

1 **On the inconsistency of pollinator species traits for predicting**
2 **either response to agricultural intensification or functional**
3 **contribution.**

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12 **Running headline:** Pollinator traits are not predictive

13 **Abstract:**

14 The response and effect trait framework, if supported empirically, would
15 provide for powerful and general predictions about how biodiversity loss will
16 lead to loss in ecosystem function. This framework proposes that species
17 traits will explain how different species respond to disturbance (i.e. response
18 traits) as well as their contribution to ecosystem function (i.e. effect traits).
19 However, predictive response and effect traits remain elusive for most
20 systems. Here, we present detailed data on crop pollination services provided
21 by native, wild bees to explore the role of six commonly used species traits in
22 determining how crop pollination is affected by increasing agricultural
23 intensification. Analyses were conducted in parallel for three crop systems
24 (watermelon, cranberry, and blueberry) located within the same geographical
25 region (mid-Atlantic USA). Bee species traits did not strongly predict species'
26 response to agricultural intensification, and the few traits that were weakly
27 predictive were not consistent across crops. Similarly, no trait predicted
28 species' overall functional contribution in any of the three crop systems,
29 although body size was a good predictor of per capita efficiency in two
30 systems. So far, most studies looking for response or effect traits in pollination
31 systems have found weak and often contradicting links. Overall we were
32 unable to make generalizable predictions regarding species responses to
33 land-use change and its effect on the delivery of ecosystem services.
34 Pollinator traits may be useful for understanding ecological processes in some
35 systems, but thus far the promise of traits-based ecology has yet to be fulfilled
36 for pollination ecology.

37 **Keywords:** Biodiversity, bees, ecosystem services, ecosystem function,
38 response traits, effect traits, body size, diet specialism.

39

40 **Introduction**

41 Land-use change, along with other human-induced global change drivers, is
42 accelerating the rates of extinction of most taxa (Ellis et al. 2010). At the same
43 time, humanity relies on ecosystem services that wild species deliver, such as
44 pollination and pest control by insects, and nutrient cycling by microorganisms
45 (Cardinale et al. 2012). Thus, it is important to understand the relationship
46 between biodiversity loss and ecosystem service delivery (Schwartz et al.
47 2000). In particular, making generalizable predictions regarding how the
48 decline or local extinction of taxa will affect ecosystem services will allow for
49 targeted conservation actions to ameliorate negative impacts of land-use
50 change.

51 One avenue for predicting the functional consequences of biodiversity loss is
52 the response and effect trait framework (Lavorel and Garnier 2002, Naeem
53 and Wright 2003, McGill et al. 2006). Local extinction does not occur at
54 random because extinction risk is dependent on the species' characteristics.
55 Identifying which traits govern species responses to particular threats
56 ('response traits') would provide the first step for predicting future species
57 loss. Furthermore, the magnitude by which ecosystem function declines when
58 a species is lost depends on that species' functional contribution. This, too, is

59 likely to be mediated by the species' traits ('effect traits'). Therefore, the
60 relationship between response and effect traits will mediate the magnitude of
61 the impact of human disturbance on ecosystem services (Schleuning et al.
62 2015). For example, if the same species traits that are associated with high
63 function are also most sensitive to disturbance, ecosystem function would be
64 predicted to decline rapidly (Larsen et al. 2005).

65 However, for the response-effect trait framework to be useful, it is first
66 necessary to identify response and effect traits that are both explanatory and
67 possible to measure in the field (Cadotte et al. 2011). While a few generalities
68 have emerged as to which traits make animal species at greater risk of local
69 decline, including dietary or habitat specialization and body size (Fisher and
70 Owens 2004, Flynn et al. 2009, Öckinger et al. 2010), the correlation between
71 these response traits and extinction risk has been found to be weak, variable,
72 or context-dependent (Devictor et al. 2008, Fritz et al. 2009, Powney et al.
73 2014). Similarly, although some effect traits have been identified, they are
74 often weakly predictive, and their identity varies by function and taxonomic
75 group (Gagic et al. 2015). Lastly, within the functional trait field as a whole,
76 most progress has been made in identifying functional traits for plants (Diaz et
77 al. 2016), while little is known for animals (Didham et al. 2016).

78 Here, we seek to identify response and effect traits for wild bee species
79 providing a key ecosystem service, crop pollination. The yield of most crop
80 plants increases with animal pollination (Klein et al. 2007). While managed
81 honey bees are a leading crop pollinator, wild insects contribute more than

82 half of pollinator visits to crop flowers across more than 40 crop systems
83 worldwide (Rader et al. 2016). A major threat to pollinators is habitat
84 destruction, primarily conversion to agriculture (Garibaldi et al. 2011), which is
85 also a leading cause of species loss worldwide (Pereira et al. 2010). Thus
86 agricultural land use has the potential to affect one of the ecosystem services
87 upon which agriculture itself depends (Deguines et al. 2014).

88 Our data sets were collected and analyzed in parallel and come from three
89 crop systems (watermelon, cranberry and blueberry) located within the same
90 geographical region (mid-Atlantic USA), but pollinated by distinct bee
91 communities. We determined whether six commonly-used species traits can
92 predict 1) species' responses to agricultural intensification (response traits)
93 and/or 2) species' contributions to crop pollination (effect traits) and discuss
94 our results on the light of recently published studies on pollinator
95 environment–trait and pollination-trait associations.

96 **Material and methods:**

97 *Study system*

98 We selected 49 sites across three study systems that were located throughout
99 New Jersey and eastern Pennsylvania (USA). Watermelon sites (N = 17)
100 were located in 90 x 60 km region central New Jersey and Eastern
101 Pennsylvania, where the main types of land use are agriculture and suburban
102 development, interspersed with highly fragmented deciduous forest.
103 Cranberry and blueberry sites (N = 16 each) were both located within a 35 x

104 55 km area in southern New Jersey, where the main land cover types are
105 pine-oak ericaceous heath and agriculture. All sites in all systems were
106 separated by at least 1 km (range, watermelon: 2-90 km, cranberry: 1-32 km,
107 blueberry 1-38 km).

108 All three crops are highly dependent upon bee pollination for marketable fruit
109 production (Klein et al. 2007). Commercial honey bees are used in most of our
110 study fields. However, honey bees are primarily managed hives, moved
111 throughout the region, and only found on sites during bloom. Therefore, honey
112 bees are not influenced by land cover in the same manner as wild bees and
113 are not used in our analyses. Wild bees are important pollinators in all three
114 systems (mean percentage of wild bee visits: 73% watermelon, 25%
115 cranberry, and 14% blueberry).

116 *Data collection:*

117 At all sites on all three crops, we used hand-netting to measure overall bee
118 abundance and species richness. To collect bees, we walked along a fixed
119 50-200 m² transect at standard times of day and collected all bees observed
120 to be visiting flowers. In watermelon and blueberry, bees were netted three
121 times throughout the day for 20 minutes per transect (60 minutes per date per
122 site) and twice each day in cranberry for 30 minutes per transect (120 minutes
123 per date per site). Data were collected during the peak bloom in 2010
124 (watermelon: July, cranberry: late-May-early July, blueberry: April-early May).
125 Data were collected on three days per site for watermelon and blueberry and
126 two days per site for cranberry. Detailed methods can be found in Cariveau et

127 al. (2013), Benjamin et al. (2014) and Winfree et al. (2015).

128 *Land cover characteristics of sites*

129 To relate pollinator response traits to agricultural intensification we used a
130 commonly used land-use variable, percent land cover in agricultural
131 production surrounding sites (Fahrig 2013). For this end, we required high-
132 quality land cover data for each pollinator collection site. For the cranberry
133 and blueberry sites in New Jersey, we used a continuous polygon layer
134 classified by visual photograph interpretation into 60 categories, at a minimum
135 mapping unit of 4047 m² (1 acre; GIS Data provided by the New Jersey
136 Department of Environmental Protection). For watermelon sites that extend
137 from central New Jersey into Pennsylvania, we created a similar land cover
138 data layer by manually digitizing Google Earth imagery and visually classifying
139 15 categories, at a minimum mapping unit of 5,000 m² (1.24 acres). As each
140 crop was analyzed separately, our results are robust to using different land
141 cover data. However, to simplify the interpretation of results for the three
142 crops, we reclassified all land cover data into the following 7 broad categories:
143 agriculture, open managed (for example, mowed grass), open natural or semi-
144 natural (for example, old fields), semi-urban (<30% impervious surface), urban
145 (>30% impervious surface), wooded, and open water.

146 For each data collection site, we calculated two land cover variables: percent
147 agriculture and percent natural and semi-natural open habitat. We used
148 agricultural land cover as our primary land-use change variable as it is the
149 dominant anthropogenic habitat type in all three study systems

150 (Supplementary Table A1). In addition, we also measured percent of open
151 natural/semi-natural habitat, which although it accounts for only a small
152 proportion of the total land cover (Supplementary Table A1), might be
153 disproportionately important as forage and nesting habitat for bees (Kleijn et
154 al. 2006). We calculated values for this two land cover variables at both a
155 small scale (300 m radius) and a large scale (1500 m radius), which
156 correspond to typical flight distances of small- and large-bodied bees,
157 respectively (Greenleaf et al. 2007).

158 *Pollinator function*

159 To estimate the pollination services provided per bee species, we measured
160 two variables in the field, flower visitation frequency and per visit efficiency. As
161 variation in visitation frequency may be a function of land use at individual
162 farms, we use species abundances for each species at the site with its highest
163 abundance for each crop. Hence, we assess visitation frequency at its
164 maximum, which represents the optimal visitation frequency for each species.

165 To measure the pollination efficiency, we quantified single-visit pollen
166 deposition by presenting virgin flowers to individual bees foraging on the
167 target crop. After visitation, we counted the number of pollen grains deposited
168 per flower visit (watermelon) or the number of pollen tetrads with pollen tubes
169 per flower visit (cranberry and blueberry). Because species identification in the
170 field is not possible for most bees and net collecting immediately after visits is
171 generally not possible, for the measurement of pollination efficiency we
172 grouped bees in species groups. Each group consisted of between one and

173 27 species, with the median number of species per group being 4 species
174 (Supplementary Table A2). Control flowers were left bagged until the end of
175 the field day, and contained few pollen grains (watermelon mean = 3 grains, N
176 = 40 stigmas; cranberry mean = 0 tetrads, N = 82 stigmas; blueberry mean =
177 2 tetrads, N= 734 stigmas). We used mean number of pollen grains deposited
178 by a single visit group and assigned that value to each of the species in the
179 single visit group. For detailed methods see Cariveau et al. (2013), Benjamin
180 et al. (2014), Winfree et al. (2015).

181 *Species traits*

182 Bee species vary in a number of traits that are associated with their response
183 to land-use change (Williams et al. 2010). Moreover, these traits will likely
184 affect the pollinator contribution to function, either by modifying its abundance
185 or because they are related to its per capita effectiveness. We obtained
186 detailed natural history data on 6 traits for the 90 bee species in our study: a)
187 sociality (solitary, facultative social, eusocial), b) nesting placement (hole,
188 cavity, stem, wood, ground), c) brood parasite (yes, no), d) body size, e) diet
189 breadth (level of specialization) and f) tongue length.

