

1 Parasite-driven replacement of a sexual by a closely related asexual taxon  
2 in nature

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20

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25 **Abstract**

26 Asexual species are thought to suffer more from coevolving parasites than related sexuals. Yet, this  
27 prediction may be modulated by the fact that closely related sexuals and asexuals often differ in  
28 respects other than reproductive mode. Here, we follow the frequency dynamics of sexual and  
29 asexual *Daphnia pulex* in a natural pond that was initially dominated by sexuals. However,  
30 coinciding with an epidemic of a microsporidian parasite infecting both sexuals and asexuals, the  
31 pond was rapidly taken over by the initially rare asexuals. We experimentally confirm that asexuals  
32 are less susceptible and also suffer less from the parasite once infected. These results show the  
33 ecological replacement of a sexual taxon by a closely related asexual taxon, as driven by parasites.  
34 We suggest that this replacement is, however, not directly connected with the reproductive mode, but  
35 rather due to the recent introduction and invasive nature of the asexuals studied.

36

## 37 **Introduction**

38 Theory on host-parasite interactions predicts that asexual species should suffer more from parasites  
39 compared to their sexual relatives. This is due to their lower genetic diversity and reduced ability to  
40 generate new allelic combinations, which are predicted to lead to a higher average parasite load  
41 (Jaenike 1978; Lloyd 1980; Hamilton 1980; Hamilton et al. 1990). This is thought to prevent  
42 asexuals from replacing sexuals within populations, because asexuals are thought to become  
43 increasingly parasitized as they become more common. The prediction that asexuals should be  
44 overparasitized has been investigated empirically in many study systems, with some studies finding  
45 evidence in support of the prediction (Lively & Jokela 2002; Kumpulainen et al. 2004; Vergara et  
46 al. 2014) and others not (Hanley et al. 1995; Ben-Ami & Heller 2005; Elzinga et al. 2012).

47

48 While such variable results are not necessarily inconsistent with theory (asexuals are only predicted  
49 to be overparasitized on average, and not in every single case), they may also be due to differences  
50 between sexuals and asexuals, other than the reproductive mode. In fact, asexual taxa may only  
51 rarely differ from their closely related sexual relatives in reproductive mode alone (e.g. Lehto &  
52 Haag 2010; Gilabert et al. 2014; Kotusz et al. 2014). This is because the evolution of a new  
53 reproductive mode is often followed by secondary adaptations and differentiation, involving traits  
54 unrelated to reproductive mode (Meirmans et al. 2012; Gilabert et al. 2014). Furthermore, asexuals  
55 may have, from the outset, different genetic backgrounds, as they often arise via hybridization or  
56 polyploidy (Simon et al. 2003). These differences may add strong variation to the predictions of who  
57 should be more parasitized, and, in some cases, perhaps even lead to the inverse prediction.

58

59 One such case where asexuals might suffer less from parasites than sexuals is in the context of  
60 biological invasions. Asexuals are thought to be particularly prone to become invasive (Vandel 1928;  
61 Baker 1955; Peck et al. 1998; Haag & Ebert 2004), and indeed there is empirical data showing that

62 asexuals are overrepresented among invasive species (Ellstrand & Schierenbeck 2000; Sakai *et al.*  
63 2001). Additionally, invasive species do not share evolutionary history with the native parasite  
64 species in their invasive ranges. This can provide a parasite-mediated competitive advantage for  
65 invaders over residents, as local adaptation of parasites to resident hosts is thought to be common  
66 (Torchin *et al.* 2003). Hence, considering that asexuals are more common among invasive species  
67 and that parasites are likely more adapted to residents than to invaders, sexuals rather than asexuals  
68 may be predicted to suffer more from parasites during the early phases of invasion.

69

70 Here, we document the rapid replacement of a sexual by a closely related asexual taxon in nature and  
71 show experimentally that this replacement was very likely driven by parasites. We discuss the results  
72 with respect to sex-asex predictions, as well as with respect to invasion biology (the asexual taxon is  
73 invasive). The aim of this study is not to pinpoint the ultimate driver behind parasite-driven  
74 competition in sex-asex systems, but rather to showcase a specific empirical example where  
75 asexuals, rather than sexuals, enjoy a parasite-mediated advantage. We then discuss the implications  
76 of these findings for other empirical sex-asex comparisons.

77

78 Specifically, we studied cyclical and obligate parthenogenetic lineages of the freshwater cladoceran  
79 *Daphnia pulex* that occur in small rock pools to form dynamic metapopulations across the skerry  
80 archipelago of Southern Finland (Pajunen 1986; Pajunen & Pajunen 2003; Lehto & Haag 2010). In  
81 winter the ponds freeze to the bottom and only diapause stages survive. In our study area, about 5 to  
82 10 % of the rock pools harbour *D. pulex*. Cyclical and obligate parthenogenetic *D. pulex* often  
83 inhabit different ponds, in part due to chance events caused by the dynamics of extinction and  
84 recolonization and in part due to different niche preferences with respect to water chemistry. In  
85 ponds with intermediate water chemistry, cyclical and obligate parthenogenetic populations coexist  
86 (Lehto & Haag 2010).

87

88 Obligate asexual lineages belong to the North American clade of *D. pulex* and have a history of  
89 introgression and contagious asexuality with cyclical parthenogens of the same clade, as well as the  
90 closely related *Daphnia pulicaria* (Innes & Hebert 1988; Tucker *et al.* 2013; Xu *et al.* 2015). It is  
91 unknown exactly when the asexuals were introduced to Europe, but it was probably near the  
92 beginning of the 20<sup>th</sup> century in the ballast water of transport ships returning from the Laurentian  
93 Great Lakes, a transport route common to many aquatic invasive species (Carlton 1985; Williams *et*  
94 *al.* 1988). The cyclical parthenogenetic *D. pulex* in our study system belong to the European clade,  
95 which, however, may represent a different species (Mergeay *et al.* 2008; Markova *et al.* 2013). In the  
96 absence of a formal description of the two taxa as different species, we here still use the species  
97 name *D. pulex* for both, but refer to them as closely related taxa. Obligate asexuals (hereafter  
98 “asexuals”) produce both live-born offspring and diapausing stages asexually, whereas cyclical  
99 parthenogens (hereafter “sexuals”) have several asexual generations (production of live-born  
100 offspring) and typically one sexual generation (production of diapause stages) per year (Innes &  
101 Hebert 1988; Paland *et al.* 2005; Lynch *et al.* 2008; Heier & Dudycha 2009). Asexuals are  
102 genetically highly uniform, whereas sexuals show moderate levels of genetic variation, typical of  
103 other cyclical parthenogenetic *Daphnia* species (Ward *et al.* 1994; Haag *et al.* 2005; Walser & Haag  
104 2012).

