Bridging multilocus species delimitation and DNA barcoding through target enrichment of

UCEs: a case study with Mexican highland frogs

Eugenia Zarza[1], Elizabeth M. Connors[1], James M. Maley[1], Whitney L.E. Tsai[1], Peter Heimes[2],

Moises Kaplan[3] & John E. McCormack[1,*]

[1] Moore Laboratory of Zoology, Occidental College, Los Angeles, CA 90041

[2] Playa Miramar 441, Colonia Marte 08830, Mexico City, Mexico

[3] 1508 Brooklyn Ave, Ann Arbor, Michigan 48104

*Corresponding author: mccormack@oxy.edu

Abstract

Species delimitation has been divided by two approaches: DNA barcoding that focuses on

standardization of the genetic marker and multilocus methods that place a premium on

genomic coverage and conceptual rigor in modeling the divergence process. Most multilocus

methods fail as barcodes, however, because few assay the same marker set and are therefore not

readily comparable across studies and databases. We introduce ultraconserved elements (UCEs)

as potential genomic barcodes that allow rigorous species delimitation and a bridge to DNA

barcoding database to allow both rigorous species delimitation and standardized identification

of delimited taxa. UCEs query thousands of loci across the nuclear genome in way that is

replicable across broad taxonomic groups (i.e., vertebrates). We apply UCEs to species

delimitation in a species complex of frogs found in the Mexican Highlands. *Sarcohyla* contains

24 described species, many of which are critically endangered and known only from their type

localities. Evidence suggests that one broadly distributed member of the genus, *S. bistincta*,

might contain multiple species. We generated data from 1,891 UCEs, which contained 1,742

informative SNPs for *S. bistincta* and closely related species. We also captured mitochondrial

genomes for most samples as off-target bycatch of the UCE enrichment process. Phylogenies

from UCEs and mtDNA agreed in many ways, but differed in that mtDNA suggested a more

complex evolutionary history perhaps influenced by reticulate processes. The species

delimitation method we used identified eight putative species (which we call lineages pending

further study) within *S. bistincta*. Being able to compare linked mtDNA data to existing

sequences on Genbank allowed us to identify one of these lineages nested within *S. bistincta* as

an already-described species, *S. pentheter*. Another lineage nested within *S. bistincta* is currently

being described as a new species (referred to here as *sp. nov.*). The remaining six lineages fell

into two non-sister clades, one containing the core *S. bistincta* mostly in Oaxaca and Guerrero,

and another in the Transvolcanic Belt. The latter clade, at 10% divergence in mtDNA and

paraphyletic with respect to other *S. bistincta*, is a clear candidate for species status. Our study

demonstrates not only that UCEs can be used as effective genomic DNA barcodes, but that

combining multilocus genomic data with mtDNA is a powerful approach for both delimiting

species and identifying them in poorly described and phenotypically challenging groups.

Keywords: ultraconserved elements, genomics, population genetics, phylogeography,

phylogenetics, systematics, species limits, Hylinae

Introduction

Species are the fundamental biological unit and their identification and description is critical to ecology, evolution, and conservation (DeSalle and Amato 2004; Isaac et al. 2004; Mace 2004; Wiens 2007). Identifying species has always been a challenging issue in biology, especially when lineages are near the cusp of speciation (De Queiroz 2007). Yet, for the vast majority of researchers working on poorly described groups under conservation threat, conceptual debates over identifying the precise moment when one species becomes two have been superseded by the practical need for methods to quickly assess and identify the basic units of biodiversity (Hey et al. 2003; Tautz et al. 2003; Bickford et al. 2007). As a result, species delimitation methods have multiplied over the last ten years, especially those using DNA data (Pons et al. 2006; Knowles and Carstens 2007; O'Meara 2009; Yang and Rannala 2010; Fujita et al. 2012; Leaché et al. 2014).

While there are now a wide variety of methods that employ different DNA sources and algorithms to delimit species, these methods can be broadly placed into two frameworks. DNA barcoding focuses on species delimitation through the application of thresholds of sequence divergence, usually at a single mitochondrial DNA (Hebert et al. 2003) or chloroplast locus (Taberlet et al. 2007). DNA barcoding is relatively cheap and easy to implement, and by focusing on a single marker, a large database has been built by researchers, which allows for comparison of newly delimited species to already-identified species that have associated DNA. With its large user base, DNA barcoding has made substantial contributions to taxonomy and conservation of poorly known groups and geographic areas (Hebert et al. 2004; Ward et al. 2005; Witt et al. 2006; Lahaye et al. 2008). The very simplicity of DNA barcoding has made it controversial since its inception (Moritz and Cicero 2004; DeSalle et al. 2005; Ebach and Holdrege 2005; Will et al. 2005). Critics argue that it oversimplifies complex divergence processes with arbitrary thresholds that vary from one organismal group to another. Also, the history of a single gene will often fail to reflect the evolutionary history of populations and

species (Edwards and Bensch 2009; Galtier et al. 2009), thereby misleading delimitation efforts

in some cases.

Multilocus species delimitation, on the other hand, addresses many of the perceived

weaknesses of DNA barcoding by incorporating a process of gene inheritance (coalescence) and

by using information from many genomic locations to more accurately reflect the history of

lineages and how they diverged. Multilocus species delimitation, however, has been criticized

on the practical grounds that large amounts of genomic data are usually not needed to delimit

species, and the marginal gains more loci might afford are outstripped by the computational

burden of modeling complex speciation processes with large data sets (Collins and Cruickshank

2014). Another key criticism, which we focus on specifically, is that the lack of standardized

multilocus marker sets inhibits the broad adoption of multilocus delimitation methods because

newly delimited taxa cannot be readily compared to an existing database to determine if the

putative new species have already been identified and named and, if not, how they relate to

other species in an evolutionary context.

Our goal is to introduce ultraconserved elements (UCEs) as a candidate for a

standardized genomic marker set for multilocus species delimitation and identification and to

show how this kind of multilocus data can be even more powerful when linked to DNA

barcoding-type data (i.e., mtDNA). UCEs are appealing as genomic barcodes because the same

loci are found across major branches of the tree of life, where they act as anchors for variable

DNA in flanking regions (Faircloth et al. 2012). For instance, it is possible to capture the same

set of 1000 or more UCEs across all mammals (McCormack et al. 2012), all reptiles (Crawford et

al. 2012), or hundreds of UCEs across arachnid lineages separated by hundreds of millions of

years (Starrett et al. 2016). While the power of UCEs for deep-level systematics is clear, their

utility at shallower scales around the species level has only recently been demonstrated (Smith

et al. 2014; McCormack et al. 2016; Zarza et al. 2016). An added benefit of the UCE enrichment

process (and all so-called "sequence capture" methods) is that whole mtDNA genomes are often

captured as off-target "bycatch" (do Amaral et al. 2015), allowing for each individual to have

associated nuclear and mtDNA data (e.g., Zarza et al. 2016).

We apply UCEs as genomic barcodes to species delimitation in treefrogs in the genus

Sarcohyla in the northern Mexican Highlands. This genus, which was recently split from

Plectrohyla to reflect those species west of the Isthmus of Tehuantepec (Duellman et al. 2016),

contains 24 described species, many of them critically endangered and many that have never

been seen after their original discovery (references compiled in Stuart et al. 2008). Some species

are thought to be in serious decline or extinct (Lips et al. 2004). Species limits and relationships

within *Sarcohyla* are not well known because few species have associated DNA sequences, and

the lack of voucher material means that the tadpole stages and geographic ranges for each

species are not well characterized (Duellman 2001; Faivovich et al. 2009; Duellman et al. 2016).

We focus specifically on the *Sarcohyla bistincta* complex and closely related species. *Sarcohyla*

bistincta is one of the more broadly distributed members of the genus and might comprise

multiple species. Especially in its early developmental stages, *S. bistincta* looks similar to closely

related species, meaning that identification from tadpoles can be challenging. Thus, *Sarcohyla*

frogs present a good test case of whether UCE data are effective at delimiting species, and how

linking UCE and mtDNA data is a powerful approach for helping identify delimited taxa by

comparison to existing genetic databases.

Methods

Sampling and Ingroup Determination

MK collected tadpoles from January to June 2004 across most of the range where *Sarcohyla*

bistincta are known to exist (Duellman 2001) in the Transvolcanic Belt of Michoacán, Morelos,

and the state of México, the Sierra Madre del Sur of Guerrero, and the highlands of Oaxaca

stretching into Puebla and Veracruz (Fig. 1; Table 1). Unsampled parts of the *S. bistincta* range

include the far west Transvolcanic Belt in Michoacán and Jalisco, the far northwest in the Sierra

Madre Occidental (Nayarit, Durango, and Sinaloa), and the far northeast in Hidalgo (see Fig. S1

for sampled and unsampled locations and known ranges of all *Sarcohyla* species). Tadpoles

were targeted to improve sampling efficiency. After collection of many tadpoles from a

sampling location with a dip net, they were separated by species based on morphology to the

extent possible and reared to subadults in the laboratory prior to vouchering. Species

identification was based on the most recent diagnosis of *S. bistincta* and other closely related

species (Duellman 2001). One tadpole was chosen for the tissue voucher, while the other

individuals became physical vouchers with museum catalog numbers. Thus, we provide both

field numbers and catalog numbers in Table 1 to provide a link to both the exact genetic

material and the associated voucher for that genotype. Before limiting our taxonomic sampling

to 38 *S. bistincta* samples and two close outgroups, we ran preliminary phylogenetic analyses

also including samples from 45 *Sarcohyla* individuals and the outgroup genus *Exerodonta* to

ensure we had correctly identified the ingroup and closest outgroups (Table S1).

