



1 **Title: Thermal mismatches explain how climate change and infectious disease**
2 **drove widespread amphibian extinctions**

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13
14 **One Sentence Summary:** Climate change and infectious disease interacted to cause widespread
15 amphibian extinctions among species adapted to cool conditions.

16
17 **Author contributions:** All authors contributed ideas, J.M.C. and M.D.V. wrote proposals to
18 acquire animals, J.M.C., M.D.V. and T.A.M. conducted disease experiments, J.M.C. assembled
19 climate database, J.M.C., D.J.C. and J.R.R. conducted statistical analyses, J.M.C. and J.R.R.
20 wrote the paper, and all authors provided editorial advice.

21 **Abstract:**

22 Global temperatures and infectious disease outbreaks are simultaneously increasing, but linking
23 climate change and infectious disease to modern extinctions remains difficult. The *thermal*
24 *mismatch hypothesis* predicts that hosts should be vulnerable to disease at temperatures where
25 the performance gap between themselves and parasites is greatest. This framework could be used
26 to identify species at risk from a combination of climate change and disease because it suggests
27 that extinctions should occur when climatic conditions shift from historical baselines. We
28 conducted laboratory experiments and analyses of recent extinctions in the amphibian genus
29 *Atelopus* to show that species from the coldest environments experienced the greatest disease
30 susceptibility and extinction risk when temperatures rapidly warmed, confirming predictions of
31 the *thermal mismatch hypothesis*. Our work provides evidence that a modern mass extinction
32 was likely driven by an interaction between climate change and infectious disease.

33 **Main Text:**

34 Global climate change and emerging infectious diseases represent two of the most
35 formidable ecological challenges in modern times, but controversy exists over whether they are
36 causally linked (1, 2). Climatic conditions often directly influence disease outbreaks (3), and
37 many predictive models and experiments have revealed that climate change and infectious
38 diseases can independently drive current and future declines in biodiversity (4, 5). However,
39 there is surprisingly little concrete evidence that interactions between climate change and
40 infectious disease are causing widespread biodiversity losses (1, 6), possibly because of a lack of
41 theoretical frameworks, supported by a combination of experiments and field data, that can relate
42 climatic factors to host-parasite interactions to account for shifts in biodiversity (1). Such
43 frameworks would be valuable in establishing causal links between climate change and
44 extinctions mediated by disease.

45 A recently proposed hypothesis, the *thermal mismatch hypothesis* (7), suggests that
46 infectious disease outbreaks are likely to occur at temperatures where the performance gap
47 between pathogens and their hosts is greatest. Because parasites generally have broader thermal
48 performance breadths than hosts (8, 9), and both hosts and parasites might be locally adapted to
49 climatic conditions in their ranges and limited by extreme conditions, the hypothesis posits that
50 hosts adapted to cooler climates should be especially susceptible to disease under unusually
51 warm conditions, and vice versa (Fig. 1). Importantly, the predictions of the thermal mismatch
52 hypothesis are robust to relaxing several of its assumptions, such as local adaptation of host and
53 parasite and the degree and direction of the skew of the performance curves (Fig. S1). Therefore,
54 the *thermal mismatch hypothesis* provides a framework to predict which species might be most
55 likely to experience disease-driven declines under warming, and thus might be able to explain

56 patterns in species declines associated with climate-related outbreaks of emerging infectious
57 diseases.

58 Although recent experimental and field evidence support the predictions of the *thermal*
59 *mismatch hypothesis* (7), the hypothesis has not yet been applied to predict widespread host
60 declines associated with climate change and infectious disease. In perhaps the most
61 comprehensive spatiotemporal dataset describing a modern mass extinction, La Marca *et al.* (10)
62 provide dates of approximately 60 extinctions in the amphibian genus *Atelopus* putatively caused
63 by the chytrid fungus *Batrachochytrium dendrobatidis* (*Bd*), a pathogen that infects the
64 epidermal layer of adult amphibians and is implicated in worldwide amphibian declines (11;
65 *Table S1*). These *Atelopus* declines have been linked to climate change (12, 13), occurring in
66 years with warm or highly variable temperatures (12, 13) despite *Bd* growing best in culture
67 under cool or moderate conditions (7). *Bd* is sensitive to environmental conditions (14), can be
68 locally adapted (15), and has a broad thermal breadth (16), fulfilling the assumptions of the
69 *thermal mismatch hypothesis*. Thus, this *Atelopus* extinction dataset provides a unique
70 opportunity to examine whether the *thermal mismatch hypothesis* can accurately predict the
71 timing and location of disease-driven extinctions and to compete the *thermal mismatch*
72 *hypothesis* against alternative hypotheses for these climate-related declines. Although there have
73 been previous analyses associating climate change and *Atelopus* spp. extinctions, they relied
74 exclusively on correlations based on broad-scale, regional climate data instead of data from
75 individual species' ranges (12, 13; *Supplement*) and thus they failed to account for spatial
76 heterogeneity in climate change and host adaptations to local climates. Therefore, previous
77 analyses could not determine whether *Atelopus* spp. declines were likely caused by climate
78 change alone or an interaction between climate change and disease. Thus, we do not yet have

