

Honey Bee Diversity is Swayed by Migratory Beekeeping and Trade Despite Conservation Practices: Genetic Evidences for the Impact of Anthropogenic Factors on Population Structure

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11 Abstract

12 Intense admixture of honey bee (*Apis mellifera* L.) populations is mostly attributed to migratory
13 beekeeping practices and replacement of queens and colonies with non-native races or hybrids of
14 different subspecies. These two practices are also heavily carried out in Anatolia and Thrace where 5
15 subspecies reside naturally.

16 Here, we carried out an analysis of population structure of honey bees sampled from six different
17 regions (n = 250) in order to test the genetic impacts of migratory beekeeping, queen and colony
18 trade and conservation efficacy of isolated regions. A total of 30 microsatellite markers were used in
19 four multiplex reactions.

20 Direct genetic impact of migratory beekeeping was demonstrated first time based on a comparison of
21 assignment of individuals to their geographically native populations where migratory colonies
22 showed less fidelity. We found genetic evidence for them acting as a hybrid zone mobile in space and
23 time, becoming vectors of otherwise local gene combinations.

24 The effects of honey bee trade were revealed by the presence of very high introgression levels from
25 the highly commercial Caucasian bees naturally limited to a narrow range. We also measured the
26 direction and magnitude of this gene flow connected with bee trade.

27 Comparison between regions that are either open to migratory beekeeping or not let us evaluate the
28 status of isolated regions as centers of limited gene flow and showed the importance of establishing
29 such regions.

30 Despite signs of gene flow, our findings confirm high levels of geographically structured genetic
31 diversity of four subspecies of honey bees in Turkey and emphasize the need to develop policies to
32 maintain this diversity.

33 Our overall results might potentially bear a wider interest to the community since they constitute an
34 important attempt to quantify the effects of anthropogenic impacts on established patterns of honey
35 bee diversity. Our measurable and justified findings on migratory beekeeping, queen and colony
36 replacements as well as conservation implications will hopefully be of use for the decision makers
37 and other stakeholders.

38 **Introduction**

39 The Western honey bee, *Apis mellifera* L., is a species which plays role together with other
40 pollinators in pollination of wild and cultivated plants while the species also have significant
41 economic importance in terms of honey and other bee products output (Morse 1991; Breeze et al.
42 2011). In addition to its ecological and economic importance, it is a model study organism both for
43 evolution of eusociality and sophisticated cognitive abilities (Weinstock et al. 2006).

44 Natural distribution of *A. mellifera* includes Central and Southwest Asia, Europe and Africa but the
45 species was also introduced to East and Southeast Asia, Australia and the Americas mainly on
46 purpose for its economic benefits (Ruttner 1988). Morphological and molecular studies point to four
47 major lineages of numerous –more than 20- subspecies (Ruttner, 1988; Whitfield et al. 2006). The
48 four widely recognized lineages are A (Africa), M (western and northern Europe), O (Near East and
49 Central Asia) and C (Eastern Europe) lineages.

50 Although bearing controversies, studies with Single Nucleotide Polymorphisms (SNPs) in the past
51 decade supported the hypothesis that *A. mellifera* have originated in the tropics or subtropics in
52 Africa and colonized its natural range by two main routes: one through Gibraltar and one through
53 Suez and then Bosphorus, ending up with a secondary contact between the highly divergent A and C
54 lineages around Alps (Whitfield et al. 2006; Han et al. 2012; Walberg et al. 2014; Harpur et al. 2014;
55 Cridland et al. 2017).

56 Both the honey bees and wild pollinators are thought to be on decline (locally and/or globally
57 depending on the species and region of concern) due to factors some of them relating closely to
58 human activities. Among them, destruction and fragmentation of natural habitats, toxicity caused by
59 pollution and pesticides –as such widely used neonicotinoids-, diseases and their spread getting
60 easier, invasive species are leading the way (Meffe 1998; Brown & Paxton 2009; Van Engelsdorp &
61 Meixner 2010; Blacquiere et al. 2012). Honey bees also, especially wild populations that are not
62 managed by beekeepers (including the feral populations), take their share from the situation like the
63 other species in the genus *Apis* –namely *Apis cerana*, *Apis florea*, *Apis dorsata* and other native bees
64 of Asia (Oldroyd 2007; Dietemann et al. 2009; Van Engelsdorp et al. 2009; Genersch 2010; Evans &
65 Schwarz 2011).

66 Besides such negative consequences created by human activities; the genetic admixture of honey bee
67 populations due to bee trade, including complete replacement of local bees with non-natives and
68 beekeeping practices involving movement of colonies from one region to the other impose another
69 kind of pressure on the species: the loss and/or swamping of locally adapted gene combinations and
70 local or global extinctions of native honey bees (De la Rua et al. 2009).

71 All those factors and their interactions, including genetic and environmental ones, when combined,
72 may have an increased adverse effect on honey bees and may be the reasons behind continuous or
73 discrete events of sudden colony losses with rapid depletion of worker bees while the queen
74 continues laying eggs accompanied by lack of dead bees in and around the hive; the syndrome called

75 as Colony Collapse Disorder (CCD) or Colony Depopulation Syndrome (CDS) (Van Engelsdorp et
76 al. 2009; Neumann & Carreck 2010).

77 Resilience of the honey bees may be lying in the adaptations they accumulated over thousands of
78 years, and new potentials reside in their genetic diversity. It is highly probable that a combination of
79 many above mentioned factors/threats are taking their places in the recent declines by weakening the
80 colonies step by step. Due to altered rankings of performance of subspecies in varying environments,
81 it is generally accepted that honey bees' resistance or tolerance to these factors differ greatly and
82 locally adapted variants may be encountering less stress, thus remain standing strong (Büchler et al.
83 2015). Hence, research on honey bee diversity in the global context and at various levels (genetic,
84 individuals, colonies, populations, ecotypes and subspecies) is of great importance for maintaining
85 the species' and ecosystem services they provide as well as their economic usefulness.

86 In recent years' research conducted on honey bee population structure in European countries, it was
87 shown that the past structure was lost or strongly disturbed (Dall'Olio et al. 2007; Canovas et al.
88 2011; Bouga et al. 2011). Introgression of non-native DNA was monitored in wild populations of
89 Sudan (El-Niweiri & Moritz 2010). Among the anthropogenic effects, mainly queen and colony trade
90 and replacement of native honey bees with non-natives as well as migratory beekeeping were the
91 usual suspects.

92 Despite grounded suspicions there are very few studies that investigate and test the direct genetic
93 consequences of human practices on honey bee diversity. Therefore, the aims of this series of
94 experiments were testing different hypotheses about recent heavy/any admixture of honey bee
95 populations across four subspecies by making use of microsatellite markers as well as i) evaluating
96 the status of isolated regions as a conservation implication where migratory beekeeping is prohibited,
97 restricted or very scarce due to lack of preference of migratory beekeepers or attitude of local
98 beekeepers ii) acquiring and demonstrating the direct genetic outcomes of migratory beekeeping by a
99 series of comparisons between migratory and stationary colonies iii) seeking for the effects of
100 unregulated queen and colony trade by figuring out the origin, extent and direction of introgression
101 between populations.

102 With five subspecies dwelling within its borders and with a variety of beekeeping strategies, Turkey
103 makes a good stage for chasing genetic evidences for the impact of anthropogenic factors on one of
104 the most important crop and wild plant pollinators. Beekeeping is an old tradition in those lands
105 which dates back to 6600 BC and Hittites civilization (Akkaya & Serhat 2007), while still intensively
106 practiced in Turkey where there are more than 8 million hives distributed all over the country. This is
107 the third highest number in the world, alone tripling those of the USA and reaching the half of the EU
108 total (USDA NASS 2019, European Parliament 2017).

109 Corresponding to one-fourth to one-fifth of all recognized subspecies of *A. mellifera*; *A. m. meda*, *A.*
110 *m. syriaca*, *A. m. caucasica*, *A. m. anatoliaca* from the O-lineage and an ecotype from C subspecies
111 group exist in Turkey (Kandemir et al. 2005). Even A-lineage genetic material was also characterized
112 in native bees from the Levantine coast of Turkey (Kandemir et al. 2006) bringing together genetic
113 elements from three continents. Major subspecies found in and around Anatolia are shown in Fig. 1a.

114 Anatolia and Thrace, when taken together, harbor a vast diversity: honey bees belonging to three
115 different lineages meet, exchange genes and adapt to local conditions determined by diverse climatic,
116 topographical and floristic variations available (Bouga et al. 2011). Refugial status of Anatolia during
117 the ice ages contributed present enhanced levels of biodiversity (Hewitt 1999). Studies concerning

118 honey bee populations of Turkey (Bodur et al. 2007; Kence et al. 2009) demonstrated high genetic
119 structuring among them and confirmed the presence of divergent populations pointing to different
120 subspecies. They, all together, drew attention to this rich diversity hotspot present and particularly
121 under threat in Anatolia and Thrace as well as importance to its conservation.