190 We obtained the trait data as follows. Species sociality level, nesting behavior
191 and brood parasite status were extracted from the literature (Bartomeus et al.
192 2013a). Body size (estimated from intertegular span, IT; Cane 1987) was
193 measured in the lab using collected specimens that had been identified to the
194 species level by professional taxonomists. Multiple specimens were measured
195 per species (mean = 6.6 specimens \pm 3 S.E.) and the mean across the

196 measured specimens was used as the value for the species. Bee body size
197 also correlates strongly with foraging distance (Greenleaf et al. 2007), and
198 thus is ecologically related to mobility. Tongue length was measured in the lab
199 for 7.7 ± 1.2 SE specimens per species, and the mean across the measured
200 specimens is used. For the 40 specimens for which we cannot obtain a
201 tongue measure, we estimated tongue length from the species' body size and
202 phylogeny using an allometric equation (Cariveau et al. 2016).

203 Diet breadth was calculated using six independent datasets previously
204 collected at 139 sites throughout the study region by the Winfree laboratory
205 group. Each data set consists of individual pollinator specimens that were net-
206 collected while foraging on a flowering plant species; both pollinator and plant
207 were then identified to the species level. Those datasets comprise overall 393
208 pollinator species, and 392 plant species, with 3890 plant-pollinator
209 interactions (Supplementary Text A1). Prior to calculating diet breadth, we
210 rarefied the data to 20 visitation records per bee species, to avoid
211 confounding rarity with specialization (Blüthgen et al. 2008; Winfree et al.
212 2014). Nine species had fewer than 20 records and we were unable to
213 estimate diet breadth in the manner described above. Five of these species
214 are known to be specialized and we simulated the diet breadth index of 20
215 individuals visiting the known host plants. The four other species are known to
216 be generalists and we therefore used the mean diet breadth of its genus.
217 These four species were extremely rare (< 5 records each) in our analyzed
218 dataset.

219 To calculate diet breadth for each bee species, we considered the number of
220 plants species as well as the phylogenetic breadth that the bees fed upon by
221 using a rarefied phylogenetic diversity index (Nipperess and Matsen 2013). To
222 determine phylogenetic distances among plants, we first constructed a
223 general phylogenetic tree using the PHYLOMATIC “megatree” (version
224 R201120829, Chamberlain and Szöcs 2013) which defines relationships
225 between higher plants (Webb et al. 2008). We then dated nodes across this
226 tree according to Wikström et al. (2001) and used the branch-length
227 adjustment algorithm BLADJ to estimate the age of all remaining, undated
228 nodes. Though this procedure implies that ages within our phylogenies should
229 be treated as approximations (Beaulieu et al. 2007), previous analysis
230 indicates marked improvements of phylogenetic analyses when even a limited
231 number of nodes are properly dated (Webb 2000).

232 *Statistical analysis*

233 **Response traits:** To investigate which traits are associated with
234 environmental variables related to agricultural intensification, we used a
235 model-based approach to the fourth-corner problem (Brown et al. 2014). The
236 fourth-corner problem highlights the difficulty of studying the environment–trait
237 associations and can be conceptualized as a set of four matrices: abundances
238 by species, trait data by species, environmental data by sites, and
239 environmental data by traits, being the relationships of this last corner the
240 ones to be estimated (Legendre et al. 1997). The core idea of the model-
241 based approach is to fit a predictive model for species abundance as a

242 function of environmental variables, species traits and their interaction. The
243 environment-trait interaction coefficients can be understood as the fourth
244 corner and describes how environmental response across taxa varies as traits
245 vary. The size of coefficients is a measure of importance and are interpreted
246 as the amount by which a unit (1 sd) change in the trait variable changes the
247 slope of the relationship between abundance and a given environmental
248 variable. To estimate these coefficients, we used a LASSO-penalised
249 negative binomial regression (R package “mvabund”, Wang et al. 2012). The
250 LASSO penalty simplifies interpretation because it automatically does model
251 selection by setting to zero any interaction coefficients that do not help reduce
252 BIC. A species effect is included in the model (i.e. a different intercept term for
253 each species), so that traits are used to explain patterns in relative abundance
254 across taxa not patterns in absolute abundance. Pseudo-R² is calculated as
255 the R² of the predicted against the observed abundance values for each
256 species at each site.

257

258 **Effect trait analysis:** To determine which traits influenced functional
259 contribution of each species, we ran separate linear models with either
260 visitation or per capita efficiency as response variables. Species traits were
261 predictors. The best model based on AICc was selected. When differences
262 between the best models were less than 2 we selected the simpler model.
263 The analysis for efficiency was done at the species group level (see above:
264 pollination function section). To obtain traits at the species-group level, we
265 calculated the mean values over species belonging to the same group,

266 weighted by the species mean abundance within the group. For categorical
267 variables we chose the dominant level, again weighted by species
268 abundance. This way, we assure that while species within a functional group
269 are selected to be functionally similar, the average traits used reflects species
270 composition.

271 All residuals were visually inspected to validate model assumptions. All
272 statistical analyses were performed in R (version 3.0.3, <www.r-project.org>).

273 **Results**

274 **Response traits:** Overall, we did not find a strong correlation between any
275 ecological traits and the environmental variables analyzed despite finding a
276 general response of species abundance to change with one or more land use
277 variables (watermelon: estimate of percentage open habitat at 300m = 0.12;
278 blueberry: estimate of percentage agricultural habitat at 300 m= -0.26 and at
279 1500m = -0.12; cranberry: estimate percentage agricultural percentage habitat
280 at 1500m = -0.23. Supplementary Table A3). Traits do not modify these
281 slopes in most instances, and despite some traits exhibiting weak responses
282 to land use in some cases, these responses were not consistent across crops
283 (Fig 1). For watermelon (overall pseudo-R² = 0.54), small bees and parasites
284 tended to decline with increasing percentage of agriculture at 300m radius
285 (Interaction estimate of % agriculture at 300m with body size = 0.19, Fig 1D;
286 and with Parasitism = 0.10) and parasites also declined with increasing open
287 areas at 1500m radius (interaction estimate = 0.13). For blueberry (overall
288 pseudo-R² = 0.22) short-tongued species increased with increasing

289 agriculture at 1500m (interaction estimate = -0.30). In cranberry (overall
290 pseudo-R² = 0.59), bees nesting in wood and generalist bees tended to
291 increase with increasing open areas at 300 m (interaction estimate = 0.14 and
292 0.11 respectively) and bees nesting in soil and bigger bees tend to increase
293 with increasing open areas at 1500 m buffer (interaction estimate = 0.14). A
294 complete list of all comparisons is presented in Supplementary material (Table
295 A3).

296 **Effect traits:** As for response traits, no traits were highly predictive of either
297 visitation frequency or per visit efficiency across crops. For watermelon, the
298 best model for visitation frequency does not includes any trait. However, per
299 visit efficiency was positively correlated with body size and tongue length (R²
300 = 0.75, F_{2,9} = 17.07, p < 0.001, Fig 2A). For cranberry, visitation frequency
301 was positively related to cavity nesters (R² = 0.38, F_{4,36} = 7.1, p < 0.0001,
302 Fig 2B). This result was driven by *Bombus* species, which are the only cavity
303 nesters in this data set. In cranberry per visit efficiency was not related to any
304 trait. For blueberry, visitation frequency was positively related to diet
305 specialism (R² = 0.37, F_{1,20} = 13.5, p = 0.001, Fig 2C), while efficiency per
306 visit is positively related to tongue length (R² = 0.70, F_{1,5} = 14.9, p = 0.01,
307 Fig 2D). Model selection, can be found in Supplementary material (Table A4).

308 **Discussion:**

309 Identifying traits that characterize which species are more sensitive to land-
310 use change or those that are functionally important is complex. We found
311 some evidence for response and effect traits but they differed among crop

312 species as well as landscape variable used. Therefore, while some traits may
313 be important in some contexts, no traits were generalizable enough to be
314 used to predict how land-use change will influence the delivery of pollination
315 services across these systems. Further, the relationships identified were
316 weak. This does not negate the importance of traits for understanding which
317 mechanisms underlie species responses to land-use change or pollination
318 effectiveness, but it does suggest that traits commonly used for wild bees
319 might not be suitable for predicting which species will decline or how land-use
320 change will influence the delivery of ecosystem services. In fact, the trait-
321 based literature in general is characterized by weak and/or idiosyncratic
322 relationships between traits and either species responses and functional
323 effects (Tables 1 and 2).

324 Being able to identify strong response traits would be a key tool for
325 understanding extinction risk, and an asset for conservation planning.
326 However, characterizing extinction risk based on traits is challenging. Despite
327 some generalities that emerge across taxa, with rare species, big species,
328 specialists, and higher trophic levels being in general more sensitive to
329 disturbances (Fisher and Owens 2004), there is a large variation in the
330 response of the species with those traits (Fritz et al. 2009; Seguin et al. 2014).
331 Work specifically on native bees has found that traits such as specialization,
332 body size, and sociality may predict responses to land use (Table 1; Winfree
333 et al. 2009, Bommarco et al. 2010, Williams et al. 2010, Bartomeus *et al.*
334 2013b, Kremen and M’Gonigle 2015, De Palma et al. 2015, Carrié et al.
335 2016). However, studies often find contrasting results (Table 1). For example,

336 De Palma et al. (2015) analyzed over 70,000 wild bee records and found that
337 small species were most sensitive to agricultural land use, while others have
338 found that larger species are more sensitive to agricultural land use and/or
339 environmental change generally (Larsen et al. 2005; Bartomeus et al. 2013b),
340 and some have found little effect of body size (Williams et al. 2010). Here, we
341 found a weak trend for small species to be more sensitive to local agricultural
342 intensification in watermelon, but this trend disappears when land use is
343 measured at larger scales. Another trait, dietary specialization, is one of the
344 few traits that has been generally linked to increased species sensitivity to
345 environmental change (Table 1, Williams et al. 2010; Scheper et al. 2014; De
346 Palma et al. 2015), but here we found that floral specialist bees did not decline
347 with intensifying agriculture. If anything, one of the most abundant bee
348 species in the cranberry system (*Mellita americana*) is a specialist on
349 cranberry (*Vaccinium macrocarpon*). Specialist bees observed in crop
350 systems are likely to be specialized on the crop plant family as was the case
351 in our data (e.g. *Mellita americana* in cranberry, but also *Habropoda sp.* and
352 *Andrena bradleyi* in blueberry and *Peponapis pruinosa* in watermelon). We
353 would expect different responses from study designs that include natural
354 habitat and a larger range of specialist host plants (Forrest et al. 2015,
355 Bartomeus and Winfree 2013).

356

357 Alternatively, the lack of strong trait-environment associations may be due to
358 the variables used to measure agricultural intensification being too coarse to
359 detect common responses. While finer-resolution studies will undoubtedly be

360 informative, they are unlikely to lead to a greater likelihood of predicting how
361 changes in biodiversity affects the delivery of ecosystem services if these
362 measures are difficult to quantify or are context dependent.

