

1 Environmental unpredictability shapes glucocorticoid regulation across populations of tree swallows

2

3 Cedric Zimmer¹, Conor C. Taff^{1,2}, Daniel R. Ardia³, Alexandra P. Rose⁴, David A. Aborn⁵, L. Scott Johnson⁶

4 and Maren N. Vitousek^{1,2}

5

6 ¹ Department of Ecology and Evolutionary Biology, Cornell University, Ithaca, NY 14853, USA

7 ² Cornell Lab of Ornithology, Ithaca, NY 14850, USA

8 ³ Department of Biology, Franklin and Marshall College, Lancaster, PA 17604, USA

9 ⁴ Institute of Arctic and Alpine Research, University of Colorado, Boulder, CO 80303, USA

10 ⁵ Biology, Geology and Environmental Science, The University of Tennessee Chattanooga, Chattanooga,
11 TN 37403, USA

12 ⁶ Department of Biological Sciences, Towson University, Towson, MD 21252, USA

13

14 Correspondence: Cedric Zimmer, cgg.zimmer@gmail.com, 813-974-4105

15 **Abstract**

16 Responding appropriately to challenges is an important contributor to fitness. Variation in the regulation
17 of glucocorticoid hormones, which mediate the phenotypic response to challenges, can therefore
18 influence the ability to persist in a given environment. We compared stress responsiveness in four
19 populations of tree swallows (*Tachycineta bicolor*) along an environmental gradient to evaluate support
20 for different selective pressures in driving the evolution of glucocorticoid regulation. In accordance with
21 the environmental unpredictability hypothesis, stronger stress responses were seen in more
22 unpredictable environments. Contrary to the reproductive value hypothesis, the stress response was not
23 lower in populations engaging in more valuable reproductive attempts. Populations with stronger stress
24 responses also had stronger negative feedback, which supports a “mitigating” rather than a
25 “magnifying” effect of negative feedback on stress responses. These results suggest that combining a
26 robust stress response with strong negative feedback may be important for persisting in unpredictable
27 or rapidly changing environments.

28

29 **Introduction**

30 Global environmental changes are altering the habitats of many species (Peñuelas et al., 2013). For
31 species with wide geographic ranges, intra-specific variation in life-history strategies resulting from
32 historical selection might predispose some populations to be more or less susceptible to increasing
33 environmental changes (Candolin, 2019; Debecker and Stoks, 2019; Harding et al., 2019). Environmental
34 variation favors individuals that differentially allocate time and energy to reproduction and self-
35 maintenance in order to maximize lifetime fitness (Stearns, 1989; Stearns, 1992; Ricklefs and Wikelski,
36 2002). Thus, characterizing differences in the regulation of this trade-off across environments is critical
37 for understanding the mechanisms that have shaped phenotypic responses and that may allow
38 successful adaptation and population persistence under rapid global changes. While the ultimate
39 reasons for variation in life-history traits across environments and latitudes have been well studied, we
40 have a limited understanding of the proximate mechanisms that underlie this variation (Eikenaar et al.
41 2012, Atwell et al. 2014). Because environmental factors may influence the evolution of life-history
42 traits by acting on physiological systems that integrate external conditions, hormones have been
43 proposed to play a crucial role as mediators of life-history trade-offs (Zera and Harshman, 2001; Ricklefs
44 and Wikelski, 2002; Eikenaar et al., 2012).

45

46 The hypothalamic-pituitary-adrenal (HPA) axis is a fundamental component of the endocrine system
47 that forms an interface between an animal and its environment (Wingfield and Sapolsky, 2003; Breuner,
48 2011; Hau et al., 2016). The HPA axis coordinates the response to energetic and other challenges mainly
49 by regulating the production and release of glucocorticoids (Wingfield et al., 1998; Sapolsky et al., 2000).
50 In non-chronically or acutely stressed individuals, glucocorticoids are usually maintained at low levels to
51 regulate energy balance and mediate foraging and other locomotor activities (Sapolsky et al., 2000;
52 Landys et al., 2006). When facing unpredictable challenges, circulating glucocorticoids increase
53 dramatically, promoting a suite of processes that facilitate responding to and recovering from these
54 challenges (Wingfield et al., 1998; Sapolsky et al., 2000). When sustained, this stress response can
55 trigger an emergency life-history stage in which breeding activities are usually reduced, and energy is
56 redirected toward survival (Wingfield et al., 1998). Thus, glucocorticoids – particularly in the presence of
57 a stressor – have been widely predicted to mediate life history trade-offs between current and future
58 reproduction.

59
60 While models of the stress response make clear predictions about redirecting effort, empirical studies
61 have proved equivocal in linking environmental conditions to appropriate responses. One limitation of
62 much empirical work to date is that it has focused on the relationship between glucocorticoids and
63 fitness in a single context (i.e., a single population, year, or environment; Hau et al., 2016; Guindre-
64 Parker, 2018). The factors that shape these relationships are best understood by measuring
65 glucocorticoids across different environmental and life-history contexts (Apfelbeck et al., 2017; Schoenle
66 et al., 2018; Vitousek et al., 2019a). Because the costs and benefits of mounting a robust glucocorticoid
67 stress response are likely to differ across environments and species, populations are predicted to differ
68 in how they regulate glucocorticoids (Schoenle et al., 2018; Vitousek et al., 2019b). For instance, a
69 comparative study in birds found a positive association between latitude and the glucocorticoid
70 response to a standardized restraint stressor (acute challenge that activates the HPA axis resulting in
71 stress-induced glucocorticoids increase) during breeding (Bókony et al., 2009). However, willow warblers
72 (*Phylloscopus trochilus*) breeding in northern Sweden—where the breeding season is shorter and
73 environmental conditions less predictable—show lower stress-induced glucocorticoid levels than those
74 breeding in southern Sweden, where conditions are more predictable (Silverin et al., 1997). Breuner and
75 colleagues (2003) compared males from three populations of white-crowned sparrows (*Zonotrichia*
76 *leucophrys*) breeding at different latitudes from California to Alaska. They found that males had similar
77 stress-induced glucocorticoid levels, but differed in their corticosteroid-binding globulin and intracellular

78 receptor affinity, suggesting that HPA axis regulation varied across the populations. Thus, while the
79 stress response appears to vary with latitude, additional research is needed to determine which of the
80 many potential selective pressures that covary with latitude are driving this variation.

81

82 Two factors thought to be particularly important in shaping the costs and benefits of stress
83 responsiveness are the frequency with which organisms face major unpredictable changes in their
84 environment (the “environmental unpredictability” hypothesis), and how valuable each breeding
85 attempt is to lifetime reproductive effort (the “brood value” or “reproductive value” hypothesis). The
86 “environmental unpredictability” hypothesis predicts that species or populations in less predictable
87 environments will have higher stress-induced glucocorticoid levels to facilitate responding effectively to
88 challenges (Wingfield, 2013; Vitousek et al., 2019a). The reproductive value hypothesis predicts that the
89 stress response differs based on the proportion of lifetime reproductive effort represented by a single
90 breeding attempt. According to this hypothesis, because high glucocorticoid levels can be deleterious to
91 reproduction by diverting energy away from breeding activities, organisms engaging in more valuable
92 reproductive attempts should have lower stress-induced glucocorticoids to avoid jeopardizing the
93 current breeding attempt (Bókony et al., 2009). The relative value of a single reproductive event
94 depends on both the length of the breeding season, which influences the number of potential
95 reproductive attempts an individual can engage in per season, and its lifespan. A shorter breeding
96 season results in less opportunity to engage in multiple successive reproductive attempts or to restart a
97 breeding attempt in case of failure of the initial one, which increases the relative value of each
98 reproductive attempt (Breuner et al., 2003; Wingfield and Sapolsky, 2003; Hau et al., 2016). Therefore,
99 the reproductive value hypothesis predicts that populations living in environments with shorter
100 breeding seasons will show lower stress-induced glucocorticoid levels during breeding (Bókony et al.,
101 2009; Breuner, 2011). This desensitization of the HPA axis to challenges facilitates not interrupting
102 valuable breeding attempts (Wingfield and Sapolsky, 2003; Breuner, 2011).

103

104 Previous comparative studies investigating the role of environmental predictability in shaping
105 glucocorticoid concentrations have predominantly used latitude as a proxy for environmental
106 predictability or harshness (e.g., Bókony et al., 2009; Jessop et al., 2013). Variability in environmental
107 conditions, especially temperature, typically increases with increasing latitude and elevation (Breuner
108 2011); however, these patterns are not always consistent. Furthermore, as many factors (including the
109 duration of the breeding season) also commonly covary with latitude, analyses that use latitude alone

110 could conflate the selective pressures imposed by living in an unpredictable environment with other
111 selective pressures. We are aware of only one previous study that has directly quantified and tested the
112 role of environmental variation in predicting variation in stress-induced glucocorticoids. A phylogenetic
113 comparative analysis of the relative support for these and other factors in shaping glucocorticoid
114 variation across vertebrates found that reproductive value better predicted large-scale variation in
115 stress-induced glucocorticoids than environmental variability (Vitousek et al., 2019a). However, the
116 relative roles of different selective pressures in shaping glucocorticoid variation are expected to vary
117 across species. Disentangling the relative roles of reproductive value and environmental unpredictability
118 in shaping glucocorticoid regulation will require comparing populations of the same species inhabiting
119 different environments.