Sequence capture and next-generation sequencing

We extracted genomic DNA from tissue using a Qiagen (Valencia, CA) DNAeasy Blood and

Tissue extraction kit. We visualized extractions on an agarose gel to ensure fragments were

larger than 200 base pairs (bp) and quantified the resulting double-stranded DNA using a Qubit

2.0 Fluorometer (Carlsbad, CA). For each sample, we sheared 100 μl of 20ng/μl concentration

DNA to a size distribution with its peak between 400 and 600 bp using a Bioruptor

ultrasonicator (Diagenode). We prepared libraries for each sheared sample with a KAPA

(Boston, MA) LTP library preparation kit for the Illumina platform, attaching custom indexing

tags (Faircloth and Glenn 2012) to each sample to allow sample pooling.

We enriched pools of eight samples using a set of synthetic RNA probes that target 5,060

tetrapod UCEs (MYbaits_Tetrapods-UCE-5K kit, Mycroarray) following the standard UCE

enrichment protocol (Faircloth et al. 2012) with one modification. Amphibians have large and

variable genome sizes with a high percentage of repetitive DNA (Olmo 1991). While we do not

have information about the genome size and composition of *Sarcohyla* specifically, we wanted to
decrease the potential risk of the probes hybridizing to repetitive elements (McCartney-Melstad
et al. 2016). We thus increased by 6X the amount of the Cot-1 blocker, a synthetic DNA derived
from chicken that binds to repetitive regions. After enrichment and recovery PCR, we verified
the library size range with an Agilent 2100 Bioanalyzer (Palo Alto, CA). We quantified the
enriched pools using qPCR and combined them in equimolar ratios before sequencing on an
Illumina HiSeq 2000 lane (100-bp paired-end cycle) at the University of California Santa Cruz
Genome Technology Center.

Bioinformatics of next-generation sequencing data

We demultiplexed the Illumina raw reads and converted them to FASTQ format with the
program bcl2fastq v.1.8.4 (Illumina, Inc.). We created quality control reports with FASTQC
v0.10.1 to detect overrepresented sequences and adapter contamination. To eliminate adapter
contamination and low quality bases, we trimmed the FASTQC output using illumiprocessor
(Faircloth 2012). We trimmed and assembled these reads into contigs with Trinity (Haas et al.
2013) and ABySS (Simpson et al. 2009), both of which are built into the PHYLUCE pipeline
(Faircloth 2015). PHYLUCE uses LASTZ (Harris 2007) to align all assembled contigs to UCE
probe sequences in order to isolate only UCE contigs and to identify and eliminate any paralogs
(MATCH_CONTIGS_TO_PROBES script). We then turned this filtered set of UCE contigs into a
data matrix describing which UCE loci are present in which samples (GET_MATCH_COUNTS
script).

Phylogenetic trees from concatenated UCE data

We extracted contigs into a single FASTA file (GET_FASTAS_FROM_MATCH_COUNTS script)
and aligned the output for each locus (SEQCAP_ALIGN script) using MAFFT (Katoh et al.
2005). We then required that 75% of the samples needed to have data for a given locus to be
included in the final concatenated matrix (GET_LOCI_WITH_MIN_TAXA script). We

converted the NEXUS file into PHYLIP format (FORMAT_NEXUS_FILES_FOR_RAxML script)

and constructed a maximum-likelihood (ML) tree in RAxML v8.0.19 (Stamatakis 2014) under

the GTRGAMMA model of evolution with 100 bootstrap searches, followed by a search for the

tree with the highest likelihood.

Mitochondrial DNA assembly and analysis

We identified and assembled mtDNA genomes from off-target, trimmed Illumina reads using

the reference genome of a closely related species, *Hyla annectans* (Genebank accession number

KM271781; Ye et al. 2016). We used MITObim 1.7 (Hahn et al. 2013), a Perl wrapper for MIRA

4.0.2 (Chevreux et al. 1999), that takes a baiting and iterative mapping approach for assembly.

We conducted *de novo* annotation of the assembled mtDNA regions with the MITOchondrial

genome annotation Server, MITOS (Bernt et al. 2013). We selected for phylogenetic analysis

only those individual genomes with MIRA quality score grater than 30. We aligned each

protein-coding region separately in Geneious vR8 (Kearse et al. 2012) using the integrated

MUSCLE (Edgar 2004) plugin. We corrected the alignments manually when necessary and

constructed a concatenated mtDNA matrix, which we also ran in RAxML v8.0.19.

Additionally, we melded our mtDNA data with existing *Sarcohyla* and *Plectrohyla*

mtDNA data on Genbank to determine whether any of the lineages we uncovered in *S. bistincta*

relate to already-described species. We determined that *cytochrome b* is the best-represented on

Genbank in this group. We downloaded all existing *cytochrome b* sequences from *Sarcohyla* and

Plectrohyla taxa. We combined these sequences with those from a subset of our *S. bistincta*

samples, choosing the individual with the most raw reads from each major genetic lineage in

the UCE tree. We used *Exerodonta* as an outgroup. Here, we aligned the trimmed, filtered reads

for each individual to a *Sarcohyla cytochrome b* reference sequence. We formed a consensus

sequence for each individual from the mapped reads. We then created an alignment and

generated a phylogeny using BEAST v2.4.2 (Bouckaert et al. 2014). Because we were particularly

interested in how *S. pentheter* related to our lineages, but there are no *cytochrome b* sequences for

S. pentheter on Genbank, we also compared our samples to Genbank samples using the *16S*

gene.

Calling SNPs from UCE loci

We called SNPs from UCE loci so that we could run genetic clustering tests and infer a species

tree. Calling SNPs requires a reference sequence, and we chose the sample with the most UCE

contigs recovered within the ingroup (UMMZ 239727). We then used BWA (Li and Durbin

2009) to map the reads of each sample to this reference. We used SAMtools (Li et al. 2009) to

sort the reads, and Picard (available at http://broadinstitute.github.io/picard) to identify and

remove PCR duplicates. We realigned the mapped reads to minimize mismatched bases due to

indels, and we removed indels using the Genome Analysis Toolkit 3.2 (GATK; McKenna et al.

2010) and a custom script (indelrealigner.sh), as suggested by the Best Practices workflow

(DePristo et al. 2011; van der Auwera et al. 2013).

There is no SNP database available for treefrogs, so we followed best practices for base

recalibration for non-model organisms suggested by GATK (McKenna et al. 2010). This consists

of (1) doing an initial round of calling SNPs on the original, uncalibrated data, (2) selecting the

SNPs with the highest confidence (a minimum emission and call quality of 40 or more), and (3)

using these SNPs as the database of known SNPs. We executed four rounds of base

recalibration on the original data to filter out systematic error using a custom script (genotype-

recal.sh). We called genotypes on the last recalibrated BAM file. We used vcf-tools (Danecek et

al. 2011) to select one SNP per UCE and produce two data sets, one allowing 25% missing data

for STRUCTURE v 2.3.4 (Pritchard et al. 2000), and one with no missing data, which is a

requirement for SNAPP (Bryant et al. 2012) species tree analysis, implemented in BEAST v2.2.1

(Bouckaert et al. 2014).

STRUCTURE analyses

The species delimitation method we used requires some *a priori* information about possible

clustering of individuals into species. We used STRUCTURE v2.3.4 as an unbiased way to

assess the limits of fine-scale genetic structure in our data. Thus, our intent was not to

determine the single most likely number of genetic clusters. Rather, our goal was to determine

the maximum number of genetic clusters in our data. We began by analyzing all individuals of

S. bistincta plus two outgroup species *S. chryses* and *S. hazelae* under $K=4$, reasoning that this

would likely split out the two outgroups as well as revealing one division within the ingroup.

After this, each identified genetic cluster was further analyzed at $K=2$ until no further structure

was evident.

SNAPP tree and species delimitation

We generated a species tree from the SNP matrix using SNAPP 1.1.10 (Bryant et al. 2012). This

analysis included all putative *S. bistincta* samples and one outgroup, *S. chryses.* For this run, we

made no a priori assumptions about how individuals grouped into species and allowed each

individual to be considered its own "species" (i.e., terminal tip). We ran two instances of

SNAPP for seven million generations using default priors. We combined tree and parameter

files from both runs with LogCombiner 2.1.3 and displayed the full set of likely species trees

with Densitree v2.2.1 (Bouckaert et al. 2014).

We then applied Bayes Factor Species Delimitation (BFD; Grummer et al. 2014; Leaché et

al. 2014) to SNAPP analyses to determine the most likely number of species. We used the

Structure results to generate hypotheses for how individuals might cluster into species. We

tested five scenarios with BFD that split or lumped populations identified in the Structure

analysis, including the hypothesis that all putative *S. bistincta* group into a single species. We

assessed which scenario (2, 6, 7, 9, or 10 species) was favored by comparing Bayes factors using

marginal likelihoods from the SNAPP analysis.

