1   **Linking the chemistry and reactivity of dissolved organic matter from low-latitude glaciers**

2   **and rock glaciers**

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Abstract

As glaciers in the western United States thaw in response to warming, they release dissolved

organic matter (DOM) to alpine lakes and streams. Biological availability of DOM from small

mountain glaciers is unknown. Differences in DOM bioavailability between glacier types like

rock and ice glaciers remains undefined, yet rock glaciers outnumber ice glacier approximately

ten to one at low latitudes. To assess which components of aquatic DOM are most reactive and

the potential for glacial DOM from low latitude glaciers to subsidize heterotrophy in alpine

headwaters we evaluated reactivity and molecular composition of DOM from ice glaciers and

rock glaciers from four paired catchments (each with a glacier and rock glacier at their

headwaters). Biological reactivity was linked to molecular composition by evaluating the

chemical characteristics of each DOM pool pre- and post-incubation using common microbial

community laboratory assays paired with untargeted mass spectrometry-based metabolomics.

Glacier and rock glacier DOM was similar in concentration and chemodiversity, but differed in

composition. When incubated with a common microbial community, DOM from ice glacier

meltwaters contained a higher proportion of bioavailable DOM (BDOM), and resulted in greater

bacterial growth efficiency. Differences in DOM reactivity between glacier types was

determined by differences in the relative abundance of only a few dozen compounds. Though

BDOM was lower in rock glaciers, because rock glaciers are more abundant and are expected to

have greater longevity, we propose that both glacial types will be important sources of

bioavailable DOM to alpine headwaters over the coming years to decades.

Introduction

Glaciers bridge the atmospheric-terrestrial-aquatic interface, integrating atmospherically deposited chemicals with weathering products and releasing reactive solutes to adjacent surface waters (Williams et al. 2007; Dubnick et al. 2010; Fellman et al. 2010; Stibal et al. 2010; Stubbins et al. 2012). Between continental and mountain glaciers, mountain glaciers release the largest flux of carbon from melting ice annually (Hood et al. 2015). Dissolved organic matter (DOM) from these large mountain glaciers of the European Alps and Alaska has been shown to stimulate heterotrophic respiration (Hood et al. 2009; Singer et al. 2012). Glaciers and rock glaciers, each with distinct geophysical attributes, are common in many of the alpine headwaters of the western U.S. (Fegel et al. 2016). Glaciers are massive ice bodies that form and persist in areas where annual snow accumulation is greater than annual snow ablation at decadal or longer time spans. Rock glaciers are flowing bodies of permafrost, composed of coarse talus and granular regolith both bound and lubricated by interstitial ice (Berthling, 2011). In the western United States, rock glaciers are an order of magnitude more abundant than ice glaciers and are more resistant to warming temperatures than ice glaciers (Fegel et al. 2016). Yet little is known about the quantity or quality of DOM being released from smaller glaciers or rock glaciers in mountain headwater ecosystems or their potential to affect low-latitude alpine ecosystems.

The chemical composition of DOM from natural aquatic systems is complex (Hedges et al. 2000; Kim et al. 2006; Hockaday et al. 2009), and glacier and rock glacier meltwaters are not likely to be an exception. Previously, assessment of the bioavailability of DOM relied on bioassays that measure the rate and amount of DOM consumed over time (e.g. Amon and Benner 1996; del Giorgio and Cole 1998; Guillemette and del Giorgio 2011). While some coarse characterizations of molecular composition (Benner 2002; Berggren and del Giorgio 2015) or the

lability of individual compounds (del Giorgio and Cole 1998) have previously been applied to

characterize bioavailable DOM (BDOM), only recently has the research community begun to

apply high-resolution analytical chemistry to assess the molecular characteristics of

environmental DOM (e.g. Kellerman et al. 2014, 2015; Andrilli et al. 2015). Whereas no single

method can identify the entire spectrum of compounds present in an environmental DOM pool

(Derenne and Tu 2014), mass spectrometry allows for molecular identification of thousands of

specific DOM compounds, simultaneously yielding a more specific and broader metric for DOM

analysis compared to spectroscopic techniques such as fluorescence or ultraviolet absorption.

Simultaneous identification and quantification of total BDOM is not an easy task. Difficulties

in experimentally connecting molecular characterization of DOM pools to their bioavailability

are partly due to the highly diverse constituent compounds that compose natural DOM (Derenne

and Tu 2014). The effects of individual compounds identified to be bioavailable (e.g. amino

acids in glacial meltwaters (Fegel et al. 2016) may not be representative of the total BDOM pool.

It is also possible that DOM with a high diversity of labile compounds may have positive

feedbacks on the bioavailability of the total DOM pool by asserting priming effects on the total

pool (Guenet et al. 2010). In addition, certain metabolites within DOM pools may not be

bioavailable individually, but may act as cofactor metabolites that allow for a mutualistic

increase in bioavailability [Hilker 2014]. Thus, the relationship between DOM pool

characteristics and lability remains largely unknown in most natural systems. However, some

patterns are beginning to emerge, like the consistently high lability of proteinaceous DOM that

has been found in glacial, estuarine, and marine environments [Andrilli et al. 2015]. Defining

these fundamental relationships between chemical composition of DOM and lability has the

potential to provide a better understanding for how DOM pools contribute to heterotrophy across

a broad spectrum of aquatic ecosystems.

There are defining characteristics in the origin of DOM that differentiate glaciers from rock

glaciers. DOM derived from ice glaciers is structured by in situ microbial activity and can be an

important source of chemical energy to headwater ecosystems (Hood et al. 2009; Singer et al.

2012; Fellman et al. 2015; Fegel et al. 2016). DOM derived from rock glaciers is a

conglomeration of carbon compounds percolated through from vegetation growing on the rock

glacier surface and microbial processing within the rock glacier itself (Wahrhaftig and Cox,

1959; Williams et al. 2007; Fegel et al. 2016). Previous research has shown that DOM released

from glaciers and rock glaciers in the western United States differ in their optical properties

(Fegel et al. 2016). Whether or not these differences in the optical properties of DOM between

glacial types translates to differences in their reactivity or proportion of BDOM remains

unknown.

To address this, we asked whether differences in the composition of DOM between ice

glaciers and rock glaciers affected differences in microbial metabolism and whether DOM

chemistry in ice and rock glacier meltwater in the western United States is similar to what has

been reported for other glacial meltwaters. Here we present the results of laboratory incubations

of DOM from ice glacier and rock glacier meltwaters with a common microbial community.

