Biodegradation of Chlorpyrifos and 3,5,6-trichloro-2-pyridinol (TCP) by *Ochrobactrum* sp. CPD-03: Insights from genome analysis on organophosphorus pesticides degradation, chemotaxis and PGPR activity

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Abstract

Organophosphorus pesticides (OPs) are xenobiotic chemicals, present in the environment at unnaturally high concentrations. Chlorpyrifos (CP) is a widespread OP pollutant that poses hazards to non-target organisms along with human beings. The present study reports a newly isolated strain, *Ochrobactrum* sp. CPD-03 form a paddy field in Bhubaneswar, Odisha, India which could degrade CP in 48 hours with a degradation efficiency of 85-88% in minimal salt medium. The CP degradation at various concentrations (100 mg L^{-1} to 1000 mg L^{-1}) by CPD-03 was curve fit into a pseudo-first order kinetics [$Y= (Y_0$ - Plateau)$*e^{(-K*X)}$ + Plateau)]. Whole genome analysis indicated the presence of arylesterase (EC 3.1.1.2) and showed the existence of several other genes accountable for the various other xenobiotic compounds degradation. CPD-03 showed chemotactic features towards CP at different concentrations along with other OPs including monocrotophos, coumaphos, diazinon, methyl parathion, and malathion, respectively. Moreover, few new metabolites during CP degradation by CPD-03 were found. This strain was also found to degrade a major metabolite, TCP (3,5,6-trichloro-2-pyridinol) during CP degradation. CPD-03 also displayed potential plant growth promoting activity in inoculated rice seedlings. Taken altogether, our results highlight the potentials of this new isolate *Ochrobactrum* sp. CPD-03 in bioremediation and application in OP-contaminated ecosystem.

Keywords: Biodegradation, Chlorpyrifos, *Ochrobactrum* sp. CPD-03, Organophosphorus hydrolase, Chemotaxis, Plant growth promotion

1. Introduction

Extensive use of organophosphorus pesticides (OPs) have triggered alarm over environmental pollution and food safety, apart from contact toxicity. OPs are the most toxic substances known and are widely used for insect pest control [1]. Corresponding to this, release of pesticides into the atmosphere leads microbes to either detoxify the residues of these toxic compounds and their metabolites or use them as new sources of carbon and energy sources. Mineralization of these OPs is significantly more challenging since microorganisms need to use specific enzymes to turn such pesticides into certain intermediates in a key metabolic pathway [2]. However, given adequate exposure time, microbial cells may develop pathways to mineralize pesticides that have been implemented [3].

Chlorpyrifos (*O,O*-diethyl *O*-3,5,6-trichloropyridin-2-yl phosphorothioate) is moderately toxic [4] among OPs commonly used for protecting agricultural crops from insects, fungi, weeds and other pests, which affect the economic growth of crops. It enters inside the insects through ingestion, by contact or absorbed through cuticles, gut [5] and affects the nervous system by inhibiting the acetylcholinesterase activity, ultimately causing death of the target insect [6]. Chlorpyrifos (CP) hydrolysis leads to the formation of TCP (3,5,6-trichloro-2-pyridinol), a key metabolite having considerable water solubility [7, 8]. TCP has a half-life between 65-360 days in soil [9], water solubility owing to its neutral pH [10], creates comprehensive soil and aquatic habitat contamination, posing important environmental ecotoxicological hazards relative to CP [11]. The enhanced resistance to microbial degradation of TCP is due to the existence of three chloride in its chemical structure [12, 13] and upon release of these chlorine atoms inhibits the growth of the CP degrading microbes in the neighboring environment [14].

Living species being exposed to these CP residues present in soil and water, affects

ecosystem balance [15]. CP toxicity has been shown to be liable for human health, including

nervous system disorders, immune system defects, liver damage and endocrine disruption

[16, 17]. Hence, remediation of the CP contaminated sites to mitigate the effects of these

toxic residues are highly important. Certain causes of pollution of this pesticide include

industrial effluents, product waste, leakage, accidental spills and remediation. Therefore,

techniques to degrade CP and TCP from contaminated fields need to be established in order

to reduce the adverse effects on human health, non-target species, and the ecosystem.

Bioremediation is one of the most promising methods due to its low cost, simple to use along

with eco-friendly for decontaminating harmful pesticides [18]. Microbial pesticide

degradation was identified as a primary mechanism for soil and water dissipation. Microbes

containing the appropriate metabolic pathways play a key role in CP and TCP degradation

[19, 20]. In the past few years several bacteria have been reported for degrading CP or TCP.

These include *Enterobacter* sp. [21], *Pseudomonas* sp. [22, 23], *Bacillus* sp. [24, 25],

Klebsiella sp. [26], *Sphingomonas* sp. [27], *Paracoccus* sp. [28], *Cupriavidus* sp. [29],

Brucella sp., *Klebsiella* sp., and *Serratia marcescens* [30], *Stenotrophomonas* sp. [31] and

Stenotrophomonas sp. [32]. There are some instances where the bacterial species could lose

its degradation ability under high pressure of TCP, thus making difficult for CP degradation

in contaminated areas [33]. Recently, *Ochrobactrum* sp. JAS2 was reported showing the

ability of CP & TCP degradation in spiked soil within 24 h & 94 h respectively along with

plant growth promoting ability [34].

Only a few bacterial genera have been investigated for degrading CP and TCP. The main

objective of this study was therefore to screen the active microbial strain(s) from paddy field

soil for CP and TCP degradation followed by genome analysis and the formation of new metabolite signatures by GC-MS. This analysis would deepen our understanding and would provide valuable source of information to expand its potential as a biodegrader of OPs pesticides.

2. Materials & methods

2.1 Soil Samples, cultivation media and chemicals

Soil samples were collected from a paddy field near Kalinga Institute of Social Science (KISS), Bhubaneswar Odisha, India [20.3551° N, 85.8187° E]. Preliminary screening of CP tolerant bacteria was performed using minimal salts media (MSM) [Per litre contains; NH_4Cl, 0.5g/L; $MgSO_4.7H_2O$, 0.3g/L; $KH_2PO_4.H_2O$, 0.2g/L; $FeSO_4.7H_2O$, 0.01g/L; $Ca(NO_3)_2.4H_2O$, 0.01g/L;] supplemented with CP (100 mg L^{-1}). Analytical grade CP (PESTANAL® 99.99%,) and TCP (PESTANAL® 99.99%) were procured from Sigma-Aldrich, USA. Another commercial grade CP (TRICEL 20% EC and HILBAN 50% EC) were also purchased from local market for soil enrichment purpose. Chemicals and other reagents used in this study were of analytical grade (HiMedia, India and Merck, Germany).

2.2 Screening & isolation of CP degrading bacteria

CP degrading bacteria was isolated from rhizospheric soil by culture enrichment technique. Briefly 100 g of soil treated with CP (100 mg l^{-1}) was inoculated into a 200 ml of MSM media. These soil samples were incubated and enriched for bacterial culture at 30±2 °C in a shaker incubator at 120 rpm for a week. Post incubation, 10 ml of grown culture was transferred for another week of incubation to new MSM amended with CP (200 mg L^{-1}). Likewise, three consecutive transfers in sterile MSM supplemented with CP (100 mg L^{-1}) was performed upon serially dilution and plated on solid media (MSM with CP @100 mg L^{-1}).

Post 96 h incubation, the pure colonies were obtained by repeated streaking. Isolate CPD-03

was further selected based on its CP degradation potential. Degradation efficiency was

calculated [(Residual amount in blank control − residual amount in sample)/ (Residual

amount in blank control) x 100]

2.3 Morphological, biochemical and taxonomic identification of the bacterial strain

CPD-03

Isolate CPD-03 was preliminarily characterized by morphological parameters by scanning

electron microscope (Nova^{TM} NanoSEM 450) and Gram staining (HiMedia, India) followed

by biochemical assays using BIOLOG$^®$ [35]. Genomic DNA was extracted followed by a

16S rRNA sequencing [36] was performed. The 16S rRNA gene was obtained by PCR

amplification with universal bacterial primers 27F and 1492R. PCR reaction was set at 95°C

denaturation for 5 min, followed by 30 cycles of 95°C for 45s, 55°C for 45s, and 72°C for

90s, and a final step of 72°C for 5 min. The PCR amplified product was purified and

sequenced by SciGenome, Ahmedabad, India. The obtained 16s rRNA gene sequence was

aligned with the 16S rRNA sequences of type strains with homology of 97% or higher with

CPD-03 and were obtained from EzTaxon (eztaxon-e.ezbiocloud.net). The sequence

alignment was performed using ClustalX [37]. Phylogenetic tree was constructed with

MEGA version 7.0 [38].

2.4 Inoculum preparation (Resting cell assay)

Pure cultures of bacterial strain CPD-03 were cryopreserved in 25% glycerol at -80^{0}C. The

bacterial strain was revived and inoculated in a 100 ml Erlenmeyer flask with 20 ml sterile

MSM supplemented with CP (100 mg L^{-1}) before each set of experiments. Flask was

incubated at 30 ± 2 °C with 120 rpm in a shaker incubator. The bacterial cells were harvested

after 44- 48h at 4500 rpm for 5 min followed by washing with sterile saline (0.9% NaCl). The

washed bacterial cells were resuspended in sterile MSM to obtain a cell density of $1-1.5\times10^8$

cells ml^{-1}. This bacterial resting cell suspension was used for further biodegradation

experiments, organophosphorus hydrolase (*opd*) gene localization, chemotaxis assays and

plant growth promoting response (PGPR) activities.