105

106 Here we followed the frequencies of sexuals and asexuals in one such mixed natural pond over 14  
107 years (1999-2013). In 2006, a strong epidemic caused by a highly virulent parasite, *Gurleya vavrai*,  
108 coincided with the total replacement of formerly dominating sexuals by obligate asexuals. *G. vavrai*  
109 is an endemic microsporidian parasite in Europe (Green 1974; Refardt *et al.* 2002), with no record of  
110 the parasite occurring in North America. Infections are localized in the epidermis and cause a  
111 progressive whitening of the body as the infection spreads, with high virulence (Friedrich *et al.* 1996;

112 Stirnadel & Ebert 1997; Little & Ebert 1999). Using a series of field and laboratory experiments, we  
113 assessed the relative competitiveness of sexuals and asexuals in the presence and absence of the  
114 parasite, the susceptibility of sexual and asexuals hosts, as well as life history traits of infected and  
115 uninfected individuals. The goal of these experiments was to investigate whether the reversal in  
116 relative abundance observed in the natural pond was likely caused by parasites.

117

## 118 **Materials and Methods**

### 119 **Monitoring of the natural pond**

120 One pond, SK-39 (59°49'55.7"N 23°15'17.6"E, pond surface 1.6 m<sup>2</sup>, depth 0.3 m) on the island of  
121 Skallothomen, which contained both sexuals and asexuals in 1999, was followed for the present  
122 study because it was initially uninfected by *G. vavrai*, but became infected in 2006. In 2000 and  
123 again in 2008, no animals were observed in the pond during routine visits, possibly because  
124 appropriate conditions for the hatching of diapause stages did not occur in those years. In all other  
125 years, namely 1999 to 2015, *D. pulex* were present. We monitored the frequencies of sexuals and  
126 asexuals in a total of 26 samples taken between 1999 and 2013 (total  $N = 1848$  individuals, for exact  
127 sampling dates and sample sizes per date, see data deposited on the Dryad digital repository). The  
128 breeding type (sexual, asexual) of each individual in these samples was assessed using cellulose-  
129 acetate electrophoresis (Hebert & Beaton 1993) at the PGI locus (phosphoglucose isomerase, EC  
130 5.3.1.9.), for which the sexuals and asexuals have consistently different genotypes (Lehto & Haag  
131 2010). On July 14<sup>th</sup> 2006, we noted a strong *G. vavrai* infection in our samples. Even though  
132 previous samples were not systematically checked for this parasite, a previous epidemic would not  
133 have gone unnoticed. We subsequently took additional samples to estimate the parasite prevalence in  
134 sexuals and asexuals (20, 25, and 30 July, 29 September). Prevalence was assessed by visually  
135 inspecting the carapace color with light from the top and against a dark background. Checking for a  
136 white carapace detects only late stage infections, while early stage infections are pooled with the

137 uninfected animals. Thus, all prevalence estimates are underestimates. From 2007-2013 we recorded  
138 only presence/absence of infections.

139

140 ***Outdoor experiment: Origin and handling of hosts and parasites***

141 Ten ponds composed purely of sexuals and ten ponds of purely asexuals were sampled in the  
142 Tvärminne area in May 2010 (Table 1). From each of these ponds, we randomly sampled 100  
143 individuals to capture a representation of the genetic variation present. These 20 geographically  
144 separate ponds should be representative of ponds across the study area: Hence the experiment was  
145 designed to test whether an average, uninfected asexual in this metapopulation suffers less from  
146 infection (and/or is less susceptible) than an average, uninfected sexual in this metapopulation. From  
147 each of these 20 populations, an outdoor culture was established in buckets, and left outdoors under  
148 ambient conditions on the island of Furuskär (59°49'58"N 23°15'49"E). Before introducing the  
149 *Daphnia*, each bucket was filled with 40L of 0.02 mm-double filtered water from a nearby pond, not  
150 colonized by *Daphnia*. The 100 individuals were then left for two months to reproduce and increase  
151 in number. The only exception was a bucket of asexuals from SK-39, which had already been started  
152 using a single, uninfected individual in 2007. The breeding type (sexual, asexual) of each bucket was  
153 confirmed using the PGI locus as described above. In July of 2010 these cultures were harvested and  
154 used as the experimental animals for the outdoor experiment. Individuals were checked for parasite  
155 infection to ensure the absence of any infected individuals at the start of the experiment. Four  
156 additional cultures of *D. pulex*, which had been maintained in buckets on the island of Furuskär from  
157 2007 to 2010 with *G. vavrai* infections, were used as the source material for the spore cocktail  
158 prepared for infections. In two of these cultures, *G. vavrai* was grown on sexuals, and in the two  
159 others on asexuals (Table S1). The two cultures per host taxon were combined to produce the spore  
160 cocktails for infection (one cocktail derived from sexual hosts, one from asexual hosts).

161

162 ***Outdoor experiment: Experimental set-up***

163 We released 100 sexuals and 100 asexuals, obtained as described above, into each of 20 new outdoor  
164 cultures (40L bucket cultures). This resulted in ten experimental pairings (each pair consisting of  
165 sexuals from one sexual pond and asexuals from one asexual pond, Table S1), with each pair  
166 repeated twice. One replicate per pair was exposed to *G. vavrai* spores and the second served as the  
167 control culture without parasite exposure. *D. pulex* infected with *G. vavrai* were ground up  
168 (individuals from all four cultures pooled) and distributed equally across the ten infection replicates.  
169 In the control treatment, a comparable amount of ground-up, uninfected *D. pulex* were added as a  
170 placebo.

171

172 ***Outdoor experiment: Recorded parameters***

173 The cultures were left in the field, and samples were obtained from each culture at three time points:  
174 August 2010, September 2010 and October 2010. At each sampling event, 50 individuals were  
175 removed per culture and checked visually for infection. Subsequently, the same 50 individuals were  
176 typed as sexual or asexual using the PGI locus. In this way the prevalence of infection in sexuals  
177 versus asexuals could be monitored over the course of the experiment. Note that we were mainly  
178 interested in whether the frequency changes differed between treatments. The overall frequency  
179 changes are difficult to interpret because they are influenced by water chemistry (Lehto & Haag  
180 2010). Note also that these estimates do not include potential differences in the production of  
181 diapause stages.

