# Adaptive pathways of coral populations on the Great Barrier Reef

Mikhail V. Matz[1*], Eric A. Treml[2], Galina V. Aglyamova[1], Madeleine J. H. van Oppen[2,3]
and Line K. Bay[3]

[1] Department of Integrative Biology, University of Texas at Austin, 205 W 24th St. C0990,
Austin, Texas 78712, USA
[2] School of BioSciences, University of Melbourne, Victoria 3010 Australia
[3] Australian Institute of Marine Science, QLD, Australia

* Author for correspondence, matz@utexas.edu

**Abstract**

Global warming is predicted to drive preferential survival of warm adapted genotypes that have
migrated to cooler locations, and result in an overall decline in genetic diversity due to bleaching-
related mortality. Population genomic analysis of *Acropora millepora* on the Great Barrier Reef
(GBR) revealed that most populations were demographically distinct with preferential southward
migration from lower (warmer) to higher (cooler) latitudes. Still, no recent increase in southward
migration was detectable, and predicted migration rates remained closely correlated with those
derived from a biophysical model based on ocean currents. There was also no evidence of recent
declines in genetic diversity. A multi-locus adaptation model suggested that standing genetic
variation spread across latitudes might be sufficient to fuel continuous adaptation of *A. millepora*
metapopulation over 100-200 years of gradual warming. However, the model also predicted
increase in severity of local mortality events induced by thermal anomalies.

**One-sentence summary**: Patterns of migration and genetic diversity in a Great Barrier Reef coral
favor efficient adaptation to gradual warming but do not alleviate the threat of bleaching and
mortality from acute thermal extremes.

Hot water coral bleaching, caused by global warming, is devastating coral reefs around the world
(*1*) but there is room for hope if corals can adapt to increasing temperatures. Many coral species
have wide distributions that span environments that differ dramatically in their thermal regimes,
demonstrating that efficient thermal adaptation has occurred in the past (*2*). But can coral
adaptation keep up with the unprecedentedly rapid current rate of global warming (*3*)? One way
for corals to achieve rapid thermal adaptation is through genetic rescue, involving the spread of
existing heat tolerance alleles from low-latitude, warm-adapted populations to higher-latitude,
warming regions, via larval migration (*4, 5*). We have previously demonstrated the presence of
genetic variants conferring high thermal tolerance in a low-latitude *A. millepora* population (*4*). It
can be expected that global warming will cause preferential survival of warm-adapted poleward
migrants because they will be following their thermal optimum, whereas individuals migrating in

the opposite direction would find themselves in increasingly mismatched environments (Fig. 1 A,
B). Another likely population-level effect of recent declines in coral cover (*6*) is a reduction in
overall genetic diversity, potentially limiting both the scope and the rate of adaptation.

Here, we test these predictions in *Acropora millepora*, a common reef-building coral from the
most ecologically prominent and diverse coral genus in the Indo-Pacific (staghorn corals,
*Acropora*). We have analyzed genome-wide genetic variation using 2bRAD (*7*) in five
populations of *A. millepora* along the latitudinal range of the GBR (Fig. 1 A). We genotyped 18-
28 individuals per population at >98% accuracy and with a >95% genotyping rate. Analysis of
population structure based on ~11,500 biallelic SNPs separated by at least by 2,500 bases agreed
with previous microsatellites results (*8*, *9*), and revealed very low levels of genetic divergence,
with only the Keppel Islands population being potentially different from the others (Fig. 1 D and
Fig. S1). We observed increasing genetic divergence with geographical distance ("isolation by
distance", Fig. 1 C) that supports population divergence, however, pairwise $F_{ST}$ were small and
did not exceed 0.014 even between the southernmost and northernmost populations (Keppel and
Wilkie). To gain a deeper insight into coral demography, we used Diffusion Approximation for
Demographic Inference (*dadi,* (*10*)) to more rigorously test for population subdivision and infer
pairwise migration rates among populations and population sizes. *dadi* is a coalescent-based
method that optimizes parameters of a pre-specified demographic model to maximize the
likelihood of generating the observed allele frequency spectrum (for two populations it is
essentially a two-dimensional histogram of allele frequencies, Fig. S2). Being a likelihood-based
method, *dadi* can be used to compare alternative models using likelihood ratio tests and Akaike
Information Criterion (AIC).

We used AIC to confirm that our populations are separate demographic units. For each pair of
populations we generated 120 bootstrapped datasets by resampling genomic contigs and
performed delta-AIC comparison of two demographic models, a split-with-migration model and a
no-split model (Fig. S3 B). The split-with-migration model assumed two populations that have
split some time *T* in the past, potentially have different sizes *N1* and *N2*, and exchange migrants
at different rates (*m12* and *m21*) depending on direction. The no-split model allowed for ancestral
population size to change at time *T* but not for a population split, so the experimental data were
modeled as two random samples from the same population of size *N*. The majority of bootstrap
replicates (88-100%) showed AIC advantage of the split-with-migration model for all but one
pair of populations (Sudbury-Magnetic, 39% bootstrap support; Fig. S3). This indicates that the
populations are demographically distinct despite very low $F_{ST}$. This result highlights the power of
coalescent analysis relative to classical approaches (such as $F_{ST}$) that assume genetic equilibrium,
i.e., that populations have been stable for thousands of generations.

80
Figure 1. The population setting and background for our study. (A) Locations of sampled populations
where mean midsummer month sea surface temperature differed by up to ~3°C. Inset: *Acropora millepora*.
(B) Working hypothesis under global warming: Warm-adapted low-latitude genotypes that migrate to
higher latitudes would be following their physiological optimum and hence expected to survive better than
migrants in the opposite direction. (C) Increase of pairwise $F_{ST}$ with distance, both indicating weak genetic
divergence along the GBR, and (D) principal component analysis of genome-wide genetic variation. On
87           panel D, centroid labels are initial letters of population names as in panel A.

