Genome wide sequencing provides evidence of adaptation to heterogeneous environments for the ancient relictual *Circaeaster agrestis* (Circaeasteraceae, Ranunculales)

Running title: Adaptive evolution of *Circaeaster agrestis*

Xu Zhang[1, 2, 3, #], Yanxia Sun[1, 2, #, *], Jacob B. Landis[4], Jianwen Zhang[5], Linsen Yang[6], Nan Lin[1, 2, 3],

Huajie Zhang[1, 2], Rui Guo[1, 2, 3], Lijuan Li[1, 2, 3], Yonghong Zhang[5], Tao Deng[5], Hang Sun[5, *],

Hengchang Wang[1,2, *]

[1]CAS Key Laboratory of Plant Germplasm Enhancement and Specialty Agriculture, Wuhan

Botanical Garden, Chinese Academy of Sciences, Wuhan 430074, Hubei, China.

[2]Center of Conservation Biology, Core Botanical Gardens, Chinese Academy of Sciences, Wuhan

430074, Hubei, China.

[3]University of Chinese Academy of Sciences, Beijing, 100049 China.

[4]Department of Botany and Plant Sciences, University of California Riverside, Riverside, CA,

USA.

[5]Key Laboratory for Plant Diversity and Biogeography of East Asia, Kunming Institute of Botany,

Chinese Academy of Sciences, Kunming, Yunnan, 650201 China.

[6]Hubei Key Laboratory of Shennongjia Golden Monkey Conservation Biology, Administration of

Shennongjia National Park, Shennongjia, Hubei, China.

[#]These authors contributed equally to this work.

*Correspondence: sunyanxia@wbgcas.cn; sunhang@mail.kib.ac.cn; hcwang@wbgcas.cn

Summary

● Investigating the interaction between environmental heterogeneity and local adaptation is

critical to understand the evolutionary history of a species, providing the premise for studying the

response of organisms to rapid climate change. However, for most species how exactly the spatial

heterogeneity promotes population divergence and how genomic variations contribute to adaptive

evolution remain poorly understood.

● We examine the contributions of geographical and environmental variables to population

divergence of the relictual, alpine herb *Circaeaster agrestis* as well as genetic basis of local

adaptation using RAD-seq and plastome data.

● We detected significant genetic structure with an extraordinary disequilibrium of genetic

diversity among regions, and signals of isolation-by-distance along with isolation-by-resistance.

The populations were estimated to begin diverging in the late Miocene, along with a possible

ancestral distribution of the Hengduan Mountains and adjacent regions. Both environmental

gradient and redundancy analyses revealed significant association between genetic variation and

temperature variables. Genome-environment association analyses identified 16 putatively adaptive

loci related to biotic and abiotic stress resistance.

● Our genome wide data provide new insights into the important role of environmental

heterogeneity in shaping genetic structure, and access the footprints of local adaptation in an

ancient relictual species.

Keywords: local adaptation, *Circaeaster agrestis*, RAD sequencing, genetic structure,

environmental heterogeneity, temperature variables

Introduction

A primary goal of phylogeography is understanding how geographical and ecological factors shape

the genetic structure and evolutionary history of species. Environmental factors can act in concert

and vary in space, resulting in divergent natural selection leading to genetic divergence during

adaptation to heterogeneous environments (Savolainen *et al.*, 2007). Such genetic divergence along

environmental gradients, or across varied ecological habitats, can be indicative of local adaptation, a

mechanism that is beneficial to the long persistence of a population and intensify differentiation

among populations (Kawecki & Ebert, 2004; Conover *et al.*, 2009; Colautti & Barrett, 2013;

Savolainen *et al.*, 2013; Lowry *et al.*, 2019). Understanding the genetic basis of local adaptation will

provide deeper insights into the response of organisms to ongoing climate change, aiding efforts to

the conservation of genetic resources (Aitken *et al.*, 2008; Ahrens *et al.*, 2019).

On the one hand, divergent natural selection is the driving force of genetic divergence

(Darwin, 1859; Coyne & Orr, 2004), while on the other hand, gene flow can buffer genetic

differentiation (Spieth, 1974; Goicoechea *et al.*, 2019). Evolution resulting from local adaptation

depends on both the strength of local selective pressures and the homogenizing effect of

connectivity among different localities. Additionally, genetic drift may reinforce the establishment

and maintenance of genetic divergence, particularly in populations with small effective size (Wright,

1946; Kimura & Crow, 1964). Therefore, quantifying the relative contributions among natural

selection, genetic drift and population connectivity to genetic divergence is crucial for

understanding the role of local adaptation in adaptive evolution of a species (Savolainen *et al.*,

2007; Friis *et al.*, 2018).

