Disproportional distribution of radioactive elements in the marine ecosystems surrounding

the accident site of the Fukushima Daiichi Nuclear Power Plant

Shigeharu Moriya[1, 3, *], Hideaki Otsu[2], Kumiko Kihara[1, 4], Yukari Kato[1], Misao Itouga[1], Kenji

Sakata[1], Jun Kikuchi[1, 3], Hiroshi Yamakawa[5, 6] and Jiro Suga[6]

1) RIKEN Center for Sustainable Resource Science, 1-7-22 Suehiro-cho, Tsurumi-ku,

Yokohama city, Kanagawa, 230-0045, JAPAN

Shigeharu Moriya, Kumiko Kihara, Yukari Kato, Misao Itouga, Kenji Sakata and Jun

Kikuchi

2) RIKEN Nishina Center for Accelerator-Based Science, 2-1 Hirosamwa, Wako city,

Saitama, 351-0198, JAPAN

Hideaki Otsu

3) Graduate School of Medical Life Science, Yokohama-city University, 1-7-29 Suehiro-cho,

Tsurumi-ku, Yokohama city, Kanagawa, 230-0045, JAPAN

Shigeharu Moriya and Jun Kikuchi

4) Department of Biological and Chemical Systems Engineering, National Institute of

Technology, Kumamoto College, 2627 Hirayama-Shinmachi, Yatsushiro-shi, Kumamoto

866-8501, JAPAN

Kumiko Kihara

5) Tokyo University of Marine Science and Technology, 4-5-7 Konan, Minato-ku, Tokyo,

108-8477, JAPAN

Hiroshi Yamakawa

6) Japan Academy of Underwater Sciences, 3-9-1-301 Botan, Koto-ku, Tokyo, 135-0046,

JAPAN

Hiroshi Yamakawa and Jiro Suga

*) Corresponding author: Shigeharu Moriya

RIKEN Center for Sustainable Resource Science, 1-7-22 Suehiro-cho, Tsurumi-ku,

Yokohama city, Kanagawa, 230-0045, JAPAN

Tel +81-45-508-7221, Fax +81-45-508-7363

smoriya@riken.jp

Abstract

Background

After the Fukushima Daiichi Nuclear Power Plant accident, various surveys were performed

to measure the extent of radioactive material contamination in the marine sediments,

surface waters, plankton and fish. Six months after the event, relatively high radiocesium

contamination could still be detected in the fish and sediment of the coastal demersal

environment. To determine the distribution of radioactive material in the demersal

ecosystem adjacent to the event, we sampled and analyzed the dominant macroorganisms

attached to the seafloor as well as environmental material, including biofilms, from the

coastal demersal environment around Hisanohama Port, which is less than 30 km south of

the accident site.

Results

Our results showed that variable ratios of radiosilver/radiocesium occurred in the

macrobenthos, macroalgae and sediment samples. However, the biofilm samples displayed

high radiocesium contamination but did not show radiosilver contamination.

Conclusions

These findings suggest that several different entry paths are available for radioactive

elements to access the biological components of the ecosystem, and these paths may explain

the disproportional and patchy distribution of radioactive elements among marine

ecosystems.

Keywords

Radio isotope, Marine eco-system, Biofilm

Background

Following the Great East Japan Earthquake and catastrophic tsunami in March of 2011,

the Fukushima Daiichi Nuclear Power Plant (FDNPP) accident led to the release of large

amounts of radioactive materials into the atmosphere and marine environment.

This environmental input of radioactive elements from the FDNPP introduced into the

marine waters a variety of radioisotopes, especially radiocesium isotopes. Public monitoring

showed that during the several months after the first explosion at the FDNPP, a rapid

decrease in radioisotopes was observed in the nearby seawater. A joint official survey team

from the Tokyo Power Electric Company (TEPCO), Fukushima prefectural government and

the Japanese Ministry of Education, Culture, Sports, Science and Technology (MEXT)

reported that the ^{137}Cs contamination level in the seawater quickly decreased from over 10 k

Bq/liter to less than 100 Bq/liter within a 2-month period, whereas the ^{137}Cs found in the

ocean sediments was continuously observed over this period at levels between 100 Bq/dry kg

and 1000 Bq/dry kg until at least September 2012 [1]. This disproportional distribution of

radioisotopes was also observed in marine fish by Wada et al. Wada reported that the

retention time of radiocesium appears to be remarkably longer in the larvae of demersal fish,

e.g., *Sebastes cheni*, than in surface fish, e.g., *Eugraulis japonica* [2]. Similar disproportional

distributions among radionuclides was also reported in a terrestrial ecosystem. Nakanishi

et al. reported ^{110m}Ag accumulation in the orb-web spider (*Nephila clavata*) and other

arthropods [3]. Decay-corrected data from Nakanishi's work showed that the radioactivity

ratio between ^{110m}Ag and ^{137}Cs (^{110m}Ag/^{137}Cs) in specific samples of arthropods was over 1.0,

whereas the same ratio (^{110m}Ag/^{137}Cs) in the soil was 0.0014-0.0023, which indicates that

these organisms may actively accumulate radiosilver more than radiocesium. This finding is

inconsistent with the majority of the organisms from land ecosystems, which accumulate

more radiocesium than radiosilver.

