

Adaptive pathways of coral populations on the Great Barrier Reef

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Abstract

Reef-building corals are extremely important for maintenance of marine biodiversity and coastal economy and are currently under severe threat from anthropogenic warming. Warming is predicted to drive preferential survival of warm-adapted genotypes that have migrated to cooler locations and result in an overall decline in genetic diversity due to bleaching-related mortality. To quantify these trends, we analyzed five populations of a common coral *Acropora millepora* along the latitudinal extent of the Great Barrier Reef (GBR). Population genomic analysis revealed that most populations were demographically distinct and that migration was indeed preferential southward, from lower (warmer) to higher (cooler) latitudes. However, no recent increase in southward migration was detectable, and inferred migration rates remained closely correlated with predictions of a biophysical model of larval dispersal based on ocean currents. There was also no evidence of recent declines in genetic diversity. A multi-locus adaptation model indicated that standing genetic variation spread across latitudes could be sufficient to fuel continuous adaptation of *A. millepora* metapopulation to warming over the next 100-200 years. Unexpectedly, we found that naturally low heritability of thermal tolerance in reef-building corals due to contribution from horizontally transmitted algal symbionts would facilitate longer metapopulation persistence. Still, despite good prospects for gradual adaptation, our model predicted increase in severity of mortality events due to random thermal anomalies, which could lead to much faster coral extinction if there are ecological feedbacks preventing rapid reef recovery.

Significance statement: Can long-lived organisms such as reef-building corals adapt fast enough to keep up with the historically unprecedented rate of sea surface warming? Here we combine population genomics, biophysical modeling, and evolutionary simulations to argue that populations of a common reef-building coral (*Acropora millepora*) spread across latitudes on the Great Barrier Reef could harbor sufficient genetic variation to fuel efficient adaptation to increasing temperature for another century and perhaps longer. However, corals will be increasingly more sensitive to extreme heat waves despite ongoing adaptation to gradual warming, which could precipitate their extinction much sooner. Our study underscores the key

42 role of standing genetic variation in the future persistence of coral reefs and calls for novel reef
43 management strategies to facilitate natural adaptation process.

44

45 Hot water coral bleaching, caused by global warming, is devastating coral reefs around the world
46 (1) but there is room for hope if corals can adapt to increasing temperatures (2). Many coral
47 species have wide distributions that span environments that differ dramatically in their thermal
48 regimes, demonstrating that efficient thermal adaptation has occurred in the past (3). But can
49 coral adaptation keep up with the unprecedentedly rapid current rate of global warming (4)? One
50 way for corals to achieve rapid thermal adaptation is through genetic rescue, involving the spread
51 of existing heat tolerance alleles from low-latitude, warm-adapted populations to higher-latitude,
52 warming regions, via larval migration (5, 6). We have previously demonstrated the presence of
53 genetic variants conferring high thermal tolerance in a low-latitude *A. millepora* population (5). It
54 can be expected that global warming will cause preferential survival of warm-adapted poleward
55 migrants because they will be following their thermal optimum, whereas individuals migrating in
56 the opposite direction would find themselves in increasingly mismatched environments (Fig. 1 A,
57 B). Another likely population-level effect of recent declines in coral cover (7) is a reduction in
58 overall genetic diversity, potentially limiting both the scope and the rate of adaptation.

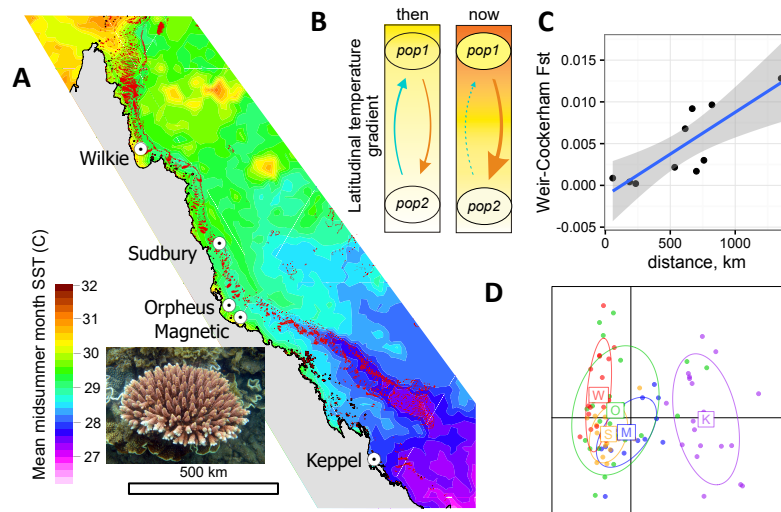
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60 Here, we test these predictions in *Acropora millepora*, a common reef-building coral from the
61 most ecologically prominent and diverse coral genus in the Indo-Pacific (staghorn corals,
62 *Acropora*). We have analyzed genome-wide genetic variation using 2bRAD (8) in five
63 populations of *A. millepora* along the latitudinal range of the GBR (Fig. 1 A). We genotyped 18-
64 28 individuals per population at >98% accuracy and with a >95% genotyping rate. Analysis of
65 population structure based on ~11,500 biallelic SNPs separated by at least by 2,500 bases agreed
66 with previous microsatellites results (9, 10), and revealed very low levels of genetic divergence,
67 with only the Keppel Islands population being potentially different from the others (Fig. 1 D and
68 Fig. S1). We observed increasing genetic divergence with geographical distance (“isolation by
69 distance”, Fig. 1 C) that supports population divergence, however, pairwise F_{ST} were small and
70 did not exceed 0.014 even between the southernmost and northernmost populations (Keppel and
71 Wilkie). To gain a deeper insight into coral demography, we used Diffusion Approximation for
72 Demographic Inference (*dadi*, (11)) to more rigorously test for population subdivision and infer
73 pairwise migration rates among populations and population sizes. *dadi* is a coalescent-based
74 method that optimizes parameters of a pre-specified demographic model to maximize the
75 likelihood of generating the observed allele frequency spectrum (for two populations it is
76 essentially a two-dimensional histogram of allele frequencies, Fig. S2). Being a likelihood-based
77 method, *dadi* can be used to compare alternative models using likelihood ratio tests and Akaike
78 Information Criterion (AIC).

79

80 We used AIC to confirm that our populations are separate demographic units. For each pair of
81 populations we generated 120 bootstrapped datasets by resampling genomic contigs and
82 performed delta-AIC comparison of two demographic models, a split-with-migration model and a
83 no-split model (Fig. S3 B). The split-with-migration model assumed two populations that have

84 split some time T in the past, potentially have different sizes $N1$ and $N2$, and exchange migrants
85 at different rates ($m12$ and $m21$) depending on direction. The no-split model allowed for ancestral
86 population size to change at time T but not for a population split, so the experimental data were
87 modeled as two random samples from the same population of size N . The majority of bootstrap
88 replicates (88-100%) showed AIC advantage of the split-with-migration model for all but one
89 pair of populations (Sudbury-Magnetic, 39% bootstrap support; Fig. S3). This indicates that the
90 populations are demographically distinct despite very low F_{ST} . This result highlights the power of
91 coalescent analysis relative to classical approaches (such as F_{ST}) that assume genetic equilibrium,
92 i.e., that populations have been stable for thousands of generations.
93



94
95 **Figure 1.** The population setting and background for our study. (A) Locations of sampled populations
96 where mean midsummer month sea surface temperature differed by up to $\sim 3^{\circ}\text{C}$. Inset: *Acropora millepora*.
97 (B) Working hypothesis under global warming: Warm-adapted low-latitude genotypes that migrate to
98 higher latitudes would be following their physiological optimum and hence expected to survive better than
99 migrants in the opposite direction. (C) Increase of pairwise F_{ST} with distance, both indicating weak genetic
100 divergence along the GBR, and (D) principal component analysis of genome-wide genetic variation. On
101 panel D, centroid labels are initial letters of population names as in panel A.
102