Results

NGS summary statistics

Detailed summary statistics for each of the 38 ingroup samples and two outgroups are described in Table 1. ABySS produced longer contigs than Trinity, and a higher number of UCE loci, so we used ABYSS contigs in all downstream analyses. Reads per sample ranged from 17,052 to 3,423,330 with an average of 1,185,165 reads. The number of identified UCEs ranged from 381 to 2,444 with an average of 1,976 UCEs. The mean length of individual UCE loci per individual ranged from 222 to 717 bp with an average of 522 bp. On average, 18% of the assembled contigs corresponded to unique UCE loci.

For SNP calling, across 40 samples of *S. bistincta* and outgroups, 9% of the trimmed reads mapped to our designated reference individual. The SNP read depth ranged from 2.4 to 35.0 with an average depth of 21.2. The recalibration and quality control steps resulted in an initial matrix of 16,578 SNPs. After removing non-biallelic loci, selecting one SNP for every UCE, and allowing 25% missing data, there were 1,742 SNPs in the STRUCTURE data set, while the 100% complete data matrix for SNAPP contained 399 SNPs.

UCE phylogeny from concatenated data

Our more taxonomically inclusive data set with all available *Sarcohyla* and outgroup *Exerodonta xera* (Table S1) contained 1,866 UCE loci and 1,030,450 bp for a concatenated analysis. The resulting ML tree (Fig. S2) showed strong support for monophyly of *Sarcohyla*, and identified *S. arborescandens* and *S. cyclada* as sister species that together form a clade sister to the rest of the *Sarcohyla* included in the study. We thus limited further analyses to a smaller data set of 40 samples with *S. hazelae* as the outgroup (Table 1). This focal data set contained 1,891 UCE loci and 1,038,600 bp. The ML tree of these 40 samples found strong support for many clades within the species currently described as *S. bistincta*, conforming to distinct geographic areas (Fig. 2a). In brief, there were three clades on relatively long branches: one distributed across the

Transvolcanic Belt (blue clade); one inhabiting two disjunct areas along the coastal slopes of the

Sierra Madre del Sur in Guerrero and Oaxaca (gray clade); and one broadly distributed in the

Sierra Madre del Sur (red and pink clades), the Oaxaca Highlands (yellow + orange clade), and

one individual in the southern portion of the Transvolcanic Belt (purple). One individual that

nested within *S. bistincta* was labeled as a different species, *S. mykter*, from Guerrero. We

suspect that this sample was mislabeled and is actually a duplicate of an *S. bistincta* sample

already included in the study because their field numbers are similar (last two digits

transposed) and the two samples grouped together in all analyses. We have left this sample in

all analyses, but labeled it as a duplicate of *S. bistincta* UMMZ 239749.

mtDNA tree

Our final concatenated mtDNA matrix was 11,269 base pairs including gaps, as coverage of the

mtDNA genome varied from sample to sample in accordance with the non-targeted nature of

the DNA collection (Table 1). Relationships in the ML tree (Fig. 2b) among the 29 individuals

with high quality scores were similar to the concatenated UCE tree with two key differences

within the broadly distributed clade in Guerrero and Oaxaca: (1) in the mtDNA tree,

individuals from Eastern and Western Guerrero (pink and red) formed a clade, whereas they

were more divergent in the UCE tree; (2) in the mtDNA tree, individual UMMZ 239731 (purple)

was still on a relatively long branch, but that branch was nested within the Guerrero clade

described above instead of being sister to a much more expansive clade, as in the UCE tree.

Structure analysis

The first run of STRUCTURE at K=4 split the two outgroup species into distinct clusters and

split the remaining individuals into two clusters. Further analysis of each cluster at K=2

revealed nine genetic clusters (Fig. 1b), which are concordant with clades observed in the UCE

and mtDNA phylogeny, as well as with the delimited species (lineages) described below.

SNAPP species tree and species delimitation

The cloudogram of posterior species trees obtained with SNAPP (Fig. 3) is consistent with the

genetic clusters in the Structure analysis and with relationships in phylogenies resulting from

the UCE and mtDNA concatenated analyses. With respect to the discrepancies between the two,

the SNAPP tree agreed with the mtDNA tree that eastern and western Guerrero individuals

form a clade, but agreed with UCE tree in the placement of individual UMMZ 239731. Bayes

factor analysis of SNAPP runs with varying number of species suggested the best model

consists of nine species. The model with nine species was supported by a factor of 10 over other

models with more and fewer species (Table 2). These putative species are labeled as lineages in

Fig. 3 and elsewhere, pending further study.

mtDNA phylogeny combining new data with Genbank sequences

Using 16S sequences, we determined that our delimited Lineage 8 matched an *S. pentheter*

sequence on Genbank. Lineage 8 individual UMMZ 239772 had one of the lowest read counts of

any samples and very few mtDNA reads. However, five reads mapped to 16S covering 421 bp

of the 681 bp reference sequence (Genbank *S. pentheter* accession number DQ055825). Over this

stretch, UMMZ 239772 was identical to the reference. As a point of comparison, UMMZ 239679

(a member of the blue *S. bistincta* Lineage 1 in the Transvolcanic Belt) had 70 differences across

the 681 bp (10.3% divergence). This DNA identification of UMMZ 239772 as *S. pentheter* was

confirmed by re-examining the subadult specimen.

After confirming UMMZ 239679 as *S. pentheter*, we generated a Bayesian tree of

cytochrome b combining our samples with Genbank sequences (Fig. 4). This tree revealed not

only that Lineage 8 (*S. pentheter*) was nested within the current *S. bistincta*, but so was another

species not included in our sampling, *S. calthula*. Relationships within *S. bistincta* were

otherwise similar to the more taxonomically inclusive mtDNA phylogeny. For example UMMZ

239731 (purple) was still nested within the Guerrero samples (red and pink). The combined

phylogeny also helped clarify relationships outside of *S. bistincta* by supporting *S. chryses + S.*

mykter to be sister to the *S. bistincta + S. pentheter + S. calthula* complex.

Discussion

One limitation of current multilocus species delimitation methods is the lack of a standardized marker set that would allow for data from delimited species to be combined with already-existing data both to identify species and place them into broader evolutionary context (Collins and Cruickshank 2014; Coissac et al. 2016). Here, we show that UCEs not only are a good candidate for this standardized "genomic barcode" for multilocus species delimitation, but also that they are especially powerful when they can be linked to mtDNA data, which provides a bridge to well-developed mtDNA databases for species identification. Having linked mtDNA data allowed us to query our lineages, delimited with multilocus genomic data, against GenBank to see if any of them corresponded to already-described species and to place them in as broad a phylogenetic context as possible. Doing so revealed that *S. bistincta* is paraphyletic, with two already-described species and one species in the process of description nested within its current species limits (Fig. 4). Our approach of delimiting species with multilocus data and identifying them with the help of mtDNA data will be especially powerful in groups where geographic ranges and subadult phenotypes are not well described or where species are phenotypically cryptic.

It was not our intention with this study to compare species delimitation methods, as this has been examined elsewhere (Camargo et al. 2012; Carstens et al. 2013; Miralles and Vences 2013). We employed one, rather new delimitation method (Leaché et al. 2014), but there are others that would surely suggest somewhat higher or lower numbers of delimited species. We are not advocating that species be delimited and described solely based on DNA. There are cases where both DNA barcoding and multilocus species delimitation will under-split species (Hime et al. 2016) and cases where coalescent methods will confound population structure with speciation, thereby recommending over-splitting species (Sukumaran and Knowles 2017). We discuss the results of our delimitation analysis below, but by calling the putative species "lineages" we acknowledge that new species descriptions should await integrative taxonomy

including, at minimum, careful study of phenotypic differences in museum specimens. Finally,

although we discuss discrepancies between the mtDNA and nuclear DNA phylogenies below,

we did not set out explicitly to compare multilocus species delimitation to DNA barcoding. As

mentioned, we feel our results demonstrate how mtDNA and nuclear data are most powerful

when used together. Despite a few notable differences, discussed below, our UCE and mtDNA

phylogenies are broadly congruent, suggesting that, if we carried out such a comparison, our

results would be in line with prior work showing that multilocus species delimitation and DNA

barcoding produce similar results (Collins and Cruickshank 2014; Dowton et al. 2014).

Rather, in addition to showing the power of linking nuclear genomic data with mtDNA

data, another goal of our study was to demonstrate that UCEs adhere to the concept of a

"extended genomic barcode" (sensu Coissac et al. 2016) and produce data sets with enough

resolving power for the necessarily recent divergences involved in species delimitation.

Another recent study applied UCEs to species delimitation in two other frog genera,

Melanophryniscus and *Brachycephalus* (Pie et al. 2017). This study used a more restrictive cut-off

for including loci and produced 800 loci (in a 100% complete data matrix) with sufficient

resolving power for species delimitation, compared to the roughly 1,900 loci (in a 75% complete

data matrix) in our study. UCE probe sets are now available for many taxonomic groups

(Faircloth et al. 2013; Faircloth et al. 2015; Starrett et al. 2016). They capture a discrete and

replicable portion of the genome, in this case a set of around 2,000 loci in frogs (from a larger set

of ~5,000 vertebrate loci) that query approximately 1,000,000 base pairs, or 0.02% of the frog

genome. The replicable nature of UCEs sets them apart from other types of genomic markers,

like RAD loci, which can vary from experiment to experiment (DaCosta and Sorenson 2014) and

find fewer orthologs with increasing phylogenetic distance (Cruaud et al. 2014).