79 clear, causal evidence in this system, or any system that we are aware of, that climate change
80 caused extinctions by facilitating infectious disease (but see 17).

81 Here, we take a hypothetico-deductive approach, linking a theoretical framework,
82 laboratory experiments, and analyses of field data to examine the relationships among
83 extinctions, climate change, and emerging infectious disease. We simultaneously tested six
84 hypotheses or predictors for the climate-related *Atelopus* spp. extinctions: 1) a null model, 2)
85 pathogen alone: temperature-dependent growth of *Bd* in culture, 3) temperature variability alone:
86 annual month-to-month variability in temperature, 4) mean climate alone: annual mean
87 temperature, 5) climate change alone: the 5-year slope of mean temperature, and 6) the
88 interaction between mean historical climate and climate change: because the *thermal mismatch*
89 *hypothesis* predicts that the effect of climate change depends on whether the host is cool or warm
90 adapted, which in turn drives the differential performance of host and pathogen.

91 Given that previous climate change analyses of the *Atelopus* dataset relied on correlating
92 extinctions with regional climate data (12, 13), we first verified that climate change was indeed
93 associated with these extinctions based on data from individual species' ranges. In the
94 geographic ranges of species that went extinct, mean temperatures in the five years leading up to
95 extinction increased ~2.5 times faster than they increased in the ranges of species that remained
96 extant (matched pairs test: $F_{1,45}=7.73$, $p<0.01$; see Methods; Fig. S2; Table S2) (see 18 for a
97 *similar approach using two species*). Hence, soon-to-be extinct species were indeed experiencing
98 conditions that were both unusually warm for them and warmer than those experienced by
99 species that remained extant, consistent with the hypothesis that climate change played a role in
100 *Atelopus* spp. declines.

101 Next, we set out to parameterize our statistical model by conducting laboratory
102 experiments to evaluate the impacts of both mean temperature and temperature variability on
103 *Atelopus* spp. mortality risk. First, we conducted a *Temperature Gradient Experiment* in which
104 we exposed *Atelopus zeteki*, which we consider to be cold-adapted for a variety of reasons (see
105 Supplement), to *Bd* in replicated temperature-controlled incubators (19; Fig. S2) across a
106 naturally relevant temperature gradient (14°, 18°, 22°, 26°, and 28°C) while simultaneously
107 maintaining unexposed frogs and growing *Bd* in liquid cultures in the same incubators. In this
108 experiment, the temperature gradient did not affect *A. zeteki* mortality in the absence of *Bd* (cox
109 proportional-hazards model: $X^2=0.54$, $p=0.46$), but mortality increased significantly with
110 temperature when *A. zeteki* was exposed to *Bd* (*Bd* x temperature: $X^2=4.41$, $p=0.036$). In fact,
111 within a week of exposure to *Bd*, frogs at 26° and 28°C experienced 69% and 78% mortality,
112 respectively, suggesting a temperature-dependent cost of exposure to *Bd* (see Supplement),
113 whereas only one *Bd*-exposed animal died at the two coldest temperatures within a week of
114 exposure (6% mortality) and only four *Bd*-negative animals died throughout the experiment
115 (20% mortality; Fig. 2; Fig. S4). Similarly, *Bd* growth rates on frogs increased with temperature
116 (Fig. 2a). In contrast, temperature-dependent *Bd* growth in culture closely followed previously
117 reported patterns with growth rates increasing as temperature increased until 18.0°C (optimum)
118 and then decreasing thereafter with little growth above 26°C (20; Fig. 2a, Fig. S5). These results
119 demonstrate that patterns of temperature-dependent *Bd* performance in culture and on hosts
120 differ sharply, a result consistent with the *thermal mismatch hypothesis*, which predicts that
121 parasites should have maximum growth on host at temperatures where they most outperform the
122 host rather than at temperatures where the parasite has the greatest absolute performance in
123 culture. The striking monotonic positive association between temperature and both *Bd* growth on

124 frogs and *Bd*-induced host mortality contradict a common assumption that *Bd* outbreaks only
125 occur at cool or moderate temperatures (14, 21). Importantly, although we only tested one
126 *Atelopus* species in this experiment, the observed patterns are likely generalizable to other
127 *Atelopus* spp., because a global analysis of *Bd* prevalence in 15,410 individuals from 598
128 amphibian populations and 1,399 species revealed that cold- and warm-adapted amphibians
129 generally have peak *Bd* prevalences during warm and cold spells, respectively (7; Fig. S6).