Incubations were bookended (i.e. analyzed before and after incubation) with non-targeted

metabolomic analysis of DOM via gas chromatography mass spectrometry (GC-MS) to

determine differences in the specific chemical compounds metabolized by microbial processing.

By exposing DOM from different sources to a common microbial community we were able to

assess the lability of each DOM source and analyze which differences in the molecular

composition of DOM affected its lability and carbon use efficiency.

Methods

Site Description

Paired ice glaciers and rock glaciers within four watersheds on the Front Range of Northern

Colorado were selected based on their size ($>0.5km^2$) and the proximity to each other within the

watershed, forming pairs of features with similar geographic parameters (Figure 1). We collected

samples of glacier meltwater in the late summer to capture the greatest contribution of ice melt

and minimize annual snowmelt contribution. We sampled the four pairs in September 2014.

Sample sites were: Isabelle Glacier (-105.640994,40.063373) and Navajo Rock Glacier (-

105.636092,40.061200), Arapaho Glacier (-105.646351,40.023378) and Arapaho Rock Glacier

(-105.637699, 40.022482), Peck Glacier (-105.663810,40.068332) and Peck Rock Glacier (-

105.664310,40.071642) in the Indian Peaks Wilderness west of Boulder, CO; and Andrews

Glacier (-105.680639,40.288370) and Taylor Rock Glacier (-105.671197,40.275568) in the Loch

Vale Watershed in Rocky Mountain National Park (Figure 1a). A complete site description for

each site is given in Fegel et al. (2016).

Field extraction of DOM

At each glacier, meltwaters were collected in the early to mid-morning (0500-1000) to

minimize diurnal variability in ice melt from solar radiation. DOM was extracted from 20 L of

meltwater collected at the terminus of each feature in the field using a slightly modified protocol

established by Dittmar et al. (2008, Supplemental Information). Briefly, meltwater samples were

passed through pre-combusted (450° C, 5hr) Whatman GF/F filters (GE Whatman, Pittsburg,

PA, USA), acidified to ~ pH 2 with 32% HCl, and concentrated using Bond Elut PPL carbon

extraction cartridges (Agilent, Santa Clara, CA, USA). Concentrated DOM was eluted in the

field with 10mL HPLC grade methanol per cartridge into cleaned, combusted, and pre-weighed

120mL borosilicate bottles.

Metabolite Analyses

Once in the laboratory, pre-incubation DOM samples were dried under a clean N_2 stream and

weighed (http://www.nrel.colostate.edu/projects/lvws/data.html). Each sample (n=8) was

prepared for metabolomic analysis by dissolving the dried OM into fresh HPLC-grade methanol

to a final concentration of 2 mg ml^{-1}.

After the microbial bioassays (in which all samples were run in duplicate), post-incubation

DOM samples were collected separately from each individual microcosm (n=16). Water from

each microcosm was filtered through pre-leached 0.2 μm Millipore filters (EMD Millipore,

Billerica, MA, USA) to remove microbial biomass, freeze-dried, and the total remaining DOM

was weighed and re-dissolved into HPLC-grade methanol

(http://www.nrel.colostate.edu/projects/lvws/data.html). In order to increase the volatility of

molecules analyzed through GC-MS, samples were derivatized with trimethylsilane (TMS) using

standard protocols (Supplemental Materials) (Pierce, 1968).

Metabolomics

Both pre- and post- incubation OM samples were analyzed with inline gas chromatography-

mass spectroscopy (GC-MS) at the Proteomics and Metabolomics Facility at Colorado State

University. Metabolites were detected using a Trace GC Ultra coupled to a Thermo ISQ mass

spectrometer (Thermo Scientific, Waltham, MA, USA). Samples were injected in a 1:10 split

ratio twice in discrete randomized blocks. Separation occurred using a 30 m TG-5MS column

(0.25 mm i.d., 0.25 μm film thickness, Thermo Scientific, Waltham, MA, USA) with a 1.2 mL

min^{-1} helium gas flow rate, and the program consisted of 80°C for 30 sec, a ramp of 15°C per

min to 330°C, and an 8 min hold. Masses between 50-650 m/z were scanned at 5 scans sec^{-1} after

electron impact ionization (Broeckling et al. 2014).

DOM lability experiments

Concentrated DOM samples from each of the eight study sites were diluted to 4 mg L^{-1} C and

incubated *in vitro* with a natural microbial community collected from The Loch, a small sub-

alpine lake in Rocky Mountain National Park, CO, USA, (-105.6455, 40.2976). Unfiltered lake

water collected from The Loch was aged for 2 years at 5 °C in order to remove the majority of

the bioavailable carbon. At the start of the experiment DOM concentration of the lake water was

0.7 mg L^{-1} C. At the initiation of the incubations 2 L of aged lake water was filtered through a

pre-combusted (450° C, 5hr) Whatman GF/C filter (1.2μm nominal pore size) (GE Whatman,

Pittsburg, PA, USA) to remove bacterial grazers (e.g. protists and metazoans). Three aliquots of

filtered-aged lake water was preserved with 2% formalin (37% formaldehyde), and set aside for

enumeration of bacteria at the initiation of the experiment (i.e. t=0), and a second set of three

aliquots (7mL) was used for initial DOC/TN analysis on a Shimadzu TOC-VWS analyzer

(Shimadzu Corp., Kyoto, Japan). To create normalized concentrations (4 mg L^{-1} C), in each

incubation bottle we added 60.96 mL of unfiltered Loch water, between 3.80 and 9.06mL of

concentrated DOC solution (depending on the initial concentration), and filled to 70mL total

volume with DI; resulting in standardized concentrations of 4 mg L^{-1} C in each incubation bottle.

During the experiment, microcosms that received DOM from ice glacier and rock glacier sites

were incubated alongside control incubations that contained lake water and the common

microbial community but with no added DOM (i.e. experimental control). In addition, an

analytical blank of MilliQ water was incubated to correct for any instrumental drift that occurred

during the experiment.

All microcosms were incubated simultaneously at 15 °C for 10 weeks. To calculate microbial

respiration, we measured changes in dissolved oxygen (DO) at 1-minute intervals in each

microcosm using an Oxy-4 fiber-optic dissolved oxygen probe (PreSens, Regensburg, Germany).

The incubation was terminated when the fastest metabolizing microcosm approached 4mg L^{-1}

DO to avoid hypoxia and the potential for anaerobic metabolism. All measurements with

amplitude less than 20000 amps were removed because of the potential for inaccurate readings of

DO. Absolute values from the raw fiber optic measurements were corrected for analytical drift

by subtraction of changes in signal from the MilliQ water analytical control over the course of

the experiment.