Bayes factor species delimitation identified nine lineages within the current *S. bistincta*

and a close outgroup, *S. chryses*. Apart from *S. chryses*, which was correctly delimited, the other

eight lineages fell into three clades (Fig. 3). One lineage turned out to be the already-described

S. pentheter of Oaxaca, which we discovered by querying its 16S mtDNA data against existing

Genbank sequences and reexamining the subadult specimen. Thus, both *S. pentheter* and *S.*

calthula are nested within the currently-described *S. bistincta*. Below, we describe the three

clades containing the eight lineages. Where warranted, formal species descriptions will follow

in later publications.

Clade 1 – Members of this clade are found in the Transvolcanic Belt of central Mexico. They are

sister to the rest of the current *S. bistincta* + *S. pentheter* + *S. calthula*. This clade is a

clear candidate for species recognition and might itself contain multiple species, as

delimited below. Based on geography, unsampled populations of *S. bistincta* in the

Sierra Madre Occidental (Fig. S1) are most likely related to this clade, but should be

included in future studies, as they might comprise their own lineage.

Lineage 1 (light blue in Fig. 1) – Michoacán to western Mexico state. There is fine-scale

genetic structure across this range, and the presence of a geographic and genetic

intermediate hints at continuity of gene flow along the distribution from sites 1 to 4

in Figure 1. Some populations in far western Michoacán (Fig. S1) are as yet

unsampled and could reveal further genetic structure.

Lineage 2 (dark blue in Fig. 1) – Morelos. Further sampling between sites 4 and 5 could

help determine whether the genetic distinctness of this individual is a true genetic

discontinuity or the result of a sampling gap.

Clade 2 – Members of this clade form the core *S. bistincta* and occur from Guerrero to Puebla

and Veracruz and south through Oaxaca.

Lineage 3 (red in Fig. 1) – Central and eastern Guerrero. Members of this lineage are

distinct from the pink lineage below and are monophyletic in the UCE trees (though

not in the mtDNA tree). Further sampling in between site 6 and site 7 would clarify

whether the genetic discontinuity between Lineages 3 and 4 results from a sampling
gap.

Lineage 4 (pink in Fig. 1) – Western Guerrero. This lineage is distinct and monophyletic
in all trees, although only a few individuals were sampled only from one locality.

Lineage 5 (orange and yellow in Fig. 1) – Puebla, Veracruz, and Oaxaca. The geographic
range of this group contains the type locality for *S. bistincta*. Central and southern
Oaxaca individuals (orange) are genetically distinct from individuals to the north
(yellow). One genetic intermediate in central Oaxaca suggests genetic continuity
across this range. One individual from southern Oaxaca (site 25 in Fig. 1) has an
unexpectedly northern genetic profile given its geographic location, suggesting a
more complex geographic structuring within this group than a simple cline from
north to south. An unsampled northern population of *S. bistincta* in Hidalgo is most
likely related to this lineage, and should be included in future studies.

Lineage 6 (purple in Fig. 1) – far northern Guerrero. Represented by a single individual,
this lineage is distinct and on a relatively long branch. However, in the mtDNA
tree, this individual is nested within Lineages 4 and 5 above. Sampling more
individuals is needed to determine how distinct this lineage might be.

Clade 3 – Pacific slope of Guerrero and Oaxaca. Sister to Clade 2, members of this clade
comprise two species, one already described and one currently being described.

Lineage 7 (dark gray in Fig. 1) – Pacific slope of Guerrero. This lineage was being
described as a new species (we call it *sp. nov.*) on the basis of phenotypic differences
before this genetic study was begun. Thus, our analysis lends support to species
status for this lineage.

<u>Lineage 8</u> (light gray in Fig. 1) – Pacific slope of Oaxaca. Genbank 16S mtDNA data and

423   the phenotype of the voucher reveal this lineage to be the already-described species,

424   *S. pentheter*, of southern and central Oaxaca.

The three clades and nearly all of the lineages were distinct in the mtDNA tree as well as

the UCE tree (Fig. 2). The mtDNA tree, however, supports a more complicated history for

Lineages 3, 4, and 6 in Clade 1 from Guerrero. It is unclear why UCE and mtDNA results

differed in this regard, but some reticulate processes might have influenced the mtDNA

genomes of these lineages, perhaps, given their close geographic proximity, mtDNA capture of

one lineage by another through ancient hybridization (e.g., Bryson Jr et al. 2010).

The phylogeny combining our mtDNA data with Genbank data (Fig. 4) is the most

taxonomically complete phylogeny of *Sarcohyla* and its sister genus *Plectrohyla* to date and

demonstrates the power of linking multilocus species delimitation with mtDNA data by

allowing delimited taxa to be placed in a broader context. This revealed that two already-

described species are nested within the current *S. bistincta*, lending support to species status of

the divergent Clade 1 from the Transvolcanic Belt (which was ~10% divergent in mtDNA from

S. pentheter). Additional insights afforded by the combined data mtDNA tree include support

for a previously hypothesized close relationship between *S. hazelae* and *S. thorectes* (Faivovich et

al. 2009), and a sister relationship between *S. mykter* and *S. chryses*. As the genus *Sarcohyla* is

very poorly represented by voucher specimens and DNA sequences (Fig. S1), a complete

understanding of the history of this genus must await more complete taxonomic and genomic

sampling. Unfortunately, there appear to be some microendemic *Sarcohyla* that might have

already gone extinct (Lips et al. 2004), especially in the Oaxacan highlands, although recent

resurveys give some cause for hope of rediscoveries (Delia et al. 2013).

By introducing UCEs as extended DNA barcodes, we are not advocating abandoning

traditional mtDNA barcoding and moving solely toward multilocus species delimitation. Even

though next-generation sequencing techniques are becoming more standardized and

widespread, they still pose a significant financial barrier and require more technical expertise

and equipment compared to mtDNA sequencing. However, there is no reason that nuclear

genomic DNA barcodes like UCEs cannot develop their own user community and databases in

parallel with mtDNA barcoding. As our study shows, mtDNA barcoding and multilocus

species delimitation are powerful when deployed together. Where they agree in this study is

that there is significant frog diversity in the Mexican Highlands that is currently unrecognized.

To the extent that these distinct lineages can be rigorously delimited and compared against

existing species, they can be added to biodiversity lists, allowing for their recognition and

protection.

Acknowledgments

458 We thank Greg Schneider and Ron Nussbaum at UMMZ for providing access to tissue samples

459 and specimens and Oscar Flores for help with permits. The project was funded by National

460 Science Foundation grant DEB-1258205 to JM.

Literature Cited

Bernt M, Donath A, Jühling F, Externbrink F, Florentz C, Fritzsch G, Pütz J, Middendorf M,
Stadler PF (2013) MITOS: Improved de novo metazoan mitochondrial genome
annotation. *Molecular Phylogenetics and Evolution*, **69**, 313-319.

Bickford D, Lohman DJ, Sodhi NS, Ng PK, Meier R, Winker K, Ingram KK, Das I (2007) Cryptic
species as a window on diversity and conservation. *Trends in Ecology & Evolution*, **22**,
148-155.

Bouckaert R, Heled J, Kühnert D, Vaughan T, Wu C-H, Xie D, Suchard MA, Rambaut A,
Drummond AJ (2014) BEAST 2: a software platform for Bayesian evolutionary analysis.
PLoS Computational Biology, **10**, e1003537.

Bryant D, Bouckaert R, Felsenstein J, Rosenberg NA, RoyChoudhury A (2012) Inferring species
trees directly from biallelic genetic markers: bypassing gene trees in a full coalescent
analysis. *Molecular Biology and Evolution*, **29**, 1917-1932.

Bryson Jr RW, De Oca ANM, Jaeger JR, Riddle BR (2010) Elucidation of cryptic diversity in a widespread Nearctic treefrog reveals episodes of mitochondrial gene capture as frogs diversified across a dynamic landscape. *Evolution*, **64**, 2315-2330.

Camargo A, Morando M, Avila LJ, Sites JW (2012) Species delimitation with ABC and other coalescent-based methods: a test of accuracy with simulations and an empirical example with lizards of the *Liolaemus darwinii* complex (Squamata: Liolaemidae). *Evolution*, **66**, 2834-2849.

Carstens BC, Pelletier TA, Reid NM, Satler JD (2013) How to fail at species delimitation. *Molecular Ecology*, **22**, 4369-4383.

Chevreux B, Wetter T, Suhai S (1999) Genome sequence assembly using trace signals and additional sequence information. *Computer Science and Biology: Proceedings of the German Conference on Bioinformatics*, **99**, 45-56.

Coissac E, Hollingsworth PM, Lavergne S, Taberlet P (2016) From barcodes to genomes: extending the concept of DNA barcoding. *Molecular Ecology*, **25**, 1423-1428.

Collins RA, Cruickshank RH (2014) Known knowns, known unknowns, unknown unknowns and unknown knowns in DNA barcoding: a comment on Dowton et al. *Systematic Biology*, **63**, 1005-1009.

Crawford NG, Faircloth BC, McCormack JE, Brumfield RT, Winker K, Glenn TC (2012) More than 1000 ultraconserved elements provide evidence that turtles are the sister group of archosaurs. *Biology Letters*, **8**, 783-786.