88
We then determined pairwise migration rates from the split-with-migration model and estimated
their confidence limits from bootstrap replicates. For all pairwise analyses except Wilkie-Sudbury
migration in southward direction exceeded northward migration, and this difference was
significant in seven out of nine cases (Fig. 2 A and Fig. S3A). Linear mixed model analysis of
direction dependent mean migration rates with a random effect of destination (to account for
variation in total migration rate) confirmed the overall significance of this southward trend
($P_{MCMC}$ <1e-4).

96
It is important to note that our pairwise migration rates captured the cumulative effect of genetic
exchange between populations, which included direct migration and the spread of alleles via other
stepping-stone populations. Such rates do not directly reflect the numbers of larvae exchanged
between populations but are very informative in the genetic rescue context. They represent the
per-generation rate of replacement of the destination population genotypes by genotypes from the
source population, which is essentially the rate at which genetic rescue could proceed.

103
To investigate whether the southward migration bias was due to higher survival of southward
migrants relative to northward migrants, as predicted under global warming (Fig. 1 B), we
developed a biophysical model of coral larval dispersal on the Great Barrier Reef. This model
quantified the per-generation migration potential among coral reef habitat patches in the GBR
based on ocean currents and parameters of larval biology (*11, 12*). We found that the genetic and

biophysical migration rates were very closely correlated (Mantel test: $r$ = 0.79, P= 0.008, Fig. 2
C). Although the biophysical model explained most of the southward migration bias in the
genetic data, the residuals were still in favor of southward migration (Fig. 2 C, inset; P = 0.058).
While this residual excess suggest preferential survival of southward migrants, as predicted by
our hypothesis (Fig. 1 B). These genetic predictions represent historical averages since the
populations split and did not resolve any potential recent migration changes.

115
116
<u>Figure 2</u>. Demography of *A. millepora* populations on the GBR. (A) Arc-plot of migration rates among
populations reconstructed from population genetic data. Inset: *dadi* model used: ancestral population splits
into two populations of unequal sizes (N1 and N2) some time T in the past, these populations exchange
migrants at different rates depending on direction. (B) Migration rates according to the biophysical model.
On panels A and B, the arcs should be read clockwise to tell the direction of migration; line thickness is
proportional to the migration rate. (C) Correlation between log-transformed biophysical and genetic
migration rates (Mantel r = 0.79, P = 0.008). Inset: box-plot of residuals from the linear regression.
Southward migration tends to exceed northward migration even after accounting for predictions of the
biophysical model (P =0.058), suggesting higher survival of southward migrants. (D) Box plot of effective
population sizes inferred by the split-with-migration model (panel A) across all population pairs and
bootstrap replicates. (E) Historical changes in effective population sizes inferred using a single-population
*dadi* model with two periods of exponential growth (*T0* and *T1*, reaching sizes *N0* and *N1*, inset), averaged
across bootstrap replicates.

130
To determine any recent changes in southward migration, we evaluated a similar basic split-with-
migration model (Fig. 2A) that allowed for a change in migration over the past 75-100 years. The
new model suggested some recent migration changes, but there was no consistent change between
northward and southward migration (Fig. S4). Delta-AIC bootstrap analysis favored the new

model over the basic one only for two pairs of populations, Wilkie-Orpheus and Wilkie-Magnetic
(85 and 60% bootstrap support, respectively). We conclude that with the current data and analysis
techniques we cannot yet detect the effect of recent warming on preferential direction of coral
migration along the GBR.
139
The GBR has already warmed by 0.8ºC since the end of last century (*13*) and may have already
reduced genetic diversity in *A.millepora* populations. We used *dadi* to infer effective population
sizes, which is a measure of genetic diversity and one of the key parameters determining the
population's adaptive potential (*14*).  The results of the split-with-migration model (Fig. 2 A)
were consistent for all population pairs and indicated that Keppel population was about one-fifth
the size of others (Fig. 2 D, E). This result was not surprising since the Keppel population
frequently suffers high mortality due to environmental disturbances and was therefore is expected
to show diminished long-term effective population size (*8*). We also used a single-population
*dadi* model that allowed for two consecutive growth/decline periods (Fig. 2 E, inset) to
reconstruct effective sizes of individual populations through time (Fig. 2 E and Fig S5). All
populations showed evidence of growth prior to the last glaciation, 500-20 thousand years ago
(Fig 2 E), which aligned well with the fossil record of rising dominance of *Acropora* corals on
Indo-Pacific reefs during this period (*15*). It has been suggested that the fast growth and early
sexual maturation of *Acropora* corals gave them an advantage relative to most other reef-building
corals during dynamic changes in the reef-forming zone due to the sea level changes
accompanying glacial cycles (*15*). Our results suggest that *A. millepora* populations have been in
stasis or slow decline since sea level changes abated (Fig. S5), although the inclusion of an
additional growth/decline period only improved the model fit significantly for the Keppel
population (Fig S6). None of the populations showed evidence of accelerated decline in effective
population size over the past few hundred years. Although our samples were collected in the
early-mid 2000s, our results are still relevant since they characterize populations only two-three
coral generations ago. Disturbances that have affected corals since then would not yet have
substantially impacted genetic diversity.
163
To evaluate whether standing genetic variation contributed by local thermal adaptation could
sustain evolution of the *A. millepora* metapopulation in response to warming, we have developed
a multi-QTL model of metapopulation adaptation in SLiM (*16*). The model was parameterized
with population sizes and migration rates inferred from the genetic analysis (Fig. 2 A, D), and
with differences in midsummer monthly mean temperature among populations (Fig. 1 A). The
number of QTLs and their effect sizes, phenotypic plasticity (standard deviation of the Gaussian
slope of fitness decline when phenotype mismatches the environment) and heritability (proportion
of phenotypic variation attributable to genetics) can all be varied in the model. It can also
incorporate climate scenarios with any combination of directional, cyclical and random changes.
The model also allows for new mutations but here the new mutation rate was set to zero. This was
to assess the contribution of only the standing genetic variation that was introduced into
populations at the start of simulation as random QTL effects. The climate scenario started with a
pre-adaptation to local thermal conditions for 2,000 generations. Assuming a generation time of

