

1 ***iucn_sim*: A new program to simulate future extinctions based on**
2 **IUCN threat status**

3 **Author information**

4 Tobias Andermann ^{1,2,*}, Søren Faurby ^{1,2}, Robert Cooke ^{1,2}, Daniele Silvestro ^{1,2,**}, Alexandre
5 Antonelli ^{1,2,3,**}

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7 ¹ Department of Biological and Environmental Sciences, University of Gothenburg,
8 Göteborg, Sweden

9 ² Gothenburg Global Biodiversity Centre, Göteborg, Sweden

10 ³ Royal Botanic Gardens, Kew, Richmond, UK

11

12 * Corresponding author. Email: tobias.andermann@bioenv.gu.se, ORCID-ID:
13 <https://orcid.org/0000-0002-0932-1623>

14 ** Joint last authorship

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16 **Abstract**

17 The ongoing environmental crisis poses an urgent need to forecast the *who*, *where*, and *when*
18 of future species extinctions, as such information is crucial for targeting conservation efforts.
19 Commonly, such forecasts are made based on conservation status assessments produced by
20 the International Union for Conservation of Nature (IUCN). However, when researchers
21 apply these IUCN conservation status data for predicting future extinctions, important
22 information is often omitted, which can impact the accuracy of these predictions.

23 Here we present a new approach and a software for simulating future extinctions based on
24 IUCN conservation status information, which incorporates generation length information of
25 individual species when modeling extinction risks. Additionally, we explicitly model future
26 changes in conservation status for each species, based on status transition rates that we
27 estimate from the IUCN assessment history of the last decades. Finally, we apply a Markov
28 chain Monte Carlo algorithm to estimate extinction rates for each species, based on the
29 simulated future extinctions. These estimates inherently incorporate the chances of
30 conservation status changes and the generation length for each given species and are specific
31 to the simulated time frame.

32 We demonstrate the utility of our approach by estimating extinction rates for all bird species.
33 Our average extinction risk estimate for the next 100 years across all birds is 6.98×10^{-4}
34 extinctions per species-year, and we predict an expected biodiversity loss of between 669 to
35 738 bird species within that time frame. Further, the rate estimates between species sharing
36 the same IUCN status show larger variation than the rates estimated with alternative
37 approaches, which reflects expected differences in extinction risk among taxa of the same
38 conservation status. Our method demonstrates the utility of applying species-specific
39 information to the estimation of extinction rates, rather than assuming equal extinction risks
40 for species assigned to the same conservation status.

41 **Keywords**

42 Aves, Bayesian, Death process, Biodiversity loss, Extinction risk, Generation length, IUCN,
43 MCMC, Conservation status, Red List

44 **Introduction**

45 We are in the middle of a massive biodiversity crisis (Barnosky et al. 2011, Davis et al. 2018,
46 Díaz et al. 2019). Extinction risks have been steadily increasing for as long as we have been
47 keeping record, with no indications of a slowdown (Ceballos et al. 2015). It is therefore
48 crucial to predict the number of future extinctions that shape the future biodiversity, whether
49 in terms of species, phylogenetic, or functional diversity (Davis et al. 2018, Cooke et al.
50 2019, Pimiento et al. 2020). An important use of such predictions is to aid conservation
51 prioritization (Mooers et al. 2008). However, all predictions require reliable estimates of
52 extinction risk.

53 The main global initiative to quantify extinction risks across animal and plant species is the
54 IUCN Red List (IUCN Red List 2019), which categorizes the conservation status of
55 organisms based on expert assessments. Since 2001, the IUCN has adopted the IUCN v3.1
56 evaluation system for determining species' conservation statuses (IUCN Species Survival
57 Commission 2001). By this standard, extant species are assessed as Least Concern (LC),
58 Near Threatened (NT), Vulnerable (VU), Endangered (EN), or Critically Endangered (CR). If
59 there is insufficient information available for a species to enable a proper status assessment,
60 the species is categorized as Data Deficient (DD). Species that have not yet been reviewed by
61 the IUCN are categorized as Non-Evaluated (NE). Species that are not found in the wild
62 anymore are labeled as Extinct in the Wild (EW), and species with no living wild or captive
63 individuals as Extinct (EX). As of the year 2020, IUCN has also introduced two additional
64 subcategories for CR species (IUCN 2020): possibly extinct [CR(PE)], and possibly extinct
65 in the wild [CR(PEW)].

66 For the IUCN to decide on assigning a species to one of the threatened categories VU, EN, or
67 CR, this species must meet at least one of five assessment criteria (A-E). One of those criteria

68 (E) is associated with a specific extinction probability, while the other criteria (A-D) mostly
69 encompass estimates of decreasing population trends and fragmentation. The IUCN
70 extinction probability thresholds defined in criterion E are as follows:

- 71 • VU: 10% extinction probability within 100 years
- 72 • EN: 20% extinction probability within 20 years or 5 generations, whichever is longer
73 (maximum 100 years)
- 74 • CR: 50% extinction probability within 10 years or 3 generations, whichever is longer
75 (maximum 100 years)

76 IUCN conservation status assessments have been used in numerous scientific studies to
77 project future biodiversity loss (e.g. Ricciardi and Rasmussen 1999, Veron et al. 2016, Davis
78 et al. 2018, Cooke et al. 2019, Oliveira et al. 2019). One critical challenge in this approach is
79 to meaningfully transform the IUCN-defined conservation statuses into explicit extinction
80 probabilities. In these previous studies, researchers have applied the extinction probabilities
81 associated with criterion E to model extinction risks for threatened species. Sometimes these
82 risks are also extrapolated to species of the statuses LC and NT in order to make it possible to
83 assign extinction probabilities to these species (Redding and Mooers 2006, Mooers et al.
84 2008, Veron et al. 2016, Davis et al. 2018).

85 Although these extinction probabilities only apply to species that are assessed under criterion
86 E (see Akçakaya et al. 2006), they are commonly applied equally to all species sharing the
87 same conservation status (e.g. Mooers et al. 2008, Davis et al. 2018). The underlying
88 assumption that the minimum extinction risks defined for criterion E can be meaningfully
89 transferred to species listed under one of the other four criteria (A-D) is difficult to test
90 empirically, but is a necessary simplification in order to model the extinction probabilities for
91 the majority of species. However, there are several other important aspects that can be easily

92 incorporated but are commonly neglected when translating IUCN conservation statuses into
93 extinction probabilities for future extinction predictions.

94 **Neglected information**

95 To the best of our knowledge, there are two key elements that are usually not incorporated
96 when using IUCN status data for future extinction predictions: generation length and
97 expected changes in conservation status (but see Monroe et al. 2019).

98 Generation length (GL) is defined as the average turnover rate of breeding individuals in a
99 population (IUCN Standards and Petitions Committee 2019) and therefore reflects the
100 turnover between generations. It is generally considered to be a more meaningful time unit
101 for modeling extinction risk than time expressed in years (Frankham and Brook 2004).

102 Generation length should not be confused with age of sexual maturity, which can be used in
103 the calculation of generation length, but is not equivalent (with age of sexual maturity always
104 being smaller than or equal to generation length). As per the IUCN definition, the extinction
105 probability for the categories EN and CR is to be understood in context of the GL of the
106 given species, if $5 \times GL$ exceeds 20 years for EN species, or if $3 \times GL$ exceeds 10 years for
107 CR species (see criterion E definitions above). We argue that including GL should be the
108 standard practice when modeling extinction risks based on IUCN data, particularly because
109 GL data is readily available for many species (Pacifci et al. 2013, BirdLife International
110 2019, IUCN Red List 2019) and can normally be approximated through a combination of
111 body size and phylogenetic information (Cooke et al. 2018, Bird et al. 2020) for species
112 missing GL data (Appendix 1).

113 A second missing element in many future predictions relates to the fact that IUCN categories
114 are generally treated as static entities that do not change over time. However, almost two

115 decades of IUCN re-assessments of species (IUCN Red List 2019), using the IUCN v3.1
116 standard, have clearly shown this not to be the case. Instead, re-assessments (Butchart et al.
117 2007, Rondinini et al. 2014) show that the conservation status of species can change
118 significantly in a relatively short time span, for instance as a result of the effectiveness of
119 conservation efforts. For a species classified as LC, the immediate extinction risk is
120 negligibly small, while for a species classified as CR, the immediate extinction risk is very
121 high. It is reasonable to assume that if we simulate, for example, 100 years into the future,
122 categories may change due to new or intensified risks or thanks to conservation efforts, which
123 will affect the extinction probabilities.

124 An example of a change in IUCN status is the Mauritian Pink Pigeon (*Nesoenas mayeri*),
125 which was listed as CR in the 1990s, with only 9 birds remaining, due to habitat loss and
126 predation by introduced species (Swinerton 2001, IUCN Red List 2019). However,
127 following an intensive conservation recovery program, the Pink Pigeon is now listed as VU,
128 with around 470 wild birds (IUCN Red List 2019). Unfortunately, most species show
129 changes with the opposite trend, for example several species of vultures, which are declining
130 due to poisoning and persecution (Green et al. 2007). There are 22 species of vulture
131 (Accipitridae: Gypaetinae, Accipitridae: Aegypiinae, and Cathartidae) according to the IUCN
132 Red List, 12 of these are classified as threatened (VU, EN or CR), including 9 CR species
133 (IUCN Red List 2019), with sharp declines in population sizes. For instance, four species of
134 vultures (the White-headed Vulture *Trigonoceps occipitalis*, White-backed Vulture *Gyps*
135 *africanus*, Hooded Vulture *Necrosyrtes monachus* and Rüppell's Vulture *Gyps rueppelli*)
136 were all listed as LC in 2004 but are now all classified as CR. Information about these
137 changes can be accessed through the IUCN history record and can then be used to inform
138 models of extinction risk.

139 While many previous studies have applied the IUCN-based extinction probabilities (criterion
140 E) outlined above to model extinction risks, a recent study by Monroe et al. (2019) has
141 presented an alternative approach, avoiding these probabilities altogether. Instead Monroe et
142 al. (2019) modeled extinction risks based on observed transitions of species to the statuses
143 EW or EX, which they then applied to model future biodiversity losses. While this approach
144 avoids the above mentioned shortcomings of the IUCN extinction probability approach,
145 namely the caveats surrounding the extrapolation of the criterion E specific extinction
146 probabilities to species listed under other criteria (A-D), it is likely limited to groups of
147 organisms with sufficient recorded transitions to EX in order to yield reliable extinction risk
148 estimates.

149 In this study we contrast different variations of both approaches, to which we refer to
150 hereafter as "critE EX mode" (approach based on IUCN criterion E extinction probabilities,
151 sensu Mooers et al. 2008) and "empirical EX mode" (approach based on historic transitions
152 towards statuses EW/EX, sensu Monroe et al. 2019). We add improvements to both
153 approaches, including the incorporation of GL data for the critE EX mode and the
154 consideration of possibly extinct taxa in the empirical EX mode. Further we present a novel
155 MCMC-based approach of estimating status transition rates from historical IUCN data, and
156 we apply these rates to simulate future status changes and extinctions. All approaches
157 presented in this study are available in the new open-access simulation program *iucn_sim*,
158 which can be run via the bash command line and is tested for compatibility in Windows,
159 MacOS, and Linux (Fig. 1, see Data availability statement).

160 **Material and methods**

161 Here we describe our approach of simulating future extinctions and IUCN status transitions
162 on the example of birds (Aves). We use the terms "reference group" and "target species" as
163 follows:

- 164 • **Reference group:** The group of species, whose IUCN history is being used to
165 estimate status transition rates, i.e. the rates at which species change between IUCN
166 statuses
- 167 • **Target species:** The group of species, for which future extinctions are being
168 simulated, while applying the estimated status transition rates

169 In this study the reference group and target species list consist of the same taxa (all extant
170 bird species), but this is not a requirement for this approach. For example, using our approach
171 one could simulate future extinctions and status transitions for a specific bird family or local
172 bird fauna, while using all birds as a larger reference group, in order to get reliable status
173 transition rate estimates.

174 To make this approach accessible and easy to use for future projects, we wrapped the
175 complete workflow described below into the open-source command line program *iucn_sim*
176 (Fig. 1, https://github.com/tobiashofmann88/iucn_extinction_simulator), which can be easily
177 installed together with all software dependencies using the conda package manager
178 (<https://docs.conda.io>). Installation instructions are available on the projects GitHub page of
179 this project. Using *iucn_sim*, it is straight-forward to 1) model future changes in IUCN status,
180 2) simulate possible times of extinction across species, and 3) estimate species-specific
181 extinction rates for any given set of species over a user-defined time span. We executed all

182 steps outlined below using *iucn_sim* (except the downloading of GL data) and we report the
183 *iucn_sim* command for each step in the Supplementary Code Sample 1.