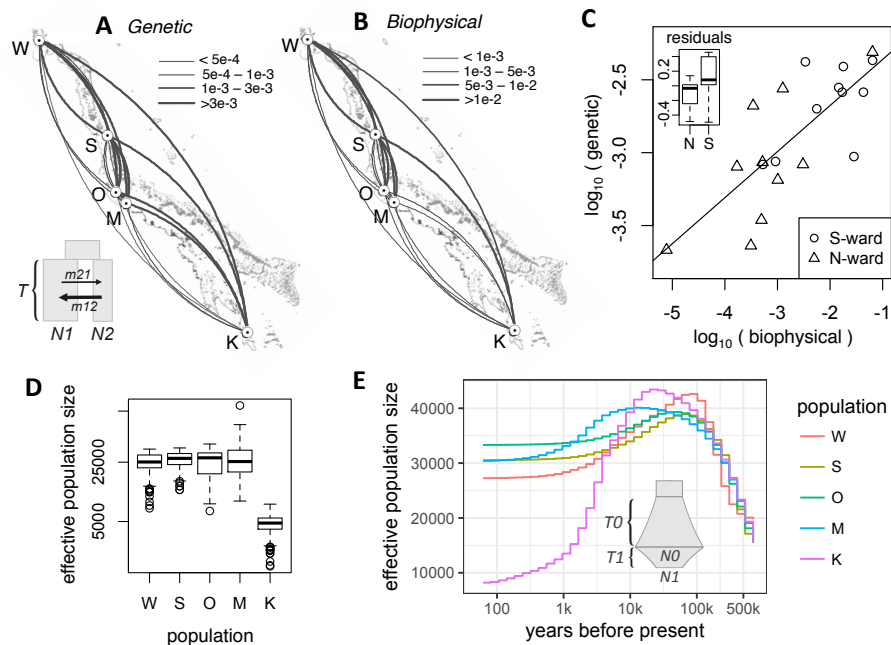
103 We then determined pairwise migration rates from the split-with-migration model and estimated
104 their confidence limits from bootstrap replicates. For all pairwise analyses except Wilkie-Sudbury
105 migration in southward direction exceeded northward migration, and this difference was
106 significant in seven out of nine cases (Fig. 2 A and Fig. S3A). Linear mixed model analysis of
107 direction dependent mean migration rates with a random effect of destination (to account for
108 variation in total migration rate) confirmed the overall significance of this southward trend
109 ($P_{\text{MCMC}} < 1e-4$).
110

111 It is important to note that our pairwise migration rates captured the cumulative effect of genetic
112 exchange between populations, which included direct migration and the spread of alleles via other

113 stepping-stone populations. Such rates do not directly reflect the numbers of larvae exchanged
 114 between populations but are very informative in the genetic rescue context. They represent the
 115 per-generation rate of replacement of the destination population genotypes by genotypes from the
 116 source population, which is essentially the rate at which genetic rescue could proceed.

117

118 To investigate whether the southward migration bias was due to higher survival of southward
 119 migrants relative to northward migrants, as predicted under global warming (Fig. 1 B), we
 120 developed a biophysical model of coral larval dispersal on the Great Barrier Reef. This model
 121 quantified the per-generation migration potential among coral reef habitat patches in the GBR
 122 based on ocean currents and parameters of larval biology (12, 13). We found that the genetic and
 123 biophysical migration rates were very closely correlated (Mantel test: $r = 0.79$, $P = 0.008$, Fig. 2
 124 C).



125

126

127 **Figure 2.** Demography of *A. millepora* populations on the GBR. (A) Arc-plot of migration rates among
 128 populations reconstructed from population genetic data. Inset: *dadi* model used: ancestral population splits
 129 into two populations of unequal sizes (N1 and N2) some time T in the past, these populations exchange
 130 migrants at different rates depending on direction. (B) Migration rates according to the biophysical model.
 131 On panels A and B, the arcs should be read clockwise to tell the direction of migration; line thickness is
 132 proportional to the migration rate. (C) Correlation between log-transformed biophysical and genetic
 133 migration rates (Mantel $r = 0.79$, $P = 0.008$). Inset: box-plot of residuals from the linear regression.

134 Southward migration tends to exceed northward migration even after accounting for predictions of the
 135 biophysical model ($P = 0.058$), suggesting higher survival of southward migrants. (D) Box plot of effective
 136 population sizes inferred by the split-with-migration model (panel A) across all population pairs and
 137 bootstrap replicates. (E) Historical changes in effective population sizes inferred using a single-population
 138 *dadi* model with two periods of exponential growth (T_0 and T_1 , reaching sizes N_0 and N_1 , inset), averaged
 139 across bootstrap replicates.

140 Although the biophysical model explained most of the southward migration bias in the genetic
141 data, the residuals were still in favor of southward migration (Fig. 2 C, inset; $P = 0.058$).

142 While this residual excess suggest preferential survival of southward migrants, as predicted by
143 our hypothesis (Fig. 1 B). These genetic predictions represent historical averages since the
144 populations split and did not resolve any potential recent migration changes.

145

146 To determine any recent changes in southward migration, we evaluated a similar basic split-with-
147 migration model (Fig. 2A) that allowed for a change in migration over the past 75-100 years. The
148 new model suggested some recent migration changes, but there was no consistent change between
149 northward and southward migration (Fig. S4). Delta-AIC bootstrap analysis favored the new
150 model over the basic one only for two pairs of populations, Wilkie-Orpheus and Wilkie-Magnetic
151 (85 and 60% bootstrap support, respectively). We conclude that with the current data and analysis
152 techniques we cannot yet detect the effect of recent warming on preferential direction of coral
153 migration along the GBR.

154

155 The GBR has already warmed by 0.8°C since the end of last century (14) and may have already
156 reduced genetic diversity in *A. millepora* populations. We used *dadi* to infer effective population
157 sizes, which is a measure of genetic diversity and one of the key parameters determining the
158 population's adaptive potential (15). The results of the split-with-migration model (Fig. 2 A)
159 were consistent for all population pairs and indicated that Keppel population was about one-fifth
160 the size of others (Fig. 2 D, E). This result was not surprising since the Keppel population
161 frequently suffers high mortality due to environmental disturbances and was therefore is expected
162 to show diminished long-term effective population size (9). We also used a single-population
163 *dadi* model that allowed for two consecutive growth/decline periods (Fig. 2 E, inset) to
164 reconstruct effective sizes of individual populations through time (Fig. 2 E and Fig S5). All
165 populations showed evidence of growth prior to the last glaciation, 500-20 thousand years ago
166 (Fig 2 E), which aligned well with the fossil record of rising dominance of *Acropora* corals on
167 Indo-Pacific reefs during this period (16). It has been suggested that the fast growth and early
168 sexual maturation of *Acropora* corals gave them an advantage relative to most other reef-building
169 corals during dynamic changes in the reef-forming zone due to the sea level changes
170 accompanying glacial cycles (16). Our results suggest that *A. millepora* populations have been in
171 stasis or slow decline since sea level changes abated (Fig. S5), although the inclusion of an
172 additional growth/decline period only improved the model fit significantly for the Keppel
173 population (Fig S6). None of the populations showed evidence of accelerated decline in effective
174 population size over the past few hundred years. Although our samples were collected in the
175 early-mid 2000s, our results are still relevant since they characterize populations only two-three
176 coral generations ago. Disturbances that have affected corals since then would not yet have
177 substantially impacted genetic diversity.

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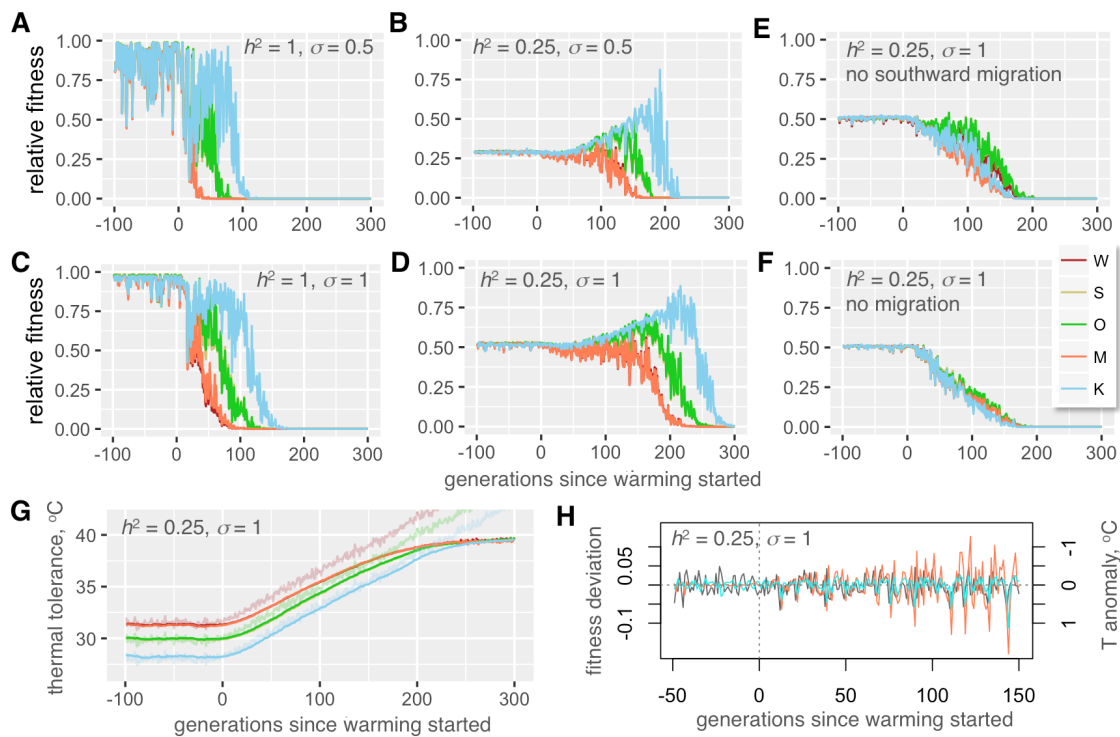
179 To evaluate whether standing genetic variation contributed by local thermal adaptation could
180 sustain evolution of the *A. millepora* metapopulation in response to warming, we have developed
181 a multi-QTL model of metapopulation adaptation in SLiM (17). The model was parameterized

