SPATIOTEMPORAL DYNAMICS OF THE BACTERIAL MICROBIOTA ON LACUSTRINE

CLADOPHORA GLOMERATA (CHLOROPHYTA)[1]

Michael J. Braus[2], Linda E. Graham

Department of Botany, University of Wisconsin-Madison, 430 Lincoln Dr., Madison, WI 53707

and Thea L. Whitman

Department of Soil Science, University of Wisconsin-Madison, 1525 Observatory Dr., 53706

[2]Corresponding author: email braus@wisc.edu, phone 608-334-2827; fax 608-265-2595

Running title: *Cladophora* bacterial spatiotemporal dynamics

ABSTRACT

The branched periphytic green alga *Cladophora glomerata*, often abundant in nearshore waters of lakes and rivers worldwide, plays important ecosystem roles, some mediated by epibiotic microbiota that benefit from host-provided surface, organic C, and O_2. Previous microscopy and high throughput sequencing studies have indicated surprising epibiont taxonomic and functional diversity, but have not included adequate consideration of sample replication or the potential for spatial and temporal variation. Here we report the results of 16S rRNA amplicon-based phylum-to-genus taxonomic analysis of *Cladophora*-associated bacterial epibiota sampled in replicate from three microsites and at six times during the open-water season of 2014, from the same lake locale (Picnic Point, Lake Mendota, Dane Co., WI, USA) explored by high throughput sequencing studies in two previous years. Statistical methods were used to test null hypotheses that the bacterial community: 1) is homogeneous across microsites tested, and 2) does not change over the course of a growth season or among successive years. Results indicated a dynamic microbial community that is more strongly influenced by sampling day during the growth season than by microsite variation. A surprising diversity of bacterial genera known to be associated with the key function of methane-oxidation (methanotrophy)–including relatively high-abundance of *Crenothrix, Methylomonas,* and *Methylocaldum*–showed intra-seasonal and inter-annual variability possibly related to temperature differences, and microsite preferences possibly related to variation in methane abundance. By contrast, a core assemblage of bacterial genera seems to persist over a growth season and from year-to-year, possibly transmitted by a persistent attached host resting stage.

Keywords: epibacterial diversity; *Cladophora*; microbiota; spatiotemporal dynamics

INTRODUCTION

The periphytic filamentous green alga *Cladophora* commonly occurs in marine and freshwaters worldwide, is particularly abundant in environments affected by eutrophication, and plays important roles in ecosystem materials cycling and providing habitat for other organisms (reviewed by Zulkifly et al. 2013). Some of these ecosystem roles are mediated by epibiotic microorganisms, which benefit from high host surface area for attachment and host-generated organic carbon and oxygen. A previous high throughput 16S rRNA gene amplicon analysis indicated that the bacterial microbiota occupying surfaces of lacustrine *Cladophora glomerata* displayed high levels of taxonomic and inferred functional diversity, including cyanobacteria and other bacterial genera, many of which are known to be associated with methane-oxidation, nitrogen-fixation, and cellulose-degradation (Zulkifly et al. 2012). A subsequent metagenomic analysis of *Cladophora* sampled from the same locale revealed high archaeal, bacterial, and eukaryotic epibiont taxonomic diversity, as well as diverse functional genes such as prokaryotic *nifH* (a marker for nitrogen-fixation), *amo* (ammonia monooxygenases), genes involved in aerobic and anaerobic biosynthesis of vitamin B_{12} required by the host, and prokaryotic and eukaryotic genes encoding putative cellulases (Graham et al. 2015a). These results indicated that the *Cladophora* host and attached microbiota modify and exchange key materials relevant to N and C cycling.

Even so, Graham et al. (2015b) pointed out the need for greater attention to spatial and temporal replication of high throughput genetic sequencing studies of algal microbiota and microbiomes. These authors also recommended that greater attention be paid to the possibility that epibiotic microbiota might provide functional redundancy that could support holobiont response to environmental change. To address these issues we used 16S rRNA gene amplicon

sequencing to classify and compare bacterial microbiota of *Cladophora glomerata* sampled in triplicate, and at six time points during the growth season of 2014, from three microsites at the same lacustrine locale employed in two previous high-throughput (but less-well replicated) studies (Zulkifly et al. 2012, Graham et al. 2015a). We used statistical methods to test the null hypotheses that the *C. glomerat*a epibiotic bacterial community: 1) was relatively homogeneous across microsites tested, and 2) did not change substantially over the course of a growth season. We aimed to determine whether seasonality or small-scale geographical variation was the more important driver of epibiotic bacterial community variation during a growth season. We also used class- and genus-level comparisons of taxonomic data for the same locale from three years (2011, 2012, 2014) to explore inter-annual variability. Additionally, we used comparisons at the genus level of *Cladophora*-associated methanotrophic bacteria among these three years to gain insight into community environmental responsiveness and functional redundancy.

