rogers, S.O. & Wang, C.J.K. (1991) Variation in ribosomal DNA among isolates of the mycorrhizal fungus *Cenococcum geophilum. Canadian Journal of Botany*, **69**, 2331-2343.

Prosser, J.I. (2010) Replicate or lie. *Environmental microbiology*, **12**, 1806-1810.

Vilgalys, R. & Hester, M. (1990) Rapid genetic identification and mapping of enzymatically amplified ribosomal DNA from several Cryptococcus species. *Journal of bacteriology*, **172**, 4238-4246.