Bacterial Cell Counts and Bulk Chemistry

From each microcosm we collected a 2 mL aliquot post-incubation and preserved it with 2%

formalin (final concentration) to assess changes in cell abundance during the course of the

incubation. Aliquots were filtered onto 0.2 μm Millipore polycarbonate filters (EMD Millipore,

Billerica, MA, USA) and stained with Acridine Orange (Hobbie et al. 1977, Supplemental

Information) for enumeration. Post-incubation we assessed TOC and TDN

(http://www.nrel.colostate.edu/projects/lvws/data.html) for each microcosm. To ensure that

filtration did not contribute significant additional carbon to the dissolved carbon pool, a subset of

blank Millipore filters flushed with MilliQ was analyzed for TOC and TN. All filtration of

microcosm water for chemical analysis was conducted on pre-leached Millipore filters that

contributed insignificant amounts of carbon to the filtrate

(http://www.nrel.colostate.edu/projects/lvws/data.html).

Data Analysis

Metabolite Analysis

Compound annotation was prioritized based on order of normalized intensity and statistical

difference between glacier types. Putative identification (candidate compounds based on

physicochemical characteristics and spectral similarity with spectral libraries (Sumner et al.

2007), was made using spectral matching to in-house, NISTv12, Golm, Metlin, and Massbank

metabolite databases. Clustered features assigned candidate compounds in our study had high

similarity (>90%) to spectra from known standard compounds within the databases used

(Supplemental Information). Molecular rank was calculated by ordering candidate compounds

by their normalized ion intensity for pre- and post- incubation sample averages for all features.

Chemical diversity was calculated using the Shannon-Wiener diversity index (Shannon 1948) by

treating each unique chemical compound identified though GC-MS as a 'species'. This was done

for DOM composition both before and after incubation in order to estimate changes in chemical

diversity through microbial metabolism.

Analysis of DOM Reactivity

Oxygen consumption rates were fit to a Berner-Multi-G two-pool decay model to estimate the

size of the labile pool (BDOM) and the recalcitrant pool. Oxygen consumption was averaged for

each glacier type and confidence intervals were calculated at $\alpha=0.05$. Data were smoothed using

a third order polynomial ($R^2>0.999$) and 95% confidence intervals were plotted. Berner's Multi-

G model was used to model carbon pool bioavailability (Berner 1980; Guillemette and del

Giorgio 2011), using SAS. Dissolved oxygen curves generated from the incubation were fit to

the Equation:

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$$Y = B_1^{kt} + B_0$$

where Y is the total carbon pool, B_1 is the bioavailable carbon pool, k is the decay rate

constant of the bioavailable pool, t is time, and B_0 is the recalcitrant carbon pool. We used a least

square means to test for statistical differences between ice glaciers and rock glaciers in the size

of the B_1 and B_0 pools. Total C consumed was calculated as the difference in pre- and post-

incubation DOC values.

To address the metabolic quality of consumed carbon we calculated Respiratory Quotient

(RQ) as the carbon consumed (in mg L^{-1}) divided by the oxygen (as O_2) consumed (in mg L^{-1}).

Bacterial growth efficiency (BGE) was calculated to examine how efficiently each carbon

source was incorporated into bacterial biomass. To calculate BGE we divided bacterial

production (BP) by the sum of bacterial respiration and bacterial production (BR + BP)(del

Giorgio and Cole 1998). Thus, BGE is a ratio of carbon incorporated into biomass relative to

total carbon consumed for the bacterial community. BP was measured as change in cell number

over time and converted to units C using an estimate of 20 fg C per bacterial cell (Borsheim and

Bratbak 1987) as a conversion. Carbon consumption rate was measured as carbon consumed

during the course of the incubation by measuring DOC content in each filtered pre- and post-

incubation sample using a TOC/TDN analyzer.

Statistical Analyses

The DOM composition for each treatment (glaciers vs. rock glaciers) was compared with ANOVA for each compound using the aov function in R, and p-values were adjusted for false positives using the Bonferroni-Hochberg method in the p.adjust function. Post-incubation samples were corrected for compounds added by the lake water by subtracting the peak intensities for each chemical candidate within the lake water (i.e. experimental control) from each post incubation sample.

We identified differences in compounds present in each sample using PCA conducted on mean-centered and pareto variance-scaled data using the pcaMethods package in R. We analyzed GC-MS spectra using principal component analysis (PCA) in R (R Core Team, 2014). For each sample, raw data files were converted to .cdf format, and a matrix of molecular features as defined by retention time and mass to charge ratio (m/z) was generated using XCMS package in R for feature detection and alignment. Raw peak areas were normalized to total ion signal, outlier features were detected based on total signal and PC1 of PCA, and the mean area of the chromatographic peak was calculated among replicate injections (n=2). We grouped all spectral features based on an in-house clustering tool, RAMClustR, which uses spectra based coelution and covariance across the full dataset groups features (Broeckling et al. 2014). We used the t-test configured for non-parametric Welch-Satterthwaite test to compare differences in RQ, BGE, chemodiversity, and loss of chemodiversity between glacier types. Our use of a Welch-Satterthwaite test allowed for comparison of samples of unequal variance and distribution.

Results

Ice glacier and rock glacier DOM composition

Neither the chemical diversity (mean=2.82), nor the C:N ratio (mean=2.1) differed between

glaciers and rock glaciers (Table 1). In addition, the molecular composition was similar in each

glacier type for the top 25 most abundant molecules. Combined, the glacier and rock glacier

DOM molecular composition consisted of 2033 compounds. Each individual compound

consisted of 3 to170 individual mass spectral features, for a total of 14,571 unique mass spectral

features in the complete dataset. Of the 2033 compounds, 328 were chosen for annotation based

either on high ion intensity (n=278) or significantly different relative abundances between glacier

and rock glacier sources (n=33). Annotated compounds consisted primarily of simple sugars,

amino acids and other organic acids (http://www.nrel.colostate.edu/projects/lvws/data.html).