2.5 Biodegradation of CP & TCP in aqueous medium

Biodegradation of CP (100 mg L^{-1}) & TCP (100 mg L^{-1}) was studied (in separate and in

mixed) in 50 ml resting cell suspension (Supplementary Fig 1a&b). An additional carbon

source molasses (0.1% v/v) [39] was added to enhance the growth. MSM with the same

concentration of CP and 0.1% molasses was treated as control. All the experimental flasks

were kept for incubation at $30\pm2°C$ and 120 rpm. Degradation under various concentration of

CP (100 mg L^{-1} to 1000 mg L^{-1}) was studied in 50 ml resting cell as mentioned above were

monitored till 12 h in an interval of 2 h and later till 48h in an interval of 6 h for determining

the rate of CP degradation using high performance liquid chromatography (HPLC) (Agilent

Technology 1260 Infinity, USA).

2.6 Sample extraction and HPLC analysis

The resting cell along with CP (100 mg L^{-1} along with molasses (0.1% v/v) from

biodegradation assay, was harvested at 7500 rpm for 10 min for obtaining a cell-free media

and extracted with organic solvents. The supernatant was pulled out for extraction with an

equal volume of ethyl acetate and this step was repeated twice. The organic layer thus

obtained was aspirated, pooled and dehydrated over a column of anhydrous sodium sulfate,

evaporated at $45\pm2°C$ using rotavapor (Eyela, Japan), reconstituted in methanol (HPLC

Grade) and filtered through 0.22 μm membrane filter. Residual CP and TCP was quantified

using a reverse-phase C18 column (ZORBAX Eclipse Plus, Agilent Technologies, USA)

fitted with Agilent Technologies 1260 Infinity HPLC system equipped with a binary pump,

UV/VIS detector, thermostat column compartment (TCC) column oven and auto sampler,

with array detection based on peak area and retention time of the pure standard. The mobile

phase used for the analysis contained a mixture of methanol and water in a ratio of 85:15 at a

flow rate of 1.2 mL min^{-1}. Water (HPLC grade) was acidified to pH 3.2 using

orthophosphoric acid. Injection volume was 20 μL. The recoveries of OPs added at

concentrations of 1, 2, 5, 10, 50 mg/L, ranged from 97.0% to 100.0%. The linear range of the

calibration curves was obtained from 1 to 50 mg/L.

2.7 Degradation kinetics

CP degradation rates has been fitted with a pseudo first order kinetic model $Y = (Y_0 -$

Plateau)$*e^{(-K*X)}$ + Plateau; Y_0 = the Y value when X (time) is zero and is expressed in the

same units as Y, Plateau = the Y value at infinite times and expressed in the same units as Y.

K = rate constant and expressed in reciprocal of the X axis time units. Half-life is in the time

units of the X axis. It is computed as $\ln(2)/K$ using statistical software Prism8.

2.8 Identification of metabolites

The resting cell containing CP (100 mg L^{-1}) along with molasses (0.1% v/v) from

biodegradation assay, was also assessed for the identification of CP degradation and

simultaneous metabolites formation. The extraction was performed as mentioned above in

section 2.6. Samples were analysed in gas chromatography-mass spectrometry (GC-MS)

(Agilent Technologies, USA) fitted with a HP-5MS capillary column (30.0 m x 250 μm x

0.25 μm) in mass spectrometer (Agilent 7890B ION TRAP MS) equipped with a manual

injector containing split/split less capillary injection system along with an array detection of

total scan from 30-500 nm. The column temperature was kept for 5 min at 80^0C initially and

increased for 5 min at 8^0C min-1 to 200^0C, then increased for 5 min at 15^0C min^{-1} to 260^0C.

The ionization energy was kept at 70 eV, and the temperatures of the transfer line and the ion

trap were kept 280^0C and 230^0C, respectively. The amount of sample injection was 1.0 µL at

250^0C with 1:50 split mode. Helium was used as a carrier gas at a flow rate of 1.5 ml min^{-1}.

The CP and intermediates found by the GC-MS study matched authentic standard compounds

from the National Standards and Technology Institute (NIST, USA) database

(http://chemdata.nist.gov).

2.9 Molecular identification & localization of organophosphorus hydrolase (*opd*) gene

For identification of organophosphorus hydrolase (*opd*) gene, the resting cell (1-1.5x10^8 cells

ml^{-1}) was used for genomic DNA isolation as mentioned section 2.3. The *opd* gene of strain

CPD-03   was   amplified   by   PCR   with   forward   primer   (OPDF,   5′-

GCGGATCCATGCAAACGAGAAGGGTTGTGC-3′,   and   reverse   primer   OPDR,   5′-

GCCTCGAGTCATGACGCCCGCAAGGTC-3′; *BamHI* and *XhoI* restriction sites,

respectively, are underlined). The reagents used for PCR were obtained from Promega

(Madison, WI) and a total volume of 25 µl reaction mixture for PCR amplification was

carried out. Each reaction mixture consists of $2\times$ green reaction mixture (10 mM MgCl$_2$, 10

µM each dNTPs, 1 U of Taq DNA polymerase along with 10x buffer), 10 pmol of each

primer along with the template DNA and nuclease free water for volume makeup. The PCR

amplification program was as follows: initial denaturation at 94°C for 5 min, 32 cycles of

cyclic denaturation at 94°C for 60s followed by annealing at 55.1°C for 45s and extension at

72°C for 60s including a final extension at 72°C for 5 min.

To localize the presence of *opd* gene, CPD-03 resting cells (1-1.5x10^8 cells ml^{-1}) were

centrifuged and processed as per Deng et al., 2015 with a slight modification followed by

estimation with enzymatic assays [40]. Briefly, CPD-03 resting cells were harvested at

$8500 \times g$ for 15 in 4^0C. Cell-free extract was obtained by passing the supernatant through a

0.22µm membrane filter and hence treated as extracellular (EC) and stored at 4^0C. The cell

pellet was washed twice with normal saline in the Tris-Cl buffer solution (50 mmol / L, pH

8.0) and re-suspended to obtain cell suspension (CS). One half of the CS was stored at 4^0C,

and the other half was re-suspended in Tris-Cl buffer and sonicated (4^0C, 400W) for 8min.

The collected suspension was then centrifuged at $8000 \times g$ for 10 min in 4^0C. The resulting

supernatant was filtered through a 0.22µm to obtain crude intracellular (IC) enzyme. *Opd*

activity was calculated as the number of moles of *p*-nitrophenol produced in unit time.

2.10 Assessment of plant growth promoting activity

237　Evaluation of the effect of CPD-03 resting cell ($1-1.5 \times 10^8$ cells ml^{-1}) on plant growth was

238　done by quantitative measurement of root length in inoculated rice seedlings followed by

239　indole acetic acid (IAA) production [41], ammonia production [42], and HCN production

240　[43, 44]. The data were subjected to statistical analysis using Prism8.

2.11 Chemotaxis towards CP

The chemotactic property of CPD-03 resting cell ($1-1.5 \times 10^8$ cells ml^{-1}) towards CP along

with other OPs was examined qualitatively in swarm plate assays and quantitatively with

capillary tube assay as per the procedure described earlier [45, 46]. Citrate (0.5%) was used

as the positive control [47]. The chemotactic response was expressed as chemotaxis index

(CI), and calculated as the ratio of the number of CFUs obtained from the capillary tube

containing the test compound (CP) to CFUs obtained from a control capillary that contains

only the chemotaxis buffer (100 mM potassium phosphate pH 7.0 and 20 µM EDTA; without

any chemotactic compound).

2.12 *De novo* genome sequencing and assembly

Illumina HiSeq System 2000 (Illumina, Inc.) was used for whole-genome sequencing. Libraries were generated in a paired end manner. Data obtained from sequencing were *de novo* assembled followed by entire quality control performed using NGSQC Toolkit. Velvet (V 1.2.10) [48] was used for primary assembly that resulted a total size of the contig of 4,65,9698 (~4.6Mb). *De novo* Genome Validation and Quality Control was performed by Bowtie2 (v 2.2.2) [49]. Quality control of assembled genome based on genomic elements were performed by ARAGORN (v1.2.36) [50]. Genome Annotation and functional characterization were performed using RAST. All the bio-informatics work was performed in collaboration with Bionivid Technology Pvt. Ltd., Bangalore, India. Genome sequence including the necessary information for the CPD-03 strain was deposited in NCBI under the accession number RSEU00000000. Considering the ability of this strain for degrading CP, the whole genome of this strain was checked for the occurrence of the CP degrading core genes/proteins. The CPD-03 genome was also compared with other eight more *Ochrobactrum* strains obtained from the NCBI database.