363 Effect traits have been even harder to identify for pollinators. The limited data
364 published on particular plants suggests insects with larger bodies tend to
365 deposit more pollen per flower visit, but this pollen was not well distributed on
366 the stigma (Table 2; Hoehn et al. 2008), or that the correlation between body
367 size and per visit pollination function is low (Larsen et al. 2005). Our study
368 supports the positive correlation between body size and per-visit pollen
369 deposition in both watermelon and blueberry (although note that tongue
370 length is correlated with body size in blueberry $r = 0.76$), but not for cranberry.
371 Hence, generality is difficult to achieve because a single pollinator trait, like
372 big body size, may not lead to high pollination function in all contexts. Rather it
373 seems likely that the most efficient trait will depend on the crop (Garibaldi et
374 al. 2015). Moreover, the total pollination provided by a pollinator species is the
375 product of visitation frequency and per capita efficiency (Kremen et al. 2005),
376 two processes that may be governed by different traits.

377 If generalizable response and effect traits can be found, the final step will be
378 to link response and effects to predict changes in ecosystem services. A
379 positive association between the response and effect traits (Naeem and
380 Wright 2003) such that species with the strongest response to environmental
381 change also had the strongest effect on function, indicates the land-use
382 change has the potential for dramatic effects on ecosystem function. Whether

383 response and effect traits are in general positively, negatively, or uncorrelated
384 is an important question that has not yet been answered (Larsen et al. 2005).
385 Despite the conceptual elegance of the response-effect trait framework, it is
386 only effective if it is predictive, and strong evidence for the generality of traits
387 has not yet been found. For example, even the very thorough and rigorously
388 analyzed study of response-effect relationships by Larsen et al. (2005) is
389 based on a non-significant weak relationship between pollinator per visit
390 efficiency and body size. Similarly, the marginal R^2 (i.e. variance explained by
391 fixed effects) of the best model including traits in the comprehensive analysis
392 done by De Palma et al. (2015) is lower than 0.1. Similarly, in our study, even
393 the strongest correlations found for watermelon, where big species are less
394 sensitive to local agricultural intensification and more efficient per visit, but not
395 more frequent flower visitors than smaller species are too weak to be useful
396 for predictive purposes.

397 Predictive response and/or effect traits are often assumed in the larger
398 literature as well. For example, recent re-evaluations of community stability in
399 food webs shows that using body size as proxy of extinction risk changes the
400 outcome of the stability simulations (Brose et al. 2016). However, the
401 assumption that body size is a good predictor of extinction risk is not directly
402 validated. Given the correlation showing that bigger species are more
403 sensitive is usually weak (Fisher and Owens 2004), these kind of approaches
404 could produce misleading outcomes.

405 Currently trait data may be too coarse to reveal ubiquitous response and

406 effect traits for four reasons. First, some traits may simply reflect identity of
407 genera or higher taxonomic groups. For example, some bumble bee species
408 in our three systems (especially *B. impatiens*) are common, functionally
409 dominant, and robust to extinction (Cariveau et al. 2013, Winfree et al. 2015).
410 Some of the important response and effect traits that we found, such as cavity
411 nesting and body size, may simply be proxies for bumble bees. Bumble bee
412 species also share other traits (e.g. sociality) that are commonly used in trait
413 analyses. Therefore, studies that don't include phylogenetic correlations may
414 be simply characterizing the general relationship between disturbance and the
415 functionally dominant taxa. As there is a great variability in the responses to
416 disturbance among bumble bee species (Cameron et al. 2011; Bartomeus et
417 al. 2013b, Persson et al. 2015) this may also explain why some studies find
418 big species to be more sensitive to land-use change (Larsen et al. 2005) and
419 other studies find the opposite (Rader et al. 2014, this study for watermelon).
420 Second, traits may interact in complex ways and single traits may be not able
421 to capture responses and functional contributions across species (e.g.
422 Bommarco et al. 2010). Third, phenotypic variability within species, usually
423 ignored in trait-based approaches, may play a more important role than
424 previously though (Bolnik et al. 2011). Finally, the most important traits may
425 not have been studied. Response traits such as dispersal ability, fecundity,
426 and nest microclimate/soil type, and effect traits like floral visitation behavior
427 or hairiness (Stavert et al. 2016) may be better predictors than the traits we
428 have now. However, if these traits are not easy to measure across bee
429 species, they may be of little use. Traits databases that include an increasing

430 number of traits and agreed-upon measurement techniques similar to those
431 used in plant ecology (Kattge et al. 2011) but that are also open-access may
432 lead to significant advancements in functional trait ecology in wild bees.

433 There is a call to be more predictive in ecology (Petchey et al. 2015, Houlahan
434 et al. 2017). The use of traits to predict species responses and subsequent
435 changes in ecosystem services is a potentially powerful approach. This is
436 especially the case for organisms such as insects where species identification
437 is challenging and detailed species-level natural history information is lacking.
438 The ability to effectively use a trait framework is becoming controversial
439 because studies thus far have not clearly related specific traits to specific
440 threats or functions (Didham et al. 2016; Shipley et al. 2016). A growing
441 number of studies are working to address the complexity and increase the
442 predictability of this framework (e.g. Laughlin and Messier 2015). However,
443 until these approaches yield consistent patterns across systems, site-specific
444 species identity and monitoring may at present be the best measure for
445 predicting changes in ecosystem services as a result of land-use change. A
446 few dominant species often drive ecosystem functioning (Kleijn et al. 2015;
447 Winfree et al. 2015). Identifying the sensitivity of the functionally dominant
448 species may be the best proxy thus far for predicting effects of species loss in
449 ecosystem function.

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454 **Data Accessibility:** All data and code used in this manuscript is accessible in

455 github (https://github.com/ibartomeus/RE_traits) and will be archived on

456 acceptance in figshare.

457 **References:**

- 458 Bartomeus, I. and Winfree, R. 2013. Pollinator declines: reconciling scales
459 and implications for ecosystem services. - *F1000Res.* 2: 146.
- 460 Bartomeus, I. et al. 2013a. Data from: Historical changes in northeastern US
461 bee pollinators related to shared ecological traits.
- 462 Bartomeus, I. et al. 2013b. Historical changes in northeastern US bee
463 pollinators related to shared ecological traits. - *Proc. Natl. Acad. Sci. U. S.*
464 *A.* 110: 4656–4660.
- 465 Beaulieu, J. M. et al. 2007. Genome size evolution in relation to leaf strategy
466 and metabolic rates revisited. - *Ann. Bot.* 99: 495–505.
- 467 Benjamin, F. E. et al. 2014. Pollinator body size mediates the scale at which
468 land use drives crop pollination services. - *J. Appl. Ecol.* 51: 440–449.
- 469 Blüthgen, N. et al. 2008. What do interaction network metrics tell us about
470 specialization and biological traits? - *Ecology* 89: 3387–3399.
- 471 Bolnick, D. I. et al. 2011. Why intraspecific trait variation matters in community
472 ecology. - *Trends Ecol. Evol.* 26: 183–192.
- 473 Bommarco, R. et al. 2010. Dispersal capacity and diet breadth modify the
474 response of wild bees to habitat loss. - *Proc. Biol. Sci.* 277: 2075–2082.
- 475 Brose, U. et al. 2016. Predicting the consequences of species loss using size-
476 structured biodiversity approaches. - *Biol. Rev. Camb. Philos. Soc.* in
477 press.
- 478 Brown, A. M. et al. 2014. The fourth-corner solution – using predictive models
479 to understand how species traits interact with the environment. - *Methods*
480 *Ecol. Evol.* 5: 344–352.
- 481 Cadotte, M. W. et al. 2011. Beyond species: functional diversity and the
482 maintenance of ecological processes and services. - *J. Appl. Ecol.* 48:
483 1079–1087.
- 484 Cameron, S. A. et al. 2011. Patterns of widespread decline in North American
485 bumble bees. - *Proc. Natl. Acad. Sci. U. S. A.* 108: 662–667.
- 486 Cane, J. 1987. Estimation of bee size using intertegular span (Apoidea). - *J.*
487 *Kans. Entomol. Soc.* 60: 145–147.
- 488 Cane, J. et al. 2006. Multiple response of desert bee guild (Hymenoptera:
489 Apiformes) to urban habitat fragmentation. - *Ecol. Appl.* 16: 632–644.
- 490 Cardinale, B. J. et al. 2012. Biodiversity loss and its impact on humanity. -
491 *Nature* 486: 59–67.

- 492 Cariveau, D. P. et al. 2013. Response diversity to land use occurs but does
493 not consistently stabilise ecosystem services provided by native
494 pollinators. - *Ecol. Lett.* 16: 903–911.
- 495 Cariveau, D. P. et al. 2016. The allometry of bee proboscis length and its uses
496 in ecology. - *PLoS One* 11: e0151482.
- 497 Carrié, R. et al. 2017. Relationships among ecological traits of wild bee
498 communities along gradients of habitat amount and fragmentation. -
499 *Ecography* 40: 85–97.
- 500 Chamberlain, S. A. and Szöcs, E. 2013. taxize: taxonomic search and
501 retrieval in R. - *F1000Res.* 2: 191.
- 502 Deguines, N. et al. 2014. Large-scale trade-off between agricultural
503 intensification and crop pollination services. - *Front. Ecol. Environ.* 12:
504 212–217.
- 505 De Palma, A. et al. 2015. Ecological traits affect the sensitivity of bees to land-
506 use pressures in European agricultural landscapes. - *J. Appl. Ecol.* 52:
507 1567–1577.
- 508 Devictor, V. et al. 2008. Distribution of specialist and generalist species along
509 spatial gradients of habitat disturbance and fragmentation. - *Oikos* 117:
510 507–514.
- 511 Díaz, S. et al. 2016. The global spectrum of plant form and function. - *Nature*
512 529: 167–171.
- 513 Didham, R. K. et al. 2016. Circle the bandwagons – Challenges mount against
514 the theoretical foundations of applied functional trait and ecosystem
515 service research. - *Insect Conserv. Divers.* 9: 1–3.
- 516 Ellis, E. C. et al. 2010. Anthropogenic transformation of the biomes, 1700 to
517 2000. - *Glob. Ecol. Biogeogr.* 19: 589–606.
- 518 Fisher, D. O. and Owens, I. P. F. 2004. The comparative method in
519 conservation biology. - *Trends Ecol. Evol.* 19: 391–398.
- 520 Flynn, D. F. B. et al. 2009. Loss of functional diversity under land use
521 intensification across multiple taxa. - *Ecol. Lett.* 12: 22–33.
- 522 Forrest, J. R. K. et al. 2015. Contrasting patterns in species and functional-
523 trait diversity of bees in an agricultural landscape. - *J. Appl. Ecol.* 52: 706–
524 715.
- 525 Fritz, S. A. et al. 2009. Geographical variation in predictors of mammalian
526 extinction risk: big is bad, but only in the tropics. - *Ecol. Lett.* 12: 538–549.
- 527 Gagic, V. et al. 2015. Functional identity and diversity of animals predict
528 ecosystem functioning better than species-based indices. - *Proceedings*
529 *of the Royal Society of London B: Biological Sciences* 282: 20142620.
- 530 Garibaldi, L. A. et al. 2011. Stability of pollination services decreases with
531 isolation from natural areas despite honey bee visits. - *Ecol. Lett.* 14:
532 1062–1072.
- 533 Garibaldi, L. A. et al. 2015. Trait matching of flower visitors and crops predicts
534 fruit set better than trait diversity. - *J. Appl. Ecol.* 52: 1436–1444.
- 535 Greenleaf, S. S. et al. 2007. Bee foraging ranges and their relationship to
536 body size. - *Oecologia* 153: 589–596.
- 537 Hoehn, P. et al. 2008. Functional group diversity of bee pollinators increases
538 crop yield. - *Proc. Biol. Sci.* 275: 2283–2291.