184 **Generation length estimates**

185 We downloaded GL estimates for all bird species from Bird et al. (2020), following the
186 IUCN 2019-v2 taxonomy of extant bird species. Since the collecting of GL data can be
187 challenging for some groups, we provide instructions how to generate GL data via
188 phylogenetic imputation on the example of birds in Appendix 1 in the Supplementary
189 Material. See Cooke et al. (2018) and Bird et al. (2020) for more detailed instructions and
190 information on generating GL estimates for species.

191 **Downloading IUCN data**

192 We downloaded the complete IUCN history for the reference group (class Aves) from the
193 year 2001 onward, to ensure compatibility with the IUCN v3.1 standard (IUCN Species
194 Survival Commission 2001), using the *rl_history* function of the R-package rredlist
195 (Chamberlain 2017) and IUCN v2019-2. The taxon list for this download was generated by
196 scanning through the entire IUCN Red List catalogue for species assigned to the class Aves,
197 using the *rl_sp* function. In addition to the historic data, we extracted the current status (most
198 recent status assessment) of all target species (all Aves species) using the *rl_search* function.
199 For all following operations we set the status of all EW species to EX.

200 **Status transition rates**

201 Based on the IUCN history data, we counted all types of status changes that have occurred in
202 the IUCN history of birds (Table 1), as well as the cumulative amount of time spent in each
203 status across all bird species. For instance, if a given species was classified as NT from 2001

204 to 2005 and then EN from 2005 to today (2020), this species contributes 1 status change from
205 NT to EN and 4 years in NT and 16 years in EN.

206 From these counts we estimated the rates of transitions between all pairs of statuses using
207 Bayesian sampling. For example, if N_{ij} transitions were observed from status i to status j and
208 the cumulative time spent in i across all species in the reference group is t_i , we used a
209 Markov chain Monte Carlo (MCMC) algorithm to sample the annual transition rate q_{ij} from
210 the following posterior:

$$P(q_{ij}|N_{ij}, t_i) \propto P(N_{ij}, t_i|q_{ij}) \times P(q_{ij}) \quad \text{Eq. 1}$$

211 where the log likelihood function is that of a Poisson process describing status change

$$\log P(N_{ij}, t_i|q_{ij}) \propto N_{ij} \log(q_{ij}) - q_{ij} t_i \quad \text{Eq. 2}$$

212 and $P(q_{ij}) \sim \mathcal{U}[0,1000]$ is a vague uniform prior on the transition rate.

213 To incorporate the uncertainties in the transition rate estimates, we took 100 samples from the
214 posterior distribution of the rate estimates (Eq. 1) for each transition type (Table 2). More
215 specifically we populated 100 q-matrices containing the sampled rates for each type of status
216 transition. These q-matrices were used for future simulations, allowing us to simulate
217 potential future status changes of any species, given its starting status (see more detailed
218 explanation below). In addition, we sampled 100 transition rates for all transition types from
219 DD to any of the statuses LC, NT, VU, EN, and CR, which we used during the future
220 simulations to draw a new valid status for DD species.

221 Finally, we modelled the transition rates towards extinction (EX) from any extant status i
 222 ($q_{i \rightarrow EX}$). Modelling these transition rates towards EX is non-trivial and we used two
 223 approaches to estimate these rates, which we refer to as the critE EX mode and the empirical
 224 EX mode. The critE EX mode is based on the IUCN criterion E extinction probabilities
 225 defined for threatened statuses (sensu Mooers et al. 2008), whereas the empirical EX mode is
 226 based on empirically observed transitions towards EW/EX in the IUCN history (sensu
 227 Monroe et al. 2019).

228 The final q-matrices contain all status transition rates, including the rates towards extinction
 229 for each given status, determined with either the critE EX mode or the empirical EX mode
 230 outlined below (last column). Since in our simulations, species are not allowed to reappear
 231 after extinction, we set the rates from EX to any other status equal to 0:

$$\begin{pmatrix} & \mathbf{LC} & \mathbf{NT} & \mathbf{VU} & \mathbf{EN} & \mathbf{CR} & \mathbf{EX} \\ \mathbf{LC} & - & q_{LC \rightarrow NT} & q_{LC \rightarrow VU} & q_{LC \rightarrow EN} & q_{LC \rightarrow CR} & q_{LC \rightarrow EX} \\ \mathbf{NT} & q_{NT \rightarrow LC} & - & q_{NT \rightarrow VU} & q_{NT \rightarrow EN} & q_{NT \rightarrow CR} & q_{NT \rightarrow EX} \\ \mathbf{VU} & q_{VU \rightarrow LC} & q_{VU \rightarrow NT} & - & q_{VU \rightarrow EN} & q_{VU \rightarrow CR} & q_{VU \rightarrow EX} \\ \mathbf{EN} & q_{EN \rightarrow LC} & q_{EN \rightarrow NT} & q_{EN \rightarrow VU} & - & q_{EN \rightarrow CR} & q_{EN \rightarrow EX} (GL) \\ \mathbf{CR} & q_{CR \rightarrow LC} & q_{CR \rightarrow NT} & q_{CR \rightarrow VU} & q_{CR \rightarrow EN} & - & q_{CR \rightarrow EX} (GL) \\ \mathbf{EX} & 0 & 0 & 0 & 0 & 0 & - \end{pmatrix} \quad \text{Eq. 3}$$

232 1) CritE EX mode

233 In the critE EX mode approach (*iucn_sim* setting: `--extinction_probs_mode 0`) we
 234 transformed the extinction probabilities (E_t) associated with threatened IUCN statuses (see
 235 Introduction), defined over specific time frames (t), into annual status-specific EX transition
 236 rates ($q_{i \rightarrow EX}$), using the formula provided by (Kindvall and Gärdenfors 2003):

$$q_{i \rightarrow EX} = 1 - \sqrt[t]{1 - E_t} \quad \text{Eq. 4}$$

237 Since the IUCN extinction probabilities are only defined for the statuses VU, EN, and CR,
238 we extrapolated the annual EX transition rates for the remaining statuses LC and NT by
239 fitting a power function to the calculated extinction rates for the statuses VU, EN, and CR,
240 estimating the parameters a and b (Appendix 1):

$$241 \quad q_{i \rightarrow EX} = ax^b$$

242 where x represents the index of the IUCN category, sorted by increasing threat (i.e. $x_{LC} =$
243 $1, x_{NT} = 2, \dots, x_{CR} = 5$). After estimating the parameters a and b , we calculated the annual
244 transition rates to EX for statuses LC and NT, using the above power function.

245 According to the IUCN definition, the extinction probabilities linked to the IUCN categories
246 EN and CR for individual species are dependent on the GL of these species. In order to
247 properly model the EX transition rates for these statuses on a species-specific basis, we
248 applied GL data to adjust the time frame (t) associated with the extinction probabilities. For
249 example for a species with a GL of 5 years, which is categorized as CR (IUCN definition:
250 "50% extinction probability within 10 years or 3 generations, whichever is longer"), the
251 annual EX transition rate according to Eq. 4 is $q_{CR \rightarrow EX} = 1 - \sqrt[3*GL]{1 - 0.5} = 0.045$, whereas
252 for a species with a GL of 2 years with status CR the EX transition rate is $q_{CR \rightarrow EX} = 1 -$
253 $\sqrt[10]{1 - 0.5} = 0.067$, because in the latter case $3 * GL < 10 \text{ years}$. From this follows that
254 when ignoring GL information and setting $t = 10$ for all species (e.g. Mooers et al. 2008),
255 the extinction risk for species with moderate or long generation times (>3.33 years) will be
256 overestimated (Fig 2).

257 We therefore applied the GL estimates of all individual bird species (Bird et al. 2020) for the
258 calculation of the yearly EX transition rates for the statuses EN and CR of each species. We
259 then added these GL-adjusted EX transition rates to the q-matrices containing the extant

260 status transition rates (Eq. 3) to simulate future status changes and extinctions. Although the
261 EX transition rates derived in this manner and the extant status transition rates are modelled
262 based on two different data sources, they can be combined in the same q-matrix since all of
263 these rates are expressed in the same unit (transition events per species and year). To evaluate
264 how the presence of GL data affects future extinction predictions, we produced an additional
265 set of q-matrices where we did not apply GL data ("no GL" scenario). Further, to evaluate the
266 effect of modeling future status changes, we produced additional sets of q-matrices where all
267 transition rates between extant statuses were set to 0 ("no status change" scenarios), for both
268 the GL and no GL scenario.

269 2) Empirical EX mode

270 In the empirical EX mode approach (*iucn_sim* setting: `--extinction_probs_mode 1`) we
271 estimated EX transition rates based on the observed transitions from any given extant status
272 to EX in the IUCN history of birds, *sensu* Monroe et al. (2019). Following the same
273 procedure we used to infer transition rates between other statuses, we counted the transitions
274 from any status to EX (Table 1) and used MCMC to sample from the posterior transition rate
275 distribution (Eq. 1), of which we randomly selected 100 samples for each type of transition.
276 In contrast to the approach of Monroe et al. (2019), we estimated a specific transition rate
277 from any given status to EX using our MCMC based approach, instead of only allowing
278 transitions from CR to EW/EX. However, due to no observed occurrences of transitions of
279 the statuses LC, NT, VU, and EN to EX in the IUCN history of birds, the estimated rates for
280 these types of transitions are negligible, effectively making these transitions events very
281 unlikely in our simulations of future extinctions. A further difference is that we did not
282 distinguish between the statuses EW and EX, but instead treated both statuses as extinct. We
283 inserted the 100 sampled rates into the last column of the q-matrix (Table 2). Since no GL

284 data replicates or other species-specific data were used in this approach, the same 100 q-
285 matrices were applied to all species for future simulations.

286 The empirical EX mode approach is likely to underestimate the true transition rate, as for
287 several threatened species there is insufficient evidence to categorize them as EX, although
288 they are likely to be extinct (IUCN 2020). In order to better account for this underestimation
289 bias, we generated another set of q-matrices, where we incorporated information on possibly
290 extinct species [CR(PE) *sensu* IUCN 2020]. Prior to determining the numbers of transitions,
291 we modeled these species as EX, starting from the date that they were categorized as PEX
292 (the empirical EX mode + PEX approach). This modeling of PEX species was only done for
293 the purpose of estimating EX transition rates, while these species were classified as CR as a
294 departure point for future simulations (see below).

295 **Simulating future extinctions**

296 We used the transition rates from the estimated q-matrices (Eq. 3) to simulate for all bird
297 species future transitions between extant statuses or to toward EX. Before simulating into the
298 future, each species was assigned its current IUCN status as starting status. For all species
299 currently assigned as DD, we randomly drew a new status based on a probability vector
300 consisting of the estimated transition rates leading from DD to the valid statuses LC, NT,
301 VU, EN, or CR. For all NE species we drew a new valid IUCN status based on the
302 frequencies of valid IUCN statuses across all bird species. We treated species that are
303 categorized as EW or EX by IUCN as irrevocably extinct and therefore did not include these
304 species in future simulations.

305 We modeled transitions as a Poisson process, by generating time-forward simulations for
306 each species based on exponentially distributed waiting times between transition events. For
307 a given current status i the waiting time until the next event is

308
$$\Delta t \sim \text{Exp} \left(\sum_{j \in S \setminus i} q_{ij} \right)$$

309 where $S \setminus i$ is the set of statuses excluding the current status i . The type of transition after the
310 waiting time Δt is then sampled randomly with probabilities proportional to the rates in $S \setminus i$.

311 We repeated these time-forward simulations for each species up to a pre-defined time of
312 $t_{max} = 100$ years after present, producing 10,000 simulations for all 6 approaches outlined
313 above: i) critE EX mode, ii) critE EX mode no GL, iii) critE EX mode no status change, iv)
314 critE EX mode no GL and no status change, v) empirical EX mode, vi) empirical EX mode +
315 PEX. For each simulation replicate, we repeated the drawing of a valid IUCN status for DD
316 and NE species, thus incorporating this uncertainty in the simulations.

317 From the simulation output we extracted for each species a) the extinction times $t_{\{EX\}}$ for
318 those simulation replicates where $t_{EX} < t_{max}$ or b) the waiting times of length t_{max} for those
319 simulation replicates where the species did not go extinct within the specified time window.

320 Next we used these extinction times and waiting times to estimate species-specific annual
321 extinction rates averaged across the simulated time window.