130 Second, we conducted a *Temperature Shift Experiment* in which we exposed *A. zeteki* to
131 *Bd* at either constant or variable temperatures to evaluate how temperature variability affects host
132 susceptibility. *A. zeteki* were exposed to *Bd* at 14°, 17°, 23°, or 26°C immediately following
133 either two weeks of acclimation to these temperatures or two weeks of acclimation to 20°C, so
134 that all the hosts experienced either constant or shifted temperatures before *Bd* exposure,
135 respectively. As with the previous experiment, *Bd*-induced mortality increased with temperature
136 (cox-proportional hazards model and ANOVA: $X^2=4.08$, $p<0.05$; Table S3). At the same *Bd*
137 exposure temperatures, frogs that experienced temperature shifts had higher *Bd* loads than those
138 that did not experience shifts (ANOVA: $F_{1,34}=8.78$, $p=0.005$), consistent with the findings of
139 previous studies (19, 22). However, we did not observe any significant effect of the temperature
140 shift treatment on mortality (Shift treatment: $X^2=0.84$, $p=0.36$; Shift \times temperature: $X^2=1.03$,
141 $p=0.31$), and the temperature gradient accounted for >6 times the variance in *Bd*-induced
142 mortality as temperature variability (Table S4).

143 Given the results of our two laboratory experiments, we hypothesized that *Bd* growth in
144 culture, temperature variability, and mean temperature alone would be poor predictors of
145 *Atelopus* extinctions in the wild relative to the *thermal mismatch hypothesis*, which posits that as
146 temperature increases, disease and extinction risk should be most pronounced among *Atelopus*

147 spp. from cooler regions because they should experience a larger performance gap relative to *Bd*
148 than species from warmer regions (Fig. 1). This prediction of the *thermal mismatch hypothesis*
149 would manifest as a statistical interaction between the temperature to which a species is adapted
150 (50-year mean temperature in a species' geographic range) and the level of climate change it has
151 experienced because cold-adapted species should experience disease-associated declines when
152 temperatures increase, whereas warm-adapted species should not. To test these hypotheses, we
153 utilized a time-dependent cox-proportional hazards survival model (23, see *Methods*) that
154 concurrently evaluated the following predictors of the occurrence and timing of extinctions: *Bd*
155 growth in culture, temperature variability, mean temperature, climate change, and the *thermal*
156 *mismatch hypothesis* (see *Methods*). Given that extinction probabilities have repeatedly been
157 shown to be negatively dependent on geographic range size (24), range size was included as a
158 crossed factor with each predictor in our model. The model also controlled for two precipitation
159 variables and altitude, which have been associated with *Atelopus* spp. extinction probabilities
160 (12).

161 Consistent with our experiments, *Atelopus* spp. extinction risk was not significantly
162 explained by interactions between geographic range size and *Bd* growth in culture or temperature
163 variability but was significantly explained by the *thermal mismatch hypothesis* (Table 1).
164 Species with large range sizes rarely experienced extinctions and thus were not strongly
165 impacted by climate or disease. In contrast, species with smaller range sizes showed extinction
166 patterns consistent with the *thermal mismatch hypothesis*. Increasing temperatures associated
167 with climate change (positive slope five years before extinction) were positively associated with
168 the occurrence and timing of the extinction of cold-adapted *Atelopus* spp. (Fig. 3a,c), whereas
169 climate change did not predict the occurrence and timing of declines of warm-adapted species

170 (Range size \times temperature shift \times 40-year mean temperature; $\beta=11.5$, $df=22$, $p=0.02$, Table 1,
171 Fig. 3b,d). In fact, in the absence of any climate change, warm-adapted species were more likely
172 to experience extinctions in cool rather than warm years (Fig. 3b,d), also consistent with the
173 *thermal mismatch hypothesis* (Fig. 1). The model testing the *thermal mismatch hypothesis*
174 explained about 2.5 times more of the variance in extinctions than a model that did not contain
175 the interaction (Nagelkerke's pseudo- $R^2=0.466$ and 0.189 , respectively).