Cruaud A, Gautier M, Galan M, Foucaud J, Sauné L, Genson G, Dubois E, Nidelet S, Deuve T, Rasplus J-Y (2014) Empirical assessment of RAD sequencing for interspecific phylogeny. *Molecular Biology and Evolution*, **31**, 1272-1274.

DaCosta JM, Sorenson MD (2014) Amplification biases and consistent recovery of loci in a double-digest RAD-seq protocol. *PLoS One*, **9**, e106713.

Danecek P, Auton A, Abecasis G, Albers CA, Banks E, DePristo MA, Handsaker RE, Lunter G, Marth GT, Sherry ST (2011) The variant call format and VCFtools. *Bioinformatics*, **27**, 2156-2158.

De Queiroz K (2007) Species concepts and species delimitation. *Systematic Biology*, **56**, 879-886.

Delia JR, Whitney JL, Burkhardt T (2013) Rediscovery of 'lost' treefrogs from the Oaxacan highlands of Mexico. *Biodiversity and Conservation*, **22**, 1405-1414.

DePristo MA, Banks E, Poplin R, Garimella KV, Maguire JR, Hartl C, Philippakis AA, Del Angel G, Rivas MA, Hanna M (2011) A framework for variation discovery and genotyping using next-generation DNA sequencing data. *Nature Genetics*, **43**, 491-498.

DeSalle R, Amato G (2004) The expansion of conservation genetics. *Nature Reviews Genetics*, **5**,
702-712.

DeSalle R, Egan MG, Siddall M (2005) The unholy trinity: taxomony, species delimitation and
DNA barcoding. *Philosophical Transactions of the Royal Society B: Biological Sciences*, **360**,
1905-1916.

do Amaral FR, Neves LG, Resende Jr MF, Mobili F, Miyaki CY, Pellegrino KC, Biondo C (2015)
Ultraconserved elements sequencing as a low-cost source of complete mitochondrial
genomes and microsatellite markers in non-model amniotes. *PLoS One*, **10**, e0138446.

Dowton M, Meiklejohn K, Cameron SL, Wallman J (2014) A preliminary framework for DNA
barcoding, incorporating the multispecies coalescent. *Systematic Biology*, **63**, 639-644.

Duellman WE (2001) Hylid frogs of Middle America. Society for the Study of Amphibians and
Reptiles, Ithaca, New York.

Duellman WE, Marion AB, Hedges SB (2016) Phylogenetics, classification, and biogeography of
the treefrogs (Amphibia: Anura: Arboranae). *Zootaxa*, **4104**, 1-109.

Ebach MC, Holdrege C (2005) DNA barcoding is no substitute for taxonomy. *Nature*, **434**, 697-
697.

Edgar RC (2004) MUSCLE: multiple sequence alignment with high accuracy and high
throughput. *Nucleic Acids Research*, **32**, 1792-1797.

Edwards S, Bensch S (2009) Looking forwards or looking backwards in avian phylogeography?
A comment on. *Molecular Ecology*, **18**, 2930-2933.

Faircloth BC (2015) PHYLUCE is a software package for the analysis of conserved genomic loci.
Bioinformatics, **32**, 786-788.

Faircloth BC, Branstetter MG, White ND, Brady SG (2015) Target enrichment of ultraconserved
elements from arthropods provides a genomic perspective on relationships among
Hymenoptera. *Molecular Ecology Resources*, **15**, 489-501.

Faircloth BC, Glenn TC (2012) Not all sequence tags are created equal: designing and validating
sequence identification tags robust to indels. *PLoS One*, **7**, e42543.

Faircloth BC, McCormack JE, Crawford NG, Harvey MG, Brumfield RT, Glenn TC (2012)
Ultraconserved elements anchor thousands of genetic markers spanning multiple
evolutionary timescales. *Systematic Biology*, **61**, 717-726.

Faircloth BC, Sorenson L, Santini F, Alfaro ME (2013) A phylogenomic perspective on the
radiation of ray-finned fishes based upon targeted sequencing of ultraconserved
elements (UCEs). *PLoS One*, **8**, e65923.

Faivovich J, Haddad CF, Garcia PC, Frost DR, Campbell JA, Wheeler WC (2009) Systematic
review of the frog family Hylidae, with special reference to Hylinae: phylogenetic

analysis and taxonomic revision. *Bulletin of the American Museum of Natural History*, **294**,

1-240.

Fujita MK, Leaché AD, Burbrink FT, McGuire JA, Moritz C (2012) Coalescent-based species

delimitation in an integrative taxonomy. *Trends in Ecology & Evolution*, **27**, 480-488.

Galtier N, Nabholz B, Glémin S, Hurst G (2009) Mitochondrial DNA as a marker of molecular

diversity: a reappraisal. *Molecular Ecology*, **18**, 4541-4550.

Grummer JA, Bryson RW, Reeder TW (2014) Species delimitation using Bayes factors:

simulations and application to the *Sceloporus scalaris* species group (Squamata:

Phrynosomatidae). *Systematic Biology*, **63**, 119-133.

Haas BJ, Papanicolaou A, Yassour M, Grabherr M, Blood PD, Bowden J, Couger MB, Eccles D,

Li B, Lieber M (2013) De novo transcript sequence reconstruction from RNA-seq using

the Trinity platform for reference generation and analysis. *Nature Protocols*, **8**, 1494-1512.

Hahn C, Bachmann L, Chevreux B (2013) Reconstructing mitochondrial genomes directly from

genomic next-generation sequencing reads—a baiting and iterative mapping approach.

Nucleic Acids Research, **41**, e129.

Harris RS (2007) Improved pairwise alignment of genomic DNA. Pennsylvania State University,

PhD Thesis.

Hebert PD, Cywinska A, Ball SL (2003) Biological identifications through DNA barcodes.

Proceedings of the Royal Society of London B: Biological Sciences, **270**, 313-321.

Hebert PD, Penton EH, Burns JM, Janzen DH, Hallwachs W (2004) Ten species in one: DNA

barcoding reveals cryptic species in the neotropical skipper butterfly *Astraptes fulgerator*.

Proceedings of the National Academy of Sciences of the United States of America, **101**, 14812-

14817.

Hey J, Waples RS, Arnold ML, Butlin RK, Harrison RG (2003) Understanding and confronting

species uncertainty in biology and conservation. *Trends in Ecology & Evolution*, **18**, 597-

603.

Hime PM, Hotaling S, Grewelle RE, O'Neill EM, Voss SR, Shaffer HB, Weisrock DW (2016) The

influence of locus number and information content on species delimitation: an empirical

test case in an endangered Mexican salamander. *Molecular Ecology*, **25**, 5959-5974.

Isaac NJ, Mallet J, Mace GM (2004) Taxonomic inflation: its influence on macroecology and

conservation. *Trends in Ecology & Evolution*, **19**, 464-469.

Kearse M, Moir R, Wilson A, Stones-Havas S, Cheung M, Sturrock S, Buxton S, Cooper A,

Markowitz S, Duran C (2012) Geneious Basic: an integrated and extendable desktop

software platform for the organization and analysis of sequence data. *Bioinformatics*, **28**,

1647-1649.

Knowles LL, Carstens BC (2007) Delimiting species without monophyletic gene trees. *Systematic*
Biology, **56**, 887-895.

Lahaye R, Van der Bank M, Bogarin D, Warner J, Pupulin F, Gigot G, Maurin O, Duthoit S,
Barraclough TG, Savolainen V (2008) DNA barcoding the floras of biodiversity hotspots.
Proceedings of the National Academy of Sciences, **105**, 2923-2928.

Leaché AD, Fujita MK, Minin VN, Bouckaert RR (2014) Species delimitation using genome-wide
SNP data. *Systematic Biology*, **63**, 534-542.

Li H, Durbin R (2009) Fast and accurate short read alignment with Burrows–Wheeler transform.
Bioinformatics, **25**, 1754-1760.

Li H, Handsaker B, Wysoker A, Fennell T, Ruan J, Homer N, Marth G, Abecasis G, Durbin R
(2009) The sequence alignment/map format and SAMtools. *Bioinformatics*, **25**, 2078-2079.

Lips KR, Mendelson Iii JR, Munoz-Alonso A, Canseco-Márquez L, Mulcahy DG (2004)
Amphibian population declines in montane southern Mexico: resurveys of historical
localities. *Biological Conservation*, **119**, 555-564.

Mace GM (2004) The role of taxonomy in species conservation. *Philosophical Transactions of the*
Royal Society of London B: Biological Sciences, **359**, 711-719.

McCartney-Melstad E, Mount GG, Shaffer HB (2016) Exon capture optimization in amphibians
with large genomes. *Molecular Ecology Resources*, **16**, 1084-1094.

McCormack J, Tsai WLE, Faircloth BC (2016) Sequence capture of ultraconserved elements from
bird museum specimens. *Molecular Ecology Resources*, **16**, 1189-1203.

McCormack JE, Faircloth BC, Crawford NG, Gowaty PA, Brumfield RT, Glenn TC (2012)
Ultraconserved elements are novel phylogenomic markers that resolve placental
mammal phylogeny when combined with species-tree analysis. *Genome Research*, **22**,
746-754.

McKenna A, Hanna M, Banks E, Sivachenko A, Cibulskis K, Kernytsky A, Garimella K,
Altshuler D, Gabriel S, Daly M (2010) The Genome Analysis Toolkit: a MapReduce
framework for analyzing next-generation DNA sequencing data. *Genome Research*, **20**,
1297-1303.