182 with population sizes and migration rates inferred from the genetic analysis (Fig. 2 A, D), and
183 with differences in midsummer monthly mean temperature among populations (Fig. 1 A). The
184 number of QTLs and their effect sizes, phenotypic plasticity (standard deviation of the Gaussian
185 slope of fitness decline when phenotype mismatches the environment) and heritability (proportion
186 of phenotypic variation attributable to genetics) can all be varied in the model. It can also
187 incorporate climate scenarios with any combination of directional, cyclical and random changes.
188 The model also allows for new mutations but here the new mutation rate was set to zero. This was
189 to assess the contribution of only the standing genetic variation that was introduced into
190 populations at the start of simulation as random QTL effects. The climate scenario started with a
191 pre-adaptation to local thermal conditions for 2,000 generations. Assuming a generation time of
192 of 5 years in *A. millepora* (18) this corresponded to the period of stable temperature since the last
193 deglaciation. After pre-adaptation, the temperature was increased at a rate of 0.05°C/generation in
194 all populations, corresponding to the projected 0.1°C warming per decade (19). Throughout the
195 simulation temperature was allowed to fluctuate randomly between generations to approximate El
196 Nino Southern Oscillation (ENSO): the temperature deviations were drawn from a normal
197 distribution with a standard deviation of 0.25°C. The size of populations was kept constant
198 throughout the pre-adaptation period and scaled linearly with the populations' relative fitness
199 (mean current fitness divided by the mean fitness at the end of pre-adaptation period) during
200 warming. Migration rates from a population also scaled linearly with the population's fitness. In
201 this way, a population declining in fitness would shrink in size and stop contributing migrants to
202 other populations.

203
204 Our model suggested that, with only ten thermal QTLs, under all combinations of heritability and
205 plasticity the pre-adapted metapopulation would be able to persist through the warming for at
206 least 50-100 generations and, in some realistic cases, much longer (Fig. 3 and Figs. S7-S8).
207 Migration in general and southward migration in particular substantially contributed to this
208 persistence (Fig. 3 E, F), underscoring the importance of the spread of warm-adapted genotypes
209 from lower to higher latitudes (5).

210
211 Predictably, higher phenotypic plasticity promoted population persistence and stability against
212 random thermal anomalies, but we were rather surprised to observe a similar positive effect of
213 lower heritability, set to the values observed in coral quantitative genetics experiments (0.25-0.5,
214 (17); Fig 3, Fig. S7). One specific reason why corals are expected to show low heritability of
215 thermal tolerance is that much of natural variation in this trait in corals is due to the type of algal
216 symbionts (*Symbiodinium* spp. (20)). Photo-symbionts are not transmitted from parent to
217 offspring in the majority of coral species (21), and although host genetics can have some effect on
218 the choice of *Symbiodinium* in the next generation (22) environment has a very strong effect on
219 this association (20, 23). Higher persistence under low heritability and high plasticity is most
220 likely explained by the fact that they both allow for higher standing genetic variation to be
221 retained in populations (Fig. S9). During warming, this variation lasts longer as a source of
222 adaptive genetic variants, enabling up to 5°C increase in mean thermal tolerance over 150
223 generation (Fig. 3 G and Fig. S7). Higher plasticity partially rescued the drop in fitness due to low

224 heritability (Fig. 3 B and D, Fig. S7). Another notable tendency observed with all parameter
 225 settings was that during warming the fitness (and hence the size) of adapting populations began to
 226 fluctuate following random thermal anomalies, and the amplitude of these fitness fluctuations
 227 increased as the warming progressed even though the amplitude of thermal anomalies did not
 228 change (Fig. 3 H). These fluctuations correspond to severe mortality events induced by thermal
 229 extremes due to ENSO and affected warm-adapted populations most, which very much resembles
 230 the situation currently observed throughout the world (1).
 231



232
 233 **Figure 3.** Modeling coral metapopulation persistence under global warming. (A-D) Fitness of modeled
 234 populations depending on heritability of thermal tolerance (h^2 , proportion of tolerance variation
 235 explained by genetics), phenotypic plasticity (σ , standard deviation of the Gaussian slope of fitness decline away from
 236 the phenotypic optimum, in degrees C), and presence-absence of migration (E, F). On panels A-F, y-axis is
 237 observed fitness relative to maximal fitness at the genetically determined optimum, averaged over all
 238 individuals in a population. Warm-adapted populations (W and M) are shown as red-tint traces, populations
 239 from mild thermal regime (S and O) are green-tint traces, and the cool-adapted population (K) is the blue
 240 trace. Note nearly complete overlap between traces for pairs of populations pre-adapted to the same
 241 temperature (W,M and S,O). (G) Thermal tolerances of evolving populations. Thin noisy lines are modeled
 242 temperatures at different locations. (H) Modeled random temperature anomalies (grey line) and fluctuations
 243 in populations' fitness (the colored lines are residuals from loess regression over fitness traces on panel D;
 244 Wilkie: orange line, Keppel: blue line). Note the inverse sign of temperature anomalies: this more clearly
 245 shows the correspondence between rise in temperature and drop in fitness in the next generation. As
 246 warming progresses, populations (especially originally warm-adapted ones) become increasingly sensitive
 247 to random temperature fluctuations.
 248

249 There are several uncertainties in our model associated with coral biology. Higher number of
250 QTLs and/or their larger effect sizes would promote higher genetic variation and lead to longer
251 population persistence. To keep the analysis conservative, our model included only ten QTLs,
252 which is likely much fewer than the actual number of thermal QTLs in acroporid corals (20). We
253 also kept the distribution of QTL effect sizes narrow: with the current settings and ten QTLs, at
254 the start of simulation only about 2% of corals deviated from the mean thermal tolerance by more
255 than 1.5°C in either direction. Such narrow variation makes adaptation to the thermal gradient of
256 ~3°C along the GBR non-trivial, but still, at present there is no experimental data to evaluate
257 whether even such narrow variation is realistic. Our model was also conservative in using
258 effective population sizes suggested by genetic analysis as census sizes. In highly fecund marine
259 organisms census sizes tend to substantially exceed effective population sizes, sometimes by
260 orders of magnitude (24), which would strongly promote higher genetic diversity and population
261 persistence. Moreover, we modeled only our five populations rather than the whole GBR, which
262 would have resulted in much higher standing genetic variation in the metapopulation, promoting
263 longer persistence.

264
265 As for phenotypic plasticity, in simulations shown on Fig. 3, $\sigma = 0.5$ and $\sigma = 1$ corresponded to
266 86% and 40% decline in fitness if the individual's phenotype mismatched the environment by
267 1°C. The existing data on the issue of coral thermal plasticity are somewhat conflicting. One
268 study shows that acroporid corals can successfully acclimatize to environments differing in
269 maximum temperatures by as much as 2°C (25); however, another study found that coral grew
270 52-80% more slowly when transplanted among locations differing by 1.5°C average temperature,
271 (26). Although it is not possible to directly place these results into our quantitative plasticity
272 framework, the former study supports the higher plasticity setting ($\sigma = 1$) while the latter study
273 supports $\sigma = 0.5$. It must also be noted that both these studies involved *in situ* transplantations and
274 hence the effect of temperature remains confounded with other local fitness-affecting
275 environmental parameters. Also, in adult corals plasticity is likely lower than in larvae and
276 recruits, which are expected to exhibit non-reversible developmental plasticity associated with
277 metamorphosis and establishment within a novel environment (27). Future experiments that
278 expose multiple genetically distinct coral individuals to a range of temperatures under controlled
279 laboratory settings are required to rigorously quantify variation in thermal optima and plasticity in
280 natural populations.

281
282 In conclusion, we found that genetic diversity and migration patterns of our study species were
283 not yet affected by global warming and were well positioned to facilitate persistence of the GBR
284 metapopulation for a century or more. However, despite ongoing adaptation to gradual
285 temperature increase, corals will become increasingly more sensitive to local thermal anomalies,
286 especially among the originally warm-adapted populations. The 10-85% mortality in the Northern
287 GBR as a result of 2016 bleaching event (28) could be a particularly sobering recent
288 manifestation of this trend. Our model assumed that recovery from such mortality events would
289 depend solely on the demographic exchange between coral populations. However, ecological
290 feedbacks such as shifts to an alternative ecological stable state (29) might substantially decrease

291 the rate of reseeded and recovery of affected reefs. In that case, the increase in severity of
292 bleaching-related mortality might lead to much faster coral extinction than predicted by our
293 model.

294
295 More research into phenotypic plasticity and genetic variation in coral thermal tolerance and its
296 genetic architecture (number of QTLs and their effect sizes) is needed to further improve the
297 predictive power of our model. The estimated migration in the order of 10 - 100 migrants per
298 generation could be feasibly facilitated by assisted gene flow efforts (30) without risking
299 disruption of the natural local adaptation patterns (31). Corals are declining on reef world-wide
300 and there is an urgent need to develop new solutions to effectively manage the impacts of global
301 processes such as climate change at local management scales. The broad characterization of
302 genetic diversity, local thermal adaptation and migration pathways in multiple reef-building coral
303 species would greatly inform both traditional spatial management and novel assisted gene flow
304 approaches and should therefore be given high priority.