There were significant differences between glacier types in the molecular composition of less

abundant compounds (Table 2), although the C:N ratio, chemical diversity, and compounds with

greatest ion intensity were similar. PCA analysis of the complete dataset suggested 33 key

compounds, ordinated on axes three and five of our PCA analysis (Figure 2A), were responsible

for differences in the ice and rock glacier DOM pools (Table 2). Of those 33 compounds, 5 were

assigned putative structures (Table 2, also see

http://www.nrel.colostate.edu/projects/lvws/data.html) with ice glacier meltwaters significantly

enriched in the simple sugar maltose and the amino acid glutamate, and rock glacier meltwaters

enriched in the primary organic acids, glycolate, threonate and quinate (Table 2). The other 28

compounds represented components of the DOM pool that differentiated DOM composition

from each glacier type, but were not present in the databases we employed and had no known

structural analogs.

Incubations

The bacterial respiration rates of DOM from each glacier type were not significantly different,

as estimated by the decay constant k (Table 3). However, a significantly larger portion of ice

glacier DOM was bioavailable, (i.e. BDOM, $B_1 = 58.8 \pm 9.7\%$) compared to rock glacier DOM

($B_1 = 37.3 \pm 10.2\%$, $p < 0.01$, Table 3). Bacterial growth efficiency (BGE) was higher for

microbial communities incubated with ice glacier DOM compared to rock glacier DOM (G =

0.26 ± 0.13, RG = 0.16 ± 0.16, Table 3). In general, more oxygen was consumed per mg organic

carbon metabolized in incubations that contained glacial DOM compared to those that contained

rock glacier DOM, even though the amount of carbon consumed between treatments was similar

(Figure 4, Table 3). This resulted in a lower respiratory quotient (RQ) for glacial derived DOM

compared to rock glacier derived DOM overall.

Post-incubation DOM analysis

Analysis of DOM after the incubation period allowed us to assess how microbial metabolism

altered the composition of DOM. Incubation with a common microbial community both rarified

and homogenized DOM between glacier types, resulting in fewer compounds with high ion

intensities, and more compounds with low ion intensities (Figure 3). This reorganized the

molecular rank-abundance curve, resulting in a different set of compounds with the highest

relative abundance in the pre vs. post-incubation dataset (Table 4). However, similar to pre-

incubation DOM composition, post-incubation ice glacier and rock glacier derived DOM shared

the same abundant compounds (Figure 2B)

(http://www.nrel.colostate.edu/projects/lvws/data.html). Interestingly, the chemical diversity of

DOM between glacier types diverged during the course of the incubation. At the end of the

incubation glacial DOM had significantly higher diversity and rock glacier DOM had

significantly lower diversity compared to pre-incubation DOM for the same glacial type (Table

1, Figure 2C). Combined with molecular rank abundance data (ordered on ion intensity), this

change in chemical diversity resulted in an increase in the richness of glacier DOM pools by

microbial metabolism and a decrease in richness in rock glacier DOM during the incubation. For

both DOM sources, many of the organic acids and sugars present pre-incubation were consumed

while amino acids that were present at lower intensities in pre-incubation DOM pools increased

in relative abundance in post-incubation DOM pools. While there were differences in the

chemical diversity of rock and ice glacier derived DOM post-incubation, PCA suggested were no

differences in the overall molecular composition between glacier types (Figure 2B). This trend

did not follow for all rock glaciers, as one replicate of Taylor Rock Glacier (lowest left point,

Figure 2A & 2B) saw few changes in its principal components after the incubation. Interestingly,

while no significant difference was seen in GC-MS compounds between glacier type post

incubation (Figure 2B), both replicates of Peck Rock Glacier (lower rightmost points Figure 2B)

showed separation along the y-axis (PC1) of their PCA, a separation that was not seen before

microbial metabolism (Figure 2A). The reasons for this remain unclear, but may be due

geographical separation (climatological differences) between Peck Rock Glacier and all other

sites.

Discussion

Our results demonstrate that chemically complex DOM released from Rocky Mountain ice

glaciers and rock glaciers stimulated bacterial respiration and productivity. Each DOM source

had unique molecular characteristics that reflected their quality as an energy source for microbial

metabolism. Glacier and rock glacier DOM shared many of the same organic compounds, but

differences in the relative concentrations of just 33 compounds appeared to drive differences in

the bioavailability of DOM between glacier types. Enrichment in simple sugars of glacier DOM

contributed to higher bioavailability of glacier DOM as measured by size of the BDOM pool and

bacterial growth efficiency (BGE). The DOM from rock glaciers, was enriched in less

bioavailable, primary organic acids. In spite of these differences, the bacteria homogenized

DOM from both sources during the incubations, resulting in more similar DOM composition at

the end of the incubation compared to the composition of the DOM before the incubation. In our

experiment, the source of the DOM affected its composition and reactivity and interaction with

the microbial community metabolism re-structured its chemistry.

Glaciers are a source of labile DOM

The results from our study of glaciers in Colorado showed similar patterns to previous studies

of glacial DOM in the Arctic and the European Alps, where bioavailable carbon from glaciers

also stimulated microbial production (Hood et al. 2009; Singer et al. 2012; Fellman et al. 2015).

Carbon concentrations in meltwaters from ice and rock glaciers in our study were low (0.5- 1.5

mg C L^{-1}), but similar to the global average of 0.97 mg C L^{-1} that has been previously reported

for glaciers (Dubnick et al. 2010; Stubbins 2012; Singer et al. 2012; Hood et al. 2015). The

percentage of DOM that was bioavailable (BDOM) from ice glacier DOM in our study (~50%)

was comparable to values of BDOM seen in DOM derived from glaciers in the European Alps

(58%), and within the range of values reported from glaciers in the Alaskan Range (23-66%)

(Hood et al. 2009; Singer et al. 2012). In all of these studies, including the results reported here,

glaciers had higher BDOM concentrations than other surficial freshwaters (Volk et al. 1997).

Previous reports on freshwater noted 16.5-34.5% of BDOM, which more closely resembled the

amount of BDOM we report from rock glacier BDOM (37% on average).

We propose that ice glaciers have higher proportions of BDOM compared to rock glaciers for

two reasons. First, rock glaciers host mosses, lichens, and vascular plants, including woody

shrubs and trees (Wahrhaftig and Cox, 1959; Cannone and Gerdol 2003; Burga et al. 2004).

Dissolved organic compounds released from these allochthonous phototrophs include a wide

variety of complex organic acids and polymers that are potentially difficult for microbes in

freshwaters to metabolize, resulting in a higher proportion of recalcitrant DOM within rock

glaciers (Wetzel 1992, Rovira and Vallejo 2002). The potential for the development of protosoils

within the rock glacier matrix may promote the presence of liquid pore water. DOM from rock

glaciers may already be partially processed by microbes within this protosoil environment by the

time it reaches the sub-rock-glacial surface water. This pre-processing may result in DOM with a

level of lability more similar to what is delivered to inland surface waters and derived from the

soil SOM matrix (Fellman et al. 2010). Conversely, BDOM in ice glaciers comes from

autocthonous sources (i.e. microbial production and processing on the ice surface and within the

ice interstitial space (Fellman et al. 2009) and atmospherically deposited aerosols (Hood et al.