of 5 years in *A. millepora* (*17*) this corresponded to the period of stable temperature since the last
deglaciation. After pre-adaptation, the temperature was increased at a rate of 0.05°C/generation in
all populations, corresponding to the projected 0.1°C warming per decade (*18*). Throughout the
simulation temperature was allowed to fluctuate randomly between generations to approximate El
Nino Southern Oscillation (ENSO): the temperature deviations were drawn from a normal
distribution with a standard deviation of 0.25°C. The size of populations was kept constant
throughout the pre-adaptation period and scaled linearly with the populations' relative fitness
(mean current fitness divided by the mean fitness at the end of pre-adaptation period) during
warming. Migration rates from a population also scaled linearly with the population's fitness. In
this way, a population declining in fitness would shrink in size and stop contributing migrants to
other populations.
188
Our model suggested that, with only ten thermal QTLs, under all combinations of heritability and
plasticity the pre-adapted metapopulation would be able to persist through the warming for at
least 50-100 generations and, in some realistic cases, much longer (Fig. 3 and Figs. S7-S8).
Migration in general and southward migration in particular substantially contributed to this
persistence (Fig. 3 E, F), underscoring the importance of the spread of warm-adapted genotypes
from lower to higher latitudes (*4*).
195
Predictably, higher phenotypic plasticity promoted population persistence and stability against
random thermal anomalies, but we were rather surprised to observe a similar positive effect of
lower heritability, set to the values observed in coral quantitative genetics experiments (0.25-0.5,
(*17*); Fig 3, Fig. S7). One specific reason why corals are expected to show low heritability of
thermal tolerance is that much of natural variation in this trait in corals is due to the type of algal
symbionts (*Symbiodinium* spp.  (*19*)). Photo-symbionts are not transmitted from parent to
offspring in the majority of coral species (*20*), and although host genetics can have some effect on
the choice of *Symbiodinium* in the next generation (*21*) environment has a very strong effect on
this association (*19*, *22*). Higher persistence under low heritability and high plasticity is most
likely explained by the fact that they both allow for higher standing genetic variation to be
retained in populations (Fig. S9). During warming, this variation lasts longer as a source of
adaptive genetic variants, enabling up to 5°C increase in mean thermal tolerance over 150
generation (Fig. 3 G and Fig. S7). Higher plasticity partially rescued the drop in fitness due to low
heritability (Fig. 3 B and D, Fig. S7). Another notable tendency observed with all parameter
settings was that during warming the fitness (and hence the size) of adapting populations began to
fluctuate following random thermal anomalies, and the amplitude of these fitness fluctuations
increased as the warming progressed even though the amplitude of thermal anomalies did not
change (Fig. 3 H). These fluctuations correspond to severe mortality events induced by thermal
extremes due to ENSO and affected warm-adapted populations most, which very much resembles
the situation currently observed throughout the world (*1*).
216

217    Figure 3. Modeling coral metapopulation persistence under global warming. (A-D) Fitness of modeled
218
populations depending on heritability of thermal tolerance ($h^2$, proportion of tolerance variation explained
by genetics), phenotypic plasticity (σ, standard deviation of the Gaussian slope of fitness decline away from
the phenotypic optimum, in degrees C), and presence-absence of migration (E, F). On panels A-F, y-axis is
observed fitness relative to maximal fitness at the genetically determined optimum, averaged over all
individuals in a population. Warm-adapted populations (W and M) are shown as red-tint traces, populations
from mild thermal regime (S and O) are green-tint traces, and the cool-adapted population (K) is the blue
trace. Note nearly complete overlap between traces for pairs of populations pre-adapted to the same
temperature (W,M and S,O). (G) Thermal tolerances of evolving populations. Thin noisy lines are modeled
temperatures at different locations. (H) Modeled random temperature anomalies (grey line) and fluctuations
in populations' fitness (the colored lines are residuals from loess regression over fitness traces on panel D;
Wilkie: orange line, Keppel: blue line). Note the inverse sign of temperature anomalies: this more clearly
shows the correspondence between rise in temperature and drop in fitness in the next generation. As
warming progresses, populations (especially originally warm-adapted ones) become increasingly sensitive
232                           to random temperature fluctuations.

There are several uncertainties in our model associated with coral biology. Higher number of
QTLs and/or their larger effect sizes would promote higher genetic variation and lead to longer
population persistence. To keep the analysis conservative, our model included only ten QTLs,
which is likely much fewer that the actual number of thermal QTLs in acroporid corals (20). We
also kept the distribution of QTL effect sizes narrow: with the current settings and ten QTLs, at
the start of simulation only about 2% of corals deviated from the mean thermal tolerance by more
than 1.5°C in either direction. Such narrow variation makes adaptation to the thermal gradient of
~3°C along the GBR non-trivial, but still, at present there is no experimental data to evaluate