182

183 ***Laboratory experiment: Origin and handling of hosts and parasites***

184 In June 2013, *D. pulex* were collected from various ponds in the Tvärminne area. Both sexuals and  
185 asexuals were collected, as well as *G. vavrai*-infected individuals of both taxa (Table S1). All  
186 individuals were transported live to the laboratory in Switzerland for immediate use. In the



187 laboratory, one clonal line (lines started by single females and maintained exclusively by clonal  
188 reproduction in both the sexuals and asexuals) was established from each of the pond samples. Each  
189 line was propagated in a 400 ml glass jar filled with a water medium designed for *Daphnia* culturing  
190 (ADaM; Klüttgen *et al.* 1994) and fed with unicellular green algae (*Scenedesmus obliquitus*) *ad*  
191 *libitum* under a summer photo-period of 16:8 light:dark and a temperature of 20 °C. The lines were  
192 maintained in this way for two weeks, and their breeding type was confirmed using the PGI locus.

193

194 Two separate *G. vavrai* cultures were established in the laboratory: one obtained from and grown on  
195 sexuals and a second obtained from and grown on asexuals. These parasite types are referred to as  
196 sexual and asexual *G. vavrai* isolates, respectively, referring to their host type. In both cases the *D.*  
197 *pulex* hosts on which the spores were grown were those on which the parasite was collected, which  
198 were different clones from those used in the experiments (Table S1).

199

#### 200 ***Laboratory experiment: Experimental procedures***

201 For the infection experiment, we used eight clonal lines (four sexual lines and four asexual lines)  
202 with 30 replicate individuals per line and each line originating from a different pond (Fig. S1; Table  
203 S1). We started by isolating 50 individuals from each of the clonal lines (each placed individually in  
204 a 50 ml falcon tube) to ensure we would have at least 30 individuals at the beginning of the  
205 experiment. The animals were fed daily with 2,500 cells of the algae *S. obliquitus* and were kept in a  
206 climate chamber with a photo-period of 16:8 light:dark and a temperature of 20 °C. The ADaM  
207 culture medium was changed three times per week (Monday, Wednesday and Friday). The animals  
208 were passed through three generations under these pre-experimental conditions in order to remove  
209 maternal effects. Day zero of the experiment was when the fourth-generation offspring (third-clutch  
210 offspring of the third-generation females) were isolated into new tubes.

211

212 For infection, two separate spore cocktails were prepared: one from ground-up sexual hosts infected  
213 with *G. vavrai* and another from ground up infected asexual hosts infected with *G. vavrai*. Infected  
214 hosts came from two separate ponds for both the sexual and asexual cultures (Table S1). The total  
215 number of spores was equalized between the two treatments (spore numbers were estimated using a  
216 Neubauer improved counting chamber) and then distributed equally over the respective tubes. This  
217 resulted in approximately 60,000 *G. vavrai* spores being added to each tube in the parasite  
218 treatments. For the control treatment, a cocktail of ground up uninfected *D. pulex* was added to the  
219 tubes.

220

221 Offspring from 30 mothers per clone (and eight clones: four sexual, four asexual) were used as the  
222 experimental animals, resulting in a cohort of 240 animals at the start of the experiment. Each clonal  
223 line then received three different treatments, with ten replicates per treatment: (1) 60,000 spores of  
224 *G. vavrai* from asexual hosts, (2) 60,000 spores of *G. vavrai* from sexual hosts, (3) ground-up,  
225 uninfected *D. pulex* (control).

226

### 227 ***Laboratory experiment: Recorded parameters***

228 Recorded parameters from the experiment were age at first reproduction, reproductive output (total  
229 number of offspring recorded across all changing events), age at death, as well as infection status and  
230 number of spores at death. To determine the number of spores at death, each individual was  
231 homogenized in 0.3 ml of medium and the concentration of spores was determined using a Neubauer  
232 Improved counting chamber.

233

### 234 **Data analysis**

235 All statistical analyses were performed using the software R (R Core team 2013). For the outdoor  
236 competition experiment, we used a generalized linear-mixed model with a binomial error distribution

237 to look at differences in the frequency of the sexuals vs. asexuals (R program lme4). For data  
238 analyses, each sexual individual was coded as 1, and each asexual individual as 0. We used treatment  
239 (control, parasite-exposed) and sampling time point as fixed factors, whereas replicate culture and  
240 pond pair were treated as random factors, with replicate culture nested within treatment:  
241  $\text{breedingtype} \sim \text{treatment} * \text{time} + (1 | \text{treatment/replicate}) + (1 | \text{pair of ponds})$ , family = binomial. We  
242 used a similar generalized linear mixed model to investigate differences in parasite prevalence  
243 (proportion of individuals infected, each individual being counted as either infected or uninfected)  
244 between the sexual and asexual cultures, but here we used the parasite-exposed treatment only:  
245  $\text{infection\_status} \sim \text{breedingtype} * \text{time} + (1 | \text{replicate})$ , family = binomial. Before running the analysis,  
246 we verified model assumptions and absence of over-dispersion by calculating the sum of the squares  
247 of the Pearson residuals and comparing them with the residual degrees of freedom using a chi-  
248 squared test.

249

250 Data on age at death, reproductive output and age at first reproduction from the laboratory  
251 experiment were evaluated using linear mixed models in R, with the nlme package. Breeding type  
252 (asexual or asexual), and treatment (control, infected) were set as fixed factors and clone as a random  
253 factor:  $\text{trait} \sim \text{breedingtype} * \text{treatment}$ ,  $\text{random} = (1 | \text{breedingtype/clone})$ . Only those individuals that  
254 successfully became infected were used for comparison between treatments. We then used the same  
255 analysis exclusively on the infection treatment (and those individuals that actually became infected),  
256 but now used spore origin instead of treatment to test whether the different life history traits were  
257 affected by whether the *G. vavrai* spores were derived from sexual or asexual host clones. Using the  
258 same data, we also tested for differences in spore load at death using the same model as for the life  
259 history traits, as well as for differences in parasite prevalence (number of exposed individuals that  
260 became infected). For the latter, we used a generalized linear mixed model, with infection success as  
261 the response variable (1 = successfully infected, 0 = infection failed) and a binomial error

262 distribution. Breeding type and spore type (sexual or asexual host origin) were held as fixed factors,  
263 whereas clone was a random factor nested within breeding system: infection status ~  
264 breedingtype\*sporetype, random= ~ 1|breedingtype/clone, family = binomial. As above, we verified  
265 model assumption and absence of over-dispersion before running the analyses.

266

## 267 **Results**

### 268 *Dynamics in the natural pond*

269 There were strong seasonal dynamics in the relative frequencies of sexuals and asexuals (Fig. 1).  
270 Asexuals were more frequent early in the season, and sexuals dominated in late season. Yet, across  
271 years, the starting frequencies of sexuals increased (reaching 93 % on June 1<sup>st</sup> 2006), as the pond  
272 became more and more dominated by sexuals. By the end of June 2006, the frequency of sexuals had  
273 increased to 98 %. Then, on July 14<sup>th</sup> 2006 we detected many individuals (sexuals and asexuals not  
274 distinguished) with *Gurleya vavrai* infections, and at the same time a dramatic decrease in the  
275 frequency of sexuals to 41 %. This decline continued over the course of the growing season down to  
276 0 % (9 % on July 20<sup>th</sup>, 4 % on July 25<sup>th</sup>, 2 % on July 30<sup>th</sup>, and 0 % on September 28<sup>th</sup>). In the  
277 following years (two samples in 2007, and one in each of 2010, 2012, and 2013), we found only  
278 asexuals (Fig. 1).