Miralles A, Vences M (2013) New metrics for comparison of taxonomies reveal striking
discrepancies among species delimitation methods in Madascincus lizards. *PLoS One*, **8**,
e68242.

Moritz C, Cicero C (2004) DNA barcoding: promise and pitfalls. *PLoS Biology*, **2**, e354.

O'Meara BC (2009) New heuristic methods for joint species delimitation and species tree
inference. *Systematic Biology*, **59**, 59-73.

Olmo E (1991) Genome variations in the transition from amphibians to reptiles. *Journal of*
Molecular Evolution, **33**, 68-75.

Pie MR, Bornschein MR, Ribiero LF, Faircloth BC, McCormack JE (2017) Phylogenomic species
delimitation in microendemic frogs Of the Brazilian Atlantic forest. *BioRxiv*, doi:
https://doi.org/10.1101/143735.

Pons J, Barraclough TG, Gomez-Zurita J, Cardoso A, Duran DP, Hazell S, Kamoun S, Sumlin
WD, Vogler AP (2006) Sequence-based species delimitation for the DNA taxonomy of
undescribed insects. *Systematic Biology*, **55**, 595-609.

Pritchard JK, Stephens M, Donnelly P (2000) Inference of population structure using multilocus
genotype data. *Genetics*, **155**, 945-959.

Simpson JT, Wong K, Jackman SD, Schein JE, Jones SJ, Birol I (2009) ABySS: a parallel assembler
for short read sequence data. *Genome Research*, **19**, 1117-1123.

Smith BT, Harvey MG, Faircloth BC, Glenn TC, Brumfield RT (2014) Target capture and
massively parallel sequencing of ultraconserved elements for comparative studies at
shallow evolutionary time scales. *Systematic Biology*, **63**, 83-95.

Stamatakis A (2014) RAxML version 8: a tool for phylogenetic analysis and post-analysis of
large phylogenies. *Bioinformatics*, **30**, 1312-1313.

Starrett J, Derkarabetian S, Hedin M, Bryson RW, McCormack JE, Faircloth BC (2016) High
phylogenetic utility of an ultraconserved element probe set designed for Arachnida.
Molecular Ecology Resources, DOI: 10.1111/1755-0998.12621.

Stuart SN, Hoffmann M, Chanson J, Cox N, Berridge R, Ramani P, Young B (2008) Threatened
amphibians of the world. Lynx Edicions, Barcelona, Spain.

Sukumaran J, Knowles LL (2017) Multispecies coalescent delimits structure, not species.
Proceedings of the National Academy of Sciences, **114**, 1607-1612.

Taberlet P, Coissac E, Pompanon F, Gielly L, Miquel C, Valentini A, Vermat T, Corthier G,
Brochmann C, Willerslev E (2007) Power and limitations of the chloroplast trnL (UAA)
intron for plant DNA barcoding. *Nucleic Acids Research*, **35**, e14-e14.

Tautz D, Arctander P, Minelli A, Thomas RH, Vogler AP (2003) A plea for DNA taxonomy.
Trends in Ecology & Evolution, **18**, 70-74.

van der Auwera GA, Carneiro MO, Hartl C, Poplin R, del Angel G, Levy-Moonshine A, Jordan
T, Shakir K, Roazen D, Thibault J (2013) From FastQ data to high-confidence variant
calls: the genome analysis toolkit best practices pipeline. *Current Protocols in*
Bioinformatics, **11: 11.10**, 11.10. 11-11.10. 33.

Ward RD, Zemlak TS, Innes BH, Last PR, Hebert PD (2005) DNA barcoding Australia's fish
species. *Philosophical Transactions of the Royal Society of London B: Biological Sciences*, **360**,
1847-1857.

Wiens JJ (2007) Species delimitation: new approaches for discovering diversity. *Systematic*
Biology, **56**, 875-878.
Will KW, Mishler BD, Wheeler QD (2005) The perils of DNA barcoding and the need for
integrative taxonomy. *Systematic Biology*, **54**, 844-851.
Witt JD, Threloff DL, Hebert PD (2006) DNA barcoding reveals extraordinary cryptic diversity
in an amphipod genus: implications for desert spring conservation. *Molecular Ecology*,
15, 3073-3082.
Yang Z, Rannala B (2010) Bayesian species delimitation using multilocus sequence data.
Proceedings of the National Academy of Sciences, **107**, 9264-9269.
Ye L, Zhu C, Yu D, Zhang Y, Zhang J (2016) The complete mitochondrial genome of *Hyla*
annectans (Anura: Hylidae). *Mitochondrial DNA Part A*, **27**, 1593-1594.
Zarza E, Faircloth BC, Tsai WL, Bryson RW, Klicka J, McCormack JE (2016) Hidden histories of
gene flow in highland birds revealed with genomic markers. *Molecular Ecology*, **25**, 5144-
5157.

Tables

Table 1. Specimen information and summary statistics for *Sarcohyla bistincta* and closely related species.

Map Number[1]	Field Number[2]	Catalog Number[3]	Current Taxonomy	State	Latitude	Longitude	Trimmed Paired Reads	UCEs	Average UCE Length	SNP Read Depth	mtDNA Reads	mtDNA Average Coverage	mtDNA Average Quality
1	MK 618	UMMZ 239796	*S. bistincta*	MICHOACAN	19.7911	-100.6605	1,889,670	2,323	529.7	31.8	3,381	20.8	86
2	MK 627-31	UMMZ 239683	*S. bistincta*	MICHOACAN	19.4266	-102.0736	1,933,019	2,254	370.9	21.6	3,415	20.8	85
3	MK 666	UMMZ 239745	*S. bistincta*	MICHOACAN	19.3452	-100.3128	2,186,871	2,350	621.3	33.9	3,900	23.4	83
4	MK 600	UMMZ 239679	*S. bistincta*	MEXICO	19.1501	-100.1469	932,448	2,174	500.8	24.4	2,377	15.7	83
4	MK 600 (1)	UMMZ 239678	*S. bistincta*	MEXICO	19.1501	-100.1469	1,217,987	2,267	532.8	21.9	2,570	16.4	84
5	MK 645	UMMZ 239749	*S. bistincta*	MORELOS	18.9224	-99.2442	2,217,054	2,406	532.5	34.2	3,494	21.8	86
5	MK 645 dup	UMMZ 239749	*S. bistincta*	MORELOS	18.9224	-99.2442	548,545	1,964	513.0	17.7	841	7.4	60
6	MK 759	UMMZ 239701	*S. bistincta*	GUERRERO	18.0013	-101.1716	1,224,835	2,199	557.6	25.0	2,522	16.2	84
6	MK 760	UMMZ 239705	*S. bistincta*	GUERRERO	18.0013	-101.1716	1,074,315	2,203	559.4	24.3	743	6.9	58
6	MK 760 (2)	UMMZ 239704	*S. bistincta*	GUERRERO	18.0013	-101.1716	927,774	2,246	579.2	26.3	906	9.4	6
7	MK 691 (5)	UMMZ 239744	*S. bistincta*	GUERRERO	17.5324	-99.8994	2,084,203	2,353	672.5	27.2	1,678	11.9	80
8	MK 650 (1)	UMMZ 239725	*S. bistincta*	GUERRERO	17.6843	-99.8034	941,101	2,124	556.6	22.6	1,278	9.7	74
8	MK 650 (2)	UMMZ 239726	*S. bistincta*	GUERRERO	17.6843	-99.8034	2,224,898	2,394	506.4	28.4	1,737	12.2	79
8	MK 652	UMMZ 239727	*S. bistincta*	GUERRERO	17.6843	-99.8034	3,423,330	2,444	526.0	29.8	1,396	10.4	74
9	MK 671 (4)	UMMZ 239733	*S. bistincta*	GUERRERO	17.6407	-99.6797	1,012,300	2,107	557.0	22.5	842	7.4	62
9	MK 672	UMMZ 239738	*S. bistincta*	GUERRERO	17.6407	-99.6797	297,782	1,667	439.8	11.9	400	5.1	41
10	MK 656 (1)	UMMZ 239729	*S. bistincta*	GUERRERO	17.5526	-99.6626	580,194	1,950	497.6	18.4	547	5.9	42
11	MK 674 (1)	UMMZ 239690	*S. bistincta*	GUERRERO	17.5087	-99.1258	659,418	1,941	565.4	18.8	1,086	8.7	68
11	MK 675 (2)	UMMZ 239879	*S. bistincta*	GUERRERO	17.5087	-99.1258	336,474	1,565	519.1	10.4	227	4.2	24
12	MK 662	UMMZ 239731	*S. bistincta*	GUERRERO	18.6359	-99.6491	1,137,742	2,130	548.7	23.5	796	7.2	56
13	MK 697 (3)	UMMZ 239789	*S. bistincta*	VERACRUZ	18.6585	-97.1574	1,524,182	2,214	625.6	24.3	2,479	16.0	82
13	MK 699 (1)	UMMZ 239791	*S. bistincta*	VERACRUZ	18.6477	-97.1574	1,342,010	2,206	610.9	24.4	1,769	12.3	79
14	MK 700 (2)	UMMZ 239750	*S. bistincta*	PUEBLA	18.3220	-97.0285	2,203,360	2,298	697.8	27.6	4,155	24.8	86
15	MK 705 (1)	UMMZ 239862	*S. bistincta*	OAXACA	18.1576	-96.8684	2,529,703	2,417	644.6	30.3	3,164	19.6	85
16	MK 715	UMMZ 239755	*S. bistincta*	OAXACA	17.2390	-97.0032	347,152	1,712	437.7	12.1	254	4.3	31
17	MK 716 (1)	UMMZ 239758	*S. bistincta*	OAXACA	17.3036	-96.7930	52,461	594	267.4	2.7	8	3.0	1
18	MK 718 (2)	UMMZ 239765	*S. bistincta*	OAXACA	17.4211	-96.6876	606,978	2,014	568.6	19.7	366	4.9	37
19	MK 755 (1)	UMMZ 239786	*S. bistincta*	OAXACA	17.4153	-96.5671	871,034	2,062	563.2	22.0	1,022	8.4	69
20	MK 751	UMMZ 239785	*S. bistincta*	OAXACA	17.3160	-96.4435	1,577,190	2,342	621.5	28.5	2,376	15.4	84
22	MK 748 (2)	UMMZ 239780	*S. bistincta*	OAXACA	16.9791	-96.1364	91,912	1,009	302.5	4.2	23	3.1	3
22	MK 748 (4)	UMMZ 239782	*S. bistincta*	OAXACA	16.9791	-96.1364	1,058,426	2,170	553.8	23.7	499	5.7	42
22	MK 767	UMMZ 239788	*S. bistincta*	OAXACA	16.9859	-96.1358	1,083,058	2,206	524.7	24.3	617	6.3	52
23	MK 721	UMMZ 239767	*S. bistincta*	OAXACA	16.7377	-97.0384	146,471	1,215	376.0	6.1	62	3.3	10
24	MK 766	UMMZ 239794	*S. bistincta*	OAXACA	16.2522	-97.1536	994,521	2,166	577.4	23.3	467	5.5	43
25	MK 730 (2)		*P. bistincta*	OAXACA	16.0186	-96.5301	2,967,630	2,404	716.9	31.1	3,581	21.7	84
26	MK 685 (2)	UMMZ 239739	*S. sp. nov.*	GUERRERO	17.3812	-100.2009	269,729	1,645	450.6	10.8	184	4.0	20
27	MK 689 (2)	UMMZ 239740	*S. sp. nov.*	GUERRERO	17.3000	-100.2792	67,313	502	221.6	2.6	32	3.1	6
28	MK 727 (2)	UMMZ 239772	*S. pentheter*	OAXACA	16.1916	-97.0958	17,052	381	247.1	2.4	10	3.0	2
29	MK 691 (3)	UMMZ 239651	*S. chryses*	GUERRERO	17.5324	-99.8994	2,267,946	2,347	663.6	35.0	4,919	33.4	83
30	MK 770	UMMZ 239802	*S. hazelae*	OAXACA	17.2216	-96.5839	538,500	2,066	526.2	19.8	557	5.9	44