whether even such narrow variation is realistic. Our model was also conservative in using
effective population sizes suggested by genetic analysis as census sizes. In highly fecund marine
organisms census sizes tend to substantially exceed effective population sizes, sometimes by
orders of magnitude (*23*), which would strongly promote higher genetic diversity and population
persistence. Moreover, we modeled only our five populations rather than the whole GBR, which
would have resulted in much higher standing genetic variation in the metapopulation, promoting
longer persistence.
249
As for phenotypic plasticity, in simulations shown on Fig. 3, $\sigma = 0.5$ and $\sigma = 1$ corresponded to
86% and 40% decline in fitness if the individual's phenotype mismatched the environment by
$1^{o}C$. The existing data on the issue of coral thermal plasticity are somewhat conflicting. One
study shows that acroporid corals can successfully acclimatize to environments differing in
maximum temperatures by as much as $2^{o}C$ (*24*); however, another study found that coral grew
52-80% more slowly when transplanted among locations differing by 1.5ºC average temperature,
(*25*). Although it is not possible to directly place these results into our quantitative plasticity
framework, the former study supports the higher plasticity setting ($\sigma = 1$) while the latter study
supports $\sigma = 0.5$. It must also be noted that both these studies involved *in situ* transplantations and
hence the effect of temperature remains confounded with other local fitness-affecting
environmental parameters. Also, in adult corals plasticity is likely lower that in larvae and
recruits, which are expected to exhibit non-reversible developmental plasticity associated with
metamorphosis and establishment within a novel environment (*26*). Future experiments that
expose multiple genetically distinct coral individuals to a range of temperatures under controlled
laboratory settings are required to rigorously quantify variation in thermal optima and plasticity in
natural populations.
266
In conclusion, we found that genetic diversity and migration patterns of our study species were
not yet affected by global warming and were well positioned to facilitate persistence of the GBR
metapopulation for a century or more. However, local mortality events induced by thermal
anomalies will be increasingly severe, especially among the originally warm-adapted populations.
The 10-85% mortality in the Northern GBR as a result of 2016 bleaching event (*27*) could be a
particularly sobering recent manifestation of this trend.
273
More research into phenotypic plasticity and genetic variation in coral thermal tolerance and its
genetic architecture (number of QTLs and their effect sizes) is needed to further improve the
predictive power of our model. The estimated migration in the order of 10 - 100 migrants per
generation could be feasibly facilitated by assisted gene flow efforts (*28*) without risking
disruption of the natural local adaptation patterns (*29*). Corals are declining on reef world-wide
and there is an urgent need to develop new solutions to effectively manage the impacts of global
processes such as climate change at local management scales. The broad characterization of
genetic diversity, local thermal adaptation and migration pathways in multiple reef-building coral
species would greatly inform both traditional spatial management and novel assisted gene flow
approaches and should therefore be given high priority.

284
285 **Methods**
286
287 *Genotyping*
288
This study relied predominantly on samples described by van Oppen et al (*9*), with addition of
several samples from Orpheus and Keppel islands that were used in the reciprocal transplantation
experiment described by Dixon et al (*30*). The samples were genotyped using 2bRAD (*7*)
modified for Illumina sequencing platform; the latest laboratory and bioinformatics protocols are
available at https://github.com/z0on/2bRAD_GATK. BcgI restriction enzyme was used and the
samples retained for this analysis had 2.3-20.2 (median: 7.45) million reads after trimming and
quality filtering (no duplicate removal was yet implemented in this 2bRAD version). The reads
were mapped to the genome of the outgroup species, *Acropora digitifera* (*31*, *32*), to polarize the
allelic states into ancestral (as in *A. digitifera*) and derived, e.g., (*33*, *34*). Genotypes were called
using GATK pipeline (*35*).
299
Preliminary analysis of sample relatedness using vcftools (*36*) revealed that our samples included
several clones: four repeats of the same genotype from the Keppel Island (van Oppen et al (*9*)
samples K210, K212, K213 and K216), another duplicated genotype from Keppel (samples K211
and K219), and one duplicated genotype from Magnetic Island (samples M16 and M17). All
other samples were unrelated. We took advantage of these clonal replicates to extract SNPs that
were genotyped with 100% reproducibility across replicates and, in addition, appeared as
heterozygotes in at least two replicate pairs (script replicatesMatch.pl with hetPairs=2 option).
These 7,904 SNPs were used as "true" SNP dataset to train the error model to recalibrate variant
quality scores at the last stage of the GATK pipeline. During recalibration, we used the transition-
transversion (Ts/Tv) ratio of 1.438 determined from the "true" SNPs to assess the number of false
positives at each filtering threshold (as it is expected that an increase of false positive calls would
decrease the Ts/Tv ratio towards unity). We chose the 95% tranche, with novel Ts/Tv = 1.451.
After quality filtering that restricted the calls to only bi-allelic polymorphic sites, retained only
loci genotyped in 95% or more of all individuals, and removed loci with the fraction of
heterozygotes exceeding 0.6 (possible lumped paralogs), we ended up with 25,090 SNPs. In total,
2bRAD tags interrogated 0.18% of the genome. The genotyping accuracy was assessed based on
the match between genotyped replicates using script repMatchStats.pl. Overall agreement
between replicates was 98.7% or better with the heterozygote discovery rate (fraction of matching
heterozygote calls among replicates) exceeding 96%.
319
320 *Genome-wide genetic divergence*
321
To begin to characterize genome-wide divergence between populations we used pairwise
genome-wide Weir and Cockerham's $F_{ST}$ calculated by vcftools (*36*), principal component
analysis (PCA) using R package adegenet (*37*), and ADMIXTURE (*38*). For PCA and
ADMIXTURE, the data were thinned to keep SNPs separated by 5kb on average and by at least