279

280 No quantitative measure of prevalence was recorded on July 14<sup>th</sup> 2006, but we noted many infected  
281 individuals. On July 20<sup>th</sup>, many dead, infected individuals were found in the pond, the breeding type  
282 of which could not be assessed. Among those living, we found 10 infected individuals in a sample of  
283 54 (prevalence = 19 %). Of the 10 infected individuals, 5 were sexuals and 5 were asexuals, whereas  
284 all of the uninfected individuals were asexuals (Fig. 1; Fisher's exact test,  $P < 0.0001$ ). Prevalence  
285 then rapidly dropped to low levels (2 % on July 25<sup>th</sup>, 4 % on July 30<sup>th</sup>, and 7 % on September 28<sup>th</sup>).  
286 The higher prevalence in sexuals was still significant on July 25<sup>th</sup> (infected: 3 sexuals, 5 asexuals,

287 uninfected: no sexuals, 72 asexuals,  $P = 0.0007$ ), but not anymore on July 30<sup>th</sup>, when all 13 infected  
288 individuals were asexuals, whereas 2 among 74 uninfected individuals were sexuals ( $P > 0.5$ ). The  
289 latter result shows that not all sexuals became infected and died during the epidemic. From 2007-  
290 2013, when only asexuals were left in the pond, we recorded infected individuals in all samples,  
291 though the infections were never as clearly abundant as during the beginning of the epidemic.

292

### 293 ***Outdoor experiment***

294 Asexuals performed relatively better in the competition experiment in the presence of the parasite *G.*  
295 *vavrai* than in their absence (Fig. 2). The infected cultures had higher frequencies of asexuals than  
296 the control cultures ( $z = -3.04$ ,  $df = 2$ ,  $P = 0.002$ ). This was true at both sampling time points and thus  
297 sampling time was not a significant factor in the linear mixed model ( $z = 0.60$ ,  $df = 2$ ,  $P = 0.551$ ). In  
298 addition, in the parasite treatment, a greater number of sexuals became infected, as opposed to  
299 asexuals ( $z = 2.21$ ,  $df = 2$ ,  $P = 0.038$ ; Fig. 2). Again sampling time point was not significant ( $z = 0.41$ ,  
300  $df = 2$ ,  $P = 0.389$ ).

301

### 302 ***Laboratory experiment***

303 In line with the results from the outdoor experiment, 76.3 % of the sexual individuals exposed to *G.*  
304 *vavrai* spores became infected, whereas only 25.0 % of the asexuals did so ( $z = 4.96$ ,  $df = 2$ ,  $P <$   
305  $0.001$ ; Fig. 3a). The comparison of the infected individuals (spore origins not distinguished) with the  
306 uninfected controls revealed clear negative fitness effects of infection: infected individuals died  
307 sooner ( $t = -5.14$ ,  $df = 149$ ,  $P < 0.001$ ) and were less fecund ( $t = -5.90$ ,  $df = 149$ ,  $P < 0.001$ ), while  
308 there was no significant effect for age at maturity ( $t = 1.47$ ,  $df = 149$ ,  $P = 0.143$ ; Fig. 4). In no case  
309 was breeding system a significant factor on its own, suggesting that there was no clear main  
310 difference in life-history traits between sexuals and asexuals under the experimental conditions (age  
311 at death:  $t = 0.52$ ,  $df = 6$ ,  $P = 0.620$ ; age at maturity:  $t = 0.04$ ,  $df = 6$ ,  $P = 0.972$ ; reproductive output:

312  $t = 1.60$ ,  $df = 6$ ,  $P = 0.163$ ). However, there was a significant interaction between breeding type and  
313 infection for reproductive output ( $t = -3.10$ ,  $df = 149$ ,  $P = 0.002$ ), suggesting that infection reduced  
314 the reproductive output of sexuals more strongly than that of asexuals, whereas asexuals had a  
315 somewhat lower reproductive output in the controls. The interaction was non-significant for age at  
316 death ( $t = -1.02$ ,  $df = 149$ ,  $P = 0.307$ ) and for age at maturity ( $t = -0.51$ ,  $df = 149$ ,  $P = 0.614$ ).

317

318 Looking only at the infected individuals, there was no significant difference in spore load at death  
319 between sexual and asexual *Daphnia* ( $t = -0.60$ ,  $df = 71$ ,  $P = 0.553$ ; Fig. 3b). Furthermore, the spore  
320 origin did not appear to affect infection success or spore load, either overall or differently between  
321 sexuals and asexuals (non-significant main effects: infection success:  $z = 1.03$ ,  $df = 2$ ,  $P = 0.304$ ;  
322 spore load at death:  $t = 0.20$ ,  $df = 150$ ,  $P = 0.845$ ; and non-significant interactions with breeding  
323 type: infection success:  $z = -1.28$ ,  $df = 2$ ,  $P = 0.200$ ; spore load at death:  $t = 0.64$ ,  $df = 71$ ,  $P = 0.522$ ).  
324 However, spore origin affected the reproductive output. Spores obtained from asexual hosts reduced  
325 the reproductive output of asexuals more than spores obtained from sexual hosts, whereas no such  
326 effect was observed in the sexuals (breeding type x spore type interaction:  $t = -2.85$ ,  $df = 71$ ,  $P =$   
327  $0.006$ ). Finally, spore origin did not affect age at death, or age at maturity (breedingtype\*sporetype  
328 interaction; age at death:  $t = -1.51$ ,  $df = 71$ ,  $P = 0.135$ ; age at maturity:  $t = -0.25$ ,  $df = 71$ ,  $P = 0.805$ ;  
329 Fig. 4).