Phenotypic patterns that are associated with divergent selection have been observed in model

plants (Mitchell-Olds *et al.*, 2007; Anderson *et al.*, 2011; Leinonen *et al.*, 2013; Zou *et al.*, 2017).

However, identification of genetic loci that govern phenotypic variations in response to

environmental pressures is a challenging task, especially for non-model species with limited

genomic resources. Facilitated by advancements in sequencing technology and population

genomics, it is now possible to study hundreds of thousands of loci improving our understanding of

genome-wide effects of accumulating genetic divergence and genomic properties that influence the

process of local adaptation (McCormack *et al.*, 2013; Ellegren, 2014; Seehausen *et al.*, 2014;

Weigel & Nordborg, 2015). In particular, highly divergent loci identified by Genome-Environment

Association (GEA) analyses can be interpreted as potential targets of divergent selection associated

with population-specific environmental covariables (Coop *et al.*, 2010; Rellstab *et al.*, 2015; Hoban

et al., 2016; Forester *et al.*, 2018). In this context, plant studies are poised to reveal the genomic

signatures underlying the evolution of local adaptation (e.g. Seehausen *et al.*, 2014; Ahrens *et al.*,

2019). However, for most species, such as relictual alpine herbs, how exactly the spatial

heterogeneity promotes population diversification and how genomic variations contribute to

adaptive evolution, in particular their interaction, remains poorly understood.

In the present study, we focus on investigating the evolutionary history and local adaptation of

Circaeaster agrestis Maxim, an annual alpine herb with a relatively narrow habitat (Fu &

Bartholomew, 2001; Sun, Y *et al.*, 2017). *Circaeaster* Maxim., with its only alliance *Kingdonia*

Balf.f. & W.W. Smith, constitute the early-diverging eudicot family Circaeasteraceae (Ranunculales)

(The Angiosperm Phylogeny et al., 2016). Due to special morphological characters such as open

dichotomous leaf venation, this genus is of great interest to many botanists (Foster, 1971; Ren *et*

al., 2003). *Circaeaster* exhibits low morphological diversity with only one species, garnering a

status of critical endangerment (Wild Plants Under State Protection in China). The distributions of

C. agrestis is confined to the Qinghai-Tibet Plateau (QTP) and adjacent areas where exhibit a wide

range of elevations from 2,100-5,000 m (Fu & Bartholomew, 2001). The uplift of QTP has created

extraordinarily geomorphological and climatic diversity in this area. Due to the high mountain

barrier formed by the Himalayas in the south, the central and western QTP (i.e. the Himalayas) are

characterized by a cold and dry climate (Favre *et al.*, 2015). In contrast, the eastern QTP (i.e. the

Hengduan Mountains) and adjacent arears are associated with deep valleys and characterized

mainly by a warm and wet climate (Ren *et al.*, 2017). Given the high degree of geographical and

ecological heterogeneity, QTP comprises a promising model system for studying genetic signatures

of local adaptation.

Specifically, we generated and analyzed two types of genomic datasets: single nucleotide

polymorphisms (SNPs) derived from the restriction site associated DNA sequencing (RAD-Seq) of

18 *C. agrestis* populations and 20 plastome sequences of *C. agrestis* populations with other

Ranunculales representatives. We aim to illustrate the evolutionary patterns of population

divergence interacting with environmental variables and access the potential genetic basis of local

adaptation. Here, we raise the following questions: 1) how alpine environments affect genetic

structure within species and drive population divergence, and 2) how genetic variation is associated

with local adaptation? Answering these questions will help us understand species adaptation to

rapidly changing climate, and aid in conservation of small patchy diversity in biodiversity hotspots.

Materials and Methods

Sampling, RAD library preparation and sequencing

A total of 139 individuals of *C. agrestis* from 18 localities were sampled (Table S1). Sampling

locations were chosen based on existing occurrence records covering the geographic and climatic

distribution of the species (Fig. 1). All materials were dried and stored in silica gel, and all the

voucher specimens were deposited in Herbarium of Wuhan Botanical Garden (HIB) (Table S1).