2.5 kb, choosing SNPs with highest minor allele frequency (script thinner.pl with options
'interval=5000 criterion=maxAF').
328
329    *Demographic analysis and bootstrapping*
330
Prior to demographic analysis, Bayescan (*39*) was used to identify sites potentially under
divergent selection among populations, and 13 such sites with q-value <0.05 were removed.
Demographic models were fitted to 120 bootstrapped datasets, which were generated in two
stages. First, five alternatively thinned datasets were generated for which SNPs were randomly
drawn to be on average 5 kb apart and not closer than 2.5 kb. This time the SNPs were drawn at
random to avoid distorting the allele frequency spectrum unlike thinning for PCA and
ADMIXTURE where the highest minor allele frequency SNPs were selected. Then, 20
bootstrapped replicates were generated for each thinned dataset by resampling contigs of the
reference genome with replacement (script dadiBoot.pl). The fitted model parameters were
summarized after excluding bootstrap replicates that fell into the lowest 15% likelihood quantile
and the ones where model fitting failed to converge, leading to some parameters being
undetermined or at infinity (less than 10% of total number of runs). Delta-AIC values were
calculated for each bootstrap replicate that passed these criteria for both compared models, and
summarized to obtain bootstrap support value, the percentage of replicates favoring the
alternative model. While fitting *dadi* models, the data for each population were projected to
sample sizes maximizing the number of segregating sites in the analysis, resulting in 7000-8172
segregating sites per population.
348
349    *Unit conversion*
350
To convert *dadi*-reported coalescent parameter values ($\theta$, *T* and *M*) into time in years (*t*), effective
population sizes in number of individuals (*Ne*) and migration rates as fraction of new immigrants
per generation (*m*), we estimated the mutation rate ($\mu$) from the time-resolved phylogeny of
*Acorpora* genus based on *paxC* intron (*40*), at 4e-9 per base per year. Although *A. millepora* was
shown to start reproducing in 3 years (*17*) we assumed the generation time of 5 years reasoning
that it would better reflect the attainment of full reproductive potential as the colony grows.
Assuming a genome size of 5e+8 bases (*31*) the number of new mutations per genome per
generation is 10. Since the fraction 2bRAD-sequenced genome in our experiment was 1.8e-3, the
mutation rate per 2bRAD-sequenced genome fraction per generation is $\mu$ = 0.018. This value was
used to obtain:
-    Ancestral effective population size: $Ne = \theta / 2\mu$
-    Migration rate: $m = M / 2Ne$
-    Time in years: $t = 2TNe \cdot 5$
364
365    *Biophysical model*
366
A spatially-explicit biophysical modeling framework (*11*, *41*) was used to quantify migration

between coral reef habitats of the broader region surrounding the Great Barrier Reef, thereby
revealing the location, strength, and structure of a species' potential population connectivity. The
model's spatial resolution of ca. 8 km coincides with hydrodynamic data for the broader region
(1/12.5 degree; HYCOM+NCODA Reanalysis and Analysis product; hycom.org). Our
biophysical dispersal model relies on geographic data describing the seascape environment and
biological parameters capturing coral-specific life-histories. Coral reef habitat data are available
from the UNEP World Conservation Monitoring Centre (UNEP-WCMC; http://data.unep-
wcmc.org/datasets/1) representing a globally-consistent and up-to-date representation of coral
reef habitat. To capture specific inter-annual variability, two decades of hydrodynamic data were
used from 1992 to 2013 (*42*).
378
Coral-specific biological parameters for *A. millipora* included relative adult density (dependent
on the habitat), reproductive output, larval spawning time and periodicity (e.g., Magnetic Island
populations spawn a month earlier than the other GBR sites (*43*)), maximum dispersal duration,
pre-competency and competency periods, and larval mortality (*44*, *45*). The spatially explicit
dispersal simulations model the dispersal kernel (2-D surface) as a 'cloud' of larvae, allowing it
to be concentrated and/or dispersed as defined by the bio-physical parameters. An advection
transport algorithm is used for moving larvae within the flow fields (*46*).
386
Simulations were carried out by releasing a cloud of larvae into the model seascape at all
individual coral reef habitat patches and allowing the larvae to be transported downstream by the
currents. Ocean current velocities, turbulent diffusion, and larval behavior move the larvae
through the seascape at each time-step. Larval competency, behavior, density, and mortality
determine when and what proportion of larvae settle in habitat cells at each time step. When
larvae encounter habitat, the concentration of larvae settling with the habitat is recorded at that
time-step. From the dispersal data, we derived the coral migration matrix representing the
proportion of settlers to each destination patch that came from a source patch, which is analogous
to the source distribution matrix (*47*) and is equivalent to migration matrices derived from
population genetic analysis. It is important to note that migration matrices extracted for the field
sites represent the potential migration through all possible stepping-stones.
398
399 *Metapopulation adaptation model*
400
The model was implemented in SLiM the forward evolutionary simulator, by modifying the
provided recipe "Quantitative genetics and phenotypically-based fitness". The model simulates
Fisher-Wright populations with discreet generations. At the start of the simulation, populations
were established at specified population sizes and pairwise migration rates (genetic replacement
rates), and all QTLs in all individuals were given a mutation with the effect size drawn from a
normal distribution with mean zero and specified standard deviation, to create standing genetic
variation. The phenotype of each individual was calculated as the sum of QTL effects plus
random noise to simulate desired heritability. Then, fitness of each individual was calculated
based on the difference between the individual's phenotype (thermal optimum), temperature of

the environment, and the setting for phenotypic plasticity, modeled as the standard deviation of
the Gaussian slope of fitness decline with increasing distance between phenotype and
environment. Then, parents were chosen to produce the next generation according to their fitness;
parents for immigrant individuals are chosen from among individuals in the source population.
New mutations at QTLs happened at the specified rate when transitioning to the next generation
and the effect of a new mutation replaced the previous QTL effect.
416
417  Our code was designed for general modeling of multilocus adaptation in metapopulations and can
418  be customized easily. It is possible to adjust:
419
-   Number of populations, their sizes, and pairwise migration rates. We modeled our five
populations with effective population sizes and pairwise migration rates inferred by *dadi*.
-   Number of QTLs and the distribution of their effect sizes. To keep the model conservative,
we modeled only ten QTLs with normal distribution of effect sizes with a standard deviation
of $0.2^{o}$C. With ten QTLs, this setting implied that at the start of simulation only about 2% of
corals deviated from mean thermal tolerance by more than $1.5^{o}$C in either direction. Since
thermal differences between our populations exceeded $3^{o}$C, this narrow variation made local
adaptation rather non-trivial.
-   Dominance of QTLs (set to 0.5 in our simulation).
-   Phenotypic plasticity. We modeled three plasticity settings, 0.5, 1 and 2, which corresponded
to 86%, 40% and 13% fitness drop when the individual's phenotypic optimum (calculated
based on QTLs and heritability setting) mismatched the environment by $1^{o}$C.
-   Heritability (proportion of phenotypic variation explained by genetics). We examined values
1, 0.5, 0.25 and 1e-5, the latter to confirm that no adaptation or evolution was observed when
the trait was not heritable.
-   Mutation rate, which was set to zero because we wanted to explore only the role of standing
genetic variation.
-   Environmental changes, modeled as identical trends across populations with population-
specific offsets. The trends can be any combination of linear, cyclical and random
components. In the current simulation, environmental trends were offset by $+1.6^{o}$C in Wilkie
and Magnetic populations and by $-1.8^{o}$C in the Keppel population, to model differences in
midsummer monthly mean temperature among populations (Fig. 1). The trends included
only random thermal anomalies (drawn from a normal distribution with a standard deviation
of $0.25^{o}$C, to approximate ENSO events) for the first 2000 generations, after which a linear
increase at $0.05^{o}$C per generation was added to simulate warming.
445
446  To better model population dynamics, we implemented linear scaling of the population size and
447  immigration rates with the population's mean fitness post pre-adaptation period (the initial 2000
448  generations with no linear change in environment). In this way, a population declining in fitness
449  shrinks in size and stops contributing migrants to other populations.
450
All combinations of parameter settings were run ten times to ensure consistency. We found that