330

### 331 **Discussion**

332 This study documents the replacement of an initially dominating sexual taxon by an initially rare  
333 asexual taxon in nature. This replacement did not occur gradually, as would be expected if the  
334 asexuals had an overall higher fitness. Rather, the replacement happened rapidly and was tightly  
335 linked with an epidemic caused by a virulent parasite, which infected both sexuals and asexuals.  
336 Subsequent field and laboratory experiments strongly support a causal role of the parasite in this

337 replacement: Asexuals were less susceptible to infection, suffered less from infection than sexuals,  
338 and their relative performance in a competition experiment was enhanced in the presence of  
339 parasites. Independent of the reproductive mode of the two competitors, this is an example showing  
340 that parasites can strongly alter interactions between closely related taxa. Parasites have been  
341 implicated in ecological replacements between closely related species several times before, including  
342 in the replacement of residents by invaders (Thomas *et al.* 2005; Hatcher *et al.* 2006; Hatcher *et al.*  
343 2008). Though, to our knowledge, in none of these previous cases could the replacement be  
344 monitored so closely in nature, or coupled with experimental evidence for a causal role of the  
345 parasite.

346

347 Our results also show that parasites can lead to rapid changes in the ecological frequency dynamics  
348 of coexisting sexual and asexual taxa. Discussion on competition between sexuals and asexuals with  
349 regards to parasites is often framed within Red Queen dynamics, whereby asexuals are predicted to  
350 become overinfected with parasites (at least on average), as they cannot evolve as fast as their sexual  
351 competitors (Jaenike 1978; Lloyd 1980; Hamilton 1980; Hamilton *et al.* 1990). In contrast, here we  
352 find that asexuals replace sexuals (and not the other way around) as a consequence of a parasite-  
353 driven advantage.

354

355 While our data are not suitable to pinpoint the ultimate cause behind this parasite-mediated  
356 advantage of asexuals, it seems likely that factors other than reproductive mode have played an  
357 important role. First, due to the invasive nature of the asexuals in this region, they do not share a long  
358 evolutionary history with the parasite (Innes & Hebert 1988; Ward *et al.* 1994). The parasite is  
359 endemic to Europe (Green 1974; Friedrich *et al.* 1996) and may thus not have had enough time to  
360 adapt to the asexuals. Second (and perhaps also due to their recent invasion of the region), asexuals  
361 are relatively rare in Southern Finland compared to sexuals (Lehto & Haag 2010). This is important

362 because theory on host-parasite coevolution predicts that parasites should specialize on the most  
363 common genotypes (Jaenike 1978; Decaestecker *et al.* 2007; Salathé *et al.* 2008). However, the  
364 advantage of asexuals in the presence of parasites may only be transitory: once asexuals become  
365 abundant, parasites are predicted to adapt to them (Morran *et al.* 2011). Indeed, in our laboratory  
366 experiment, the spores obtained from asexual hosts were more virulent to asexuals than spores from  
367 sexual hosts. While this test was not replicated (we only tested one mixture of two parasite isolates  
368 from each of the two host types), this result suggests that some effects of parasite adaptation towards  
369 asexual hosts may have started to become visible. Finally, we cannot entirely exclude that the higher  
370 fitness of the asexuals in the presence of the parasite is due to the adaptation of asexuals to the  
371 parasite (rather than of the parasite to sexual hosts). However, this seems less plausible given the  
372 likely lack of a shared long-term evolutionary history.

373

374 As outlined in the introduction, it is also possible that our results are indirectly linked to reproductive  
375 mode. Asexuals are overrepresented among invasive species (Ellstrand & Schierenbeck 2000; Sakai  
376 *et al.* 2001), and obligate asexual *Daphnia* have been shown to be effective invaders who have  
377 rapidly replaced resident populations of closely related sexuals in other parts of the world, not only  
378 in Finland (Mergeay *et al.* 2006; Fadda *et al.* 2011; Duggan *et al.* 2012; So *et al.* 2015). Even though  
379 reproductive assurance (Baker 1955) does not differ between sexual and asexual *Daphnia* (due to  
380 cyclical parthenogenesis, also a single sexual individual can establish a population on its own), the  
381 success of obligate asexual *Daphnia* as invaders may still be linked to reproductive mode (though  
382 sexual *Daphnia* species can be invasive too; see Searle *et al.* 2016). First, obligate asexuality may  
383 allow a particularly successful invasive genotype to be “frozen” and thus shielded from segregation  
384 and mixing with other genotypes. In fact, what is essentially a single clone of *D. pulex* (or a group of  
385 closely related clones) is apparently responsible for the invasion of freshwater habitats in Africa and  
386 Southern Europe (Mergeay *et al.* 2006; Fadda *et al.* 2011). Second, colonization by a single cyclical



387 parthenogenetic individual leads to within-clone mating during the production of diapause stages,  
388 and within-clone mating is known to lead to strong inbreeding depression in this and other *Daphnia*  
389 species (Deng & Lynch 1997; Lohr & Haag 2015), also specifically with regards to parasites (Haag  
390 et al. 2003).

391

392 Thus, our results suggest that parasites may have played an important role in the rapid establishment  
393 of these obligate asexual invaders, perhaps not only in Southern Finland but also elsewhere. In stark  
394 contrast, in the native range of asexual *D. pulex*, they are mostly outcompeted by sexuals if the two  
395 co-occur locally (Innes & Gin 2014). Though the possible involvement of parasites in the latter  
396 pattern is not known, this is potentially consistent with the “enemy release” advantage for invasive  
397 species, whereby invaders leave behind natural enemies from their native range and suffer less from  
398 newly encountered enemies in their introduced range (Keane & Crawley 2002; Torchin & Mitchell  
399 2004).

400

401 Our study shows that asexuals do not always suffer more from parasites than sexuals. On the  
402 contrary, under certain circumstances, such as in the invasion context studied here, asexual can  
403 benefit from parasitism. Thus, the ecological context can modulate the generally predicted patterns  
404 of parasitism in sexual versus asexual species. This highlights an important limitation in the  
405 interpretation of empirical comparisons between closely related sexual and asexual species. These  
406 types of sex-asex comparisons are often used to investigate hypotheses related to the maintenance of  
407 sexual reproduction (Otto & Lenormand 2002), and while such comparisons can offer key insights,  
408 sexuals and asexuals often differ in many traits other than just the reproductive mode (e.g. Lehto &  
409 Haag 2010; Gilibert *et al.* 2014; Kotusz *et al.* 2014). This study highlights that it should not be  
410 assumed that patterns found between sexual and asexual species are driven solely by the reproductive  
411 mode.

412

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421

422

423 **Tables – supplementary material**

424

425 Table S1: Host and parasite origins for the outdoor experiment (experiment 1) and the laboratory

426 experiment (experiment 2). Ponds used both for *G. vavrai* spores in experiment 2 and for *D. pulex* in

427 experiment 1 were previously uninfected (an uninfected *Daphnia* sample was obtained from these

428 ponds for experiment 1).