[1] Map number in Figure 1
[2] The first three-digit number corresponds to a sampling location. If there is a second number in parentheses, this corresponds to different aquaria where tadpoles were sorted by species before
[3] All specimens are from the University of Michigan Museum of Zoology

Table 2. Bayes factor species delimitation results.

Model	Species	MLE	Rank	*ln*(BF)
runA	9	-2832.33	1	10.52
runB	10	-2837.59	2	132.52
runC	7	-2903.85	3	69.36
runD	6	-2938.53	4	3822
runE	2	-4849.53	5	-

Figure 1. (A) Map of central Mexico showing sampling sites for *S. bistincta* and close outgroups,

with numbers matching localities listed in Table 1 and colors matching Structure results below.

Unsampled parts of the distribution of *S. bistincta* are shown in Fig. S1; (B) *S. bistincta* individual

from near site 1; (C) Composite results of repeated Structure runs at *K* = 2 showing the finest

detectable structure in the genetic data. Each vertical line represents an individual labeled with

its UMMZ catalog number above and, in descending order below, the site number and the

lineages indicated by the species delimitation analysis.

Figure 2. (A) UCE tree; (B) mtDNA tree. Colors match Structure groups identified in Fig. 1. Tips

are labeled with their UMMZ catalog number.

Figure 3. Cloudogram of the posterior distribution of SNAPP trees from 399 high-quality SNPs

mined from UCE loci. Tip labels are UMMZ catalog numbers. Colored bars show the eight

lineages identified by the species delimitation analysis in addition to the outgroup *S. chryses*.

Clades are discussed in text. Colors match genetic clusters from Figure 1.

Figure 4. Phylogeny of mtDNA for a subset of the individuals from this study (labeled with red

UMMZ catalog numbers) combined with existing sequences from Genbank (labeled with

accession numbers). Colored boxes relate to genetic lineages in prior figures. The tree was

rooted with *Exerodonta xera*. Photo credit: Peter Heimes.

Table S1. Information and summary statistics on all 45 samples used to determine the ingroup

for this study.