- 539 Houlahan et al. 2017. The priority of prediction in ecological understanding. -
540 *Oikos* 126: 1-7.
- 541 Jauker B. et al. 2013. Linking life history traits to pollinator loss in fragmented
542 calcareous grasslands. - *Landscape Ecol.* 28:107-120.
- 543 Kattge, J. et al. 2011. TRY--a global database of plant traits. - *Glob. Chang.*
544 *Biol.* 17: 2905–2935.
- 545 Kleijn, D. et al. 2006. Mixed biodiversity benefits of agri-environment schemes
546 in five European countries: Biodiversity effects of European agri-
547 environment schemes. - *Ecol. Lett.* 9: 243–254.
- 548 Kleijn, D. et al. 2015. Delivery of crop pollination services is an insufficient
549 argument for wild pollinator conservation. - *Nat. Commun.* 6: 7414.
- 550 Klein, A.-M. et al. 2007. Importance of pollinators in changing landscapes for
551 world crops. - *Proc. Biol. Sci.* 274: 303–313.
- 552 Klein, A.-M. et al. 2008. Advances in pollination ecology from tropical
553 plantation crops. - *Ecology* 89: 935–943.
- 554 Kremen, C. 2005. Managing ecosystem services: what do we need to know
555 about their ecology? - *Ecol. Lett.* 8: 468–479.
- 556 Kremen, C. and M'Gonigle, L. K. 2015. Small-scale restoration in intensive
557 agricultural landscapes supports more specialized and less mobile
558 pollinator species. - *J. Appl. Ecol.* 52: 602–610.
- 559 Larsen, T. H. et al. 2005. Extinction order and altered community structure
560 rapidly disrupt ecosystem functioning. - *Ecol. Lett.* 8: 538–547.
- 561 Laughlin, D. C. and Messier, J. 2015. Fitness of multidimensional phenotypes
562 in dynamic adaptive landscapes. - *Trends Ecol. Evol.* 30: 487–496.
- 563 Lavorel, S. and Garnier, E. 2002. Predicting changes in community
564 composition and ecosystem functioning from plant traits: revisiting the
565 Holy Grail. - *Funct. Ecol.* 16: 545–556.
- 566 Legendre, P. et al. 1997. Relating behavior to habitat: Solutions to the fourth-
567 corner problem. - *Ecology* 78: 547–562.
- 568 McGill, B. J. et al. 2006. Rebuilding community ecology from functional traits. -
569 *Trends Ecol. Evol.* 21: 178–185.
- 570 Naeem, S. and Wright, J. P. 2003. Disentangling biodiversity effects on
571 ecosystem functioning: deriving solutions to a seemingly insurmountable
572 problem. - *Ecol. Lett.* 6: 567–579.
- 573 Nipperess, D. A. and Matsen 4, F. A. 2013. The mean and variance of
574 phylogenetic diversity under rarefaction. - *Methods Ecol. Evol.* 4: 566–
575 572.
- 576 Öckinger, E. et al. 2010. Life-history traits predict species responses to habitat
577 area and isolation: a cross-continental synthesis. - *Ecol. Lett.* 13: 969–
578 979.
- 579 Pereira, H. M. et al. 2010. Scenarios for global biodiversity in the 21st century.
580 - *Science* 330: 1496–1501.
- 581 Persson, A. S. et al. 2015. Bumble bees show trait-dependent vulnerability to
582 landscape simplification. - *Biodivers. Conserv.* 24: 3469–3489.
- 583 Petchey, O. L. et al. 2015. The ecological forecast horizon, and examples of
584 its uses and determinants. - *Ecol. Lett.* 18: 597–611.

- 585 Powney, G. D. et al. 2014. Can trait-based analyses of changes in species
586 distribution be transferred to new geographic areas? - *Glob. Ecol.*
587 *Biogeogr.* 23: 1009–1018.
- 588 Rader, R. et al. 2014. The winners and losers of land use intensification:
589 pollinator community disassembly is non-random and alters functional
590 diversity. - *Divers. Distrib.* 20: 908–917.
- 591 Rader, R. et al. 2016. Non-bee insects are important contributors to global
592 crop pollination. - *Proc. Natl. Acad. Sci. U. S. A.* 113: 146–151.
- 593 Scheper, J. et al. 2014. Museum specimens reveal loss of pollen host plants
594 as key factor driving wild bee decline in The Netherlands. - *Proc. Natl.*
595 *Acad. Sci. U. S. A.* 111: 17552–17557.
- 596 Schleuning, M. et al. 2015. Predicting ecosystem functions from biodiversity
597 and mutualistic networks: an extension of trait-based concepts to plant-
598 animal interactions. - *Ecography* 38: 380–392.
- 599 Schwartz, M. W. et al. 2000. Linking biodiversity to ecosystem function:
600 implications for conservation ecology. - *Oecologia* 122: 297–305.
- 601 Séguin, A. et al. 2014. Body size as a predictor of species loss effect on
602 ecosystem functioning. - *Sci. Rep.* 4: 4616.
- 603 Shipley, B. et al. 2016. Reinforcing loose foundation stones in trait-based
604 plant ecology. - *Oecologia* 180: 923–931.
- 605 Stavert, J. R. et al. 2016. Hairiness: the missing link between pollinators and
606 pollination. - *PeerJ* 4: e2779.
- 607 Wang, Y. et al. 2012. mvabund - an R package for model-based analysis of
608 multivariate abundance data: The mvabund R package. - *Methods Ecol.*
609 *Evol.* 3: 471–474.
- 610 Webb 2000. Exploring the phylogenetic structure of ecological communities:
611 An example for rain forest trees. - *Am. Nat.* 156: 145.
- 612 Webb, C. O. et al. 2008. Phylocom: software for the analysis of phylogenetic
613 community structure and trait evolution. - *Bioinformatics* 24: 2098–2100.
- 614 Wikstrom, N. et al. 2001. Evolution of the angiosperms: calibrating the family
615 tree. - *Proceedings of the Royal Society B: Biological Sciences* 268:
616 2211–2220.
- 617 Williams, N. M. et al. 2010. Ecological and life-history traits predict bee
618 species responses to environmental disturbances. - *Biol. Conserv.* 143:
619 2280–2291.
- 620 Winfree, R. et al. 2009. A meta-analysis of bees' responses to anthropogenic
621 disturbance. - *Ecology* 90: 2068–2076.
- 622 Winfree, R. et al. 2014. Species abundance, not diet breadth, drives the
623 persistence of the most linked pollinators as plant-pollinator networks
624 disassemble. - *Am. Nat.* 183: 600–611.
- 625 Winfree, R. et al. 2015. Abundance of common species, not species richness,
626 drives delivery of a real-world ecosystem service. - *Ecol. Lett.* 18: 626–
627 635.
- 628

629

630 Tables and Figures

631 **Table 1:** Summary of some recent studies identifying response traits and its
 632 relationship with environmental variables. Environmental variables have been
 633 grouped in two main categories because each study uses different metrics.
 634 Habitat loss (e.g. isolation, % natural habitat, habitat fragment size) and
 635 Agricultural intensification (e.g. natural vs agricultural, % agriculture). Only the
 636 direction of the response is indicated, as the different analysis makes any
 637 comparison of effect sizes meaningless. Note that in addition, most of this
 638 relationships are weakly predictive. See text for details.

| Trait | Environment | Relationship direction | | | Context dependent |
|------------------------------|------------------------------|------------------------|------------------|----------------------|-------------------|
| | | Positive | Neutral | Negative | |
| Body size | Habitat loss | 0 | 1 ⁹ | 2 ^{1,10} | 1 ² |
| | Agricultural intensification | 2 ^{7,11} | 2 ^{3,6} | 4 ^{1,4,5,8} | 0 |
| Diet specialization | Habitat loss | 1 ³ | 0 | 2 ^{1,9} | 1 ² |
| | Agricultural intensification | 0 | 1 ^{8,6} | 4 ^{3,7} | 0 |
| Sociality (social) | Habitat loss | 0 | 1 ² | 1 ³ | 1 ¹⁰ |
| | Agricultural intensification | 4 ^{3,4,8} | 2 ^{6,7} | 0 | 0 |
| Nest location (below-ground) | Habitat loss | 0 | 1 ³ | 0 | 0 |
| | Agricultural intensification | 0 | 2 ^{3,6} | 2 ^{8,11} | 1 ⁷ |

639 ¹Larsen *et al.* 2005, ²Bommarco *et al.* 2010, ³Williams *et al.* 2010, ⁴Rader *et al.* 2014,

640 ⁵Benjamin *et al.* 2014, ⁶Forrest *et al.* 2015, ⁷De Palma *et al.* 2015, ⁸Carrié *et al.* 2016, ⁹Cane

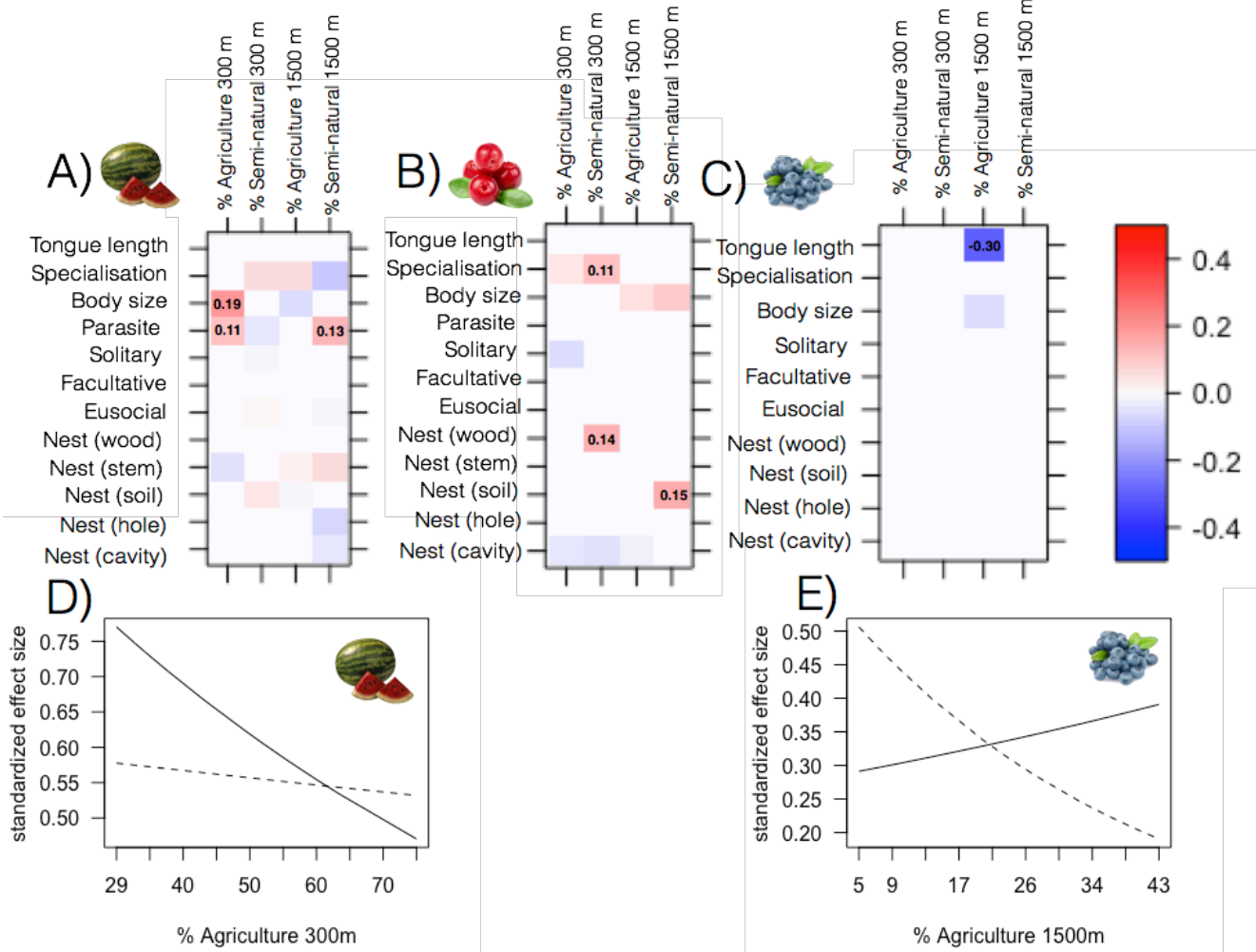
641 et al. 2006, ¹⁰Jauker et al. 2013, ¹¹Klein et al 2008.