Total DNA was extracted from silica gel-dried leaves with a modified CTAB (Cetyl

trimethylammonium bromide) method (Yang *et al.*, 2014). For RAD library construction and

sequencing, genomic DNA was digested with the restriction enzyme EcoRI in a 30 ul reaction. T4

ligase was used to ligate a P1 adapter to the fragments. The adapter-ligated fragments were

subsequently pooled, randomly sheared, and size-selected with a length of 350–550 bp. DNA was

then ligated to a second adapter (P2), a Y adapter hat has divergent ends to the chosen DNA

fragments. The ligation products were purified and PCR amplified, followed by gel purification,

and finally size selected to 350-550 bp. Agilent 2100 Bioanaylzer and qPCR were used to qualify

and quantify the sample library. Paired-end sequencing was performed on two lanes of Illumina

HiSeq 2000 platform at BGI-Shenzhen (Shenzhen, Guangdong, China).

For plastome sequencing, 20 samples covering all 18 localities were chosen. For each sample,

a 500-bp DNA TruSeq Illumina (Illumina Inc., San Diego, CA, USA) sequencing library was

constructed using 2.5-5.0 ng sonicated DNA as input. Then the libraries were quantified using an

Agilent 2100 Bioanalyzer (Agilent Technologies, Santa Clara, CA, USA) and real-time quantitative

PCR. Libraries were multiplexed and sequenced using a 2×125 bp run on one lane of Illumina

HiSeq 2000 platform at BGI-Shenzhen (Shenzhen, Guangdong, China).

Processing of Illumina data

For RAD-sequencing, pair-end Illumina reads were processed into RAD-tags using the STACKS

v.2.30 software pipeline (Catchen *et al.*, 2013). Samples were initially demultiplexed with

PROCESS_RADTAGS. Reads with an average Phred score of at least 30, an unambiguous barcode,

and restriction cut site were retained. Then the perl wrapper DENOVO_MAP.PL was used to

identify optimal parameters by executing USTACKS, CSTACKS, and SSTACKS (Catchen *et al.*,

2013). Following the suggested protocol, we investigated a range of parameter values with M and n

values ranging from 1 to 9 (fixing $M = n$) and $m = 3$ (Rochette & Catchen, 2017). We plotted two

statistics that were useful in making the appropriate parameter choice: the number of polymorphic

loci shared across most samples (the r80 loci) and the distribution of the number of SNPs per locus.

We filtered the catalog of reads using the POPULATIONS module in the STACKS pipeline to

produce datasets for downstream population genetic analyses. Polymorphic RAD loci that

presented in all 139 individuals were retained for subsequent analyses. To validate the influence of

missing data, we also employed a loose parameter of - $r = 0.8$, in which at least 80% of individuals

in a population were required to process a locus for that population. Potential homologs were

excluded by removing loci showing heterozygosity > 0.5 within samples. We further filtered our

dataset with a minor allele frequency (MAF) > 0.01 and kept only biallelic SNPs. Finally, per locus,

we filtered our dataset by selecting the most informative SNP based on the number of minor

alleles.

For genome-skimming sequencing, raw sequence reads were subsequently filtered using

Trimmomatic v.0.36 (Bolger *et al.*, 2014) by removing duplicate reads and adapter-contaminated

reads. Remaining high-quality reads were directly mapped to the reported plastome of *C. agrestis*

(NCBI accession number: KY908400) using Geneious v9.0.2 (Kearse *et al.*, 2012). Initial

annotations were implemented in the Plastid Genome Annotator (PGA) (Qu *et al.*, 2019), and then

were refined by manual corrections in Geneious. The annotated plastomes were deposited in

GenBank under accession numbers xxx-xxx.

Genetic diversity and population structure

We estimated genetic diversity indices including nucleotide diversity (π), expected heterozygosity

(H_e), observed heterozygosity (H_o) and inbreeding coefficients (F_{IS}) using POPULATIONS based

on the RAD-seq data. The polymorphic sites in *Phylip* format were exported from POPULATIONS

and were used for calculating haplotypes in DnaSP v6.12 (Rozas *et al.*, 2017). The *hierfstat* package

(Goudet, 2005) in R v3.6.1 (R Team, 2014) was used to estimate population and group level F_{ST}

using Weir and Cockerham's method (Weir & Cockerham, 1984).