122 Despite that, still prevail the arguments in popular opinion -with significant effect on decision makers
123 and other stakeholders- that honey bee ecotypes are inevitably lost due to gene flow facilitated by
124 anthropogenic factors, so the relevance of investing in a strategy involving conservation of locally
125 adapted variants are unremittingly questioned. This study aimed to quantify and weigh the impacts of
126 anthropogenic factors and conservation efforts on the present condition of honey bee genetic
127 diversity.

128 **Materials and Methods**

129 Sampling

130 We sampled a total of 250 honey bees each from different colonies from 18 provinces during the
131 period of March 2010 and August 2012. Of those 250 honey bees, 174 were from apiaries that were
132 stationary and 76 were from migratory ones. Beekeepers declared that they used original honey bees
133 from stocks native to the area and that they did not purchase non-native queens or colonies in the last
134 ten years.

135 We grouped samples from provinces with small sample sizes together with nearby provinces to form
136 10 major localities: Kırklareli, Edirne+ (Edirne and Tekirdağ), Muğla, Eskişehir+ (Eskişehir,
137 Kütahya and Bilecik), Düzce+ (Düzce, Zonguldak and Bolu), Ankara, Hatay, Bitlis+ (Bitlis, Elazığ,
138 Erzurum and Ordu), Ardahan, and Artvin. Those localities correspond to natural distribution range of
139 four subspecies. Those subspecies are *A. m. syriaca* in Hatay, *A. m. caucasica* in Ardahan and Artvin,
140 *A. m. anatoliaca* in Düzce, Eskişehir+, Muğla and Ankara from the O lineage and an ecotype from C
141 subspecies group in Kırklareli and Edirne+ by excluding the fifth subspecies *A. m. meda*. We carried
142 out combinations of locations according to geographical proximity; similarity in terms of climatic,
143 topographic and floral variables; results of previous studies as well as results of preliminary analysis
144 of this study. Sampling sites and sample sizes can be seen in Fig. 1b.

145 The samples were kept in -80 °C until genetic analysis.

146 Genotyping

147 We isolated DNA from bee heads by QIAGEN DNeasy Blood and Tissue Kit following the
148 procedure of the producer for insect samples with slight modifications. We grouped a set of 30
149 microsatellite loci into four clusters for two 7-plex (set 1: AP218, A113, AB024, AP249, A088,
150 AP001, AP043 and set 2: AP049, AP238, AC006, AP243, AP288, HBC1602, A107) and two 8-plex
151 (set 3: A079, AC306, AP226, A007, HBC1601, AP068, A014, AP223 and set 4: AP019, AB124,
152 A043, A076, AP273, AP289, HBC1605, A028) polymerase chain reactions (Estoup et al. 1995;
153 Solignac et al. 2003; Bodur et al. 2007; Shaibi et al. 2008; Tunca et al. 2009). A software program,
154 Multiplex Manager 1.2 (Holleley & Geerts 2009), was used for constructing the multiplex groups.
155 Information on primer pairs, fluorescent dyes and PCR conditions are provided in the supplementary
156 file.

157 Detection of microsatellite allele sizes was achieved by capillary electrophoresis with ABI 3730XL
158 sequencing machines. We were not able to amplify locus A076 consistently across the samples thus
159 we definitely excluded it from the data set and the downstream analysis.

160 Population structure

161 We calculated pairwise F_{ST} values by Arlequin 3.5 (Excoffier et al. 2005), Mantel test was applied to
162 account for isolation by distance procedure. Pairwise population distances were calculated (Reynolds
163 et al. 1983) by Populations 1.2.32 software (Langella 2011) and visualized by the online tool
164 Interactive Tree of Life v4 (Letunic and Bork 2019). We used PAST4 and PCAGEN software to plot
165 populations on a two-dimensional space by a PCA based on correlation matrix between groups
166 (Goudet 1999; Hammer 2001).

167 Population structure was estimated by Structure 2.3.3 (Pritchard et al. 2000), K values of distinct
168 populations were analyzed by Structure Harvester software (Earl & von Holdt 2012), and we used
169 Clumpp software (Jakobsson & Rosenberg 2007) to permute the membership coefficients of
170 individuals determined by Structure 2.3.3 and Distruct software to (Rosenberg 2004) visualize the
171 results obtained by Clumpp.

172 Other population genetic parameters and diversity indicators were also estimated and they are
173 provided as supplementary file. These parameters and indicators contain frequency of null alleles,
174 allelic richness and diversities, inbreeding and prevalence of close relatives, number effective alleles,
175 levels of heterozygosity, deviations from Hardy-Weinberg and linkage disequilibrium, bottlenecks,
176 effective population sizes and microsatellite information index.

177 Statistical analyses

178 We then used membership coefficients obtained, to test hypotheses about beekeeping practices,
179 isolated regions and queen/colony trade. For the analysis, we first arcsine root square transformed the
180 coefficients since the data was composed of proportions and non-normally distributed. Then we
181 carried out Shapiro, Mann-Whitney U , Kruskal-Wallis, Dunn's, F , ANOVA, Tukey's and t tests
182 wherever necessary and applicable to compare mean membership coefficients and estimated Cohen's
183 d to determine effect sizes. Those were carried out in R statistical software using packages pwr,
184 effsize, dunn.test and dabestr (R Core Team 2013, Torchiano 2016; Dinno 2017; Champely et al.
185 2018; Ho et al. 2019). R code is provided as a supplementary material.

186 We made use of estimation plots to visualize untransformed data for membership coefficients and
187 impact of various factors on them. This is a less conventional method when compared to bar or
188 boxplots and reporting of significance tests but much more convenient and powerful method to
189 summarize the whole data in an unbiased way by displaying all measurements and effect sizes as
190 well as precision of estimates and distribution of mean differences (Ho et al. 2019).

191 Beekeeping practice: migratory vs stationary

192 For the first hypothesis to be tested, we compared membership coefficients of migratory and
193 stationary colonies in Ankara, Muğla and Hatay separately, for the three provinces combined and for
194 the total data set. If the migratory colonies acted as a potential vector of foreign alleles then they
195 would have much lower probabilities of being assigned to their own clusters.

196 Isolated regions as a conservation practice

197 The second hypothesis was about isolated regions. If the isolated regions were efficient in preserving
198 genetic diversity by preventing gene flow between different clusters then one would expect to see
199 higher membership coefficients for stationary individuals belonging to these regions and lower for
200 stationary individuals that belong to regions open to migratory beekeeping.

201 Kırklareli is a province that is declared officially as an isolated region where migratory beekeepers
202 could not visit for years at first thanks to local beekeepers' negative attitude towards them. The
203 region is home to a Carniolan ecotype carefully maintained by local beekeepers. Ardahan is legally
204 declared a conservation and breeding area for *A. m. caucasica* so migratory beekeepers cannot enter
205 the province and queen import from other subspecies is forbidden. Parts of Artvin province are also
206 officially declared as isolated regions for conservation of *A. m. caucasica* as a pure race. The
207 province in general is rarely visited by migratory beekeepers for geographical reasons and beekeepers
208 there, dealing with mass queen breeding, do not use non-native queens. We compared these three
209 provinces with the other six regions (Edirne+, Muğla, Düzce+, Eskişehir+, Ankara and Hatay) where
210 migratory beekeeping and bee trade are freely exercised.

211 Effect of queen and colony trade

212 Third set of tests were about the impacts of honey bee trade. We compared the estimated proportion
213 of genomes assigned to a different cluster than the native cluster among individuals of the total data
214 set to find out which cluster contributed most to other clusters' gene pools.

215 Ardahan and Artvin provinces host the *A. m. caucasica* subspecies which is also widely used for
216 commercial purposes and the *caucasica* queens and their hybrids are sold all over the country. But
217 these provinces are also limited to a very narrow range in the Northeast of the country and are
218 declared isolated regions. So, a possible high introgression of their alleles would mostly, if not
219 completely, be due to replacement of queens and colonies.

220 We also investigated further patterns across populations to understand the magnitude and direction of
221 the gene flow by tracing the signs of those misassigned proportions within localities.

222 **Results**

223 We calculated F_{ST} values by using both the frequencies obtained in the study and by using the null
224 allele corrected frequencies. We calculated for the stationary ($n = 174$) colonies an overall F_{ST} of
225 0.065 and an F_{ST} of 0.067 after null allele corrections. For migratory colonies the values were 0.011
226 and 0.015 respectively and for all the 250 samples the values were 0.046 and 0.047.