305

306 **Methods**

307

308 *Genotyping*

309

310 This study relied predominantly on samples described by van Oppen et al (10), with addition of
311 several samples from Orpheus and Keppel islands that were used in the reciprocal transplantation
312 experiment described by Dixon et al (32). The samples were genotyped using 2bRAD (8)
313 modified for Illumina sequencing platform; the latest laboratory and bioinformatics protocols are
314 available at https://github.com/zoon/2bRAD_GATK. BcgI restriction enzyme was used and the
315 samples retained for this analysis had 2.3-20.2 (median: 7.45) million reads after trimming and
316 quality filtering (no duplicate removal was yet implemented in this 2bRAD version). The reads
317 were mapped to the genome of the outgroup species, *Acropora digitifera* (33, 34), to polarize the
318 allelic states into ancestral (as in *A. digitifera*) and derived, e.g., (35, 36). Genotypes were called
319 using GATK pipeline (37).

320

321 Preliminary analysis of sample relatedness using vcftools (38) revealed that our samples included
322 several clones: four repeats of the same genotype from the Keppel Island (van Oppen et al (10)
323 samples K210, K212, K213 and K216), another duplicated genotype from Keppel (samples K211
324 and K219), and one duplicated genotype from Magnetic Island (samples M16 and M17). All
325 other samples were unrelated. We took advantage of these clonal replicates to extract SNPs that
326 were genotyped with 100% reproducibility across replicates and, in addition, appeared as
327 heterozygotes in at least two replicate pairs (script replicatesMatch.pl with hetPairs=2 option).
328 These 7,904 SNPs were used as “true” SNP dataset to train the error model to recalibrate variant
329 quality scores at the last stage of the GATK pipeline. During recalibration, we used the transition-
330 transversion (Ts/Tv) ratio of 1.438 determined from the “true” SNPs to assess the number of false
331 positives at each filtering threshold (as it is expected that an increase of false positive calls would
332 decrease the Ts/Tv ratio towards unity). We chose the 95% tranche, with novel Ts/Tv = 1.451.

333 After quality filtering that restricted the calls to only bi-allelic polymorphic sites, retained only
334 loci genotyped in 95% or more of all individuals, and removed loci with the fraction of
335 heterozygotes exceeding 0.6 (possible lumped paralogs), we ended up with 25,090 SNPs. In total,
336 2bRAD tags interrogated 0.18% of the genome. The genotyping accuracy was assessed based on
337 the match between genotyped replicates using script repMatchStats.pl. Overall agreement
338 between replicates was 98.7% or better with the heterozygote discovery rate (fraction of matching
339 heterozygote calls among replicates) exceeding 96%.

340

341 *Genome-wide genetic divergence*

342

343 To begin to characterize genome-wide divergence between populations we used pairwise
344 genome-wide Weir and Cockerham's F_{ST} calculated by vcftools (38), principal component
345 analysis (PCA) using R package adegenet (39), and ADMIXTURE (40). For PCA and
346 ADMIXTURE, the data were thinned to keep SNPs separated by 5kb on average and by at least
347 2.5 kb, choosing SNPs with highest minor allele frequency (script thinner.pl with options
348 'interval=5000 criterion=maxAF').

349

350 *Demographic analysis and bootstrapping*

351

352 Prior to demographic analysis, Bayescan (41) was used to identify sites potentially under
353 divergent selection among populations, and 13 such sites with q-value <0.05 were removed.
354 Demographic models were fitted to 120 bootstrapped datasets, which were generated in two
355 stages. First, five alternatively thinned datasets were generated for which SNPs were randomly
356 drawn to be on average 5 kb apart and not closer than 2.5 kb. This time the SNPs were drawn at
357 random to avoid distorting the allele frequency spectrum unlike thinning for PCA and
358 ADMIXTURE where the highest minor allele frequency SNPs were selected. Then, 20
359 bootstrapped replicates were generated for each thinned dataset by resampling contigs of the
360 reference genome with replacement (script dadiBoot.pl). The fitted model parameters were
361 summarized after excluding bootstrap replicates that fell into the lowest 15% likelihood quantile
362 and the ones where model fitting failed to converge, leading to some parameters being
363 undetermined or at infinity (less than 10% of total number of runs). Delta-AIC values were
364 calculated for each bootstrap replicate that passed these criteria for both compared models, and
365 summarized to obtain bootstrap support value, the percentage of replicates favoring the
366 alternative model. While fitting *dadi* models, the data for each population were projected to
367 sample sizes maximizing the number of segregating sites in the analysis, resulting in 7000-8172
368 segregating sites per population.

369

370 *Unit conversion*

371

372 To convert *dadi*-reported coalescent parameter values (θ , T and M) into time in years (t), effective
373 population sizes in number of individuals (N_e) and migration rates as fraction of new immigrants
374 per generation (m), we estimated the mutation rate (μ) from the time-resolved phylogeny of

375 *Acorpora* genus based on *paxC* intron (42), at $4e-9$ per base per year. Although *A. millepora* was
376 shown to start reproducing in 3 years (18) we assumed the generation time of 5 years reasoning
377 that it would better reflect the attainment of full reproductive potential as the colony grows.
378 Assuming a genome size of $5e+8$ bases (33) the number of new mutations per genome per
379 generation is 10. Since the fraction 2bRAD-sequenced genome in our experiment was $1.8e-3$, the
380 mutation rate per 2bRAD-sequenced genome fraction per generation is $\mu = 0.018$. This value was
381 used to obtain:

- 382 - Ancestral effective population size: $Ne = \theta / 2\mu$
- 383 - Migration rate: $m = M / 2Ne$
- 384 - Time in years: $t = 2TNe \cdot 5$

385

386 *Biophysical model*

387

388 A spatially-explicit biophysical modeling framework (12, 43) was used to quantify migration
389 between coral reef habitats of the broader region surrounding the Great Barrier Reef, thereby
390 revealing the location, strength, and structure of a species' potential population connectivity. The
391 model's spatial resolution of ca. 8 km coincides with hydrodynamic data for the broader region
392 ($1/12.5$ degree; HYCOM+NCODA Reanalysis and Analysis product; [hycom.org](http://data.unep-wcmc.org/datasets/1)). Our
393 biophysical dispersal model relies on geographic data describing the seascape environment and
394 biological parameters capturing coral-specific life-histories. Coral reef habitat data are available
395 from the UNEP World Conservation Monitoring Centre (UNEP-WCMC; [http://data.unep-](http://data.unep-wcmc.org/datasets/1)
396 [wcmc.org/datasets/1](http://data.unep-wcmc.org/datasets/1)) representing a globally-consistent and up-to-date representation of coral
397 reef habitat. To capture specific inter-annual variability, two decades of hydrodynamic data were
398 used from 1992 to 2013 (44).

399

400 Coral-specific biological parameters for *A. millipora* included relative adult density (dependent
401 on the habitat), reproductive output, larval spawning time and periodicity (e.g., Magnetic Island
402 populations spawn a month earlier than the other GBR sites (45)), maximum dispersal duration,
403 pre-competency and competency periods, and larval mortality (46, 47). The spatially explicit
404 dispersal simulations model the dispersal kernel (2-D surface) as a 'cloud' of larvae, allowing it
405 to be concentrated and/or dispersed as defined by the bio-physical parameters. An advection
406 transport algorithm is used for moving larvae within the flow fields (48).

407

408 Simulations were carried out by releasing a cloud of larvae into the model seascape at all
409 individual coral reef habitat patches and allowing the larvae to be transported downstream by the
410 currents. Ocean current velocities, turbulent diffusion, and larval behavior move the larvae
411 through the seascape at each time-step. Larval competency, behavior, density, and mortality
412 determine when and what proportion of larvae settle in habitat cells at each time step. When
413 larvae encounter habitat, the concentration of larvae settling with the habitat is recorded at that
414 time-step. From the dispersal data, we derived the coral migration matrix representing the
415 proportion of settlers to each destination patch that came from a source patch, which is analogous
416 to the source distribution matrix (49) and is equivalent to migration matrices derived from

417 population genetic analysis. It is important to note that migration matrices extracted for the field
418 sites represent the potential migration through all possible stepping-stones.

419

420 *Metapopulation adaptation model*

421

422 The model was implemented in SLiM the forward evolutionary simulator, by modifying the
423 provided recipe “Quantitative genetics and phenotypically-based fitness”. The model simulates
424 Fisher-Wright populations with discreet generations. At the start of the simulation, populations
425 were established at specified population sizes and pairwise migration rates (genetic replacement
426 rates), and all QTLs in all individuals were given a mutation with the effect size drawn from a
427 normal distribution with mean zero and specified standard deviation, to create standing genetic
428 variation. The phenotype of each individual was calculated as the sum of QTL effects plus
429 random noise to simulate desired heritability. Then, fitness of each individual was calculated
430 based on the difference between the individual’s phenotype (thermal optimum), temperature of
431 the environment, and the setting for phenotypic plasticity, modeled as the standard deviation of
432 the Gaussian slope of fitness decline with increasing distance between phenotype and
433 environment. Then, parents were chosen to produce the next generation according to their fitness;
434 parents for immigrant individuals are chosen from among individuals in the source population.
435 New mutations at QTLs happened at the specified rate when transitioning to the next generation
436 and the effect of a new mutation replaced the previous QTL effect.