MATERIALS AND METHODS

Sampling locale and microsites. The sampling locale was Picnic Point, a shoreline peninsular feature of Lake Mendota, Dane Co., WI (43°4.53' N, W 89°35.5' W, Supplementary Figure 1). This locale was chosen because two previous high throughput studies of *Cladophora* microbiota (Zulkifly et al. 2013, Graham et al. 2015a) were based on samples obtained at the same place in 2011 and 2012, potentially allowing inter-annual comparisons. The locale was suitable for all three studies because it is relatively undisturbed by shoreline development, and is representative of other relatively undisturbed shorelines of the same lake (Graham et al. 2015a). Attached biomass dominated by the species *C. glomerata* (Graham et al. 2015a) was sampled six times between Julian days 172-214 (June 21st to August 2nd), 2014 from three peninsular microsites, a

northern aspect (N), southern aspect (S), and point (P), ~100m apart. The point microsite was the

least shaded by shoreline vegetation. Although *Cladophora* occurs on substrata located near the

water surface as well as substrata located more deeply in the water column, for this study,

sampling was restricted to depths of 1.0-1.5m. Water temperature at the depth of algal collection

increased over the open-water period; water temperature was 21°C on Julian day 160 and peaked

at 26°C on Julian day 220 (Supplementary Figure 2). Near-shore regions of Picnic Point and

other areas of Lake Mendota examined in 2014 were supersaturated with methane; methane

concentrations varied from 190 ppm at mid-summer, peaked at 350 ppm in late August, then

dropped to 20 ppm in October (Luke Loken, University of Wisconsin Center for Limnology,

personal communication).

  Sample collection, transport and processing. Using ethanol-sterilized tools, we removed three

entire highly-branched thalli, including holdfasts, from submersed substrata and placed them in

sterile Whirl-Pak bags (Nasco, Fort Atkinson, WI) that also contained local lake water and a

headspace of air. Samples were transported to the lab in an insulated container, then processed

the same day. To remove planktonic and loosely-associated materials, as in previous high

throughput sequencing studies of *Cladophora* microbiota (Zulkifly et al. 2012, Graham et al.

2015a), algal filaments were washed three times in freshly made, autoclave-sterilized SD11, a

defined medium closely resembling the chemical composition and pH of Lake Mendota water

(Graham et al. 1996). The washing process involved using sterile forceps to transfer algal thalli

into a sterile plastic petri dish containing sterile medium, swirling the dish for ~1 minute, and

repeating this process twice more, using fresh petri dishes and medium. *Cladophora* thalli were

trimmed to retain only the most basal two cm, which included older basal cells and holdfasts: 1)

to consistently sample host material of similar age, 2) because young host cells at growing tips

are much less highly colonized, and 3) to facilitate comparisons with samples from two previous

years, which had been similarly collected, processed, and analyzed (Zulkifly et al. 2012, Graham

et al. 2015a). Cleaned, trimmed algal materials collected in 2014 were placed into 1.5 mL

Eppendorf Safe-Lock tubes. We processed three technical replicates from each of the three

microsites that were sampled once per week during two three-week periods, one period near the

beginning and the other toward the end of the 2014 growth season (Julian days 172, 178, 185,

199, 206, and 214). Two of the 54 possible samples, one sample from site Point and one sample

from site South on Julian day 206 (July 25th), were excluded from further processing due to

contamination by soil from nearby shoreline, resulting in 52 final samples.

   DNA extraction, amplification, and sequencing. Total genomic DNA was extracted on the

same day as collection and processing using the DNeasy Plant Mini Kit (Qiagen) amended with

100 µL of lysozyme (Sigma-Aldrich) of concentration 100 mg/ml, to increase access to DNA of

bacterial cells having recalcitrant cell walls (Yuan et al. 2012). Samples included DNA from the

host, epibionts DNA, and DNA of intracellular bacteria potentially present. All DNA was stored

in a freezer at -20°C from the date of collection/extraction until September 16, 2014, when all

DNA samples were transported to the University of Wisconsin-Madison Biotechnology Center

for 16S rRNA gene amplification and amplicon sequencing. 16S rRNA gene is both highly

conserved across the domain Bacteria and contains variable regions, making amplified reads of

this gene ideal for conducting a broad taxonomic census of bacterial microbiota (Ward et al.

1990).