176 To gather further support for the notion that an interaction between climate change and
177 *Bd* drove *Atelopus* extinctions, we compared the magnitude of climate change and extinctions
178 experienced by genus *Atelopus*, which is believed to have been widely exposed to *Bd* and is
179 found in a region of South America where *Bd* has been detected as early 1894 (25), to
180 amphibians in Madagascar and Scandinavia, regions historically considered to be free of *Bd* (26-
181 28). Compared to *Atelopus* spp., amphibian species in Madagascar and Scandinavia experienced
182 similar and more climate change between 1950 and 2004, respectively (T-test: $T=0.118$,
183 $p=0.906$; $T=-5.59$, $p<0.0001$; Fig. S7). However, unlike genus *Atelopus*, there were no
184 amphibian extinctions in these areas during this time (29). This suggests that, in the absence of
185 *Bd*, the same level of climate change experienced in Latin America was insufficient to cause
186 amphibian extinctions in Madagascar and Scandinavia. Although there are major differences in
187 taxonomy and life history between these groups of amphibians, as well as differences in the
188 potential stressors they experienced, these results are consistent with the *thermal mismatch*
189 *hypothesis* and our laboratory findings, which suggest that *Atelopus* declines were caused by an
190 interaction between climate change and *Bd* rather than either stressor alone. However, we
191 caution that these analyses assume a consistent response to climate change across regions, which

192 may not be realistic given that a degree change in temperature has a greater metabolic impact on
193 tropical species than it does on temperate species (30).

194 Our experiments and analyses of field data together suggest that *Atelopus* spp. from
195 cooler environments are more vulnerable to mortality from chytridiomycosis under warmer
196 conditions than those from warmer environments and thus, climate change poses a greater threat
197 to cold-adapted *Atelopus* spp. Therefore, the *thermal mismatch hypothesis* was a useful
198 framework for predicting which species were most likely to be impacted by an interaction
199 between climate change and infectious disease outbreaks in this system. However, the generality
200 of the *thermal mismatch hypothesis* is unknown, as it has only been evaluated in systems with
201 ectothermic hosts and directly transmitted pathogens, which may be especially sensitive to
202 environmental conditions. Further large-scale analyses of disease datasets are needed to test how
203 well the *thermal mismatch hypothesis* applies across host-parasite systems that vary in host
204 thermal biology or mode of transmission.

205 As global temperatures and infectious disease outbreaks have increased, these two crises
206 have been repeatedly correlated by researchers to explain species declines and extinctions.
207 However, evidence that they interact to cause declines has been elusive, possibly because
208 researchers have tried to simplistically correlate increases in temperature with infectious disease,
209 rather than looking for more nuanced patterns that depend on the host-parasite interaction (1).
210 Here, we apply the *thermal mismatch hypothesis*, a framework that can relate environmental
211 temperature to disease patterns while accounting for host-level variation in adaptation to climate
212 to predict which host species are most vulnerable to infectious disease with global warming. By
213 combining experiments with field patterns to examine how mean temperature and temperature
214 variability impact susceptibility to *Bd* in the amphibian genus *Atelopus*, we provide the first

215 evidence that one of the greatest modern day mass extinctions was likely driven by an interaction
216 between climate change and infectious disease.

217

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228 **References:**

229 1. J. R. Rohr *et al.*, *Trends Ecol Evol* **26**, 270 (Jun, 2011).

230 2. C. D. Harvell *et al.*, *Science* **296**, 2158 (Jun 21, 2002).

231 3. M. Pascual, L. F. Chaves, B. Cash, X. Rodo, M. Yunus, *Clim Res* **36**, 131 (Apr 30, 2008).

232 4. X. Liu, J. R. Rohr, Y. M. Li, *Proc. R. Soc. Lond. Ser. B-Biol. Sci.* **280**, (Feb, 2013).

233 5. J. A. Patz, D. Campbell-Lendrum, T. Holloway, J. A. Foley, *Nature* **438**, 310 (Nov,

234 2005).