with population sizes in thousands, such as in our case, the results were very consistent among

independent runs. We therefore did not aggregate results over many replicated runs but show one

randomly chosen run for each tested parameter combination.

**Acknowledgements**

We wish to thank Ryan Gutenkunst and Benjamin Haller for their continuous support of *dadi* and

SLiM users, respectively. The bioinformatics analysis was accomplished using computational

resources of the Texas Advanced Computer Center. This study has been supported by NSF

(DEB-1054766) grant to M.V.M, ARC (LP120200245) and University of Melbourne ECR grants

to E. A.T., a Coral Reef Alliance grant ("Coral Adaptation Challenge") to E.A.T and M.V.M,

Queensland Government funding to L.K.B and AIMS funding to L.K.B. and M.J.V.O

**Data and code availability**

The finalized genotyping dataset in VCF format, detailed bioinformatic walkthrough, accessory

formatting and plotting scripts, *dadi* scripts and the SLiM model code are available from

https://github.com/z0on/Adaptive-pathways-of-coral-populations-on-the-Great-Barrier-Reef. Raw

sequencing data has been deposited to National Center for Biotechnology Information's Short

Reads Archive (accession number pending).

**References**

1. A. C. Baker, P. W. Glynn, B. Riegl, Climate change and coral reef bleaching: An ecological assessment of long-term impacts, recovery trends and future outlook. *Estuar. Coast. Shelf Sci.* **80**, 435–471 (2008).

2. T. P. Hughes *et al.*, Climate change, human impacts, and the resilience of coral reefs. *Science.* **301**, 929–933 (2003).

3. J. M. Pandolfi, S. R. Connolly, D. J. Marshall, A. L. Cohen, Projecting coral reef futures under global warming and ocean acidification. *Science.* **333**, 418–22 (2011).

4. G. B. B. Dixon *et al.*, Genomic determinants of coral heat tolerance across latitudes. *Science.* **348**, 1460–1462 (2015).

5. P. K. Ingvarsson, Restoration of genetic variation lost – the genetic rescue hypothesis. *Trends Ecol. Evol.* **16**, 62–63 (2001).

6. G. De'ath, K. E. Fabricius, H. Sweatman, M. Puotinen, The 27-year decline of coral cover on the Great Barrier Reef and its causes. *Proc. Natl. Acad. Sci. U. S. A.* **109**, 17995–9 (2012).

7. S. Wang, E. Meyer, J. K. McKay, M. V Matz, 2b-RAD: a simple and flexible method for genome-wide genotyping. *Nat. Methods.* **9**, 808–810 (2012).

8. M. J. H. van Oppen, V. Lukoschek, R. Berkelmans, L. M. Peplow, A. M. Jones, A population genetic assessment of coral recovery on highly disturbed reefs of the Keppel Island archipelago in the southern Great Barrier Reef. *PeerJ.* **3**, e1092 (2015).

9. M. J. H. Van Oppen, L. M. Peplow, S. Kininmonth, R. Berkelmans, Historical and contemporary factors shape the population genetic structure of the broadcast spawning coral, Acropora millepora, on the Great Barrier Reef. *Mol. Ecol.* **20**, 4899–4914 (2011).

10. R. N. Gutenkunst, R. D. Hernandez, S. H. Williamson, C. D. Bustamante, Inferring the joint demographic history of multiple populations from multidimensional SNP frequency data. *PLoS Genet.* **5**, e1000695 (2009).

11. E. A. Treml, J. Roberts, P. N. Halpin, H. P. Possingham, C. Riginos, The emergent geography of biophysical dispersal barriers across the Indo-West Pacific. *Divers. Distrib.* **21**, 465–476 (2015).

12. E. A. Treml *et al.*, Reproductive Output and Duration of the Pelagic Larval Stage Determine Seascape-Wide Connectivity of Marine Populations. *Integr. Comp. Biol.* **52**, 525–537 (2012).

13. J. Lough, A. Hobday, Observed climate change in Australian marine and freshwater environments. *Mar. Freshw. Res.* **62**, 984–999 (2011).

14. B. Charlesworth, Fundamental concepts in genetics: Effective population size and patterns of molecular evolution and variation. *Nat. Rev. Genet.* **10**, 195–205 (2009).

15. W. Renema *et al.*, Are coral reefs victims of their own past success? *Sci. Adv.* **2**, e1500850 (2016).

16. B. C. Haller, P. W. Messer, SLiM 2: Flexible, interactive forward genetic simulations. *Mol. Biol. Evol.* **34**, 230–240 (2017).

17. M. Vanessa, B. Baria, D. W. Dela Cruz, R. D. Villanueva, J. R. Guest, Spawning of three-year-old Acropora millepora corals reared from larvae in Northern Philippines. *Bull. Mar. Sci.* **88**, 61–62 (2012).

18. IPCC, *Climate Change 2007 - The Physical Science Basis. Contribution of Working Group I to the Fourth Assessment Report of the IPCC* (Cambridge University Press, New York, NY, 2007).