437

438 Our code was designed for general modeling of multilocus adaptation in metapopulations and can
439 process matrices of population sizes and migration rates for an arbitrary number of populations.
440 We modeled our five populations with effective population sizes and pairwise migration rates
441 inferred by *dadi*. Within the code, it is also possible to adjust:

442

- 443 - Number of QTLs and the distribution of their effect sizes. To keep the model conservative,
444 we modeled only ten QTLs with normal distribution of effect sizes with a standard deviation
445 of 0.2°C. With ten QTLs, this setting implied that at the start of simulation only about 2% of
446 corals deviated from mean thermal tolerance by more than 1.5°C in either direction. Since
447 thermal differences between our populations exceeded 3°C, this narrow variation made local
448 adaptation rather non-trivial.
- 449 - Dominance of QTLs (set to 0.5 in our simulation).
- 450 - Phenotypic plasticity. We modeled three plasticity settings, 0.5, 1 and 2, which corresponded
451 to 86%, 40% and 13% fitness drop when the individual’s phenotypic optimum (calculated
452 based on QTLs and heritability setting) mismatched the environment by 1°C.
- 453 - Heritability (proportion of phenotypic variation explained by genetics). We examined values
454 1, 0.5, 0.25 and 1e-5, the latter to confirm that no adaptation or evolution was observed when
455 the trait was not heritable.
- 456 - Mutation rate, which was set to zero because we wanted to explore only the role of standing
457 genetic variation.

458

459 To better model population dynamics during warming period, we implemented linear scaling of
460 the population size and immigration rates with the population's mean fitness. In this way, a
461 population declining in fitness shrinks in size and stops contributing migrants to other
462 populations.

463
464 Environmental conditions are supplied to our model as a table of values for each population in
465 every generation and can be arbitrary. Here we modeled identical thermal trends across
466 populations with population-specific offsets. During pre-adaptation period lasting 2000
467 generations, the temperature was constant on average but experienced random fluctuations
468 across generations drawn from a normal distribution with a standard deviation of 0.25°C (to
469 approximate ENSO events). The temperature was offset by +1.6°C in Wilkie and Magnetic
470 populations and by -1.8°C in the Keppel population, to model differences in midsummer
471 monthly mean temperature among populations (Fig. 1). After 2000 generations a linear increase
472 at 0.05°C per generation was added to simulate warming.

473
474 All combinations of parameter settings were run ten times to ensure consistency. We found that
475 with population sizes in thousands, such as in our case, the results were very consistent among
476 independent runs. We therefore did not aggregate results over many replicated runs but show one
477 randomly chosen run for each tested parameter combination.

478 479 **Acknowledgements**

480
481 We wish to thank Ryan Gutenkunst and Benjamin Haller for their continuous support of *dadi* and
482 SLiM users, respectively. The bioinformatics analysis was accomplished using computational
483 resources of the Texas Advanced Computer Center. This study has been supported by NSF
484 (DEB-1054766) grant to M.V.M, ARC (LP120200245) and University of Melbourne ECR grants
485 to E. A.T., a Coral Reef Alliance grant ("Coral Adaptation Challenge") to E.A.T and M.V.M,
486 Queensland Government funding to L.K.B and AIMS funding to L.K.B. and M.J.V.O

487 **Data and code availability**

488 The finalized genotyping dataset in VCF format, detailed bioinformatic walkthrough, accessory
489 formatting and plotting scripts, *dadi* scripts and the SLiM model code are available from
490 <https://github.com/z0on/Adaptive-pathways-of-coral-populations-on-the-Great-Barrier-Reef>. Raw
491 sequencing data has been deposited to National Center for Biotechnology Information's Short
492 Reads Archive (accession number pending).

493

494 **References**

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- 496 1. A. C. Baker, P. W. Glynn, B. Riegl, Climate change and coral reef bleaching: An ecological assessment
497 of long-term impacts, recovery trends and future outlook. *Estuar. Coast. Shelf Sci.* **80**, 435–471 (2008).
- 498 2. C. A. Logan, J. P. Dunne, C. M. Eakin, S. D. Donner, Incorporating adaptive responses into future
499 projections of coral bleaching. *Glob. Chang. Biol.* **20**, 125–139 (2014).
- 500 3. T. P. Hughes *et al.*, Climate change, human impacts, and the resilience of coral reefs. *Science*. **301**,
501 929–933 (2003).
- 502 4. J. M. Pandolfi, S. R. Connolly, D. J. Marshall, A. L. Cohen, Projecting coral reef futures under global
503 warming and ocean acidification. *Science*. **333**, 418–22 (2011).
- 504 5. G. B. B. Dixon *et al.*, Genomic determinants of coral heat tolerance across latitudes. *Science*. **348**,
505 1460–1462 (2015).
- 506 6. P. K. Ingvarsson, Restoration of genetic variation lost – the genetic rescue hypothesis. *Trends Ecol.*
507 *Evol.* **16**, 62–63 (2001).
- 508 7. G. De’ath, K. E. Fabricius, H. Sweatman, M. Puotinen, The 27-year decline of coral cover on the Great
509 Barrier Reef and its causes. *Proc. Natl. Acad. Sci. U. S. A.* **109**, 17995–9 (2012).
- 510 8. S. Wang, E. Meyer, J. K. McKay, M. V Matz, 2b-RAD: a simple and flexible method for genome-wide
511 genotyping. *Nat. Methods*. **9**, 808–810 (2012).
- 512 9. M. J. H. van Oppen, V. Lukoschek, R. Berkelmans, L. M. Peplow, A. M. Jones, A population genetic
513 assessment of coral recovery on highly disturbed reefs of the Keppel Island archipelago in the southern
514 Great Barrier Reef. *PeerJ*. **3**, e1092 (2015).
- 515 10. M. J. H. Van Oppen, L. M. Peplow, S. Kininmonth, R. Berkelmans, Historical and contemporary factors
516 shape the population genetic structure of the broadcast spawning coral, *Acropora millepora*, on the
517 Great Barrier Reef. *Mol. Ecol.* **20**, 4899–4914 (2011).
- 518 11. R. N. Gutenkunst, R. D. Hernandez, S. H. Williamson, C. D. Bustamante, Inferring the joint
519 demographic history of multiple populations from multidimensional SNP frequency data. *PLoS Genet.*
520 **5**, e1000695 (2009).
- 521 12. E. A. Trembl, J. Roberts, P. N. Halpin, H. P. Possingham, C. Riginos, The emergent geography of
522 biophysical dispersal barriers across the Indo-West Pacific. *Divers. Distrib.* **21**, 465–476 (2015).
- 523 13. E. A. Trembl *et al.*, Reproductive Output and Duration of the Pelagic Larval Stage Determine Seascape-
524 Wide Connectivity of Marine Populations. *Integr. Comp. Biol.* **52**, 525–537 (2012).
- 525 14. J. Lough, A. Hobday, Observed climate change in Australian marine and freshwater environments. *Mar.*
526 *Freshw. Res.* **62**, 984–999 (2011).
- 527 15. B. Charlesworth, Fundamental concepts in genetics: Effective population size and patterns of molecular
528 evolution and variation. *Nat. Rev. Genet.* **10**, 195–205 (2009).
- 529 16. W. Renema *et al.*, Are coral reefs victims of their own past success? *Sci. Adv.* **2**, e1500850 (2016).
- 530 17. B. C. Haller, P. W. Messer, SLiM 2: Flexible, interactive forward genetic simulations. *Mol. Biol. Evol.*
531 **34**, 230–240 (2017).
- 532 18. M. Vanessa, B. Baria, D. W. Dela Cruz, R. D. Villanueva, J. R. Guest, Spawning of three-year-old
533 *Acropora millepora* corals reared from larvae in Northern Philippines. *Bull. Mar. Sci.* **88**, 61–62 (2012).
- 534 19. IPCC, *Climate Change 2007 - The Physical Science Basis. Contribution of Working Group I to the*
535 *Fourth Assessment Report of the IPCC* (Cambridge University Press, New York, NY, 2007).
- 536 20. R. Berkelmans, M. J. H. van Oppen, The role of zooxanthellae in the thermal tolerance of corals: a
537 “nugget of hope” for coral reefs in an era of climate change. *Proc. R. Soc. B-Biological Sci.* **273**, 2305–
538 2312 (2006).
- 539 21. A. H. Baird, J. R. Guest, B. L. Willis, Systematic and biogeographical patterns in the reproductive
540 biology of scleractinian corals. *Annu. Rev. Ecol. Evol. Syst.* **40**, 551–571 (2009).
- 541 22. K. Quigley, B. Willis, L. Bay, Heritability of the Symbiodinium community in vertically- and
542 horizontally-transmitting broadcast spawning corals. *bioRxiv*. (2017).
- 543 23. E. J. Howells *et al.*, Coral thermal tolerance shaped by local adaptation of photosymbionts. *Nat. Clim.*
544 *Chang.* **2** (2011), pp. 116–120.
- 545 24. F. P. Palstra, D. J. Fraser, Effective/census population size ratio estimation: a compendium and
546 appraisal. *Ecol. Evol.* **2**, 2357–2365 (2012).