3. Results & discussion

3.1 Isolation and identification of CP degrading bacterial strain CPD-03

Several bacterial strains were isolated from rhizosphere soil of paddy fields near KIIT University area which could utilize CP as carbon source for their growth. CPD-03 was distinguished from other isolates for its ideal growth response during isolation. Further CPD-03 (1.2×10^8 cfu/ml) showed a CP degradation efficiency of 82 ±2% in MSM supplemented with CP (100 mg L^{-1}) (Supplementary Fig.1a) and hence selected for further study. CPD-03

was able to degrade TCP (100 mg L^{-1}) separately with a degradation efficiency of 60 ±2%

(Supplementary Fig.1b).

CPD-03 is Gram-negative and showed regular rod shaped cells which was observed in

scanning electron microscopy (Fig. 1a). A detailed comparison of the physiological and

biochemical properties of CPD-03 with its closest phylogenetic species of *Ochrobactrum* was

performed (Supplementary Table 1a-c). A phylogenetic tree including strain CPD-03 and

related type strains within the clade of *Ochrobactrum* (obtained from EzBioCloud) was also

compared (Fig. 1b). 16S rRNA gene sequence analysis illustrated that, CPD-03 shared

maximum homology of 99.8% with *Ochrobactrum intermedium* LMG 3301^{T} followed by

99.2% with *Ochrobactrum ciceri* Ca-34^{T} of the genus *Ochrobactrum* [51], hence designated

as *Ochrobactrum* sp. CPD-03. The 1396 bp 16S rRNA nucleotide sequence of CPD-03 strain

was submitted in GenBank under Accession No. KT366923.

Fig.1. a. Scanning electron microscopy of CPD-03; b. phylogenetic tree of CPD-03 based on

comparison of 16S rRNA sequence similarity of closely related *Ochrobactrum* sp. conducted

in MEGA version 7.0

3.2 Biodegradation of CP and TCP

Degradation of CP (100 mg L^{-1}) and TCP (100 mg L^{-1}) when added together were assessed using CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in MSM along with molasses (0.1% v/v). This resting cell approach showed CP and TCP degradation efficiency of 88±2% and 62±2%, respectively in 48 h (Fig. 2a). TCP degradation was monitored separately using CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in MSM along with molasses (0.1% v/v) with a degradation efficiency of 62.2±2% (Fig. 2b). Earlier reports have shown TCP can be degraded by *Alcaligenes* sp. [52], *Burkholderia* sp. [53], *Pseudomonas* sp. [54], *Pseudomonas* sp., *Bacillus* sp. and *Agrobacterium* sp. [55], *Trichosporon* sp. [56], and *Pseudomonas* sp. [57] when used as sole carbon source. Previous report has shown *Ochrobactrum* sp. JAS2 had degraded not only CP and TCP in aqueous medium within 96 h when used together. The present study showed, CPD-03 could degrade CP at various concentrations (100 mg L^{-1} to 1000 mg L^{-1}) and it was clearly evident that higher concentration reduced the CP degradation efficiency (Fig. 2c). This correlates with a previous report where higher concentration of pesticides has inhibited the bacterial growth hence decreasing the degradation rate.

Fig.2. a. Degradation of CP (100 mg L^{-1}) and TCP (100 mg L^{-1}) by CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in a MSM supplemented molasses (0.1% v/v) showing formation of its intermediate TCP and its further degradation; b. degradation of TCP (100 mg L^{-1}) by CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in a MSM supplemented molasses (0.1% v/v); c. degradation of CP at various concentrations (100 mg L^{-1}, 200 mg L^{-1},400 mg L^{-1},600 mg L^{-1},800 mg L^{-1}, and 1000 mg L^{-1}) by CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in MSM supplemented molasses (0.1% v/v); d. comparision of CP (100 mg L^{-1}) degradation by CPD-03 resting cells (1-1.5x10^8 cells ml^{-1}) and LMG 3301^T (1.2-1.4x10^8 cells ml^{-1}). Data were represented in the mean of three replicates and analysis were performed using ANOVA with the Prism8. All were tested at the $p < 0.005$*** significance level

CP degradation using CPD-03 resting cell in MSM supplemented with molasses (0.1%) revealed very good compliance with pseudo first order kinetics. Various pesticide

degradation studies were previously reported following a pseudo first order kinetics [58, 59].

A recent study had reported the CP degradation by *Ochrobactrum* sp. JAS2 also followed a

pseudo first order kinetics [34]. CPD-03 degraded 100 mg L^{-1} CP with a half-life of 6.559 h

and TCP 100 mg L^{-1} with a half-life of 12.08 h. CPD-03 was compared with its

phylogenetically closest neighbor *Ochrobactrum intermedium* LMG 3301^{T} for its ability to

degrade CP (100 mg L^{-1}) and was found to show higher degradation efficiency of 88.34% as

compared to LMG 3301^{T} with a degradation efficiency of 60.34% (Fig. 2d).

3.3 Identification of metabolites

Degradation of CP by *Ochrobactrum* sp. CPD-03 was monitored in GC-MS, has resulted in

some new metabolites (Fig. 3). Previous studies have shown the formation of TCP upon CP

mineralization and its immediate degradation upon subsequent hydrolysis steps through ring

cleavage [15]. However, in our study, few new downstream metabolites like (a) cis-Vaccenic

acid, (b) Octadecanoic acid, (c) Benzoic acid, 4-hydroxyl 3-methoxy, (d) n-Hexadecanoic

acid, (e) 9-Hexadecanoic acid were obtained after CP mineralization which could add new

insights into the metabolic pathways. cis-Vaccenic acid is a byproduct from the activity of 3-

oxoacyl-[acyl-carrier protein] reductase which was present in CPD-03 genome. The presence

of this enzyme has previously been shown in case of poly aromatic hydrocarbon (PAH)

degradation [60]. Similarly, Octadecanoic acid was found to be a metabolite during

biodegradation of endocrine-disrupting chemicals in earlier reports [61], Benzoic acid, 4-

hydroxyl 3-methoxy was reported in beta-cypermethrin degradation [62], n-Hexadecanoic

acid production and degradation due to microbial activity has been studied [63] and 9-

Hexadecanoic acid was reported to produce during bioremediation of *n*-alkanes [64]. This

could possibly draw new details into the metabolic pathway in terms of pesticide degradation.

Fig.3 GC-MS chromatograms showing the metabolites formed during CP (100 mg L^{-1})

degradation by CPD-03 resting cells (1-1.5x10^8 cells ml^{-1}). A: 0 h, B: 2 h, C: 4 h, D: 6 h, E: 8

h, and F: 10 h respectively. 1: chlorpyrifos, 2: cis-Vaccenic acid, 3: Octadecanoic acid, 4:

Benzoic acid, 4-hydroxyl 3-methoxy, 5: n-Hexadecanoic acid, 6: 9-Hexadecanoic acid,

respectively.

3.4 Identification and localization of *opd* gene in CPD-03

CPD-03 was found to contain the gene encoding organophosphorus hydrolase (*opd*, ~1098 bp), which was successfully amplified using the primers OPDF and OPDR (Fig. 4a). This genes could possibly transform the complex OPs to generate *p*-nitrophenol have been previously reported [65]. Since CPD-03 is a Gram-negative bacteria, organophosphorus hydrolases potentially are secreted into the extracellular space from periplasmic region, hence can be extracted in more quantities [66]. A maximum amount of *p*-nitrophenol (2±0.2 mmol/L) produced by treatment of the CPD-03 resting cells (extracellular cell free content, EC) with CP for 30 min (Fig. 4B) as compared to the intracellular and cell suspension. This result is consistent with the earlier report of *Flavobacterium* sp. MTCC 2495 [67].

Fig.4. a. Amplification of ~1098 bp *opd* gene; b. enzyme activity of CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}); where IC: intracellular, CS: cell suspension and EC: Extracellular. Data analysis were performed using ANOVA with the Prism8. All were tested at the p<0.005*** significance level

3.5 CPD-03 genome analysis and presence of arylesterases gene

The draft genome sequence of *Ochrobactrum* CPD-03 has been submitted in GenBank under the accession RSEU00000000. The version described in this paper is version RSEU01000000 and comprises 4,660,061bp (~4.6 Mb), 31 contigs and 25 scaffolds (N50: 403.18 kb) (Supplementary Table 2). A comparative genomics approached was analyzed using KAAS (KEGG Automatic Annotation Server) [68] for functional annotation of proteins and metabolic pathways present in CPD-03. The minimal sets of metabolic pathways were further annotated and reconstructed against the predicted protein families using a web-based tool MinPath [69] and were clustered using ClustVis [70]. In addition, comparative genome analysis of other *Ochrobactrum* genomes were assessed for the capability to degrade the aromatic and relative compounds (Supplementary Fig. 2). This feature could also demonstrate the potential genes of CPD-03 strain for possible application in case of aromatic compound degradation.

Fig. 5 Genome map of *Ochrobactrum* sp. CPD-03

Although microorganisms have been known to possess numerous enzymes with capability to

degrade several xenobiotic compounds, only ten organophosphorus hydrolase genes were

reported to degrade CP. Organophosphorus hydrolase (OPH; EC 3.1.8.1) is an ortholog of

arylesterase (EC 3.1.1.2), are known to catalyze the hydrolysis of a wide range of OPs and

are produced by few microbes. OPH uses OPs as substrates rather than inhibitors or antigens.