Pond	Coordinates	Breeding type	<i>D. pulex</i> used in	<i>G. vavrai</i> spores used in
AB2-1	59°49'51.6"N 23°13'57.0"E	Sexual	Experiment 1, Pair 1	
ALB-7	59°49'53.1"N 23°14'02.3"E	Sexual	Experiment 1, Pair 2	
ALO-1	59°50'06.3"N 23°15'58.5"E	Sexual	Experiment 1, Pair 3	Experiment 2
F-4	59°49'46.6"N 23°14'44.1"E	Sexual		Experiment 2
KV1-1	59°50'59.8"N 23°15'15.9"E	Sexual	Experiment 1, Pair 4	
LAG-3	59°49'39.6"N 23°11'41.3"E	Sexual	Experiment 1, Pair 5	
LH-1	59°51'02.7"N 23°15'00.2"E	Sexual		Experiment 2
M-1	59°49'26.7"N 23°14'50.8"E	Sexual		Experiment 1
M-60	59°49'19.2"N 23°14'57.5"E	Sexual	Experiment 1, Pair 6	Experiment 2
N-69	59°49'19.6"N 23°15'36.8"E	Sexual	Experiment 1, Pair 7	
N-74	59°49'19.2"N 23°15'35.9"E	Sexual		Experiment 2
RO1-3	59°50'17.4"N 23°15'07.5"E	Sexual	Experiment 1, Pair 8	
RO1-6	59°50'14.1"N 23°15'04.7"E	Sexual		Experiment 1
SK-47	59°49'57.0"N 23°15'19.7"E	Sexual		Experiment 2
VIN-4	59°49'53.5"N 23°12'26.5"E	Sexual	Experiment 1, Pair 9	
VIW-3	59°49'49.8"N 23°12'01.1"E	Sexual	Experiment 1, Pair 10	
FU1-57	59°49'59.2"N 23°16'08.5"E	Asexual	Experiment 1, Pair 1	
FU2-83	59°50'10.6"N 23°16'34.8"E	Asexual	Experiment 1, Pair 2	
M-69	59°49'10.8"N 23°14'49.6"E	Asexual	Experiment 1, Pair 3	Experiment 2
N-80	59°49'20.3"N 23°15'35.9"E	Asexual		Experiment 1*
SK-30	59°49'52.1"N 23°15'19.6"E	Asexual	Experiment 1, Pair 4	
SK-39	59°49'55.7"N 23°15'17.6"E	Asexual	Experiment 1, Pair 5	Experiment 1 Experiment 2
SK-42	59°49'55.8"N 23°15'18.4"E	Asexual		Experiment 2
SK-44	59°49'55.9"N 23°15'18.8"E	Asexual	Experiment 1, Pair 6	
SK-45	59°49'55.9"N 23°15'18.9"E	Asexual	Experiment 1, Pair 7	Experiment 2
SYN-3	59°52'23.2"N 23°14'45.6"E	Asexual	Experiment 1, Pair 8	Experiment 2
SYN-4	59°52'23.1"N 23°14'45.7"E	Asexual	Experiment 1, Pair 9	
SYN-5	59°52'23.8"N 23°14'45.0"E	Asexual		Experiment 2
SYN-6	59°52'24.0"N 23°14'44.0"E	Asexual	Experiment 1, Pair 10	

429

430 The \* indicates that *G. vavrai* was cultured on *D. pulex* from this pond prior to the experiment, but

431 the spores were originally obtained from the pond SK-39.

432 **Figure legends**

433 Figure 1. Frequency of sexuals in pond SK-39. Samples taken in the same year are connected with a  
434 line. The error bars correspond to 95 % confidence intervals according to the modified Wald method  
435 (Agresti & Coull 1998). The four pie charts represent data on prevalence whenever systematically  
436 recorded (red: sexuals, blue: asexuals, dark: infected, light: uninfected).

437

438 Figure 2. Outcome of the competition experiment between asexual and sexual *Daphnia pulex*  
439 exposed to *G. vavrai* in the summer of 2010. In August, the starting point of the experiment, there  
440 were an equal number of sexual and asexual individuals. Line graphs show the change in the  
441 frequency of sexuals and asexuals over the three month experiment. Error bars represent standard  
442 errors of the means across ten replicates. Pie charts show the proportion of the sexuals and asexuals  
443 infected with *G. vavrai* (red: sexuals, blue: asexuals, dark: infected, light: uninfected).

444

445 Figure 3. (a) Spore load at death and (b) parasite prevalence for the four asexual and four sexual  
446 clones used in the laboratory experiment. Error bars show the standard error of the mean values.

447

448 Figure 4. Life-history traits of asexual and sexual *Daphnia*, from either the control treatment,  
449 infection with spores of asexual origin treatment, or infection with spores of sexual origin treatment  
450 for (a) age at death, (b) age at first reproduction and (c) reproductive output (number of offspring per  
451 mother). Error bars show the standard error of the mean values. \*P < 0.05, \*\*P < 0.005.

452

453 Figure S1. Experimental design of the laboratory experiment. Four sexual and four asexual clones  
454 were each established from an independent pond within the Finnish archipelago.

455

456

457 **References**

- 458 Agresti, A. & Coull, B.A. (1998). Approximate is better than “exact” for interval estimation of  
459 binomial proportions. *Am. Stat.*, 52, 119-126.
- 460 Baker, H.G. (1955). Self-compatibility and establishment after 'long-distance' dispersal. *Evolution*,  
461 9, 347-349.
- 462 Ben-Ami, F. & Heller, J. (2005). Spatial and temporal patterns of parthenogenesis and parasitism in  
463 the freshwater snail *Melanooides tuberculata*. *J. Evol. Biol.*, 18, 138-146.
- 464 Bengtsson, J. & Ebert, D. (1998). Distributions and impacts of microparasites on *Daphnia* in a  
465 rockpool metapopulation. *Oecologia*, 115, 213-221.
- 466 Carlton, J.T. (1985). Transoceanic and interoceanic dispersal of coastal marine organisms: the  
467 biology of ballast water. *Oceanogr. Mar. Biol. Annu. Rev.*, 23, 313-374.
- 468 Decaestecker, E., Gaba, S., Raeymaekers, J.A.M., Stoks, R., Van Kerckhoven, L., Ebert, D. *et al.*  
469 (2007). Host-parasite “Red Queen” dynamics archived in pond sediment. *Nature*, 450, 870-  
470 873.
- 471 Deng, H. & Lynch, M. (1997). Inbreeding depression and inferred deleterious-mutation parameters  
472 In *Daphnia*. *Genetics*, 147: 147-155.
- 473 Duggan, I.C., Robinson, K.V., Burns, C.W., Banks, J.C. & Hogg, I.D. (2012). Identifying  
474 invertebrate invasions using morphological and molecular analyses: North American *Daphnia*  
475 ‘pulex’ in New Zealand fresh waters. *Aquat. Invasions*, 7, 585-590.
- 476 Ellstrand, N.C. & Schierenbeck, K.A. (2000). Hybridization as a stimulus for the evolution of  
477 invasiveness in plants? *Proc. Natl. Acad. Sci.*, 97, 7043–7050.
- 478 Elzinga, J.A., Chevasco, V., Mappes, J. & Grapputo, A. (2012). Low parasitism rates in  
479 parthenogenetic bagworm moths do not support the parasitoid hypothesis for sex. *J. Evol.*  
480 *Biol.*, 25, 2547-2558.
- 481 Fadda, A., Marková, S., Kotlík, P., Lugliè, A., Padedda, B., Buscaruni, *et al.* (2011). First record of