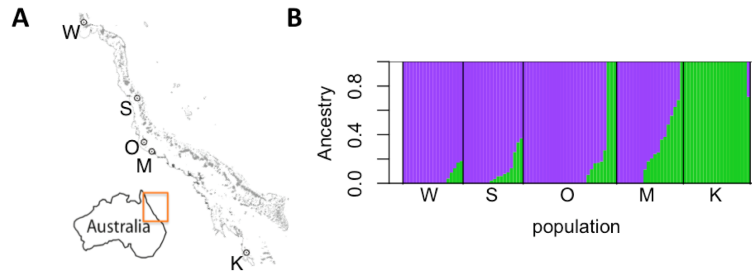
- 547 25. S. R. Palumbi, D. J. Barshis, N. Traylor-Knowles, R. A. Bay, Mechanisms of reef coral resistance to
548 future climate change. *Science*. **344**, 895–8 (2014).
- 549 26. E. J. Howells, R. Berkelmans, M. J. H. van Oppen, B. L. Willis, L. K. Bay, Historical thermal regimes
550 define limits to coral acclimatization. *Ecology*. **94**, 1078–1088 (2013).
- 551 27. D. Nettle, M. Bateson, Adaptive developmental plasticity: what is it, how can we recognize it and when
552 can it evolve? *Proc. R. Soc. B Biol. Sci.* **282**, 20151005 (2015).
- 553 28. Great Barrier Reef Marine Park Authority, *Interim report on the environmental impacts of the 2016*
554 *coral bleaching event*.
- 555 29. T. J. Done, Phase shifts in coral reef communities and their ecological significance. *Hydrobiologia*. **247**,
556 121–132 (1992).
- 557 30. O. Hoegh-Guldberg *et al.*, Assisted colonization and rapid climate change. *Science*. **321**, 345–346
558 (2008).
- 559 31. S. N. Aitken, M. C. Whitlock, Assisted Gene Flow to Facilitate Local Adaptation to Climate Change.
560 *Annu. Rev. Ecol. Evol. Syst.* **44**, 367–388 (2013).
- 561 32. G. B. Dixon, L. K. Bay, M. V. Matz, Bimodal signatures of germline methylation are linked with gene
562 expression plasticity in the coral *Acropora millepora*. *BMC Genomics*. **15**, 1109 (2014).
- 563 33. C. Shinzato *et al.*, Using the *Acropora digitifera* genome to understand coral responses to environmental
564 change. *Nature*. **476**, 320-U82 (2011).
- 565 34. M. J. H. van Oppen, B. J. McDonald, B. Willis, D. J. Miller, The Evolutionary History of the Coral
566 Genus *Acropora* (Scleractinia, Cnidaria) Based on a Mitochondrial and a Nuclear Marker: Reticulation,
567 Incomplete Lineage Sorting, or Morphological Convergence? *Mol. Biol. Evol.* **18**, 1315–1329 (2001).
- 568 35. B. F. Voight, S. Kudaravalli, X. Wen, J. K. Pritchard, A map of recent positive selection in the human
569 genome. *PLoS Biol.* **4**, e72 (2006).
- 570 36. I. K. Jordan *et al.*, A universal trend of amino acid gain and loss in protein evolution. *Nature*. **433**, 633–
571 8 (2005).
- 572 37. A. Mckenna *et al.*, The Genome Analysis Toolkit: A MapReduce framework for analyzing next-
573 generation DNA sequencing data. *Genome Res.* **20**, 1297–1303 (2010).
- 574 38. P. Danecek *et al.*, The variant call format and VCFtools. *Bioinformatics*. **27**, 2156–8 (2011).
- 575 39. T. Jombart, adegenet: a R package for the multivariate analysis of genetic markers. *Bioinformatics*. **24**,
576 1403–5 (2008).
- 577 40. D. H. Alexander, J. Novembre, K. Lange, Fast model-based estimation of ancestry in unrelated
578 individuals. *Genome Res.* **19**, 1655–64 (2009).
- 579 41. T. Günther, G. Coop, Robust identification of local adaptation from allele frequencies. *Genetics*. **195**,
580 205–20 (2013).
- 581 42. Z. T. Richards, D. J. Miller, C. C. Wallace, Molecular phylogenetics of geographically restricted
582 *Acropora* species: Implications for threatened species conservation. *Mol. Phylogenet. Evol.* **69**, 837–851
583 (2013).
- 584 43. E. A. Treml, P. N. Halpin, D. L. Urban, L. F. Pratson, Modeling population connectivity by ocean
585 currents, a graph-theoretic approach for marine conservation. *Landsc. Ecol.* **23**, 19–36 (2008).
- 586 44. E. P. Chassignet *et al.*, The HYCOM (HYbrid Coordinate Ocean Model) data assimilative system. *J.*
587 *Mar. Syst.* **65**, 60–83 (2007).
- 588 45. R. C. Babcock *et al.*, Synchronous Spawning of 105 Scleractinian Coral Species on the Great-Barrier-
589 Reef. *Mar. Biol.* **90**, 379–394 (1986).
- 590 46. S. W. Davies, E. A. Treml, C. D. Kenkel, M. V. Matz, Exploring the role of Micronesian islands in the
591 maintenance of coral genetic diversity in the Pacific Ocean. *Mol. Ecol.* **24**, 70–82 (2015).
- 592 47. S. R. Connolly, A. H. Baird, Estimating dispersal potential for marine larvae: dynamic models applied
593 to scleractinian corals. *Ecology*. **91**, 3572–3583 (2010).
- 594 48. P. K. Smolarkiewicz, J. Szmelter, An MPDATA-based solver for compressible flows. *Int. J. Numer.*
595 *Methods Fluids*. **56**, 1529–1534 (2008).
- 596 49. R. K. Cowen, G. Gawarkiewicz, J. Pineda, S. Thorrold, F. Werner, Population Connectivity in Marine
597 Systems: An Overview. *Oceanography*. **20**, 14–21 (2007).
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599 **Supplemental Figures**

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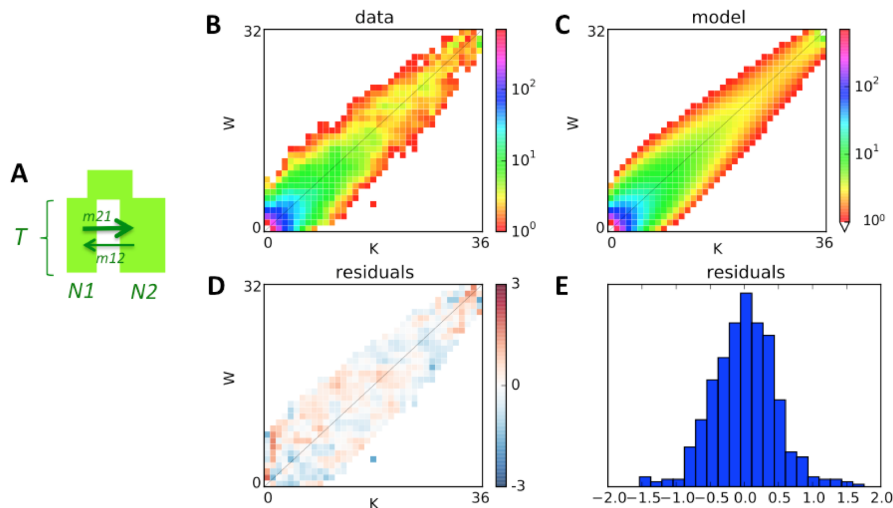
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Figure S1. ADMIXTURE analysis of genetic differentiation between populations. (A) Map of sampled locations with one-letter population identifiers. (B) ADMIXTURE plot of ancestry proportions with $K = 2$ (optimal K was 1).