642 **Table 2:** Summary of some recent studies identifying effect traits and its
 643 relationship with ecosystem functioning. Only body size is included, as other
 644 traits are rarely measured (but see Stavert et al 2016 for hairiness). Only the
 645 direction of the response is indicated, as the different analysis makes any
 646 comparison of effect sizes meaningless. Note that in addition, most of these
 647 relationships are weakly predictive. See text for details.

| Trait | Function | Relationship direction | | | Context dependent |
|-----------|-------------------|------------------------|----------------|----------------|-------------------|
| | | Positive | Neutral | Negative | |
| Body size | Pollen deposition | 2 ^{1,2} | 1 ³ | 0 | 0 |
| | Visitation rate | 0 | 0 | 1 ² | 0 |
| | Fruit set | 0 | 0 | 0 | 2 ^{4,5} |

648 ¹Larsen *et al.* 2010, ²Hohen et al. 2008, ³Stavert et al. 2016, ⁴Garibaldi et al. 2015, ⁵Gagic et
 649 al. 2015

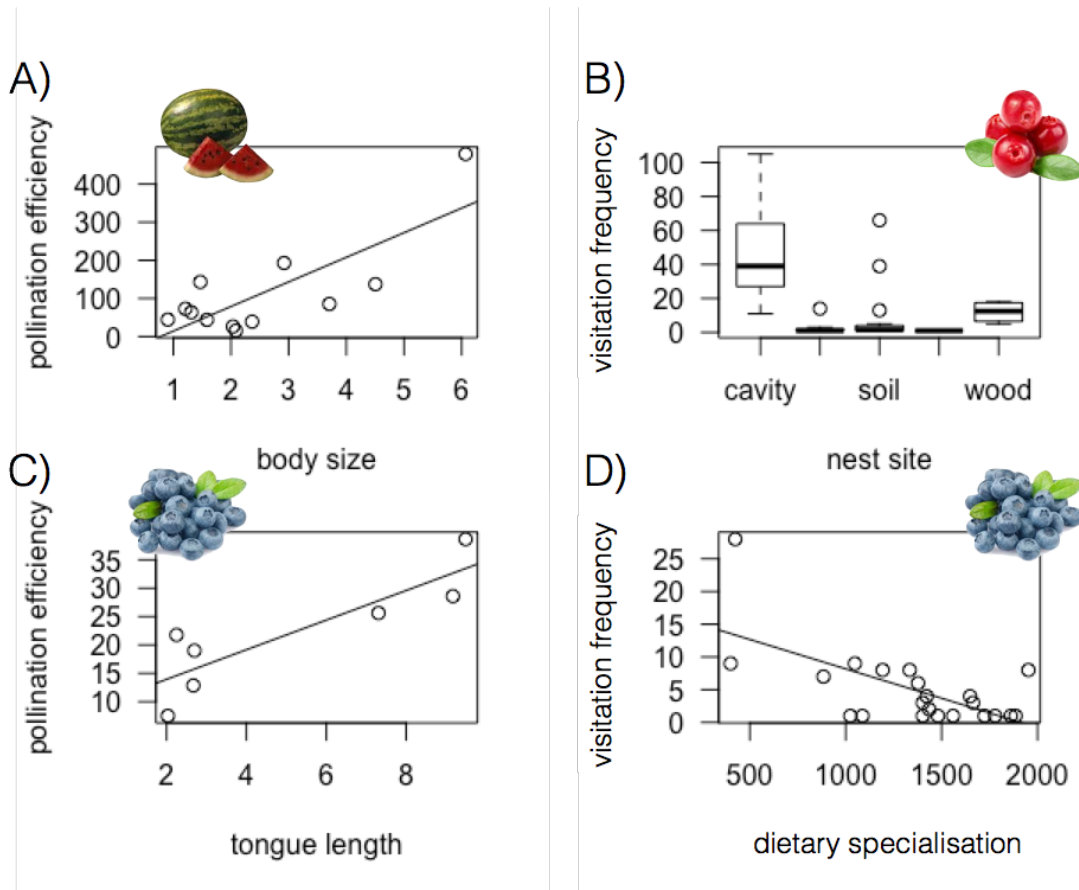
650 **Fig. 1: Relationships between traits and environmental variables** for A)
 651 watermelon, B) blueberry and c) cranberry. Positive estimates are in red and
 652 negative estimates in blue. Note that the LASSO penalty has set many
 653 estimates to zero. D) Detail of the two stronger interactions between body size
 654 and percentage of agriculture at 300 meter radii for watermelon and tongue
 655 length and % of agriculture at 1500 meter for blueberry. The solid line is the
 656 prediction for the 25 percentile of body size and tongue length, while the
 657 dashed line is the prediction for the 75% of body size and tongue length for
 658 watermelon and blueberry respectively.



659

660

661 **Fig. 2: Multipanel plot showing the relationships between species traits**
662 **and pollination function**, which is decomposed into efficiency (pollen
663 deposited per flower visit) and frequency of flower visits. A) watermelon, B)
664 cranberry, C-D) blueberry.



665

666

667 **Supplementary material**

668 **Text A1:** Datasets used for calculating dietary specialization: Six datasets

669 were used to create the phylogenetic distance index. All data were collected in

670 the region of the crop study. Specimens were collected using a hand net and

671 the bee species and plant species were recorded. This resulted in a total of

672 18,733 bee x plant interactions for species that were also in the crop dataset.

673 The number of species, sites, and years of collection are as follows: 1) Pine

674 barrens in 2003: 280 bee x plant interactions. Habitat types were extensive

675 pine-oak forest (14 sites), forest fragments (14 sites), suburban back yards (7

676 sites), and agricultural field borders (5 sites) in New Jersey (Winfree et al.

677 2007). Bees were collected in temporally stratified sampling rounds between

678 April and September. 2) NJPA: 3906 bee x plant interactions. Data collected

679 on watermelon field margins at a total of 20 sites. Farm types included small-

680 scale mixed farming, both crops and field margins, both organic and low-

681 pesticide-input conventional. All bees were collected in three temporally

682 stratified sampling rounds in July, in each of 3 years. 3) NFWF 3906 bee x

683 plant interactions. Habitat types were old fields. Bees were collected in May

684 through Sept at 25 sites for two years. *Lasioglossum* species were not

685 included for this dataset due to recent changes in its taxonomy. 4) NSF 2006

686 666 bee x plant interactions. Habitat types were deciduous forest fragments

687 (13 sites), and suburban / urban yards (3 sites) and sites with extensive

688 forests with diverse wildflower communities (4 sites). All bees were collected

689 in sampling rounds between April and early June. 5) CIG 4600 bee x plant

690 interactions. Sites were comprised of old fields as well as pollinator

691 enhancement sites. Bees using were collected using a hand net from a total of
 692 a total 18 sites in 2011-2013. For each bee specimen, the plant species was
 693 recorded. 6) Cape May 5858 bee x plant interactions. This study included only
 694 one site. The habitat was an old field that had been planted in 20 species of
 695 native perennial plants. Sampling took place over 3 years in sampling rounds
 696 that occurred in May through September.

697 Winfree, R. Griswold, T. and Kremen, C. (2007). Effect of human disturbance
 698 on bee communities in a forested ecosystem. *Conservation Biology*. 21: 213-
 699 223.

700

701 **Table A1:** Range of variation in agricultural and semi-natural land cover for
 702 three crop systems

| Crop | Percent Agriculture | | | | Percent Natural, Open | | | |
|------------|---------------------|-----|---------------|-----|-----------------------|-----|---------------|-----|
| | Radius 300 m | | Radius 1500 m | | Radius 300 m | | Radius 1500 m | |
| | Min | Max | Min | Max | Min | Max | Min | Max |
| Watermelon | 29 | 75 | 5 | 42 | 0 | 20 | 1 | 15 |
| Blueberry | 37 | 100 | 13 | 81 | 0 | 16 | 0.5 | 9 |
| Cranberry | 37 | 99 | 5 | 48 | 0 | 38 | 4 | 13 |

703

704 **Table A2:** Equivalencies between species and groups used for single visit
 705 data.

| Crop | Species | Single Visit Group | Percentage |
|------|---------|--------------------|------------|
|------|---------|--------------------|------------|