227 Phylogenetic tree we constructed by using pairwise population distances based on stationary colonies
228 only resolves four distinct branches corresponding to four subspecies (Fig. 2b). Thracian samples
229 constitute the extreme end of the unrooted tree. The other end is divided to three almost equidistant
230 branches of Caucasian, Levantine and Anatolian samples.

231 We plotted stationary colonies, migratory colonies and the overall data for the regions of sampling on
232 2D spaces by carrying out Principle Component Analysis (Fig. 2a) which showed a similar pattern
233 with the UPGMA tree. First axis designating the first principle component differentiated samples
234 those in Thrace whereas the second one corresponding to the second component differentiated
235 subspecies in Anatolia (*syriaca*, *anatoliaca* and *caucasica*). The x and y axes explained 41.8% and
236 32.1% of the variance within the samples.

237 Genetic distances in stationary colonies showed significant correlation ($p < 0.001$) with geographic
238 distance but those of migratory colonies were not correlated with geographic distances. Results of
239 Mantel test point to an isolation by distance pattern in stationary colonies that is lost in migratory
240 ones.

241 Concerning the Structure results, the best K values were selected by the Structure Harvester program
242 as 2 and 4 with similar outcomes, K=2 being slightly likelier than K=4 which hint for lineage level
243 diversification of C and O ancestries. We calculated membership coefficients of individuals to the
244 observed clusters in K=4 since it can be biologically attributed to relevant subspecies under
245 investigation and we used them for further hypothesis testing. Clustering analyses showed no
246 population structuring for migratory colonies (Fig. 3a) in contrast to stationary colonies and the
247 overall data (Fig. 3b and 3c). Concerning the overall data, however, distortion in the population
248 structure caused by migratory colonies is evident in higher admixture levels observed.

249 We compared individuals from stationary and migratory colonies according to their membership
250 coefficients belonging to their native clusters (or it can be called their expected natural populations
251 alike). The mean values and effect sizes as well as the significance level of the differences were
252 summarized in Table 1. Boxplots contrasting the arcsine root square transformed membership
253 coefficients for migratory and stationary colonies are shown in Fig. 4a and scatter plots are very
254 much similar but visualizing raw membership coefficients for each sample are shown in Fig. 4b.

255 Estimation plots not only fairly visualize the real distribution of the data but also let us compare the
256 effect sizes and their precision. Stationary colonies are annotated as <Group name> 0 and migratory
257 colonies are as <Group name> 1 (Fig. 4). Bars right to the data points refer to the 25% and 75%
258 quartiles and the gap between them is the median value for the sample. The zero line below
259 correspond to the mean membership coefficients of stationary colonies in each pairwise comparison.
260 The Euclidean distances from those means for the migratory colonies are shown as dots with a 95%
261 confidence interval bar around. Also, distributions of the estimation statistics are included. So that we
262 can comprehensively compare the strength of the drift for different populations and subsets of the
263 data.

264 Stationary colonies from Muğla and Hatay were quite more likely to be assigned to their own clusters
265 than the migratory colonies from these provinces, the same held when we compared the combined
266 data from the three provinces or all the migratory and stationary colonies. However, the situation was
267 the reverse in Ankara possibly due to factors we discuss below. Stationary colonies from that
268 province reflected patterns of high admixture. The difference between stationary colonies and
269 migratory in Muğla are much less when compared against the ones in Hatay, signaling for a possible
270 higher level of admixture in Muğla.

271 For that first comparison we used the complete ($n = 250$) data set to be able to quantify the
272 differences in membership coefficients for migratory and stationary colonies. But for the rest of the
273 analysis we used the subset of data which is only composed of stationary colonies ($n = 174$) since this
274 would better reflect the population structure.

275 In the first scatter and the corresponding boxplots (respectively Fig. 5b for raw membership
276 coefficients and Fig. 5a for transformed values) one can observe that within each locality samples are
277 assigned with high proportions to their native clusters despite some admixed individuals. Also, one
278 can see through observation of unpaired mean differences that Kırklareli, Ardahan, Artvin, Hatay and

279 to a lesser extent Düzce play role as centers of genuine subspecies diversity with exceptionally high
280 levels and few individuals of admixed origin.

281 But when we compared isolated regions (Kırklareli, Ardahan, Artvin) and regions open to migratory
282 beekeeping (consisting of Edirne+, Muğla, Ankara, Düzce+, Eskişehir+ and Hatay provinces in our
283 sample) in terms of their arcsine transformed membership coefficients (Table 1 for means, effect
284 sizes and significance of the difference and Fig. 5a second boxplot) we witnessed that -as expected-
285 stationary colonies within isolated regions showed significantly higher fidelity to the original
286 clusters. This is also obvious in the estimation plot in Fig. 5c where the mean membership
287 coefficients of samples that are from regions open to migratory beekeeping (green colored group
288 designated as <0>) fall beyond the 95% confidence interval of the estimated mean of the samples
289 from the isolated regions (orchid colored group designated as <1>).

290 Even if the individuals are assigned with high probability to their own clusters, let's say with a 90%
291 of probability, this means that 10% of their genome still belongs to other clusters. Given that there
292 are four clusters, we investigated if any of these misassigned genome parts were enriched for any of
293 them. Mean transformed values for Thracian cluster misassignments among individuals of the other
294 populations were 0.16 and 0.25 for Anatolian cluster, 0.26 for the Caucasian and 0.20 for the
295 Levantine (Fig. 6a).

296 A significant Kruskal-Wallis test ($p < 0.001$) and a *post hoc* Dunn's test, accompanied by a
297 significant ANOVA result ($p < 0.001$) followed by a Tukey's test, showed that misassignments to *A.*
298 *m. caucasica* and *A. m. anatoliaca* clusters were significantly more frequent than the others ($p < 0.001$
299 for both subspecies against C-lineage Thracian bees and $p < 0.05$ against *syriaca* group). The effect
300 sizes according to Cohen's *d* varied from 0.34 to 0.54 with estimation plots verifying the precision of
301 the difference observed (Fig. 6b). Despite observation of the highest values in *A. m. caucasica*
302 misassignments, the results were not significant between *A. m. caucasica* and *A. m. anatoliaca*
303 clusters.

304 We checked if those differences result from many individuals with high admixture levels but such
305 data only constituted the 7.5% of all the observations. This is with a threshold level of 0.5 for
306 transformed values which corresponds to a second hybrid with a 25% contribution of non-native
307 origin. So, we concluded that rather the main effect is due to consistent mid to low subspecies'
308 contributions to other populations.

309 We also investigated if these small drifts in admixture proportions were more prominent in some
310 localities and if populations differed in the subspecies they are receiving gene flow. This led us to
311 comprehend the extent, magnitude and direction of the patterns of gene flow among the subspecies
312 with a particular sensitivity to the populations. The results are summarized in Fig 7. We applied
313 Dunn's test for each pairwise comparison between the populations. Significance of results will be
314 mentioned but details of the test results can be found in the supplementary file.

315 Contributions from the Thracian cluster seem to be high in Düzce+ and Eskişehir+ which reside in
316 the southeast of Marmara Sea across the Bosphorus and also there is some non-significant surplus in
317 Muğla province in the Aegean coast. This is consistent with the Structure result for $K = 2$. While
318 Thracian populations of Kırklareli and Edirne+ receive most gene flow from the Anatolian cluster
319 these are not significantly different than Anatolian contributions to other regions which points to a
320 balanced, uniform contribution of this subspecies to each group. Although non-significant, Edirne+
321 receives much gene flow from that cluster. Caucasian cluster on the other hand contributes most to

322 Ankara, Muğla, Eskişehir+ and Düzce+ populations. Only significant differences are observed
323 between Ankara and Muğla receiver populations and Kırklareli in terms of admixture with Caucasian
324 populations. The same populations also significantly differed from Kırklareli in their admixture
325 levels with the Levantine cluster. In both cases Kırklareli population in Thrace had lower gene flow.
326 This is interesting since Muğla in the southwest and Caucasus region at the northeast lie at the
327 different extremes of the country.

328 Discussion

329 F_{ST} values obtained were highly significant but they were lower than what Bodur et al. (2007)
330 estimated -a total F_{ST} of 0.077 together with higher values for pairwise comparisons among
331 populations- by samples collected ten years before ours. This may indicate a recent increased gene
332 flow and can be an alarm signal for a trend. Constant monitoring studies are needed in the future to
333 see if it is a persistent trend really. The high degree of structuring in stationary colonies according to
334 F_{ST} results was lost in migratory ones, meaning they are less differentiated from each other due to
335 high degree of gene flow.