Due to its bioremediation potential, OPH has been used in industrial, health care and military

purposes [71]. As of now, many arylesterases were isolated from mammals, known as serum

paraoxonases responsible for hydrolyzing paraoxon along with other organophosphorous

esters [72]. Previous studies have reported arylesterases from insect pests displaying

resistance against oxo-analogs of organophosphorous insecticides such as parathion and

chlorpyrifos [73].

Comparative genomics of CPD-03 genome with other *Ochrobactrum* genomes showing

presence of arylesterase was performed. Top 25 hits for Arylesterase sequences were

obtained from UniProt (https://www.uniprot.org/) and analyzed against the same present in

CPD-03 strain. A phylogenetic tree based on arylesterase sequences was constructed to

validate the closeness of this enzyme among top hits from the database. The highest sequence

homology   was   with   *Ochrobactrum   intermedium*   LMG   3301^T which   was   also

phylogenetically closest (Supplementary Fig. 3). Comparison of *opd* sequence obtained from

whole genome with the amplified *opd* gene sequence revealed that these two shared

similarity in genetic and protein level (Supplementary Fig. 4). The position of arylesterase

was annotated on genome of CPD-03 using a rapid prokaryotic genome annotation tool,

Prokka [74] (Supplementary Fig. 5). CPD-03 shared numerous putative pathways for aerobic

degradation of aromatic compounds (Supplementary Table 3). As of now, only two reports

are available on TCP biodegradation genes [75]. Based on the previous reports along with the

CPD-03 genome, a metabolic pathway has been proposed (Supplementary Fig. 6) which indicated the presence of gene clusters responsible for CP degradation by CPD-03. The characteristic feature of the strain CPD-03 as revealed by in silico analysis, showed its ability to degrade benzoate, aminobenzoate, chloroalkane and chloroalkene, chlorocyclohexane and chlorobenzene, suggesting that it might have acquired genes and pathways for mineralization of these xenobiotic compounds from the soil ecosystem.

3.6 Chemotactic response of strain CPD-03 towards various OPs

Bacterial chemotactic movement under a chemoattractant helps bacterium find metabolizable substrates for growth and survival. The chemoattractant is often a compound that also acts as a source of carbon and energy [76]. Chemotaxis has recently become a widespread phenomenon of bacterial metabolism against various xenobiotic compounds, and the use of these bacteria in bioremediation may therefore be beneficial [77].

Qualitative and quantitative assays for chemotaxis were performed with resting cells of CPD-03 with CP along with various other OPs like monocrotophos, methyl parathion, diazinon, malathion, and coumaphos separately. The strongest chemotactic response was observed for CP as compared to other OPs (Fig. 6a). Moreover, various concentration of CP also had the similar effect of chemotaxis (Fig. 6b, Supplementary Fig. 7). Additionally, putative gene clusters were also found in the genome of *Ochrobactrum* CPD-03, validating the chemotactic effects towards chemoattractant (Supplementary Table 3). To the best of our knowledge, our results suggested the first report of the genus *Ochrobactrum* for chemotaxis towards CP along with other OPs.

Fig.6 a. Chemotaxis index of CPD-03 resting cells CPD-03 (1-1.5x10^8 cells ml^{-1}) towards CP, TCP, monocrotophos, methyl parathion, coumaphos, diazione and malathione as competitive attractant at various concentrations (from 200 mg L^{-1} to 2000 mg L^{-1}). Values are presented as arithmetic means and error bars indicate standard deviations based on three independent replicates. b. migrated cell width of chemotactic movement by CPD-03 towards various concentration of CP (100 mg L^{-1}, 200 mg L^{-1}, 400 mg L^{-1}, 600 mg L^{-1}, 800 mg L^{-1}, 1000 mg L^{-1}, 1500 mg L^{-1} and 2000 mg L^{-1}) along with citrate (0.5%) as positive control. Data analysis were performed using ANOVA with the Prism8. All were tested at the p<0.005*** significance level

3.7 Assessment of Plant growth promoting ability of CPD-03

Rhizospheric bacteria live on and around the paddy fields have been contributing towards degradation of toxic xenobiotic compounds in contaminated ecosystem and shown to have potential for accelerating degradation process [78]. Paddy field environment comprised of phosphorus in soils (in both organic and inorganic forms) [79]. As a result, microbes associated with phosphate solubilizing ability would give an insight into the available forms of phosphate to the plants. In our present study, CPD-03 was found to solubilize the tricalcium phosphate post incubation, which was a qualitative analysis i.e. the agar media of

the pikovskaya containing bromophenol turned from blue to yellow due to a decrease in the

pH of the medium confirming a positive feature of phosphate solubilization (Fig. 7a). In

plants, production of ammonia is accountable for the indirect plant growth promotion via

pathogens control and d HCN is a key compound during this metabolism [80]. CPD-03

showed a substantial increase in ammonia production as compared to the control (*E.coli*)

(Fig. 7b). CPD-03 had shown to produce HCN (Fig. 7c). It is reported that, microbes isolated

from various agricultural sites have the ability to produce phytohormone auxin indole-3-

acetic acid (IAA) as secondary metabolites [81]. Generally, IAA secreted by microbes

stimulate seed germination followed by increase the rate of xylem and root development,

controls vegetative growth process, regulates photosynthesis, formation of pigments,

production of several metabolites, and resistance to stress [82]. In our study, CPD-03

produced 40.33 mg L^{-1} IAA as compared to the control (*E.coli*) which had produced 22.36

mg L^{-1} of IAA, respectively (Fig. 7d). CPD-03 inoculation was also positively affected the

increased root length (Fig. 7e). This result is consistent with the previous report on

Ochrobactrum sp. JAS2. Genome analysis had shown putative gene clusters present in

Ochrobactrum CPD-03, validating the potential effects of PGPR activity. However, the gene

clusters for HCN production, phosphate solublization, ACC deaminase activity, and ammonia

production were not found. The genome shows the presence of several hypothetical proteins

which might be contributing for these above mechanisms or presence of ortholog sequences

(Supplementary Table 4) which has still not been studied but may be of interest in future.

Fig.7(a-e). a. Phosphate solubilization on Pikovskaya's agar supplemented with

tricalciumphosphate by CPD-03 resting cells ($1\text{-}1.5\text{x}10^8$ cells ml^{-1}); b. quantification of

ammonia production; c. HCN production and d. quantification of IAA production by CPD-03

resting cells ($1\text{-}1.5\text{x}10^8$ cells ml^{-1}); e. Effect of root length upon CPD-03 inoculation on rice

seed; -ve control: culture free, +ve control: *Bacillus* sp., lab stock for positive PGPR activity.

E.coli was used as a negative control. Data analysis were performed using ANOVA with the

Prism8. All were tested at the $p < 0.005$*** significance level

4. Conclusion

A new CP degrading bacterium of *Ochrobactrum* genus was isolated and characterized as

Ochrobactrum sp. CPD-03 for its ability to degrade CP & its major metabolite TCP with

chemotactic ability towards other OPs and additional plant growth promoting abilities. Our

study highlights as the second report on a bacterial strain from the genus *Ochrobactrum*

possessing the ability to degradation of both CP and TCP. Whole genome assessment

revealed the presence of gene clusters responsible for CP biodegradation by CPD-03. In

addition to this, several other genes were found responsible for degradation of aromatic and

relevant compounds suggested a versatile role for CPD-03 in the degradation of xenobiotic

compounds. Based on phylogenetic and genomic analyses, CPD-03 was identified as

belonging to the genus *Ochrobatrum*. Thus, the genome sequence of CPD-03 is a valuable

source of information to understand the potential of this paddy field isolate to study its

importance in hydrolyzing the OP compounds. This study provides the genetic information

for its application for the degradation of OP pollutants in natural environment for further

development of a biostimulation process. CP degrading ability makes this strain CPD-03

useful for decontamination of the pesticide polluted sites. Moreover, the plant growth

promoting abilities could be of great potential for improved and persistent crop production

during CP induced stressed agricultural ecosystem.

Authors' contributions

TN & VR designed the research and outline. TN performed the experiments and generated

the result. BD performed microscopic analysis. TN & ANP performed genome analysis. TN

& ANP performed comparative genomics. TN wrote the manuscript. TKA & VR critically

reviewed the manuscript with appropriate technical inputs. Authors would like to thank Dr.

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References

[1] P. Abhilash, N. Singh, Pesticide use and application: an Indian scenario, Journal of

hazardous materials, 165 (2009) 1-12.

[2] S.D. Copley, Evolution of efficient pathways for degradation of anthropogenic chemicals,

Nature chemical biology, 5 (2009) 559.

[3] R. Lal, C. Dogra, S. Malhotra, P. Sharma, R. Pal, Diversity, distribution and divergence of

lin genes in hexachlorocyclohexane-degrading sphingomonads, Trends in biotechnology, 24

(2006) 121-130.

[4] L. Mhadhbi, R. Beiras, Acute toxicity of seven selected pesticides (alachlor, atrazine,

dieldrin, diuron, pirimiphos-methyl, chlorpyrifos, diazinon) to the marine fish (Turbot, Psetta

maxima), Water, Air, & Soil Pollution, 223 (2012) 5917-5930.