These disproportional distributions may occur via partitioning in complex ecosystems and

food webs through processes that include biological concentration. To better understand the

fate of radioisotopes in nearshore marine environments, the present study sampled water,

sediment, biofilms, macrobenthos and macroalgae as basic environmental materials and

entry points of organic/inorganic materials into the ecosystem. In the present study,

samples were collected using a self-contained underwater breathing apparatus (SCUBA) at

a depth of 5-10 m along the nearshore area around Hisanohama Port. Sampling was

essentially performed weekly from the end of November 2011 to the beginning of February

2012 at 6 different points. Our results showed that macroalgae and macrobenthos

accumulate and retain both radiocesium and radiosilver. The observed ratios of

radiosilver/radiocesium are variable among the organisms, and they may be dependent on

the specific usage and behavior of the organisms with respect to metals, such as specific

cofactors, oxygen acceptors, metal uptake and other metabolic features. Although sediment

also retains both radiocesium and radiosilver, biofilms contain only radiocesium. The

microbial community structure is likely related to the radiocesium accumulation and

retention in microbes.

Materials and Methods

Sampling

Sampling was performed from 11 November 2011 to 4 February 2012. The sampling site

was the ocean along the coastline of Hisanohama, Iwaki City, Fukushima, Japan. We

collected samples 10 times within the established time period at 11 Nov (week 1), 20 Nov

(week 2), 26 Nov (week 3), 3 Dec (week 4), 10 Dec (week 5), 17 Dec (week 6), 14 Jan (week 7),

21 Jan (week 8), 29 Jan (week 9) and 4 Feb (week 10). Sampling was performed at 6

different areas near Hisanohama Port: Area 1 (37.153762N 141.006167E, weeks 1 and 5),

Area 2 (approximately 37.149863N 141.001017E, weeks 2, 8 and 9), Area 3 (approximately

37.146579N 141.002390E, week 4), Area 3' (approximately 37.147503N 141.008570E, week

10), Area 4 (approximately 37.133887N 141.002390E, weeks 3 and 7) and Area 4'

(approximately 37.142953N 141.003592E, week 6). These sampling sites are located 30 km

south of the Fukushima Nuclear Power Plant #1. Sampling was performed by diving with a

SCUBA unit. Divers sampled the bottom water and surface water (5 L) and collected bottom

stones to obtain biofilms and locally observed macrobenthos and macroalgae.

Sediment core samples were collected at Area 2 using a piston corer prototype. The "core

sampler" was composed of a 4.5 cm diameter/28 cm height cylinder and inscribed piston.

Core samples were collected during a separate dive in weeks 6 (core sample 1) and 7 (core

samples 2 – 4).

Samplings were performed with permission from Fukushima Prefecture and the fishery

association of Iwaki City.

Sample preparation

The seawater samples were immediately filtered with two 24-mm-wide 0.22 μm Durapore

filters (Millipore, MA, USA) sample until they clogged. All samples were frozen at -20°C in a

field freezer prior to transportation back to laboratory.

In the laboratory, the samples were dried at 105°C overnight except for the filters, water

and sediment core samples. The dried samples were crushed by a food mill and settled into a

U8 container, and the weight and height were measured in the container. In the case of

biofilms, half of the samples was dried as mentioned above and the other half was captured

on 0.2 μm pore size membrane filters and stored at -20°C for ribosomal RNA gene analysis.

The water samples were dried by boiling. The remaining salts were settled into a U8

container, and the weight and height were measured in the container.

Sediment core samples for radioactivity measurements were stored at -80°C. The cores were

sliced by a clean saw at 5 cm intervals from the surface side of the sediment to the bottom

side. Each piece of the core sample was freeze dried and then settled into a U8 container,

and the weight and height measured in the container.