322 For a given set of extinction times and waiting times simulated for species i , we applied a
323 MCMC to obtain posterior samples of the extinction rate μ_i using the likelihood function of a
324 death process (Silvestro et al. 2019):

$$P(w|\mu_i) \propto \mu_i^D \times \exp(-\mu_i \sum_{j \in w} (w_j)) \quad \text{Eq. 5}$$

325 where D is the number of instances in which $w \leq t_{max}$, i.e. the number of simulation
326 replicates in which the species went extinct within the considered time window.

327 We sampled estimates of μ_i for each species throughout the MCMC from the posterior
328 distribution:

$$P(\mu_i|w) \propto P(w|\mu_i) \times P(\mu_i) \quad \text{Eq. 6}$$

329 where $P(\mu_i)$ is a uniform prior distribution ($\mathcal{U}[0,1000]$) set on the extinction rate. For each
330 bird species we exported the mean and the 95% HPD interval of the posterior extinction rate
331 estimates.

332 To compare the extinction rates between different approaches, we calculated the average
333 extinction rate for each approach by running an MCMC with the death process likelihood
334 function (Eq. 5), based on the simulated extinction dates of all bird species across 10,000
335 simulation replicates.

336 **Testing accuracy with synthetic data**

337 In addition to the empirical bird data, we validated our approach on simulated data to
338 determine the accuracy of the estimated transition rates and extinction rates.

339 To test the accuracy of the transition rates estimated from the IUCN history and the effect of
340 the size of the chosen reference group on these estimates, we simulated status transitions data
341 mimicking the empirical IUCN history data. We then simulated status changes over a time
342 period of 20 years for reference groups of 100, 1,000, and 10,000 species. The starting status

343 for each species was drawn randomly, based on the empirical frequencies of the current
344 IUCN status distribution across all birds. To produce realistic transition rates to use for our
345 simulations, we randomly drew these rates from a uniform range in log-space, ranging
346 between the minimum and the maximum empirical rate estimated for birds. We drew 30 rates
347 to reflect the 30 possible transition types between the six main IUCN statuses (LC, NT, VU,
348 EN, CR, and DD). We then simulated the change of IUCN statuses through time in the same
349 manner as described above for the future simulations for the empirical bird data, with the
350 difference that no extinction events were being modeled. We then used the simulated IUCN
351 history for all species to infer transition rates using MCMC as done for the empirical bird
352 data.

353 To evaluate the accuracy of the species-specific extinction rates estimated with our approach,
354 we simulated extinction times for 1,000 species under known extinction rates. The extinction
355 rates (μ) that were used for these simulations were randomly drawn from a uniform range (in
356 log-space) with a minimum and maximum value derived from the EX transition rates of the
357 statuses LC and CR respectively, as modeled in this study with the outlined IUCN extinction
358 probabilities approach. Based on the chosen number of simulation replicates, N extinction
359 time replicates (t_e) were drawn randomly from an exponential distribution with mean μ^{-1} for
360 each species. This simulation was repeated for 100, 1,000, and 10,000 replicates, in order to
361 test how many replicates are necessary for an accurate rate estimation.

362 **Results**

363 **Transition rates**

364 We counted a total of 919 status transitions between extant IUCN statuses in the IUCN
365 history data of birds between the years 2001 and 2020 (Table 1). Additionally, we counted 6

366 transitions from CR to EX. This count increased to 20 when additionally modeling the PEX
367 taxa as EX. The mean transition rates estimated from these counts, averaged across all 100 q-
368 matrix replicates for all species, can be found in Table 2. With our transition rate estimation
369 method, even transition types with zero-counts in the IUCN history are assigned a positive
370 transition rate, although these rates will be very small. Differences between estimated rates
371 can occur even if based on identical counts because of differences in the cumulative times
372 spent in each category (see Eq. 2), as can be seen when for example comparing the transition
373 rate from LC to EX with that from EN to EX, which differ by orders of magnitude (Table 2).
374 A comparison between the estimated EX transition rates based on the two main approaches
375 tested in this study (critE EX mode and empirical EX mode) show, that the empirical EX
376 mode, which uses empirical extinction events, leads to lower average transition rates towards
377 status EX for the threatened categories (Table 2).

378 The estimates based on our synthetic status transition data demonstrate that our approach
379 accurately recovers the transition rates that were used to simulate the data, yet the precision
380 of these estimates is strongly dependent on the size of the reference group (Fig. 3a). These
381 results suggest that it is recommendable to choose reference groups of preferably more than
382 1,000 species, because stochastic fluctuations of status counts below that threshold preclude
383 the estimation of transition rates with meaningful precision, particularly so for low rates.

384 **Future extinctions**

385 Our future simulations for birds based on the critE EX mode approach resulted in 738
386 predicted species extinction within the next 100 years (95% confidence interval: 669-809
387 species, Fig. 2). In comparison the empirical EX mode approach resulted in 57 predicted
388 extinctions within the same time frame (21-93), which increased to 127 (82-182) when

389 accounting for PEX species. The species-specific extinction rates for all birds are available in
390 the Supplementary Data.

391 The estimations of species-specific extinction rates from the simulated extinction times (Eq.
392 5) produces accurate rate estimates, yet it requires around 10,000 future simulation replicates
393 to ensure this accuracy also for very low rates, such as species starting as LC (Fig. 3b). These
394 species-specific rates differ significantly between the approaches tested in this study (Fig. 4).
395 The empirical EX mode consistently leads to lower rate estimates than the critE EX mode,
396 which is a direct result of the differences in EX transition rates in the q-matrix between these
397 two approaches (Table 2).

398 The average rate estimated across all birds for the critE EX mode was 6.98×10^{-4}
399 extinctions per species-years (ESY) (95% credible interval $6.97 - 6.98 \times 10^{-4}$). This rate is
400 to be understood as the average bird extinction rate expected over the next 100 years. The
401 rate for the empirical EX mode was estimated to be significantly lower at 5.09×10^{-5}
402 ($5.08 - 5.11 \times 10^{-5}$). When in addition modeling PEX taxa as extinct, the rate increased to
403 1.16×10^{-4} E/SY ($1.15 - 1.16 \times 10^{-4}$). These rate estimates fall within the same level of
404 magnitude as previous estimates for birds, such as the 2.17×10^{-4} E/SY estimated by
405 Monroe et al. (2019).

406 To further compare our results with those of Monroe et al. (2019), we additionally simulated
407 extinctions and estimated extinction rates within a time window of 500 years, to match the
408 time window addressed in their study, and we used the empirical EX mode setting to match
409 the approach taken in their study. This resulted in an average rate estimate of 1.37×10^{-4}
410 E/SY ($1.369 - 1.370 \times 10^{-4}$) for the empirical EX mode approach, which represents the
411 average rate expected for the next 500 years under this model. These rate estimates are
412 significantly lower than the 2.17×10^{-4} E/SY estimated by Monroe et al. (2019). Yet, the

413 predicted number of extinctions under our approach ranged between 271 and 791, which
414 largely overlaps with the 226 to 589 extinctions predicted by Monroe et al. (2019).

415 This discrepancy in rate estimates reflects a difference in how the rates are estimated and
416 what they represent. Monroe et al. (2019) calculated their rate as the inverse of the average
417 expected longevity (time until extinction) based on all birds. This corresponds to the average
418 extinction rate of a process running until the extinction of all species. Our rate estimate, on
419 the other hand, is based on simulated extinction events over the next 500 years and therefore
420 reflects the average extinction rate within that time frame. Because in both approaches
421 species threat statuses evolve according to an asymmetric transition matrix (Eq. 3, Table 2),
422 the extinction process is not time-homogenous, as also noted by Monroe et al. (2019).

423 Extinction rates increase through time as a consequence of the trend towards increasing
424 frequencies of high threat statuses (Fig. 2). Consequently, rates averaged over a shorter time
425 window (as the 500 years simulated in our case) are expected to be lower than rates averaged
426 over the much longer time window reaching until the time of the expected extinction of all
427 birds (Monroe et al. 2019).

428 Our rate estimates provide a representation of the extinction process specifically within the
429 time window of interest and they change accordingly to the chosen simulation time. This is
430 the reason why the rate estimates for the empirical EX mode approach, reported above, differ
431 between the 100 year and the 500 years simulations (5.09×10^{-5} , and 1.37×10^{-4}
432 respectively).

433 To better compare the estimated rates between Monroe et al. (2019) and our study, we
434 calculated the expected number of extinctions under each rate for the time interval of 500
435 years. Using the properties of a death process and assuming time-homogeneous rates within

436 this interval we can compute the expected number of extinctions (D) based on a number N of
437 initial species within a time interval t as:

$$438 \quad D = N \times (1 - \exp(-\mu t))$$

439 where the second term of the multiplication is the probability of surviving until time t given
440 the extinction rate μ . Using this formula with $N = 10,961$ (number of extant bird species
441 according to IUCN 2019-v2) and $t = 500$ years with our extinction rate ($\mu = 1.37 \times 10^{-4}$)
442 we obtain 726 expected number of extinctions, which is well within the range obtained from
443 our simulations (271 - 791). In contrast, the 2.17×10^{-4} rate of Monroe et al. (2019) predicts
444 1,127 extinctions for the same time frame. This differs from their reported range of 226 to
445 589 expected extinctions, which was not estimated based on that reported rate, but derived
446 from the expected longevities of all species based on an IUCN status transition q-matrix,
447 similar to the one used in our study. Their rate estimate was calculated subsequently as the
448 inverse of the average longevity across all birds and thus represents an overall rate averaged
449 across the complete time frame until the extinction of all birds. Our rate estimate on the other
450 hand is specific to the chosen simulation time window and describes more adequately the
451 extinction process within that window. This demonstrates the utility of our **iucn_sim**
452 program, which can be applied in future studies to predict extinction rates for specified time
453 frames for any organism group or for individual species.

454 **Effect of modeling status change and GL data**

455 Our empirical results show that accounting for GL data decreases the resulting extinction rate
456 estimates (Fig. 5). As an example we highlight this effect for the Red-headed Vulture
457 (*Sarcogyps calvus*), which is categorized as CR and has a relatively long generation length of
458 15 years (IUCN Red List 2019). The reduction of extinction probability when including GL

459 is expected to be particularly strong for CR species with long GL times, since the immediate
460 extinction probability applied in the simulations for EN and CR species decreases when
461 incorporating the GL information, according to IUCN definition (critE EX mode). But the
462 GL effect will also apply to LC species, as highlighted for the Turkey Vulture (*Cathartes*
463 *aura*, GL = 9.9 years), where incorporating GL data leads to a small decrease in the
464 extinction rate estimates, since occasionally these species will transition to the categories EN
465 or CR in the future simulations, when allowing for future status changes (Fig. 5a). Overall,
466 accounting for GL data leads to a decrease in the number of predicted extinctions across the
467 whole target group (Fig. 6).

468 The effect of modeling future IUCN status changes can increase or decrease the estimated
469 extinction rates of a species, depending on its current status and on the transition rates
470 between statuses. Therefore, this effect is expected to be dependent on the chosen reference
471 group. However, for LC species this generally appears to lead to an increase in the estimated
472 extinction rates (Fig. 5c), which is likely because these species can only change to a more
473 threatened status (LC being the least threatened status). Similarly, for CR species, the effect
474 of modeling future status changes generally leads to a decrease in extinction rates (Fig. 5d),
475 since species can only switch to less threatened categories in the future (CR being the most
476 threatened status). Overall, modeling future status changes leads to a sharp increase in the
477 number of predicted extinctions across the whole target group (Fig. 6), compared to the
478 scenario with no future status changes.

479 **Discussion**

480 **Utility of the *iucn_sim* program**

481 With our open-source program *iucn_sim* that accompanies this study, we are presenting
482 improved versions of the two main approaches of previous studies for modeling future
483 biodiversity losses based on IUCN status assessments (Fig. 1). Through this, we hope to
484 facilitate future studies to apply these workflows for generating future diversity predictions
485 and for estimating extinction rates for whole groups or individual species. The program is
486 easy to use and to simulate future extinctions it requires only a list of target species names, or
487 even just the name of the taxonomic group, as it automatically retrieves all available IUCN
488 information. Moreover, *iucn_sim* also allows for additional data input for more specific
489 estimates, such as GL data, that the user can choose to provide.

490 One of the main outputs of the program is the predicted number of future species extinctions
491 for a given group of species, as well as the future changes of the IUCN status distribution
492 within the group (Fig. 2). The program also calculates the probabilities of each species being
493 extinct by a user-defined date. Finally, the program estimates the extinction rates based on
494 the simulated extinction dates separately for each species (Fig. 4). These species-specific
495 extinction rates can be of interest for downstream analyses where species are required to be
496 modeled individually based on biological or geographic data, and where the extinction
497 dynamics of specific species or groupings of species are of interest (Davis et al. 2018, Cooke
498 et al. 2019, Pimiento et al. 2020).