19. R. Berkelmans, M. J. H. van Oppen, The role of zooxanthellae in the thermal tolerance of corals: a "nugget of hope" for coral reefs in an era of climate change. *Proc. R. Soc. B-Biological Sci.* **273**, 2305–2312 (2006).

20. A. H. Baird, J. R. Guest, B. L. Willis, Systematic and biogeographical patterns in the reproductive biology of scleractinian corals. *Annu. Rev. Ecol. Evol. Syst.* **40**, 551–571 (2009).

21. K. Quigley, B. Willis, L. Bay, Heritability of the Symbiodinium community in vertically- and horizontally-transmitting broadcast spawning corals. *bioRxiv.* doi: https://doi.org/10.1101/100453 (2017).

22. E. J. Howells *et al.*, Coral thermal tolerance shaped by local adaptation of photosymbionts. *Nat. Clim. Chang.* **2** (2011), pp. 116–120.

23. F. P. Palstra, D. J. Fraser, Effective/census population size ratio estimation: a compendium and

appraisal. *Ecol. Evol.* **2**, 2357–2365 (2012).

24.   S. R. Palumbi, D. J. Barshis, N. Traylor-Knowles, R. A. Bay, Mechanisms of reef coral resistance
to future climate change. *Science*. **344**, 895–8 (2014).

25.   E. J. Howells, R. Berkelmans, M. J. H. van Oppen, B. L. Willis, L. K. Bay, Historical thermal
regimes define limits to coral acclimatization. *Ecology*. **94**, 1078–1088 (2013).

26.   D. Nettle, M. Bateson, Adaptive developmental plasticity: what is it, how can we recognize it and
when can it evolve? *Proc. R. Soc. B Biol. Sci.* **282**, 20151005 (2015).

27.   Great Barrier Reef Marine Park Authority, *Interim report on the environmental impacts of the 2016*
*coral bleaching event*.

28.   O. Hoegh-Guldberg *et al.*, Assisted colonization and rapid climate change. *Science*. **321**, 345–346
(2008).

29.   S. N. Aitken, M. C. Whitlock, Assisted Gene Flow to Facilitate Local Adaptation to Climate
Change. *Annu. Rev. Ecol. Evol. Syst.* **44**, 367–388 (2013).

30.   G. B. Dixon, L. K. Bay, M. V. Matz, Bimodal signatures of germline methylation are linked with
gene expression plasticity in the coral Acropora millepora. *BMC Genomics*. **15**, 1109 (2014).

31.   C. Shinzato *et al.*, Using the Acropora digitifera genome to understand coral responses to
environmental change. *Nature*. **476**, 320-U82 (2011).

32.   M. J. H. van Oppen, B. J. McDonald, B. Willis, D. J. Miller, The Evolutionary History of the Coral
Genus Acropora (Scleractinia, Cnidaria) Based on a Mitochondrial and a Nuclear Marker:
Reticulation, Incomplete Lineage Sorting, or Morphological Convergence? *Mol. Biol. Evol.* **18**,
1315–1329 (2001).

33.   B. F. Voight, S. Kudaravalli, X. Wen, J. K. Pritchard, A map of recent positive selection in the
human genome. *PLoS Biol.* **4**, e72 (2006).

34.   I. K. Jordan *et al.*, A universal trend of amino acid gain and loss in protein evolution. *Nature*. **433**,
633–8 (2005).

35.   A. Mckenna *et al.*, The Genome Analysis Toolkit: A MapReduce framework for analyzing next-
generation DNA sequencing data. *Genome Res.* **20**, 1297–1303 (2010).

36.   P. Danecek *et al.*, The variant call format and VCFtools. *Bioinformatics*. **27**, 2156–8 (2011).

37.   T. Jombart, adegenet: a R package for the multivariate analysis of genetic markers. *Bioinformatics*.
**24**, 1403–5 (2008).

38.   D. H. Alexander, J. Novembre, K. Lange, Fast model-based estimation of ancestry in unrelated
individuals. *Genome Res.* **19**, 1655–64 (2009).

39.   T. Günther, G. Coop, Robust identification of local adaptation from allele frequencies. *Genetics*.
**195**, 205–20 (2013).

40.   Z. T. Richards, D. J. Miller, C. C. Wallace, Molecular phylogenetics of geographically restricted
Acropora species: Implications for threatened species conservation. *Mol. Phylogenet. Evol.* **69**,
837–851 (2013).

41.   E. A. Treml, P. N. Halpin, D. L. Urban, L. F. Pratson, Modeling population connectivity by ocean
currents, a graph-theoretic approach for marine conservation. *Landsc. Ecol.* **23**, 19–36 (2008).

42.   E. P. Chassignet *et al.*, The HYCOM (HYbrid Coordinate Ocean Model) data assimilative system.
*J. Mar. Syst.* **65**, 60–83 (2007).

43.   R. C. Babcock *et al.*, Synchronous Spawnings of 105 Scleractinian Coral Species on the Great-
Barrier-Reef. *Mar. Biol.* **90**, 379–394 (1986).

44.   S. W. Davies, E. A. Treml, C. D. Kenkel, M. V. Matz, Exploring the role of Micronesian islands in
the maintenance of coral genetic diversity in the Pacific Ocean. *Mol. Ecol.* **24**, 70–82 (2015).

45.   S. R. Connolly, A. H. Baird, Estimating dispersal potential for marine larvae: dynamic models
applied to scleractinian corals. *Ecology*. **91**, 3572–3583 (2010).

46.   P. K. Smolarkiewicz, J. Szmelter, An MPDATA-based solver for compressible flows. *Int. J.*
*Numer. Methods Fluids*. **56**, 1529–1534 (2008).

47.   R. K. Cowen, G. Gawarkiewicz, J. Pineda, S. Thorrold, F. Werner, Population Connectivity in
Marine Systems: An Overview. *Oceanography*. **20**, 14–21 (2007).

 **Supplemental Figures**

581     <u>Figure S1</u>. ADMIXTURE analysis of genetic differentiation between populations. (A) Map of sampled

582     locations with one-letter population identifiers. (B) ADMIXTURE plot of ancestry proportions with $K = 2$

583     (optimal $K$ was 1).