| | | | within group |
|-----------|---------------------------------|------------|-------------------------|
| Blueberry | <i>Andrena_banksi</i> | MED_AND | 0.00 |
| Blueberry | <i>Andrena_barbara</i> | MED_AND | 0.00 |
| Blueberry | <i>Andrena_bradleyi</i> | MED_AND | 0.93 |
| Blueberry | <i>Andrena_carlini</i> | LG_AND | 0.12 |
| Blueberry | <i>Andrena_carolina</i> | LG_AND | 0.01 |
| Blueberry | <i>Andrena_cressonii</i> | LG_AND | 0.01 |
| Blueberry | <i>Andrena_fenningeri</i> | MED_AND | 0.01 |
| Blueberry | <i>Andrena_hilaris</i> | MED_AND | 0.00 |
| Blueberry | <i>Andrena_ilicis</i> | MED_AND | 0.00 |
| Blueberry | <i>Andrena_imitatrix</i> | MED_AND | 0.01 |
| Blueberry | <i>Andrena_mandibularis</i> | MED_AND | 0.00 |
| Blueberry | <i>Andrena_morrisonella</i> | MED_AND | 0.01 |
| Blueberry | <i>Andrena_screpteroopsis</i> | MED_AND | 0.00 |
| Blueberry | <i>Andrena_vicina</i> | LG_AND | 0.86 |
| Blueberry | <i>Augochlora_pura</i> | Green | 0.56 |
| Blueberry | <i>Augochlorella_aurata</i> | Green | 0.44 |
| Blueberry | <i>Bombus_bimaculatus</i> | Bom_Q | 0.24 |
| Blueberry | <i>Bombus_griseocollis</i> | Bom_Q | 0.49 |
| Blueberry | <i>Bombus_impatiens</i> | Bom_Q | 0.18 |
| Blueberry | <i>Bombus_perplexus</i> | Bom_Q | 0.08 |
| Blueberry | <i>Ceratina_calcarata/dupla</i> | Dialictus | 0.22 |
| Blueberry | <i>Colletes_inaequalis</i> | Coll | 0.26 |
| Blueberry | <i>Colletes_thoracicus</i> | Coll | 0.09 |
| Blueberry | <i>Colletes_validus</i> | Coll | 0.65 |
| Blueberry | <i>Habropoda_laboriosa</i> | HAB | 1.00 |
| Blueberry | <i>Lasioglossum_acuminatum</i> | Small-dark | 0.02 |
| Blueberry | <i>Lasioglossum_coeruleum</i> | Small-dark | 0.02 |
| Blueberry | <i>Lasioglossum_fuscipenne</i> | Small-dark | 0.02 |
| Blueberry | <i>Lasioglossum_leucocomum</i> | Small-dark | 0.02 |
| Blueberry | <i>Lasioglossum_oblongum</i> | Small-dark | 0.10 |
| Blueberry | <i>Lasioglossum_pilosum</i> | Small-dark | 0.15 |
| Blueberry | <i>Lasioglossum_versatum</i> | Small-dark | 0.02 |
| Blueberry | <i>Lasioglossum_weemsi</i> | Small-dark | 0.07 |
| Blueberry | <i>Lasioglossum_zephyrum</i> | Small-dark | 0.05 |
| Blueberry | <i>Nomada_luteola</i> | Small-dark | 0.02 |
| Blueberry | <i>Osmia_taurus</i> | Small-dark | 0.02 |
| Blueberry | <i>Osmia_cornifrons</i> | Small-dark | 0.02 |
| Blueberry | <i>Osmia_pumila</i> | Small-dark | 0.07 |
| Blueberry | <i>Sphecodes_aroniae</i> | Small-dark | 0.02 |
| Blueberry | <i>Sphecodes_stygius</i> | Small-dark | 0.02 |
| Blueberry | <i>Xylocopa_virginica</i> | XYL | 1.00 |
| Cranberry | <i>Agapostemon_splendens</i> | Green | 0.02 |
| Cranberry | <i>Andrena_cressonii</i> | Med_and | 0.01 |

| | | | |
|-----------|------------------------------------|---------------------|------|
| Cranberry | <i>Andrena_imitatrix</i> | Med_and | 0.01 |
| Cranberry | <i>Andrena_morrisonella</i> | Med_and | 0.01 |
| Cranberry | <i>Andrena_spiraeana</i> | Med_and | 0.01 |
| Cranberry | <i>Andrena_vicina</i> | Med_and | 0.02 |
| Cranberry | <i>Augochlora_pura</i> | Green | 0.16 |
| Cranberry | <i>Augochlorella_aurata</i> | Green | 0.63 |
| Cranberry | <i>Augochloropsis_metallica</i> | Green | 0.16 |
| Cranberry | <i>Augochloropsis_sumptuosa</i> | Green | 0.03 |
| Cranberry | <i>Bombus_bimaculatus</i> | Bombus_bimaculatus | 1.00 |
| Cranberry | <i>Bombus_citrinus</i> | Bombus_spp | 0.50 |
| Cranberry | <i>Bombus_griseocollis</i> | Bombus_griseocollis | 1.00 |
| Cranberry | <i>Bombus_impatiens</i> | Bombus_impatiens | 1.00 |
| Cranberry | <i>Bombus_perplexus</i> | Bom_pervag | 0.85 |
| Cranberry | <i>Bombus_sandersoni</i> | Bombus_spp | 0.50 |
| Cranberry | <i>Bombus_vagans</i> | Bom_pervag | 0.15 |
| Cranberry | <i>Ceratina_calcarata/dupla</i> | Small_black | 0.12 |
| Cranberry | <i>Coelioxys_immaculata</i> | Megachile | 0.02 |
| Cranberry | <i>Coelioxys_porterae</i> | Megachile | 0.04 |
| Cranberry | <i>Coelioxys_sayi</i> | Megachile | 0.02 |
| Cranberry | <i>Colletes_consors</i> | Megachile | 0.02 |
| Cranberry | <i>Halictus_rubicundus</i> | Small_black | 0.02 |
| Cranberry | <i>Heriades_carinatus</i> | Osmia | 0.08 |
| Cranberry | <i>Hoplitis_truncata</i> | Osmia | 0.12 |
| Cranberry | <i>Hylaeus_affinis</i> | Small_black | 0.08 |
| Cranberry | <i>Lasioglossum_apopkense</i> | Small_black | 0.01 |
| Cranberry | <i>Lasioglossum_coeruleum</i> | Small_black | 0.01 |
| Cranberry | <i>Lasioglossum_creberrimum</i> | Small_black | 0.02 |
| Cranberry | <i>Lasioglossum_fuscipenne</i> | Small_black | 0.05 |
| Cranberry | <i>Lasioglossum_georgeickworti</i> | Small_black | 0.05 |
| Cranberry | <i>Lasioglossum_lineatum</i> | Small_black | 0.01 |
| Cranberry | <i>Lasioglossum_oblongum</i> | Small_black | 0.15 |
| Cranberry | <i>Lasioglossum_pilosum</i> | Small_black | 0.02 |
| Cranberry | <i>Lasioglossum_planatum</i> | Small_black | 0.02 |
| Cranberry | <i>Lasioglossum_subviridatum</i> | Small_black | 0.17 |
| Cranberry | <i>Lasioglossum_trigeminum</i> | Small_black | 0.02 |
| Cranberry | <i>Lasioglossum_versatum</i> | Small_black | 0.04 |
| Cranberry | <i>Megachile_addenda</i> | Megachile | 0.22 |
| Cranberry | <i>Megachile_gemula</i> | Megachile | 0.32 |
| Cranberry | <i>Megachile_mendica</i> | Megachile | 0.28 |
| Cranberry | <i>Megachile_texana</i> | Megachile | 0.08 |
| Cranberry | <i>Melitta_americana</i> | Melitta | 0.95 |
| Cranberry | <i>Nomada_bella/lepida</i> | Small_black | 0.01 |
| Cranberry | <i>Nomada_pygmaea</i> | Small_black | 0.01 |
| Cranberry | <i>Nomada_rodecki</i> | Small_black | 0.07 |

| | | | |
|------------|-------------------------------------|-------------|------|
| Cranberry | <i>Osmia_ inspergens</i> | Osmia | 0.07 |
| Cranberry | <i>Osmia_ pumila</i> | Osmia | 0.13 |
| Cranberry | <i>Osmia_ virga</i> | Osmia | 0.60 |
| Cranberry | <i>Panurginus_ atramontensis</i> | Small_black | 0.09 |
| Cranberry | <i>Sphecodes_ aroniae</i> | Small_black | 0.03 |
| Cranberry | <i>Sphecodes_ fattigi</i> | Small_black | 0.01 |
| Cranberry | <i>Xylocopa_ virginica</i> | XYL | 1.00 |
| Watermelon | <i>Agapostemon_ sericeus</i> | Large_Green | 0.25 |
| Watermelon | <i>Agapostemon_ texanus</i> | Large_Green | 0.11 |
| Watermelon | <i>Agapostemon_ virescens</i> | Large_Green | 0.52 |
| Watermelon | <i>Anthidium_ oblongatum</i> | LDS | 0.03 |
| Watermelon | <i>Augochlora_ pura</i> | Small_Green | 0.80 |
| Watermelon | <i>Augochlorella_ aurata</i> | Small_Green | 0.20 |
| Watermelon | <i>Augochloropsis_ metallica</i> | Large_Green | 0.11 |
| Watermelon | <i>Bombus_ bimaculatus</i> | BOM | 0.01 |
| Watermelon | <i>Bombus_ fervidus</i> | BOM | 0.00 |
| Watermelon | <i>Bombus_ griseocollis</i> | BOM | 0.01 |
| Watermelon | <i>Bombus_ impatiens</i> | BOM | 0.98 |
| Watermelon | <i>Bombus_ perplexus</i> | BOM | 0.00 |
| Watermelon | <i>Bombus_ vagans</i> | BOM | 0.00 |
| Watermelon | <i>Calliopsis_ andreniformis</i> | SD | 0.03 |
| Watermelon | <i>Ceratina_ calcarata/dupla</i> | CER | 0.81 |
| Watermelon | <i>Ceratina_ strenua</i> | CER | 0.19 |
| Watermelon | <i>Halictus_ confusus</i> | HAL_MDS | 0.83 |
| Watermelon | <i>Halictus_ ligatus</i> | HAL_MDS | 0.16 |
| Watermelon | <i>Halictus_ parallelus</i> | LDS | 0.03 |
| Watermelon | <i>Halictus_ rubicundus</i> | LDS | 0.56 |
| Watermelon | <i>Hoplitis_ pilosifrons</i> | HAL_MDS | 0.00 |
| Watermelon | <i>Hoplitis_ producta</i> | HAL_MDS | 0.00 |
| Watermelon | <i>Hylaeus_ affinis</i> | TD | 0.01 |
| Watermelon | <i>Lasioglossum_ admirandum</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_ albipenne</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_ atwoodi</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_ bruneri</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_ callidum</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_ cattellae</i> | TD | 0.00 |
| Watermelon | <i>Lasioglossum_ cinctipes</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_ coreopsis</i> | TD | 0.00 |
| Watermelon | <i>Lasioglossum_ coriaceum</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_ cressonii</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_ ellisiae</i> | TD | 0.01 |
| Watermelon | <i>Lasioglossum_ ephialtum</i> | SD | 0.02 |
| Watermelon | <i>Lasioglossum_ georgeickworti</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_ gotham</i> | SD | 0.00 |

| | | | |
|------------|-----------------------------------|-----|------|
| Watermelon | <i>Lasioglossum_illinoense</i> | TD | 0.05 |
| Watermelon | <i>Lasioglossum_imitatum</i> | TD | 0.59 |
| Watermelon | <i>Lasioglossum_laevisissimum</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_leucomomum</i> | SD | 0.02 |
| Watermelon | <i>Lasioglossum_leucozonium</i> | LDS | 0.09 |
| Watermelon | <i>Lasioglossum_lineatulum</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_mitchelli</i> | TD | 0.17 |
| Watermelon | <i>Lasioglossum_nymphaearum</i> | SD | 0.04 |
| Watermelon | <i>Lasioglossum_oblongum</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_obscurum</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_paradmirandum</i> | TD | 0.04 |
| Watermelon | <i>Lasioglossum_pectinatum</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_pectorale</i> | SD | 0.02 |
| Watermelon | <i>Lasioglossum_pilosum</i> | SD | 0.28 |
| Watermelon | <i>Lasioglossum_planatum</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_platyparium</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_rozeni</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_smilacinae</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_subviridatum</i> | SD | 0.00 |
| Watermelon | <i>Lasioglossum_tegulare</i> | TD | 0.07 |
| Watermelon | <i>Lasioglossum_trigeminum</i> | SD | 0.02 |
| Watermelon | <i>Lasioglossum_truncatum</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_versatum</i> | SD | 0.44 |
| Watermelon | <i>Lasioglossum_viridatum</i> | SD | 0.01 |
| Watermelon | <i>Lasioglossum_weemsi</i> | TD | 0.05 |
| Watermelon | <i>Lasioglossum_zephyrum</i> | SD | 0.04 |
| Watermelon | <i>Megachile_brevis</i> | LDS | 0.06 |
| Watermelon | <i>Megachile_mendica</i> | LDS | 0.18 |
| Watermelon | <i>Megachile_rotundata</i> | LDS | 0.03 |
| Watermelon | <i>Megachile_sculpturalis</i> | LDS | 0.03 |
| Watermelon | <i>Melissodes_bimaculata</i> | MEL | 0.99 |
| Watermelon | <i>Melissodes_trinodis</i> | MEL | 0.00 |
| Watermelon | <i>Nomada_articulata</i> | TRI | 0.01 |
| Watermelon | <i>Peponapis_pruinosa</i> | PEP | 1.00 |
| Watermelon | <i>Ptilothrix_bombiformis</i> | MEL | 0.01 |
| Watermelon | <i>Triepeolus_lunatus</i> | TRI | 0.01 |
| Watermelon | <i>Triepeolus_remigatus</i> | TRI | 0.98 |
| Watermelon | <i>Xylocopa_virginica</i> | XYL | 1.00 |