   Variable regions V3 and V4 of the 16S rRNA gene were targeted as part of the 16S

Metagenomic Sequencing Library Preparation Protocol, Par #15044223 Rev. B (Illumina Inc.,

San Diego, California, USA), using forward primer S-D-Bact-0341-b-S-17 and reverse primer S-

D-Bact-0785-a-A-21 (Klindworth et al. 2013). Paired end, 250 base pair sequencing was

performed using the Illumina MiSeq Sequencer and a MiSeq 500 bp (v2) sequencing cartridge.

Amplicons were sequenced at the University of Wisconsin-Madison Biotechnology Center using

the standard Illumina Pipeline, version 1.8.2, and these sequences have been deposited at the

National Center for Biotechnology Information (NCBI) Short Reads Archive (SRA) under

BioProject ID PRJNA360140.

   Bioinformatic processing. Forward and reverse paired ends were merged with PEAR (Zhang et

al. 2014), and QIIME v.1.9.1 (Caporaso et al. 2010) was used to cluster reads with the UCLUST

algorithm (Edgar 2010) and annotate resulting operational taxonomic units (OTUs) at 97%

minimum similarity with the Silva database v. 128 (Quast et al. 2013). A total of 11,541,927

reads were obtained from amplicon sequencing. 985,143 reads were removed due to insufficient

length or quality for assigning taxonomy. Taxonomy was assigned to the remaining 10,556,784

reads (a mean of $203,015 \pm 47,471$ reads per sample). Reads from Zulkifly et al. (2012), which

was likewise a 16S rDNA amplicon study, were reanalyzed with the following bioinformatic

processing for accurate comparison to the reads of the samples in this study.

   Reads in each sample were filtered to represent only OTUs identified as bacterial in origin and

singletons were removed. In an effort to reduce PCR bias that could inflate downstream OTU

abundance metrics and indices, OTUs were normalized by relative abundance by dividing the

sum of reads of each OTU by the sum of all reads in the sample. The Shannon index of alpha

diversity, which incorporates both relative abundance and evenness across OTUs in a community

(Sager & Hasler 1969), was calculated to measure site- and time-specific bacterial diversity. We

chose the Shannon index for this study because the potentially low abundances of many rare

OTUs detected from all bacteria associated with the alga influence the quantification of diversity

(Wilhm 1970). The test for non-parametric multivariate analysis of variance (NPMANOVA) (Anderson 2001) was used to detect significant differences among microbial communities. Non-metric multi-dimensional scaling (NMDS) was used to cluster microbial communities, using a calculated Bray-Curtis dissimilarity distance matrix. All calculations and tests were completed using the R packages Phyloseq (McMurdie & Holmes 2013) and vegan (Dixon 2003).

Comparisons of 2014 results to those reported from the same locale during 2011 and 2012. To gain insight into inter-annual variability, taxonomic results for samples collected in 2011 (Zulkifly et al. 2012) and 2014 (this study), which were obtained using similar amplicon methods, were compared at the class level. We also compared lists of bacterial genera reported from: 1) a 16S rRNA gene amplicon analysis of *Cladophora* microbiota sampled in 2011 (Zulkifly et al. 2012), 2) an analysis of samples collected from the same locale in 2012 using 16S rRNA gene sequences filtered from whole metagenomic shotgun sequence (Graham et al. 2015a), and 3) the present analysis of *Cladophora*, sampled in 2014. This effort was designed to reveal possible core bacterial epibionts that may persist from year to year. However, in this process, the number of shared genera was almost certainly underestimated because public reference sequence databases grew during the 2011-2017 period encompassing the three contrasted studies.

We also employed the bacterial taxonomic data to infer the spatio-temporal dynamics of *Cladophora*-associated methanotrophic bacteria, whose activity in freshwater aquatic systems is of increasing interest. Because multiple methanotrophic bacterial genera–possibly reflecting functional redundancy–were inferred for *Cladophora* microbiota sampled from the same site in 2011, yet a metagenomic analysis of *Cladophora* microbiota sampled in 2012 did not reveal this guild, we also examined the occurrence of methanotrophs in 2014.

RESULTS

Epibiotic bacterial communities sampled from Lake Mendota (WI, USA) *Cladophora*

glomerata at six time points and three microsites during the growing season of 2014 included

57,901 total possible OTUs present in 52 separate assemblages of 16S rRNA gene sequenced

reads. The dominant bacterial phyla present in the *Cladophora* microflora, in decreasing rank by

mean relative abundance, were *Bacteriodetes, Proteobacteria, Deinococcus-Thermus,*

Cyanobacteria, Acidobacteria, and *Actinobacteria*, but *Bacteriodetes* and *Proteobacteria*

accounted for most of the relative abundance (Supplementary Figure 3). Spatio-temporal

occurrence of the 30 most abundant bacterial genera can be found in Supplementary Fig. 4.