Measurement of radioactivity

The gamma ray activity of the samples in the containers was measured by a Germanium

(Ge) detector. The detector setup is identical to those described in [4]. The energy resolution

of 2 keV full width at half maximum (FWHM) was achieved at 1332.5 keV gamma rays. The

sample location and the detector were surrounded with 15 cm lead blocks to reduce the

background radiation in the gamma ray spectroscopy. The detection efficiency of the gamma

rays was calibrated with an accuracy of 2 - 10% using a multiple gamma ray standard

source, which ranged from 88 keV to 1836 keV. A calibrated ^{134g}Cs source was also used to

correct for the coincidence summing for radioactivity determinations of ^{134g}Cs.

Every sample was contained in a U8 plastic container. The U8 container is a standard

container in Japan that is used for measurements of absolute radioactivity with high

accuracy by a Ge detector. The geometry of the container is cylindrical, with a diameter of 47

mm and height of 60 mm. The volume is approximately 100 mL. The efficiency calibration

source is available within the container with a uniform density in the container.

Every sample was measured by the Ge detector for at least 8 hours to obtain sufficient

statistics for each gamma ray peak. The relative geometry between the sample and the Ge

detector was carefully reproduced by a guide apparatus at the sample location.

Gamma rays from ^{134,137}Cs and ^{110m}Ag were observed for the majority of the samples. The

radioactivity levels for every isotope were determined from the yield of the gamma rays

corrected by the detection efficiency of the Ge detector and a geometrical acceptance

between the sample and the detector. The geometry difference in the height of the sample

from the calibration source was simulated by several electromagnetic simulation codes,

including Geant4 [5,6] and EGS5 [7].

The radioactivity of the marine animals, macroalgae and biofilm samples were normalized

by weight, and the radioactivity of the water samples and sediment core samples were

normalized by volume.

Element analysis

To quantitatively analyze the content of 37 elements (^{7}Li, ^{9}Be, ^{23}Na, ^{24}Mg, ^{27}Al, ^{39}K, ^{43}Ca,

^{45}Sc, ^{47}Ti, ^{51}V, ^{52}Cr, ^{55}Mn, ^{57}Fe, ^{59}Co, ^{60}Ni, ^{63}Cu, ^{66}Zn, ^{85}Rb, ^{88}Sr, ^{98}Mo, ^{111}Cd, ^{133}Cs, ^{139}La, ^{140}Ce,

^{141}Pr, ^{146}Nd, ^{152}Sm, ^{153}Eu, ^{158}Gd, ^{159}Tb, ^{164}Dy, ^{165}Ho, ^{166}Er, ^{169}Tm, ^{172}Yb, ^{175}Lu, and ^{202}Hg) in

the samples (i.e., surface salts, surface microbes, biofilms), the samples were predigested

with 5 mL of concentrated HNO_3 for 1 h at room temperature. Next, the organic components

were completely decomposed by wet-ashing using a microwave sample preparation system

(Multi-Wave-3000, Perkin Elmer, MA USA) [8]. The digested samples were brought up to a

volume of 50 mL with MilliQ water (MQW) and filtered through 5B filter paper (Advantec,

Tokyo, Japan). The concentrations of the mineral elements were determined by Inductively

Coupled Plasma Mass Spectrometry (ICP-MS, NexION300, Perkin Elmer). For the ICP-MS

analysis, a portion of the filtrated samples was diluted appropriately with 0.01 mol L-1 HCl

[8]. The mineral nutrient concentration in the samples was calculated as a unit (mg g^{-1} in

dry weight).

The obtained ICP-MS data were normalized by the root-sum-of-squares (RSS) levels among

each element. Count data of radioactive cesium (^{134}Cs and ^{137}Cs) were treated with the same

calculation for normalization and then merged with the normalized ICP-MS dataset

(elements table).

ssrDNA analysis

Genomic DNA was prepared from stored biofilm samples by bead-beating with a

phenol:chloroform:isoamyl alcohol (PCI) extraction. Bead-beating was performed at 3000

rpm for 5 min (TOMY, Tokyo Japan), and then nucleic acids were recovered by ethanol

precipitation.

PCR amplification was performed with the extracted genomic DNA with two primer sets for

16S ssrDNA, 515F/806R [9,10], and 18S ssrDNA, TAReuk454FWD1/TAReukREV3 [11]. The

primers were modified for Illumina sequencing with a multi-index kit for the Nextera XT

sequencing library construction kit (Illumina, CA, USA). We added a "read1" sequence (5'-

TCG TCG GCA GCG TCA GAT GTG TAT AAG AGA CAG -3') to the forward primers and a

"read1" sequence (5'- GTC TCG TGG GCT CGG AGA TGT GTA TAA GAG ACA G -3') to the

reverse primers.