590     <u>Figure S2</u>. Example of two-population *dadi* model fit. (A) The model: ancestral population splits into two

591     populations of unequal sizes (N1 and N2) some time T in the past, which exchange migrants with different

592     rates depending on direction. (B) Observed allele frequency spectrum comparing Wilkie (W) and Keppel

593     (K) populations. (C) Allele frequency spectrum generated by the fitted model. (D, E) Map and histogram of

594     residuals (absolute scale).

599
<u>Figure S3</u>. Bootstrap analysis of migration rates and population subdivision using *dadi*. (A) Migration
among population pairs, with bootstrap-derived 95% confidence intervals. The pairs are identified on the x-
axis and sorted by increasing geographical distance. Black bars – southward migration, grey bars –
northward migration. (B) Models being compared: the full model (left) implies populations' split into two
different sizes (N1 and N2) at time T in the past, since when they exchanged migrants at unequal rates
depending on direction. Reduced model allows for population size change at time T in the past but does not
include population split: the two genotyped groups (p1 and p2) are regarded as two samples from the same
population. (C-L) Histograms of delta-AIC values for 100 bootstrap replicates (bootstrap was performed
over genomic contigs of the draft genome of *A. digitifera*. Positive numbers indicate support for the full
model. The letters on top of each panel identify compared populations, the number is the proportion of
positive bootstrap replicates (i.e., bootstrap support for the full model). The only comparison that did not
receive >50% bootstrap support for population split is between S and M populations (panel H).
612

613
614
615
616

617
<u>Figure S4.</u> Migration rates inferred by the *dadi* model allowing for the change in migration rates over the
last 0.01 T units (15-20 generations or 75-100 years, in our case). Box plots show historical (N, S) and
recent (Nr, Sr) migration rates inferred among pairs of population across 100 bootstrap replicates. Numbers
in the top left corner of the WO and WM plots are delta-AIC bootstrap support values for the model with
the recent change in migration when compared to the split-with-migration model with no recent change
(Fig. 1A). All other pairs had less than 50% delta-AIC bootstrap support. There is no consistent recent
change in the preferential direction of migration.
625
626
627
628

629
630     Figure S5. Population history. (A-E) Historical population sizes with bootstrap-derived 95% confidence
631    intervals, according to the two-growth model (Fig. S6 A). (F) Sea level with shaded area corresponding to
632                  standard error (*41*).

634
Figure S6. Delta-AIC bootstrap analysis of single-population models. (A) Models compared. The full
model (left) includes two exponential growth periods (any of which could be growth or decline), the
reduced model (right) has only one growth period. (B-F) Histograms of delta-AIC values for 100 bootstrap
replicates. Positive numbers indicate support for the full model. The letter on top of each panel identify the
population, the number is the proportion of positive bootstrap replicates (i.e., bootstrap support for the full
model). The two-growth model is strongly supported for population K (panel F) and marginally supported
641                     for population W (panel B).

642
643

644
<u>Figure S7</u>. Fitness of modeled populations after pre-adaptation period and under warming, depending on
heritability of thermal tolerance ($h^2$, proportion of phenotypic variation explained by genetics) and
phenotypic plasticity ($\sigma$, standard deviation of the Gaussian slope of fitness decline away from the
phenotypic optimum, in degrees C). X-axis is generations; warming starts at generation 0. Y-axis is fitness
relative to maximal fitness at the genetically determined optimum. Warm-adapted populations (W and M)
are shown as red-tint traces, populations from mild thermal regime (S and O) are green-tint traces, and the
cool-adapted population (K) is the blue trace. Pairs of traces for warm- and mild-adapted populations
largely overlap. (A, C, E): $h^2$=1. (B, D, F): $h^2$=0.25. (A, B): $\sigma$ = 0.5. (C, D): $\sigma$ = 1. (E, F): $\sigma$ = 2.  Higher
plasticity facilitates metapopulation persistence during warming and confers stability against random
fluctuations. Higher plasticity also partially rescues the drop in fitness achievable under low heritability
655                 (compare pre-warming generations, from -100 to 0, on panels B, D and F).
656

658
Figure S8. Higher plasticity and lower heritability promote longer and more extensive evolution in response
to warming. The graphs show mean thermal tolerance of modeled populations after pre-adaptation period
and under warming, depending on heritability of thermal tolerance ($h^2$, proportion of phenotypic variation
explained by genetics) and phenotypic plasticity ($\sigma$, standard deviation of the Gaussian slope of fitness
decline away from the phenotypic optimum, in degrees C). X-axis is generations; warming starts at
generation 0. Y-axis is thermal tolerance (mean phenotype of the population). Warm-adapted populations
(W and M) are shown as red-tint traces, populations from mild thermal regime (S and O) are green-tint
traces, and the cool-adapted population (K) is the blue trace. Thin noisy lines are modeled temperatures at
the corresponding locations. Pairs of traces for warm- and mild-adapted populations largely overlap. (A, C,
668            E): $h^2$=1. (B, D, F): $h^2$=0.25. (A, B): $\sigma$ = 0.5. (C, D): $\sigma$ = 1. (E, F): $\sigma$ = 2.
669
670

<u>Figure S9</u>. Higher plasticity (σ) and lower heritability ($h^2$) promote retention of higher genetic variation in
thermal tolerance. The scatterplots show the dependence of the number of individuals in a population
bearing a mutation at a thermal QTL locus on the mutation's effect size (change in thermal tolerance, in °C)
at the end of the pre-adaptation period (2000 generations with no directional change in temperature). The
starting standing genetic variation was the same in all simulations. (A,D,E): $h^2$=1. (B,E,H): $h^2$=0.5. (C,F,I):
$h^2$=0.25.  (A-C): σ = 0.5. (D-F): σ = 1. (G-I): σ = 2. Populations are colored according to the color scheme
used in Figures 3, S9 and S10 (see legend).