- 482 planktonic crustaceans in Sardinian reservoirs. *Biologia*, 66, 856.
- 483 Friedrich, C., Winder, O., Schaffler, K. & Reinthaler, F.F. (1996). Light and electron microscope  
484 study on *Gurleya daphniae* sp. nov. (Microspora, Gurleyidae), a parasite of *Daphnia pulex*  
485 (Crustacea, Phyllopora). *Eur. J. Protistol.*, 32, 116-122.
- 486 Gilabert, A., Simon, J.C., Dedryver, C.A. & Plantegenest, M. (2014). Do ecological niches differ  
487 between sexual and asexual lineages of an aphid species?. *Evol. Ecol.*, 28, 1095-  
488 1104.
- 489 Green, J. (1974). Parasites and epibionts of Cladocera. *Trans. Zool. Soc. London*, 32, 417-515.
- 490 Haag, C.R., Sakwińska, O. & Ebert, D. (2003). Test of synergistic interaction between infection and  
491 inbreeding in *Daphnia magna*. *Evolution*, 57, 777-783.
- 492 Haag, C.R. & Ebert, D. (2004). A new hypothesis to explain geographic parthenogenesis. *Ann.*  
493 *Zool. Fenn.*, 41, 539-544.
- 494 Haag, C.R., Riek, M., Hottinger, J.W., Pajunen, V.I. & Ebert, D. (2005). Genetic diversity and  
495 genetic differentiation in *Daphnia* metapopulations with subpopulations of known  
496 age. *Genetics*, 170, 1809-1820.
- 497 Hamilton, W.D. (1980). Sex versus non-sex versus parasite. *Oikos*, 35, 282-290.
- 498 Hamilton, W.D., Axelrod, R. & Tanses, R. (1990). Sexual reproduction as an adaptation to resist  
499 parasites (a review). *Proc. Natl. Acad. Sci.*, 87, 3566-3573.
- 500 Hanley, K.A., Fisher, R.N. & Case, T.J. (1995). Lower mite infestations in an asexual gecko  
501 compared with its sexual ancestors. *Evolution*, 49, 418-426.
- 502 Hatcher, M.J., Dick, J.T.A. & Dunn, A.M. (2006). How parasites affect interactions between  
503 competitors and predators. *Ecol. Lett.*, 9, 1253-1271.
- 504 Hatcher, M.J., Dick, J.T.A. & Dunn, A.M. (2008). A keystone effect for parasites in intraguild  
505 predation? *Biol. Lett.*, 4, 534-537.

- 506 Hebert, P.D.N. & Beaton, M.J. (1993). Methodologies for Allozyme Analysis Using Cellulose  
507 Acetate Electrophoresis. Helena Laboratories, Beaumont, TX, USA.
- 508 Heier, C.R. & Dudycha, J.L. (2009). Ecological speciation in a cyclic parthenogen: sexual  
509 capability of experimental hybrids between *Daphnia pulex* and *Daphnia pulicaria*. *Limnol.*  
510 *Oceanogr.*, 54, 492-502.
- 511 Innes, D.J. & Gin, M. (2014). A population of sexual *Daphnia pulex* resists invasion by asexual  
512 clones. *Proc. R. Soc. B*, 281, 20140564.
- 513 Innes, D.J. & Hebert, P.D.N. (1988). The origin and genetic basis of obligate parthenogenesis in  
514 *Daphnia pulex*. *Evolution*, 42, 1024-1035.
- 515 Jaenike, J. (1978). A hypothesis to account for the maintenance of sex within populations. *Evol.*  
516 *Theor.*, 3, 191-194.
- 517 Keane, R.M. & Crawley, M. (2002). Exotic plant invasions and the enemy release hypothesis.  
518 *Trends Ecol. Evolut.*, 17, 167-170.
- 519 Kotusz, J., Popiołek, M., Drozd, P., De Gelas, K., Šlechtová, V. & Janko, K. (2014). Role of parasite  
520 load and differential habitat preferences in maintaining the coexistence of sexual and asexual  
521 competitors in fish of the *Cobitis taenia* hybrid complex. *Biol. J. Linn. Soc.*, 113, 220-235.
- 522 Klüttgen, B., Dülmer U., Engels, M. & Ratte H.T. (1994). ADaM, an artificial freshwater for the  
523 culture of zooplankton. *Water Res.*, 28, 743-746.
- 524 Kumpulainen, T., Grapputo, A. & Mappes, J. (2004). Parasites and sexual reproduction in psychid  
525 moths. *Evolution*, 58, 1511-1520.
- 526 Lehto, M.P. & Haag, C.R. (2010). Ecological differentiation between coexisting sexual and  
527 asexual strains of *Daphnia pulex*. *J. Anim. Ecol.* 79, 1241-1250.
- 528 Little, T.J. & Ebert, D. (1999). Associations between parasitism and host genotype in natural  
529 populations of *Daphnia* (Crustacea: Cladocera). *J. Anim. Ecol.*, 68, 134-149.
- 530 Lively, C.M. & Jokela, J. (2002). Temporal and spatial distributions of parasites and sex in a