120

121 Central to the hormonal mediation of life history is an understanding of the potential cost of the stress
122 response. The costs of the stress response likely depend not only on maximum glucocorticoid levels but
123 also on the duration of exposure to high levels, which is influenced by the strength of negative feedback.
124 Negative feedback is triggered after activation of the HPA axis and is coordinated by glucocorticoids
125 binding to receptors in the brain inducing a decrease in circulating glucocorticoids (de Kloet et al., 1998;
126 Breuner and Orchinik, 2001; Romero, 2004). Despite increasing evidence that differences in the strength
127 of negative feedback affect aspects of health and performance (Sapolsky, 1983; Dallman et al., 1992;
128 Weaver et al., 2004), its functional effects have been largely neglected in free-living organisms (but see
129 (Romero and Wikelski, 2010; Zimmer et al., 2019). We are not aware of any previous studies that have
130 assessed how negative feedback varies across environments. We hypothesized that differences in
131 negative feedback efficacy could either mitigate or magnify the costs of a stress response. Strong
132 negative feedback could serve to mitigate the costs of mounting a robust stress response by inducing a
133 fast decrease in circulating glucocorticoids. This could be particularly important when a strong stress
134 response is required to cope effectively with frequent unpredictable short-term challenges. In this case,
135 strong negative feedback may make it possible to avoid the negative effects of sustained glucocorticoid
136 elevation and hence recover quickly and resume critical activities such as breeding. A recent study
137 within a single population of tree swallows (*Tachycineta bicolor*) supported the mitigating hypothesis:
138 incubating females that exhibited both a robust stress response and strong negative feedback were less
139 likely to abandon reproductive attempts when facing stressors (Zimmer et al., 2019). Alternatively,
140 coupling an elevated stress response with weak negative feedback could serve to magnify the effects of
141 the stress response. This could be adaptive if mounting a longer stress response facilitates avoiding or

142 alleviating severe challenges; for example, if a more robust response enhances sensitivity to
143 environmental cues (Wingfield, 2013; Vitousek et al., 2019b). Accordingly, individuals breeding in highly
144 variable environments are predicted to show elevated stress responses followed by strong negative
145 feedback.

146

147 Here, we compared the support for two sets of predictions about how HPA axis regulation differs across
148 populations and environments in breeding tree swallows. First, we asked whether variation in the
149 glucocorticoid stress response (peak stress-induced glucocorticoids) across populations is better
150 predicted by the reproductive value hypothesis or the environmental unpredictability hypothesis.
151 Second, we assessed whether negative feedback varies across populations, and if so, whether the
152 patterns suggested a magnifying or mitigating effect on the stress response. The tree swallow, a
153 common passerine bird that breeds across much of North America, is an ideal species in which to test
154 these predictions as it breeds along an expansive latitudinal as well as elevational gradient within the
155 temperate zone. As such, different populations face different amounts of time available for
156 reproduction as well as differing levels of environmental unpredictability. We compared populations
157 breeding in Tennessee, New York, Wyoming and Alaska. We assessed females' HPA axis activity by
158 measuring baseline glucocorticoids, stress responses, and negative feedback during two life history
159 stages: incubation and nestling rearing. In order to test whether between-population differences in the
160 stress response are better predicted by differences in the time available for reproduction or
161 environmental unpredictability, we characterized both parameters in each population. We first
162 determined breeding synchrony and the length of the breeding season in each population. Breeding
163 season length is a good proxy for differences in reproductive value across populations breeding at
164 different latitudes (Breuner, 2011). Because increasing reproductive value is generally associated with
165 higher parental investment (Williams, 1966; Ardia, 2005; Ardia and Clotfelter, 2007) and thus potentially
166 with higher reproductive success, we also determined breeding effort and success. To characterize
167 differences in environmental predictability, we calculated an index of unpredictability for different
168 weather variables using historical weather data for each site. We predicted that if variation in the
169 magnitude of the glucocorticoid stress response is predominantly shaped by reproductive value, females
170 breeding at the two sites with a relatively short breeding season (Alaska and Wyoming) would mount a
171 lower acute stress response (Bókony et al., 2009; Vitousek et al., 2019a). Conversely, if environmental
172 unpredictability plays a greater role, we expected the opposite pattern: females breeding in the more
173 unpredictable environments of Alaska and Wyoming would have higher stress-induced corticosterone

174 levels. Concerning negative feedback, the mitigation hypothesis predicts strong negative feedback in
175 populations with greater stress responses. The magnifying hypothesis, in contrast, predicts weaker
176 feedback in the face of greater stress responses.

177

178 Differences in HPA axis regulation across populations could also be associated with downstream
179 physiological costs. Glucocorticoids affect both redox balance and glucose metabolism (Sapolsky et al.,
180 2000; Costantini et al., 2011). We also tested whether oxidative stress, another potential physiological
181 mediator of the trade-off between reproduction and survival (Casagrande and Hau, 2018), or glucose
182 levels varied across populations. We predicted that oxidative stress and circulating glucose would be
183 higher in populations breeding in more unpredictable environments and/or with higher reproductive
184 value, and in those with higher baseline glucocorticoid levels.

185

186 **Results**

187 *Environmental unpredictability*

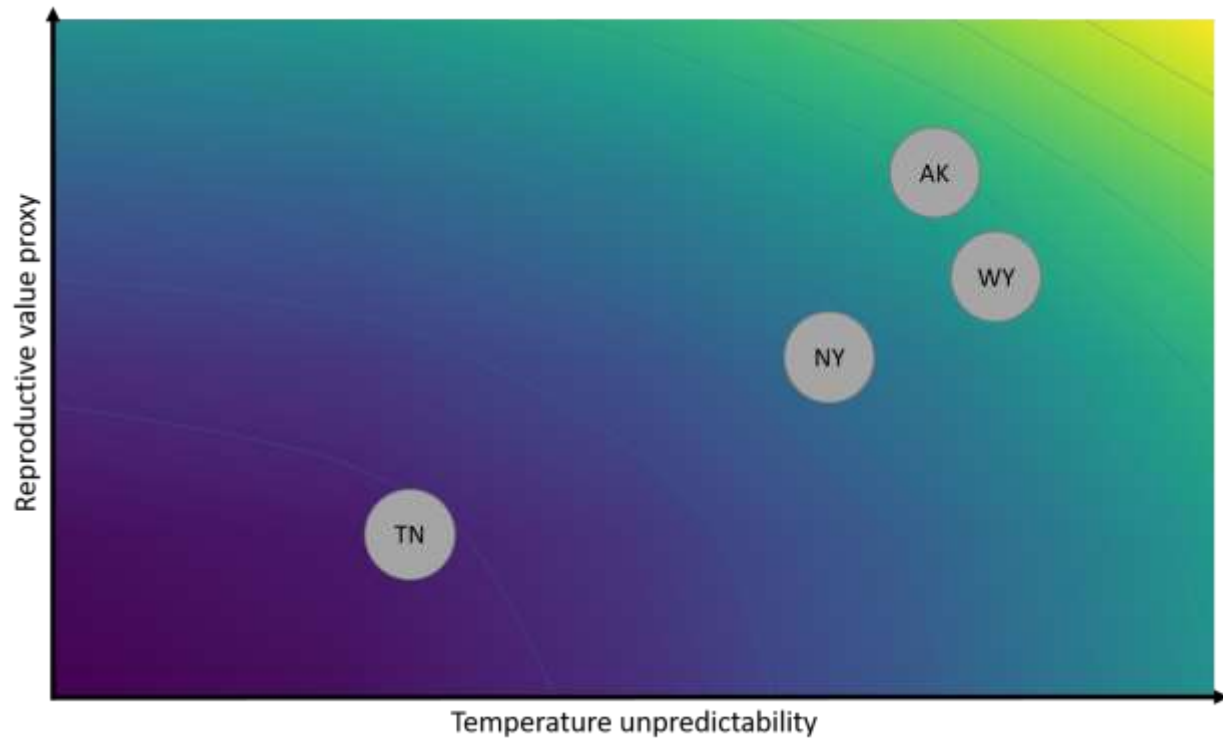
188 The unpredictability of temperature during the breeding season was the lowest in Tennessee and the
189 highest in Wyoming and Alaska (Table 1, Fig 1). Temperature unpredictability in New York was
190 intermediate, but closer to Alaska and Wyoming than to Tennessee (Table 1).

191

192 Table 1: Unpredictability of daily average temperature, daily average active time (between 0600 and 2000)
193 temperature, daily maximum temperature and daily total precipitation at the field sites in Tennessee, New York,
194 Wyoming and Alaska. Higher SD_{res} value indicates greater unpredictability.

	Tennessee	New York	Wyoming	Alaska
SD_{res} average daily temperature	0.18	0.33	0.41	0.38
SD_{res} average daytime temperature	0.21	0.42	0.53	0.50
SD_{res} maximum daily temperature	0.21	0.38	0.51	0.47

195



196

197 Figure 1: Plot illustrating the difference between the four populations (Tennessee (TN), New York (NY), Wyoming
198 (WY) and Alaska (AK) along a gradient of increasing temperature unpredictability (x axis) and a proxy of
199 reproductive value based on breeding season length (y axis). Warmer color indicates higher environmental
200 unpredictability and higher reproductive value.