499 We note that the actual extinction rates of a given species or group are expected to vary over
500 time as a function of changes in the IUCN status, while the extinction rates inferred by
501 *iucn_sim* are a time-averaged proxy of this process. Particularly during the current human-

502 induced wave of extinctions, extinction rates are expected to vary within relatively short time
503 frames of at least 100s of years (Ceballos et al. 2015). Therefore, our approach presented here
504 may not be suitable for estimating extinction rates based on simulations that span across
505 several hundred years or more.

506 Our method further allows for modeling DD species for which IUCN statuses are imputed
507 based on historical transition rates that reflect how frequently DD species change to other
508 statuses. Similarly, a new status for NE species is modeled based on the current status
509 distribution of the reference group. As a status is imputed for DD and NE species at each
510 simulation replicate, our method incorporates the full uncertainty concerning their true status.
511 The *iucn_sim* program flags and prints to the screen the names of those species that cannot
512 be found in the IUCN taxonomy and produces a warning for the user to revise the taxonomy.
513 If these cases cannot be fixed by the user, they will be treated as NE. This approach enables
514 future simulations even for groups where it is difficult to match the taxonomy with that of
515 IUCN, yet we recommend thoroughly revising the taxonomies to minimize the number of
516 taxonomic mismatches. Never the less, species unknown to IUCN, which are modeled in this
517 manner, are not expected to bias the overall future biodiversity predictions (under the
518 assumption that these taxa constitute a random sample of the target species group), due to
519 their status being repeatedly resampled based on the empirical status distribution of the
520 reference group. While these species are not expected to affect the overall predictions for the
521 target group, the resulting species-specific extinction rates for these taxa on the other hand
522 may be misrepresentative. To address this issue, the user can manually change the status of
523 NE species to a status they deem more representative for the species, by altering the status in
524 the *species_data.txt* text file produced by *iucn_sim*.

525 Our *iucn_sim* program further allows the simulation of different future conservation
526 scenarios, through simple q-matrix modifications. For example, one can simulate an increase
527 of conservation efforts by providing a specific conservation factor. This factor is then applied
528 to all transition rates in the q-matrix, leading to an improvement in conservation status for
529 each species. Similarly, one can simulate increased threats by providing a threat factor, which
530 is then applied to all threat-increasing transition rates in the q-matrix. These factors can also
531 be set to 0 to simulate scenarios that do not allow for future improvements or increased
532 threats. This flexibility of *iucn_sim* makes it easy to simulate and compare different future
533 scenarios and their expected effect on biodiversity.

534 **Comparing approaches to simulate future extinctions**

535 The critE EX mode and empirical EX mode approaches that were applied in this study
536 represent different ways of modeling the EX transition rates, which are the rates at which
537 species transition towards extinction in our future simulations. These are not to be confused
538 with the species-specific extinction rates, which are instead estimated from the simulated
539 extinction times and describe the extinction risk of individual species.

540 The critE EX mode makes use of extinction probabilities that are defined by the IUCN as one
541 of several criteria for species assessments of threatened species. Although widely used in the
542 scientific literature for modeling species' extinction risks (Veron et al. 2016, Davis et al.
543 2018, Cooke et al. 2019, Oliveira et al. 2019), these probabilities are not originally intended
544 for this purpose and per definition only apply to the subset of threatened species that was
545 assessed under criterion E (Akçakaya et al. 2006). The simulated extinctions resulting from
546 this approach are alarmingly high and the estimated extinction rates are in most cases more
547 than an order of magnitude higher than those estimated with the empirical EX mode
548 approach, even when accounting for PEX taxa in the latter approach.

549 The empirical EX mode, on the other hand, will likely lead to an underestimation of the true
550 extinction rates, because it is directly dependent on the number of observed transitions from
551 extant categories to EW or EX in the IUCN history, and these documented numbers are likely
552 a significant underestimate (IUCN 2020). This underestimation bias is due to rather strict
553 requirements to classify species as EW or EX. In 2020, IUCN therefore released a list of
554 species that are possibly extinct (PEX species), but do not qualify as EX according to the
555 IUCN guidelines. Making use of this information (which is available in *iucn_sim*) and
556 modeling these taxa as extinct, usually leads to more observed status transitions towards EX
557 within the last 20 years of IUCN history and therefore leads to higher EX transition rate
558 estimates that are expected to better reflect the true extinction risk within the studied group.
559 However, this approach is expected to be sensitive towards small reference groups with very
560 few or no observed extinctions, which will lead to high uncertainties in the estimated rates.
561 Given these significant differences in predicted future estimates between the two approaches,
562 it is important to consider that these approaches are based on different assumptions and while
563 both can theoretically be applied for any organism group, their utility varies depending on the
564 group and purpose of the future simulations.

565 If the primary aim is to conservatively model future biodiversity losses for a given group of
566 species, and if this group can be meaningfully represented by a reference group that is a) well
567 represented in the IUCN Red List (i.e. many assessed species), b) has a high species
568 diversity, and c) has several recorded extinctions throughout the last 20 years, then the
569 empirical EX mode including PEX taxa may be the most suitable choice, as in this case EX
570 transition rates can be meaningfully modeled for the specific reference group, rather than
571 being based on general pre-defined extinction probabilities.

572 If, on the other hand, the primary aim is to produce species-specific extinction rates for
573 downstream analyses, and GL data is available or can be modeled for the group of species,
574 then the critE EX mode approach may be the more appropriate choice, as it leads to a larger
575 variation of rate estimates. This variation is expected to reflect differences in how threatened
576 with extinction species of the same category are, based on their differences in GL. In the
577 empirical EX mode approach on the other hand all species belonging to a given category are
578 modeled equally, only leading to small stochastic differences between the rates of species
579 belonging to the same status.

580 **Choice of reference group**

581 Essential to both approaches discussed above is the choice of the reference group, because
582 the precision and accuracy of the estimated transition rates depends on the number of species
583 in the reference group (Fig. 3). There are two main considerations to make when choosing a
584 reference group: 1) Is the chosen group expected to reflect the trends of status change for the
585 target species that are being simulated? and 2) Does the reference group contain a sufficient
586 number of species so that stochastic effects do not overrule the actual trends for that group?

587 These two objectives can conflict, as illustrated by the example of simulating future
588 extinctions for vultures. In that case, using all birds (class Aves) as reference group (~ 11,000
589 species) provides a large enough group of sufficient size for accurate transition rate
590 estimations. However, given the notable recent worsening of almost all vulture species'
591 conservation status (e.g. Green et al. 2007), the trends observed over all birds may not be
592 representative of this group.

593 The species in the reference group do not necessarily have to form a monophyletic clade,
594 although phylogenetically related taxa are likely to provide a suitable reference group if there

595 is any phylogenetic signal in extinction risk. More importantly, a suitable reference group
596 consists of species that are expected to share the same extinction threats as the group of target
597 species for which to simulate future extinctions, so that representative status transition rates
598 can be inferred. A reference group could include species that share a similar ecology and are
599 similarly affected by habitat losses or pollution, or it could include species from the same
600 biogeographic area as the target species if they are expected to share common threats, such as
601 is the case for many island faunas. For these reasons, the reference group should also always
602 contain all of the target species, although this is not an analytical requirement.

603 **Conclusions**

604 In this study, we demonstrated that modeling future changes in IUCN conservation status and
605 incorporating generation length data has a substantial effect on future extinction predictions.
606 In addition, we encountered significant differences in extinction rate predictions when
607 comparing different approaches of modeling extinction risks. This shows that the results of
608 future projections are strongly dependent on the selected method and the selected reference
609 group.

610 The aim of this study was to make the simulation of future extinctions under different
611 approaches accessible for future projects. Further, this study can provide a starting point for
612 researchers to decide which approach to choose for their specific target group and research
613 objective.

614 Our software *iucn_sim* is designed for ease of use and contains many options for adjusting
615 the simulation approach for different types of projects. The source code on GitHub is open
616 for contributions and feedback from users, which hopefully will lead to the incorporation of
617 further improvements for predicting future extinctions. Future additions to the program could

618 for example include more specific future modeling of species based on similarities in
619 biological traits, geographic location, or niche space.

620 **Data availability statement**

621 All source code, input files used in this study, and the output produced by *iucn_sim* are
622 available on the project's GitHub repository at
623 https://github.com/tobiashofmann88/iucn_extinction_simulator. The estimated extinction
624 rates for all bird species, a Supplementary Code Sample describing the *iucn_sim* workflow,
625 and Appendix 1, can be directly downloaded from the Supplementary Material
626 accompanying this study.

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- 691
- 692

693 **Tables**

694 **Table 1:** Status transitions counted in the IUCN history of birds (class Aves) between 2011-
695 2020. For example, the empirical count of transitions from status LC to NT is 182, while the
696 count of transitions from NT to LC is 112. The count for transitions from CR to EX changes
697 from 6 to 20 when modeling species that are possibly extinct according to IUCN (PEX) as
698 EX.

	LC	NT	VU	EN	CR	DD	EW/EX
LC	0	182	77	18	3	1	0
NT	112	0	73	24	3	1	0
VU	18	82	0	99	13	1	0
EN	1	15	69	0	50	0	0
CR	0	2	9	42	0	0	6 (+14)
DD	8	12	5	2	0	0	0
EW/EX	0	0	0	0	2	0	0

699

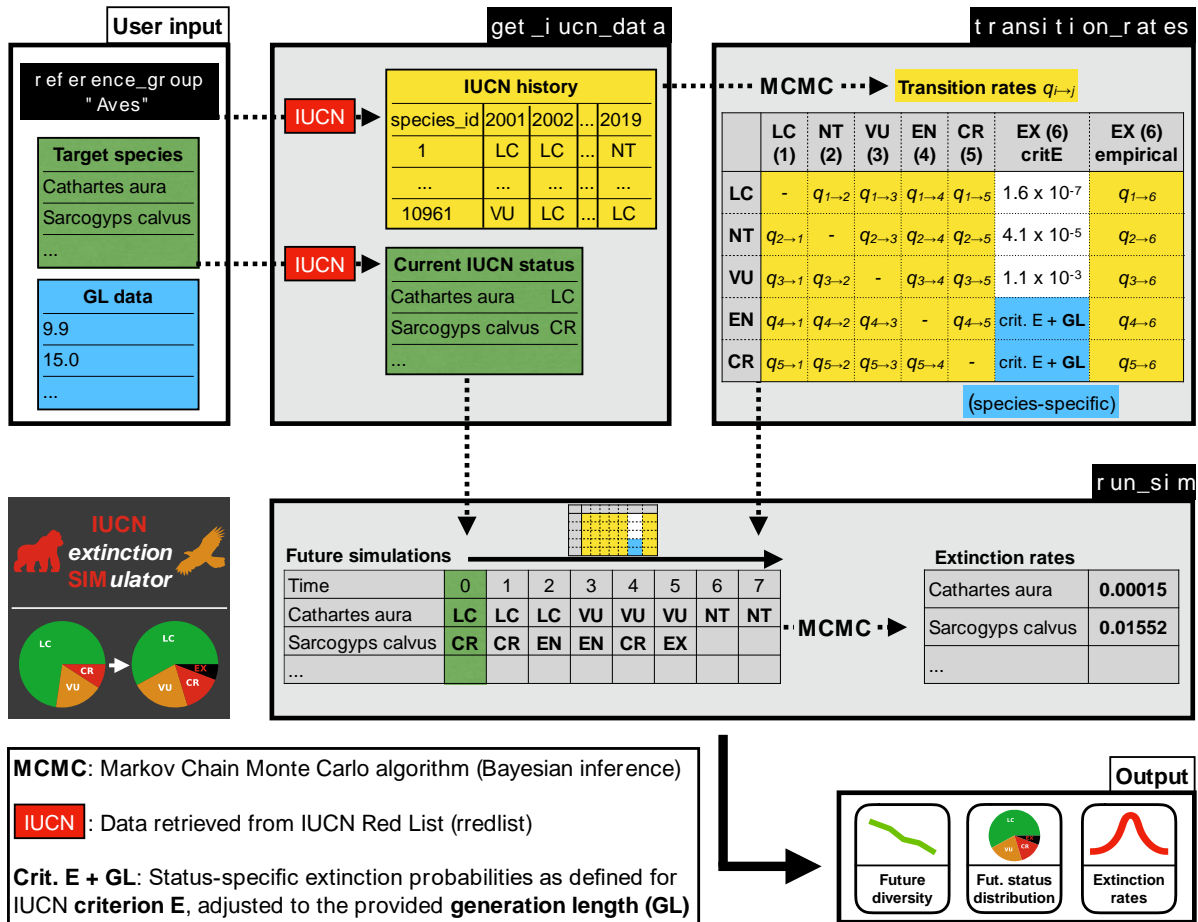
700

701 **Table 2:** Status transitions rates estimated for birds (class Aves) that were used for future
702 simulations. The q-matrix below shows the mean of the transition rate estimates across the q-
703 matrix replicates for all bird species, scaled in transitions per species-year (T/SY). The
704 transition rates between all extant statuses were estimated from the counted transitions in the
705 IUCN history of birds (**Table 1**), taking into account the total cumulative time across all bird
706 species spent in each category. The EX transition rates are shown for both approaches, the
707 critE EX mode and the empirical EX mode (including PEX taxa), respectively.