In contrast to null hypotheses of spatial and temporal homogeneity, analysis of all reads

identified as bacterial in origin indicated substantial temporal and spatial variation. The non-

parametric multivariate analysis of variance test showed that the bacterial microbiota associated

with Picnic Point, Lake Mendota *Cladophora glomerata* differed significantly over the growth

season (pseudo F = 26.05, R^2 = 0.438, p < 0.001) and between microsites (pseudo F = 30.92, R^2

= 0.208, p < 0.001). Because there was a significant interaction between sampling date and

microsite (pseudo F = 7.11, R^2 = 0.239, p < 0.001), Shannon diversity indices linear regressions

were separated by microsite. The Point microsite showed a gradual net diversity increase over

seasonal time (slope = 0.0144/day, R^2 = 0.919, starting at 5.0, ending at 5.7). The South

microsite more sharply increased in diversity over the growth season (slope = 0.0252/day, R^2 =

0.9214, starting at 5.3, ending at 6.4) than did the North and Point microsites. The North

microsite diversity decreased from 5.7 to 5.2 and ended at 5.6, but this trend did not fit a linear

pattern (slope = 0.00263/day, R^2 = 0.0851) (Fig. 1). 2014 samples clustered equally by collection

date and site (Fig. 2), suggesting nearly equal contribution to differences in composition.

However, in this study the collection period was relatively long, but distances between microsites were relatively small. The data indicate that samples from early-summer collection dates clustered separately from later-summer samples (Fig. 2).

At the class level, we observed considerable variability in relative bacterial abundance over time and space (Fig. 3). Notable trends of dominant classes were an abundance of *Alphaproteobacteria* at the Point microsite, and a general decline over time of *Acidomicrobia, Actinobacteria, Armatimonadia, Cytophagia,* and *Flavobacteria.* Re-analysis and comparison of class-level data originally obtained in 2011 (Zulkifly et al. 2012) suggested that abundances of taxa in classes *Acidomicrobiia, Actinobacteria, Clostridia, Deinococci, Deltaproteobacteria, Gammaproteobacteria, Gemmatimonadetes, Nitrospira,* and *Sphingobacteria* were lower in 2011 than nearly all sampled dates in 2014 (Fig. 3). The same contrast indicated that abundances of classes *Anaerolineae, Betaproteobacteria,* and *Flavobacteria* were all higher in 2011 than at any time in 2014 (Fig. 3).

A total of 33 consistently detected bacterial genera associated with *Cladophora* were found in 2011 (Zulkifly et al. 2012), 2012 (Graham et al. 2015), and 2014 (this study), and these genera are listed in Supplementary Table 1. Seven putatively methanotrophic genera were inferred to occur as epibionts on *Cladophora* sampled in 2014, listed here in order of greatest to least mean relative abundance: *Crenothrix, Methylobacter, Methylovulum, Methylocaldum, Methylocella, Methylocystis,* and *Methylomonas*. Relative abundances of these genera varied through the growth season and across microsites. Methanotrophic bacterial genera were inferred to be most abundant at the South microsite throughout the summer of 2014 and in early summer at the North microsite, but were nearly undetectable at the Point microsite (Fig. 4).

DISCUSSION

The goals of this study were to statistically test the null hypotheses that lacustrine *Cladophora glomerata*-associated bacterial communities were relatively homogeneous across the growing season of 2014, and that communities were relatively homogenous among relatively close microsites at the same sampling locale. Published taxonomic data from two previous years were also used together with 2014 information to gain insight into inter-annual variation, with a particular focus on methanotrophs, because methane-oxidation is known to be a key ecosystem function in lakes (Khmelenina et al. 2010, Deutzmann et al. 2014, Billard et al. 2015, Zigah et al. 2015, Yang et al. 2016, Roland et al. 2017, Samad and Bertilsson 2017).

Temporal and spatial variation. Our results indicated that populations of relatively shallow lacustrine *Cladophora* support a highly dynamic bacterial community whose composition within a growing season was influenced twice as much by sampling day as by closely-placed microsites. The results indicated that future studies of periphytic algal microbiota should include more than two technical replicates per sample site as well as sampling at multiple times during a growth season. Variation also occurred among epibiotic bacterial communities of closely-spaced microsites, but optimal spacing of sampling sites may depend upon the specific hypotheses being tested and upon the degree of variation in locale features such as substratum quality and shading, and therefore may need to be determined for individual ecosystems.

The consistent increase in alpha diversity we observed across two microsites during the course of a growth season suggests a pattern of accumulation of bacterial taxa. Even so, observed increases and decreases in relative abundance at all taxonomic levels suggests that complex successions of microbial guilds may surge and retreat in response to microenvironment changes. Environmental factors other than water temperature were not assessed in our study, but future investigations of periphytic algal microbiota might also include monitoring of variation in

sunlight, wave action, shoreline aspect, human use, and shoreline runoff, to begin to parse the drivers of community change over time.