PCR was performed with a standard reaction of ExTaq DNA polymerase according to the

manufacturer's instructions (Takara Bio, Kyoto Japan) with 25 pmol of each primer/50 µl

reaction. The reactions were performed as a two-step PCR (10 cycles of 94°C for 30 s, 55°C

for 45 s and 72°C for 60 s followed by 20 cycles of 94°C for 30 s and 72°C for 60 s) for 16S and

3-step PCR (30 cycles of 94°C for 30 s, 55°C for 45 s and 72°C for 60 s) for 18S. The amplified

products were purified by 1% agarose gel electrophoresis. The obtained amplified products

were applied to the NexteraXT (Illumina CA USA) sequencing library construction process

without the "tagmentation" process.

A multi-index sequencing library was applied to the MiSeq sequencing kit v1 and read by

Illumina MiSeq according to the instructions in the paired-end sequencing mode. Data were

treated by MiSeq reporter software, and the reads were obtained from individual samples.

After the extraction of reads, reads from weeks 4, 5 and 8 for the 18S ssrDNA were

insufficient for analysis. Therefore, the following analysis was performed using reads from

weeks 1, 2, 3, 6, 7, 9 and 10.

The obtained read data were mapped onto the ribosomal RNA sequence library SILVA

release 108 by the program package QIIME [12] using "Prefix-suffix OTU picking", and the

absolute abundance OTU table at the family level was constructed. Count zero data were

manually removed from the table, and the relative abundance values were calculated. 16S

data and 18S data were treated independently and then merged into a single OTU table

with the relative abundance values.

Sequence data have been deposited in the DDBJ sequence read archive with accession

number DRA004367.

Statistical analysis

For the statistical analysis, we used the program package R [13]. First, we drew a heatmap

chart with the hierarchical cluster analysis (HCA). The heatmap.3 function in the GMD

package, which is a package for non-parametric distance measurements between two

discrete frequency distributions [14], was used to draw the heatmap with the HCA.

Independent heatmaps were drawn from each element table and the OTU table with a

two-axis HCA. A scaling option was used for the OTU table treatment according to the row

direction but was not used for the elements table treatment.

A principal components analysis (PCA) was performed with the merged dataset, including

both the out table and elements table, using the prcomp function in R. Score plots were

drawn with principal component 1 (PC1) (x-axis) and PC2 (y-axis). The loading plots for

quadrants II and III were manually drawn.

Spearman's rank correlation coefficient between the elements dataset and the OTUs dataset

was calculated with the cor function in R. A correlation matrix was used to draw the

heatmap chart with the HCA using the heatmap.3 function of R.

Results and Discussion

Heterogeneous distribution of artificial radioisotopes in the coastal marine environment

around Hisanohama

According to our gamma ray energy spectra, we found significant peaks corresponding to

^{134}Cs and ^{137}Cs in addition to natural radioisotope signals, such as ^{40}K (data not shown), for

all samples with the exception of the seawater samples. We also detected significant ^{110m}Ag

signals from the macrobenthos, macroalgae and sediment samples. Significant signals of the

radioisotopes ^{134}Cs, ^{137}Cs and ^{110m}Ag were not detected from the control samples taken from

Nishi-kawana inside Tokyo Bay, which is approximately 280 kilometers from the FDNPP.

Although ^{134}Cs (half-life of 2.0652 years) and ^{137}Cs (half-life of 30.1 years) from prior nuclear

weapons tests in the Pacific Ocean and the Chernobyl disaster might have persisted in the

seabed around Japan, our analysis of samples collected near Nishi-kawana did not detect

these signals. Therefore, we concluded that the radioisotopes detected in the samples in this

study originated from the FDNPP disaster rather than the environmental background or

past nuclear weapon tests.

As presented in Table 1, among the results from the radioactivity measurements of a variety

of samples, we noticed a heterogeneous distribution between radiocesium (^{134}Cs and ^{137}Cs)

and radiosilver (^{110m}Ag). The retention of radiocesium was observed in all sediments,

macrobenthos, macroalgae and biofilm samples, whereas non-detectable signals of these

isotopes were observed in the surficial and bottom water. This result is surprising because

the ^{134}Cs radioactivity detected in the Pacific Ocean near the accident site was 3900 Bq/m^3

(3.9 Bq/liter) in June 2011, which is one of the highest levels ever recorded [15]. Although

our results do not directly indicate that marine organisms and sediments are able to

accumulate radioisotopes at the concentrations found in this environment, we must consider

that they are able to retain radioisotopes for at least up to 9 months, which represents the

time lapse between the disaster and sample collection.