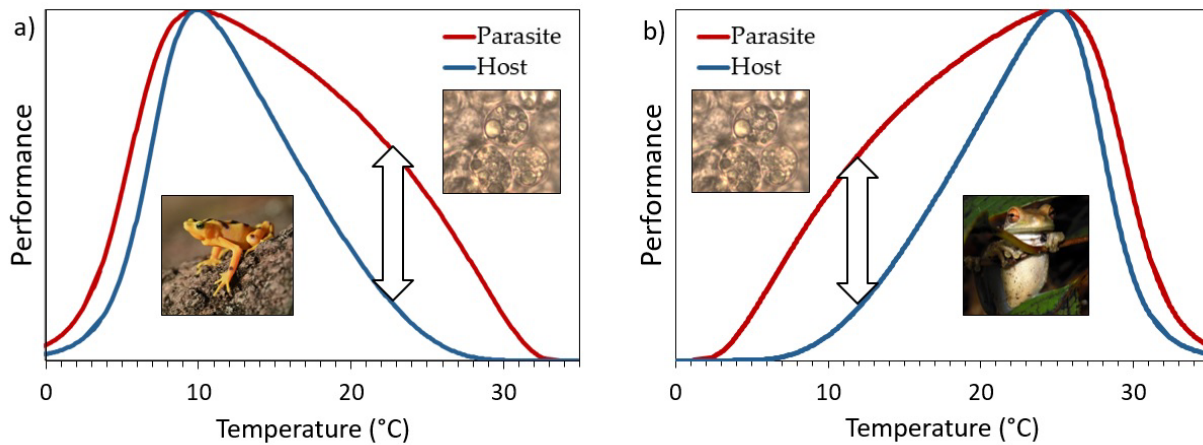
235 6. A. E. Cahill *et al.*, *Proceedings of the Royal Society B-Biological Sciences* **280**, (Jan 7,

236 2013).

- 237 7. J. Cohen *et al.*, *Ecology Letters* **20**, 184 (2017).
- 238 8. J. Rohr *et al.*, (in review).
- 239 9. L. G. M. Baas-Becking, *Geobiologie; of inleiding tot de milieukunde*. (WP Van Stockum
240 & Zoon NV, 1934).
- 241 10. E. La Marca *et al.*, *Biotropica* **37**, 190 (Jun, 2005).
- 242 11. J. P. Collins, *Proceedings of the National Academy of Sciences of the United States of*
243 *America* **110**, 9193 (Jun 4, 2013).
- 244 12. J. A. Pounds *et al.*, *Nature* **439**, 161 (Jan 12, 2006).
- 245 13. J. R. Rohr, T. R. Raffel, *Proceedings of the National Academy of Sciences of the United*
246 *States of America* **107**, 8269 (May 4, 2010).
- 247 14. A. M. Kilpatrick, C. J. Briggs, P. Daszak, *Trends Ecol Evol* **25**, 109 (2010).
- 248 15. L. A. Stevenson *et al.*, *PLoS ONE* **8**, (Sep 4, 2013).
- 249 16. J. Voyles *et al.*, *Oecologia*, (2017).
- 250 17. T. Ben-Horin, H. S. Lenihan, K. D. Lafferty, *Ecology* **94**, 161 (Jan, 2013).
- 251 18. R. A. Alford, K. S. Bradfield, S. J. Richards, *Nature* **447**, E3 (May 31, 2007).
- 252 19. T. R. Raffel *et al.*, *Nature Climate Change* **3**, 146 (Feb, 2013).
- 253 20. D. C. Woodhams, R. A. Alford, C. J. Briggs, M. Johnson, L. A. Rollins-Smith, *Ecology*
254 **89**, 1627 (2008).
- 255 21. L. Berger *et al.*, *Fungal Ecol* **19**, 89 (Feb, 2016).
- 256 22. T. R. Raffel, N. T. Halstead, T. A. McMahon, A. K. Davis, J. R. Rohr, *Proc. R. Soc.*
257 *Lond. Ser. B-Biol. Sci.* **282**, (Feb 22, 2015).
- 258 23. T. Therneau, (2014).
- 259 24. C. D. Thomas *et al.*, *Nature* **427**, 145 (Jan 8, 2004).