Population genetic structure was estimated using two methods, Bayesian clustering and

principal coordinate analysis (PCoA). Bayesian clustering was performed in STRUCTURE v2.3.4

(Pritchard *et al.*, 2000) with the admixture model. As STRUCTURE analysis assumes that loci are

unlinked, we thus filtered out linked SNPs using the *–write_single_snp* option in the

POPULATIONS module (Catchen *et al.*, 2013). To determine the optimal number of groups (*K*), we

ran STRUCTURE 10 times for each *K* value from 1 through 10. Each run was performed for

200,000 Markov Chain Monte Carlo (MCMC) generations with a burn-in period of 100,000

generations. The optimal *K* was chosen using both delta-*K* method implemented in STRUCTURE

HARVESTER (Earl & vonHoldt, 2012) and cross-entropy criterion implemented in R package *LEA*

v2.8.0 (*snmf* function) (https://bioconductor.org/packages/LEA/). The coefficient for cluster

membership of each individual was averaged across the 10 independent runs using CLUMPP

(Jakobsson & Rosenberg, 2007) and plotted using DISTRUCT (Rosenberg, 2004). We performed

PCoA analyses using the *dartR* v1.1.11 package (*gl.pcoa* function) in R. We tested the overall

population structure using the G-statistic test (Goudet *et al.*, 1996) by *gstat.randtest* function in

adegenet v2.1.1 package (Jombart, 2008). Additionally, an analysis of molecular variance

(AMOVA) was performed to quantify genetic differentiation among and within populations and

genetic groups (as identified by STRUCTURE and PCoA) using Arlequin v3.5 (Excoffier *et al.*,

2007) with significance tests of variance components based on 10,000 permutations.

Population divergence history

We adopted a two-step strategy for investigating the divergent time of *C. agrestis*. First, we used

plastome sequences assembled from Ranunculales and newly sampled *C. agrestis* populations to

estimate the diversification of *C. agrestis*. A total of 47 plastomes were included (Table S1). The 79

protein coding regions (CDS) were extracted and aligned using MAFFT v7.313 (Katoh & Standley,

2013) and were then concatenated using PhyloSuite v1.1.15 (Zhang *et al.*, 2019). The program

BEAST2 v2.5.2 (Bouckaert *et al.*, 2014) was employed for molecular dating with concatenated

matrix using a GTR+G+I substitution model selected by jModelTest v2.0.1 (Darriba *et al.*, 2012)

under the Bayesian information criterion (BIC). A log-normal relaxed-clock mode and birth-death

process were applied. According to fossil records, the time to most recent common ancestors

(TMRCA) of Ranunculales and Berberidaceae were constrained to the minimum age of 112 Ma and

72 Ma, respectively. We set both calibrations as an exponential distribution (Ho & Phillips, 2009)

with a mean of 1 and an offset (hard bound constraint) that equaled the minimum age of the

calibrations. The Markov Chain Monte Carlo (MCMC) run was set as 100 million generations with

sampling every 1,000 generations. Tracer v1.7.1 (Rambaut *et al.*, 2018) was used to assess the

effective sample size (ESS > 200) of each parameter. A maximum clade credibility tree was built by

TreeAnnotator v1.8.0 (Rambaut & Drummond, 2010) with the initial 20% of trees discarded as

burn-in.

Second, we estimated divergent time between haplotypes of *C. agrestis* using the SNP

sequences generated from RAD-seq. All the parameters of BEAST2 were kept the same as above,

with the exception of the TMRCA of *C. agrestis* populations was constrained based on the previous

results. To infer the biogeographic history, ancestral distributions of C. agrestis were estimated by

statistical dispersal-extinction-cladogenesis (S-DEC) (Ree & Smith, 2008), as implemented in

Reconstruct Ancestral State in Phylogenies (RASP) v4.0 (Yu *et al.*, 2015), using 50,000 randomly

sampled trees from posterior distribution generated by BEAST2. We defined the following five

areas according to geographic distributions of *C. agrestis*: A, Qinling Mountain and adjacent

regions; B, Northwestern Sichuan-South of Gansu; C, North of Qinghai; D, Hengduan Mountains;

E, Himalayas. To detect potential demographic expansions, temporal changes in the effective

population size (Ne) were inferred with Extended Bayesian skyline plots (EBSP) as implemented in

BEAST2 (Heled & Drummond, 2008). We used the SNP sequences of 139 individuals and the

default substitution model. The MCMC was set to 50 million generations with sampling every 5000

generations, and the first 20% was discarded as burn-in. Final graphs were produced with the

custom R script provided by Heled (2010), with a burn-in cutoff of 20%.