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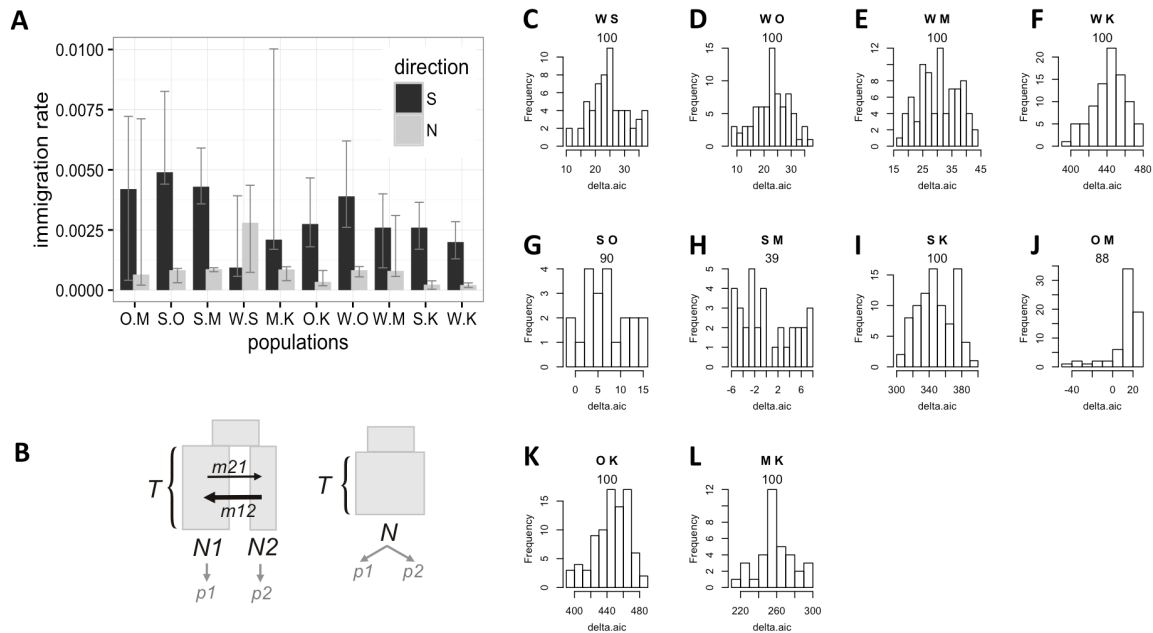
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Figure S2. Example of two-population *dadl* model fit. (A) The model: ancestral population splits into two populations of unequal sizes ($N1$ and $N2$) some time T in the past, which exchange migrants with different rates depending on direction. (B) Observed allele frequency spectrum comparing Wilkie (W) and Keppel (K) populations. (C) Allele frequency spectrum generated by the fitted model. (D, E) Map and histogram of residuals (absolute scale).

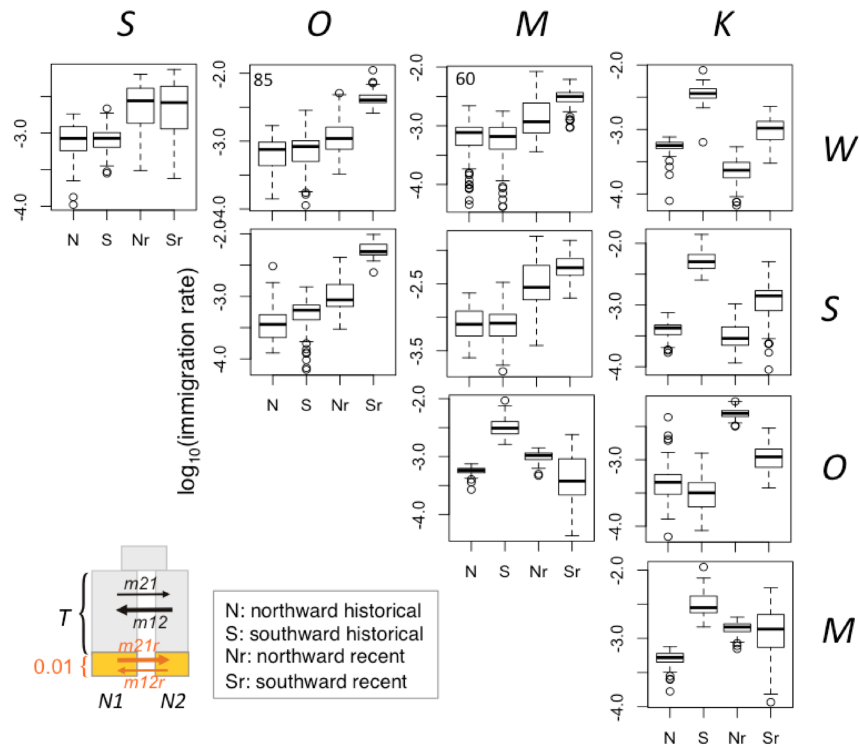
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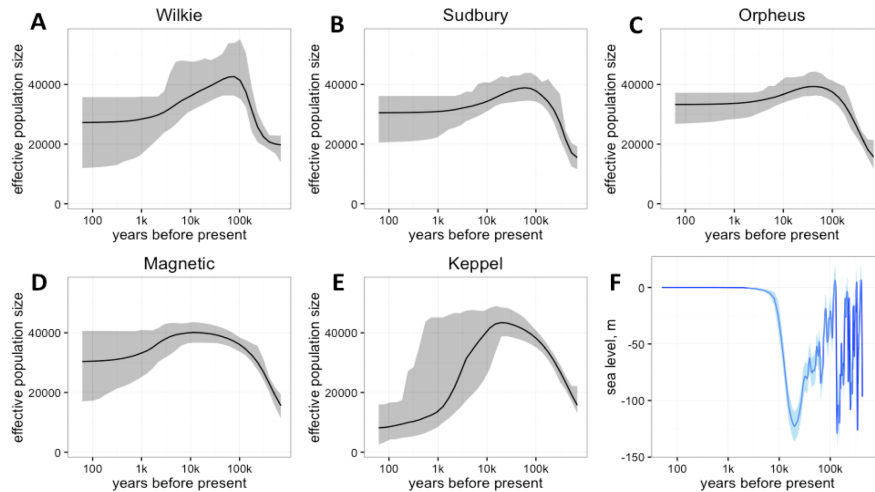
Figure S3. Bootstrap analysis of migration rates and population subdivision using *dadi*. (A) Migration among population pairs, with bootstrap-derived 95% confidence intervals. The pairs are identified on the x-axis and sorted by increasing geographical distance. Black bars – southward migration, grey bars – northward migration. (B) Models being compared: the full model (left) implies populations' split into two different sizes ($N1$ and $N2$) at time T in the past, since when they exchanged migrants at unequal rates depending on direction. Reduced model allows for population size change at time T in the past but does not include population split: the two genotyped groups ($p1$ and $p2$) are regarded as two samples from the same population. (C-L) Histograms of delta-AIC values for 100 bootstrap replicates (bootstrap was performed over genomic contigs of the draft genome of *A. digitifera*). Positive numbers indicate support for the full model. The letters on top of each panel identify compared populations, the number is the proportion of positive bootstrap replicates (i.e., bootstrap support for the full model). The only comparison that did not receive >50% bootstrap support for population split is between S and M populations (panel H).

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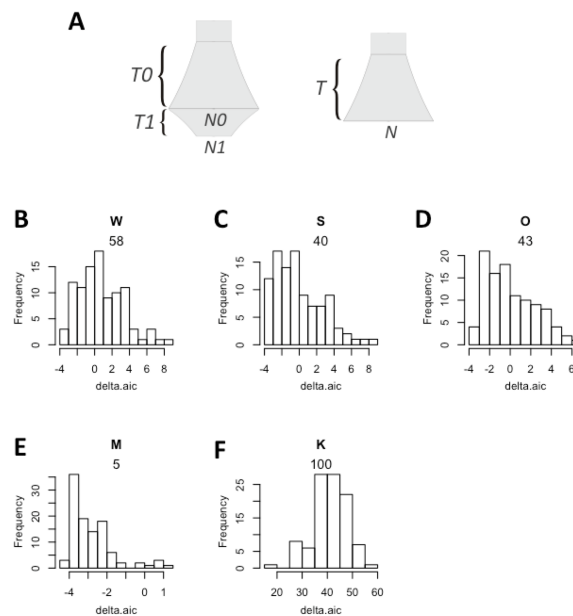
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Figure S4. Migration rates inferred by the *dadi* model allowing for the change in migration rates over the last 0.01 T units (15-20 generations or 75-100 years, in our case). Box plots show historical (N, S) and recent (Nr, Sr) migration rates inferred among pairs of population across 100 bootstrap replicates. Numbers in the top left corner of the WO and WM plots are delta-AIC bootstrap support values for the model with the recent change in migration when compared to the split-with-migration model with no recent change (Fig. 1A). All other pairs had less than 50% delta-AIC bootstrap support. There is no consistent recent change in the preferential direction of migration.