	LC	NT	VU	EN	CR	EX
LC	-	0.00146420	0.00063157	0.00015742	0.00003312	0.00000016 / 0.00000794
NT	0.00755463	-	0.00491179	0.00161410	0.00025690	0.00004155 / 0.00006458
VU	0.00164326	0.00691254	-	0.00831167	0.00114747	0.00105305 / 0.00007353
EN	0.00030274	0.00248436	0.01058826	-	0.00757056	0.00900926 / 0.00014377
708 CR	0.00033174	0.00100060	0.00319970	0.01443416	-	0.04767378 / 0.00654700

709

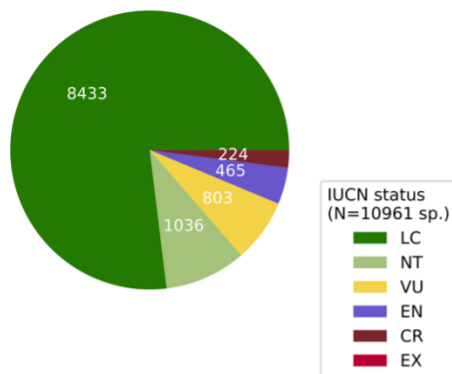
710 **Figures**



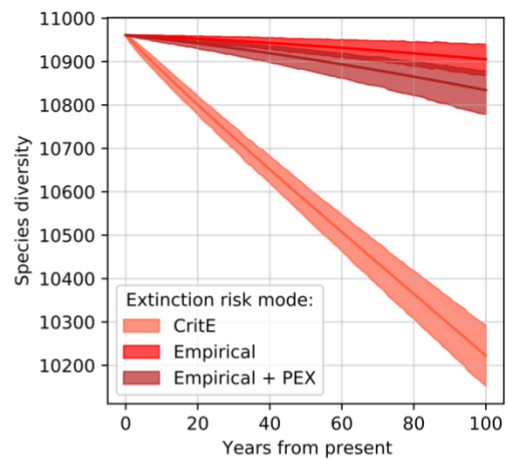
711

712 **Figure 1:** Workflow of *iucn_sim* to simulate future extinctions and estimate extinction rates.
 713 The only required input by the user is a) the list of target species whose future extinctions are
 714 supposed to be simulated and b) the name of a reference group, which will be used to
 715 estimate status transition rates based on the recorded IUCN history of this group. Optionally
 716 the user can provide generation length (GL) estimates for each target species, which will be
 717 considered when calculating the extinction risks associated with the statuses EN and CR,
 718 according to IUCN criterion E (critE EX mode). Alternatively, the user can choose the
 719 empirical EX mode, in which case extinction risks will be estimated from the empirically
 720 observed extinctions in the IUCN history of the reference group. The modeled extinction
 721 risks and the status transition rates will be stored in a q-matrix, which is used to simulate
 722 future status changes and extinctions for all target species. Finally, the program estimates
 723 species-specific extinction rates from the simulated extinction times (typically from multiple
 724 simulation replicates) and produces various summary statistics and plots as output, including
 725 the simulated future status distribution of the target group, the future diversity trajectory, and
 726 the histograms of simulated extinction times and extinction rate estimates for each species.

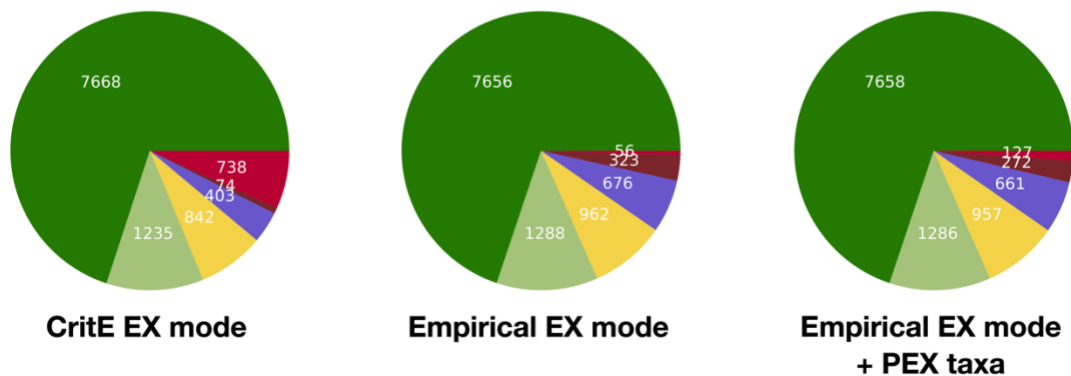
a) Present status distribution



b) Simulated diversity trajectory

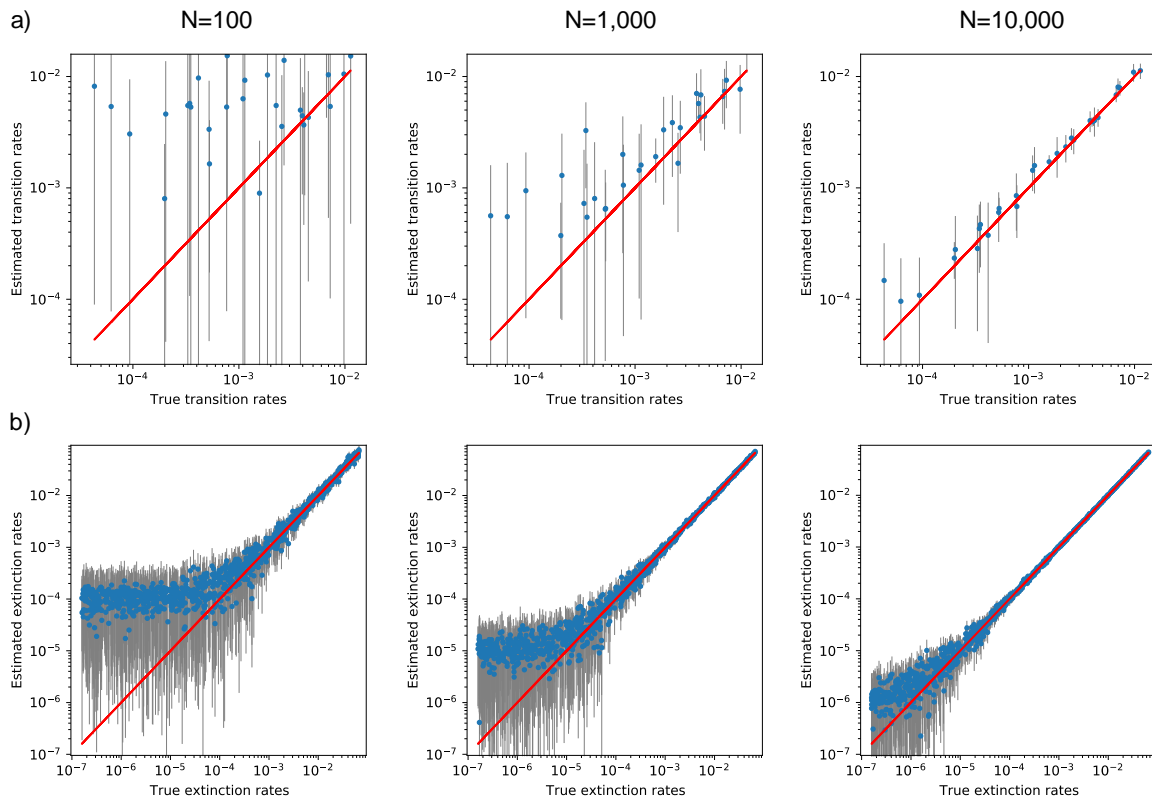


c) Status distribution in 100 years



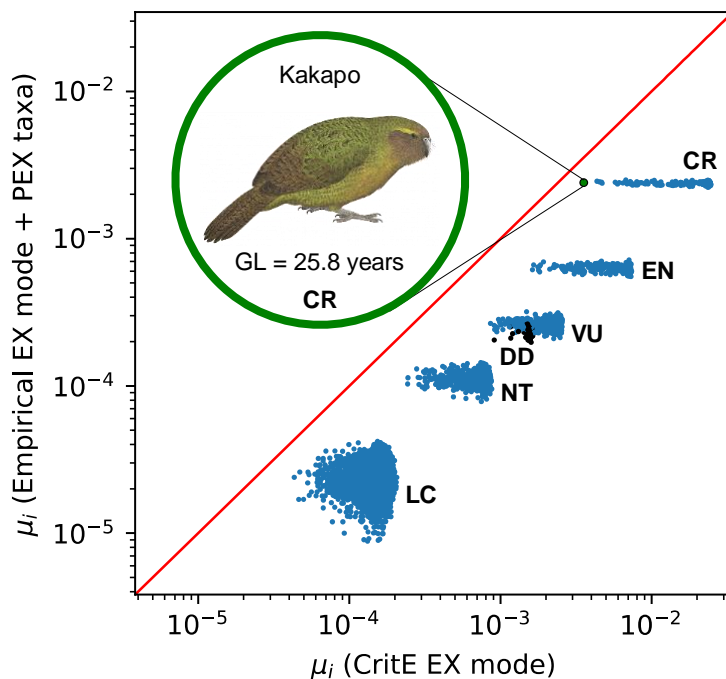
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728 **Figure 2:** Future diversity trajectory and IUCN status distribution for birds. We simulated
 729 future extinctions with three different approaches of modeling extinction risks: the critE EX
 730 mode, the empirical EX mode, and the empirical EX mode including the modeling of PEX
 731 species as extinct. Panel a) shows the current IUCN threat status distribution of all bird
 732 species. Panel b) shows the future diversity trajectory over the next 100 years, based on
 733 future extinctions simulated with *iucn_sim* under the 3 different extinction risk scenarios.
 734 Panel (c) shows the future IUCN status distribution in 100 years simulated with *iucn_sim*.
 735 Note that the y-axis in the diversity through time plots only displays a selected diversity
 736 range starting at 10,000 (displayed range does not cover the value 0). All panels represent
 737 graphic output options available in *iucn_sim*.