201

202 *Reproductive value*

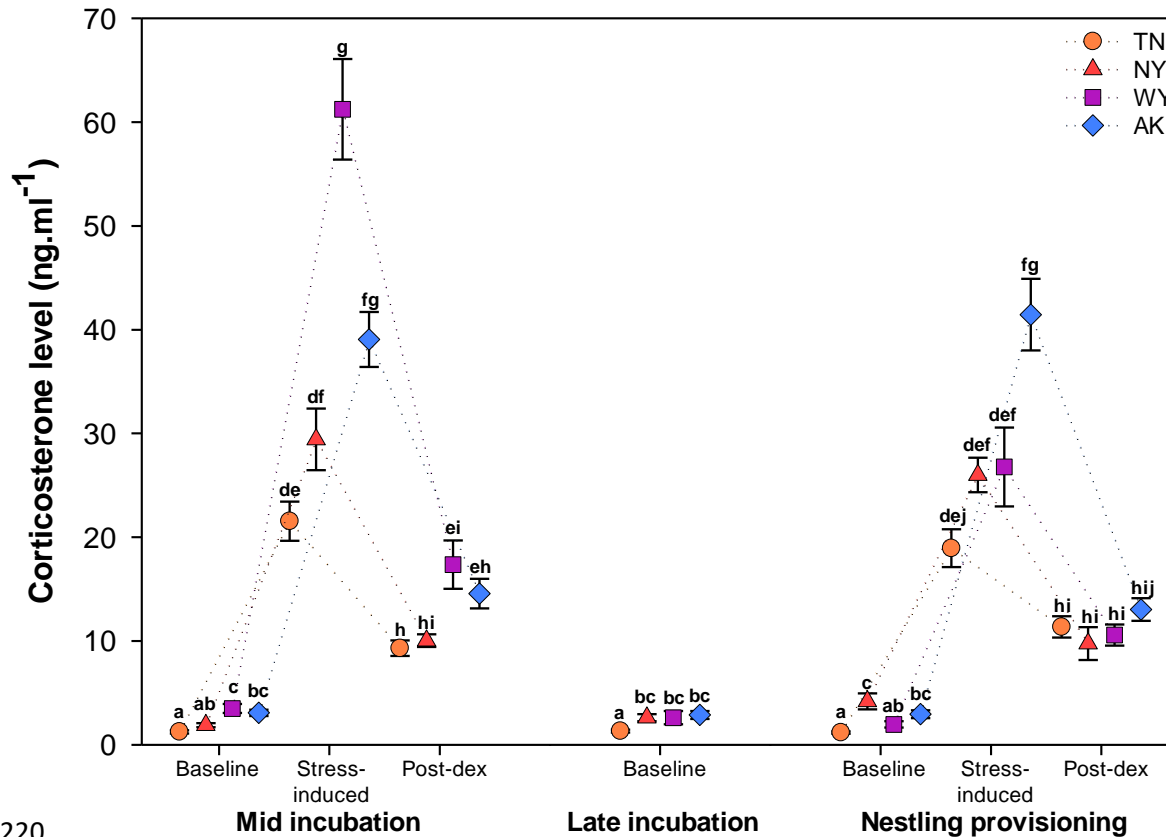
203 The synchrony with which females laid first clutches differed between populations ($\chi^2_{3,250} = 89.51$, $p <$
204 0.0001) with females in TN (17.7 ± 1 days) showing a more extended clutch initiation period than
205 females in the three other populations (NY: 8.9 ± 1.1 days, WY: 9.2 ± 0.7 days, AK: 8.8 ± 0.6 days; $z \geq$
206 6.68 , $p < 0.0001$). Total length of the breeding season also differed between populations ($\chi^2_3 = 8.49$, $p =$
207 0.037) with a longer breeding season in TN (99 days) and the shortest in AK (66 days). Breeding season
208 lengths in NY (74 days) and WY (70 days) were intermediate, but more similar to AK than TN. Overall,
209 these patterns suggest that reproductive value was the lowest in TN and relatively similar across the
210 other three populations (Fig 1).

211

212 *Corticosterone regulation*

213 Females' corticosterone phenotypes differed between populations. This difference was influenced by
214 the life history substage during which the female was captured and corticosterone sample type

215 (baseline, stress-induced, post-dexamthasone (post-dex); population x life history substage x sample:
 216 $F_{12,1263} = 3.92, p < 0.0001$; Fig 2).
 217
 218 Across sampling periods, baseline corticosterone levels tended to be lower in Tennessee than in the
 219 other three populations (Fig. 2, see supplementary material for statistical analyses).



220
 221 Figure 2: Corticosterone levels (mean \pm SE) of females when captured at three points during the reproductive
 222 cycle (mid incubation: day 6-7 of incubation, late incubation: day 12-13 of incubation, and nestling provisioning:
 223 day 6 post-hatch) in Tennessee (TN), New York (NY), Wyoming (WY), and Alaska (AK). At all captures baseline
 224 corticosterone levels were measured. During the first and third captures, stress-induced corticosterone after 30
 225 minutes of restraint and 30 minutes after injection of dexamethasone (post-dex) were also measured. Different
 226 letters indicate significant differences.

227
 228 Circulating stress-induced corticosterone levels generally increased with temperature unpredictability
 229 (Fig 2). The exception to this pattern was in WY during mid incubation, when stress-induced
 230 corticosterone levels were significantly higher than all other populations except for AK (see

231 supplementary material for statistical analyses). Corticosterone stress responses (the difference
232 between stress-induced and baseline corticosterone) showed a similar pattern. Overall, stress responses
233 differed between populations and life history substages (population x life history substage: $F_{3,220.1} = 9.45$,
234 $p < 0.0001$; Fig 3a). During mid-incubation, stress responses were highest in WY ($t \geq 5.08$, $p < 0.0001$; Fig
235 3a) and intermediate in AK ($t \geq 2.71$, $p \leq 0.044$; Fig 2a). During nestling provisioning, females in AK had a
236 stronger stress response than females in all other populations ($t \geq 3.01$, $p \leq 0.048$; Fig 3a).

237
238 Circulating corticosterone levels after dexamethasone injection did not differ between the four
239 populations ($t \leq 1.73$, $p \geq 0.99$; Fig 2). However, negative feedback, i.e., the decrease in corticosterone
240 following dexamethasone injections, differed between populations and life history substages
241 (population x life history substage: $F_{3,236.3} = 8.30$, $p < 0.0001$; Fig 3b). During mid-incubation, negative
242 feedback was stronger in WY than in all other populations ($t \geq 5.36$, $p < 0.0001$; Fig 3b). In the three
243 other populations, negative feedback was highest in AK, intermediate in NY, and weakest in TN (pairwise
244 comparisons: AK vs. TN: $t = 3.60$, $p = 0.016$, AK vs. NY: $t = 0.69$, $p = 0.99$, TN vs. NY: $t = 1.55$, $p = 0.77$; Fig
245 3b). During the nestling provisioning period, females in AK had stronger negative feedback than females
246 in all other populations ($t \geq 3.16$, $p \leq 0.04$; Fig 3b); negative feedback did not differ among the other
247 populations ($t \leq 1.24$, $p \geq 0.91$; Fig 3b). Overall, negative feedback was stronger in populations that
248 mounted a strong stress response.

249
250 Within populations, the magnitude of the stress response did not change between substages of the
251 reproductive period ($t \leq 1.49$, $p \geq 0.81$; Fig 3a), except in WY where it decreased between incubation and
252 nestling provisioning ($t = 7.09$, $p < 0.0001$; Fig 3a). Negative feedback efficacy also did not change
253 between life history substages within populations ($t \leq 1.28$, $p \geq 0.90$; Fig 3b), except in WY where it
254 decreased between mid-incubation and nestling provisioning ($t = 6.08$, $p < 0.0001$; Fig 3b). See
255 supplementary material for detailed information about within population corticosterone changes across
256 the breeding season.

257
258 Correlations between the different measures of the HPA axis differed between populations and life
259 history substages. Overall, the different aspects of the HPA axis are not correlated or weakly to
260 moderately correlated (Table 2). However, the strength of the stress response and the efficacy of
261 negative feedback are positively correlated in all populations at both life history substages (Table 2).

262

263 Table 2: Correlations between baseline, stress-induced and post-dexamethasone (post-dex) glucorticoid levels and
 264 between the strength of the stress response and the efficacy of negative feedback within each population. Data
 265 are R², p-value. Bold indicates significant correlations.

