Type of article: Letter

Symmetric assembly and disassembly processes in an ecological

network

Jason M. Tylianakis[1,2,3], Laura B. Martínez-García[4,5], Sarah J. Richardson[4], Duane A.

Peltzer[4], Ian A. Dickie[3]

[1] Centre for Integrative Ecology, School of Biological Sciences, University of Canterbury, Private Bag

4800, Christchurch 8140, New Zealand.

[2] Department of Life Sciences, Imperial College London, Silwood Park Campus, Buckhurst Road, Ascot,

Berkshire SL5 7PY, United Kingdom

[3] Bio-protection Research Centre, School of Biological Sciences, University of Canterbury, Private Bag

4800, Christchurch 8140, New Zealand.

[4] Landcare Research, PO Box 69040, Lincoln 7640, New Zealand

[5] Department of Soil Quality, Wageningen University. P.O. Box 47. Wageningen 6700 AA, The

Netherlands

Statement of authorship

J.M.T. and I.A.D. designed the study, in discussion with all other authors. L.B.M.-G.

and I.A.D. conducted molecular analyses. J.M.T. conducted statistical analyses. S.J.R.

and D.A.P. collected plant abundance data. All authors contributed to field sampling

and writing of the manuscript.

Running title: Network assembly and disassembly

Keywords: community assembly, succession, ecosystem development, retrogression,

mutualist network, mycorrhizal symbiosis, preferential attachment

Number of words in abstract:150

Number of words in the main text (excluding abstract, acknowledgements,

references, table and figure legends: 4,981

Number of references: 48

Number of figures: 2

Corresponding author: Jason M. Tylianakis, School of Biological Sciences, University

of Canterbury, Private Bag 4800, Christchurch 8140, New Zealand. Email:

jason.tylianakis@canterbury.ac.nz   Ph: +64 3 369 5379 Fax: +64 3 3642590

Abstract

The processes whereby ecological networks emerge, persist and decay throughout ecosystem development are largely unknown. Here we study networks of plant and arbuscular mycorrhizal fungal (AMF) communities along a 120,000 yr soil chronosequence, as they undergo assembly (progression) and then disassembly (retrogression). We found that network assembly and disassembly were symmetrical, self-reinforcing processes that together were capable of generating key attributes of network architecture. Plant and AMF species that had short indirect paths to others in the community (i.e. high centrality), rather than many direct interaction partners (i.e. high degree), were best able to attract new interaction partners and in the case of AMF species, also to retain existing interactions with plants during retrogression. We then show using simulations that these non-random patterns of attachment and detachment promote nestedness of the network. These results have implications for predicting extinction sequences, identifying focal points for invasions, and suggesting trajectories for restoration.

Introduction

The arrangement of species interactions in ecological networks can determine the stability and functioning of ecosystems (Bastolla *et al.* 2009; Thebault & Fontaine 2010; Thompson *et al.* 2012). However, the processes that generate and maintain network structure during ecosystem development are unresolved. The generally accepted hypothesis for complex network assembly is that new nodes (e.g., species) preferentially connect with those that are linked to many immediate neighbors (Barabási & Albert 1999) (i.e. generalist nodes with high 'degree'). This 'preferential attachment' was predicted (Jordano *et al.* 2003), then shown empirically (Olesen *et al.* 2008), to occur during the assembly of pollination networks within a season. Moreover, preferential attachment could in theory promote nestedness, a network property that relates to network stability (Bastolla *et al.* 2009; Thebault & Fontaine 2010), such that networks would become more stable over time (Ponisio *et al.* 2017). In contrast, long-term changes to interaction networks have shown contradictory patterns, such that new interactions can primarily involve more-connected species in ant-plant networks (Díaz-Castelazo *et al.* 2010), less-connected species in pollination networks (Burkle *et al.* 2013), or simply be opportunistic based on available partners (Ponisio *et al.* 2017). Thus, the importance of preferential attachment based on species degree remains contentious for ecological networks.

Although studies on the generation of network architecture have largely focused on node attachment during network assembly, an opposing process of preferential detachment can theoretically produce the same architecture as preferential attachment (Salathé *et al.* 2005; König *et al.* 2012). This possibility is strengthened by empirical findings that pollination networks show preferential loss of less-connected species and their interactions during network decay (Aizen *et al.* 2012; Burkle *et al.* 2013). It is also possible that network architecture may be generated by preferential attachment during community assembly, and maintained by preferential detachment during disassembly. This would imply that the same rules may operate during both the assembly and disassembly of ecological networks (Bascompte & Stouffer 2009), rather than via different pathways, as would be expected if networks

exhibit hysteresis (Suding & Hobbs 2009). To our knowledge, whether networks

assemble and disassemble via the same mechanisms remains untested.

Here we study interaction network assembly and disassembly in a community of

plants and arbuscular mycorrhizal fungi (AMF) (Montesinos-Navarro *et al.* 2012)

throughout 120,000 years of ecosystem development. Arbuscular mycorrhizal fungi

are essential mutualists of most terrestrial plants (Brundrett 2009), and have typical

within-site diversity comparable to plant communities (Öpik *et al.* 2009). Moreover,

the high importance of symbiotic interactions for determining AMF community

composition (Martínez-García *et al.* 2015; Vályi *et al.* 2015) demands a focus on

interactions (Powell *et al.* 2015) for these systems, within the context of soil

nutrients which can also be important drivers of community structure (Krüger *et al.*

2015). First, we test whether network assembly and disassembly processes are

random or related to species' direct or indirect connections to others the network.