Functional inferences. Although there are limitations to inferring bacterial functions from taxonomic information (Langille et al. 2013), inferring function from generic classification can be justified when a metabolic function is highly conserved at the generic level. Our observations of bacterial communities associated with lacustrine *Cladophora* during the growing season of 2014 reveals patterns of change in bacterial genera that might be ecologically significant and thus warrant further investigation. For example, the studied eutrophic lake is rich in dissolved organic carbon, phosphorus, and nitrogen (Beversdorf et al. 2013; Carpenter et al. 2014; Karatayev et al. 2013), which may foster the occurrence of hardy copiotrophs such as *Meiothermus* (Tindall et al. 2010), *Sediminibacterium* (Kang et al. 2014), *Pseudomonas* (Klausen et al. 2003), and *Flectobacillus* (Hwang et al. 2006), all of which occurred at high relative abundance on *Cladophora*. The phototrophic cyanobacterium *Leptolyngbya* (Shimura et al. 2015) occurred in greater relative abundance at the least-shaded microsite, but *Rhodobacter*, commonly a non-oxygenic photosynthesizer (Hiraishi et al. 1996), did not seem to have a preferred microsite.

Temporal and spatial patterns of methanotrophs may be of particular functional importance. Although most studies of lacustrine methanotrophs have focused on suspended or benthic taxa (Khmelenina et al. 2010, Deutzmann et al. 2014, Billard et al. 2015, Zigah et al. 2015, Yang et al. 2016, Roland et al. 2017, Samad and Bertilsson 2017), our past (Zulkifly et al. 2012) and present evidence for multiple genera of putatively methanotrophic bacteria in *Cladophora* epibiotic communities indicates that littoral methanotrophs may also be important. Our inferences of rapid increase followed by slow decrease in the relative abundances of the putative

methanotrophs *Crenothrix* (Stoecker et al. 2006), *Methylocaldum* (Islam et al. 2015), and

Methylomonas (Ogiso et al. 2012) suggest that littoral methane concentrations may have changed

over time, possibly reflecting differences in emissions from sediments. The putative

methanotrophs *Methylomicrobium* (Brantner et al. 2002) and *Methylosinus* (Fox et al. 1989)

were found in lower relative abundance, but these genera contributed to distinctive community

responses at the South microsite. These results indicate that greater attention should be paid to

methanotrophs associated with periphytic algae, particularly for lakes having a high proportion

of littoral zone occupied by *Cladophora* and other periphytic algal hosts.

   Inter-annual comparisons. Consistent with our two previous, less well-replicated studies that

were conducted at the same locale (Zulkifly et al. 2012, Graham et al. 2015a), the present

assessment indicated that the microbiota of *Cladophora glomerata* includes representatives of

diverse bacterial phyla, but that *Bacteroidetes* and *Proteobacteria* dominated. Although we

inferred presence of similar bacterial classes during the 2011 and 2014 seasons, differences in

mean relative abundance were observed in the two years, suggesting inter-annual variation.

However, because our results indicated that bacterial communities can vary even among closely-

located microsites sampled on the same day, slight differences in collection location employed in

the different years might explain some of this variation. In addition, less well-replicated

sampling in the earlier year might have resulted in time- or site-biased collection bias. Even so,

future explorations of inter-annual variation in *Cladophora* epibiotic bacterial communities and

possible environmental correlates seem justified in view of potentially powerful ecosystem

functions such as methanotrophy.

   Our observation that >30 bacterial genera, including functionally important representatives

(Zulkifly et al. 2012), were present on *Cladophora* in all three study years suggests the

occurrence of a core microbiota that might be transmitted from one year to the next. The mechanism of transmission might be persistent thick-walled akinetes that remain attached to submergent substrata from year to year (Zulkifly et al. 2013). To test this hypothesis, future investigations might explore the microbiota of *Cladophora* akinetes collected during late fall, winter, and early spring, to check for the ~30 putative core bacterial genera.

The contrast between inferred presence of methanotrophic bacteria in epibiotic *Cladophora* communities in 2011 and 2014 and complete lack of detection of methanotrophs or sequences encoding enzymes that mark methane oxidation (e.g. *pMMO*) in similarly collected and processed samples from 2012 suggests an important environmental difference between these years. Graham et al. (2015a) suggested that an earlier than usual rise in lake temperature coupled with a longer than usual period of sustained maximal lake temperature that occurred in 2012 may have been detrimental to the *Cladophor*a-methanotroph association. However, because the present study indicated that putative methanotrophs can be abundant at one microsite, yet undetectable at a nearby microsite, it is possible that 2012 spatial sampling missed methanotrophs that might have been present elsewhere. These observations suggest that future efforts to monitor the abundances, activities, and spatial and temporal heterogeneity of *Cladophora*-associated methanotrophs are warranted. Even though conspicuous *Cladophora* growths in freshwater bodies are often maligned, it is becoming increasingly more evident that the holobiont serves important ecosystem functions such as methanotrophy.