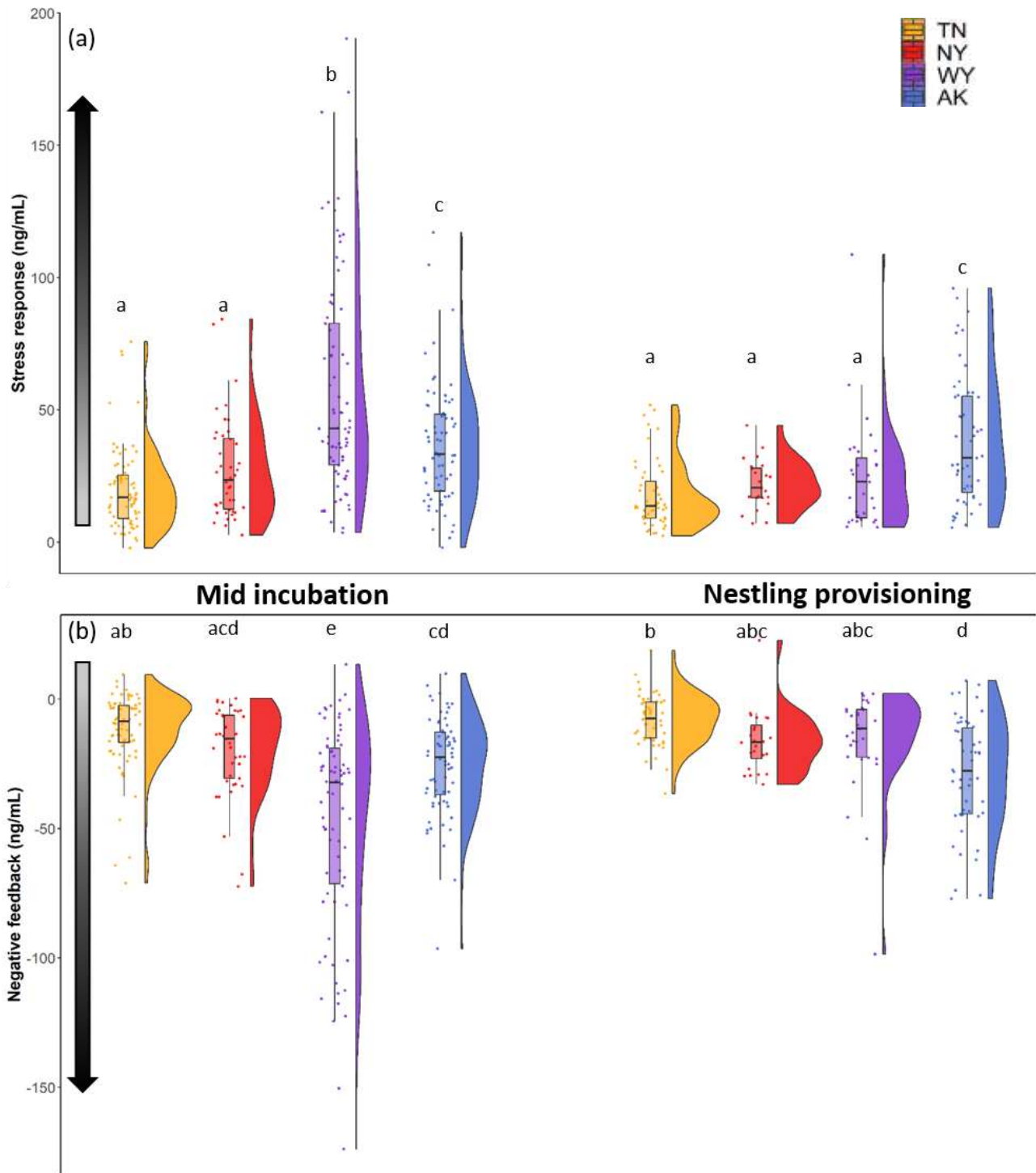
		Mid incubation			Nestling provisioning		
		stress- induced (R ² , p-value)	post-dex (R ² , p-value)	negative feedback (R ² , p-value)	stress- induced (R ² , p-value)	post-dex (R ² , p-value)	negative feedback (R ² , p-value)
Tennessee	baseline	0.001, 0.94	0.005, 0.54		0.09, 0.03	0.18, 0.002	
	stress-induced		0.08, 0.02			0.27, 0.0001	
	stress response			0.84, <0.0001			0.70, <0.0001
New York	baseline	0.02, 0.43	0.02, 0.41		0.001, 0.89	0.03, 0.43	
	stress-induced		0.15, 0.01			0.01, 0.62	
	stress response			0.96, <0.0001			0.50, 0.0008
Wyoming	baseline	0.07, 0.02	0.005, 0.55		0.11, 0.07	0.01, 0.60	
	stress-induced		0.001, 0.91			0.08, 0.12	
	stress response			0.91, <0.0001			0.92, <0.0001
Alaska	baseline	0.001, 0.78	0.003, 0.66		0.01, 0.43	0.007, 0.54	
	stress-induced		0.03, 0.15			0.25, 0.0001	
	stress response			0.69, <0.0001			0.90, <0.0001

266

267

268 Resistance to oxidative stress and reactive oxygen metabolite levels did not differ between populations
 269 (Supplementary material). Baseline and stress-induced glucose did not differ between females in TN, NY
 270 and AK, but baseline glucose was lower in WY than in the other populations (Supplementary material).

271



272

273 Figure 3: Females' glucocorticoid stress responses (a) and the strength of negative feedback (b) in Tennessee
274 (orange), New York (red), Wyoming (purple) and Alaska (blue) during mid incubation and nestling provisioning.

275 Points overlaying the boxplot show raw data and the half-split violin shows the probability density function.

276 Shaded arrows along the y-axis denote increasing strength of the stress response and negative feedback with

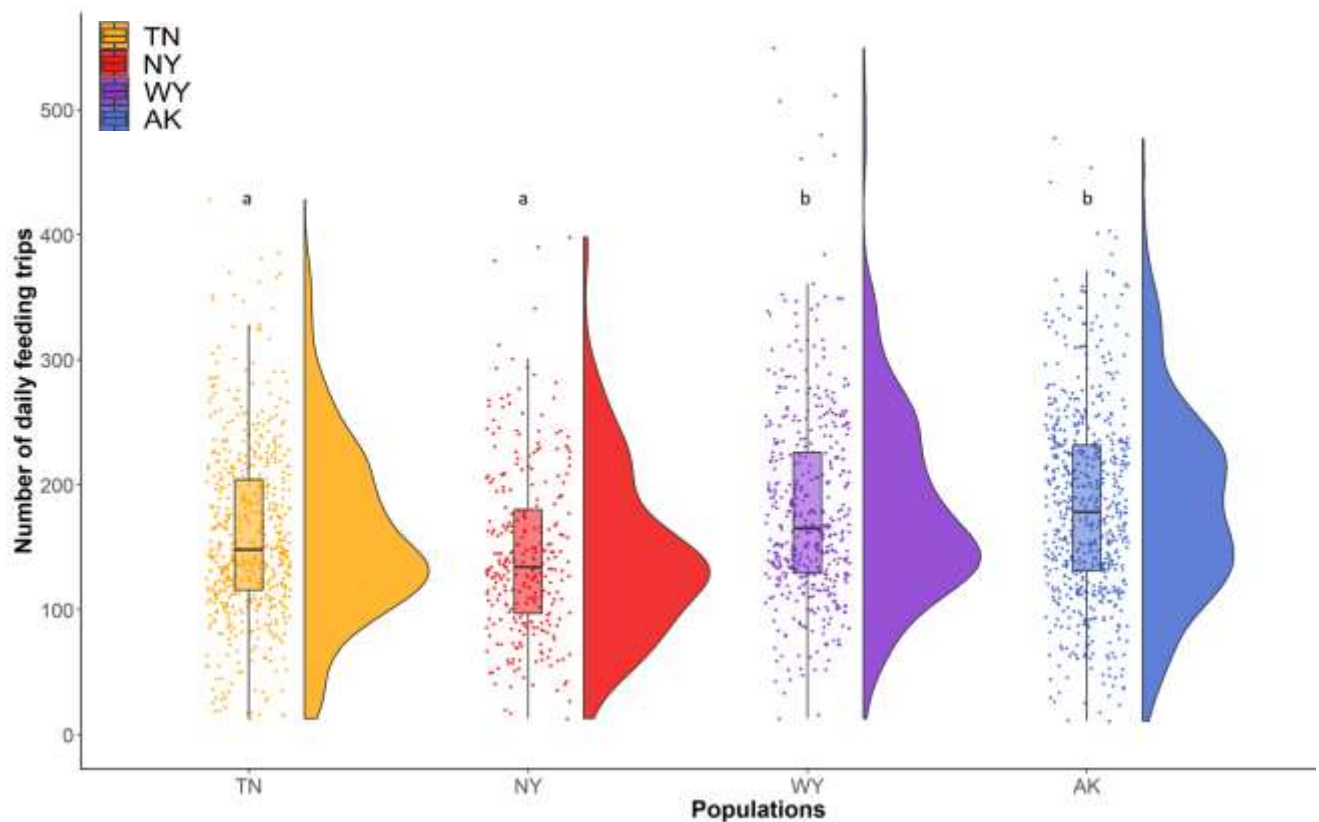
277 increasing darkness. Different letters indicate significant differences.

278

279 *Reproductive effort and success*

280 There were no differences in clutch size or brood size at hatching between populations ($\chi^2 \leq 5.90$, $p \geq$
281 0.12, see supplementary material). Because of an extended period of cold, wet weather that occurred
282 during the incubation stages of most females, hatching success ($\chi^2_{3,250} = 29.24$, $p < 0.0001$) was lower in
283 WY (50.7 %, 38 of 75 nests) than in the other three populations (TN: 82.2 %, 60 of 73 nets; NY: 79.5 %,
284 31 of 39 nests; AK: 84.4 %, 54 of 64 nests; $z \geq 3.58$, $p \leq 0.002$). Provisioning effort differed across
285 populations ($F_{3,155.4} = 6.00$, $p = 0.0007$). Females in AK and WY provisioned at higher rates than those in
286 NY and TN (Fig 4).

287



288

289 Figure 4: Number of daily feeding trips for females in Tennessee (orange), New York (red), Wyoming (purple), and
290 Alaska (blue), calculated from the total number of raw visits recorded (329,047). Points overlaying the boxplot
291 show raw data and the half-split violin shows the probability density function. Different letters indicate significant
292 differences between populations.

293

294 Nestling body mass differed across populations, mirroring the observed differences in female
295 provisioning rates ($F_{3,128.5} = 4.02$, $p = 0.009$). On day 12 post-hatching, nestlings were heaviest in AK (21.2

296 ± 0.2 g) and WY (20.6 ± 0.2 g), and lightest in NY (18.7 ± 0.5 g) and TN (19.6 ± 0.2 g). Among nests in
297 which one or more eggs hatched, the number of nests in which one or more nestlings fledged also
298 differed among populations ($\chi^2_{3,250} = 22.49$, $p < 0.0001$); fledging success was lower in NY (35.5 %, 11 of
299 31 nests) than in the other populations (TN: 76.7 %, 46 of 60 nests, AK: 80.4 %, 41 of 51 nests, WY: 80.0
300 %, 28 of 35 nests; $t \geq 3.30$, $p \leq 0.005$).