Concomitantly, we test the hypothesis (Bascompte & Stouffer 2009) that the

processes governing attachment during assembly mirror those for detachment

during disassembly. Finally, we use simulations to determine how non-random

assembly and disassembly processes impact upon key aspects of network

architecture.

The first requirement for an interaction to take place is an encounter, and if co-

occurrence rates between AMF propagules and plants were merely stochastic

(Encinas-Viso *et al.* 2015), we would expect to see more frequent than random

attachment to the most abundant partner species. In addition to such random

encounter probabilities, species' abundances may convey benefits to their

interaction partners. For example, an established AMF species that has used its

associations with many plant species to develop an extensive mycelium network

could be an attractive partner for newly-arriving plants, because they would not

need to pay the initial construction costs for the mycelium network but they would

gain immediate access to a large soil volume (Newman 1988; Simard & Durall 2004).

Likewise, high AMF diversity can reduce competition among plant species and

provide insurance against variability in conditions (Moeller & Neubert 2016), such

that plant species that interact with many AMF experience increased biomass (Wagg

et al. 2011). This greater biomass could make AMF-generalist plants more attractive

partners because they have the capacity to provide more carbon across a wider

range of conditions. Selection for partner quality could therefore lead to traditional

preferential attachment for abundant, generalist species (Barabási & Albert 1999;

Olesen *et al.* 2008), and a mechanism for this selection could be post-infection

selectivity by plants and AMF in their use of interaction partners (Kiers *et al.* 2011).

In fact, preferential attachment to generalists could also occur if generality is simply

an indicator of low partner selectivity.

The benefits of interacting with generalists may also depend in part on whether they

share their interaction partners with other species. Although AMF typically interact

with many plant species, there is growing evidence that their composition (Martínez-

García *et al.* 2015) and fitness can differ across plant hosts (Ehinger *et al.* 2009).

Host-plant selectivity may not necessarily be species specific, but rather groups of

AMF can associate with ecological groups of plants (Davison *et al.* 2011). Closely-

related AMF also tend to co-occur in a given location (Horn *et al.* 2014), and closely-

related organisms tend to share interaction partners (Gómez *et al.* 2010). We

therefore hypothesize that large interacting consortia of ecologically-similar, and in

some cases phylogenetically related, plants and fungi will form at a given

successional stage, with the dominant consortium potentially differing between

early-successional generalist species and later-successional forest species (Öpik *et al.*

2009; Davison *et al.* 2011). Thus, at each stage of ecosystem development, newly-

arriving fungi would have a higher probability of interacting with plants in the

dominant ecological group at that stage, which will already interact with species

from the dominant fungal group. These plant species would not necessarily be the

most generalist individually, but they would share AMF partners with many other

plant species in the dominant group at that stage (i.e. with the majority of species in

the site). This high interaction-partner overlap with others in the network may also

bring fitness benefits by reducing the distance that positive indirect effects must

travel (Bastolla *et al.* 2009), which may make species with many short indirect paths

to others more reliable interaction partners. We therefore hypothesise that the

attractiveness of a species to new arrivals will depend on its distance (i.e. path

length) from other species in the network, as a measure of the extent to which it

could receive positive indirect effects by sharing partners with species from the

dominant consortium (see Fig. S1 for our approach to measuring this path length).

To test these hypotheses, we used a dataset (Martínez-García *et al.* 2015) of 10

plant-AMF networks that were sampled at different sites along a long-term

(>120,000 yr) chronosequence comprising all stages of ecosystem development,

including retrogression (Peltzer *et al.* 2010). Chronosequences represent a powerful

tool for understanding long-term co-ordinated changes amongst species and their

interactions, resource availability and ecosystem processes (Peltzer *et al.* 2010).

Strong gradients from initial N-limitation to eventual strong P-limitation of

ecosystem processes drive turnover of plant and AMF species (Richardson *et al.*

2004; Martínez-García *et al.* 2015). We treated ecosystem progression (during which

plant species richness and biomass increase; Richardson *et al.* 2004) as the assembly

phase, and ecosystem retrogression (during which strong P-limitation drives declines

in plant diversity and biomass; Richardson *et al.* 2004) as the network disassembly

phase. We defined attachment as the first ecosystem stage at which a plant or AMF

species was observed along the chronosequence, and detachment as the final

ecosystem stage in which a species was observed.

Materials and Methods

<u>Dataset</u>

We generated networks of interactions between vascular plants and arbuscular

mycorrhizal fungi (AMF) using data from a study that examined AMF beta diversity

and the importance of soil age vs. plant host for explaining AMF community

structure (Martínez-García *et al.* 2015). Sampling and sequencing methods can be

found in that study or in the Supporting Information (Appendix 1), but we summarise

the key points here. Sampling was conducted along the Franz Josef soil

chronosequence, on the southern west coast of the South Island, New Zealand.