706

707 **Table A3:** Response trait model estimates for all variables, including the
708 fourth corner interactions. Note that many coefficients are set to zero due to

709 the lasso penalty which acts as model selection.

| crop | variable | estimates |
|-------------|-------------------------------------|------------------|
| watermelon | Intercept | -0.576421649 |
| watermelon | sppAgapostemon_sericeus | -0.009279838 |
| watermelon | sppAgapostemon_texanus | -0.014124596 |
| watermelon | sppAgapostemon_virescens | 0 |
| watermelon | sppAugochlora_pura | 0.558159608 |
| watermelon | sppAugochlorella_aurata | 0.252162681 |
| watermelon | sppAugochloropsis_metallica | -0.102126726 |
| watermelon | sppBombus_bimaculatus | -0.068328543 |
| watermelon | sppBombus_griseocollis | -0.044189731 |
| watermelon | sppBombus_impatiens | 0.725048904 |
| watermelon | sppCalliopsis_andreniformis | -0.029828203 |
| watermelon | sppCeratina_calcarata_dupla_miqmaki | 0.280782801 |
| watermelon | sppCeratina_strenua | 0 |
| watermelon | sppHalictus_confusus | 0.18712962 |
| watermelon | sppHalictus_ligatus | 0.135271953 |
| watermelon | sppHalictus_parallelus | -0.096935391 |
| watermelon | sppHalictus_rubicundus | 0 |
| watermelon | sppLasioglossum_admirandum | -0.129706998 |
| watermelon | sppLasioglossum_albipenne | -0.118065531 |
| watermelon | sppLasioglossum_bruneri | 0 |
| watermelon | sppLasioglossum_coriaceum | -0.107890698 |
| watermelon | sppLasioglossum_cressonii | -0.071238004 |
| watermelon | sppLasioglossum_ephialtum | 0 |
| watermelon | sppLasioglossum_illinoense | 0.036021268 |
| watermelon | sppLasioglossum_imitatum | 0.55394426 |
| watermelon | sppLasioglossum_leucomum | 0.002812834 |

| | | |
|------------|-------------------------------|--------------|
| watermelon | sppLasioglossum_mitchelli | 0.283412671 |
| watermelon | sppLasioglossum_nymphaearum | 0 |
| watermelon | sppLasioglossum_oblongum | -0.118783972 |
| watermelon | sppLasioglossum_obscurum | -0.07245479 |
| watermelon | sppLasioglossum_paradmirandum | 0.133974695 |
| watermelon | sppLasioglossum_pectorale | 0.041004684 |
| watermelon | sppLasioglossum_pilosum | 0.337276238 |
| watermelon | sppLasioglossum_rozeni | -0.150752269 |
| watermelon | sppLasioglossum_smilacinae | -0.106925589 |
| watermelon | sppLasioglossum_tegulare | 0.069406914 |
| watermelon | sppLasioglossum_trigeminum | -0.021964911 |
| watermelon | sppLasioglossum_truncatum | 0 |
| watermelon | sppLasioglossum_versatum | 0.428595442 |
| watermelon | sppLasioglossum_weemsi | 0.203014094 |
| watermelon | sppLasioglossum_zephyrum | 0 |
| watermelon | sppMegachile_mendica | -0.040327507 |
| watermelon | sppMelissodes_bimaculata | 0.338962415 |
| watermelon | sppPeponapis_pruinosa | 0.298377021 |
| watermelon | sppTriepeolus_remigatus | 0.142618209 |
| watermelon | sppXylocopa_virginica | 0 |
| watermelon | agriculture_300 | -0.033515551 |
| watermelon | semi-natural_300 | 0.124754762 |
| watermelon | agriculture_1500 | -0.047821623 |
| watermelon | semi-natural_1500 | 0.083440715 |
| watermelon | agriculture_300.squ | -0.042700664 |
| watermelon | semi-natural_300.squ | -0.055003683 |
| watermelon | agriculture_1500.squ | -0.096777603 |
| watermelon | semi-natural_1500.squ | 0 |
| watermelon | Nest_placecavity | 0 |

| | | |
|------------|-------------------------------------|--------------|
| watermelon | Nest_placehole | 0 |
| watermelon | Nest_placesoil | 0 |
| watermelon | Nest_placestem | 0.198778917 |
| watermelon | Nest_placewood | 0 |
| watermelon | SocialityEusocial | 0 |
| watermelon | Socialityfac_social | 0 |
| watermelon | SocialitySolitary | 0 |
| watermelon | ParasiticYes | 0 |
| watermelon | body_size | -0.173420908 |
| watermelon | specialization | 0 |
| watermelon | tongue | 0 |
| watermelon | body_size.squ | 0 |
| watermelon | specialization.squ | 0 |
| watermelon | tongue.squ | 0 |
| watermelon | agriculture_300:Nest_placecavity | 0 |
| watermelon | agriculture_300:Nest_placehole | 0 |
| watermelon | agriculture_300:Nest_placesoil | 0 |
| watermelon | agriculture_300:Nest_placestem | -0.049978079 |
| watermelon | agriculture_300:Nest_placewood | 0 |
| watermelon | agriculture_300:SocialityEusocial | 0 |
| watermelon | agriculture_300:Socialityfac_social | 0 |
| watermelon | agriculture_300:SocialitySolitary | 0 |
| watermelon | agriculture_300:ParasiticYes | 0.108922029 |
| watermelon | agriculture_300:body_size | 0.192003954 |
| watermelon | agriculture_300:specialization | 0 |
| watermelon | agriculture_300:tongue | 0 |
| watermelon | semi-natural_300:Nest_placecavity | 0 |
| watermelon | semi-natural_300:Nest_placehole | 0 |
| watermelon | semi-natural_300:Nest_placesoil | 0.038326704 |

| | | |
|------------|---------------------------------------|--------------|
| watermelon | semi-natural_300:Nest_placestem | 0 |
| watermelon | semi-natural_300:Nest_placewood | 0 |
| watermelon | semi-natural_300:SocialityEusocial | 0.00808636 |
| watermelon | semi-natural_300:Socialityfac_social | 0 |
| watermelon | semi-natural_300:SocialitySolitary | -0.013199001 |
| watermelon | semi-natural_300:ParasiticYes | -0.039553688 |
| watermelon | semi-natural_300:body_size | 0 |
| watermelon | semi-natural_300:specialization | 0.061463166 |
| watermelon | semi-natural_300:tongue | 0 |
| watermelon | agriculture_1500:Nest_placecavity | 0 |
| watermelon | agriculture_1500:Nest_placehole | 0 |
| watermelon | agriculture_1500:Nest_placesoil | -0.005213896 |
| watermelon | agriculture_1500:Nest_placestem | 0.017326634 |
| watermelon | agriculture_1500:Nest_placewood | 0 |
| watermelon | agriculture_1500:SocialityEusocial | 0 |
| watermelon | agriculture_1500:Socialityfac_social | 0 |
| watermelon | agriculture_1500:SocialitySolitary | 0 |
| watermelon | agriculture_1500:ParasiticYes | 0 |
| watermelon | agriculture_1500:body_size | -0.063254087 |
| watermelon | agriculture_1500:specialization | 0.061183593 |
| watermelon | agriculture_1500:tongue | 0 |
| watermelon | semi-natural_1500:Nest_placecavity | -0.036150993 |
| watermelon | semi-natural_1500:Nest_placehole | -0.0710914 |
| watermelon | semi-natural_1500:Nest_placesoil | 0 |
| watermelon | semi-natural_1500:Nest_placestem | 0.056228857 |
| watermelon | semi-natural_1500:Nest_placewood | 0 |
| watermelon | semi-natural_1500:SocialityEusocial | -0.010382428 |
| watermelon | semi-natural_1500:Socialityfac_social | 0 |
| watermelon | semi-natural_1500:SocialitySolitary | 0 |

| | | |
|------------|----------------------------------|--------------|
| watermelon | semi-natural_1500:ParasiticYes | 0.134685819 |
| watermelon | semi-natural_1500:body_size | 0 |
| watermelon | semi-natural_1500:specialization | -0.099793932 |
| watermelon | semi-natural_1500:tongue | 0 |
| blueberry | Intercept | -1.105790242 |
| blueberry | sppAndrena_bradleyi | 0.42049863 |
| blueberry | sppAndrena_cressonii | 0 |
| blueberry | sppAndrena_fenningeri | 0 |
| blueberry | sppAndrena_morrisonella | 0 |
| blueberry | sppAndrena_vicina | 0.019815503 |
| blueberry | sppAugochlora_pura | 0 |
| blueberry | sppAugochlorella_aurata | 0 |
| blueberry | sppBombus_bimaculatus | 0 |
| blueberry | sppBombus_griseocollis | 0 |
| blueberry | sppBombus_impatiens | 0 |
| blueberry | sppBombus_perplexus | 0 |
| blueberry | sppColletes_inaequalis | 0.126105594 |
| blueberry | sppColletes_thoracicus | 0 |
| blueberry | sppColletes_validus | 0.083273954 |
| blueberry | sppHabropoda_laboriosa | 0 |
| blueberry | sppLasioglossum_leucocomum | 0 |
| blueberry | sppLasioglossum_oblongum | 0 |
| blueberry | sppLasioglossum_pilosum | 0 |
| blueberry | sppLasioglossum_versatum | 0 |
| blueberry | sppLasioglossum_weemsi | 0 |
| blueberry | sppOsmia_pumila | 0 |
| blueberry | sppXylocopa_virginica | 0 |
| blueberry | agriculture_300 | -0.26658714 |
| blueberry | semi-natural_300 | 0 |