301

302 **Discussion**

303 In the current context of global changes, understanding how organisms have evolved in response to
304 environmental challenges is critical. Our results demonstrate that glucocorticoid regulation across
305 different environmental and life-history contexts may underlie successful adaptation and population
306 persistence. First, we showed that female tree swallows breeding in environments with higher
307 temperature unpredictability secrete more glucocorticoids in response to acute challenges during
308 breeding. This pattern was strong even though at these sites birds also experienced shorter breeding
309 seasons and would normally be predicted to skew investment towards offspring. Thus, if elevated
310 glucocorticoids triggered reproductive failure or delay, birds at those sites would have less time to
311 complete reproduction and re-nest if necessary. This seemingly contrary result reveals environmental
312 unpredictability may be a stronger force than the value of the current breeding attempt in determining
313 stress-induced glucocorticoid levels in female tree swallows. Second, we found that populations with
314 higher stress responses also have stronger negative feedback, which likely functions to limit potential
315 damage caused by high glucocorticoid levels. The combination of high stress-induced glucocorticoid
316 levels followed by the induction of strong negative feedback may allow females to cope effectively with
317 frequent unpredictable challenges, and to recover faster in order to continue breeding activities. This
318 phenotype may be most strongly favored in environments with both high unpredictability and greater
319 time constraints on reproduction.

320

321 *Life history and environmental unpredictability*

322 Across the four study populations we found differences in key selective pressures. Birds breeding in
323 Alaska and in the mountains of Wyoming faced more unpredictable temperatures than birds in
324 Tennessee; New York conditions were intermediate. The time constraints on reproduction differed
325 similarly across populations, resulting in higher apparent reproductive value in Alaska and Wyoming and
326 lowest apparent reproductive value in Tennessee. As predicted by life history theory (and consistent
327 with previous comparative work in tree swallows: Ardia, 2006; Ardia, 2007; Rose and Lyon, 2013; Akçay

328 et al., 2016), birds in Alaska and Wyoming showed higher parental investment (measured as offspring
329 feeding rate) and reared larger offspring than those in New York and Tennessee.

330

331 *Glucocorticoid regulation*

332 At our two sites with stronger time constraints on reproduction and high temperature unpredictability
333 (AK and WY) we found that the magnitude of the glucocorticoid stress response was greatest. This result
334 provides the first empirical support for the idea that environmental unpredictability could be a stronger
335 force shaping the peak glucocorticoid response to challenges than brood value, which should select for
336 suppressed stress responses in the face of low residual reproductive value. In contrast with this finding,
337 a recent large scale phylogenetic comparative analysis found greater support for reproductive value
338 shaping glucocorticoid stress responses across vertebrates than environmental variation (Vitousek et al.
339 2019b). Although other studies have not directly compared these effects, a number of analyses have
340 supported reproductive value as a driver of variation in the strength of glucocorticoid stress responses
341 within and across species (Silverin et al., 1997; Bókony et al., 2009; Breuner, 2011; Hau et al., 2016, but
342 see Breuner et al., 2003; Krause et al., 2016). However, previous studies have also found that latitude
343 (which is often used as a proxy for environmental variation or harshness: e.g., Bókony et al., 2009;
344 Jessop et al., 2013) covaries positively with stress-induced glucocorticoids. We predict that differences in
345 the relative importance of environmental unpredictability and reproductive value in shaping stress
346 responsiveness is likely to vary across species. Specifically, we predict that the stress responsiveness of
347 income breeders will be more strongly influenced by environmental unpredictability than that of capital
348 breeders. Similarly, we predict that stress responsiveness will be particularly closely tied to
349 environmental unpredictability in species that rely on critical food sources impacted by short-term
350 environmental fluctuations (e.g., aerial insectivores).

351

352 A robust stress response could allow appropriate responses to challenges, but might also result in
353 associated reproductive costs (Wingfield and Sapolsky, 2003). However, we found no evidence that
354 maintaining a robust stress response had negative effects on reproduction. Reproductive success did not
355 vary consistently across populations; instead, the differences that we saw in hatching success (lowest in
356 WY) and fledging success (lowest in NY) mirrored temporary periods of inclement weather in those
357 populations. Consistent with life history theory, females in Alaska and Wyoming showed a higher
358 investment in nestlings. Nevertheless, it is possible that the costs of maintaining an elevated stress

359 response would manifest under more prolonged stressful conditions, or that it imposes longer-term
360 costs (e.g., accelerated telomere shortening or senescence).

361

362 We found support for a mitigating effect of negative feedback on stress responsiveness: female tree
363 swallows breeding in Alaska and Wyoming, where both time constraints on reproduction and
364 temperature unpredictability were the highest, showed both a high stress response and strong negative
365 feedback. Stronger negative feedback allows for a faster decrease in circulating glucocorticoids,
366 potentially reducing the costs associated with sustained glucocorticoid elevation. Thus, strong negative
367 feedback should be particularly beneficial for individuals with a strong stress response (Zimmer et al.,
368 2019). We recently demonstrated that variation in negative feedback efficacy is related to the speed of
369 recovery of the HPA axis from repeated transient stressors, stress resilience, and reproductive success in
370 tree swallows (Taff et al., 2018; Vitousek et al., 2019b; Zimmer et al., 2019). These results indicate that
371 regulation of negative feedback is crucial for coping with perturbances, particularly in more challenging
372 environments.

373

374 Taken together, our results suggest that when environmental conditions become more variable female
375 tree swallows couple high magnitude stress responses and strong negative feedback, instead of
376 decreasing the hormonal response to challenges. Through strong negative feedback individuals could
377 limit the reproductive costs of glucocorticoid exposure, and resume reproductive activities as soon as
378 the stressor has passed. Thus, this combination of glucocorticoid regulatory traits may enable individuals
379 living in harsher environments to balance the challenges of high environmental unpredictability and a
380 brief reproductive period in order to maximize survival and reproductive success.

381

382 Our findings are also consistent with the idea that interactions among glucocorticoid regulatory
383 elements may be important for appropriately responding to challenges, and for fitness (Vitousek et al.,
384 2018b; Vitousek et al., 2019b; Zimmer et al., 2019). In all four populations the strength of the stress
385 response and the efficacy of negative feedback positively covary; however, there were weakly positive
386 or no phenotypic correlations between stress-induced and post-dexamethasone corticosterone levels.
387 Covariation in components of glucocorticoid regulation could result from similar regulatory pathways, as
388 negative feedback is regulated by glucocorticoids binding to glucocorticoid receptors (de Kloet et al.,
389 1998; Romero, 2004). As such, higher glucocorticoid levels could activate more receptors, inducing
390 faster negative feedback (Breuner and Orchinik, 2001; Romero, 2004). However, it has been suggested

391 that different components of the HPA axis are modulated independently (e.g., Romero, 2004).
392 Therefore, the phenotypic correlations seen here could also result from selection favoring combinations
393 of these traits. Determining whether stress-induced and post-dexamethasone corticosterone levels are
394 genetically correlated and whether glucocorticoid profiles covary with receptor expression could help to
395 illuminate the flexibility and physiological underpinnings of these traits and their potential to respond to
396 selection.

397

398 Baseline glucocorticoid levels also differed across populations. However, as both the reproductive value
399 hypothesis and the environmental unpredictability hypothesis predict higher baseline glucocorticoid
400 levels in Alaska and Wyoming, where the season is short and weather unpredictable, these patterns do
401 not allow us to differentiate between the role of these forces in shaping glucocorticoid evolution.
402 Elevated baseline glucocorticoid levels have been shown to support energetically demanding activities,
403 and are associated with more challenging conditions and increased investment in reproduction within
404 and across species (Hau et al., 2010; Bonier et al., 2011; Jessop et al., 2013; Apfelbeck et al., 2017;
405 Vitousek et al., 2019a). The particularly high baseline (and stress-induced) corticosterone levels in
406 Wyoming during incubation may also have reflected a temporary upregulation in HPA axis activation
407 because of a period of unusually cold and wet weather that occurred during this time. In contrast, in
408 Tennessee, in which tree swallows experience a long breeding season and more predictable conditions
409 baseline glucocorticoid levels were consistently low throughout the reproductive period.

410

411 Contrary to our predictions that elevated stress levels would incur physiological costs, we did not detect
412 differences between populations in measures of oxidative damage, antioxidant capacity, or glucose
413 metabolism. Investment in reproduction is not always associated with higher oxidative stress (Garratt et
414 al., 2011; Speakman and Garratt, 2014). However, within individuals both oxidative stress and glucose
415 have been shown to increase with increasing glucocorticoids levels or workload (Sapolsky et al., 2000;
416 Costantini et al., 2011; Metcalfe and Monaghan, 2013). The lack of these relationships at the population
417 level could reflect different trade-offs operating across populations and within individuals.

418

419 *Conclusion*

420 These findings provide support for the hypotheses that environmental unpredictability may be a critical
421 factor in shaping glucocorticoid stress responses, and that selection favoring strong negative feedback in
422 more stress responsive individuals could serve as a mechanism to mitigate the costs of mounting a

423 strong stress response. Our results are also in accordance with the hypothesis that negative feedback
424 and the dynamic regulation of glucocorticoids are important for coping with challenging conditions
425 (Romero and Wikelski, 2010; Taff and Vitousek, 2016; Vitousek et al., 2019b). In the current context of
426 global changes, intraspecific differences in the response to stressors may be particularly important for
427 survival or for the ability to adapt to new conditions (Angelier and Wingfield, 2013; Harding et al., 2019).
428 Our results suggest that this phenotype (elevated stress response and strong negative feedback) has
429 been selected for in unpredictable environments and might therefore be expected to become
430 increasingly common over time, assuming genetic variation exists. However, as climate change affects
431 both the length of the breeding season (Dunn and Winkler, 2010) and environmental predictability
432 (Thornton et al., 2014), populations may face rapidly changing regimes of selection on glucocorticoid
433 regulation outside the bounds of evolutionary history. Confirming that selection is occurring in these
434 populations will require testing whether among individual differences in glucocorticoid phenotype affect
435 fitness. Ultimately, determining the evolutionary causes and consequences of differences in
436 glucocorticoid levels within and among populations will help to reveal how selection drives HPA axis
437 regulation and whether the history of selection on hormonal regulation influences the ability to cope
438 with unpredictable or changing environments.