Table 1. Peak detection in the samples

Material	Peak detection		
	^{134}Cs	^{137}Cs	^{110m}Ag
Surface water	−	−	−
Bottom water	−	−	−
Particles in surface water *	−	−	−
Particles in bottom water *	−	−	−
Sediments	++	++	++
Macrobenthos	++	++	+

Macroalgae	++	++	+
Biofilms on bottom rocks	++	++	−

++: significant peak detected from all samples

+: significant peak detected from certain samples (not all samples)

-: significant peak not detected from any sample

*: Particles were filtered from water using a 0.22 μm filter, and the radioactivity was

measured

Sediment core samples were obtained from the sampling point "Area 2" by the core sampler.

The analysis of these sediment core samples indicated the presence of both radiocesium and

radiosilver. The measured radioactivities are shown in Additional file 1: Supplementary

Table 1. To better understand the radioactivity in the same volume of space, we employ the

unit Bq/liter to express the radioactivity of the core samples because the core samples

contain different materials along their depth profile.

In Figure 1, the distribution of radioactivity along the sediment column is shown. Both

radiocesium and radiosilver are distributed similarly in the surficial part of the core

samples, which contains a large amount of radioactivity. The calculated regression line for

the data showed that both radiocesium and radiosilver occur at depths above approximately

155 mm. The similar distribution of radiocesium and radiosilver suggest that both

radioisotopes start to precipitate at almost the same time and they share a similar

precipitation pattern. The ratio of radioactivity between ^{110m}Ag and ^{137}Cs is 0.0046 on

average, which is based on the detected radiosilver of each 5 cm slice of the core samples

(Fig. 1b).

Figure 1

Radioactivity of the sliced sediment core. Panel a shows the results of ^{134}Cs, ^{137}Cs and ^{110m}Ag

measurements. Open circles indicate each sliced core sample. All 4 core samples were

collected inside Hisanohama Port (Area 2). Radioactivity is expressed as Bq/liter material.

The regression lines are calculated as linear approximation lines, and the R^2 value is

indicated beside each line. Panel b shows the ^{110m}Ag/^{137}Cs ratios. Panel a shows the

individual ^{110m}Ag/^{137}Cs ratio of each core slice.

In the macrobenthos and macroalgae samples, we also detected radiosilver and radiocesium.

However, the ratio between ^{110m}Ag and ^{137}Cs varied among the samples.

As shown in Figure 2a and Additional file 2: Supplementary Table 2, macroalgae contained

a relatively low ratio of radiosilver relative to radiocesium. The ratio between ^{110m}Ag and

^{137}Cs averaged 0.06 to 0.07 among brown, red and calcareous algae (Fig. 2a); nevertheless, it

varied among individuals (Fig. 2b). Except for green algae, the ratio of ^{110m}Ag/^{137}Cs of

macroalgae was approximately 10-fold larger than that of the surrounding sediments. Green

algae contained almost no ^{110m}Ag.

Figure 2

Radioactivity of ^{134}Cs, ^{137}Cs and ^{110m}Ag in the macroalgae samples. Panel a shows the

average ^{110m}Ag/^{137}Cs ratio in the macroalgae sample, and panel b shows the individual

results for the macroalgae. The measured radioactivity is expressed by the bar graph.

Closed bars show ^{134}Cs, open bars show ^{137}Cs and gray bars show ^{110m}Ag. Radioactivity is

expressed as Bq/kg of dried sample (Bq/kg dry biomass).

Figure 3

Radioactivity of ^{134}Cs, ^{137}Cs and ^{110m}Ag in the macrobenthos samples. Panel a shows the

average ^{110m}Ag/^{137}Cs ratio in the macrobenthos sample, and panel b shows the individual

results for the macrobenthos. The measured radioactivity is expressed by the bar graph.

Closed bars represent ^{134}Cs, open bars represent ^{137}Cs and gray bars represent ^{110m}Ag.

Radioactivity is expressed as Bq/kg of the dry weight of the samples (Bq/kg dry biomass).

The arrowhead indicates the relatively high concentration of radiosilver found in the sponge

samples (i.e., the main body).

As shown in Figures 3a and 3b and Additional file 3: Supplementary Table 3, the

macrobenthos samples showed a variable ratio between ^{110m}Ag and ^{137}Cs.

Our results indicated that sea snails (Buccinidae) and oysters (*Crassostrea gigas*)

consistently presented a high ratio of ^{110m}Ag/^{134}Cs, which was also observed for the sponge

Cliona chilensis. Samples of the Ascidian *Halocynthia roretzi* showed a relatively high ratio

of ^{110m}Ag/^{134}Cs (Fig. 3a).