Field Number	UMMZ Number	Current Taxonomy	State	Locality	Latitude	Longitude	Trimmed Paired Reads	UCEs	Average UCE length	mtDNA reads	mtDNA average coverage	mtDNA average quality	Fig. 1 Map Number	Accession number
MK 618	UMMZ 239796	Sarcohyla bistincta	MICHOACAN	LOS AZUFRES / SAN PEDRO ROAD	19.791051	-100.660542	1,889,670	2323	529.7	3381	20.8	86	1	
MK 627-31	UMMZ 239683	Sarcohyla bistincta	MICHOACAN	PARQUE BARRANCA DEL CUPATITZIO URUAPAN	19.426646	-102.073574	1,933,019	2254	370.9	3415	20.8	85	2	
MK 666	UMMZ 239745	Sarcohyla bistincta	MICHOACAN	13 KMS ROAD ZITACUARO MACHERO (LITTLE TOWN WERE THERE IS A MONARCH BUTTERFLY SANCTUARY)	19.3452	-100.3128	2,186,871	2350	621.3	3900	23.4	83	3	
MK 600	UMMZ 239679	Sarcohyla bistincta	MEXICO	CARRETERA VALLE DE BRAVO-SAN PEDRO TENEYAC ARROYO 1 (6860 FEET) (POR ALBARRADA) BOSQUE BIEN CONSERVADO Y HUMEDO	19.150083	-100.1469	932,448	2174	500.8	2377	15.7	83	4	
MK 600 (1)	UMMZ 239678	Sarcohyla bistincta	MEXICO	CARRETERA VALLE DE BRAVO-SAN PEDRO TENEYAC ARROYO 1 (6860 FEET) (POR ALBARRADA) BOSQUE BIEN CONSERVADO Y HUMEDO	19.150083	-100.1469	1,217,987	2267	532.8	2570	16.4	84	4	
MK 645	UMMZ 239749	Sarcohyla bistincta	MORELOS	2 KMS OUT OF CUERNAVACA ON THE CUERNAVACA CHALMA RD. (PASSING CONJUNTO CERRADO) OF LA BARRANCA WHERE THE SALTO SAN ANTONIO IS	18.922402	-99.244151	2,217,054	2406	532.5	3494	21.8	86	5	
MK 645 dupe	UMMZ 239749	Sarcohyla bistincta	MORELOS	2 KMS OUT OF CUERNAVACA ON THE CUERNAVACA CHALMA RD. (PASSING CONJUNTO CERRADO) OF LA BARRANCA WHERE THE SALTO SAN ANTONIO IS	18.922402	-99.244151	548,585	1964	513.0	841	7.4	60	5	
MK 759	UMMZ 239701	Sarcohyla bistincta	GUERRERO	CA 100 DEG KMS FROM CIUDAD ALTAMIRANO VIA IXTAPA ZIHUATANEJO	18.0013	-101.1716	1,224,835	2199	557.6	2522	16.2	84	6	
MK 760	UMMZ 239705	Sarcohyla bistincta	GUERRERO	CA 100 DEG KMS FROM CIUDAD ALTAMIRANO VIA IXTAPA ZIHUATANEJO	18.0013	-101.1716	1,074,315	2203	559.4	743	6.9	58	6	
MK 760 (2)	UMMZ 239704	Sarcohyla bistincta	GUERRERO	CA 100 DEG KMS FROM CIUDAD ALTAMIRANO VIA IXTAPA ZIHUATANEJO	18.0013	-101.1716	927,774	2246	579.2	906	9.4	6	6	
MK 691 (5)	UMMZ 239744	Sarcohyla bistincta	GUERRERO	2-3 KMS IN THE ROAD TO JALEACA FROM POINT WHERE ROAD TRIFURCATE TO PTO. DEL GALLO / YEXTLA / ANDJALEACA.COMMING FROM CARRIZAL DE BRAVO / IN RIVER UNDER THE BRIDGE	17.5324	-99.8994	2,084,203	2353	672.5	1678	11.9	80	7	
MK 650 (1)	UMMZ 239725	Sarcohyla bistincta	GUERRERO	MOUNTAINS W. OF CHILPANCINGO / TOWN LOS MORROS	17.6843	-99.80339	941,101	2124	556.6	1278	9.7	74	8	
MK 650 (2)	UMMZ 239726	Sarcohyla bistincta	GUERRERO	MOUNTAINS W. OF CHILPANCINGO / TOWN LOS MORROS	17.6843	-99.80339	2,224,898	2394	506.4	1737	12.2	79	8	
MK 652	UMMZ 239727	Sarcohyla bistincta	GUERRERO	MOUNTAINS W. OF CHILPANCINGO / TOWN LOS MORROS	17.6843	-99.80339	3,423,330	2444	526.0	1396	10.4	74	8	
MK 671 (4)	UMMZ 239733	Sarcohyla bistincta	GUERRERO	BEHIND CHICHIHUALCO / ON ROAD TO CARRIZAL DE BRAVO / 2 KMS FROM ENTRONQUE	17.6407	-99.6797	1,012,300	2107	557.0	842	7.4	62	9	
MK 672	UMMZ 239738	Sarcohyla bistincta	GUERRERO	BEHIND CHICHIHUALCO / ON ROAD TO CARRIZAL DE BRAVO / 2 KMS FROM ENTRONQUE	17.6407	-99.6797	297,782	1667	439.8	400	5.1	41	9	
MK 656 (1)	UMMZ 239729	Sarcohyla bistincta	GUERRERO	3 KMS FROM THE TOWN OF OMILTEMI IN THE ROAD OMILTEMI CHILPANCINGO	17.552603	-99.662569	580,194	1950	497.6	547	5.9	42	10	
MK 674 (1)	UMMZ 239690	Sarcohyla bistincta	GUERRERO	ON ATZACUALOYA HUEYCATENANGO RD.	17.5087	-99.1258	659,418	1941	565.4	1086	8.7	68	11	
MK 675 (2)	UMMZ 239879	Sarcohyla bistincta	GUERRERO	ON ATZACUALOYA HUEYCATENANGO RD.	17.5087	-99.1258	336,474	1565	519.1	227	4.2	24	11	
MK 662	UMMZ 239731	Sarcohyla bistincta	GUERRERO	1.6 KMS FROM THE TOWN OF TETIPAC ON THE TETIPAC TAXCO ROAD (MAYBE ARROYO LAS DAMAS)	18.635895	-99.6491	1,137,742	2130	548.7	796	7.2	56	12	
MK 697 (3)	UMMZ 239789	Sarcohyla bistincta	VERACRUZ	ON ATZOMPA XOXOCOTLA RD. 1.5 KMS FROM XOXOCOTLA	18.6585	-97.1574	1,524,182	2214	625.6	2479	16.0	82	13	
MK 699 (1)	UMMZ 239791	Sarcohyla bistincta	VERACRUZ	ON STREAM CROSSING THE TOWN OF XOXOCOTLA	18.6477	-97.1574	1,342,010	2206	610.9	1769	12.3	79	13	
MK 700 (2)	UMMZ 239750	Sarcohyla bistincta	PUEBLA	IN THE STREAM LOCATED AFTER THE TOWN OF ZOQUITLAN TURNING DOWN AT THE CENTRO DE SALUD	18.322	-97.0285	2,203,360	2298	697.8	4155	24.8	86	14	
MK 705 (1)	UMMZ 239862	Sarcohyla bistincta	OAXACA	56 KMS FROM TEOTITLAN VIA HUAUTLA	18.1576	-96.8684	2,529,703	2417	644.6	3164	19.6	85	15	
MK 715	UMMZ 239755	Sarcohyla bistincta	OAXACA	NEAR "EL TEJOCOTE" (FRENTE AL KINDER)	17.239	-97.0032	347,152	1712	437.7	254	4.3	31	16	
MK 716 (1)	UMMZ 239758	Sarcohyla bistincta	OAXACA	ON ROAD BETWEEN THE TOWNS OF SAN JUAN DEL ESTADO AND SAN MIGUEL ALOAPAN	17.3036	-96.793	52,461	594	267.4	8	3.0	1	17	
MK 718 (2)	UMMZ 239765	Sarcohyla bistincta	OAXACA	PASSING SAN MIGUEL ALOAPAN	17.4211	-96.6876	606,978	2014	568.6	366	4.9	37	18	
MK 755 (1)	UMMZ 239786	Sarcohyla bistincta	OAXACA	ON ROAD SAN JUAN ATEPEC SAN MIGUEL ABEJONES	17.4153	-96.5671	871,034	2062	563.2	1022	8.4	69	19	
MK 751	UMMZ 239785	Sarcohyla bistincta	OAXACA	3.8 KMS PASSING "RANCHO TEXAS" ON THE ROAD FROM THE TOWN OF IXTLAN DE JUAREZ	17.316	-96.4435	1,577,190	2342	621.5	2376	15.4	84	20	
MK 748 (2)	UMMZ 239780	Sarcohyla bistincta	OAXACA	CA 37 KMS FROM MITLA ON THE ROAD MITLA AYUTLA	16.9791	-96.1364	91,912	1009	302.5	23	3.1	3	22	
MK 748 (4)	UMMZ 239782	Sarcohyla bistincta	OAXACA	CA 37 KMS FROM MITLA ON THE ROAD MITLA AYUTLA	16.9791	-96.1364	1,058,426	2170	553.8	499	5.7	42	22	
MK 767	UMMZ 239788	Sarcohyla bistincta	OAXACA	9.2 KMS E STA MARIA ALBARRADAS / SIERRA MIXE	16.985888	-96.135816	1,055,985	2206	524.7	617	6.3	52	22	
MK 721	UMMZ 239767	Sarcohyla bistincta	OAXACA	2 KMS N. THE TOWN OF STA. MARIA LAXICHIO VIA THE TOWN OF SAN SEBASTIAN RIO DULCE	16.7377	-97.0384	146,471	1215	376.0	62	3.3	10	23	
MK 766	UMMZ 239794	Sarcohyla bistincta	OAXACA	CERRO DE VIDRIO VIA A PUERTO ESCONDIDO	16.25216	-97.15359	994,521	2166	577.4	467	5.5	43	24	
MK 730 (2)		Sarcohyla bistincta	OAXACA	SLIGHTLY NORTH OF THE TOWN JALATENGO ON HWY 175	16.0186	-96.5301	2,967,630	2404	716.9	3581	21.7	84	25	
MK 685 (2)	UMMZ 239739	Sarcohyla sp. nov.	GUERRERO	ON ATOYAC PTO DE GALLO RD / BETWEEN 10 - 20 KMS NORTH OF THE TOWN "EL PARAISO"	17.3812	-100.2009	269,729	1645	450.6	184	4.0	20	26	
MK 689 (2)	UMMZ 239740	Sarcohyla sp. nov.	GUERRERO	ON ATOYAC PTO DE GALLO RD. / 500 M. NORTH OF THE TOWN OF SAN VICENTE	17.3	-100.2792	67,313	502	221.6	32	3.1	6	27	
MK 727 (2)	UMMZ 239772	Sarcohyla pentheter	OAXACA	RIO "EL SALADO" / 8 KMS N SAN JUAN LACHAO ON HWY 135	16.1916	-97.0958	17,052	381	247.1	10	3.0	2	28	
MK 691 (3)	UMMZ 239651	Plectrohyla chryses	GUERRERO	2-3 KMS IN THE ROAD TO JALEACA FROM POINT WHERE ROAD TRIFURCATE TO PTO. DEL GALLO / YEXTLA / ANDJALEACA.COMMING FROM CARRIZAL DE BRAVO / IN RIVER UNDER THE BRIDGE	17.5324	-99.8994	2,267,946	2347	663.6	4919	33.4	83	29	
MK 770	UMMZ 239802	Plectrohyla hazelae	OAXACA	EL PUNTO SIERRA JUAREZ	17.22156	-96.58386	538,500	2066	526.2	557	5.9	44	30	
MK 667	UMMZ 239952	Sarcohyla arborescandens	VERACRUZ	PUERTO DEL AIRE (ARRIBA DE ALCUTZINGO)	18.6787	-97.3485	923,330	2116	511.8	1154	9.0	69		
MK 700 (1)	UMMZ 239813	Sarcohyla cyclada	PUEBLA	IN THE STREAM LOCATED AFTER THE TOWN OF ZOQUITLAN TURNING DOWN AT THE CENTRO DE SALUD	18.322	-97.0285	2,490,616	2385	661.4	2961	18.4	82		
MK 701	UMMZ 239814	Sarcohyla cyclada	OAXACA	24 KMS FROM THE TOWN OF TEOTITLAN DE FLORES MAGON VIA HUAUTLA DE JIMENEZ	18.1781	-97.0054	1,166,574	2129	565.5	1061	8.7	69		
MK 742 (1)	UMMZ 239954	Sarcohyla arborescandens	VERACRUZ	LEFT ROAD BIFURCATING FROM ROAD TO THE TOWN OF "LAS MINAS"	19.6758	-97.1751	1,074,225	2240	553.8	1750	12.0	79		
MK 768	UMMZ 239833	Exerodonta xera	PUEBLA	5 KM SW ZAPOTITLAN DE SALINAS	18.311958	-97.51266	1,257,938	2288	581.5	3506	21.4	84		

697   Fig. S1. Sampled and unsampled parts of *S. bistincta* range in relation to known distributions (or

698   localities, where distributional information is lacking) of other *Sarcohyla* species.

700   Fig. S2. UCE tree of 45 samples of *Sarcohyla* and outgroup *Exerodonta xera* used to determine the

701   ingroup.