An important feature of this chronosequence, which led us to select this site, is that

strong soil nutrient gradients (Appendix 1 in the Supporting Information) are

associated with pronounced shifts in ecological community composition, structure

and function, such that ecosystem development exhibits a clear progressive phase

up to the surface soils aged 5,000 - 12,000 years, then retrogression in the older sites.

We treated the sites up to 12,000 years old as the assembly phase, because this was

the site of peak tree biomass and height, and a retrogressive phase thereafter in

which tree basal area (biomass) declines about three-fold (Richardson *et al.* 2004;

Wardle *et al.* 2008) (the disassembly phase). We defined assembly and disassembly

phases *a priori* according to well-characterized ecosystem stages of progression and

retrogression (Richardson *et al.* 2004; Wardle *et al.* 2008), rather than by using the

size of the sampled plant-AMF network, for two reasons. First, younger sites have

many herbaceous plants of low biomass (so many can fit into a given sample area),

which can make sampled species richness per plot appear lower in later sites (see

Table S1) because the roots of large-biomass trees are more likely to be repeatedly

sampled at random. Second, many of the diverse species that were sampled at

young sites were not important components of our plant-AMF network, which is

evidenced by the sampled plant-fungal associations showing a significant pattern of

accumulation up to the site of peak biomass then subtraction during retrogression

(described below).

We sampled fungal communities on roots at ten surfaces of the following ages (in

years): <5, 15, 70, 290, 500, 1000, 5000, 12,000, 60,000, 120,000 (Supporting

Information: Appendix 1 and Table S1). At each site we sampled 50 root fragments,

which then underwent a molecular analysis to identify the plant species, as well as

any AMF present inside the root (see Martínez-García *et al.* 2015 or Appendix 1 in

the Supporting Information). We did not assign weights to these links, as the

hypotheses we test here relate to the initial formation of interactions, rather than

their strength or frequency once formed.

<u>Data analysis</u>

We hypothesized that species entering the network for the first time during the

progressive phase would associate non-randomly with other existing species based

on their direct or indirect connections to other species within the network (i.e.

preferential attachment; Barabási & Albert 1999). We also hypothesized that the

same would occur during the disassembly phase, such that the probability of species

going locally extinct from the network would depend on the network position of

species with which they associate (i.e. preferential detachment (Salathé *et al.* 2005;

Aizen *et al.* 2012)). We tested for preferential formation of interactions (attachment)

by new species during ecosystem progression (5 - 12,000 year-old) and preferential

disappearance (detachment) during ecosystem retrogression (12,000 − 120,000

year-old), representing respectively the assembly and disassembly of the network. A

species was deemed to have attached during the progressive phase when it first

appeared in our samples of the chronosequence. During the progressive phase, once

a species had attached (i.e. a link was formed), it was deemed to remain part of the

network for the remainder of ecosystem progression. This assumption was made to

ensure that any species that appeared, disappeared then reappeared during the

progressive phase were not counted twice, because this would give any species-

specific interaction preferences of these species a disproportionate weighting in

analyses. Also, it gave the largest sample size with which to estimate each species'

number of interaction partners (our predictor variable) in each stage.

Detachment was deemed to have occurred during the retrogressive phase when a

species that had been present during the peak of the progressive phase was absent

from a given site and all remaining sites along the sequence. Note that, because by

definition attachment and detachment processes must occur during the interval

between two chronosequence stages, we treated the 12,000 year-old site (which

represented the community peak in diversity and biomass) as the 'peak' community,

comprising the last stage of assembly and the first stage from which disassembly

occurred. Although there were only two sites older than this 'peak' community, the

statistical power for the test came from the number of species within each site,

because preferential detachment was tested across species, within sites.

An underpinning assumption of this analysis (and of our selection of the 12,000 year

site as the peak) is that interactions have the tendency to accumulate during the

assembly phase then be lost during the disassembly phase. In other words, the

interactions present in sites preceding or following the peak community would be

nested subsets of those in the peak site. This pattern has been seen previously

during the disassembly of pollination networks (Aizen *et al.* 2012). To test for such a

pattern, we generated two matrices in which we ranked the sites according to their

age (rows), either becoming progressively older or younger than the site of peak

biomass, which was in the top row. Cells of the matrices depicted the presence or

absence of each pairwise plant-AMF association (columns). These matrices included

all interactions, not just those present in the peak-biomass site. We tested each

matrix for nestedness, using the nestedness metric based on overlap and decreasing

fill (NODF), with 999 permutations and a swap algorithm, using the oecosimu

function in the vegan (Oksanen *et al.* 2008) package for R (R Core Team 2013). This

analysis revealed a pattern whereby the peak biomass site had a set of interactions,

of which progressively older or younger sites tended to have decreasing subsets ($P <$

0.001 for rows, columns and overall NODF statistic for both assembly and

disassembly), such that interactions, on average, tended to progressively accumulate

during progression but be sequentially lost during retrogression.

Preferential attachment and detachment

We analyzed preferential attachment and detachment processes using generalized

linear mixed effects models with binomial errors and the canonical logit link function,

conducted in the lme4 package (Bates *et al.* 2014) in R. We used separate models to

test for attachment processes during progression vs. detachment processes during

retrogression, and for plants vs. AMF. The response variable in each model was

binary, whereby each existing species in the network for each site was coded with a

value of one if a newly-arriving species attached to it, or detached from it, in that site,

otherwise zero. Site (surface along the chronosequence) was included as a random

factor, such that preferential attachment was tested according to the relative

network roles of the various species within a site, rather than comparing species

across sites, which could be influenced by differences in the sizes of networks.