336 Phylogenetic tree clearly showed that Thracian samples were completely distinct from others
337 pointing to an early division of populations and limited gene flow. This supports the hypothesis for a
338 Carniolan (C-lineage) descent of Thracian bees in Turkey. Directly including samples from the major
339 C-lineage subspecies would confirm the subspecies of these bees highly differentiated from
340 Anatolian samples. In our initial observations of another research, Thracian samples grouped with C-
341 lineage European breeds rather than the samples throughout Anatolia (Kükreker, unpublished data).
342 This is a challenging point to Ruttner's claim (Ruttner 1988) that Thracian bees belong to *anatoliaca*
343 subspecies and needs further investigation.

344 West Anatolian, Levant and Caucasian populations did also form separate clusters in the tree. PCA
345 results confirmed those 4 different clusters inferred from tree topology. Bitlis+ samples resided with
346 Central and West Anatolian populations in both phylogenetic tree and PCA results (supplementary
347 file) but it should be kept in mind that all samples from that locality belonged to migratory colonies
348 so resampling with inclusion of stationary colonies from East Anatolia would be beneficial to
349 understand the real phylogenetic relations.

350 The two most possible K values in structure analysis for the whole sample and the stationary colonies
351 were $K = 2$ and $K = 4$, both results supporting the hypotheses of populations belonging to 2 separate
352 lineages (C and O) and 4 distinct subspecies (a Carniolan ecotype in Thrace, *A. m. caucasica* in
353 Artvin and Ardahan, *A. m. syriaca* in Hatay and *A. m. anatoliaca* widely distributed covering the rest
354 of the country). In contrast to the expectations of migratory beekeepers of making use of native
355 stocks, results involving migratory beekeepers' samples lacked any population structuring in the
356 cluster analysis further clarifying the highly hybridized status of migratory apiaries.

357 Stationary apiaries, as expected, yielded highly structured groups where all the subspecies could be
358 detected. When K was 2, the structure analysis of two distinct clusters showed that there was a
359 transition zone between Thracian and Anatolian samples around Marmara Sea and Aegean. This may
360 be a hybrid zone between the C and O lineages like the ones identified before between M and C
361 lineages in Alps and Apennine Peninsula and between A and M lineages at the Iberian Peninsula and
362 Mediterranean islands (De la Rua et al. 2009). An analysis of ecological niches under species
363 distribution models suggest an intersection of habitat suitability of both subspecies within the
364 aforementioned geographic area (Kükreker, unpublished data).

365 When K was considered as 4, all four subspecies were easily differentiated from each other, in
366 accordance with the expectances. The significance of two distinct clusters ($K = 2$) was higher than
367 four ($K = 4$) which means that the differences between the populations belonging to C (in Thrace)
368 and O (in Anatolia) lineages are more clear-cut than differences between the populations of four
369 different subspecies.

370 *A. m. anatoliaca* samples fell in the middle of the other subspecies in ordinations, being similar to all
371 other populations according to FST values despite being a distinct cluster in structure analysis which
372 may point to a significant historical contribution to *A. m. anatoliaca* populations from the
373 neighboring regions. Another explanation can be that *anatoliaca* subspecies' putative basal position
374 for O-lineage honey bees places it as a center of genetic diversity. With *anatoliaca* bees exhibiting a
375 distinct identity, the situation was quite different than what was observed in all-migratory Bitlis+
376 samples where a mixture of different clusters surpassed instead of a separate identity.

377 A better understanding in terms of phylogenetic relationships between the populations in Turkey can
378 be developed if populations neighboring Anatolia and Thrace in Balkans, Iran, Caucasus and the
379 Middle East are also sampled. This can be a direction for future research, for shedding light on the
380 complicated taxonomic status within and between the C and O lineages and for drawing edges and
381 transition zones of the subspecies present across the whole region.

382 Results from different analyses conducted here confirmed the presence of clusters but also, they all
383 together pointed to the status of migratory colonies: they might be acting as a hybrid zone mobile in
384 space and time, being at one region in spring and at others in summer and fall, becoming vectors of
385 otherwise local gene combinations. Statistical results concerning a comparison between migratory
386 and stationary colonies confirmed the significant gene flow towards the migrants from local bees.

387 A significant gene flow towards local bees was also observed by a comparison between isolated
388 regions and those are not. This result, derived from directly contrasting two settings in an
389 experimental framework, is pointing to the vitality of establishing areas away and free from
390 migratory beekeeping for preservation of honey bee genetic diversity in conclusion with other studies
391 on conservation practices (Pinto et al. 2015; Oleksa et al. 2015).

392 One interesting point in the results was that the trend of the stationary colonies in Ankara. They had a
393 significantly lower probability of being assigned to their own clusters than the migratory colonies of
394 their province. This may be related with the regions migratory beekeepers of Ankara prefer to visit
395 during their migratory cycle or due to the insistent preference of using native queen bees by
396 migratory beekeepers. The low assignment degree of stationary colonies in Ankara may also be
397 related with Kazan apiary of TKV (Development Foundation of Turkey) placed there where hundreds
398 of colonies of Caucasian bees are raised and sold around for more than 30 years. The same practice is
399 also carried out by many queen bee breeders in Kazan region. Gene flow through these apiaries and
400 queen bees distributed locally by trade may contribute quietly to such an admixture observed in
401 stationary colonies in Ankara. The high misassignment probability of colonies in Ankara to the
402 Caucasian cluster also revealed such a process as probable.

403 It's hard to directly quantify the effect of queen and colony trade but unique features of Anatolia and
404 Thrace by availability of a number of naturally occurring subspecies renders possible the
405 understanding of their relative roles. Honeybees from stationary colonies were assigned more often to
406 their native clusters but they were also assigned to other clusters with lower probabilities. Samples in

407 the whole range of the study misassigned to Caucasian cluster more often than they were misassigned
408 to others.

409 This is most probably due to wide distribution of Caucasian queen bees by trade. Migratory
410 beekeeping is not practiced in Ardahan and Artvin where highly commercial Caucasian bees are
411 native. Hence no bees go in or leave out the region as migratory colonies. So, the observed
412 introgression of Caucasian alleles to the stationary colonies elsewhere whose beekeepers let them
413 change their queens on their own rather than purchasing queens of different origins, could mainly be
414 attributed to frequent queen bee and colony replacements in neighboring apiaries within those
415 regions.

416 It is shown here that practicing of honey bee replacements increase the level of admixture within the
417 gene pool. As previously discussed, a very high level of Caucasian introgression was observed in
418 Ankara. *A. m. anatoliaca* alleles also showed high introgression especially in Edirne+ of Thrace
419 region but also at average levels in other regions. These high levels may be related to this subspecies'
420 geographical proximity to other populations which might have led to historical and recent gene
421 exchange. By another explanation it can be related to the widespread practicing of migratory
422 beekeeping by Western and Central Anatolian beekeepers throughout Turkey, rather than queen or
423 colony replacements since there are very few commercial queen breeders within the distribution
424 range of *A. m. anatoliaca*.

425 Results of the various statistical tests carried out and analysis applied in this study clearly showed
426 that the genetic structure of honeybee populations in Turkey were highly conserved and still
427 maintained. But it doesn't mean that the structure and diversity observed is secure. Rather it should
428 be considered under threat since the anthropogenic factors leading to gene flows are still underway
429 and keep admixing the populations.

430 A quiet interesting point was that, the preservation of population structure was achieved despite a
431 very high number of colonies moved from one location to the other by migratory beekeeping practice
432 and despite unregulated and frequent queen and colony sales. Future research may also need to focus
433 on how this biodiversity and its structuring were preserved and its relation to natural selection.
434 Further hypothesis can be formulated to distinguish the relative effects of natural selection and gene
435 flow, the former could be so significant that it could potentially counterbalance the latter.

436 Genetic variation eventually leading to local adaptations with such significant outweighing effect can
437 be considered as a valuable resource for honey bee populations in the global context at this time of
438 unusual bee losses as well as global climate change. So, a better understanding of both present
439 adaptation to local climate and geographic conditions as well as adaptive capacity to future changes
440 would better be developed for the sake of the bees and their beneficiaries. A fair amount of effort
441 should be invested on more studies focusing on candidate functional variants at the genome level that
442 play role in due process in different parts of the world. Novel and innovative ways of coping with
443 environmental and climatic stressors developed by honey bee populations or exploration of
444 interesting patterns of convergent evolution are waiting ahead to be yet discovered.

445 Importance of establishing isolated regions was highlighted with genetic data. The results of the
446 statistical tests showed a significant difference between the conservation of identity in and out of
447 isolated regions with isolated regions staying purer in terms of subspecies composition. Such regions
448 were proven to be effective in conservation of unique diversity present within.