Effects of heterogeneous landscapes on genetic structure

To illustrate the effects of the heterogeneous landscapes on shaping genetic structure, we conducted Mantel tests for the presence of isolation-by-distance (IBD) and isolation-by-resistance (IBR). To avoid the potential correlation between geographical distance and resistance distance, a partial Mantel test was also conducted to test IBD and IBR by controlling the resistance and geographical distance, respectively. GenAlEx v6.5 (Peakall & Smouse, 2012) was employed to compute the pairwise geographic distance among 18 populations. Resistance distance was generated in CIRCUITSCAPE v4.0.5 based on circuit theory (McRae & Beier, 2007). Circuit resistance provide models of connectivity or resistance to predict patterns of dispersal among sites in a heterogeneous landscape (Dickson *et al.*, 2019). The current ecological niche model (ENM) of environmental niche was calculated in MAXENT (Phillips & Dudik, 2008). We used 19 bioclimatic variables available from the WorldClim2 database (Fick & Hijmans, 2017) at a 30 arc□seconds resolution (Table S1), and actual evapotranspiration (AET; http://www.physicalgeography.net/fundamentals/8j.html) as environmental predictors. To avoid multicollinearity, we ran a Pearson correlation analysis to eliminate one of the variables in each pair with a correlation value higher than 0.9 paired. A total 12 climatic layers was retained for analyses (Table S1).

We used environmental niches representing connectivity between populations as conductance grids to produce pairwise resistance distance, as high habitat suitability was assumed to have low resistance (Nowakowski *et al.*, 2015).Tests of the significance for the relationship between geographical/resistance distances and genetic distance among populations were implemented in the R package *ade4* v1.7 (https://CRAN.R-project.org/package=ade4) using *mantel.rtest* function with 9,999 permutations.

Environmental variables associated with genetic structure

To estimate the contributions of environmental variables to genetic differentiation and to understand the turnover of allele frequencies along an environmental gradient, a gradient forest (GF) analysis

was performed using the R package *gradientForest* v0.1(http://gradientforest.r-forge.r-project.org/).

GF is a nonparametric, machine‐learning regression tree approach that allows for exploration of

nonlinear associations of spatial, environmental, and allelic variables (Gugger *et al.*, 2017; Bay *et*

al., 2018). The analysis partitions the allele frequency data at split values along the environmental

gradients and defines the amount of variation explained as 'split importance' values (Jiang *et al.*,

2019). The overall predictor importance plot of GF shows the mean importance weighted by allele

R^2, and the cumulative plot for each predictor shows cumulative change along the environmental

gradient. The split importance values are added cumulatively to produce a step-like curve, thus,

areas with large steps in a row indicate significant influence on allelic change. The 12 climatic

layers used in above analysis was employed as environmental variables in the GF analysis.

A redundancy analysis (RDA) was performed to understand the associations between genetic

structure and environmental variables. We estimated the proportion of genetic variance in the

populations that is explained by environmental variables using six important variables identified by

GF (bio03: isothermality; AET: actual evapotranspiration; bio04: temperature seasonality; bio07:

temperature annual range; bio12: annual precipitation and bio15: seasonality precipitation, see

results). In our analysis, we constrained the dependent variables (individuals) by the explanatory

variables (climate). The RDA analysis was performed using the *rda* function in the *vegan* v2.5

package (Oksanen *et al.*, 2018) in R (http://CRAN.R-project.org/package=vegan). The *anova.cca*

function was used to calculate overall significance of the RDA plot, and if this was found to be

significant, then significance of each climate variable was calculated, using a significance tests with

9,999 permutations.