[5] M. Harishankar, C. Sasikala, M. Ramya, Efficiency of the intestinal bacteria in the

degradation of the toxic pesticide, chlorpyrifos, 3 Biotech, 3 (2013) 137-142.

[6] Y. Abreu-Villaca, E.D. Levin, Developmental neurotoxicity of succeeding generations of

insecticides, Environment international, 99 (2017) 55-77.

[7] S. Anwar, F. Liaquat, Q.M. Khan, Z.M. Khalid, S. Iqbal, Biodegradation of chlorpyrifos

and its hydrolysis product 3, 5, 6-trichloro-2-pyridinol by Bacillus pumilus strain C2A1,

Journal of Hazardous Materials, 168 (2009) 400-405.

[8] S. Das, T.K. Adhya, Degradation of chlorpyrifos in tropical rice soils, Journal of

Environmental Management, 152 (2015) 36-42.

[9] J. Li, J. Liu, W. Shen, X. Zhao, Y. Hou, H. Cao, Z. Cui, Isolation and characterization of

3, 5, 6-trichloro-2-pyridinol-degrading Ralstonia sp. strain T6, Bioresource technology, 101

(2010) 7479-7483.

[10] L. Fang, T. Shi, Y. Chen, X. Wu, C. Zhang, X. Tang, Q.X. Li, R. Hua, Kinetics and

catabolic pathways of the insecticide chlorpyrifos, annotation of the degradation genes and

characterization of enzymes TcpA and Fre in Cupriavidus nantongensis X1T, Journal of agricultural and food chemistry, (2019).

[11] M. Watts, Chlorpyrifos as a possible global POP, Pesticide Action Network North America, Oakland, CA. www. pan-europe. info/News/PR/121009_Chlorpyrifos_as_POP_final. pdf, (2012).

[12] B.K. Singh, A. Walker, J.A.W. Morgan, D.J. Wright, Effects of soil pH on the biodegradation of chlorpyrifos and isolation of a chlorpyrifos-degrading bacterium, Appl. Environ. Microbiol., 69 (2003) 5198-5206.

[13] H. Jabeen, S. Iqbal, S. Anwar, Biodegradation of chlorpyrifos and 3, 5, 6-trichloro-2-pyridinol by a novel rhizobial strain M esorhizobium sp. HN3, Water and Environment Journal, 29 (2015) 151-160.

[14] K.D. Racke, Environmental fate of chlorpyrifos, in: Reviews of environmental contamination and toxicology, Springer, 1993, pp. 1-150.

[15] S. Chen, C. Liu, C. Peng, H. Liu, M. Hu, G. Zhong, Biodegradation of chlorpyrifos and its hydrolysis product 3, 5, 6-trichloro-2-pyridinol by a new fungal strain Cladosporium cladosporioides Hu-01, PloS one, 7 (2012) e47205.

[16] S. Chandra, A.N. Mahindrakar, L. Shinde, Determination of cypermethrin and chlorpyrifos in vegetables by GC-ECD, Int. J. Chem Tech Res, 2 (2010) 908-911.

[17] F. Muhammad, M. Akhtar, Z. Rahman, H. Farooq, T. Khaliq, M. Anwar, Multi-residue determination of pesticides in the meat of cattle in Faisalabad-Pakistan, Egypt. Acad. J. Biol. Sci, 2 (2010) 19-28.

[18] S. Rayu, D.G. Karpouzas, B.K. Singh, Emerging technologies in bioremediation: constraints and opportunities, Biodegradation, 23 (2012) 917-926.

[19] R.R.M. Thengodkar, S. Sivakami, Degradation of Chlorpyrifos by an alkaline

phosphatase from the cyanobacterium Spirulina platensis, Biodegradation, 21 (2010) 637-

644.

[20] D. Singh, J. Khattar, J. Nadda, Y. Singh, A. Garg, N. Kaur, A. Gulati, Chlorpyrifos

degradation by the cyanobacterium Synechocystis sp. strain PUPCCC 64, Environmental

Science and Pollution Research, 18 (2011) 1351-1359.

[21] B.K. Singh, A. Walker, J.A.W. Morgan, D.J. Wright, Biodegradation of chlorpyrifos by

Enterobacter strain B-14 and its use in bioremediation of contaminated soils, Appl. Environ.

Microbiol., 70 (2004) 4855-4863.

[22] M. Farhan, A. Khan, A. Wahid, M. Ahmad, F. Ahmad, Biodegradation of chlorpyrifos

using indigenous Pseudomonas sp. isolated from industrial drain, Pak. J. Nutr, 11 (2012)

1183-1189.

[23] N. Chawla, S. Suneja, K. Kukreja, Isolation and characterization of chlorpyriphos

degrading bacteria, Indian Journal of Agricultural Research, 47 (2013).

[24] Z.Y. Liu, X. Chen, Y. Shi, Z.C. Su, Bacterial degradation of chlorpyrifos by Bacillus

cereus, in: Advanced Materials Research, Trans Tech Publ, 2012, pp. 676-680.

[25] E.R. El-Helow, M.E. Badawy, M.E. Mabrouk, E.A. Mohamed, Y.M. El-Beshlawy,

Biodegradation of chlorpyrifos by a newly isolated Bacillus subtilis strain, Y242,

Bioremediation journal, 17 (2013) 113-123.

[26] I. Ghanem, M. Orfi, M. Shamma, Biodegradation of chlorphyrifos byKlebsiella sp.

isolated from an activated sludge sample of waste water treatment plant in damascus, Folia

Microbiologica, 52 (2007) 423-427.

[27] X. Li, J. He, S. Li, Isolation of a chlorpyrifos-degrading bacterium, Sphingomonas sp.

strain Dsp-2, and cloning of the mpd gene, Research in Microbiology, 158 (2007) 143-149.

[28] G. Xu, W. Zheng, Y. Li, S. Wang, J. Zhang, Y. Yan, Biodegradation of chlorpyrifos and

3, 5, 6-trichloro-2-pyridinol by a newly isolated Paracoccus sp. strain TRP, International

Biodeterioration & Biodegradation, 62 (2008) 51-56.

[29] P. Lu, Q. Li, H. Liu, Z. Feng, X. Yan, Q. Hong, S. Li, Biodegradation of chlorpyrifos

and 3, 5, 6-trichloro-2-pyridinol by Cupriavidus sp. DT-1, Bioresource technology, 127

(2013) 337-342.

[30] C.V. Lakshmi, M. Kumar, S. Khanna, Biotransformation of chlorpyrifos and

bioremediation of contaminated soil, International Biodeterioration & Biodegradation, 62

(2008) 204-209.

[31] K.K. Dubey, M. Fulekar, Chlorpyrifos bioremediation in Pennisetum rhizosphere by a

novel potential degrader Stenotrophomonas maltophilia MHF ENV20, World Journal of

Microbiology and Biotechnology, 28 (2012) 1715-1725.

[32] C. Yang, N. Liu, X. Guo, C. Qiao, Cloning of mpd gene from a chlorpyrifos-degrading

bacterium and use of this strain in bioremediation of contaminated soil, FEMS microbiology

letters, 265 (2006) 118-125.

[33] M. Yadav, A.K. Shukla, N. Srivastva, S.N. Upadhyay, S.K. Dubey, Utilization of

microbial community potential for removal of chlorpyrifos: a review, Critical reviews in

biotechnology, 36 (2016) 727-742.

[34] J. Abraham, S. Silambarasan, Biodegradation of chlorpyrifos and its hydrolysis product

3, 5, 6-trichloro-2-pyridinol using a novel bacterium Ochrobactrum sp. JAS2: a proposal of

its metabolic pathway, Pesticide biochemistry and physiology, 126 (2016) 13-21.

[35] B. Engelen, K. Meinken, F. Von Wintzingerode, H. Heuer, H.-P. Malkomes, H.

Backhaus, Monitoring impact of a pesticide treatment on bacterial soil communities by

metabolic and genetic fingerprinting in addition to conventional testing procedures, Appl.

Environ. Microbiol., 64 (1998) 2814-2821.

[36] W.G. Weisburg, S.M. Barns, D.A. Pelletier, D.J. Lane, 16S ribosomal DNA

amplification for phylogenetic study, Journal of bacteriology, 173 (1991) 697-703.

[37] M.A. Larkin, G. Blackshields, N. Brown, R. Chenna, P.A. McGettigan, H. McWilliam,

F. Valentin, I.M. Wallace, A. Wilm, R. Lopez, Clustal W and Clustal X version 2.0,

bioinformatics, 23 (2007) 2947-2948.

[38] S. Kumar, G. Stecher, K. Tamura, MEGA7: molecular evolutionary genetics analysis

version 7.0 for bigger datasets, Molecular biology and evolution, 33 (2016) 1870-1874.

[39] Y.T. He, C. Su, Use of Additives in Bioremediation of Contaminated Groundwater and

Soil, in: Advances in Bioremediation of Wastewater and Polluted Soil, IntechOpen, 2015.

[40] J. Yashphe, H. Chikarmane, M. Iranzo, H.O. Halvorson, Phosphatases ofAcinetobacter

lwoffi. Localization and regulation of synthesis by orthophosphate, Current Microbiology, 20

(1990) 273-280.