449 In the light of this study we propose a renewed effort to address the need for massive establishing of
450 such regions for conserving locally adapted native bees throughout the whole natural distribution of
451 the species. This especially holds for underrepresented regions in terms of local diversity hotspots. A
452 gap analysis aiming for complementarity in the planning of systematic conservation efforts are
453 urgently needed globally.

454 In such isolated regions, naturally, migratory beekeeping as well as replacement of queen bees with
455 non-native ones must be strictly prohibited and checked by relevant molecular monitoring
456 techniques. However, these isolated regions should also be wide enough involving additional buffer
457 zones where further restrictions on migratory beekeeping and bee trade are applied for efficient
458 isolation and for fulfilling sufficient effective population sizes.

459 Thanks to increasing awareness in the last decade within the industry, now there are at least 11
460 isolated regions in service or being established in Turkey through significant efforts of scientists and
461 their collaboration with Turkish Beekeepers Association. There is an ever-growing need for
462 establishing closer links with decision makers and stakeholders and necessity of investing more effort
463 in communicating results of scientific studies in order to make the most out of them.

464 Queen bee trade is not currently subjected to any restrictions or regulations in Turkey and there are
465 still very few pioneering measures within the natural distribution range, obviously not enough to
466 guarantee the realized preservation in the next decades. Such measures should be applied from a
467 conservation perspective to avoid extinction of native races, ecotypes and diversity present in these
468 populations. Genetic similarity of donor and recipient populations should be considered while
469 determining migration routes for migratory beekeepers and determining permissions for bee sales.

470 Central and western Anatolian populations suffer from significant gene flow from Caucasian
471 populations as demonstrated by our results. Muğla and Ankara especially showed alarming levels of
472 significant gene flow from other subspecies. This is not unexpected since the former receives
473 millions of migratory colonies during the honeydew season.

474 Despite its wide range of distribution spanning Anatolia from one side to other, special consideration
475 should be taken for preserving *A. m. anatoliaca* subspecies. The large and heterogenous native range
476 of this subspecies permitted the evolution of numerous ecotypes including those in coastal, inner step
477 or rainy forest ecosystems with noteworthy adaptations linked to their local environments.

478 The case with Hatay's *syriaca* populations too, can get worse and worse since the migratory
479 beekeeping practice is heavily carried out in the region and queen bee replacement with non-native
480 races was frequent throughout the last decade. This is mainly due to aggressive behavior, high
481 swarming tendency and an infame for low levels of honey piling but the subspecies is also shown to
482 exhibit some natural forms of varroa resistance (Kence et al. 2013). In the future this may end up in
483 *A. m. syriaca* colonies getting limited to a few localities and apiaries since the range of the subspecies
484 in Turkey is very narrow. A long-term conservation program considering improvement of traits that
485 result in beekeepers staying away from that subspecies should be actualized immediately in this
486 region too.

487 Thracian populations show a significant differentiation from the rest of the bees in Anatolia but the
488 subspecies which they belong to is not characterized on a strong basis yet and this unique population
489 is not registered officially like the case with *A. m. syriaca* of Hatay. Only subspecies officially
490 recognized in Turkey is *A. m. caucasica* so identification and registration procedures for the others
491 should be put into practice as soon as possible.

492 An improvement based on molecular genetic techniques can be applied to the ongoing conservation
493 programs for the *A. m. caucasica* subspecies. It is interesting to note that we even detected hybrid
494 individuals within the range of largest, oldest and heavily invested conservation area. This proposal
495 for application of molecular monitoring techniques holds for other subspecies too.

496 Recently a registration procedure for Muğla bees as an Aegean coastal ecotype of *anatoliaca*
497 subspecies with specific adaptations to resource phenology in the form of availability of honeydew
498 obtained from scale insect *M. hellenica* of Turkish red pine *P. brutia* is under process. During the
499 conservation and breeding efforts, an adequate level of use of molecular markers was achieved
500 (Kükrer, unpublished data). More attention should be paid to genetically characterize *A. m. meda*
501 subspecies that was out of the reach of this study and which can be threatened by anthropogenic
502 factors listed and studied here.

503 Rather than queen bee replacement, it should be encouraged to use native bees improved for desired
504 characters which are also locally adapted by definition. Such improved breeds would be used locally
505 and not distributed in a country-wide manner so that local adaptations would still be preserved while
506 bees are selected for resistance to pests and pathogens, hygienic behavior, reduced aggressiveness,
507 reduced tendency for swarming, higher winter survival, higher productivity or for increased
508 pollination. For obtaining better results in that, research concerning the smoothing, development and
509 extension of breeding locally adapted native bees and artificial insemination techniques should be
510 given higher priority and be adopted globally throughout the natural distribution range of local
511 subspecies.

512 Our overall results answer arguments about the present situation of honey bee subspecies in Turkey
513 but they also bear a wider interest to the community since they constitute an important pioneer
514 attempt to quantify the effects of human impact. Our measurable and justified scientific findings on
515 migratory beekeeping, queen and colony trade as well as conservation implications will hopefully be
516 of some use for the decision makers and other stakeholders.

517 **Conflict of Interest**

518 The authors declare that the research was conducted in the absence of any commercial or financial
519 relationships that could be construed as a potential conflict of interest.

520 **Author Contributions**

521 All authors conceived and planned the experiments and contributed to the field work. Mert Kükrer
522 carried out the experiments and statistical analyses. All authors contributed to the interpretation of the
523 results. Mert Kükrer took the lead in writing the manuscript. All authors provided critical feedback
524 and helped shape the research, analysis and manuscript. Aykut Kence was in charge of overall
525 direction and planning.

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533 makers and beekeepers to the conservation of locally adapted native bees as a precious legacy of our
534 kind. He will be remembered also for defending theory of evolution to be taught in science
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547 the corresponding authors thesis (Kükrer 2013).

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687 Tables

688 **Table 1.** Genetic impact of beekeeping and conservation practices on (arcsine root square
689 transformed) membership coefficients to native clusters (** $p < 0.05$, * $p < 0.01$ and *** $p < 0.001$).

690

	n	n				
	Migratories	Stationaries	Stationaries	Migratories	Cohen's <i>d</i>	<i>U</i> and <i>t</i> tests significance
Beekeeping Practice						
Ankara	9	18	0,82	1,11	-1,00	*
Muğla	15	21	0,93	0,70	0,89	*
Hatay	13	23	1,20	0,66	2,01	***
Combined	37	62	1,00	0,79	0,66	**
Overall	76	174	1,06	0,72	1,22	***
Conservation Practice						
Isolated	NA	79	NA	1,21	0,49	***
Not						
Isolated	NA	95	NA	1,08		

691

692 Supplementary material

693 R codes and Supplementary File.

694 Figure legends

695 Figure 1. Geographic distribution of (a) major honey bee (*A. mellifera*) subspecies in and around
696 Anatolia (b) sampling sites and sample sizes.

697 Figure 2. (a) PCA of stationary colonies, Component 1: 41.8%, Component 2: 32.1% (b) UPGMA
698 tree of honey bee populations based on Reynolds' 1983 genetic distances (orange: Thracian, yellow:
699 Anatolian, blue: Caucasian, violet: Levantine clusters).

700 Figure 3. Estimated population structure and clustering of honeybees in Anatolia and Thrace for (a)
701 migratory colonies (b) stationary colonies (c) the whole sample (orange: Thracian, yellow: Anatolian,
702 blue: Caucasian, violet: Levantine clusters).

703 Figure 4. Comparison between stationary (Sta_) and migratory (Mig_) colonies in Ankara, Muğla
704 and Hatay, as well as these three provinces combined and the whole data set, n = 250 (a) boxplot
705 display of arcsine root square transformed membership coefficients (b) scatter plot with estimations
706 of mean differences based on raw individual membership coefficients (yellow: Ankara and Muğla
707 belonging to Anatolian cluster, violet: Levantine cluster, coral: for a combination of three provinces,
708 firebrick: whole data).

709 Figure 5. Comparison between isolated regions and regions that are open to migratory beekeeping (a)
710 first boxplot display the arcsine root square transformed membership coefficients for 9 populations
711 whereas the second one presents a comparison of samples within isolated regions and those are not
712 (b) scatter plot with estimations of mean differences based on raw individual membership
713 coefficients to the native clusters (c) scatter plot contrasting individual raw membership coefficients
714 with an estimation of mean differences (orange: Thracian, yellow: Anatolian, blue: Caucasian, violet:
715 Levantine clusters, orchid and “1”: isolated regions, green and “0”: regions open to migratory
716 beekeeping).