Identity of the attaching species, nested within site, was also included as a random

factor to control for the non-independence of a single plant or AMF-OTU attaching

or detaching from multiple partners. The fixed predictor variable was either degree

or centrality (defined below).

To measure the path length among species of a trophic level, which is related to the

number of species through which positive indirect effects (Bastolla *et al.* 2009) must

travel, we used closeness centrality (hereafter 'centrality', defined in Fig. S1) in the

unipartite projection of the bipartite interaction network (Gómez & Perfectti 2012).

Centrality is often correlated with degree (Martín González *et al.* 2010) (see also Fig.

S3), but carries additional information by measuring the number of links from a focal

species to all others in the network (not just immediate neighbors), and has been

used as a measure of species' importance (Martín González *et al.* 2010) or functional

specialization (Dalsgaard *et al.* 2008) in mutualistic networks. In the unipartite

projection, species at a given trophic level are linked if they share an interaction

partner (Fig. S1), so high centrality indicates that a species shares partners with

many other species that also share partners with many others. In this sense, high

centrality in the unipartite projection (Fig. S1) indicates that positive indirect effects

of sharing mutualists with other species at the same trophic level (Bastolla *et al.*

2009) would have a short distance to travel. In contrast, species with low centrality

tend to interact with partners that are not used by the dominant consortium of

tightly-interacting plants and AMF. Although centrality may be implicitly used to

infer the nature of flows through a network (e.g. paths vs. walks, replication vs.

spread), we do not make any assumption about the nature of flows, if any, in the

plant-AMF network. Rather, we use centrality simply as a measure of the length of

indirect paths within a trophic level. Centrality was calculated in the 'sna' package

(Butts 2014) in R, scaled within a given network (i.e. within each site).

We tested whether attachment to a node during assembly, or detachment from a

node during disassembly, depended on that node's centrality (Fig. S1) or degree (the

number of partners with which the species interacts, normalized within networks of

a given site age). We used the centrality/degree of a node in the chronosequence

stage immediately prior to that in which the new node attached or an existing node

detached. This allowed us to compare the fit of models with centrality vs. degree as

the predictor using the Akaike Information Criterion (AIC). Thus, the predictor

variable in each model was either the degree or centrality of each existing species in

the site, to which newly-arriving species either attached vs. did not, or detached vs.

did not in the next stage of the chronosequence. For completeness, we also ran

models with both of these predictors, to determine if any potential collinearity

qualitatively altered their separate effects. None of our models showed any signs of

overdispersion. We repeated these analyses controlling for species abundances in

case sampling effort generated any spurious effects (see Controlling for the effect of

species abundances in Appendix 1 of the Supporting Information).

We omitted from this analysis any links representing new species attaching to other

newly-arriving species, because we could not ascribe a value of network position for

those species to use as the predictor variable. However, these pairings among two

newly-arriving species were rare, and the species involved typically also interacted

with other species already present in the network. Across all sites, 20 newly-arriving

plant species were associated with newly-arriving AMF, and of these, only 4 did not

also attach to AMF species that were already present in the network, for which

network position could be calculated. Similarly, of the 15 newly-arriving AMF species

that attached to newly-arriving plants, 7 did not also attach to existing species.

<u>Simulations</u>

To explore the consequences of preferential attachment and detachment based on

centrality vs. degree, we conducted a set of simulations of network assembly and

disassembly scenarios (more detail provided in Appendix 1 of the Supporting

Information). At each time step, newly-arriving plant or AMF species would interact

with one existing species, with a probability proportional to the existing species'

degree (i.e. the Barabási & Albert (1999) model) or distance from other species at

the same trophic level (closeness centrality). We also ran a third scenario whereby

attachment was random (equally probable attachment to any existing species), to

provide a null point of comparison.

This assembly phase continued until networks contained 50 species of plant and 50

of AMF. Subsequently, each network was subjected to a disassembly phase, using

the same scenario (degree-based, centrality-based or random) as was used during

that network's assembly. Alternating plant and AMF species were removed (i.e. went

extinct) with a probability that was inversely proportional to the degree or closeness

centrality of the species with which they interacted, such that species that interacted

with species of high degree or high centrality were less likely to go extinct. As with

the assembly phase, we ran a scenario of the disassembly phase whereby extinction

was random.

Each scenario of the assembly and disassembly phases was run for 1000 replicates

(each comprising multiple time steps). At each time step within each replicate, after

species arrived or went extinct, we recorded the nestedness of the network using

NODF, calculated using the nestednodf function in the vegan package (Oksanen *et al.*

2008) for R. Nestedness is a common feature of mutualistic networks (Bascompte *et*

al. 2003), which has also been observed recently in plant-mycorrhizal networks

(Montesinos-Navarro *et al.* 2012) and has been shown to increase the persistence of

networks (Bastolla *et al.* 2009; Thebault & Fontaine 2010). We scaled the observed

NODF by the distribution of NODF values obtained by randomizing the adjacency

matrix using a null model (Appendix 1 of the Supporting Information).