439

440 **Methods**

441 *Populations*

442 Field data were collected from 2016 to 2018 in four different populations of tree swallows breeding in
443 nest-boxes. Populations were located in Chattanooga, Tennessee (TN) (35.1°N, 85.2°W, 206m
444 elevation), Ithaca, New York (NY) (42.5°N, 76.5°W, 340m elevation), Burgess Junction, Wyoming (WY)
445 (44.5°N, 107.3°W, 2451m elevation) and in McCarthy, Alaska (AK) (61.4°N, 143.3°W, 445m elevation).
446 Tree swallows are widely distributed across North America and breed in a variety of environments.
447 These populations were chosen to allow for comparisons among populations breeding in environments
448 with different degrees of environmental predictability and reproductive value. Populations at higher
449 latitude (Alaska) or elevation (Wyoming) are expected to experience cooler and more unpredictable
450 weather conditions and a shorter breeding season starting later in the year (late May). In Tennessee,
451 near the Southern edge of the breeding distribution, tree swallows are expected to experience warmer
452 and more predictable weather conditions and a long breeding season with the first egg usually laid
453 earlier in the season (early to mid-April). In New York, environmental conditions and breeding season
454 length are expected to be intermediate, with the first egg usually laid in early May.

455

456 *General field methods and stress manipulation*

457 In the four populations, nests were monitored every 1-2 days throughout the breeding season from the
458 initiation of activity at each site to fledging, except for the last week of nestlings' development (to avoid
459 inducing premature fledging). For every active nest, we recorded clutch initiation date and completion
460 dates, clutch size, hatch date, brood size, and the number of nestlings fledged. Nestling fates were
461 determined by checking boxes 22-24 days after hatching. We installed radio frequency identification
462 (RFID) units on each box on the fourth day of incubation (see below). Birds were captured at their nest
463 boxes by hand or using a manually activated trap. All birds were captured and sampled on specific days
464 of life history substages, and during a set time of day, to reduce the variation in circulating
465 glucocorticoid hormones resulting from circadian rhythms. Adult females were captured between 0700
466 and 1000h in NY, TN and WY and between 0600 and 0900h in AK to compensate for the earlier start
467 of activity due to the increased day length compared to the other populations.

468

469 Females were initially captured 6 or 7 days after clutch completion (capture number 1). At this capture,
470 we took a first blood sample within 3 min of initial disturbance to measure baseline circulating
471 corticosterone levels. A second blood sample was taken after 30 min of restraint in a cloth bag to
472 measure stress-induced corticosterone levels. Immediately after this sample was taken, females were
473 injected with dexamethasone (dex) ($0.5 \mu\text{l.g}^{-1}$, Dexamethasone Sodium Phosphate, Mylan Institutional
474 LLC), a synthetic glucocorticoid that binds to receptors within the HPA axis, in order to induce negative
475 feedback (Zimmer et al., 2019). A final blood sample was taken 30 min after dex injection to measure
476 the degree of down-regulation in circulating corticosterone (a measure of negative feedback). Between
477 samples, we weighed the females, and measured the length of their skull from the back of the head to
478 the bill tip (head-bill) and flattened wing length. Non-banded individuals received USGS leg bands and a
479 celluloid color band with attached passive integrated transponder (PIT) tag encoding a 10-digit
480 hexadecimal string (Cyntag, Cynthiana, KY). Female age was determined based on plumage coloration
481 and characterized as second year (SY) or after second year (ASY) (Hussell, 1983).

482

483 As part of a separate study, at their first capture, adult females were allocated to one of the
484 experimental groups: control, feather restraint (in which three primaries were reversibly attached to
485 alter flight ability and thereby increasing the cost of foraging), or predator exposure (see Zimmer et al.,

486 2019 for details on both experimental treatments). Treatments started after the first capture and lasted
487 for 5-6 days.

488
489 Females were then recaptured 5-6 days later (on incubation day 12 or 13; capture 2), at the end of the
490 experimental treatments (see above). At this capture, we only took a baseline blood sample and
491 weighed the bird before release. Finally, we recaptured females again 6-8 days after eggs hatched
492 (capture 3). We followed the same procedure as in capture 1, taking a baseline, restraint stress-induced,
493 and post-dex blood samples, and again weighed each female.

494
495 Twelve days after eggs hatched, each nestling received an USGS leg band, was weighed and had head-
496 bill and flat wing length measured.

497
498 All blood samples were collected from the alar vein, in heparinized microhematocrit capillary tubes.
499 Glucose levels in baseline and stress-induced blood samples were determined in the field (see
500 supplementary). Blood samples were then transferred to microcentrifuge tubes, and kept on ice until
501 centrifugation (within 4h). After separation, the plasma was stored at -20°C in the field and then at -80
502 °C in the lab until analysis. All methods were approved by Cornell IACUC and conducted with
503 appropriate state and federal permits.

504
505 *Provisioning behavior*

506 Number of feeding trips for females from nestling ages 1-18 was recorded using radio-frequency
507 identification (RFID) devices (Cellular Tracking Technologies; Rio Grande, NJ, USA) (Bridge and Bonter,
508 2011). RFID units were installed on each active box on day 4 of incubation. Antennae were fastened
509 around each entrance hole so that birds had to pass directly through an antenna to enter or exit the
510 box. We programmed our RFID units to sample for PIT tags every second between 0500 and 2200h each
511 day as tree swallows are not very active at night. Poll time was set at 500, and cycle time at 1000. The
512 delay time (minimum period of time between successive tag recordings) was set to 1s. RFID boards were
513 powered by 12V 5Ah (PS-1250, PowerSonic, San Diego, CA) batteries that were replaced every five days.
514 At the first capture, each bird was fitted with a PIT tag attached to a color band. Each PIT tag encoded a
515 unique 10-digit hexadecimal string that was recorded, along with a time stamp, when birds passed
516 through or perched on the antenna (see Vitousek et al., 2018a for more details). From the raw RFID

517 records, we determined the number of daily feeding trips for each female through 18 days of age for the
518 brood using an algorithm validated in the New York population (Vitousek et al., 2018a).

519

520 *Corticosterone assay*

521 Steroids were extracted from plasma samples using a triple ethyl acetate extraction and then
522 corticosterone levels were determined using an enzyme immunoassay kit (DetectX Corticosterone,
523 Arbor Assays: K014-H5) previously validated for tree swallows (Taff et al., 2019). Samples were run in
524 duplicate and all samples from an individual were run on the same plate. In total we ran 47 assays with
525 an average extraction efficiency of 92.8 % and a detection limit of 0.47 ng.ml⁻¹. The intra-assay variation
526 was 8.88 % and the inter-assay variation was 11.1 %.

527

528 *Data analysis*

529 To characterize the degree of environmental unpredictability at the different field sites we calculated
530 the unpredictability of temperature variables. We obtained historical weather data for each site over as
531 long a yearly range as possible. For New York we obtained data from the North East Climate Center
532 (<http://www.nrcc.cornell.edu/>) for the Game Farm road weather station (from 1983, located about 7km
533 from field sites) and from the Western Regional Climate Center (<https://wrcc.dri.edu/>) for the Prentice
534 Cooper State Forest station in Tennessee (from 2003, located about 16 km from field sites), the May
535 Creek station in Alaska (from 1990 located about 19 km from field sites) and the Burgess station in
536 Wyoming (from 1992, located about 5 km from field sites). From these data, we extracted the average
537 daily temperature, the daytime average daily temperature (between 0600 and 2200) which is the period
538 when the swallows are the most active and the daily maximum temperature, which is known to affect
539 flying insects' activity and therefore food availability (Winkler et al., 2013).

540

541 We quantified the unpredictability of these temperatures variables during the breeding season: from
542 April to June in Tennessee and New York and from May to July in Alaska and Wyoming. We calculated
543 unpredictability using a general additive model (GAM) following the methods described in Franch-Gras
544 et al. (2017). For each site, these variables were divided by their mean to normalize them before
545 analysis (Franch-Gras et al., 2017). This model considers the dispersion of the data in the time series
546 around a typical curve of the normalized variable. For each weather variable, the typical curve was fitted
547 in a GAM model in relation to the day of the year using the gam function in the mgcv package in R 3.5.3
548 (R Core Team, 2019). As suggested by Franch-Gras (2017), we fitted the GAMs using cubic splines as

549 smoothing function to not *a priori* constrain the shape of the curve. The standard deviation of the
550 residuals of the fitted model (SD_{res}) represents an index of unpredictability for each variable (Franch-
551 Gras et al., 2017). This index gives an overall measure of unpredictability using the historical records and
552 is not intended to indicate variation in weather in the particular years of study at each site.

553

554 We compared females' corticosterone levels by fitting a generalized linear mixed model (GLMM) with a
555 gamma distribution that included population, capture number, sample (baseline, stress-induced and
556 post-dex), female age and their interactions as fixed factors, relative clutch initiation date as a covariate
557 and female identity as random factor. We did not include treatment upon first capture (i.e., feather
558 restraint, predator exposure, or control) as an independent variable in the above analyses as it did not
559 affect HPA axis regulation, i.e., corticosterone levels (treatment x sample: $F_{4,1202} = 0.60$, $p = 0.66$;
560 treatment x capture number x sample: $F_{8,1254} = 1.51$, $p = 0.15$). We further characterized females'
561 corticosterone regulation by calculating their glucocorticoid stress response as the difference between
562 stress-induced and baseline corticosterone levels and their negative feedback as the difference between
563 post-dex and stress-induced corticosterone levels. We compared the magnitude of females' acute stress
564 response and negative feedback strength using GLMMs fit with a normal distribution including
565 population, capture number, female age and their interactions as fixed factors, relative clutch initiation
566 date as a covariate and female identity as a random factor. Within each population, we determined
567 whether corticosterone level at each time point, and stress response and negative feedback were
568 correlated using Pearson correlations.