717 Figure 6. Comparison of misassignment proportions between the major clusters (a) boxplot display of
718 arcsine root square transformed membership coefficients (b) scatter plot with estimations of mean
719 differences based on raw individual membership coefficients (orange: Thracian, yellow: Anatolian,
720 blue: Caucasian, violet: Levantine clusters).

721 Figure 7. Patterns of gene flow between populations (a) boxplot display of arcsine root square
722 transformed membership coefficients (b) scatter plot with estimations of mean differences based on
723 raw individual membership coefficients (orange: Thracian, yellow: Anatolian, blue: Caucasian,
724 violet: Levantine clusters).

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729

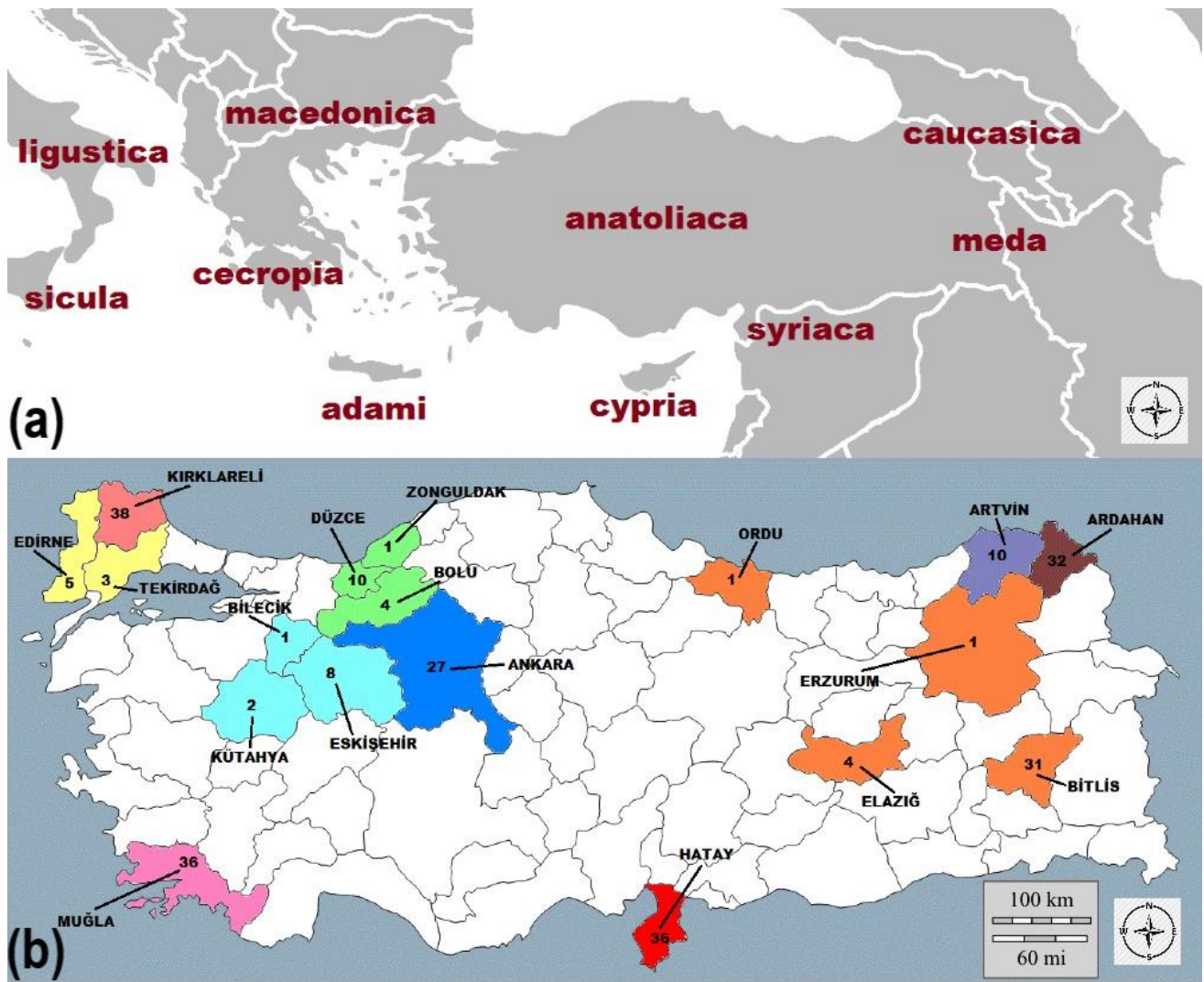
730

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732

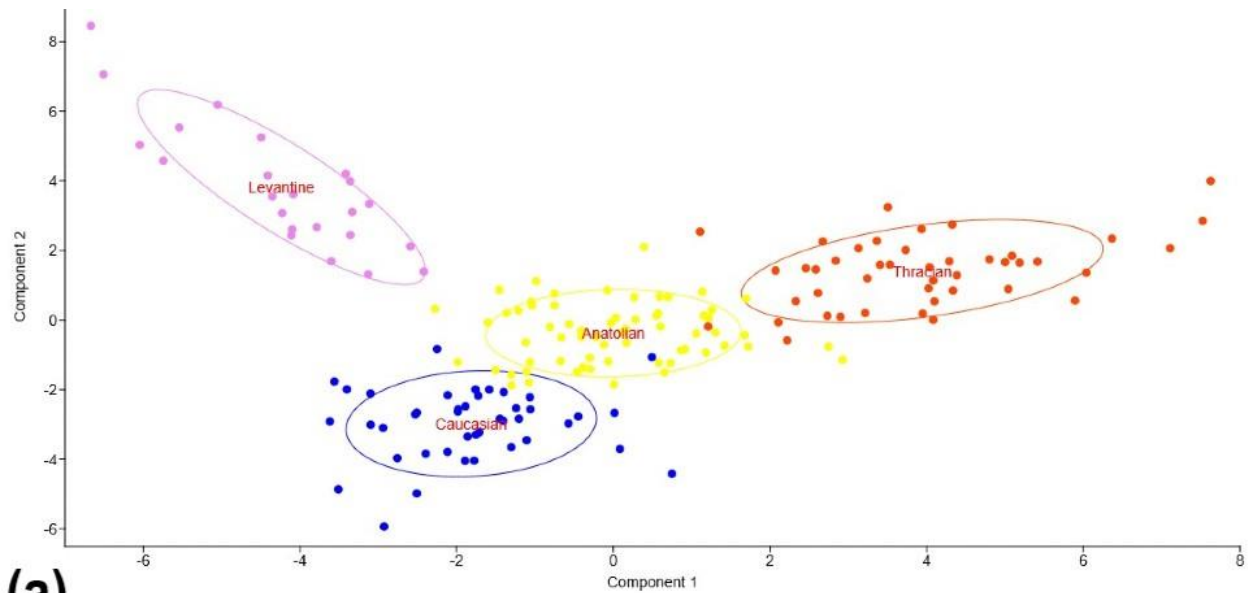
733

734

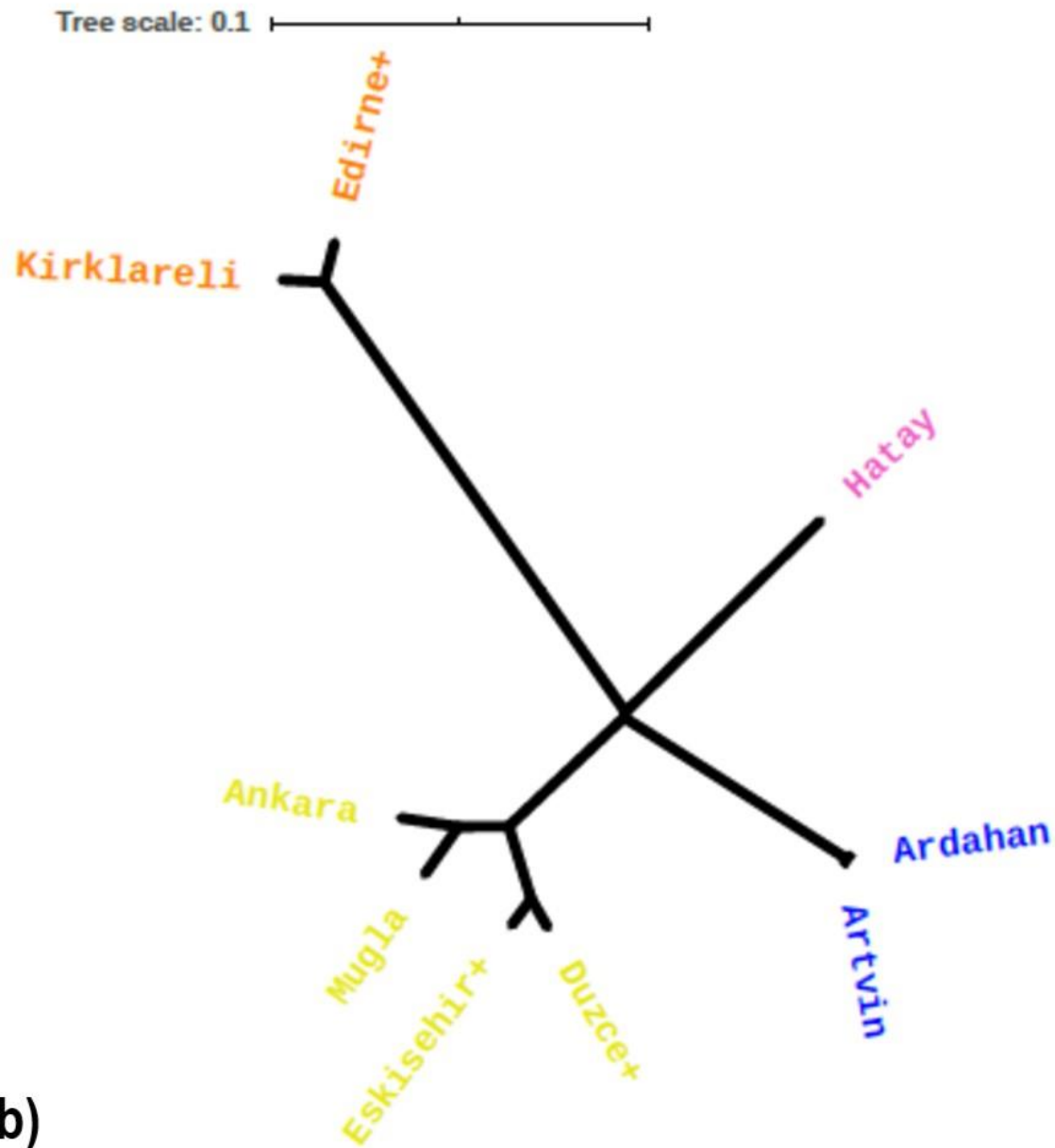


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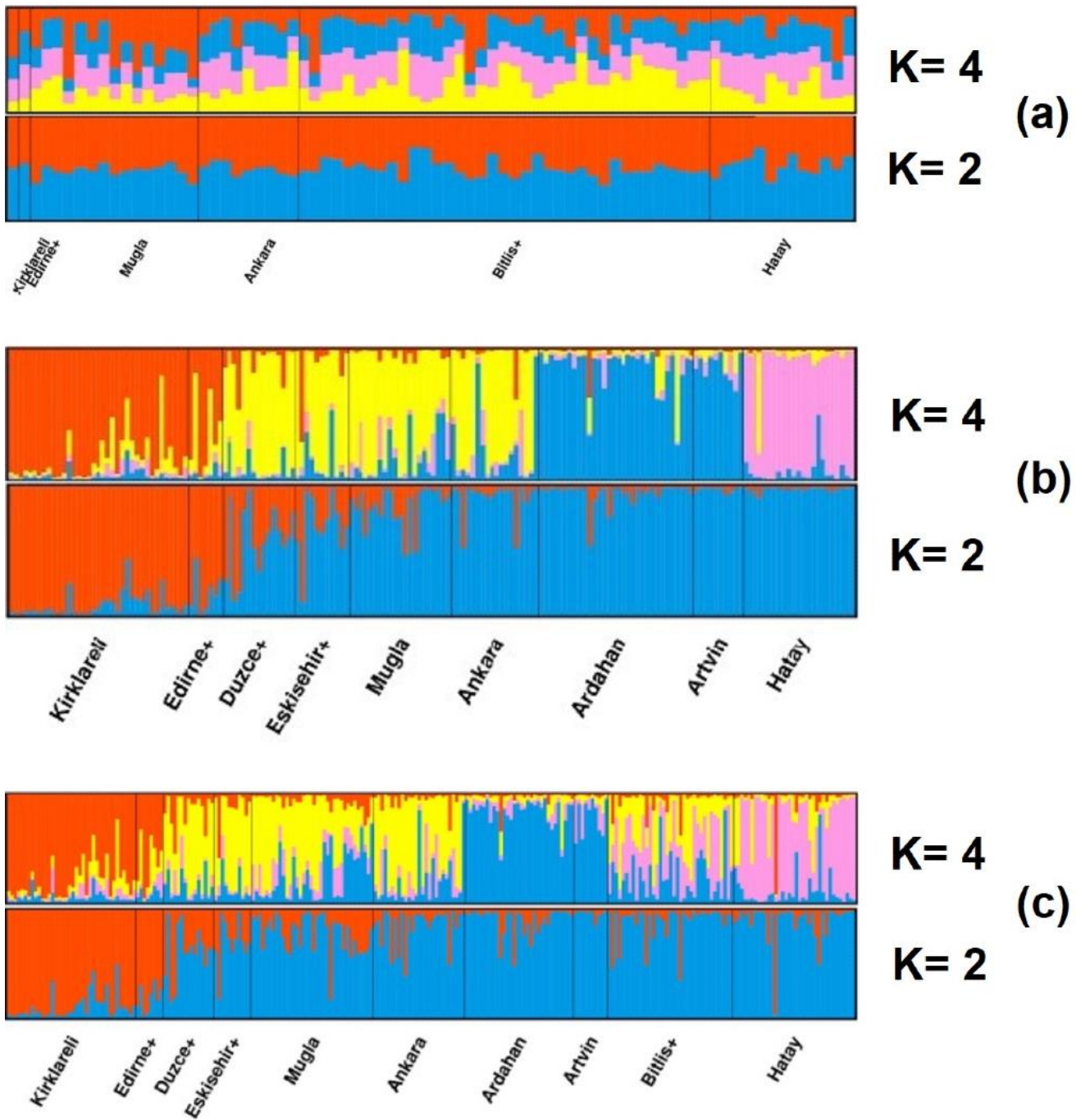
Anthropogenic factors and honey bee diversity



(a)

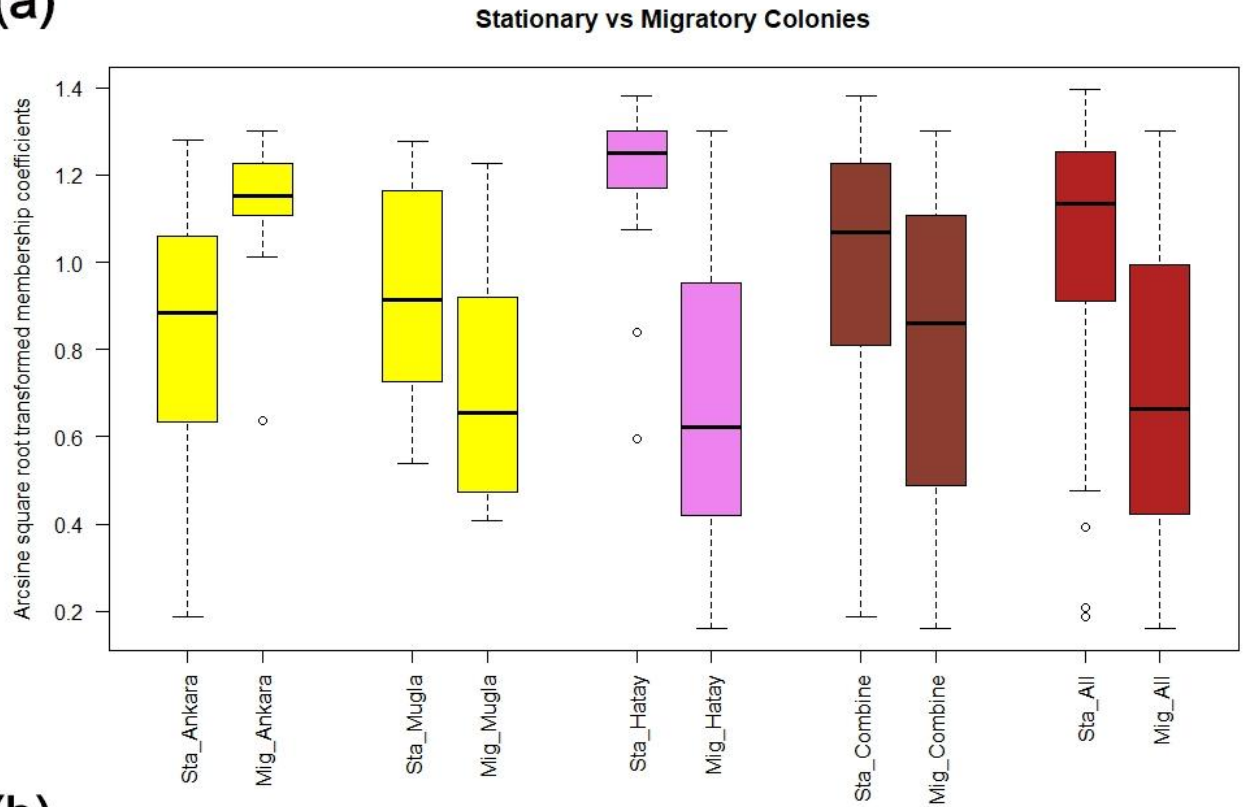


(b)