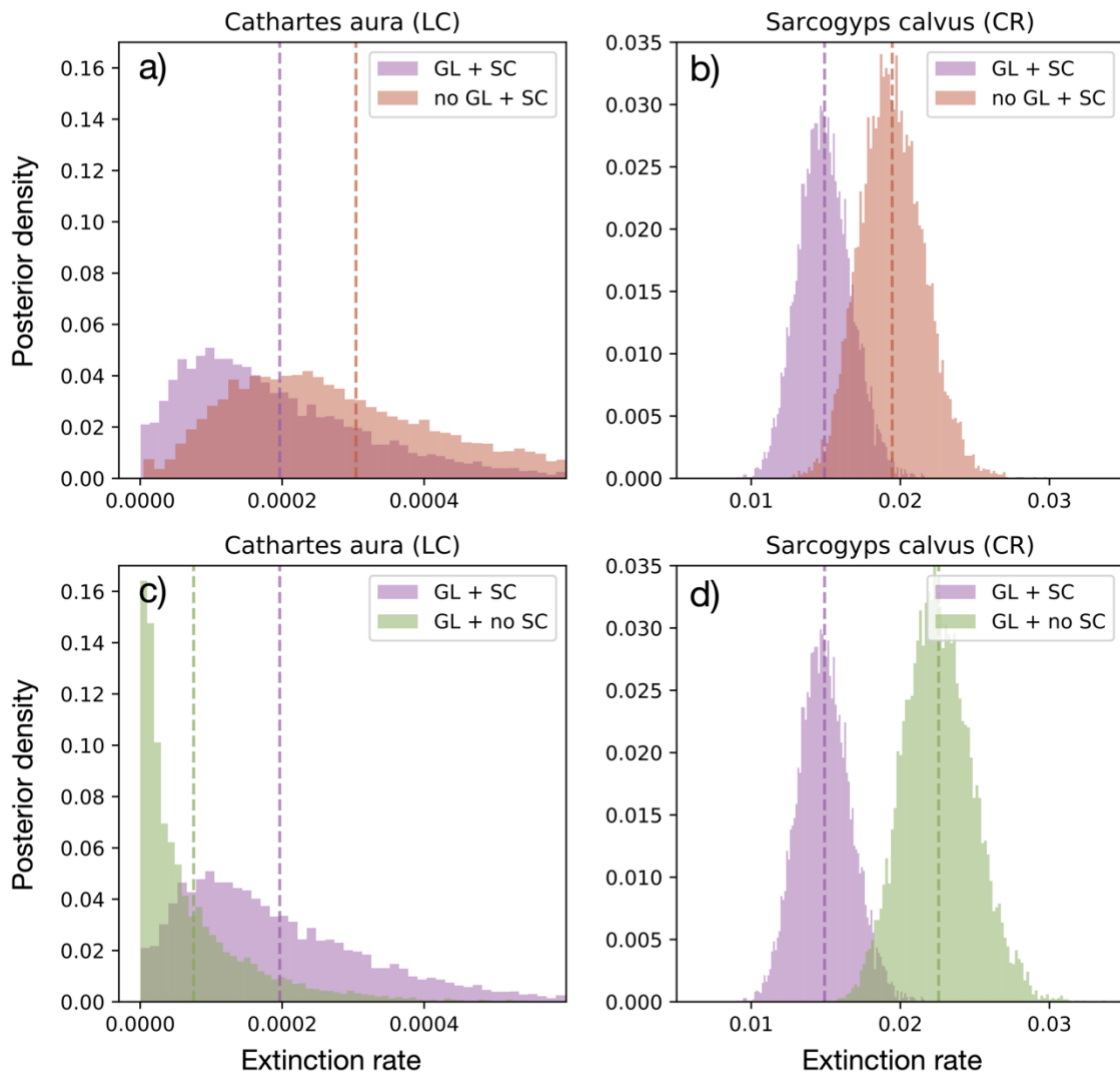
738

739 **Figure 3:** Accuracy of rate estimations improves with higher sample sizes. The scatter-plots
740 show the status transition rates (a) and the extinction rates (b), estimated from synthetic data
741 that was simulated in this study under known rates. We plotted the mean values (blue dots)
742 and the 95% credible interval (grey vertical lines) of the rates sampled by MCMC (y-axis)
743 against the true rates (x-axis) to evaluate the accuracy under different sample sizes (see plot
744 titles). The sample size in case of the status transition rates (a) constitutes the number of
745 species in the reference group, while sample size for extinction rates represents the number of
746 future simulations for each species. Rate estimates close to the diagonal red line show high
747 accuracy, while small error bars show high precision. Status transition rates estimated for
748 reference groups of only 100 species show very low accuracy and therefore it is
749 recommended to choose reference groups of at least 1,000 or more species. The dotted
750 horizontal line in the extinction rate plots (b) shows the minimum empirical extinction rate
751 estimate for the bird data ($\sim 1 \times 10^{-5}$). Extinction rates far below this line are therefore
752 unlikely to occur in empirical data sets. Running at least 10,000 simulation replicates ensures
753 accurate and precise extinction rate estimates.



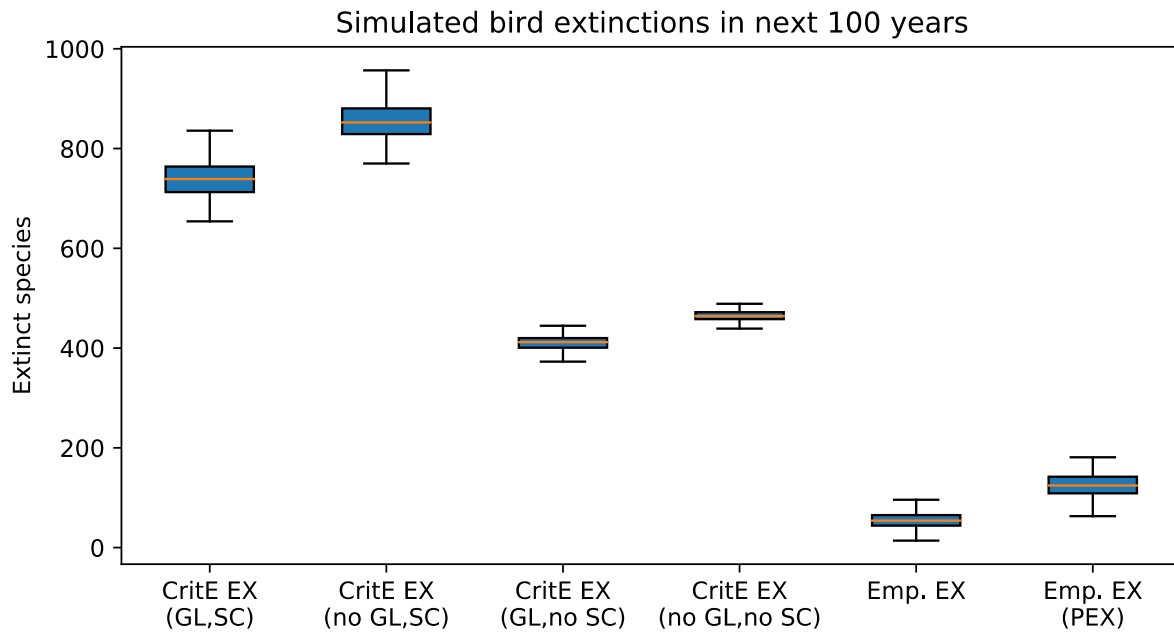
754

755 **Figure 4:** Species-specific extinction rates for the two *iucn_sim* approaches of modeling EX
756 transition rates. The x-axis shows the estimated rates for the critE EX mode approach (all
757 rates scaled in extinctions per species-years - ESY). The y-axis shows the estimated rates for
758 the empirical EX mode approach including PEX taxa modeling. The rates estimated from the
759 empirical EX mode approach are consistently lower than those from the critE EX mode
760 approach. Species with the same IUCN status at present end up with similar rate estimates,
761 forming visible clusters in the plot. However, there is some observed variation in the
762 estimates between species of the same status, which is present in the estimates of both
763 approaches (x and y-axis). This variation is partly caused by the stochasticity in our
764 simulation approach. But particularly for the more threatened categories EN and CR we find
765 additional variation among the critE EX mode rate estimates that is not present on the y-axis,
766 causing the elongated shapes of these clusters as opposed to the round shapes of the less
767 threatened status clusters. This is caused by differences in the GL values of individual
768 species, leading to smaller extinction rate estimates for species with long generation times, as
769 highlighted exemplarily for the CR Kakapo (*Strigops habroptila*), with one of the longest
770 generation lengths in our dataset (25.8 years), which places on the very low rate end of the
771 CR extinction rate cluster. In our approach DD species (black dots) are being modeled based
772 on the observed DD transition rates in the IUCN history of the reference group, which in the
773 case of birds results in extinction rate estimates similar to those of VU species. The
774 illustration was provided by the Handbook of the birds of the world alive (Collar, N. et al.
775 2020).



776

777 **Figure 5:** The effect of generation length (GL) and status-change (SC) on estimated
778 extinction rates. The plots show histograms of the posterior density of extinction rates
779 estimated with *iucn_sim* for two different species: the Turkey Vulture (*Cathartes aura*, GL =
780 9.9 years, LC), panels a) and c); and the Red-headed Vulture (*Sarcogyps calvus*, GL = 15
781 years, CR), panels b) and d). Upper panels show that the extinction rate estimates slightly
782 decrease when including GL data into the simulations (purple) compared to ignoring GL data
783 (red) for both LC and CR species. Bottom panels show that accounting for future changes of
784 IUCN statuses slightly increases the extinction rate of LC species, but leads to a decrease for
785 CR species (d). Note that the effect of future status changes on extinction rates depends on
786 the estimated status transition rates and is therefore expected to change depending on the
787 chosen reference group.



788

789 **Figure 6:** Number of predicted extinctions for birds in the next 100 years under different
790 simulation scenarios across 100 simulation replicates. The blue boxes show the lower to
791 upper quartile values of the predicted extinctions, with an orange line at the median. The
792 whiskers show the full range of the predictions. Including generation length (GL) and
793 conservation status changes (SC) into future simulations, leads to a significant increase in the
794 number of predicted extinctions, compared to ignoring this information (compare first and
795 fourth box plot column). The individual effect of adding GL information to the simulations is
796 a decrease of the predicted extinctions (third box-plot column), while only modeling SC leads
797 to very high numbers of predicted extinctions (second box-plot column). The last two
798 columns show the range of predicted extinctions for the empirical EX mode approach, with
799 (column 5) and without PEX taxa (column 6). The estimates for both empirical EX mode
800 approaches are significantly lower than those for any of the variations of the critE EX mode
801 approach (columns 1-4).

User input

reference_group
"Aves"

Target species

Cathartes aura
Sarcogyps calvus
...

GL data

9.9
15.0
...

get_iucn_data

IUCN

IUCN history

species_id	2001	2002	...	2019
1	LC	LC	...	NT
...
10961	VU	LC	...	LC

IUCN

Current IUCN status

Cathartes aura LC
Sarcogyps calvus CR
...

transition_rates

MCMC

Transition rates $q_{i \rightarrow j}$

	LC (1)	NT (2)	VU (3)	EN (4)	CR (5)	EX (6) critE	EX (6) empirical
LC	-	$q_{1 \rightarrow 2}$	$q_{1 \rightarrow 3}$	$q_{1 \rightarrow 4}$	$q_{1 \rightarrow 5}$	1.6×10^{-7}	$q_{1 \rightarrow 6}$
NT	$q_{2 \rightarrow 1}$	-	$q_{2 \rightarrow 3}$	$q_{2 \rightarrow 4}$	$q_{2 \rightarrow 5}$	4.1×10^{-5}	$q_{2 \rightarrow 6}$
VU	$q_{3 \rightarrow 1}$	$q_{3 \rightarrow 2}$	-	$q_{3 \rightarrow 4}$	$q_{3 \rightarrow 5}$	1.1×10^{-3}	$q_{3 \rightarrow 6}$
EN	$q_{4 \rightarrow 1}$	$q_{4 \rightarrow 2}$	$q_{4 \rightarrow 3}$	-	$q_{4 \rightarrow 5}$	crit. E + GL	$q_{4 \rightarrow 6}$
CR	$q_{5 \rightarrow 1}$	$q_{5 \rightarrow 2}$	$q_{5 \rightarrow 3}$	$q_{5 \rightarrow 4}$	-	crit. E + GL	$q_{5 \rightarrow 6}$

(species-specific)

run_sim

Future simulations

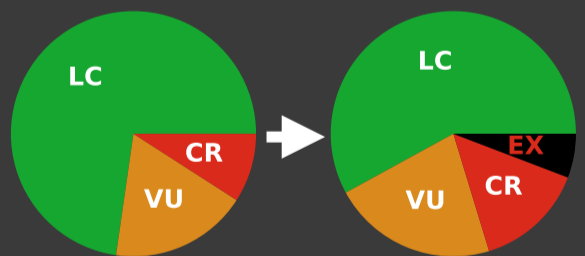
Time	0	1	2	3	4	5	6	7
Cathartes aura	LC	LC	LC	VU	VU	VU	NT	NT
Sarcogyps calvus	CR	CR	EN	EN	CR	EX		
...								

MCMC

Extinction rates

Cathartes aura	0.00015
Sarcogyps calvus	0.01552
...	