2012), both of which appear to be more biologically available than terrestrial sources. Ice glacier

BDOM may be preserved in the ice matrix and physically inaccessible to microbial degradation.

In contrast, rock glacier DOM may be in contact with intraglacial sediments and liquid water,

allowing for further processing through microbial metabolism. In addition, sub-glacial

environments below ice glaciers are often anoxic (Tranter et al. 2005), allowing some

compounds to remain in a chemically reduced state until released with ice melt. This would

allow for rapid metabolism of these energetically-preferable, reduced compounds once unlocked

from the ice and released to an oxic environment (Hood et al. 2009). Thus the production of

DOM within each feature differs and pathways by which that DOM is delivered to the glacial

meltwater differ. These mechanisms would result in the differences in the proportion of BDOM

we observed in our bioassays of DOM lability.

The molecular structure of glacial DOM from large glaciers has been mechanistically linked to

its lability and shown to be more labile than marine or freshwater derived material due to the

presence of protein and amino-sugar components of microbial origin (Andrilli et al. 2015). This

is consistent with studies from other inland waters that show that higher DOM bioavailability

correlates with enrichment of amino acids and simple sugars (Lafreniere and Sharp 2004;

Williams et al. 2007; Dubnick et al. 2010). Taken together with the results presented here, we

suggest that DOM enriched in amino acids and simple sugars may be a key trait that defines

BDOM and be characteristic of glacial meltwaters worldwide.

Glacier DOM Diversity

The diversity of compounds in organic matter released from glaciers in this study was similar

to the DOM diversity found in other freshwater systems (Dubnick et al. 2010, Guillemette and

del Giorgio 2011, Kellerman et al. 2014). We found that chemical diversity was very similar

between ice glaciers and rock glaciers. However, the diversity of DOM between glacial types

was altered by microbial metabolism in different ways when incubated. Chemical diversity in

incubations with glacial DOM increased during the course of the incubation, whereas chemical

diversity decreased in incubations with rock glacier DOM. This suggests that when BDOM is a

smaller fraction of total DOM, microbial communities may use previously processed DOM to

fuel their own heterotrophic metabolism. This would reduce chemical diversity through the

consumption of primary metabolites (i.e. simple organic acids).

Our study expands the understanding of DOM complexity in inland waters by assessing the

glacial contribution of DOM to alpine headwaters and moving beyond broad functional groups

and compound classes identification to identification of specific compounds (Dubnick et al.

2010; Fellman et al. 2010; Singer et al. 2012, Fegel et al. 2016). By using a mass spectrometry

based technique we were able to identify individual candidate compounds through metabolomics

paired with total chemical pool complexity. Our results are more detailed, but consistent with

previous analyses that used a fluorescence index (Fegel et al. 2016). Results from Fegel et al.

(2016) showed higher fluorescence index (FI) values in ice glacier meltwaters compared to that

of rock glaciers, indicating a higher presence of proteinaceous (i.e. nitrogen containing)

compounds in ice glacier meltwaters. Here, using metabolomics, we identified higher

proportions of specific amino acids within ice glacier meltwaters compared to rock glaciers that

are likely the cause of the increased FI values observed (Fegel et al. 2016). Application of a

community level analytical approach such as mass spectrometry-based metabolomics provides

the opportunity to disentangle the critically important components of complex DOM pools for

aquatic C cycling, however only a few studies have used metabolomic approaches to address the

bioavailability of DOM as a heterotrophic subsidy (see Kujawinski 2011 for a review; Logue et

al. 2015).

Whereas our results provide unambiguous evidence that differences in bioavailability are due

to chemical differences between DOM that differ in origin from glacier types, some components

of the DOM pool are likely not assessed by the GC-MS we employed. GC-MS measures

molecules up to 1200 amu with electron impact ionization, leaving larger molecules unidentified.

Further, some of the components important for glacial C processing could not be identified using

the most current databases. While we were unable to determine exactly what those compounds

were, we are able to say what they were not. Many of the unannotated compounds that were

significantly different between glaciers and rock glaciers had m/z (mass to charge ratio) values

greater than 200 (Table 2), indicative of secondary metabolites with large molecular weights

relative to primary metabolites, most of which have mass to charge (m/z) values between 60 and

205. Products of primary metabolism, e.g. amino acids, organic acids, sugars and peptides are

well represented in mass spectrometry databases and therefore more easily identified with a GC-

MS approach. Less well represented are larger molecular weight products of secondary

metabolism such as terpenes, alkaloids, polyketides, aromatic structures, and products of mixed

biosynthesis. Known metabolites are often a small portion of data obtained through mass

spectrometry (<10%), with much of the data reflecting unknown metabolites or those yet to be

verified with standards (Jansson et al. 2009). Yet, the quality of mass spectrometry databases and

representation is rapidly improving, and this will likely be a critically important source of

information for understanding the relationship between molecular composition of DOM and its

lability in future research. The high proportion of compounds that could not be annotated with

candidate assignments in our study reflects the infancy of using mass spectrometry databases in

environmental applications. It exposes the need for more environmentally derived spectra and

standards from secondary metabolism to be added to the current metabolite databases. With the

expansion of metabolic databases, and the verification of environmentally-derived spectra to

known standard compounds, metabolomic techniques will provide the ability to directly assess

the functionality of in situ metabolic pathways from natural systems.

Distribution of Glaciers and Rock Glaciers

Complete mapping of rock glaciers has only been completed for the contiguous United States

and portions of South America, yet early results suggest rock glaciers may be exceptionally more

abundant than ice glaciers in headwater ecosystems (Falaschi et al. 2013; Rangecroft et al. 2015;

Fegel et al. 2016). In addition to their ubiquity, rock glacier derived DOM may contribute to

ecosystem productivity for much longer than ice glaciers due to the slower recession of rock

glaciers compared to ice glaciers (Woo 2012). At similar carbon concentrations and with only

~15% less BDOM (~37% in rock glaciers compared to 52% in glacier derived DOM), as

observed in our study, rock glaciers may play a role in heterotrophic metabolism of alpine

headwaters. Current glacial carbon modeling neglects the contribution of rock glacial carbon

(Hood et al. 2015), yet our results suggest that rock glaciers supply a significant source of

organic matter for metabolism and will continue to do so for decades to come.