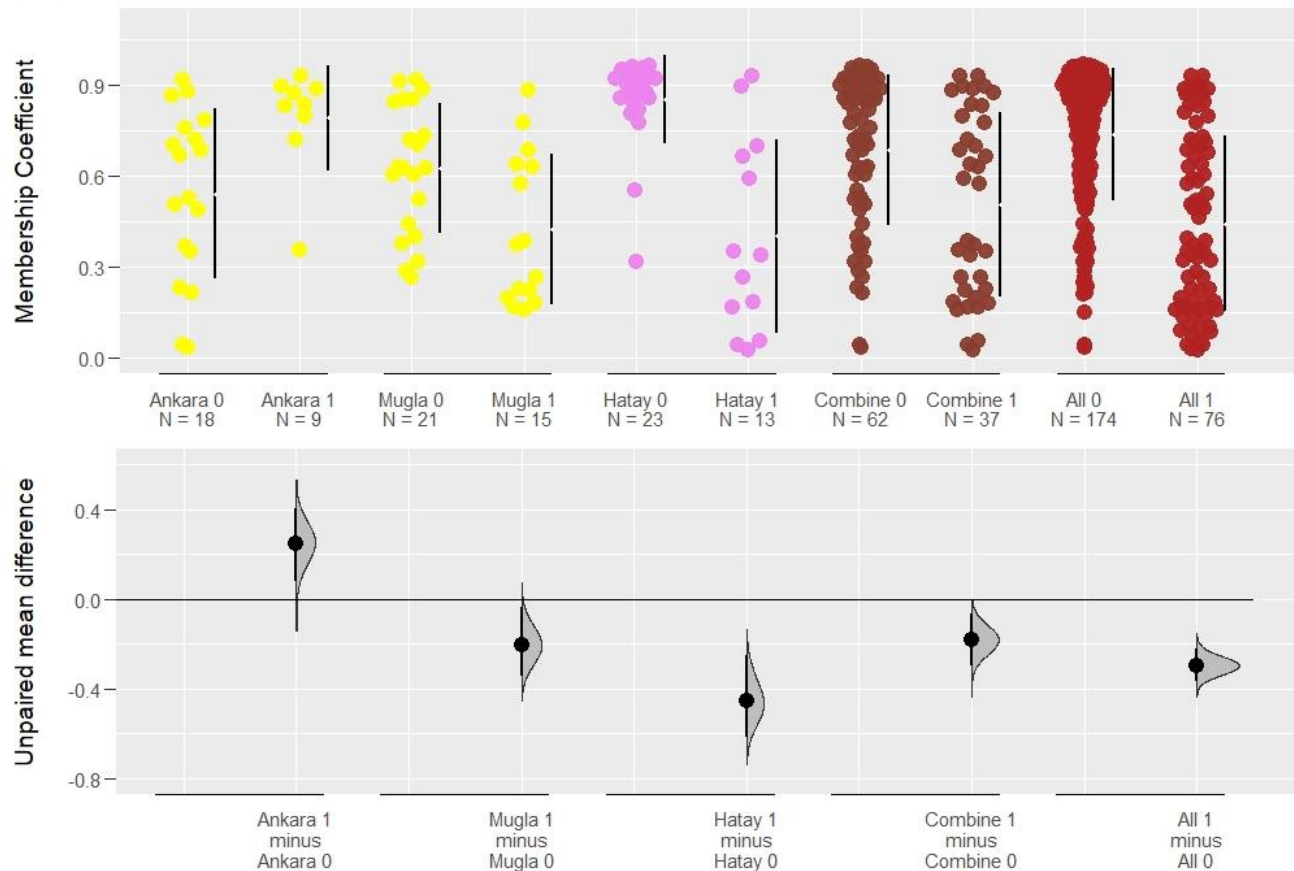


737

(a)

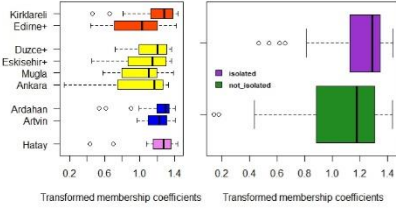


(b)

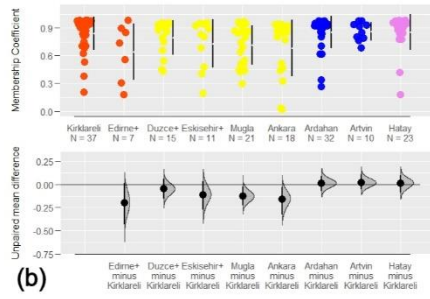


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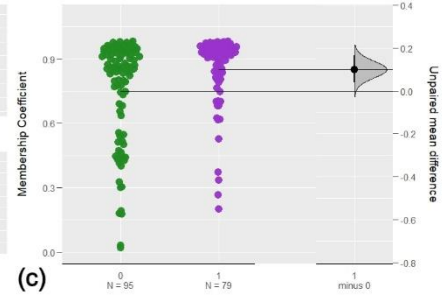
(a)



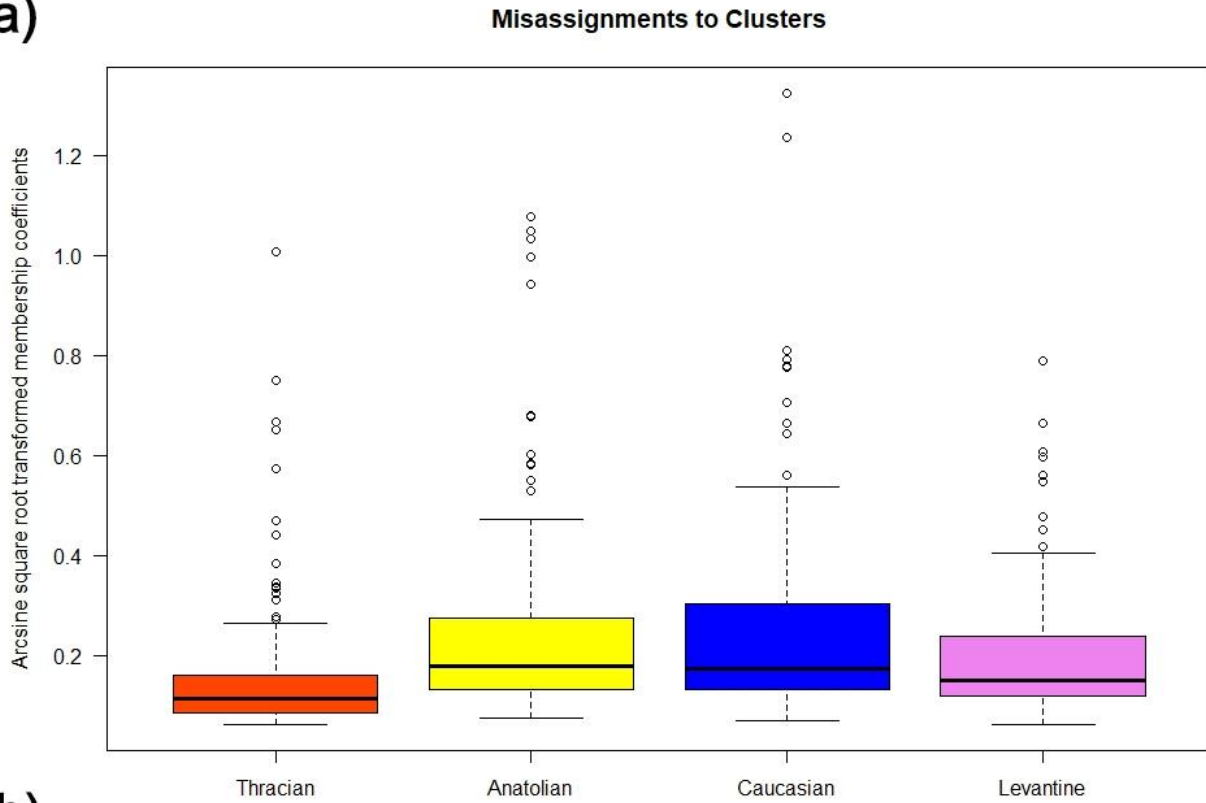
(b)



(c)



(a)



(b)