ACKNOWLEDGEMENTS

We thank the Anna Grant Birge Fund administered by the UW-Madison Center for Limnology financial support for amplicon sequencing, and UW-Madison Biotechnology Center personnel for services and expertise in Next-Gen Sequencing technology.

REFERENCES

Anderson, M.J. 2001. A new method for nonparametric multivariate analysis of variance. *Austral Ecol.* 26:32–46.

Beversdorf, L. J., Miller, T. R. & McMahon, K. D. 2013. The role of nitrogen fixation in cyanobacterial bloom toxicity in a temperate, eutrophic lake. *PLoS One* 8.

Billard, E., Domaizon, I., Tissot, N., Arnaud, F. & Lyautey, E. 2015. Multi-scale phylogenetic heterogeneity of archaea, bacteria, methanogens and methanotrophs in lake sediments. *Hydrobiologi*a 751:159-173.

Brantner, C. A., Remsen, C. C., Owen, H. A., Buchholz, L. A. & Collins, M. 2002. Intracellular localization of the particulate methane monooxygenase and methanol dehydrogenase in *Methylomicrobium album* BG8. *Arch. Microbiol.* 178:59–64.

Caporaso, J. G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F. D., Costello, E. K., Fierer, N. et al. 2010. Correspondence QIIME allows analysis of high-throughput community sequencing data intensity normalization improves color calling in SOLiD sequencing. *Nat. Publ. Gr.* 7:335–6.

Deutzmann, J. S., Stief, P., Brandes, J. & Schink, B. 2014. Anaerobic methane oxidation coupled to denitrification is the dominant methane sink in a deep lake. *Proc. Natl. Acad. Sci. USA* 111:18273-18278.

Dixon, P. 2003. VEGAN, a package of R functions for community ecology. *J. Veg. Sci.* 14:927–30.

Edgar, R. C. 2010. Search and clustering orders of magnitude faster than BLAST. 26:2460–1.

Fox, B. G., Froland, W., Dege, J. E. & Lipscomb, J. D. 1989. Methane monooxygenase from *Methylosinus trichosporium* OB3b. *J. Bacteriol.* 264:10023–33.

Graham, J. M., Arancibia-Avila, P. & Graham, L. E. 1996. Effects of pH and selected metals on growth of the filamentous green alga *Mougeotia* under acidic conditions. *Limnol. Oceanogr.* 41:263–70.

Graham, L.E., Knack, J. J., Graham, M. E., Graham, J. M. & Zulkifly, R. 2015a. A metagenome for lacustrine *Cladophora* (Cladophorales) reveals remarkable diversity of eukaryotic epibionts and genes relevant to materials cycling. *J. Phycol.* 418:408–18.

Graham, L. E., Wilcox, L. W. & Knack, J. J. 2015b. Why we need more algal metagenomes. *J. Phycol.* 51:1029-1036.

Hiraishi, A., Muramatsu, K. & Ueda, Y. 1996. Molecular genetic analyses of *Rhodobacter azotoformans* sp. nov. and related species of phototrophic bacteria. *Syst. Appl. Microbiol.* 19:168–77.

Hwang, C. Y. & Cho, B. C. 2006. *Flectobacillus lacus* sp. nov., isolated from a highly eutrophic pond in Korea. *Int. J. Syst. Evol. Microbiol.* 56:1197–201.

Islam, T., Larsen, Ø., Torsvik, V., Øvreås, L., Panosyan, H., Murrell, J., Birkeland, N.-K. *et al*. 2015. Novel methanotrophs of the family Methylococcaceae from different geographical regions and habitats. *Microorganisms* 3:484–99.

Kang, H., Kim, H., Lee, B., Joung, Y. & Joh, K. 2014. *Sediminibacterium goheungense* sp. nov., isolated from a freshwater reservoir. *Int. J. Syst. Evol. Microbiol.* 64:1328–33.

Karatayev, A. Y., Burlakova, L. E., vander Zanden, M. J., Lathrop, R. C. & Padilla, D. K. 2013. Change in a lake benthic community over a century: Evidence for alternative community states. *Hydrobiologia* 700:287–300.