IUCN extinction SIMulator

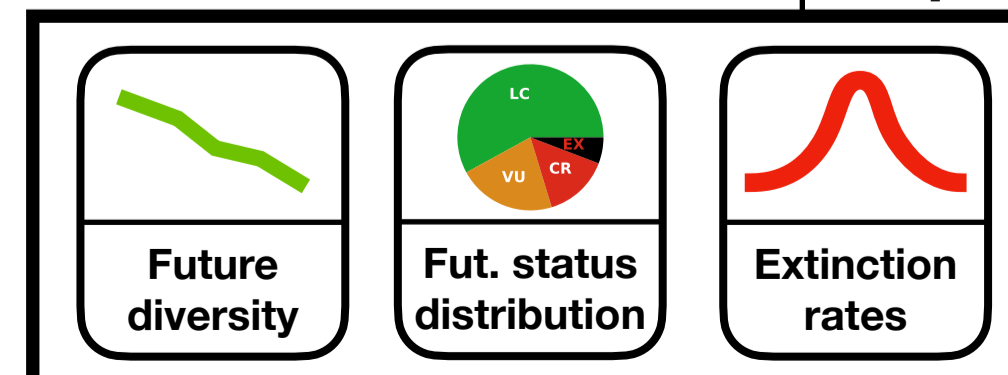


MCMC: Markov Chain Monte Carlo algorithm (Bayesian inference)

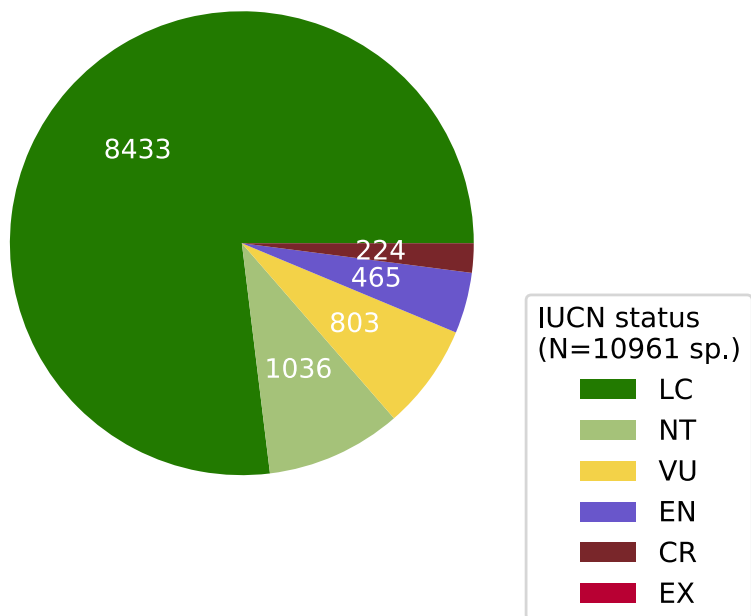
IUCN: Data retrieved from IUCN Red List (rredlist)

Crit. E + GL: Status-specific extinction probabilities as defined for IUCN criterion E, adjusted to the provided generation length (GL)

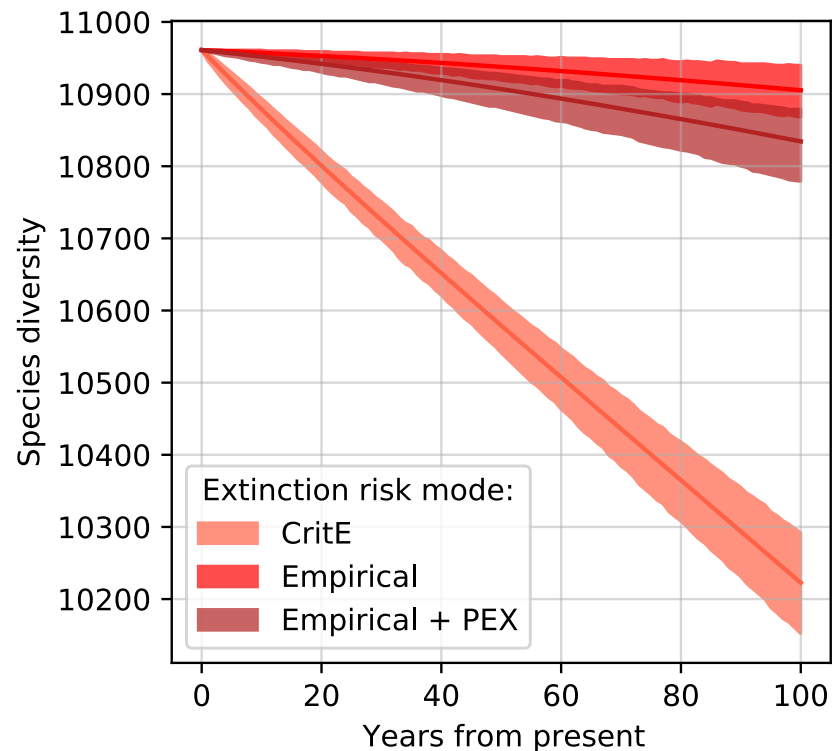
Output



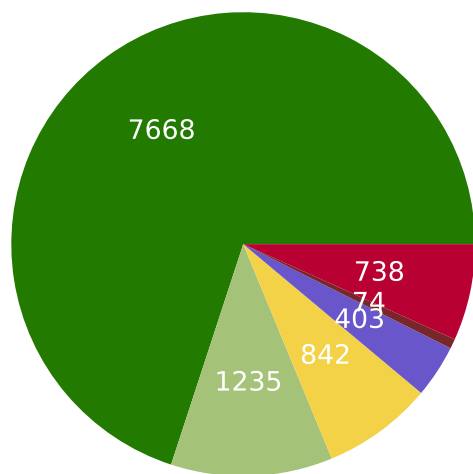
a) Present status distribution



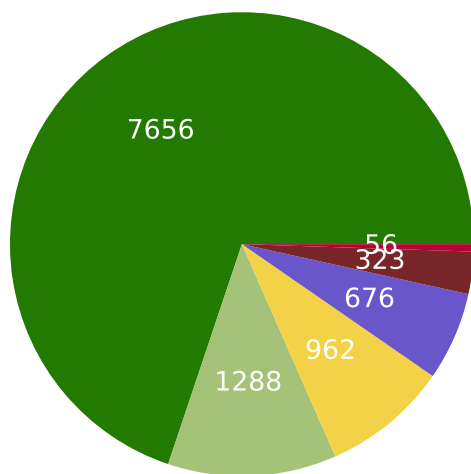
b) Simulated diversity trajectory



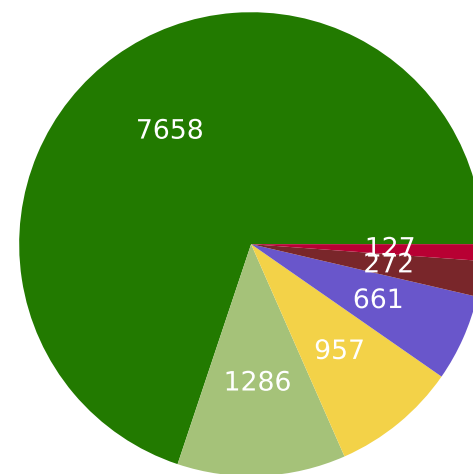
c) Status distribution in 100 years



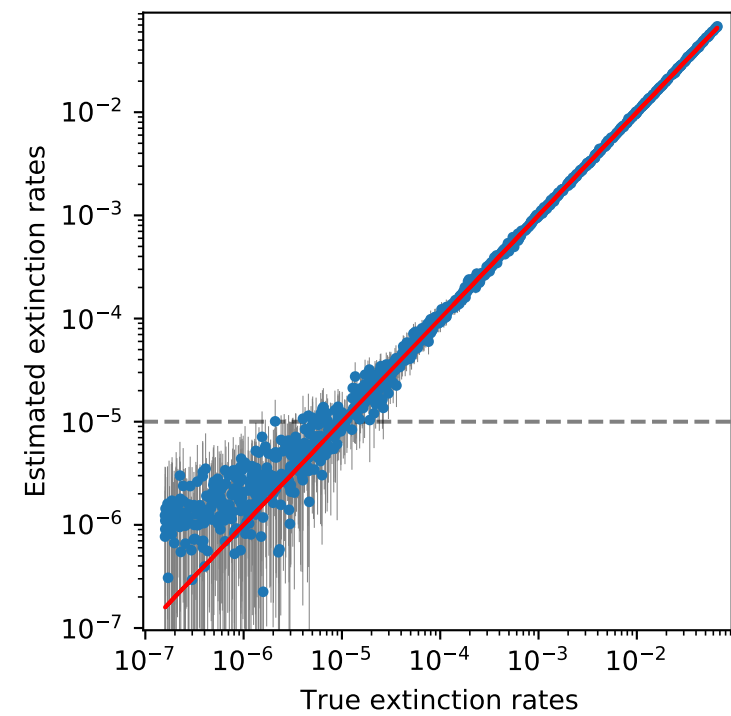
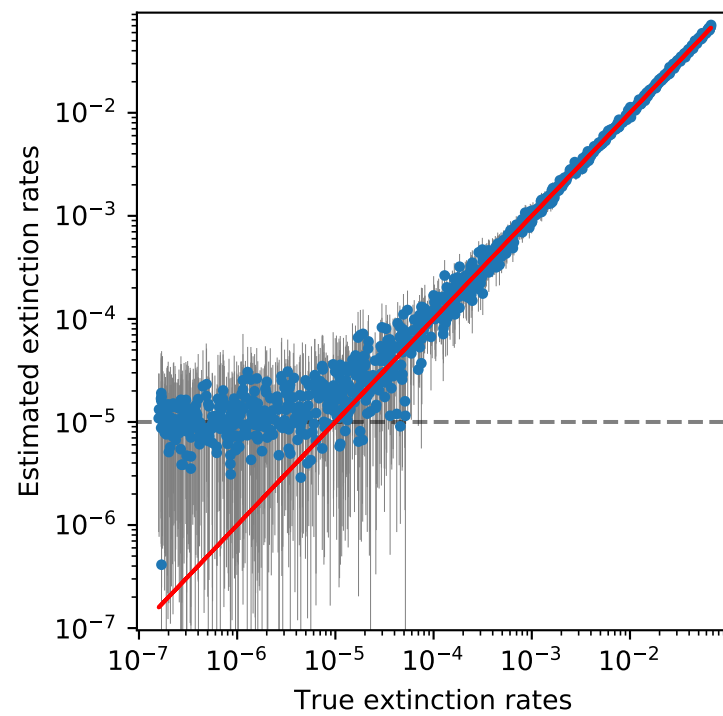
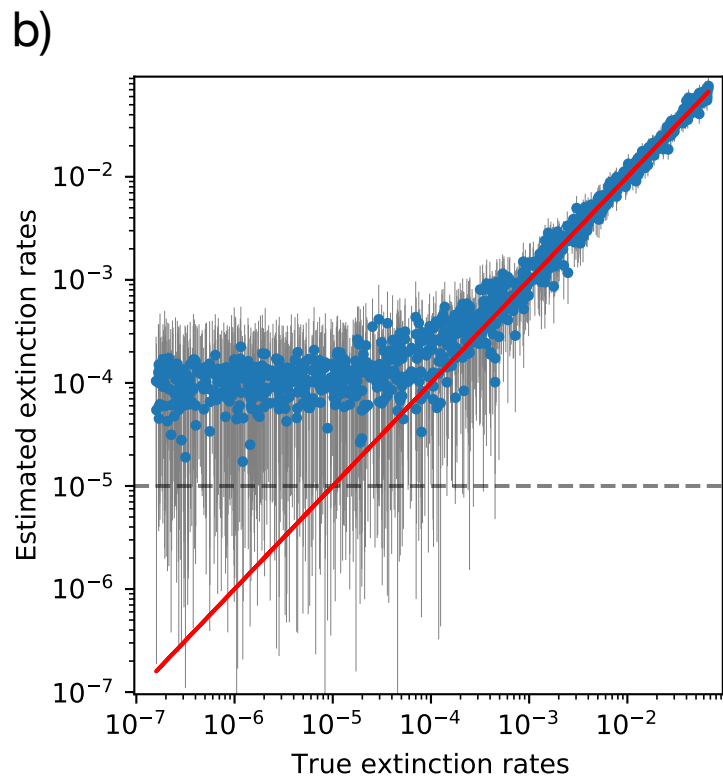
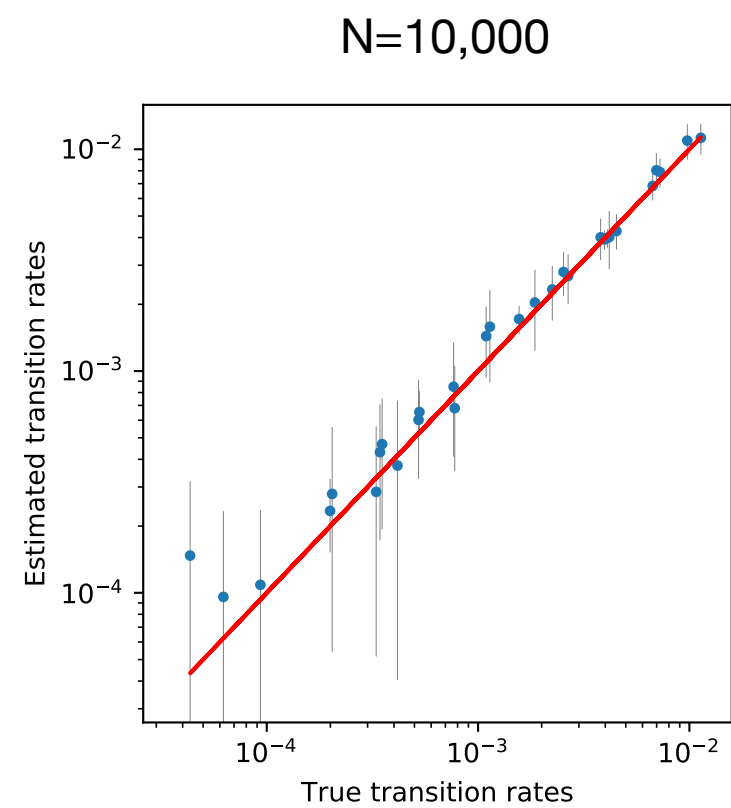
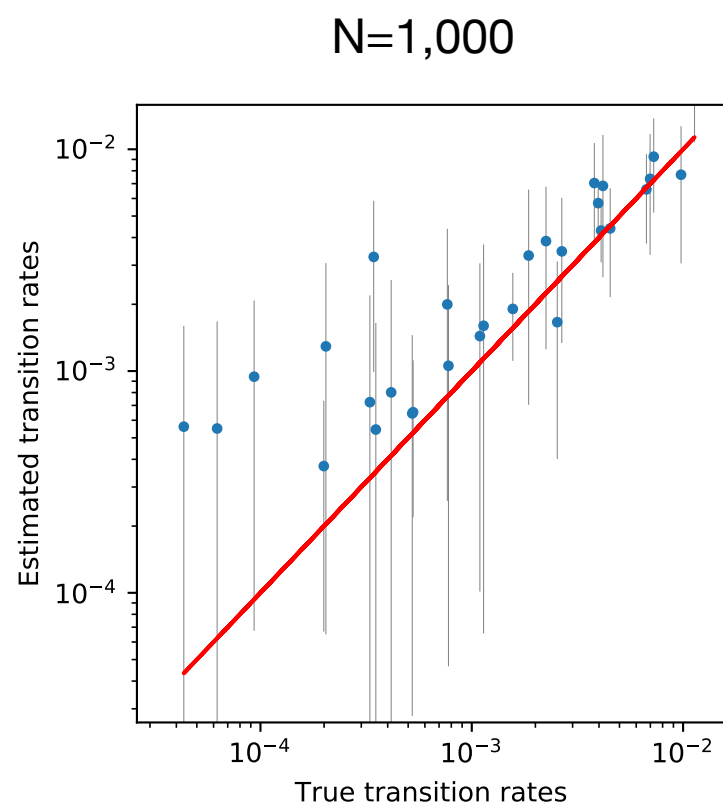
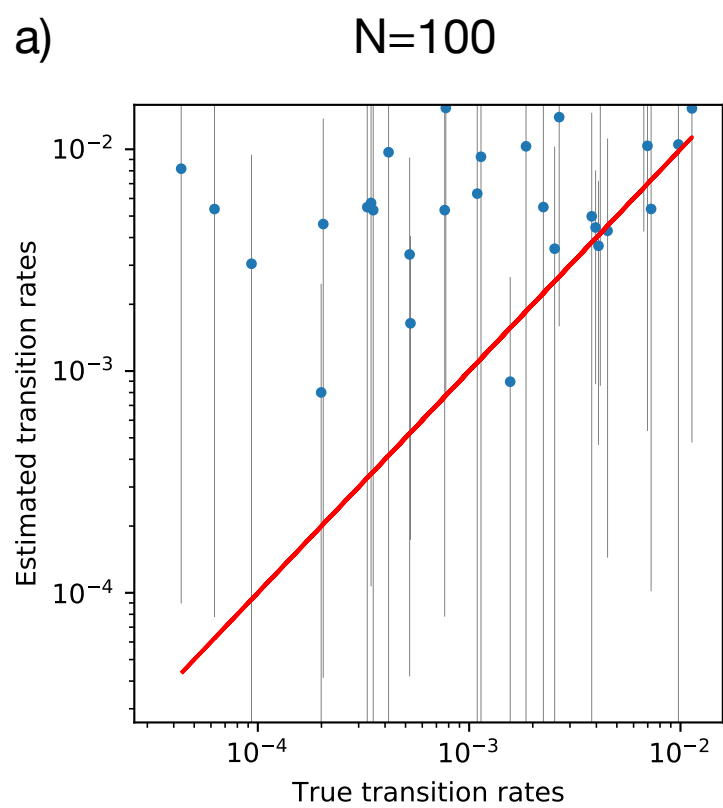
CritE EX mode