[41] J.M. Bric, R.M. Bostock, S.E. Silverstone, Rapid in situ assay for indoleacetic acid

production by bacteria immobilized on a nitrocellulose membrane, Appl. Environ.

Microbiol., 57 (1991) 535-538.

[42] D. Dye, The inadequacy of the usual determinative tests for the identification of

Xanthomonas spp, New Zealand Journal of Science, 5 (1962).

[43] A.W. Bakker, B. Schippers, Microbial cyanide production in the rhizosphere in relation

to potato yield reduction and Pseudomonas spp-mediated plant growth-stimulation, Soil

Biology and Biochemistry, 19 (1987) 451-457.

[44] S. Islam, A.M. Akanda, A. Prova, M.T. Islam, M.M. Hossain, Isolation and

identification of plant growth promoting rhizobacteria from cucumber rhizosphere and their

effect on plant growth promotion and disease suppression, Frontiers in microbiology, 6

(2016) 1360.

[45] J. Adler, A method for measuring chemotaxis and use of the method to determine

optimum conditions for chemotaxis by Escherichia coli, Microbiology, 74 (1973) 77-91.

[46] J. Pandey, N.K. Sharma, F. Khan, A. Ghosh, J.G. Oakeshott, R.K. Jain, G. Pandey,

Chemotaxis of Burkholderia sp. strain SJ98 towards chloronitroaromatic compounds that it

can metabolise, BMC microbiology, 12 (2012) 19.

[47] S. Pailan, P. Saha, Chemotaxis and degradation of organophosphate compound by a

novel moderately thermo-halo tolerant Pseudomonas sp. strain BUR11: evidence for possible

existence of two pathways for degradation, PeerJ, 3 (2015) e1378.

[48] D.R. Zerbino, E. Birney, Velvet: algorithms for de novo short read assembly using de

Bruijn graphs, Genome research, 18 (2008) 821-829.

[49] B. Langmead, S.L. Salzberg, Fast gapped-read alignment with Bowtie 2, Nature

methods, 9 (2012) 357.

[50] D. Laslett, B. Canback, ARAGORN, a program to detect tRNA genes and tmRNA genes

in nucleotide sequences, Nucleic acids research, 32 (2004) 11-16.

[51] A. Imran, F.Y. Hafeez, A. Frühling, P. Schumann, K. Malik, E. Stackebrandt,

Ochrobactrum ciceri sp. nov., isolated from nodules of Cicer arietinum, International journal

of systematic and evolutionary microbiology, 60 (2010) 1548-1553.

[52] L. Yang, Y.-h. Zhao, B.-x. Zhang, C.-H. Yang, X. Zhang, Isolation and characterization

of a chlorpyrifos and 3, 5, 6-trichloro-2-pyridinol degrading bacterium, FEMS Microbiology

Letters, 251 (2005) 67-73.

[53] J.-R. Kim, Y.-J. Ahn, Identification and characterization of chlorpyrifos-methyl and 3, 5,

6-trichloro-2-pyridinol degrading Burkholderia sp. strain KR100, Biodegradation, 20 (2009)

487-497.

[54] Y. Feng, K.D. Racke, J. Bollag, Isolation and characterization of a chlorinated-pyridinol-

degrading bacterium, Appl. Environ. Microbiol., 63 (1997) 4096-4098.

[55] K. Maya, R. Singh, S. Upadhyay, S.K. Dubey, Kinetic analysis reveals bacterial efficacy

for biodegradation of chlorpyrifos and its hydrolyzing metabolite TCP, Process Biochemistry,

46 (2011) 2130-2136.

[56] G. Xu, Y. Li, W. Zheng, X. Peng, W. Li, Y. Yan, Mineralization of chlorpyrifos by co-

culture of Serratia and Trichosporon spp, Biotechnology letters, 29 (2007) 1469-1473.

[57] Y. Feng, K. Racke, J.-M. Bollag, Use of immobilized bacteria to treat industrial

wastewater containing a chlorinated pyridinol, Applied microbiology and biotechnology, 47

(1997) 73-77.

[58] C.S. Mosquera-Vivas, E.W. Hansen, G. García-Santos, N. Obregón-Neira, R.E. Celis-

Ossa, C.A. González-Murillo, R. Juraske, S. Hellweg, J.A. Guerrero-Dallos, The effect of the

soil properties on adsorption, single-point desorption, and degradation of chlorpyrifos in two

agricultural soil profiles from Colombia, Soil Science, 181 (2016) 446-456.

[59] A.C. Affam, M. Chaudhuri, Degradation of pesticides chlorpyrifos, cypermethrin and

chlorothalonil in aqueous solution by TiO2 photocatalysis, Journal of environmental

management, 130 (2013) 160-165.

[60] Y. Lyu, W. Zheng, T. Zheng, Y. Tian, Biodegradation of polycyclic aromatic

hydrocarbons by Novosphingobium pentaromativorans US6-1, PLoS One, 9 (2014) e101438.

[61] R. Chandra, P. Sharma, S. Yadav, S. Tripathi, Biodegradation of Endocrine-Disrupting

Chemicals and Residual Organic Pollutants of Pulp and Paper Mill Effluent by

Biostimulation, Frontiers in Microbiology, 9 (2018).

[62] S. Chen, Q. Lin, Y. Xiao, Y. Deng, C. Chang, G. Zhong, M. Hu, L.-H. Zhang,

Monooxygenase, a novel beta-cypermethrin degrading enzyme from Streptomyces sp, PloS

one, 8 (2013) e75450.

[63] A. Zyakun, S. Nii-Annang, G. Franke, T. Fischer, F. Buegger, O. Dilly, Microbial

Actvity and 13C/12C Ratio as Evidence of N-Hexadecane and N-Hexadecanoic Acid

Biodegradation in Agricultural and Forest Soils, Geomicrobiology Journal, 28 (2011) 632-
647.

[64] C.C. de Carvalho, Adaptation of Rhodococcus erythropolis cells for growth and
bioremediation under extreme conditions, Research in microbiology, 163 (2012) 125-136.

[65] X. Tang, B. Liang, T. Yi, G. Manco, A. Liu, Cell surface display of organophosphorus
hydrolase for sensitive spectrophotometric detection of p-nitrophenol substituted
organophosphates, Enzyme and microbial technology, 55 (2014) 107-112.

[66] J.-C. Wang, Y. Lin, B.-L. Sun, Localization and Characterization of Fe (III) Reductase
of Desulfovibrio dechloracetivorans Strain SF3, Journal of University of Science and
Technology of China, 39 (2009) 337-343.

[67] S.K. Falahati-Pour, A.S. Lotfi, G. Ahmadian, A. Baghizadeh, R. Behroozi, F. Haghighi,
High-level extracellular secretion of organophosphorous hydrolase of Flavobacterium sp. in
Escherichia coli BL21 (DE3) pLysS, Biotechnology and applied biochemistry, 63 (2016)
870-876.

[68] Y. Moriya, M. Itoh, S. Okuda, A.C. Yoshizawa, M. Kanehisa, KAAS: an automatic
genome annotation and pathway reconstruction server, Nucleic acids research, 35 (2007)
W182-W185.

[69] Y. Ye, T.G. Doak, A parsimony approach to biological pathway reconstruction/inference
for genomes and metagenomes, PLoS computational biology, 5 (2009) e1000465.

[70] T. Metsalu, J. Vilo, ClustVis: a web tool for visualizing clustering of multivariate data
using Principal Component Analysis and heatmap, Nucleic acids research, 43 (2015) W566-
W570.

[71] C.J. Jackson, A. Carville, J. Ward, K. Mansfield, D.L. Ollis, T. Khurana, S.B. Bird, Use
of OpdA, an organophosphorus (OP) hydrolase, prevents lethality in an African green
monkey model of acute OP poisoning, Toxicology, 317 (2014) 1-5.

[72] A. Aharoni, L. Gaidukov, S. Yagur, L. Toker, I. Silman, D.S. Tawfik, Directed evolution

of mammalian paraoxonases PON1 and PON3 for bacterial expression and catalytic

specialization, Proceedings of the National Academy of Sciences, 101 (2004) 482-487.

[73] N. De Malkenson, E. Wood, E. Zerba, Isolation and characterization of an esterase of

Triatoma infestans with a critical role in the degradation of organophosphorus esters, Insect

biochemistry, 14 (1984) 481-486.

[74] T. Seemann, Prokka: rapid prokaryotic genome annotation, Bioinformatics, 30 (2014)

2068-2069.

[75] R.K. Math, S.M.A. Islam, K.M. Cho, S.J. Hong, J.M. Kim, M.G. Yun, J.J. Cho, J.Y.

Heo, Y.H. Lee, H. Kim, Isolation of a novel gene encoding a 3, 5, 6-trichloro-2-pyridinol

degrading enzyme from a cow rumen metagenomic library, Biodegradation, 21 (2010) 565-

573.

[76] G. Pandey, R.K. Jain, Bacterial chemotaxis toward environmental pollutants: role in

bioremediation, Appl. Environ. Microbiol., 68 (2002) 5789-5795.

[77] F. Gordillo, F.P. Chávez, C.A. Jerez, Motility and chemotaxis of Pseudomonas sp. B4

towards polychlorobiphenyls and chlorobenzoates, FEMS microbiology ecology, 60 (2007)

322-328.