Khmelenina, V. N., Shchukin, V. N. , Reshetnikov, A. S., Mustakhimov, I. I., Suzina, N. E., Schinimaev, B. T., & Trotsenko, Y. A.. 2010. Structural and functional features of methanotrophs from hypersaline and alkaline lakes. *Microbiology* 79:472-482.

Klausen, M., Aaes-Jørgensen, A., Molin, S. & Tolker-Nielsen, T. 2003. Involvement of bacterial migration in the development of complex multicellular structures in *Pseudomonas aeruginosa* biofilms. *Mol. Microbiol.* 50:61–8.

Klindworth, A., Pruesse, E., Schweer, T., Peplies, J., Quast, C., Horn, M. & Glöckner, F. O. 2013. Evaluation of general 16S ribosomal RNA gene PCR primers for classical and next-generation sequencing-based diversity studies. *Nucleic Acids Res.* 41:1–11.

Langille, M., Zaneveld, J., Caporaso, J. G., McDonald, D., Knights, D., Reyes, J., Clemente, J. *et al*. 2013. Predictive functional profiling of microbial communities using 16S rRNA marker gene sequences. *Nat. Biotechnol.* 31:814–21.

McMurdie, P. J. & Holmes, S. 2013. Phyloseq: An R package for reproducible interactive analysis and graphics of microbiome census data. *PLoS One* 8.

Ogiso, T., Ueno, C., Dianou, D., Van Huy, T., Katayama, A., Kimura, M. & Asakawa, S. 2012. *Methylomonas koyamae* sp. nov., a type I methane-oxidizing bacterium from floodwater of a rice paddy field. *Int. J. Syst. Evol. Microbiol.* 62:1832–7.

Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., Peplies, J. *et al*. 2013. The SILVA ribosomal RNA gene database project: Improved data processing and web-based tools. *Nucleic Acids Res.* 41:590–6.

Roland, F. A. E., F. Darchambeau, C. Morana, S. Bouillon & A. V. Borges. 2017. Emission and oxidation of methane in a meromictic, eutrophic and temperate lake (Dendre, Belgium). *Chemosphere* 168:756-764.

Sager, P. E. & Hasler, A. D. 1969. Species diversity in lacustrine phytoplankton. I. The components of the index of diversity from Shannon's formula. *Am. Nat.* 103:51–9.

Samad, M. S. & Bertilsson, S. 2017. Seasonal variation in abundance and diversity of bacterial methanotrophs in five temperate lakes. *Front. Microbiol.* 8:142.

Shimura, Y., Hirose, Y., Misawa, N., Osana, Y., Katoh, H., Yamaguchi, H. & Kawachi, M. 2015. Comparison of the terrestrial cyanobacterium *Leptolyngbya* sp. NIES-2104 and the freshwater *Leptolyngbya boryana* PCC 6306 genomes. *DNA Res.* 22:403–12.

Stoecker, K., Bendinger, B., Schöning, B., Nielsen, P. H., Nielsen, J. L., Baranyi, C., Toenshoff, E. R. *et al*. 2006. Cohn's *Crenothrix* is a filamentous methane oxidizer with an unusual methane monooxygenase. *Proc. Natl. Acad. Sci. USA* 103:2363–7.

Tindall, B. J., Sikorski, J., Lucas, S., Goltsman, E., Copeland, A., Glavina Del Rio, T., Tice, H. *et al*. 2010. Complete genome sequence of *Meiothermus silvanus* type strain (VI-R2). *Stand. Genomic Sci.* 3:37–46.

Ward, D. M., Weller, R. & Bateson, M. M. 1990. 16S rRNA sequences reveal numerous uncultured microorganisms in a natural community. *Lett. Nat.* 345:183–7.

Wilhm, J. L. 1970. Effect of sample size on Shannon's formula. *Southwest Nat.* 14:441–5.

Yang, Y., Zhao, Q., Cui, Y., Wang, Y., Xie, S. & Liu, Y. 2016. Spatio-temporal variation of sediment methanotrophic microorganisms in a large eutrophic lake. *Microb. Ecol.* 71:9-17.

Yuan, S., Cohen, D. B., Ravel, J., Abdo, Z. & Forney, L. J. 2012. Evaluation of methods for the extraction and purification of DNA from the human microbiome. *PLoS One* 7.

Zhang, J., Kobert, K., Flouri, T. & Stamatakis, A. 2014. PEAR: A fast and accurate Illumina Paired-End reAd mergeR. *Bioinformatics*. 30:614–20.

Zigah, P. K., Oswald, K., Brand, A., Dinkel, C., Wehrli, B. & Schubert, C. J. 2015. Methane oxidation pathways and associated methanotrophic communities in the water column of a tropical lake. *Limnol. Oceanogr*. 60:553-572.