Genome‐Environment Association (GEA) analysis

To access the evidence of divergent selection acting on the genome and the genetic signatures of

local adaptation, tests of associated outlier loci with environmental variation were carried out using

BAYESCENV v1.1 (de Villemereuil & Gaggiotti, 2015), which extends the capabilities of the

BAYESCAN algorithm by including a model that incorporates environmental data. BAYESCENV

has been shown to be fairly robust to isolation by distance and a hierarchically structured scenario

(de Villemereuil & Gaggiotti, 2015). For a comprehensive consideration of the environmental effect,

the environmental covariable layer was calculated with principal component analysis (PCA) of the

six important bioclimatic variables, and the first PC axe (PC1), which explained most of the

variability (48.62%), was extracted as the input for BAYESCENV. We ran 20 pilot runs of 5,000

iterations and a burn□in of 50,000 iterations, ultimately obtaining 5,000 MCMC iterations for the

analysis. Diagnostics of the log likelihoods and F_{ST} values for the 5,000 sampled iterations were

checked using the R package *coda* v0.19 (Plummer *et al.*, 2006) to confirm convergence and sample

sizes of at least 2,500.

For functional annotation, the locus consensus sequences of all 6,120 SNPs that were initially

generated by POPULATIONS were exported in *fasta* format. Candidate loci identified in GEA were

annotated with Blast2Go v2 (Gotz *et al.*, 2008), which performs automated annotation

augmentation by searching the non-redundant *Arabidopsis* protein database with records from the

NCBI ESTs databases (Blastx). All the parameters were set as default with e-value < 1.0E-5.

Results

Sequence data quality and processing

We sequenced 139 individuals of *C. agrestis* using Illumina that produced a total of 133 gigabyte

(Gb) data containing 845,396,236 raw reads (Table S2). After filtering out low quality reads, the

average number of used reads per sample was 4,542,944 (median: 4,883,567; Table S2). The depths

of coverage for processed samples were ranged from 5.77x (WLG-2) to 22.44x (HZ2-3), with a

mean coverage of 11.99x (Table S2). We optimized the parameters for STACKS analysis using five

representative samples and identified optimal parameters for our dataset to be $M = n = 3$ (Fig. S1 and

S2). There were 3,640,060 loci containing 2,519,268 variant sites after all samples *de novo*

assembling. After strict filtering (- $r = 1$), we obtained 6,959 RAD loci containing 6,120 SNPs that

were used for population genetic analyses. The mean genotyped sites per locus was 146.59bp (stderr

0.03). After a looser filtering (- $r = 0.8$), we obtained 24,247 RAD loci containing 22,915 variant

sites. The sizes of 20 sequenced plastomes from 18 populations of *C. agrestis* were ranged from

150,979 bp (ZF) to 151,056 bp (SNJ) possessing the same 79 protein coding genes arranged in the

same order (Table S3).

Genetic diversity and structure

We detected an extraordinary difference of genetic diversity at population level, especially for the

nucleotide diversity π, which ranged from 0.0008 (HZ2) to 0.1646 (DDL). The expected

heterozygosity (H_e) and observed heterozygosity (H_o) were 0.0007 (HZ2) - 0.1555 (DDL) and

0.0014 (HZ2) - 0.2650 (DDL; Table 1), respectively. When a loose filtering strategy was employed,

we obtained a similar genetic diversity pattern (Table S4). In total, 66 haplotypes from 139

individuals (Fig. 1; Table S5) were detected, and the haplotype diversity (H_d) was 0.9798. Genetic

differentiation between YL and HZ3 is the highest (F_{ST}=0.9913), and that between ZG and DDL is

the lowest (F_{ST}=0.0640; Table S6; Fig. S3a).

We conducted a Bayesian clustering analysis to explore the genetic structure of *C. agrestis*.

The delta-K method identified best-fit number for clustering (highest ΔK value) was two (Fig. S4a),

representing the East and West clades. While the method based on the cross-entropy criterion

indicated a better number of clusters was six (Fig. S4b). Considering high F_{ST} among populations,

we further subdivided two clades into two and four groups (E1, E2, W1, W2, W3 and W4),

respectively, which was also supported by the PoCA results (Fig. 2). The AMOVA revealed that

77.59% of the overall variation was distributed among six groups, with 11.27% explained by

variation among populations within groups (Table S7). Within the genetic groups, W2 exhibited the

highest nucleotide diversity (π; 0.1646), while W3 had the lowest (0.0206). All genetic groups had

lower observed heterozygosity than expected, with the exception of W2. In general, the East clade

has lower levels of observed heterozygosity (H_o, E1=0.0029, E2=0.0034, W1=0.0548, W2=0.2650,

W3=0.0134, W4=0.0088) and higher levels of inbreeding coefficient (F_{IS}, E1=0.1952, E2=0.1011,

W1=0.0384, W2=-0.1806, W3=0.0227, W4=0.0510) than the West clade (Table 1). Among genetic

groups, genetic differentiation between W3 and W4 is the highest (F_{ST}=0.8908), followed by E2

and W4 (F_{ST}=0.8451), and that between E1 and E2 is lowest (F_{ST}=0.1284; Table S8; Fig. S3b). The

G-statistic test suggested a significant population structure ($p = 0.01$**; Fig. S3c).