Certain mollusks, such as sea snails and oysters, are known to use hemocyanin as an oxygen

carrier [16]. A study on the relatively high ratio of $^{110m}Ag/^{137}Cs$ in the spider, *Atypus karschi*

indicated that the reaction center Cu in hemocyanin can be replaced by Ag, or ^{110m}Ag in this

case [3]. Similarly, the sampled mollusks may have actively incorporated ^{110m}Ag from

environmental water and taken up the radioisotope into the blood system. To corroborate

this hypothesis, our results showed a higher ratio of radiosilver compared with that of

radiocesium detected in the body ($^{134}Cs = 21.7923$, $^{137}Cs = 37.3885$, $^{110m}Ag = 134.6388$ (Bq/kg

dry biomass)) than in the shell ($^{134}Cs = 12.6245$, $^{137}Cs = 18.8071$, $^{110m}Ag = 17.7947$ (Bq/kg dry

biomass)).

A similar mechanism for accumulation and retention of the relatively high observed amount

of radiosilver in *Halocynthia roretzi* is expected. Ascidians accumulate vanadium into their

blood cells [17]. Although the function of vanadium accumulation is not yet fully understood

and the mechanism of metal uptake in ascidians is still under investigation, these

organisms may accumulate other metals using vanadium binding protein(s). Our samples of

ascidians are from *Halocynthia roretzi,* and a high specificity for vanadium has not been

reported in this species [18]. Our hypothesis is that a metal binding factor within the

Halocynthia roretzi blood system with a low specificity for vanadium is able to promote

^{110m}Ag accumulation. Although our hypothesis is based on indirect evidence, it is partially

supported by the observation that the *Halocynthia roretzi* body (including the blood system)

showed a larger amount of radiosilver than was found in the shell, which can contain

radiosilver because of attached organisms on the surface of the shell. We investigated three

individuals of *Halocynthia roretzi*, and the body measurement results are ^{134}Cs = 645,0563,

927.5638, 944.2312; ^{137}Cs = 997.5926, 1437.1948, 1588.1469; ^{110m}Ag = 295.6794, 255.9421,

213.5846 (Bq/kg dry biomass) and the shell measurement results are ^{134}Cs = 18.6579,

51.1889, 28.2958; ^{137}Cs = 30.3881, 80.5451, 43.8699; ^{110m}Ag = not detected, 3.3797, not

detected (Bq/kg dry biomass).

The sea urchin (*Anthocidaris crassispina*) samples and the majority of the sponge samples

(*Cliona chilensis*) presented ratios of radiosilver/radiocesium that were lower relative to that

of the other macrobenthos (Fig. 3a, b). Nevertheless, the feeding pattern of sea urchins and

sea snails are similar because they are both deposit feeders, and the feeding patterns of

sponges, oysters and Ascidians are also similar because they are all filter feeders.

Interestingly, 4 of the 16 specimens of sponge presented remarkably high

radiosilver/radiocesium ratios of approximately 0.5 (Fig. 3b, arrowheads). Variation were

not observed in the radiosilver/radiocesium ratio corresponding to the three color variations

of the samples of *Cliona chilensis*. A previous report showed that certain sponge species

exhibit a strong ability to incorporate a variety of metal elements, including silver [19].

Indeed, sponges are known as "a sentinel organism" that can be used as a marker to monitor

heavy metal pollution in water because they present a high capacity for specific metal

accumulation [20]. According to Genta-Jouve et al., *Acanthella acuta*, which is a species of

commonly found sponge, can incorporate ^{110m}Ag from the surrounding water and retain the

element even after moving to clean water that does not contain ^{110m}Ag [19]; however, among

the six sponge species analyzed in that study, this feature was only observed in *Acanthella*

acuta. In our study, four specimens of sponge incorporated relatively high ratios of

radiosilver relative to radiocesium (Fig. 3b arrowhead); therefore, we regarded these

sponges as a "silver accumulator" species.

Thus, macroalgae and macrobenthic species present varying capacities to incorporate and

retain radioisotopes based on their biological features. In addition, the

radiosilver/radiocesium ratios of these organisms are higher than those in the sediment.

Although we detected both radiocesium and radiosilver from almost all biological samples in

this research, an interesting exception was observed because radiosilver was not detected in

any of the biofilm samples, including biofilms collected from Area 2 (weeks 2, 8 and 9) (Table

2). This result suggested that microbes that compose biofilms specifically accumulate and

retain radiocesium but not radiosilver, even when the surface of nearby sediments contains

both radioisotopes.