[78] B. Divya, M.D. Kumar, Plant-microbe interaction with enhanced bioremediation, Res J

Biotechnol, 6 (2011) 72-79.

[79] M.S. Khan, A. Zaidi, P.A. Wani, M. Oves, Role of plant growth promoting rhizobacteria

in the remediation of metal contaminated soils, Environmental chemistry letters, 7 (2009) 1-

19.

[80] C. Blumer, D. Haas, Mechanism, regulation, and ecological role of bacterial cyanide

biosynthesis, Archives of Microbiology, 173 (2000) 170-177.

[81] C.L. Patten, B.R. Glick, Bacterial biosynthesis of indole-3-acetic acid, Canadian journal

of microbiology, 42 (1996) 207-220.

[82] M. Ahemad, M. Kibret, Mechanisms and applications of plant growth promoting

rhizobacteria: current perspective, Journal of King saud University-science, 26 (2014) 1-20.

763   Supplementary Fig.1 a. Degradation of CP (100 mg L^{-1}) by CPD-03 (1.2x10^8 cfu/ml) leads to

764   formation of TCP followed by its further degradation; b. degradation of TCP (100 mg L^{-1}) by

765   CPD-03 (1.2x10^8 cfu/ml). Data were represented in the mean of three replicates and analysis

766   were performed using ANOVA with the Prism8. All were tested at the p<0.0001***

767   significance level

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Supplementary Fig. 2. Heatmap and dendrogram of potential abilities of degradation of

Aromatic and relative compounds predicted by genomic features of CPD-03 along with other

Ochrobactrum strains (*Ochrobactrum thiophenivorans* DSM 7216[T], *Ochrobactrum*

pseudogrignonense CCUG 30717[T], *Ochrobactrum pituitosum* CCUG 50899[T], *Ochrobactrum*

grignonense OgA9a[T], *Ochrobactrum rhizosphaerae* PR17[T], *Ochrobactrum anthropi* ATCC

49188[T], *Ochrobactrum lupini* LUP21[T], *Ochrobactrum intermedium* LMG 3301[T]).

Information retrieved from MinPath Annotation Tool.

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791 Supplementary Fig. 3: Phylogenetic tree with respect to the Arylesterase enzyme present in

792 *Ochrobactrum* strains. These sequences were obtained from UniProt followed by the Top 25

793 hits were analyzed in MEGA 7.

Supplementary Fig. 4: a. Nucleotide sequence similarity of organophosphate hydrolase and

Arylesterase (EC 3.1.1.2) present in CPD-03 strain; b. Protein sequence similarity of

organophosphate hydrolase and Arylesterase (EC 3.1.1.2) present in CPD-03 strain. The

nucleotide sequence of organophosphate hydrolase gene was obtained after successful

cloning followed by Sangers sequencing and the Arylesterase sequence was obtained from

the whole genome. These sequences similarity comparison was performed in ClustalOmega.

811   Supplementary Fig. 5: Position of Arylesterase mapped on the genome of *Ochrobactrum* sp.

812   CPD-03. This was located using rapid prokaryotic genome annotation (Prokka).

Supplementary Fig. 6. Proposed pathway for metabolic degradation of CP by *Ochrobactrum*

sp. CPD-03 based on whole genome analysis

Supplementary Fig. 7. Chemotaxis of *Ochrobactrum* sp. CPD-03 resting cells of CPD-03 (1-

1.5x10^8 cells ml^{-1}) towards various concentration of CP along with citrate (0.5%) as positive

control citrate. Migrated cell width of CPD-03 showing the response towards CP at various

concentration

854   Supplementary Table. 1a.

Test	Results
Catalase, Oxidase	+
NaCl tolerance	Up to 4%
Growth temperature	25^0C - 37^0C (30 ± 2 ^0C optimum)
pH	5-8 (7 optimum)
Utilization of Dextrin, maltose, Cellobiose, Sucrose, N-Acetyl-D-Glucosamine, Citrate, Galactose, Glucose, Arabitol, myo-Inositol, Mannose, Fructose, Xylose	+

856   Notes.: (+, positive reaction). Data taken from GEN III plate of Biolog® System

857   Phenotypic characteristics of the strain *Ochrobactrum* sp. CPD-03

880   Supplementary Table. 1b.

Assimilation of	A	B	C
Tween 80	-	+	-
N-Acetyl-D-galactosamine	+	+	+
Adonitol	+	+	+
D-Arabitol	+	+	+
Cellobiose	+	+	+
Gentiobiose	+	+	+
L-Fucose	+	+	+
myo-Inositol	+	+	+
D-Mannitol	+	+	+
L-Rhamnose	+	+	-
D-Sorbitol	+	+	+
Raffinose	-	-	+
Sucrose	+	+	+
Trehalose	-	-	+
Citric acid	+	+	+
Glycerol	-	-	-

884   Differential biochemical properties of strain CPD-03 (A) and closet reference *Ochrobactrum*

885   type strains Strains: (B) Ca-34[T]; (C) LMG 3301[T]. The table is arranged according to

886   phylogenetic clustering. All data for CPD-03 are from this study and others obtained from

887   Imran et al., 2010.

896   Supplementary Table 1c.

RT	Response	Ar/Ht	RFact	ECL	Peak Name	Percent
1.4094	1143	0.011	1.180	10.6984	11:0 anteiso	1.01
1.4701	1393	0.016	1.153	11.0051	11:0	1.20
1.5098	459	0.009	1.137	11.1740	10:0 2OH	0.39
1.6360	833	0.009	1.094	11.7107	12:0 anteiso	0.68
1.8971	849	0.014	1.027	12.7126	13:0 anteiso	0.65
2.1184	628	0.014	0.987	13.4833	12:0 3OH	0.46
2.2721	782	0.015	0.965	14.0005	14:0	0.56
2.8562	1163	0.010	0.912	15.8393	Sum In Feature 3	0.79
2.9082	8304	0.008	0.910	16.0001	16:0	5.66
3.0882	1047	0.013	0.902	16.5564	17:1 anteiso w9c	0.71
3.1724	725	0.012	0.899	16.8162	17:1 w8c	0.49
3.1956	379	0.006	0.899	16.8879	17:1 w6c	0.25
3.2028	545	0.008	0.898	16.9102	17:0 cyclo	0.37
3.2327	1643	0.009	0.898	17.0024	17:0	1.10
3.4100	804	0.012	0.894	17.5528	16:0 3OH	0.54
3.5035	59266	0.009	0.893	17.8430	Sum In Feature 8	39.64
3.5342	582	0.009	0.893	17.9382	18:1 w5c	0.39
3.5529	10971	0.009	0.893	17.9963	18:0	7.34
3.5797	1234	0.011	0.893	18.0812	18:1 w7c 11-methyl	0.83
3.7517	1304	0.019	0.893	18.6277	19:0 iso	0.87
3.8096	1653	0.017	----	18.8116	unknown 18.815	----
3.8446	41251	0.009	0.894	18.9225	19:0 cyclo w8c	27.61
3.8695	998	0.015	0.894	19.0015	19:0	0.67
3.9118	5990	0.011	0.895	19.1386	18:1 2OH	4.01
4.0464	2199	0.010	0.897	19.5748	18:0 3OH	1.48
4.0939	1558	0.013	0.898	19.7286	20:2 w6,9c	1.05
4.1292	1844	0.011	0.898	19.8430	20:1 w7c	1.24
----	1163	---	----	----	Summed Feature 3	0.79
----	59266	---	----	----	Summed Feature 8	39.64

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902   Fatty Acid Methyl Ester (FAME) analysis of CPD-03

912 Supplementary Table 2.
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	Ochrobactrum sp. CPD-03
Genome Size (bp)	4,660,061bp (~4.6 Mb)
GC Content (%)	57.67 %
CDS	4566
rRNA	3
tRNA	51
No. of plasmid	0

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915 Genome features of *Ochrobactrum* sp. CPD-03