In conclusion, the work presented here suggests that as long as the cryosphere is present,

DOM entering alpine streams and lakes from both glacier and rock glaciers is likely to continue

to contribute to ecosystem heterotrophy during summer melt. We found clear differences in the

proportion of BDOM and microbial growth efficiency between glacier and rock glaciers DOM.

This suggests a higher potential for secondary production in systems with glacial headwaters

compared to those with rock glacier headwaters. The applicability of metabolomics to the

analysis of natural DOM provides the potential to expand the application of our results for the

prediction of BDOM beyond glaciated ecosystems to a broad range of inland waters. At the same

time our approach exposes the need for better metabolite database development for ecological

metabolomic approaches. In the coming decades, the DOM inputs to alpine lakes and streams

will increasingly be dominated by rock glacier-like DOM inputs, as ice glaciers are lost and rock

glaciers continue to contribute to alpine hydrology due to their slower ablation rates. Despite

differences in chemical compounds and BDOM pool sizes between glacier types, both glaciers

and rock glaciers represent an important heterotrophic subsidy of organic carbon to alpine

headwaters that will fuel ecosystems processes from the bottom up for years to come.

References

Amon, Rainer MW, and Ronald Benner. "Bacterial utilization of different size classes of

dissolved organic matter." *Limnology and Oceanography* 41.1 (1996): 41-51.

Andrilli, Juliana, et al. "An ultrahigh-resolution mass spectrometry index to estimate natural

organic matter lability." *Rapid Communications in Mass Spectrometry* 29.24 (2015): 2385-2401.

Benner, Ronald. "Chemical composition and reactivity." *Biogeochemistry of marine dissolved*

organic matter 3 (2002): 56-90.

Berggren, Martin, and Paul A. Giorgio. "Distinct patterns of microbial metabolism associated to

riverine dissolved organic carbon of different source and quality." *Journal of Geophysical*

Research: Biogeosciences (2015).

Berner, Robert A. *Early diagenesis: A theoretical approach*. No. 1. Princeton University Press,

1980.

Berthling, Ivar. "Beyond confusion: Rock glaciers as cryo-conditioned

landforms." *Geomorphology* 131.3 (2011): 98-106.

Borsheim, Knut Yngve, and Gunnar Bratbak. "Cell volume to cell carbon conversion factors for

a bacterivorous Monas sp. enriched from seawater." *Mar. Ecol. Prog. Ser* 36.17 (1987): ll.

Broeckling, Corey David, et al. "RAMClust: A Novel Feature Clustering Method Enables

Spectral-Matching-Based Annotation for Metabolomics Data."*Analytical chemistry* 86.14

(2014): 6812-6817.

Burga, Conradin A., et al. "Vegetation on Alpine rock glacier surfaces: a contribution to

abundance and dynamics on extreme plant habitats." *Flora-Morphology, Distribution,*

Functional Ecology of Plants 199.6 (2004): 505-515.

Cannone, Nicoletta, and Renato Gerdol. "Vegetation as an ecological indicator of surface

instability in rock glaciers." *Arctic, Antarctic, and Alpine Research* 35.3 (2003): 384-390.

Del Giorgio, Paul A., and Jonathan J. Cole. "Bacterial growth efficiency in natural aquatic

systems." *Annual Review of Ecology and Systematics* (1998): 503-541.

Derenne, Sylvie, and Thanh Thuy Nguyen Tu. "Characterizing the molecular structure of organic

matter from natural environments: An analytical challenge." *Comptes Rendus Geoscience* 346.3

(2014): 53-63.

Dittmar, Thorsten, et al. "A simple and efficient method for the solid-phase extraction of

dissolved organic matter (SPE-DOM) from seawater." *Limnology and Oceanography:*

Methods 6.6 (2008): 230-235.

Dubnick, Ashley, et al. "Characterization of dissolved organic matter (DOM) from glacial

environments using total fluorescence spectroscopy and parallel factor analysis." *Annals of*

Glaciology 51.56 (2010): 111-122.

Falaschi, Daniel, et al. "First glacier inventory and recent changes in glacier area in the Monte

San Lorenzo region (47 S), Southern Patagonian Andes, South America." *Arctic, Antarctic, and*

Alpine Research 45.1 (2013): 19-28.

Fegel, T. S., Baron, J. S., Fountain, A. G., Johnson, G. F., & Hall, E. K. (2016). The differing

biogeochemical and microbial signatures of glaciers and rock glaciers. *Journal of Geophysical*

Research: Biogeosciences.

Fellman, Jason B., Eran Hood, and Robert GM Spencer. "Fluorescence spectroscopy opens new

windows into dissolved organic matter dynamics in freshwater ecosystems: A

review." *Limnology and Oceanography* 55.6 (2010): 2452-2462.

Fellman, Jason B., et al. "The impact of glacier runoff on the biodegradability and biochemical

composition of terrigenous dissolved organic matter in near-shore marine ecosystems." *Marine*

Chemistry 121.1 (2010): 112-122.

Fellman, Jason B., et al. "Evidence for the assimilation of ancient glacier organic carbon in a

proglacial stream food web." *Limnology and Oceanography* (2015).

Guenet, Bertrand, et al. "Priming effect: bridging the gap between terrestrial and aquatic

ecology." *Ecology* 91.10 (2010): 2850-2861.

Guillemette, François, and Paul A. del Giorgio. "Reconstructing the various facets of dissolved

organic carbon bioavailability in freshwater ecosystems. *"Limnology and Oceanography* 56.2

(2011): 734-748.

Hedges, John I., et al. "The molecularly-uncharacterized component of nonliving organic matter

in natural environments." *Organic Geochemistry* 31.10 (2000): 945-958.

Hilker, Monika. "New synthesis: parallels between biodiversity and chemodiversity." *Journal of*

chemical ecology 40.3 (2014): 225-226.

Hobbie, J. El, R. Jasper Daley, and STTI977 Jasper. "Use of nuclepore filters for counting

bacteria by fluorescence microscopy." *Applied and environmental microbiology* 33.5 (1977):

1225-1228.

Hockaday, William C., et al. "Electrospray and photoionization mass spectrometry for the

characterization of organic matter in natural waters: a qualitative assessment." *Limnology and*

Oceanography: Methods 7.1 (2009): 81-95.

Hood, Eran, et al. "Glaciers as a source of ancient and labile organic matter to the marine

environment." *Nature* 462.7276 (2009): 1044-1047.