Zulkifly, S. B., Graham, J. M., Young, E. B., Mayer, R. J., Piotrowski, M. J., Smith, I. & Graham, L. E. 2013. The Genus *Cladophora* Kutzing (Ulvophyceae) as a globally distributed ecological engineer. *J. Phycol.* 49:1–17.

Zulkifly, S., Hanshew, A., Young, E. B., Lee, P., Graham, M. E., Piotrowski, M. & Graham, L. E. 2012. The epiphytic microbiota of the globally widespread macroalga *Cladophora glomerata* (Chlorophyta, Cladophorales). *Am. J. Bot.* 99:1541–52.

TABLE AND FIGURES

Figure 1. Mean alpha diversities, with linear regressions, as measured by the Shannon index of 16S amplicon operational taxonomic units (OTUs) that represent bacterial diversity calculated by weighting taxa richness by the evenness of the OTUs detected in the epibiota of *Cladophora* from three sites on Picnic Point, Lake Mendota, WI, during the summer of 2014. Error bars ±SE (n=3).

Figure 2. Non-metric multi-dimensional scaling (NMDS) ordination plot using Bray-Curtis distances of 16S rRNA gene amplicon reads identifed as bacterial, sequenced from samples of *Cladophora* collected from Picnic Point, Lake Mendota, WI, during the summer of 2014. Color indicates collection date (Julian days 172 to 214), and shape indicates collection site (circle = North, triangle = Point, square = South). This plot visualizes an ordination using Bray-Curtis calculated differences (k = 2) between all samples in two-dimensional space and has a sufficiently low stress of 0.1186.

Figure 3. Mean relative abundance of epibiont bacterial classes on *Cladophora* sampled from Picnic Point, Lake Mendota, WI, during the summer of 2014, inferred from 16S rRNA gene sequences. Error bars ±SE (n=3). Red dashed lines indicate values for classes detected by 16S rRNA gene sequences associated with *Cladophora* collected from the same locale in the summer of 2011 by Zulkifly et al. (2012) after reanalysis using the bioinformatic processing used in this study.

Figure 4. Mean relative abundance of methanotrophic (methane-oxidizing) genera occurring in the bacterial microbiota of *Cladophora* sampled at three microsites (North, Point, and South) from the locale Picnic Point, Lake Mendota, WI, during the summer of 2014, inferred from the 16S rRNA gene. Error bars ±SE (n=3).

Bacterial Class	Bacterial Genus
Alphaproteobacteria	*Hyphomicrobium*
	Porphyrobacter
	Sandarakinorhabdus
	Altererythrobacter
	Sphingopyxis
	Sphingomonas
	Rhodobacter
	Brevundimonas
Betaproteobacteria	*Leptothrix*
	Hydrogenophaga
	Curvibacter
	Ideonella
	Aquabacterium
Deltaproteobacteria	*Byssovorax*
	Sorangium
	Bdellovibrio
Gammaproteobacteria	*Haliea*
	Pseudomonas
	Lysobacter
	Alkaninidiges
	Rheinheimera
Sphingobacteria	*Terrimonas*
	Flavisolibacter
	Ferruginibacter
	Haliscomenobacter
Bacteroidetes	*Flavobacterium*
	Luteolibacter
	Opitutus
	Verrucomicrobium
	Prosthecobacter
Deinococcus-Thermus	*Trupera*
	Meiothermus
Gemmatimonadetes	*Gemmatimonas*

Supplementary Table 1. Recurring core bacterial genera closely associated with *Cladophora*

glomerata in Lake Mendota identified using short reads of the 16S rRNA in the growth seasons

of 2011 (Zulkifly et al. 2012), 2012 (Graham et al. 2015), and 2014 (this study).

Supplementary Figure 1. Map of collection sites North (N), Point (P), and South (S) on Lake

Mendota, WI, USA.

Supplementary Figure 2. Mean daily temperatures aggregated from the Mendota Buoy Long

Term Ecological Network database from 2006 to 2016, excluding data from 2014 when

intrumentioned malfunctioned, resulting in large temperature measurement gaps and errors.

Green dashed lines indicate sampling dates of this study.

Supplementary Figure 3. Mean relative abundances of the 16 most abundant bacterial phyla occurring in the bacterial microbiota of *Cladophora* sampled at three microsites (North, Point, and South) from the locale Picnic Point, Lake Mendota, WI, during the summer of 2014, inferred from the 16S rRNA gene. Error bars ±SE (n=3).

Supplementary Figure 4. Mean relative abundances of the 30 most abundant bacterial genera occurring in the bacterial microbiota of *Cladophora* sampled at three microsites (North, Point, and South) from the locale Picnic Point, Lake Mendota, WI, during the summer of 2014, inferred from the 16S rRNA gene gene. Error bars ±SE (n=3).