569

570 In order to compare the populations in terms of breeding synchronization, we calculated the relative
571 clutch initiation date of the first clutch as the number of days after the first laying female in the
572 population. Then, we compared relative clutch initiation date between populations using GLMs with
573 population, female age and their interaction as fixed factors. We also calculated the total breeding
574 season length as the number of days between the first clutch initiation and the last day nestlings fledged
575 at each site. We compared populations using a GLM fitted with a Poisson distribution. We also
576 compared clutch size, brood size, hatching success and fledging success between populations.
577 Population, female age and their interactions were added as fixed factors, and relative clutch initiation
578 day as a covariate. The model for relative clutch initiation date was fitted with a normal distribution,
579 models for clutch size and brood size with a Poisson distribution and models for hatching success and
580 fledging success with a binomial distribution.

581

582 We compared the number of daily feeding trips females made using a generalized linear mixed model
583 GLMM fitted with a Poisson distribution. Population, female age, their interactions and nestling age
584 were added as fixed factors and brood size at each nestling age as a covariate. We also added relative
585 clutch initiation date and brood size as covariates and nest identity as random factor. We used GLMMs
586 to compare nestlings' body mass with population, female age and their interaction as fixed factors.
587 Female identity was added as a random factor.

588

589 GLMs were run using the GENMOD procedure and GLMMs the GLIMMIX procedure in SAS University
590 Edition (SAS Institute Inc., Cary, NC). Post-hoc comparisons were performed using Tukey-Kramer
591 multiple comparison adjustment to obtain corrected p-values. Probability levels <0.05 were considered
592 significant. Data are presented as mean \pm SE.

593

594 **Acknowledgments**

595 Funding was provided by NSF IOS grant 1457151 to MV. We thank the many field and lab assistants who
596 helped who helped with data collection.

597

598 **Competing interests**

599 No competing interests.

600

601 **References**

- 602 Akçay, Ç., Lendvai, Á. Z., Stanback, M., Hausmann, M., Moore, I. T. & Bonier, F. 2016. Strategic
603 adjustment of parental care in tree swallows: life-history trade-offs and the role of
604 glucocorticoids. *Royal Society open science*, 3, 160740. 10.1098/rsos.160740.
- 605 Angelier, F. & Wingfield, J. C. 2013. Importance of the glucocorticoid stress response in a changing
606 world: Theory, hypotheses and perspectives. *General and Comparative Endocrinology*, 190, 118-
607 128. 10.1016/j.ygcen.2013.05.022.
- 608 Apfelbeck, B., Helm, B., Illera, J. C., Mortega, K. G., Smiddy, P. & Evans, N. P. 2017. Baseline and stress-
609 induced levels of corticosterone in male and female Afrotropical and European temperate
610 stonechats during breeding. *Bmc Evolutionary Biology*, 17. 10.1186/s12862-017-0960-9.

- 611 Ardia, D. R. 2005. Tree swallows trade off immune function and reproductive effort differently across
612 their range. *Ecology*, 86, 2040-2046. 10.1890/04-1619.
- 613 Ardia, D. R. 2006. Geographic variation in the trade-off between nestling growth rate and body condition
614 in the tree swallow. *The Condor*, 108, 601-611. 10.1093/condor/108.3.601.
- 615 Ardia, D. R. 2007. The ability to mount multiple immune responses simultaneously varies across the
616 range of the tree swallow. *Ecography*, 30, 23-30. 10.1111/j.0906-7590.2007.04939.x.
- 617 Ardia, D. R. & Clotfelter, E. D. 2007. Individual quality and age affect responses to an energetic
618 constraint in a cavity-nesting bird. *Behavioral Ecology*, 18, 259-266. 10.1093/beheco/arl078.
- 619 Bókony, V., Lendvai, A. Z., Liker, A., Angelier, F., Wingfield, J. C. & Chastel, O. 2009. Stress Response and
620 the Value of Reproduction: Are Birds Prudent Parents? *American Naturalist*, 173, 589-598.
621 10.1086/597610.
- 622 Bonier, F., Moore, I. T. & Robertson, R. J. 2011. The stress of parenthood? Increased glucocorticoids in
623 birds with experimentally enlarged broods. *Biology Letters*, 7, 944-946. 10.1098/rsbl.2011.0391.
- 624 Breuner, C. W. 2011. Stress and Reproduction in Birds. In: Norris, D. O. & Lopez, K. H. (eds.) *Hormones
625 and Reproduction of Vertebrates*. London: Academic Press.
- 626 Breuner, C. W. & Orchinik, M. 2001. Seasonal regulation of membrane and intracellular corticosteroid
627 receptors in the house sparrow brain. *Journal of Neuroendocrinology*, 13, 412-420.
628 10.1046/j.1365-2826.2001.00646.x.
- 629 Breuner, C. W., Orchinik, M., Hahn, T. P., Meddle, S. L., Moore, I. T., Owen-Ashley, N. T., Sperry, T. S. &
630 Wingfield, J. C. 2003. Differential mechanisms for regulation of the stress response across
631 latitudinal gradients. *American Journal of Physiology-Regulatory, Integrative and Comparative
632 Physiology*, 285, R594-R600. 10.1152/ajpregu.00748.2002.
- 633 Bridge, E. S. & Bonter, D. N. 2011. A low-cost radio frequency identification device for ornithological
634 research. *Journal of Field Ornithology*, 82, 52-59. 10.1111/j.1557-9263.2010.00307.x.
- 635 Candolin, U. 2019. Why study intraspecific variation: a comment on Harding et al. *Behavioral Ecology*,
636 30, 1515-1516. 10.1093/beheco/arz162.
- 637 Casagrande, S. & Hau, M. 2018. Enzymatic antioxidants but not baseline glucocorticoids mediate the
638 reproduction - survival trade-off in a wild bird. *Proceedings of the Royal Society B: Biological
639 Sciences*, 285, 20182141. doi:10.1098/rspb.2018.2141.
- 640 Costantini, D., Marasco, V. & Moller, A. P. 2011. A meta-analysis of glucocorticoids as modulators of
641 oxidative stress in vertebrates. *Journal of Comparative Physiology B-Biochemical Systemic and
642 Environmental Physiology*, 181, 447-456. 10.1007/s00360-011-0566-2.

- 643 Dallman, M. F., Akana, S. F., Scribner, K. A., Bradbury, M. J., Walker, C.-D., Strack, A. M. & Cascio, C. S.
644 1992. Stress, Feedback and Facilitation in the Hypothalamo-Pituitary-Adrenal Axis. *Journal of*
645 *Neuroendocrinology*, 4, 517-526. 10.1111/j.1365-2826.1992.tb00200.x.
- 646 de Kloet, E. R., Vreugdenhil, E., Oitzl, M. S. & Joëls, M. 1998. Brain corticosteroid receptor balance in
647 health and disease. *Endocrine Reviews*, 19, 269-301. 10.1210/edrv.19.3.0331.
- 648 Debecker, S. & Stoks, R. 2019. Pace of life syndrome under warming and pollution: integrating life
649 history, behavior, and physiology across latitudes. *Ecological Monographs*, 89, e01332.
650 10.1002/ecm.1332.
- 651 Dunn, P. O. & Winkler, D. W. 2010. Effects of climate change on timing of breeding and reproductive
652 succes in birds. *In: Moller, A. P., Fiedler, W. & Berthold, P. (eds.) Effects of Climate Change on*
653 *Birds*. Oxford: Oxford University Press.
- 654 Eikenaar, C., Husak, J., Escallón, C. & Moore, I. T. 2012. Variation in Testosterone and Corticosterone in
655 Amphibians and Reptiles: Relationships with Latitude, Elevation, and Breeding Season Length.
656 *The American Naturalist*, 180, 642-654. 10.1086/667891.
- 657 Franch-Gras, L., García-Roger, E. M., Franch, B., Carmona, M. J. & Serra, M. 2017. Quantifying
658 unpredictability: A multiple-model approach based on satellite imagery data from
659 Mediterranean ponds. *PLOS ONE*, 12, e0187958. 10.1371/journal.pone.0187958.
- 660 Garratt, M., Vasilaki, A., Stockley, P., McArdle, F., Jackson, M. & Hurst, J. L. 2011. Is oxidative stress a
661 physiological cost of reproduction? An experimental test in house mice. *Proceedings of the Royal*
662 *Society B: Biological Sciences*, 278, 1098-1106. 10.1098/rspb.2010.1818.
- 663 Guindre-Parker, S. 2018. The evolutionary endocrinology of circulating glucocorticoids in free-living
664 vertebrates: Recent advances and future directions across scales of study. *Integrative and*
665 *Comparative Biology*, 58, 814-825. 10.1093/icb/icy048.
- 666 Harding, H. R., Gordon, T. A. C., Eastcott, E., Simpson, S. D. & Radford, A. N. 2019. Causes and
667 consequences of intraspecific variation in animal responses to anthropogenic noise. *Behavioral*
668 *Ecology*. 10.1093/beheco/arz114.
- 669 Hau, M., Casagrande, S., Ouyang, J. Q. & Baugh, A. T. 2016. Glucocorticoid-Mediated Phenotypes in
670 Vertebrates: Multilevel Variation and Evolution. *In: Naguib, M., Mitani, J. C., Simmons, L. W.,*
671 *Barrett, L., Healy, S. & Zuk, M. (eds.) Advances in the Study of Behavior*. Academic Press.
- 672 Hau, M., Ricklefs, R. E., Wikelski, M., Lee, K. A. & Brawn, J. D. 2010. Corticosterone, testosterone and life-
673 history strategies of birds. *Proceedings of the Royal Society B: Biological Sciences*, 277, 3203-
674 3212. 10.1098/rspb.2010.0673.