Hood, Eran, et al. "Storage and release of organic carbon from glaciers and ice sheets." *Nature Geoscience* 8.2 (2015): 91-96.

Jansson, Janet, et al. "Metabolomics reveals metabolic biomarkers of Crohn's disease." *PloS one* 4.7 (2009): e6386.

Kellerman, Anne M., et al. "Chemodiversity of dissolved organic matter in lakes driven by climate and hydrology." *Nature communications* 5 (2014).

Kellerman, Anne M., et al. "Persistence of dissolved organic matter in lakes related to its molecular characteristics." *Nature Geoscience* 8.6 (2015): 454-457.

Kim, Sunghwan, Louis A. Kaplan, and Patrick G. Hatcher. "Biodegradable dissolved organic matter in a temperate and a tropical stream determined from ultra-high resolution mass spectrometry." *Limnology and Oceanography* 51.2 (2006): 1054-1063.

Kujawinski, Elizabeth B. "The impact of microbial metabolism on marine dissolved organic matter." *Annual review of marine science* 3 (2011): 567-599.

Lafrenière, Melissa J., and Martin J. Sharp. "The concentration and fluorescence of dissolved organic carbon (DOC) in glacial and nonglacial catchments: interpreting hydrological flow routing and DOC sources." *Arctic, Antarctic, and Alpine Research* 36.2 (2004): 156-165.

Logue, Jürg B., et al. "Experimental insights into the importance of aquatic bacterial community

composition to the degradation of dissolved organic matter." *The ISME journal* (2015).

Pierce, Alan Eugene. "Silylation of organic compounds." (1968).

R Core Team (2014). R: A language and environment for statistical

computing. R Foundation for Statistical Computing, Vienna, Austria. URL http://www.R-

project.org/.

Rangecroft, S., S. Harrison, and K. Anderson. "Rock glaciers as water stores in the Bolivian

Andes: an assessment of their hydrological importance." *Arctic, Antarctic, and Alpine*

Research 47.1 (2015): 89-98.

Rovira, Pere, and V. Ramón Vallejo. "Labile and recalcitrant pools of carbon and nitrogen in

organic matter decomposing at different depths in soil: an acid hydrolysis

approach." *Geoderma* 107.1 (2002): 109-141.

Shannon, C. E. "A Mathematical Theory of Communication-An Integrated Approach." (1948).

Singer, Gabriel A., et al. "Biogeochemically diverse organic matter in Alpine glaciers and its

downstream fate." *Nature Geoscience* 5.10 (2012): 710-714.

Stibal, Marek, et al. "Organic matter content and quality in supraglacial debris across the ablation

zone of the Greenland ice sheet." *Annals of Glaciology*51.56 (2010): 1-8.

Stubbins, Aron, et al. "Anthropogenic aerosols as a source of ancient dissolved organic matter in

glaciers." *Nature Geoscience* 5.3 (2012): 198-201.

Sumner, Lloyd W., et al. "Proposed minimum reporting standards for chemical

analysis." *Metabolomics* 3.3 (2007): 211-221.

Tranter, Martyn, Mark Skidmore, and Jemma Wadham. "Hydrological controls on microbial

communities in subglacial environments." *Hydrological Processes* 19.4 (2005): 995-998.

Volk, Christian J., Catherine B. Volk, and Louis A. Kaplan. "Chemical composition of

biodegradable dissolved organic matter in streamwater."*Limnology and Oceanography* 42.1

(1997): 39-44.

Wahrhaftig, Clyde, and Allan Cox. "Rock glaciers in the Alaska Range." *Geological Society of*

America Bulletin 70.4 (1959): 383-436.

Wetzel, Robert G. "Gradient-dominated ecosystems: sources and regulatory functions of

dissolved organic matter in freshwater ecosystems." *Dissolved organic matter in lacustrine*

ecosystems. Springer Netherlands, 1992. 181-198.

Williams, M. W., et al. "Nitrate content and potential microbial signature of rock glacier outflow,

Colorado Front Range." *Earth Surface Processes and Landforms* 32.7 (2007): 1032-1047.

Woo, Ming-ko. *Permafrost hydrology*. Springer Science & Business Media, 2012.

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 Figure Legends

Figure 1 A) Stars note the approximate location of each glacier feature that was sampled for this

study within the state of Colorado. In total four pairs of glaciers and rock glaciers were sampled

along the Front Range of Colorado. **(B) Isabelle Glacier and Navajo Rock Glacier** form a pair

of a glacier and a rock glacier from the same watershed are shown here to illustrate the

differences between the two types of features.

Figure 2 PCA GC-MS Analysis DOM compounds **A)** PCA plots showing separation along PC5

between glaciers (blue) and rock glaciers (red) before incubation and **B)** no significant difference

between glacier types after incubation (p-value of 0.119 and 0.0681 for PC1 and PC2,

respectively). **C)** The Shannon-Wiener Index (SW) for chemical diversity was similar between

glaciers and rock glaciers before incubation, however microbial metabolism increased chemical

diversity in glacier DOM and decreased chemical diversity in rock glacier DOM.

Figure 3 Molecular distribution of GC-MS identified compounds A) Distribution of

compounds by ion intensity pre- (orange) and post-incubation (brown). Many of the compounds

present before incubation that were of intermediate abundance were metabolized. The most

abundant molecules were different between pre- and post- incubation metabolomics analysis (see

Table 4).

Figure 4 Dissolved oxygen consumption curves from glaciated watersheds (n=8). Values were

averaged for each of the glaciers (blue) and rock glaciers (red), and the analytical control (dotted

black). All curves were smoothed using a third order polynomial regression function (R^2=0.999).

Here we show 95% Confidence Intervals in light blue for glaciers and in pink for rock glaciers.

Glaciers consumed a larger portion of available DOM during the course of the incubation.

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 Tables

Table 1 Characteristics of DOM and microbial metabolism fed DOM from glaciers and rock

glaciers. C:N is the ratio of carbon to nitrogen for pre-incubation DOM. D.O (dissolved oxygen)

and DOC (dissolved organic carbon) represent the change in concentration of each during the

course of the incubation (mg L^{-1}). RQ = respiratory quotient, SW = Shannon Wiener Diversity

Index, and ΔSW = change is SW of DOM before and after the incubation. Bold values represent

significant differences between glaciers and rock glaciers ($p<0.05$), standard deviations of mean

values are listed in parentheses.