- 531 freshwater snail. *Evol. Ecol. Res.*, 4, 219-226.
- 532 Lohr, J.N. & Haag, C.R. (2015). Genetic load, inbreeding depression, and hybrid vigor covary with  
533 population size: An empirical evaluation of theoretical predictions. *Evolution*, 69, 3109-3122.
- 534 Lynch, M., Seyfert, A., Eads, B. & Williams, E. (2008). Localization of the genetic determinants  
535 of meiosis suppression in *Daphnia pulex*. *Genetics*, 180, 317-327.
- 536 Lloyd, D.G. (1980). Benefits and handicaps of sexual reproduction. In *Evolutionary biology* (pp. 69  
537 -111). Springer US.
- 538 Marková, S., Dufresne, F., Manca, M. & Kotlík, P. (2013). Mitochondrial capture misleads about  
539 ecological speciation in the *Daphnia pulex* complex. *PLoS One*, 8, e69497.
- 540 Meirmans, S., Meirmans, P. G. & Kirkendall, L. R. (2012). The costs of sex: facing real-world  
541 complexities. *Quart. Rev. Biol.*, 87, 19-40.
- 542 Mergeay, J., Verschuren, D. & De Meester, L. (2006). Invasion of an asexual American water flea  
543 clone throughout Africa and rapid displacement of a native sibling species. *Proc. Biol. Sci.*,  
544 273, 2839-2844.
- 545 Mergeay, J., Aguilera, X., Declerck, S., Petrusek, A., Huyse, T. & De Meester, L. (2008). The  
546 genetic legacy of polyploid Bolivian *Daphnia*: the tropical Andes as a source for the  
547 North and South American *D. pulicaria* complex. *Mol. Ecol.*, 17, 1789-1800.
- 548 Morran, L.T., Schmidt, O.G., Gelarden, I.A., Parrish, R.C. & Lively, C.M. (2011). Running with the  
549 Red Queen: host-parasite coevolution selects for biparental sex. *Science*, 333, 216–218.
- 550 Otto, S.P. & Lenormand, T. (2002). Resolving the paradox of sex and recombination. *Nature Rev.*  
551 *Genet.* 3, 252–261. □
- 552 Pajunen, V.I. (1986). Distributional dynamics of *Daphnia* species in a rock-pool environment.  
553 *Ann. Zool. Fenn.*, 23, 131-140.
- 554 Pajunen, V.I. & Pajunen, I. (2003). Long-term dynamics in rock pool *Daphnia*  
555 metapopulations. *Ecography*, 26, 731-738.



- 556 Pajunen, V.I. & Pajunen, I. (2007). Habitat characteristics contributing to local occupancy and  
557 habitat use in rock pool *Daphnia* metapopulations. *Hydrobiologia*, 592, 291-302.
- 558 Paland, S., Colbourne, J.K. & Lynch, M. (2005). Evolutionary history of contagious asexuality in  
559 *Daphnia pulex*. *Evolution*, 59, 800-813.
- 560 Peck, J.R., Yearsley, J.M. & Waxman, D. (1998). Explaining the geographic distributions of sexual  
561 and asexual populations. *Nature*, 391, 889-892.
- 562 R Core Team (2013). R: A Language and Environment for Statistical Computing. R Foundation for  
563 Statistical Computing, Vienna, Austria.
- 564 Refardt, D., Canning, E.U., Mathis, A., Cheney, S.A., Lafranchi-Tristem, N.J. & Ebert, D. (2002).  
565 Small subunit ribosomal DNA phylogeny of microsporidia that infect *Daphnia* (Crustacea:  
566 Cladocera). *Parasitology*, 124, 381-389.
- 567 Sakai, A.K., Allendorf, F.W., Holt, J.S., Lodge, D.M., Molofsky, J., With, K.A. *et al.* (2001). The  
568 population biology of invasive species. *Annu. Rev. Ecol. Syst.*, 32, 305-332.
- 569 Salathé, M., Kouyos, R.D. & Bonhoeffer, S. (2008). The state of affairs in the kingdom of the Red  
570 Queen. *Trends Ecol. Evolut.*, 23, 439-445.
- 571 Searle, C.L., Cortez, M.H., Hunsberger, K.K., Grippi, D.C., Oleksy, I.A., Shaw, C.L., *et al.*  
572 (2016). Population density, not host competence, drives patterns of disease in an  
573 invaded community. *Am. Nat.*, 188, 554-566.
- 574 Simon, J.C., Delmotte, F., Rispe, C. & Crease, T. (2003). Phylogenetic relationships between  
575 parthenogens and their sexual relatives: the possible routes to parthenogenesis in animals.  
576 *Biol. J. Linn. Soc.*, 79, 151-163.
- 577 So, M., Ohtsuki, H., Makino, W., Ishida, S., Kumagai, H., Yamaki, K.G. *et al.* (2015). Invasion and  
578 molecular evolution of *Daphnia pulex* in Japan. *Limnol. Oceanogr.*, 60, 1129-1138.
- 579 Stirnadel, H.A. & Ebert, D. (1997). Prevalence, host specificity and impact on host fecundity of  
580 microparasites and epibionts in three sympatric *Daphnia* species. *J. Anim. Ecol.*,

- 581           66, 212-222.
- 582 Thomas F., Bonsall, M.B. & Dobson, A.P. (2005). Parasitism, biodiversity and conservation. In  
583 Parasitism and Ecosystems: Thomas, F., Renaud, F. & Geugan, J.F. pp. 124–139. Eds.  
584 Oxford, UK. Oxford University Press.
- 585 Torchin, M.E., Lafferty, K.D., Dobson, A.P., McKenzie, V.J. & Kuris, A.M. (2003). Introduced  
586 species and their missing parasites. *Nature*, 421, 628-630.
- 587 Torchin, M.E. & Mitchell, C.E. (2004). Parasites, pathogens, and invasion by plants and animals.  
588 *Front. Ecol. Environ.*, 2, 183-190.
- 589 Tucker, A.E., Ackerman, M.S., Eads, B.D., Xu, S. & Lynch, M. (2013). Population-genomic insights  
590 into the evolutionary origin and fate of obligately asexual *Daphnia pulex*. *Proc. Natl. Acad.*  
591 *Sci.*, 110, 15740-15745.
- 592 Vandel, A. (1928). La parthénogénèse géographique. Contribution à l'étude biologique de la  
593 parthénogénèse naturelle. *Bull. Biol. France Belg.*, 62, 164-281.
- 594 Vergara, D., Jokela, J. & Lively, C.M. (2014). Infection dynamics in coexisting sexual and  
595 asexual host populations: support for the red queen hypothesis. *Am. Nat.*, 184, S22-S30.
- 596 Walser, B. & Haag, C. R. (2012). Strong intraspecific variation in genetic diversity and genetic  
597 differentiation in *Daphnia magna*: the effects of population turnover and population  
598 size. *Mol. Ecol.*, 21, 851-861.
- 599 Ward, R.D., Bickerton, M.A., Finston, T. & Hebert, P.D. (1994). Geographical cline in breeding  
600 systems and ploidy levels in European populations of *Daphnia pulex*. *Heredity*, 73, 532-543.
- 601 Williams, R. J., Griffiths, F.B., Van der Wal, E.J. & Kelly, J. (1988). Cargo vessel ballast water as a  
602 vector for the transport of non-indigenous marine species. *Estuar. Coast. Shelf Sci.*, 26, 409-  
603 420.
- 604 Xu, S., Spitze, K., Ackerman, M., Ye, Z., Bright, L., Keith, *et al.* (2015). Hybridization and the  
605 origin of contagious asexuality in *Daphnia pulex*. *Mol. Biol. Evol.*, 32, 3215-3225.