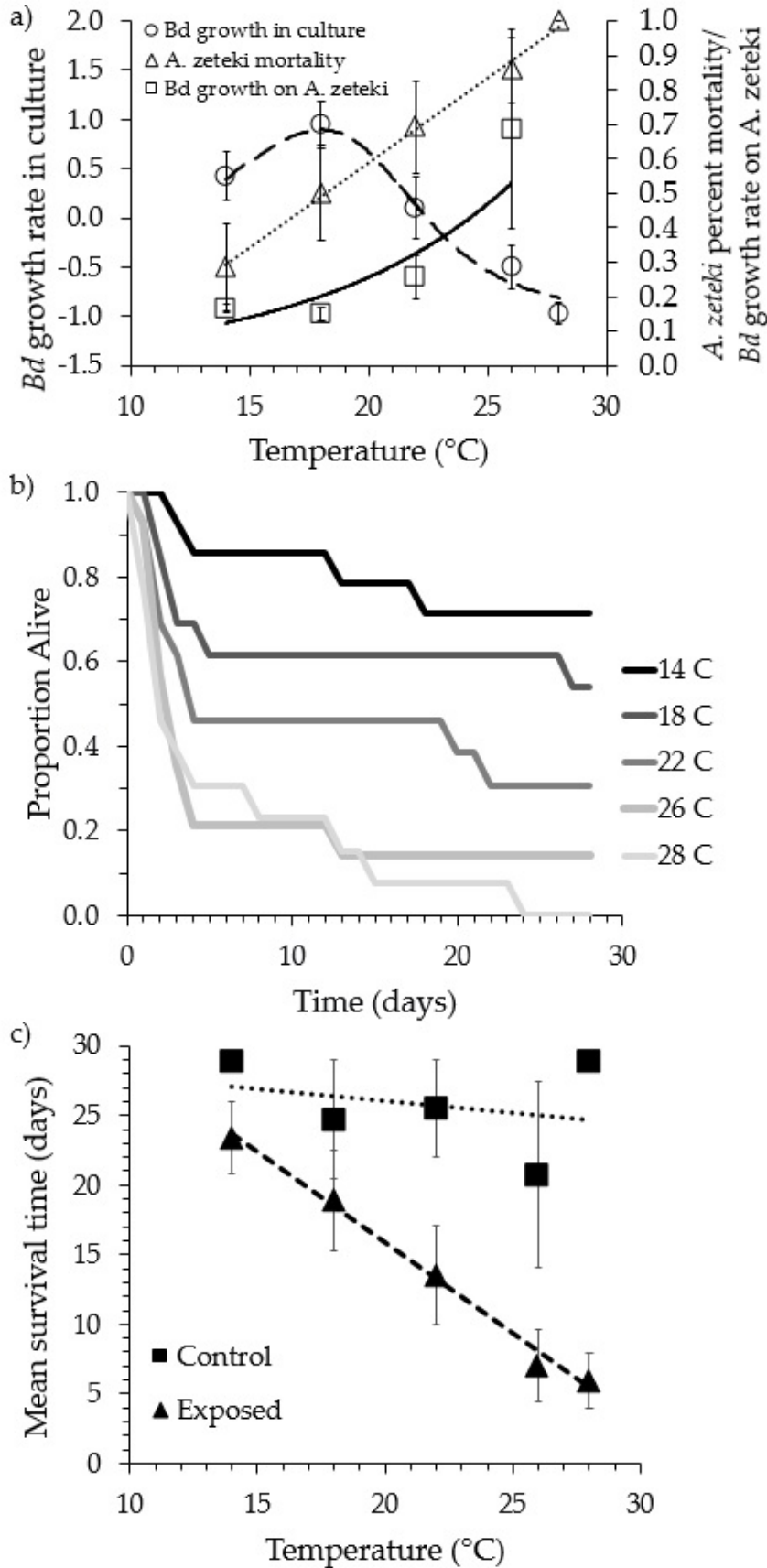
- 260 25. D. Rodriguez, C. Becker, N. Pupin, C. Haddad, K. Zamudio, *Molecular Ecology* **23**, 774
261 (2014).
- 262 26. J. E. Kolby *et al.*, *PLoS ONE* **10**, (Jul 29, 2015).
- 263 27. F. Andreone *et al.*, *Conservation Biology* **19**, 1790 (Dec, 2005).
- 264 28. www.bd-maps.net. (2017).
- 265 29. N. S. Sodhi *et al.*, *PLoS ONE* **3**, (Feb 20, 2008).
- 266 30. M. E. Dillon, G. Wang, R. B. Huey, *Nature* **467**, 704 (2010).
- 267
- 268