Results

In total our samples yielded 33 operational taxonomic units (OTUs, hereafter

referred to as species) of AMF and 53 species of plant, among which we observed

399 pairwise interactions (i.e. links, defined as the colonisation of a plant root by an

AMF species) (Fig. 1). There were an average of 16 AMF (SD = 3.33) and 10.8 (SD =

3.77) plant species, with 137.5 (SD = 24.14) interactions per site (Table S1). The AMF

community changes and turnover along the chronosequence have been described in

detail elsewhere (Martínez-García *et al.* 2015).

We first tested whether preferential attachment or detachment occurred to existing

species that had many partners or high centrality. At their earliest occurrence along

the sequence, both plant and AMF species were significantly more likely to interact

with species that shared interaction partners with many others (i.e. high centrality,

Fig. 2A, $P < 0.0001$ in both cases), but not those with many partners (i.e. high degree;

$P > 0.862$ in both cases) (Table S2). In addition to this preferential attachment, plants

also showed preferential detachment, such that they were significantly less likely to

be lost from the network during ecosystem retrogression if they interacted with AMF

species that had high centrality ($P < 0.0001$; Fig. 2D; Table S2). In contrast, the

apparent decline in AMF detachment probability with plant centrality (Fig. 2C) or

degree was non-significant (Table S2). Thus, symmetrical assembly and disassembly

processes, either of which could theoretically generate complex network

architecture (Barabási & Albert 1999; Salathé *et al.* 2005; Bastolla *et al.* 2009; König

et al. 2012), both operated on the plant community to rapidly generate architecture

that did not change significantly throughout ecosystem development (Appendix 2 of

the Supporting Information, Table S3, Fig. S2).

Although species with many direct interaction partners also had significantly shorter

indirect paths on average (i.e. degree and centrality were correlated in our networks;

Fig. S3), we found no evidence of preferential attachment or detachment of plants or

AMF based on degree alone (Table S2). Moreover, when both degree and centrality

were included in models, attachment probability *decreased* with increasing degree

(Table S2) and plant detachment probability *increased* when they interacted with

AMF of high degree (Table S2), as has been observed in a long-term study of

pollination networks (Burkle *et al.* 2013). Importantly, in all our models, centrality

explained attachment or detachment probability better than did equivalent models

with only degree as the predictor, and its effect remained qualitatively unchanged

when degree was also included in the model (Table S2). Most importantly, the

relation between AMF centrality and plant attachment and detachment probability

remained significant after controlling for abundance (Appendix 2 of the Supporting

Information, Table S4).

Our findings suggest that both attachment and detachment are key processes in

network development and, when combined, the loss or gain of links becomes self-

reinforcing (König *et al.* 2012). We simulated the processes examined here to

explore the consequences for network architecture (Appendix 1 of the Supporting

Information) and found that both preferential attachment and detachment

processes were important for determining and maintaining the nestedness of the

network (Fig. S4), a key element of ecological network architecture (Bastolla *et al.*

2009; Montesinos-Navarro *et al.* 2012) whereby specialists interact with species that

also interact with generalists. Preferential attachment (based on degree or

centrality) generated networks that were considerably more nested than random,

and preferential detachment maintained this nestedness in the face of species

extinctions (Fig. S4).

Discussion

Obviously species do not have information about the architecture of their interaction

networks, so centrality (a species' indirect proximity to others at the same trophic

level) must be associated with ecological characteristics that make species more

likely to attract and retain interaction partners. There are several potential, non-

exclusive explanations for this attractiveness. The simplest would be if abundant

species shared many partners due to a high probability of random encounters. In our

dataset, there was a weak correlation between a species' abundance and centrality

within sites (Fig. S5). Moreover, the core of interacting generalists (i.e. the species

involved in the majority of interactions in the network, and whose interaction

partners overlapped considerably with the rest of the community) was present

throughout much of ecosystem development (see dark colors in top left of matrices

in Fig. 1), and this ubiquity likely made them a reliable target for newly arriving

species. However, causality cannot be inferred from these correlations in our dataset,

as attraction of diverse fungal mutualists could equally cause longer-term

persistence of plants (Scheublin *et al.* 2007). Moreover, the significant relationships

between centrality and attachment or detachment probability remained

qualitatively unchanged when controlling for abundance (Appendix 2 of the

Supporting Information, Table S4).

Therefore, although abundance likely plays some role, the attractiveness of plant

species with high centrality to AMF, and vice versa, reflects additional benefits to

associating with those species. One such benefit of close proximity is the potential to

receive beneficial indirect effects (Bastolla *et al.* 2009); plants may benefit from the

AMF communities generated by other species (Bever 2002), and plant-plant

competition can be reduced (and plant biomass increased) by high AMF diversity

(Wagg *et al.* 2011). Thus, consortia of plants that are able to share diverse AMF

communities with one another may grow better, and therefore have the ability to

provide more carbon to AMF symbionts. These groupings may also reflect habitat

preference of the species involved, with habitat generalist plant and AMF species

interacting tightly during early succession and forest specialists during late

succession (Davison *et al.* 2011), an hypothesis congruent with findings of partner

specificity occurring at the level of ecological groups (Öpik *et al.* 2009).