Site	Type	C:N	D.O.	DOC	RQ	SW	ΔSW
Isabelle	G	1.65	6.03	2.50	0.42	2.92	0.19
Peck	G	2.02	3.95	1.89	0.48	2.89	0.88
Andrews	G	3.00	4.11	2.64	0.64	2.68	0.13
Arapaho	G	2.73	3.97	2.84	0.72	2.84	0.52
Peck	RG	2.63	3.52	2.09	0.59	2.80	0.54
Navajo	RG	1.48	2.81	1.71	0.61	2.83	-0.32
Arapaho	RG	2.45	4.12	3.00	0.73	2.79	-1.10
Taylor	RG	0.86	2.69	2.25	0.83	2.80	-0.01
Mean	G	2.35 (0.62)	4.51 (1.02)	2.47 (0.41)	0.56 (0.14)	2.83 (0.11)	**0.43**
Mean	RG	1.85 (0.83)	3.29 (0.67)	2.26 (0.54)	0.69 (0.11)	2.80 (0.02)	-0.22

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777 **Table 2** DOM compounds that defined the differences between glaciers and rock glaciers before

778 incubation along PC5 from Figure 3A. Mass to charge ratios (m/z) ratios are the largest ion

779 observed for each compound. Normalized intensities (N.I.), an indicator of compound

780 concentration, are given for glaciers and rock glaciers. Bold values indicate which glacier type

781 had a higher normalized intensity (N.I) on average.

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DOM Peak	m/z	N.I. Glacier	N.I. Rock Glacier
Maltose	568.3	**4817**	2201
Glutamate	365.3	**5393**	2704
C996	94.2	**28959**	3589
C884	565.3	**14584**	1468
C815	419.1	**2486**	569
C1890	613.2	**805**	346
C1870	642	**1516**	1229
C1840	640.1	**384**	245
C1739	489.1	**2407**	1376
C1655	480.2	**17313**	9597
C1414	507.2	**550**	479
Threonic Acid	411	117	**2730**
Quinic Acid	409.2	2136	**7891**
Glycolic Acid	279.2	2920	**16696**
C589	649.5	1605	**1868**
C531	343.2	7446	**22845**
C35	348.2	1422	**2188**
C1942	535.1	846	**1062**
C1439	406.2	598	**861**
C1376	440.2	2623	**3610**
C1226	634.1	708	**1602**
C1125	468.1	3550	**7504**

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Table 3 Results from carbon decay model where B_O represents the size of the recalcitrant pool of

DOM (mg C L^{-1}), B_1 is the size of the bioavailable pool (mg C L^{-1}), k is the decay constant (mg

C L^{-1} h^{-1}), and BGE = bacterial growth efficiency. Glaciers had a larger percentage of BDOM

compared to rock glaciers and rock glaciers had a larger proportion of recalcitrant C compared to

glaciers. BGE was also higher for glaciers than rock glaciers. Bold values represent significant

differences between glaciers and rock glaciers (p<0.05) and standard deviations of mean values

are listed in parentheses.

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Site	Type	B_0	B_1	k	% BDOM	% Recalcitrant	BGE
Andrews	Glacier	3.20	5.59	0.001	48.23	51.77	0.37
Arapaho	Glacier	3.72	5.20	0.001	47.74	52.26	0.39
Isabelle	Glacier	1.35	7.03	0.001	60.90	39.10	0.39
Peck	Glacier	3.78	5.24	0.001	46.06	53.94	0.17
Arapaho	Rock Glacier	5.32	3.80	0.002	32.05	67.95	0.13
Navajo	Rock Glacier	5.64	3.42	0.001	29.89	70.11	0.03
Peck	Rock Glacier	3.31	5.24	0.001	45.74	54.26	0.14
Taylor	Rock Glacier	4.69	3.83	0.001	36.59	63.41	0.07
Mean G	Glacier	**3.38 (1.09)**	**6.35 (1.06)**	0.001 (0.000)	**52.83 (9.73)**	**47.17 (7.03)**	**0.263 (0.13)**
Mean RG	Rock Glacier	4.74 (1.26)	4.07 (1.12)	0.001 (0.000)	37.18 (10.23)	62.82 (6.84)	0.157 (0.16)

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799 **Table 4** The top 25 compounds present in pre- and post-incubation samples organized by

800 normalized intensity. Glaciers and rock glaciers had the same top 25 compounds before and after

801 incubation. Microbial metabolism altered compounds present, resulting in a different set of

802 compounds post incubation.

Pre-Incubation	Post-Incubation
Succinic Acid	Glycerol
Octyl 2-propylpentyl ester-phthalic acid	2,4,6-triethyl-, S-(2-phenylethyl) ester-benzenecarbothioic acid
Octyl 2,4,4-trimethylpentyl ester-terephthalic acid	Pyroglutamate
3-chlorophenyl 4-methoxybenzyl ester-succinic acid	10-dioxide-2-phenoxathiinamine-10
pentadecan-1-ol	Serine
Hydroxylamine	1-Ethoxy-2-propanol
Homosalate	Palmitic Acid
p-cyanophenyl 4'-heptyl-4-biphenylcarboxylate	Sucrose
Scopoletin	Beta-alanine
Diiodoacetylene-1	2,4,6-tri-tert-butyl-phenol
Palmitic Acid	Aspartate
Diiodoacetylene	Boric Acid
Phosphoric Acid	Stearic Acid
(2-amino-5-chlorophenyl)(2-fluorophenyl)-methanone	Norleucine
Stearic Acid	Homoserine
phenyl ester-2-thiophenecarboxylic acid	Glycine
phenylethylmalonamide	n-pentadecan-1-ol
3-Methyl-3-(N-methyl-2-pyrrolyl)-1,2-diphenylcyclopropene	D-pinitol
5-amino-1,2,3,8-tetramethoxy-deibenz[d,f]cycloheptane	N-methoxycarbonyl-,isohexyl ester-l-valine
Furan-2-carboxylic acid [2-(2,2,6,6-tetramethyl-piperidin-4-yl)-ethyl]-amide	3,5-Di-tert-butyl-4-hydroxybenzoic acid
1,3-Dimethyl-2,6-dioxo-2,3,6,7-tetrahydro-1H-purine-8-carbaldehyde-oxime	1-ethyl-1H-Benzimidazole
p-Cyanophenyl p-(2-butoxyethoxy)benzoate	Phosphoric acid
2-fluorophenyl 2,4-dimethylpent-3-yl ester-succinic acid	Tyrosine
Homoserine	Sarcosine
Pre-Incubation	Post-Incubation