269 **Figures:**

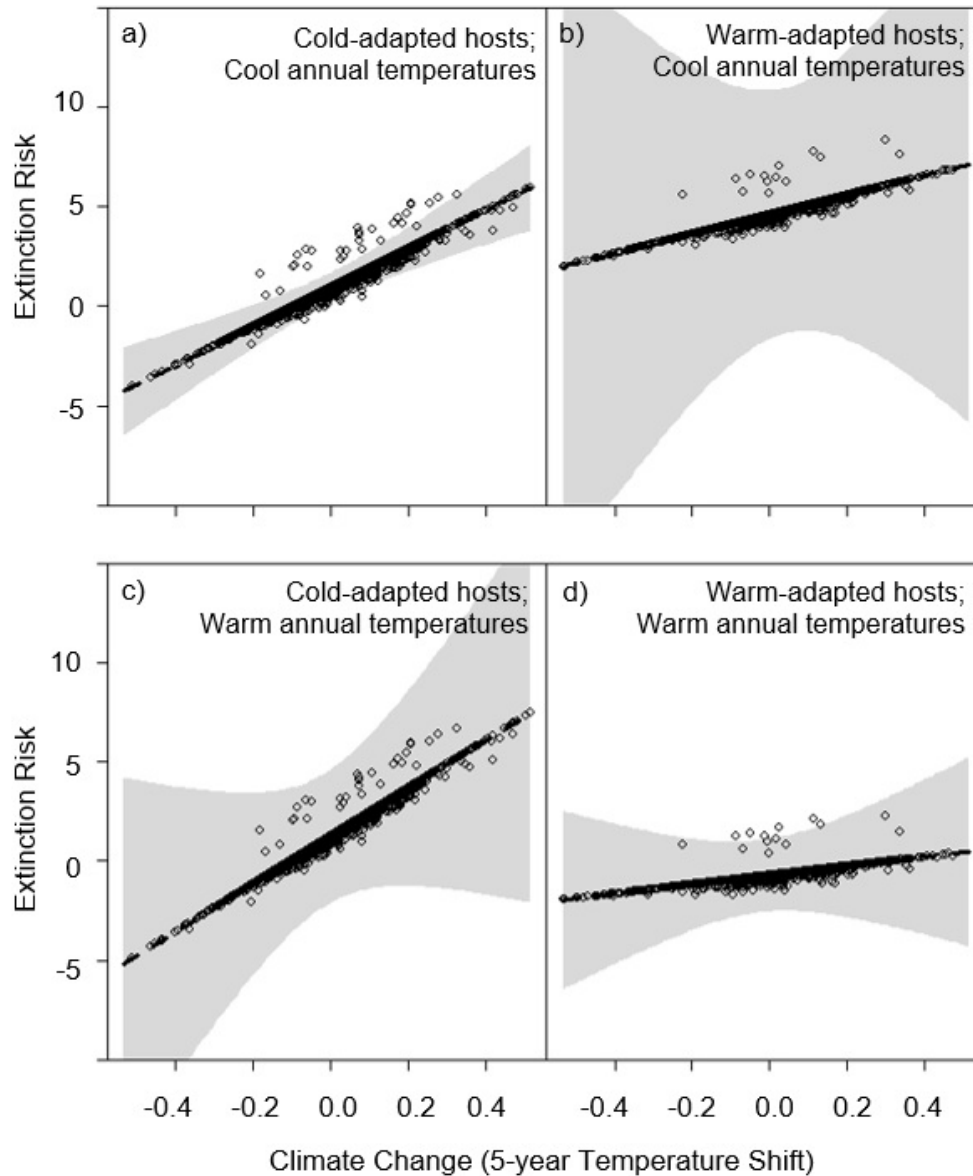


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272 **Fig. 1. Conceptual figure describing the *thermal mismatch hypothesis*.** In isolation, small
273 organisms, such as parasites (red lines), generally have broader thermal performance curves than
274 larger organisms, such as hosts (blue lines). Parasite growth on hosts is likely to occur at
275 temperatures where a parasite most outperforms its host (bidirectional arrows), and not
276 necessarily at the temperature at which a parasite performs best in isolation, providing a
277 hypothesis for the thermal performance curve of a parasite growing on the host. For interacting
278 cold-adapted hosts and parasites (**a**), parasite growth should be maximized at relatively warm
279 temperatures, whereas for interacting warm-adapted hosts and parasites (**b**), parasite growth is
280 predicted to be maximized at relatively cool temperatures.



282 **Fig. 2. Temperature-dependent patterns of *Batrachochytrium dendrobatidis* (*Bd*) growth**
283 **and *Atelopus zeteki* mortality. (a)** *A. zeteki* experienced high mortality (triangles, dotted line)
284 and high *Bd* growth (squares, solid line) at warm temperatures after *Bd* exposure, even though
285 *Bd* growth rates in culture (circles, dashed line) were low at these temperatures. We could not
286 measure *Bd* growth rates on *A. zeteki* at 28 °C because very few animals survived long enough to
287 be tested multiple times. **(b)** Survival plot for *A. zeteki* exposed to *Bd* across two temporal
288 blocks. **(c)** Mean survival time for *A. zeteki* at each of five temperatures when exposed to *Bd*
289 (triangles; both temporal blocks) or not exposed (squares; 2nd temporal block). Temperature and
290 *Bd* exposure interacted to induce high mortality in *A. zeteki* ($X^2=4.41$, $p=0.036$). Animals
291 surviving the experiment are conservatively assumed to have died on day 29 for these figures
292 only but were censored in the survival analysis. Error bars represent SEMs in all panels.