Plant species were also less likely to be lost during retrogression if they interacted

with AMF species that had high centrality (Fig. 2D). We hypothesized (based on the

benefits described above) that mutually beneficial interactions will involve species

with high centrality, and we observed that plant species that preferentially

interacted with these core species were significantly more likely to persist during

retrogression. The decline in soil P during ecosystem development likely makes

plants increasingly dependent on AMF-provided P. Likewise, declining plant

photosynthetic rates along the chronosequence (Whitehead *et al.* 2005) could

increase competition among AMF for plant-derived carbon. Thus, the need to

develop and maintain mutualists and avoid symbiotic 'cheaters' will intensify along

the sequence, particularly during retrogression. Over time, both plants and AMF can

reduce their resource allocation to less-beneficial partners (Kiers *et al.* 2011),

particularly when resources are scarce (Bever 2015). The interactions that persist

during retrogression may therefore be those that are most mutually beneficial.

Previous work has found that species in late-successional plant-AMF networks tend

to interact more frequently with a subset of their total range of partners (Bennett *et*

al. 2013), and a viable strategy may thus be to test a number of partners initially,

then gradually restrict resource allocation to a subset of these that provide the

greatest benefit (Bever 2015) as resources become limiting. This process could easily

occur at the level of plant individuals, but be reflected as a loss of links at the species

level.

Finally, if our findings are found to generalize to other ecological networks, they

would have several potential applications. First the symmetry of assembly and

disassembly processes for plants and our observation that interactions in sites

moving forward or backwards from the site of peak biomass were nested subsets of

that site (see Materials and Methods section) suggests that knowledge of assembly

order for a given system could be used to predict the interactions most at risk of

extinction, such as in using network information for conservation (Tylianakis *et al.*

2010). Species reintroductions during restoration could also follow the reverse order

of extinction and replace extinct species with others that fulfill the same network

role, or focus on species that share interaction partners with others, and will then

attract interactions with any new species that colonize. Similarly, non-native species

should interact preferentially with high-partner-overlap species in the network,

which could therefore be a focus for biosecurity monitoring to detect invasions early.

It is impossible to fully understand the architecture of complex networks in isolation

from the dynamic processes that generate and maintain them (Barabási & Albert

1999). We have shown empirically that both preferential attachment and

detachment underpinned the development of an ecological network under changing

abiotic conditions. Importantly, species with short indirect paths to others at the

same trophic level were more attractive interaction partners, congruent with

suggestions that positive indirect effects may be important for species persistence in

mutualistic networks (Bastolla *et al.* 2009). Moreover, interactions in sites moving

forward or backward in age from the site of peak biomass were nested subsets of

those in the peak biomass site. This pattern has been observed in the disassembly of

pollination networks (Aizen *et al.* 2012), and we found it to also occur during

assembly. Finally, we showed through simulation that symmetrical, self-reinforcing

processes can generate and maintain important features of interaction network

architecture (Barabási & Albert 1999; Salathé *et al.* 2005; Bastolla *et al.* 2009; König

et al. 2012), thereby linking species colonization and extinction processes to

emergent and potentially stabilizing (Bastolla *et al.* 2009; Thebault & Fontaine 2010)

ecosystem properties.

References

Aizen, M.A., Sabatino, M. & Tylianakis, J.M. (2012). Specialization and Rarity Predict

Nonrandom Loss of Interactions from Mutualist Networks. *Science*, 335,

1486-1489.

Barabási, A.-L. & Albert, R. (1999). Emergence of scaling in random networks. *Science*,

286, 509-512.

Bascompte, J., Jordano, P., Melian, C.J. & Olesen, J.M. (2003). The nested assembly

of plant-animal mutualistic networks. *Proceedings of the National Academy*

of Sciences of the United States of America, 100, 9383-9387.

Bascompte, J. & Stouffer, D.B. (2009). The assembly and disassembly of ecological

networks. *Philosophical Transactions of the Royal Society B-Biological*

Sciences, 364, 1781-1787.

Bastolla, U., Fortuna, M.A., Pascual-Garcia, A., Ferrera, A., Luque, B. & Bascompte, J.

(2009). The architecture of mutualistic networks minimizes competition and

increases biodiversity. *Nature*, 458, 1018-U1091.

Bates, D., Maechler, M., Bolker, B. & Walker, S. (2014). lme4: Linear mixed-effects

models using Eigen and S4.

Bennett, A.E., Daniell, T.J., Öpik, M., Davison, J., Moora, M., Zobel, M. *et al.* (2013).

Arbuscular mycorrhizal fungal networks vary throughout the growing season

and between successional stages. *PloS one*, 8, e83241.

Bever, J.D. (2002). Negative feedback within a mutualism: host–specific growth of

mycorrhizal fungi reduces plant benefit. *Proceedings of the Royal Society of*

London B: Biological Sciences, 269, 2595-2601.

Bever, J.D. (2015). Preferential allocation, physio‐evolutionary feedbacks, and the

stability and environmental patterns of mutualism between plants and their

root symbionts. *New Phytologist*, 205, 1503-1514.

Brundrett, M.C. (2009). Mycorrhizal associations and other means of nutrition of
vascular plants: understanding the global diversity of host plants by resolving
conflicting information and developing reliable means of diagnosis. *Plant and*
Soil, 320, 37-77.