Population divergence history

The 47-taxa 79-CDS matrix of Ranunculales used for estimating the original divergence of *C. agrestis* was 66,117 bp in length containing 12,998 parsimony-informative sites. The divergence of *C. agrestis* from its closely related *Kingdonia* was estimated to occur in the early Eocene (stem age, ca. 52.19 Ma; 95% highest posterior density [HPD] intervals, 26.12-83.13 Ma), whereas the early divergence within *C. agrestis* started in the late Miocene (ca. 6.33 Ma; 95% HPD, 1.96-23.35 Ma) (Fig. S5). The tree topology inferred by BEAST2 using the SNP sequences generated from RAD-seq identified two main clades (East and West clades) and six sub-groups (E1, E2, W1, W2, W3 and W4 groups), which was consistent with STRUCTURE and PoCA results. The divergence of two main clades was estimated at appropriately 6.71 Ma (95% HPD, 5.18-7.89 Ma). The subsequent divergences between six groups occurred mainly during the Pliocene to late Pleistocene (ca. 5.05Ma-2.90Ma) (Fig. 3a). Biogeographic analysis supported Hengduan Mountains and adjacent regions as the most likely ancestral region for C. agrestise (Fig. 3a and 3b). For the demographic inference, EBSP analysis showed no sign of population expansion or contraction in demographic history of *C. agrestis* (Fig. 3c).

Effects of topography and ecology

Given that the distribution range of *C. agrestis* is topographically complex, we evaluated the influence of the heterogeneous landscape in shaping genetic structure via Mantel test. Our analyses suggested both significant patterns of IBD ($R^2 = 0.443$, $P = 0.001**$) and IBR ($R^2 = 0.315$, $P = 0.017*$) (Fig. 4a and b), which were also supported by a partial Mantel test ($R_1^2 = 0.426$, $P = 0.001**$; $R_2^2 = 0.192$, $P = 0.011**$). In our case, geographical distance explained the genetic structure better than resistance distances based on current distribution of habitats. We further plotted the conductance grid derived from the IBR model to illustrate the connectivity among populations and thus represent the empirical costs of gene flow through a complex landscape. In general, connectivity gaps among populations were observed in the western clade, reflecting increased habitat isolation in this area resulted from complex topography. In contrast, locations of eastern clade had low resistance to dispersal thus high connectivity among populations (Fig. 4c).

Environmental variables associated with genetic structure

Among the 12 variables used for GF analysis, isothermality (bio03) was indicated as the most

important predictor. Actual evapotranspiration (AET), temperature annual range (bio07) and

temperature seasonality (bio04) were also of high importance. Seasonality precipitation (bio15),

and annual precipitation (bio12) showed moderate importance to allele frequencies. The other six

environmental variables had relatively lower contributions (Fig. 5a). Fig. 5b showed the cumulative

change in overall allele frequencies, where changes evolved with the environmental gradient.

The RDA analysis revealed that a significant amount of genetic variation among populations is

associated with the six important environmental variables (46.58%, $p = 0.001$). Each of the two axes

explained a significant amount of variation (Axis 1: 58.55%, $P = 0.001**$; Axis 2: 27.31%, $P =$

$0.001**$; Fig. 6). We also conducted RDA analysis using each six environmental variables

separately and found that all six variables were significant. The contribution on genetic variation of

each variable generally consistent with the GF analysis. Isothermality (bio03; 18.93%, $P = 0.001**$)

was the most important predictor, whereas the temperature annual range (bio07) was less important

in RDA analysis (Table 2).