293

294 **Fig. 3. Partial residual plot displaying the effects of climate change and annual mean**
295 **temperature on the extinction risk of cold- and warm-adapted *Atelopus* spp.** The partial
296 residuals are from the time-dependent cox proportional-hazards model shown in Table 1 and
297 display the significant three-way interaction among 5-year slopes in mean temperature-by-40-
298 year mean temperature-by-annual mean temperature. Points represent individual years for each
299 species and gray shading shows associated 95% confidence bands. The model suggests that
300 species from typically cooler climates (**a,c**) were at greater risk of extinction (log-odds risk ratio;
301 y-axis) after experiencing climate change (warming, or positive 5-year slope in mean
302 temperature; x-axis) than species from warmer climates (**b,d**; breaks based on 20th and 80th
303 percentiles). This pattern was consistent whether the annual mean temperature in a given year
304 was relatively cool (**a,b**) or warm (**c,d**; breaks based on 20th and 80th percentiles), although
305 species from warmer climates may be at greater risk when conditions are cooler (**b**).

306 **Tables:**

307 **Table 1.** Results of time-dependent cox-proportional hazards model predicting extinction with a
 308 four-way interaction between log-transformed range size, long-term mean temperature
 309 (40yr.meantemp), annual mean temperature (meantemp), and recent temperature shift
 310 (tempchange) across both extinct and extant *Atelopus* spp. Mortality probability based on *Bd*
 311 growth in culture (culturemortprob), log-transformed altitude (logaltitude) and a measure of
 312 temperature variability (log-transformed AVMD, absolute value of monthly difference in
 313 temperature) were also included. Bolded lines represent tests of the *thermal mismatch*
 314 *hypothesis*.

| | Coefficient | SE | Robust SE | z | p |
|--|---------------|--------------|--------------|---------------|-------------|
| Rangesize | -23.500 | 72.300 | 51.500 | -0.460 | 0.64 |
| Logaltitude | -0.124 | 0.283 | 0.147 | -0.840 | 0.39 |
| Culturemortprob (pathogen only) | -1.130 | 1.480 | 1.220 | -0.930 | 0.35 |
| Total precipitation | <0.001 | <0.001 | <0.001 | 0.840 | 0.40 |
| Frequency of wet days | <0.001 | <0.001 | <0.001 | -0.280 | 0.77 |
| Log(AVMD) (temp. variability) | 0.913 | 0.756 | 0.719 | 1.270 | 0.20 |
| Tempchange (climate change only) | 11.300 | 35.600 | 37.600 | 0.300 | 0.76 |
| 40yr.meantemp (cold or warm adapted) | 1.310 | 0.901 | 0.759 | 1.730 | 0.08 |
| Meantemp (mean temp. only) | 0.564 | 0.972 | 0.756 | 0.750 | 0.45 |
| Rangesize:Culturemortprob | -4.070 | 9.130 | 6.690 | -0.610 | 0.54 |
| Rangesize:log(AVMD) | -1.410 | 2.030 | 1.810 | -0.780 | 0.43 |
| Rangesize:Tempchange | -160.000 | 155.000 | 82.800 | -1.930 | 0.05 |
| Rangesize:40yr.meantemp | 1.460 | 4.380 | 2.890 | 0.510 | 0.61 |
| Tempchange:40yr.meantemp | -0.080 | 2.650 | 2.710 | -0.030 | 0.97 |
| Rangesize:meantemp | 3.130 | 5.610 | 4.040 | 0.770 | 0.43 |
| Tempchange:meantemp | 0.420 | 2.380 | 2.200 | 0.190 | 0.84 |
| 40yr.meantemp:meantemp | -0.051 | 0.040 | 0.033 | -1.580 | 0.11 |
| Rangesize:Tempchange:40yr.meantemp | 11.500 | 9.800 | 5.060 | 2.280 | 0.02 |
| Rangesize:Tempchange:meantemp | 5.680 | 9.480 | 6.130 | 0.930 | 0.35 |
| Rangesize:40yr.meantemp:meantemp | -0.146 | 0.280 | 0.195 | -0.750 | 0.45 |
| Tempchange:40yr.meantemp:meantemp | -0.028 | 0.102 | 0.111 | -0.250 | 0.80 |
| Rangesize:Tempchange:40yr.meantemp:meantemp | -0.464 | 0.467 | 0.224 | -2.070 | 0.03 |

315