Burkle, L.A., Marlin, J.C. & Knight, T.M. (2013). Plant-pollinator interactions over 120
years: loss of species, co-occurrence, and function. *Science*, 339, 1611-1615.

Butts, C.T. (2014). sna: Tools for Social Network Analysis.

Dalsgaard, B., Martín González, A.M., Olesen, J.M., Timmermann, A., Andersen, L.H.
& Ollerton, J. (2008). Pollination networks and functional specialization: a test
using Lesser Antillean plant–hummingbird assemblages. *Oikos*, 117, 789-793.

Davison, J., Öpik, M., Daniell, T.J., Moora, M. & Zobel, M. (2011). Arbuscular
mycorrhizal fungal communities in plant roots are not random assemblages.
FEMS Microbiology Ecology, 78, 103-115.

Díaz-Castelazo, C., Guimarães, P.R., Jordano, P., Thompson, J.N., Marquis, R.J. & Rico-
Gray, V. (2010). Changes of a mutualistic network over time: reanalysis over a
10‐year period. *Ecology*, 91, 793-801.

Ehinger, M., Koch, A.M. & Sanders, I.R. (2009). Changes in arbuscular mycorrhizal
fungal phenotypes and genotypes in response to plant species identity and
phosphorus concentration. *New Phytologist*, 184, 412-423.

Encinas-Viso, F., Alonso, D., Klironomos, J.N., Etienne, R.S. & Chang, E.R. (2015). Plant
–mycorrhizal fungus co‐occurrence network lacks substantial structure.
Oikos.

Gómez, J.M. & Perfectti, F. (2012). Fitness consequences of centrality in mutualistic
individual-based networks. *Proceedings of the Royal Society B: Biological*
Sciences, 279, 1754-1760.

Gómez, J.M., Verdú, M. & Perfectti, F. (2010). Ecological interactions are
evolutionarily conserved across the entire tree of life. *Nature*, 465, 918-921.

Horn, S., Caruso, T., Verbruggen, E., Rillig, M.C. & Hempel, S. (2014). Arbuscular
mycorrhizal fungal communities are phylogenetically clustered at small scales.
The ISME journal, 8, 2231-2242.

Jordano, P., Bascompte, J. & Olesen, J.M. (2003). Invariant properties in
coevolutionary networks of plant-animal interactions. *Ecology letters*, 6, 69-
81.
Kiers, E.T., Duhamel, M., Beesetty, Y., Mensah, J.A., Franken, O., Verbruggen, E. *et al.*
(2011). Reciprocal rewards stabilize cooperation in the mycorrhizal symbiosis.
science, 333, 880-882.
König, M., Tessone, C. & Zenou, Y. (2012). Nestedness in Networks: A Theoretical
Model and Someapplications. *SIEPR Discussion Papers*, 11, 1-61.
Krüger, M., Teste, F.P., Laliberté, E., Lambers, H., Coghlan, M., Zemunik, G. *et al.*
(2015). The rise and fall of arbuscular mycorrhizal fungal diversity during
ecosystem retrogression. *Mol. Ecol.*, 24, 4912-4930.
Martín González, A.M., Dalsgaard, B. & Olesen, J.M. (2010). Centrality measures and
the importance of generalist species in pollination networks. *Ecological
Complexity*, 7, 36-43.
Martínez-García, L.B., Richardson, S.J., Tylianakis, J.M., Peltzer, D.A. & Dickie, I.A.
(2015). Host identity is a dominant driver of mycorrhizal fungal community
composition during ecosystem development. *New Phytologist*, 205, 1565-
1576.
Moeller, H.V. & Neubert, M.G. (2016). Multiple friends with benefits: an optimal
mutualist management strategy? *The American Naturalist*, 187, E1-E12.
Montesinos-Navarro, A., Segarra-Moragues, J.G., Valiente-Banuet, A. & Verdu, M.
(2012). The network structure of plant-arbuscular mycorrhizal fungi. *New
Phytologist*, 194, 536-547.
Newman, E.I. (1988). Mycorrhizal links between plants: their functioning and
ecological significance. *Advances in Ecological Research*, 18, 243-270.
Oksanen, J., Kindt, R., Legendre, P., O'Hara, B., Simpson, G.L., Solymos, P. *et al.*
(2008). vegan: Community Ecology Package <u>http://cran.r-project.org/</u>.
Olesen, J.M., Bascompte, J., Elberling, H. & Jordano, P. (2008). Temporal dynamics in
a pollination network. *Ecology*, 89, 1573-1582.
Öpik, M., Metsis, M., Daniell, T., Zobel, M. & Moora, M. (2009). Large‐scale parallel
454 sequencing reveals host ecological group specificity of arbuscular
mycorrhizal fungi in a boreonemoral forest. *New Phytologist*, 184, 424-437.

Peltzer, D.A., Wardle, D.A., Allison, V.J., Baisden, W.T., Bardgett, R.D., Chadwick, O.A.

et al. (2010). Understanding ecosystem retrogression. *Ecological Monographs*,

80, 509-529.

Ponisio, L.C., Gaiarsa, M.P. & Kremen, C. (2017). Opportunistic attachment

assembles plant–pollinator networks. *Ecology Letters*, 20, 1261-1272.