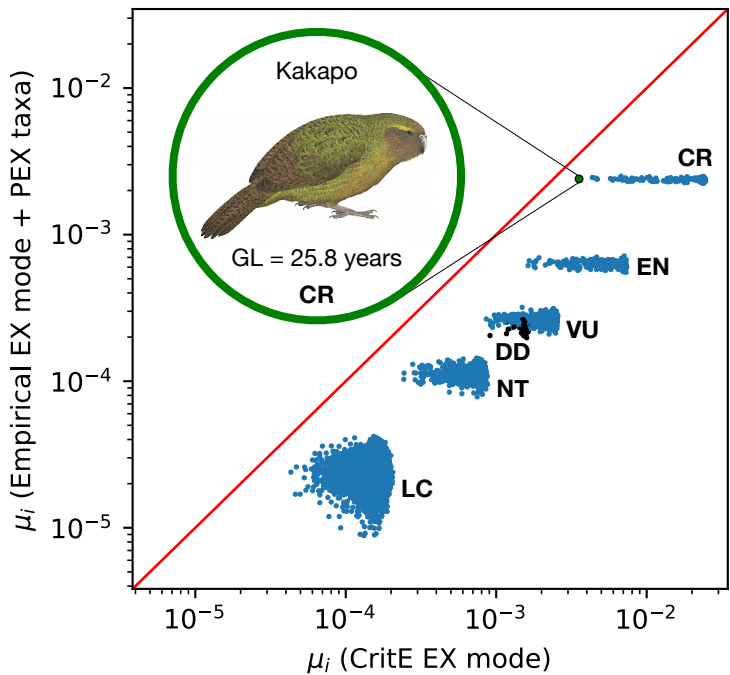


Empirical EX mode

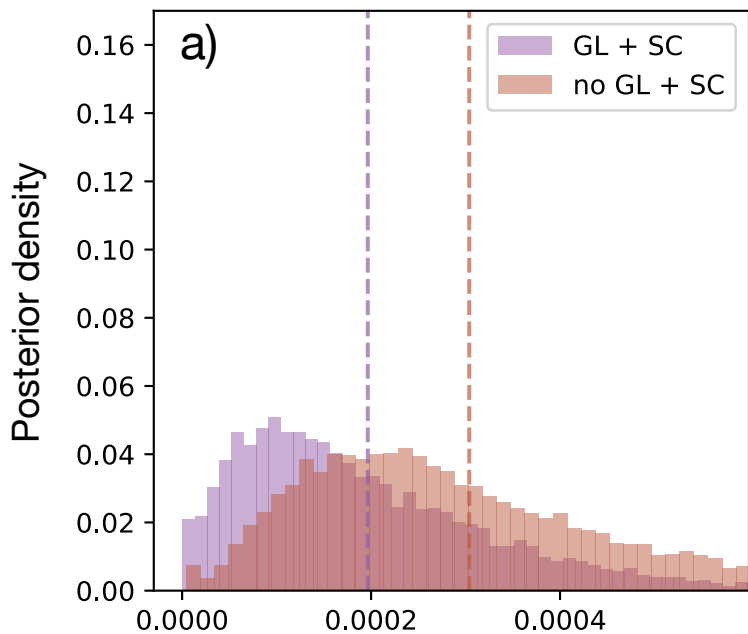


**Empirical EX mode
+ PEX taxa**

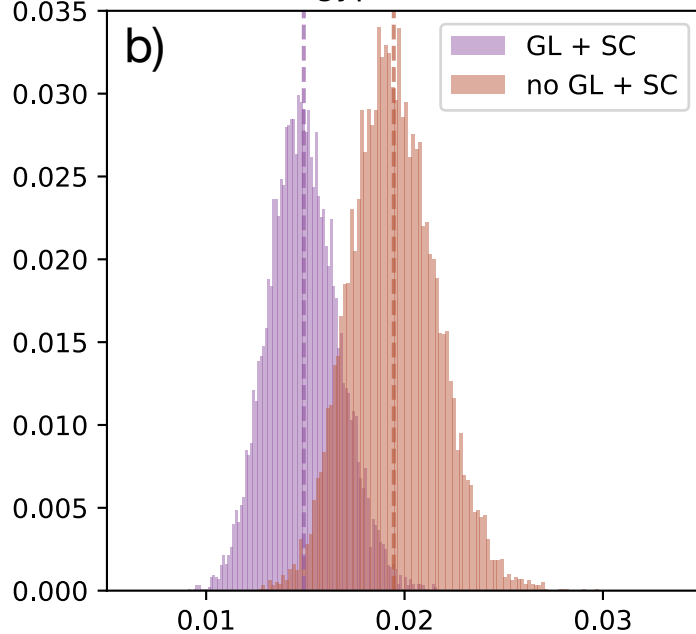




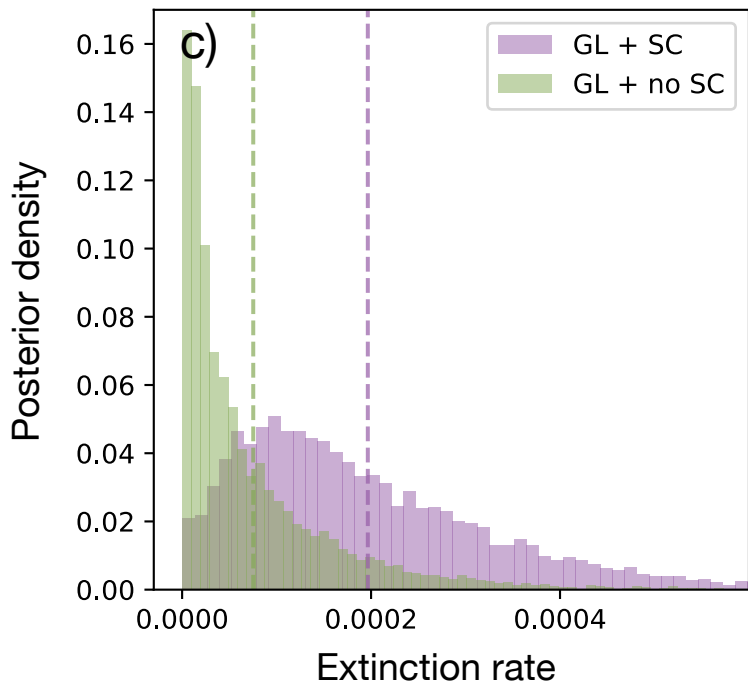
Cathartes aura (LC)



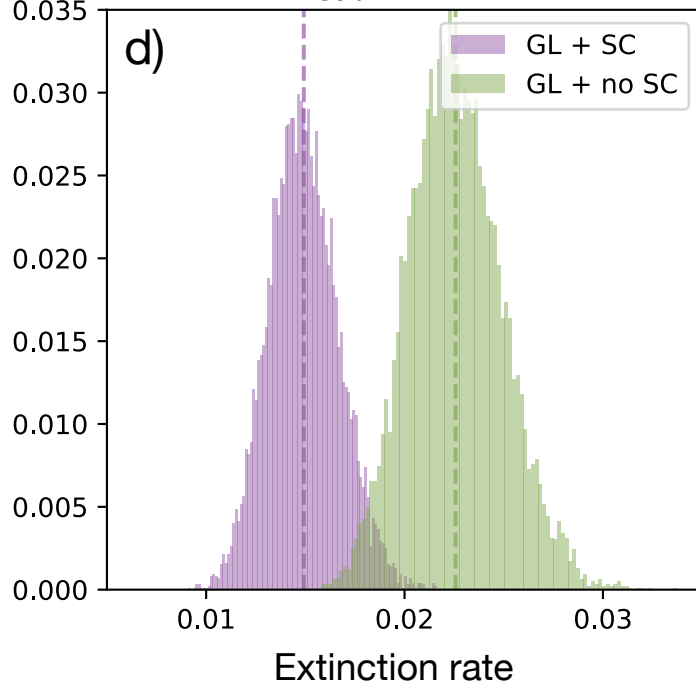
Sarcogyps calvus (CR)



Cathartes aura (LC)



Sarcogyps calvus (CR)



Simulated bird extinctions in next 100 years