954   Supplementary Table 3.

S No.	Compound Name	Gene Designation	KEGG Number	EC Number	Enzyme Name
1	Benzoate degradation	*pobA*	K00481	1.14.13.2	p-hydroxybenzoate 3-monooxygenase
		pcaG	K00448	1.13.11.3	protocatechuate 3,4-dioxygenase, alpha subunit
		pcaB	K01857	5.5.1.2	3-carboxy-cis,cis-muconate cycloisomerase
		pcaC	K01607	4.1.1.44	4-carboxymuconolactone decarboxylase
		dmpB, xylE	K00446	1.13.11.2	catechol 2,3-dioxygenase
		pcaD	K01055	3.1.1.24	3-oxoadipate enol-lactonase
		fadA, fadI	K00632	2.3.1.16	acetyl-CoA acyltransferase
		dmpB, xylE	K00446	1.13.11.2	catechol 2,3-dioxygenase
		fadJ	K01782	1.1.1.35/ 4.2.1.17/ 5.1.2.3	3-hydroxyacyl-CoA dehydrogenase / enoyl-CoA hydratase / 3-hydroxybutyryl-CoA epimerase
		fadA, fadI	K00632	2.3.1.16	acetyl-CoA acyltransferase
		gcdH	K00252	1.3.8.6	glutaryl-CoA dehydrogenase
		paaF, echA	K01692	4.2.1.17	enoyl-CoA hydratase
		paaH, hbd, fadB, mmgB	K00074	1.1.1.157	3-hydroxybutyryl-CoA dehydrogenase
		atoB	K00626	2.3.1.9	acetyl-CoA C-acetyltransferase
2	Aminobenzoate degradation	*amiE*	K01426	3.5.1.4	amidase
		bsdC	K16239	4.1.1.61	4-hydroxybenzoate decarboxylase subunit C
		paaF, echA	K01692	4.2.1.17	enoyl-CoA hydratase
3	Chloroalkane and chloroalkene degradation	*adhP*	K13953	1.1.1.1	alcohol dehydrogenase, propanol-preferring
		aldH	K00128	1.2.1.3	aldehyde dehydrogenase (NAD+)
		fdhA	K00148	1.2.1.46	glutathione-independent formaldehyde dehydrogenase
		NA	K01560	3.8.1.2	2-haloacid dehalogenase
4	Chlorocyclohexane and chlorobenzene degradation	*dmpB, xylE*	K00446	1.13.11.2	catechol 2,3-dioxygenase
		NA	K01560	3.8.1.2	2-haloacid dehalogenase
5	Xylene degradation	*dmpB, xylE*	K00446	1.13.11.2	catechol 2,3-dioxygenase
6	Nitrotoluene degradation	*nfsA*	K10678	1.5.1.34	nitroreductase
7	Ethylbenzene	*fadA, fadI*	K00632	2.3.1.16	acetyl-CoA acyltransferase
8	Styrene degradation	*amiE*	K01426	3.5.1.4	amidase
		dmpB, xylE	K00446	1.13.11.2	catechol 2,3-dioxygenase
9	Atrazine degradation	*ure*	K01427	3.5.1.5	urease
10	Dioxin degradation	NA	K00480	1.14.13.1	salicylate hydroxylase
11	Naphthalene degradation	NA	K00480	1.14.13.1	salicylate hydroxylase
		adhP	K13953	1.1.1.1	alcohol dehydrogenase, propanol-preferring
12	Polycyclic aromatic hydrocarbon degradation	NA	K00480	1.14.13.1	salicylate hydroxylase
		pcaG	K00448	1.13.11.3	protocatechuate 3,4-dioxygenase, alpha subunit

957   Putative genes of *Ochrobactrum* CPD-03 involved in degradation of Aromatic and relative

958   compounds. Information retrieved from KAAS annotation tool. [NA: Not available]

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962 Supplementary Table 4.

S No.	Activity	Gene Designation	Enzyme Name
1	IAA production	**NA***	Indole-3-glycerol phosphate synthase
2	Siderophore production	*FhuB*	Ferric hydroxamate ABC transporter, permease component
		TonB	Ferric siderophore transport system, periplasmic binding protein
		FepC	Ferric enterobactin transport ATP-binding protein @ ABC-type Fe3+-siderophore transport system, ATPase component
		FepG	Ferric enterobactin transport system permease protein
		FepB	Ferric enterobactin-binding periplasmic protein
		FepC	Ferric enterobactin transport ATP-binding protein
		FhuD	Ferric hydroxamate ABC transporter, periplasmic substrate binding protein
		FhuC	Ferric hydroxamate ABC transporter, ATP-binding protein
		NA*	ABC transporter (iron.B12.siderophore.hemin) , permease component
		FhuA	Ferric hydroxamate outer membrane receptor
		EntS	Enterobactin exporter
		NA*	ABC-type Fe3+-siderophore transport system, permease component
		FhuB	Ferric hydroxamate ABC transporter, permease component
3	Flagellar protein	*FlhB, FliL, FliP, FliQ, FlhA, FliR*	Flagellar biosynthesis protein
		FliG, FliN, FliM	Flagellar motor switch protein
		MotA, *MotB*	Flagellar motor rotation protein
		FlgF, FlgF	Flagellar basal-body rod protein
		FliI	Flagellum-specific ATP synthase
		FlgB, FlgC, FlgG	Flagellar basal-body rod protein
		FliE	Flagellar hook-basal body complex protein
		FlgA	Flagellar basal-body P-ring formation protein
		FlgI	Flagellar P-ring protein
		FlgH	Flagellar L-ring protein
		FlaA, FlaF, FliJ	Flagellin protein
		FliF	Flagellar M-ring protein
		FliK	Flagellar hook-length control protein
		FlgE	Flagellar hook protein
		FlgK, FlgL	Flagellar hook-associated protein
		FlgD	Flagellar basal-body rod modification protein
		FlgJ	Flagellar protein [peptidoglycan hydrolase]
4	Chemotaxis	**NA***	Dipeptide-binding ABC transporter, periplasmic substrate-binding component
		RbsB	Ribose ABC transporter, periplasmic ribose-binding protein
		MalE	Maltose/maltodextrin ABC transporter, substrate binding periplasmic protein
		NA*	Dipeptide-binding ABC transporter, periplasmic substrate-binding component; Putative hemin-binding lipoprotein
		CheR	Chemotaxis protein methyltransferase (EC 2.1.1.80)
		motC	Chemotaxis protein
		MalE	Maltose/maltodextrin ABC transporter, substrate binding periplasmic protein
			Methyl-accepting chemotaxis protein
5	HCN production		
6	Phosphate solublization		**NA***
7	ACC deaminase activity		
8	Ammonia production		

963 **NA*: Not assigned in the genome of CPD-03**

964 Putative genes of *Ochrobactrum* CPD-03 involved in Plant growth promoting activities and
965 chemotaxis. Information retrieved from RAST annotation tool. [NA: Not available]

Fig.1. a. Scanning electron microscopy of CPD-03; b. phylogenetic tree of CPD-03 based on comparison of 16S rRNA sequence similarity of closely related *Ochrobactrum* sp. conducted in MEGA version 7.0

Fig.2. a. Degradation of CP (100 mg L^{-1}) and TCP (100 mg L^{-1}) by CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in a MSM supplemented molasses (0.1% v/v) showing formation of its intermediate TCP and its further degradation; b. degradation of TCP (100 mg L^{-1}) by CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in a MSM supplemented molasses (0.1% v/v); c. degradation of CP at various concentrations (100 mg L^{-1}, 200 mg L^{-1},400 mg L^{-1},600 mg L^{-1},800 mg L^{-1}, and 1000 mg L^{-1}) by CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}) in MSM supplemented molasses (0.1% v/v); d. comparision of CP (100 mg L^{-1}) degradation by CPD-03 resting cells (1-1.5x10^8 cells ml^{-1}) and LMG 3301^T (1.2-1.4x10^8 cells ml^{-1}). Data were represented in the mean of three replicates and analysis were performed using ANOVA with the Prism8. All were tested at the $p < 0.005$*** significance level

Fig.3 GC-MS chromatograms showing the metabolites formed during CP (100 mg L^{-1}) degradation by CPD-03 resting cells (1-1.5x10^8 cells ml^{-1}). A: 0 h, B: 2 h, C: 4 h, D: 6 h, E: 8 h, and F: 10 h respectively. 1: chlorpyrifos, 2: cis-Vaccenic acid, 3: Octadecanoic acid, 4: Benzoic acid, 4-hydroxyl 3-methoxy, 5: n-Hexadecanoic acid, 6: 9-Hexadecanoic acid, respectively.

Fig.4. a. Amplification of ~1098 bp *opd* gene; b. enzyme activity of CPD-03 resting cell (1-1.5x10^8 cells ml^{-1}); where IC: intracellular, CS: cell suspension and EC: Extracellular. Data analysis were performed using ANOVA with the Prism8. All were tested at the p<0.005*** significance level

Fig. 5 Genome map of *Ochrobactrum* sp. CPD-03

Fig.6 a. Chemotaxis index of CPD-03 resting cells CPD-03 (1-1.5x10^8 cells ml^{-1}) towards CP, TCP, monocrotophos, methyl parathion, coumaphos, diazione and malathione as competitive attractant at various concentrations (from 200 mg L^{-1} to 2000 mg L^{-1}). Values are presented as arithmetic means and error bars indicate standard deviations based on three independent replicates. b. migrated cell width of chemotactic movement by CPD-03 towards various concentration of CP (100 mg L^{-1}, 200 mg L^{-1}, 400 mg L^{-1}, 600 mg L^{-1}, 800 mg L^{-1}, 1000 mg L^{-1}, 1500 mg L^{-1} and 2000 mg L^{-1}) along with citrate (0.5%) as positive control. Data analysis were performed using ANOVA with the Prism8. All were tested at the p<0.005*** significance level

Fig.7(a-e). a. Phosphate solubilization on Pikovskaya's agar supplemented with tricalciumphosphate by CPD-03 resting cells (1-1.5x10^8 cells ml^{-1}); b. quantification of ammonia production; c. HCN production and d. quantification of IAA production by CPD-03 resting cells (1-1.5x10^8 cells ml^{-1}); e. Effect of root length upon CPD-03 inoculation on rice seed; -ve control: culture free, +ve control: *Bacillus* sp., lab stock for positive PGPR activity. *E.coli* was used as a negative control. Data analysis were performed using ANOVA with the Prism8. All were tested at the p<0.005*** significance level