Powell, J.R., Karunaratne, S., Campbell, C.D., Yao, H., Robinson, L. & Singh, B.K.

(2015). Deterministic processes vary during community assembly for

ecologically dissimilar taxa. *Nature communications*, 6.

R Core Team (2013). R: A language and environment for statistical computing. R

Foundation for Statistical Computing, Vienna Austria.

Richardson, S.J., Peltzer, D.A., Allen, R.B., McGlone, M.S. & Parfitt, R.L. (2004). Rapid

development of phosphorus limitation in temperate rainforest along the

Franz Josef soil chronosequence. *Oecologia*, 139, 267-276.

Salathé, M., May, R.M. & Bonhoeffer, S. (2005). The evolution of network topology

by selective removal. *Journal of the Royal Society Interface*, 2, 533-536.

Scheublin, T.R., Van Logtestijn, R.S.P. & Van Der Heijden, M.G.A. (2007). Presence

and identity of arbuscular mycorrhizal fungi influence competitive

interactions between plant species. *Journal of Ecology*, 95, 631-638.

Simard, S.W. & Durall, D.M. (2004). Mycorrhizal networks: a review of their extent,

function, and importance. *Canadian Journal of Botany*, 82, 1140-1165.

Suding, K.N. & Hobbs, R.J. (2009). Threshold models in restoration and conservation:

a developing framework. *Trends in ecology & evolution*, 24, 271-279.

Thebault, E. & Fontaine, C. (2010). Stability of Ecological Communities and the

Architecture of Mutualistic and Trophic Networks. *Science*, 329, 853-856.

Thompson, R.M., Brose, U., Dunne, J.A., Hall Jr, R.O., Hladyz, S., Kitching, R.L. *et al.*

(2012). Food webs: reconciling the structure and function of biodiversity.

Trends in ecology & evolution.

Tylianakis, J., Laliberté, E., Nielsen, A. & Bascompte, J. (2010). Conservation of

species interaction networks. *Biological Conservation*, 143, 2270-2279.

Vályi, K., Rillig, M.C. & Hempel, S. (2015). Land‐use intensity and host plant identity

interactively shape communities of arbuscular mycorrhizal fungi in roots of

grassland plants. *New Phytologist*, 205, 1577-1586.

Wagg, C., Jansa, J., Stadler, M., Schmid, B. & Van Der Heijden, M.G.A. (2011).

628   Mycorrhizal fungal identity and diversity relaxes plant-plant competition.

629   *Ecology*, 92, 1303-1313.

630   Wardle, D., Bardgett, R., Walker, L., Peltzer, D. & Lagerström, A. (2008). The response

631   of plant diversity to ecosystem retrogression: evidence from contrasting long-

632   term chronosequences. *Oikos*, 117, 93 -103.

633   Whitehead, D., Boelman, N.T., Turnbull, M.H., Griffin, K.L., Tissue, D.T., Barbour,

634   M.M. *et al.* (2005). Photosynthesis and reflectance indices for rainforest

635   species in ecosystems undergoing progression and retrogression along a soil

636   fertility chronosequence in New Zealand. *Oecologia*, 144, 233-244.

Acknowledgements

We thank P.J. Bellingham, T. Fukami, J.H. Jones, S. Pawar, D.B. Stouffer and members

of the Tylianakis/Stouffer lab for critical discussions and comments on the

manuscript. N. Bolstridge and C. Mitchel provided valuable lab assistance. The

research was funded by Core funding for Crown Research Institutes from the New

Zealand Ministry of Business, Innovation and Employment's Science and Innovation

Group, a Rutherford Discovery Fellowship to JMT, and the Bio-protection Research

Centre (JMT and IAD).

Author information

The authors declare no competing financial interests. Correspondence and requests

for materials should be addressed to J.M.T. (jason.tylianakis@canterbury.ac.nz).

Supplementary Information is available online

Figure 1: Plant-mycorrhizal association matrices during assembly (N = 8 sites) and

disassembly (i.e. retrogressive, N = 3 sites) phases of ecosystem development. The

site of peak biomass represents the end of assembly and beginning of disassembly,

so is included in both figures. Darker colours indicate that the plant-mycorrhizal

association was present in a greater number of sites (i.e. it formed early during

assembly or persisted during retrogression). These darker interactions tend to

involve species that share interaction partners with many others that also share

partners with each other (i.e. species with high centrality). The network is

significantly nested (see Appendix 2 of the Supporting Information), whereby

specialists interact with proper subsets of the species that interact with generalists.

Figure 2: Probability (left vertical axis) that a new A,C) arbuscular mycorrhizal fungal

(AMF), or B,D) plant species entering the network will interact with an existing

species during assembly (top), or detach from a species during disassembly (bottom)

as a function of that species' proximity in the network to others of its trophic level

(measured as closeness centrality of the unipartite projection of the plant-AMF

network, see Fig. S1). Histograms top and bottom of each graph represent the

frequencies (right vertical axis) of ones and zeroes respectively in the raw data.

Trend lines are based on inverse-linked coefficients of a binomial linear mixed effects

model with chronosequence stage (plant-mycorrhizal network) as a random effect.

Solid lines were statistically significant (P < 0.025 in all cases), dashed line was not

significant at alpha = 0.05.