Table 2. Radioactivity of the biofilm samples

		^{134}Cs Bq/kg dry	^{137}Cs Bq/kg dry	^{110m}Ag Bq/kg dry
Week1	Area1	1141.77	1655.17	nd
Week2	Area2	671.00	932.18	nd
Week3	Area4	306.90	429.21	nd
Week4	Area3	429.21	2353.58	nd
Week5	Area1	926.40	1436.73	nd
Week6	Area4'	538.56	829.08	nd
Week7	Area4	844.49	1543.05	nd
Week8	Area2	818.66	1607.65	nd

| Week9 | Area2 | 359.77 | 591.46 | nd |
| Week10 | Area3' | 190.29 | 294.20 | nd |

nd: not detected

"Cesium-philic" biofilm

Microbial communities that grow on surfaces of solid materials in the presence of water are

known as biofilms. Our observations suggest that the biofilm community specifically

accumulates radiocesium. Therefore, we analyzed the relationship between the microbial

consortia within the biofilms and their elemental composition.

Quantified element data were combined with radioactivity data for radiocesium and utilized

for the statistical analysis. Ribosomal RNA genes were sequenced and used to construct a

dataset for the statistical analysis.

A HCA from the dataset of elements and radiocesium (Fig. 4) in which an accumulation of

radioactive cesium, Cr, Ni, Sc, Rb, Li and cesium (Cs cluster) was performed for the week 1

sample. Another remarkable profile from the elemental analysis is observed in week 3,

where high relative amounts of lanthanide elements were measured (Lanth cluster).

Figure 4

Heatmap chart of the element and radioactivity data from the biofilm samples based on the

hierarchical clustering analysis (HCA). The heatmap chart was sorted by the HCA for both

axes (elements axis and sample axis). Boxes indicate clusters of radiocesium (Cs cluster) or

lanthanides (Lanth cluster).

A PCA score plot (Fig. 5a) of the merged dataset between the elements/radiocesium and the

microbial consortia showed that the week 1 sample was placed in the PC1 negative – PC2

positive direction and the week 3 sample was placed in the PC1 negative – PC2 negative

direction. As mentioned above, the week 1 sample contained high relative amounts of the Cs

cluster signals, and the week 3 sample contained high relative amounts of the Lanth cluster

signals. Consistent with those findings, the PCA loading plot showed that the Cs cluster was

placed in the PC1 negative – PC2 positive direction (Fig. 5b, Additional file 4:

Supplementary Table 4), and the lanthanide cluster was placed in the PC1 negative – PC2

negative direction (Fig. 5b, Additional file 5: Supplementary Table 5). The results of the

HCA and PCA analyses suggested that a "cesium-philic" biofilm occurred at the sampling

point in week 1 and a "lanthanide-philic" biofilm occurred at the sampling point in week 3.

Figure 5

PCA results among the sampling sites with microbial consortia data and element data with

radiocesium. Panel a is the score plot of the PCA among sampling sites. Each colored dot

indicates each sampling site. The color code is superimposed on the PCA score plot. Panel b

indicates the loading plot of the elements data. The upper panel shows the plot of elements

that contribute to the PC1 minus and PC2 plus direction, indicating that they contribute to

separate week 1 samples in the PC1 minus and PC2 plus direction. The lower panel shows

the plot of elements that contribute to the PC1 minus and PC2 minus direction, indicating

that they contribute to separate week 3 samples in the PC1 minus and PC2 minus direction

in the PCA score plot. Panel c indicates the loading plot of the microbial consortia data. The

upper panel displays the plot of OTUs that contribute to the PC1 minus and PC2 plus

direction, indicating that they contributes to separate week 1 samples in the PCA score plot.

The lower panel shows the plot of the OTUs that contribute to the PC1 minus and PC2

minus direction, indicating that they contribute to separate week 3 samples in the PCA

score plot.

The PCA results from both datasets of elements and microbes showed candidate microbes in

the "cesium-philic" and "lanthanide-philic" biofilms on the loading plots because those plots

were placed in same loading direction, e.g., PC1 negative – PC2 positive (Cs cluster) or PC1

negative – PC2 negative (Lanth cluster) (Fig. 5c, Additional file 4 and 5: Supplementary

Table 4 and 5). Indeed, the HCA heat map of the microbial consortia showed that the PC1

negative – PC2 negative-directed microbes mainly appeared in the week 1 column (Fig. 6

blue dots). The same tendency was observed with the Lanth cluster microbes (Fig. 6 green

dots). These results suggest that these two OTU sets are positively correlated with the

Lanth cluster (week 3 sample) and Cs cluster (week 1 sample). The correlation heat map

chart between the element dataset and the OTU dataset supports this result

Figure 6

Heatmap chart of microbial consortia data for biofilms with a hierarchical clustering

analysis (HCA).