- 675 Hussell, D. J. T. 1983. Age and plumage color in female tree swallows. *Journal of Field Ornithology*, 54,
676 312-318.
- 677 Jessop, T. S., Woodford, R. & Symonds, M. R. E. 2013. Macrostress: do large-scale ecological patterns
678 exist in the glucocorticoid stress response of vertebrates? *Functional Ecology*, 27, 120-130.
679 10.1111/j.1365-2435.2012.02057.x.
- 680 Krause, J. S., Pérez, J. H., Chmura, H. E., Meddle, S. L., Hunt, K. E., Gough, L., Boelman, N. & Wingfield, J.
681 C. 2016. The stress response is attenuated during inclement weather in parental, but not in pre-
682 parental, Lapland longspurs (*Calcarius lapponicus*) breeding in the Low Arctic. *Hormones and*
683 *Behavior*, 83, 68-74. 10.1016/j.yhbeh.2016.05.018.
- 684 Landys, M. M., Ramenofsky, M. & Wingfield, J. C. 2006. Actions of glucocorticoids at a seasonal baseline
685 as compared to stress-related levels in the regulation of periodic life processes. *General and*
686 *Comparative Endocrinology*, 148, 132-149. 10.1016/j.ygcen.2006.02.013.
- 687 Metcalfe, N. B. & Monaghan, P. 2013. Does reproduction cause oxidative stress? An open question.
688 *Trends in Ecology & Evolution*, 28, 347-350. 10.1016/j.tree.2013.01.015.
- 689 Peñuelas, J., Sardans, J., Estiarte, M., Ogaya, R., Carnicer, J., Coll, M., Barbeta, A., Rivas-Ubach, A., Llusà,
690 J., Garbulsky, M., Filella, I. & Jump, A. S. 2013. Evidence of current impact of climate change on
691 life: a walk from genes to the biosphere. *Global Change Biology*, 19, 2303-2338.
692 10.1111/gcb.12143.
- 693 Ricklefs, R. E. & Wikelski, M. 2002. The physiology/life-history nexus. *Trends in Ecology & Evolution*, 17,
694 462-468. 10.1016/S0169-5347(02)02578-8.
- 695 Romero, L. M. 2004. Physiological stress in ecology: lessons from biomedical research. *Trends in Ecology*
696 *& Evolution*, 19, 249-255. 10.1016/j.tree.2004.03.008.
- 697 Romero, L. M. & Wikelski, M. 2010. Stress physiology as a predictor of survival in Galapagos marine
698 iguanas. *Proceedings of the Royal Society B: Biological Sciences*, 277, 3157-3162.
699 10.1098/rspb.2010.0678.
- 700 Rose, A. P. & Lyon, B. E. 2013. Day length, reproductive effort, and the avian latitudinal clutch size
701 gradient. *Ecology*, 94, 1327-1337. 10.1890/12-0953.1.
- 702 Sapolsky, R. M. 1983. Individual differences in cortisol secretory patterns in the wild baboon: Role of
703 negative feedback sensitivity. *Endocrinology*, 113, 2263-2267. 10.1210/endo-113-6-2263.
- 704 Sapolsky, R. M., Romero, L. M. & Munck, A. U. 2000. How do glucocorticoids influence stress responses?
705 Integrating permissive, suppressive, stimulatory, and preparative actions. *Endocr Rev*, 21, 55-89.
706 10.1210/er.21.1.55.

- 707 Schoenle, L. A., Zimmer, C. & Vitousek, M. N. 2018. Understanding context dependence in
708 glucocorticoid–fitness relationships: The role of the nature of the challenge, the intensity and
709 frequency of stressors, and life history. *Integrative and Comparative Biology*, 777–789.
710 10.1093/icb/icy046.
- 711 Silverin, B., Arvidsson, B. & Wingfield, J. 1997. The adrenocortical responses to stress in breeding Willow
712 Warblers *Phylloscopus trochilus* in Sweden: effects of latitude and gender. *Functional Ecology*,
713 11, 376-384. 10.1046/j.1365-2435.1997.00097.x.
- 714 Speakman, J. R. & Garratt, M. 2014. Oxidative stress as a cost of reproduction: Beyond the simplistic
715 trade-off model. *BioEssays*, 36, 93-106. 10.1002/bies.201300108.
- 716 Stearns, S. C. 1989. Trade-offs in life-history evolution. *Functional Ecology*, 3, 259-268.
- 717 Stearns, S. C. 1992. *The evolution of life histories*, New York, Oxford University Press.
- 718 Taff, C. C. & Vitousek, M. N. 2016. Endocrine Flexibility: Optimizing Phenotypes in a Dynamic World?
719 *Trends in Ecology & Evolution*, 31, 476-488. 10.1016/j.tree.2016.03.005.
- 720 Taff, C. C., Zimmer, C. & Vitousek, M. N. 2018. Efficacy of negative feedback in the HPA axis predicts
721 recovery from acute challenges. *Biology Letters*, 14. 10.1098/rsbl.2018.0131.
- 722 Taff, C. C., Zimmer, C. & Vitousek, M. N. 2019. Achromatic plumage brightness predicts stress resilience
723 and social interactions in tree swallows (*Tachycineta bicolor*). *Behavioral Ecology*, 30, 733-745.
724 10.1093/beheco/arz010.
- 725 Thornton, P. K., Ericksen, P. J., Herrero, M. & Challinor, A. J. 2014. Climate variability and vulnerability to
726 climate change: a review. *Global Change Biology*, 20, 3313-3328. 10.1111/gcb.12581.
- 727 Vitousek, M. N., Johnson, M. A., Downs, C. J., Miller, E. T., Martin, L. B., Francis, C. D., Donald, J. W.,
728 Fuxjager, M. J., Goymann, W., Hau, M., Husak, J. F., Kircher, B. K., Knapp, R., Schoenle, L. A. &
729 Williams, T. D. 2019a. Macroevolutionary patterning in glucocorticoids suggests different
730 selective pressures shape baseline and stress-induced levels. *The American Naturalist*, 193, 866-
731 880. 10.1086/703112.
- 732 Vitousek, M. N., Taff, C. C., Ardia, D. R., Stedman, J. M., Zimmer, C., Salzman, T. C. & Winkler, D. W.
733 2018a. The lingering impact of stress: brief acute glucocorticoid exposure has sustained, dose-
734 dependent effects on reproduction. *Proceedings of the Royal Society B: Biological Sciences*, 285.
735 10.1098/rspb.2018.0722.
- 736 Vitousek, M. N., Taff, C. C., Hallinger, K. K., Zimmer, C. & Winkler, D. W. 2018b. Hormones and fitness:
737 Evidence for trade-offs in glucocorticoid regulation across contexts. *Frontiers in Ecology and*
738 *Evolution*, 6. 10.3389/fevo.2018.00042.

- 739 Vitousek, M. N., Taff, C. C., Ryan, T. A. & Zimmer, C. 2019b. Stress resilience and the dynamic regulation
740 of glucocorticoids. *Integrative and Comparative Biology*, 59, 251-263. 10.1093/icb/icz087.
- 741 Weaver, I. C. G., Cervoni, N., Champagne, F. A., D'Alessio, A. C., Sharma, S., Seckl, J. R., Dymov, S., Szyf,
742 M. & Meaney, M. J. 2004. Epigenetic programming by maternal behavior. *Nature Neuroscience*,
743 7, 847-854. 10.1038/nn1276.
- 744 Williams, G. C. 1966. Natural selection, the costs of reproduction, and a refinement of Lack's principle.
745 *The American Naturalist*, 100, 687-690.
- 746 Wingfield, J. C. 2013. Ecological processes and the ecology of stress: the impacts of abiotic
747 environmental factors. *Functional Ecology*, 27, 37-44. 10.1111/1365-2435.12039.
- 748 Wingfield, J. C., Maney, D. L., Breuner, C. W., Jacobs, J. D., Lynn, S., Ramenofsky, M. & Richardson, R. D.
749 1998. Ecological bases of hormone-behavior interactions: The "emergency life history stage".
750 *American Zoologist*, 38, 191-206. 10.1093/icb/38.1.191.
- 751 Wingfield, J. C. & Sapolsky, R. M. 2003. Reproduction and resistance to stress: When and how. *Journal of*
752 *Neuroendocrinology*, 15, 711-724. 10.1046/j.1365-2826.2003.01033.x.
- 753 Winkler, D. W., Luo, M. K. & Rakhimberdiev, E. 2013. Temperature effects on food supply and chick
754 mortality in tree swallows (*Tachycineta bicolor*). *Oecologia*, 173, 129-138. 10.1007/s00442-013-
755 2605-z.
- 756 Zera, A. J. & Harshman, L. G. 2001. The physiology of life history trade-offs in animals. *Annual Review of*
757 *Ecology and Systematics*, 32, 95-126. doi:10.1146/annurev.ecolsys.32.081501.114006.
- 758 Zimmer, C., Taff, C. C., Ardia, D. R., Ryan, T. A., Winkler, D. W. & Vitousek, M. N. 2019. On again, off
759 again: Acute stress response and negative feedback together predict resilience to experimental
760 challenges. *Functional Ecology*, 33, 619-628. 10.1111/1365-2435.13281.
- 761