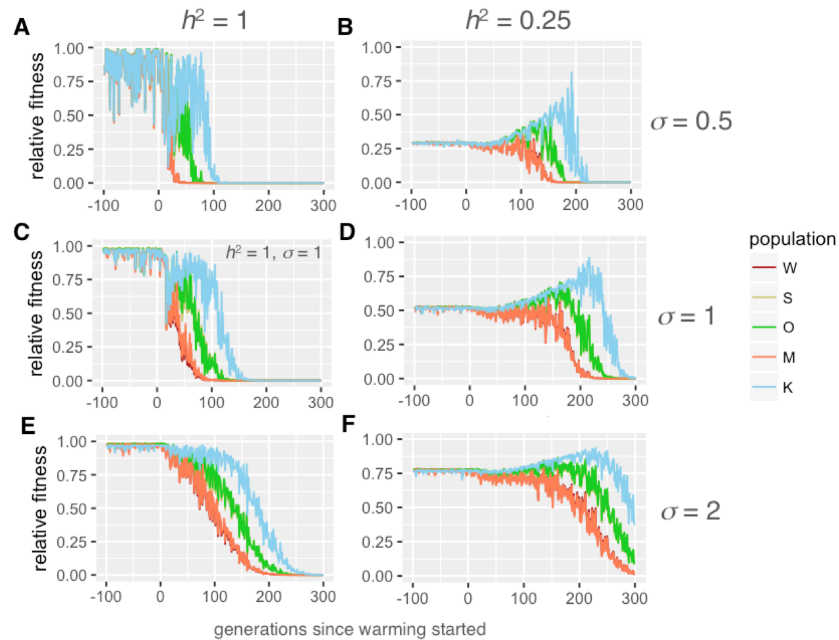
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Figure S5. Population history. (A-E) Historical population sizes with bootstrap-derived 95% confidence intervals, according to the two-growth model (Fig. S6 A). (F) Sea level with shaded area corresponding to standard error (41).



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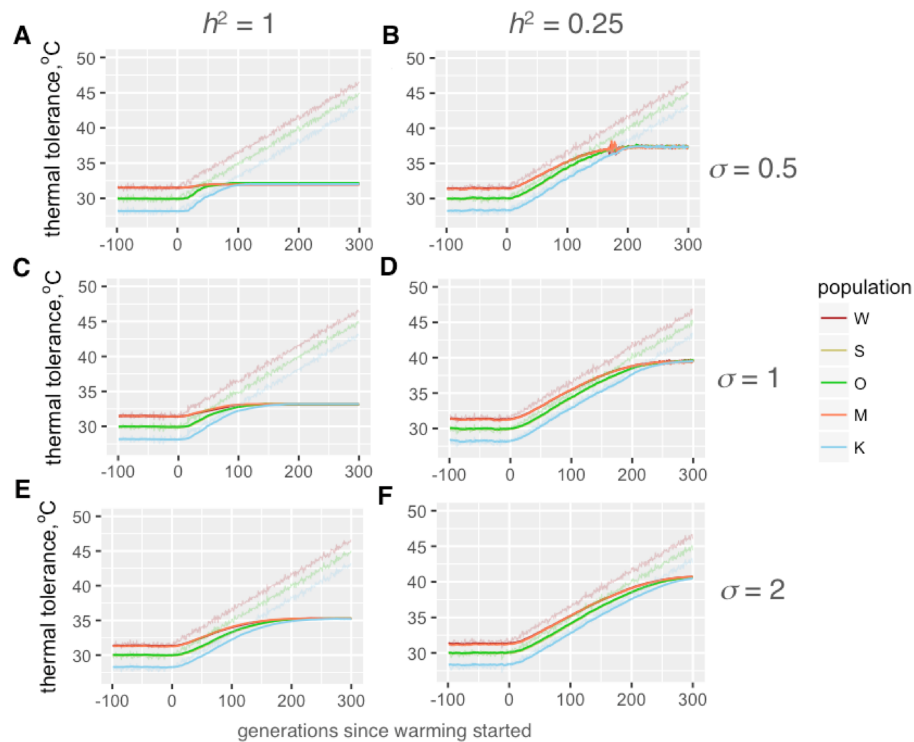
Figure S6. Delta-AIC bootstrap analysis of single-population models. (A) Models compared. The full model (left) includes two exponential growth periods (any of which could be growth or decline), the reduced model (right) has only one growth period. (B-F) Histograms of delta-AIC values for 100 bootstrap replicates. Positive numbers indicate support for the full model. The letter on top of each panel identify the population, the number is the proportion of positive bootstrap replicates (i.e., bootstrap support for the full model). The two-growth model is strongly supported for population K (panel F) and marginally supported for population W (panel B).



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Figure S7. Fitness of modeled populations after pre-adaptation period and under warming, depending on heritability of thermal tolerance (h^2 , proportion of phenotypic variation explained by genetics) and phenotypic plasticity (σ , standard deviation of the Gaussian slope of fitness decline away from the phenotypic optimum, in degrees C). X-axis is generations; warming starts at generation 0. Y-axis is fitness relative to maximal fitness at the genetically determined optimum. Warm-adapted populations (W and M) are shown as red-tint traces, populations from mild thermal regime (S and O) are green-tint traces, and the cool-adapted population (K) is the blue trace. Pairs of traces for warm- and mild-adapted populations largely overlap. (A, C, E): $h^2=1$. (B, D, F): $h^2=0.25$. (A, B): $\sigma = 0.5$. (C, D): $\sigma = 1$. (E, F): $\sigma = 2$. Higher plasticity facilitates metapopulation persistence during warming and confers stability against random fluctuations. Higher plasticity also partially rescues the drop in fitness achievable under low heritability (compare pre-warming generations, from -100 to 0, on panels B, D and F).

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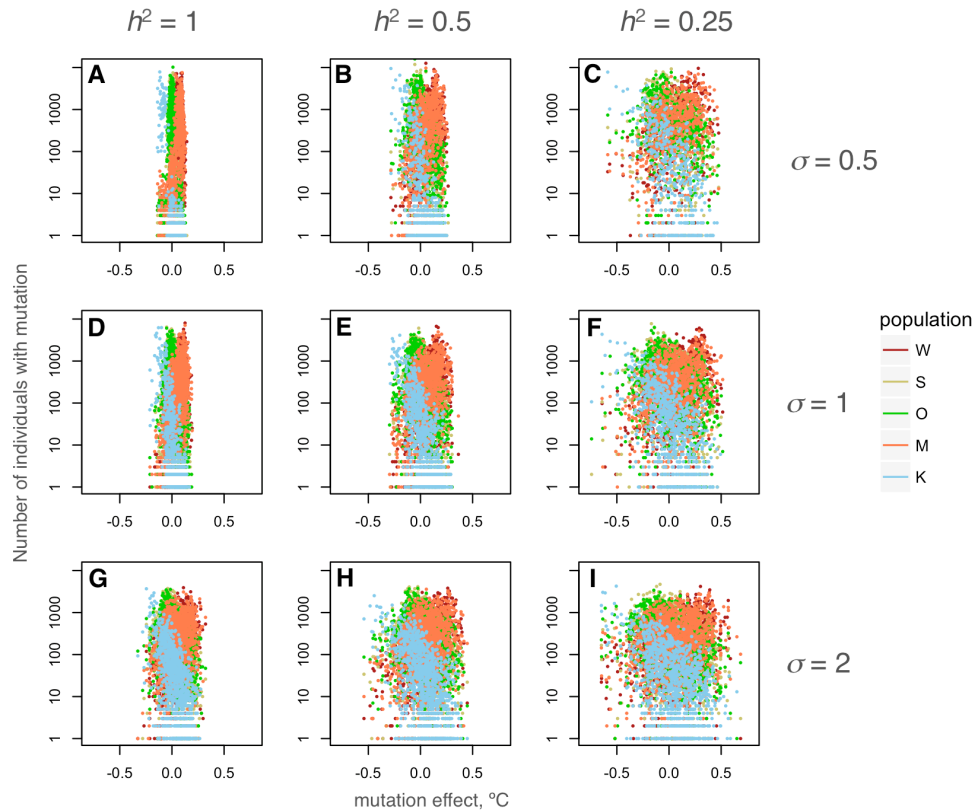
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683 Figure S8. Higher plasticity and lower heritability promote longer and more extensive evolution in response
684 to warming. The graphs show mean thermal tolerance of modeled populations after pre-adaptation period
685 and under warming, depending on heritability of thermal tolerance (h^2 , proportion of phenotypic variation
686 explained by genetics) and phenotypic plasticity (σ , standard deviation of the Gaussian slope of fitness
687 decline away from the phenotypic optimum, in degrees C). X-axis is generations; warming starts at
688 generation 0. Y-axis is thermal tolerance (mean phenotype of the population). Warm-adapted populations
689 (W and M) are shown as red-tint traces, populations from mild thermal regime (S and O) are green-tint
690 traces, and the cool-adapted population (K) is the blue trace. Thin noisy lines are modeled temperatures at
691 the corresponding locations. Pairs of traces for warm- and mild-adapted populations largely overlap. (A, C,
692 E): $h^2=1$. (B, D, F): $h^2=0.25$. (A, B): $\sigma = 0.5$. (C, D): $\sigma = 1$. (E, F): $\sigma = 2$.

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Figure S9. Higher plasticity (σ) and lower heritability (h^2) promote retention of higher genetic variation in thermal tolerance. The scatterplots show the dependence of the number of individuals in a population bearing a mutation at a thermal QTL locus on the mutation's effect size (change in thermal tolerance, in °C) at the end of the pre-adaptation period (2000 generations with no directional change in temperature). The starting standing genetic variation was the same in all simulations. (A,D,E): $h^2=1$. (B,E,H): $h^2=0.5$. (C,F,I): $h^2=0.25$. (A-C): $\sigma = 0.5$. (D-F): $\sigma = 1$. (G-I): $\sigma = 2$. Populations are colored according to the color scheme used in Figures 3, S7 and S8 (see legend).