| | | |
|-----------|-------------------------------------|--------------|
| blueberry | agriculture_1500 | -0.115548638 |
| blueberry | semi-natural_1500 | 0 |
| blueberry | agriculture_300.squ | 0.177220016 |
| blueberry | semi-natural_300.squ | 0 |
| blueberry | agriculture_1500.squ | 0 |
| blueberry | semi-natural_1500.squ | 0 |
| blueberry | Nest_placecavity | 0 |
| blueberry | Nest_placehole | 0 |
| blueberry | Nest_placesoil | 0 |
| blueberry | Nest_placewood | 0 |
| blueberry | SocialityEusocial | 0 |
| blueberry | Socialityfac_social | 0 |
| blueberry | SocialitySolitary | 0 |
| blueberry | body_size | 0.569257497 |
| blueberry | specialization | 0 |
| blueberry | tongue | 0 |
| blueberry | body_size.squ | 0 |
| blueberry | specialization.squ | 0.1567025 |
| blueberry | tongue.squ | 0 |
| blueberry | agriculture_300:Nest_placecavity | 0 |
| blueberry | agriculture_300:Nest_placehole | 0 |
| blueberry | agriculture_300:Nest_placesoil | 0 |
| blueberry | agriculture_300:Nest_placewood | 0 |
| blueberry | agriculture_300:SocialityEusocial | 0 |
| blueberry | agriculture_300:Socialityfac_social | 0 |
| blueberry | agriculture_300:SocialitySolitary | 0 |
| blueberry | agriculture_300:body_size | 0 |
| blueberry | agriculture_300:specialization | 0 |
| blueberry | agriculture_300:tongue | 0 |

| | | |
|-----------|---------------------------------------|--------------|
| blueberry | semi-natural_300:Nest_placecavity | 0 |
| blueberry | semi-natural_300:Nest_placehole | 0 |
| blueberry | semi-natural_300:Nest_placesoil | 0 |
| blueberry | semi-natural_300:Nest_placewood | 0 |
| blueberry | semi-natural_300:SocialityEusocial | 0 |
| blueberry | semi-natural_300:Socialityfac_social | 0 |
| blueberry | semi-natural_300:SocialitySolitary | 0 |
| blueberry | semi-natural_300:body_size | 0 |
| blueberry | semi-natural_300:specialization | 0 |
| blueberry | semi-natural_300:tongue | 0 |
| blueberry | agriculture_1500:Nest_placecavity | 0 |
| blueberry | agriculture_1500:Nest_placehole | 0 |
| blueberry | agriculture_1500:Nest_placesoil | 0 |
| blueberry | agriculture_1500:Nest_placewood | 0 |
| blueberry | agriculture_1500:SocialityEusocial | 0 |
| blueberry | agriculture_1500:Socialityfac_social | 0 |
| blueberry | agriculture_1500:SocialitySolitary | 0 |
| blueberry | agriculture_1500:body_size | -0.065424217 |
| blueberry | agriculture_1500:specialization | 0 |
| blueberry | agriculture_1500:tongue | -0.30068745 |
| blueberry | semi-natural_1500:Nest_placecavity | 0 |
| blueberry | semi-natural_1500:Nest_placehole | 0 |
| blueberry | semi-natural_1500:Nest_placesoil | 0 |
| blueberry | semi-natural_1500:Nest_placewood | 0 |
| blueberry | semi-natural_1500:SocialityEusocial | 0 |
| blueberry | semi-natural_1500:Socialityfac_social | 0 |
| blueberry | semi-natural_1500:SocialitySolitary | 0 |
| blueberry | semi-natural_1500:body_size | 0 |
| blueberry | semi-natural_1500:specialization | 0 |

| | | |
|-----------|-------------------------------------|--------------|
| blueberry | semi-natural_1500:tongue | 0 |
| cranberry | Intercept | -0.631992954 |
| cranberry | sppAgapostemon_splendens | 0 |
| cranberry | sppAndrena_imitatrix | -0.040613799 |
| cranberry | sppAndrena_vicina | 0 |
| cranberry | sppAugochlora_pura | 0 |
| cranberry | sppAugochlarella_aurata | 0.352482114 |
| cranberry | sppAugochloropsis_metallica | 0.236752572 |
| cranberry | sppAugochloropsis_sumptuosa | 0.121551536 |
| cranberry | sppBombus_bimaculatus | 0.080454703 |
| cranberry | sppBombus_griseocollis | 0.084130022 |
| cranberry | sppBombus_impatiens | 0.164000021 |
| cranberry | sppBombus_perplexus | 0 |
| cranberry | sppBombus_vagans | -0.086979307 |
| cranberry | sppCeratina_calcarata_dupla_miqmaki | 0 |
| cranberry | sppCoelioxys_immaculata | 0 |
| cranberry | sppCoelioxys_porterae | 0 |
| cranberry | sppHeriades_carinatus | 0 |
| cranberry | sppHoplitis_truncata | 0 |
| cranberry | sppHylaeus_affinis | 0.046287451 |
| cranberry | sppLasioglossum_apopkense | -0.015975843 |
| cranberry | sppLasioglossum_creberrimum | 0 |
| cranberry | sppLasioglossum_fuscipenne | 0 |
| cranberry | sppLasioglossum_georgeickworti | 0.103453894 |
| cranberry | sppLasioglossum_lineatulum | 0 |
| cranberry | sppLasioglossum_oblongum | 0.065553692 |
| cranberry | sppLasioglossum_pilosum | 0 |
| cranberry | sppLasioglossum_subviridatum | 0.103315494 |
| cranberry | sppLasioglossum_trigeminum | 0 |

| | | |
|-----------|-----------------------------|--------------|
| cranberry | sppLasioglossum_versatum | 0 |
| cranberry | sppMegachile_addenda | 0 |
| cranberry | sppMegachile_gemula | 0 |
| cranberry | sppMegachile_mendica | 0 |
| cranberry | sppMegachile_texana | 0 |
| cranberry | sppMelitta_americana | 0.371484288 |
| cranberry | sppNomada_rodecki | 0.28043488 |
| cranberry | sppOsmia_inspergens | 0 |
| cranberry | sppOsmia_pumila | 0 |
| cranberry | sppOsmia_virga | 0.24124345 |
| cranberry | sppPanurginus_atramontensis | 0 |
| cranberry | sppSphecodes_aroniae | 0 |
| cranberry | sppSphecodes_fattigi | 0 |
| cranberry | sppXylocopa_virginica | 0 |
| cranberry | agriculture_300 | 0 |
| cranberry | semi-natural_300 | 0 |
| cranberry | agriculture_1500 | -0.239273032 |
| cranberry | semi-natural_1500 | 0 |
| cranberry | agriculture_300.squ | 0 |
| cranberry | semi-natural_300.squ | 0 |
| cranberry | agriculture_1500.squ | 0 |
| cranberry | semi-natural_1500.squ | 0.161123069 |
| cranberry | Nest_placecavity | 0.788141301 |
| cranberry | Nest_placehole | 0 |
| cranberry | Nest_placesoil | 0 |
| cranberry | Nest_placestem | 0 |
| cranberry | Nest_placewood | 0.210443302 |
| cranberry | SocialityEusocial | 0.119359227 |
| cranberry | Socialityfac_social | 0 |

| | | |
|-----------|--------------------------------------|--------------|
| cranberry | SocialitySolitary | 0 |
| cranberry | ParasiticYes | -0.002756197 |
| cranberry | body_size | 0.360722617 |
| cranberry | specialization | 0 |
| cranberry | tongue | 0 |
| cranberry | body_size.squ | 0 |
| cranberry | specialization.squ | 0.278531163 |
| cranberry | tongue.squ | 0.236448892 |
| cranberry | agriculture_300:Nest_placecavity | -0.041561834 |
| cranberry | agriculture_300:Nest_placehole | 0 |
| cranberry | agriculture_300:Nest_placesoil | 0 |
| cranberry | agriculture_300:Nest_placestem | 0 |
| cranberry | agriculture_300:Nest_placewood | 0 |
| cranberry | agriculture_300:SocialityEusocial | 0 |
| cranberry | agriculture_300:Socialityfac_social | 0 |
| cranberry | agriculture_300:SocialitySolitary | -0.063994245 |
| cranberry | agriculture_300:ParasiticYes | 0 |
| cranberry | agriculture_300:body_size | 0 |
| cranberry | agriculture_300:specialization | 0.038303398 |
| cranberry | agriculture_300:tongue | 0 |
| cranberry | semi-natural_300:Nest_placecavity | -0.052194922 |
| cranberry | semi-natural_300:Nest_placehole | 0 |
| cranberry | semi-natural_300:Nest_placesoil | 0 |
| cranberry | semi-natural_300:Nest_placestem | 0 |
| cranberry | semi-natural_300:Nest_placewood | 0.139451554 |
| cranberry | semi-natural_300:SocialityEusocial | 0 |
| cranberry | semi-natural_300:Socialityfac_social | 0 |
| cranberry | semi-natural_300:SocialitySolitary | 0 |
| cranberry | semi-natural_300:ParasiticYes | 0 |

| | | |
|-----------|---------------------------------------|--------------|
| cranberry | semi-natural_300:body_size | 0 |
| cranberry | semi-natural_300:specialization | 0.111563737 |
| cranberry | semi-natural_300:tongue | 0 |
| cranberry | agriculture_1500:Nest_placecavity | -0.016970912 |
| cranberry | agriculture_1500:Nest_placehole | 0 |
| cranberry | agriculture_1500:Nest_placesoil | 0 |
| cranberry | agriculture_1500:Nest_placestem | 0 |
| cranberry | agriculture_1500:Nest_placewood | 0 |
| cranberry | agriculture_1500:SocialityEusocial | 0 |
| cranberry | agriculture_1500:Socialityfac_social | 0 |
| cranberry | agriculture_1500:SocialitySolitary | 0 |
| cranberry | agriculture_1500:ParasiticYes | 0 |
| cranberry | agriculture_1500:body_size | 0.053019912 |
| cranberry | agriculture_1500:specialization | 0 |
| cranberry | agriculture_1500:tongue | 0 |
| cranberry | semi-natural_1500:Nest_placecavity | 0 |
| cranberry | semi-natural_1500:Nest_placehole | 0 |
| cranberry | semi-natural_1500:Nest_placesoil | 0.150882962 |
| cranberry | semi-natural_1500:Nest_placestem | 0 |
| cranberry | semi-natural_1500:Nest_placewood | 0 |
| cranberry | semi-natural_1500:SocialityEusocial | 0 |
| cranberry | semi-natural_1500:Socialityfac_social | 0 |
| cranberry | semi-natural_1500:SocialitySolitary | 0 |
| cranberry | semi-natural_1500:ParasiticYes | 0 |
| cranberry | semi-natural_1500:body_size | 0.093262373 |
| cranberry | semi-natural_1500:specialization | 0 |
| cranberry | semi-natural_1500:tongue | 0 |

711 **Table A4:** Model selection procedure showing all models within 2 AICc

712 values.

| Crop | Measure | Model | Delta AICc |
|------------|----------------------|----------------------|------------|
| Watermelon | Visitation Frequency | ~ body size + tongue | 0 |
| Watermelon | Visitation Frequency | ~ tongue | 0.06 |
| Watermelon | Visitation Frequency | ~ . | 0.28 |
| Watermelon | Visitation Frequency | ~ body size | 1.99 |
| Watermelon | Pollen deposition | ~ body size + tongue | 0 |
| Cranberry | Visitation Frequency | ~ sociality + tongue | 0 |
| Cranberry | Visitation Frequency | ~ nest place | 1.43 |
| Cranberry | Visitation Frequency | ~ sociality | 1.56 |
| Cranberry | Pollen deposition | ~ . | 0 |
| Blueberry | Visitation Frequency | ~ specialization | 0 |
| Blueberry | Pollen deposition | ~ tongue | 0 |

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