The heatmap chart has been sorted by the HCA for both axes (OTU axis and sample axis).

Green and blue dots indicate contributing OTUs for the PC1 minus and PC2 plus direction

and the PC1/PC2 minus direction, respectively. The relative intensity value has been scaled

and centered in the row direction.

The radiocesium and Cs cluster elements are positively correlated to other multicellular

eukaryotes, such as *Hexacorallia* and *Oligochaete*. For bacteria, high relative intensities of

Bacteroidetes, *Rhodobacteraceae* and Gamma-proteobacteria were observed (Additoinal file

6: Supplementary Table 6 and Fig. 6). Although a primer set that did not target Archaea

was used, Archeal sequences were obtained in this work by sequencing amplicons generated

by the 16S ribosomal RNA primers set. An uncultured Crenarcheota in marine group I was

observed to be one of the dominant OTUs from the biofilm sample of week 1 (Additional file

6: Supplementary Table 6).

The function of these microorganisms is poorly understood. However, Alpha- and

Gamma-proteobacteria have been obtained from biofilms on the surface of nuclear fuel pools

[21]. According to Sarro et al. [21], biofilms obtained from nuclear fuel pools accumulate ^{60}Co.

Cobalt accumulation in the week 1 sample was suggested by the PCA (Fig. 5b and

Additional file 4: Supplementary Table 4) despite a lack of ^{60}Co signal indicated by

gamma-ray spectrography. The microbial community retrieved from the nuclear fuel pool

was described as capable of accumulating both ^{60}Co and ^{137}Cs as reported by Tisakova et al.

[22]. Our results suggest that the biofilm sampled in week 1 presents similar characteristics

to the microbes described by Sarro et al. and Tisakova et al.

Conclusions

In this study, we found heterogeneous distributions of radiocesium and radiosilver among a

variety of sample types from the coastal environment near the FDNPP disaster. Specific

organisms that use hemocyanin, such as shell fish and Ascidians, accumulate and retain

^{110m}Ag, and similar observations have been reported in terrestrial spiders. We also found

radiosilver in the sediment samples, suggesting that particular materials can be sources of

radiosilver for these organisms, even if particles filtered from the water have low levels of

radioactivity in our samples.

However, the biofilm samples did not display detectable signals of radiosilver, although the

amount of radiocesium in these samples was higher than that found in other marine

organisms. Biofilms may still act as an entry point for radiocesium to enter into the

ecosystem. In this paper, we could not assess the functional mechanisms by which biofilms

accumulate cesium specifically; thus, future research that includes cultivation experiments

and meta-transcriptome analyses should be performed to further investigate the microbial

communities described in this work and provide insights into this complex topic.

Declarations

Author's contributions

SM, HO and JS designed research. SM, HO, KK, HY and JS performed sampling. YK and

MI performed ICP-MS analysis and KS and JK partially performed sample preparation for

ICP-MS and advice for statistical analysis. Sample preparation and molecular biological

experiment was performed by SM and KK. Radioactivity measurement has been done by

HO. SCUBA diving safety control and sampling arrangement has been done by HY and JS.

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Competing interests

The authors declare no competing financial interests.

Availability of data and materials

The datasets used and/or analysed during the current study except for ribosomal RNA

sequences data and datasets that showed in manuscript or supplementary tables are

available from the corresponding author on reasonable request.

Ribosomal RNA sataset are available in the DDBJ sequence read archive with accession

number DRA004367.

Electronic supplementary material

Additional file 1: Supplementary Table 1: Radio activity of core samples. (XLSX 44 KB)

Additional file 2: Supplementary Table 2: Radio activity of macro algae samples. (XLSX 47

KB)

Additional file 3: Supplementary Table 3: Radio activity of macro benthos samples. (XLSX

46 KB)

Additional file 4: Supplementary Table 4: Loading plot on PC1 minus - PC2 plus. (XLSX 52

KB)

Additional file 5: Supplementary Table 5: oading plot results of PC1 minus - PC2 minus.

(XLSX 51 KB)

Additional file 6: Supplementary Table 6: Cs cluster taxon in Week1 sample. (XLSX 30 KB)

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Fig.1

Fig.2

a

b

Fig.3

a

b

Fig.4

Fig.5

Fig.6

Row direction scaled and centered relative abundance

-2   3