Genome☐Environment Association (GEA) analysis

The BAYESCENV result revealed a total of 61 loci have a significant association with

environmental variables. We then aligned consensus sequences of these candidate loci for

annotation. Sixteen of the 61 loci were successfully annotated by Blast2Go with cut-off of e-value

<1E-5 and a mean similarity more than 80 (Table 3; Table S9). Among these were genes related to

abiotic stress responses and stress-induced morphological adaptations, such as *UGT74E2* (Loci ID:

2763) regulating auxin homeostasis (Tognetti *et al.*, 2010), *SRK2E* (Loci ID: 8038) involving in

the resistance to drought by avoiding water loss (Mustilli *et al.*, 2002) and *AHK5* (Loci ID:

203056) regulating stomatal state and transmitting the stress signal (Pham *et al.*, 2012), as well as

genes required for the formation and development of leaves (*CYP71*, loci ID: 293398; *CSI1*, loci

ID: 218068) (Li & Luan, 2011). These loci also included genes related to oxidation-reduction

process involved in cellular respiration, as for example, *NDUFS7* (Loci ID: 83859) and *NDUFB3*

(Loci ID: 219291); and genes implicated in the defense of plants against pathogens, such as

Glucan endo-1,3-beta-glucosidase B (Loci ID: 296907), *RPPL1* (Loci ID:146932) and *NUDT8*

(Loci ID: 243120) (Stroud *et al.*, 2016).

Discussion

Effects of topography on genetic structure

Our study presents the genetic structure of the palaeoendemic alpine *C. agrestis*. High haplotype

diversity, uneven distribution of genetic diversity, as well as significant IBD and IBR patterns

suggest restricted gene flow is likely a key factor in the high genetic differentiation among *C.*

agrestis populations. The distribution area of *C. agrestis* is remarkable for its extraordinarily

topographic diversity with some of the world's most rugged mountain ranges and altitudinal

gradients spanning over 5,000 m (Favre *et al.*, 2015). As a direct consequence of the uplift of the

QTP, habitat fragmentation induced by complex topography and frequented climatic fluctuation

was invoked as an important mechanism for the high diversity of alpine plants in this area (Wen *et*

al., 2014; Sun, H *et al.*, 2017). Hence, dramatically topographic and climatic changes triggered by

mountain uplift probably have influenced genetic connectivity and shaped current genetic structure

of *C. agrestis*.

The high IBR pattern of western clade, reflecting the habitat isolation and perhaps caused by

complex topography, may be the cause for limited gene flow as well as observed high

differentiation. The Himalayan orogeny likely triggered the vicariance among different *C. agrestis*

groups. The north-south mountain chains and valleys may provide corridors for flora exchange

between the north and south, but represent barriers to migration between the east and west (Chen *et*

al., 2017); this likely resulted in the observed meridional differentiation of *C. agrestis* in the

western clade. Previous studies (e.g., Gao *et al.*, 2007; Li *et al.*, 2011) have also suggested

phylogeographic isolation associated with mountains and valleys with north-south direction.

Notably, the populations located in the Hengduan Mountains (HM), exhibit high levels of genetic

diversity (Table 1). As the QTP blocks the cold and dry air toward HM (Liu *et al.*, 2013), the HM

region is characterized by a much warmer and wetter climate, providing a comparatively stable

environment for species. Although the western individuals are likely to possess higher genetic

diversity, they are also likely to be more vulnerable to shifting disturbance regimes and

environmental changes due to lack of genetic connectivity resulted from habitats isolation.

Compared to the western clade, north-south mountain barriers among different latitudes may

be the reason for the divergence among the two groups in the eastern clade. A conductance grid

derived from the IBR model showed low resistance to dispersal, thus high connectivity, among

populations in the eastern clade. In addition, E1 and E2 exhibit a relatively low level of genetic

differentiation. This phenomenon is likely due to the similar local environments among different

populations resulting from the quite moderate topographical and climatic characteristics in this

region (Chen *et al.*, 2017). We detected a significantly high F_{IS} values combined with low level of

observed heterozygosity in the eastern clade (Table 1), indicating high levels of inbreeding, which

may attribute to disturbance caused by human activities or the possible bottleneck events (Rands et

al., 2010). There is consensus that breeding between related individuals can lead to a decrease in

heterozygosity and fixation of deleterious alleles that may reduce reproductive fitness and increase

the population susceptibility to environmental pressure (Kramer & Havens, 2009; Poudel *et al.*,

2014), making populations vulnerable to cope with rapid climate changes. Notably, the highest F_{IS}

value was detected in WLG (F_{IS}: 0.0621), indicating a key protected population for *C. agrestis*

populations in this region.

Impact of environmental heterogeneity in population divergence

Our molecular dating based on plastome sequences suggested an early Eocene origin of *C. agrestis*,

which is accordant with the previous estimation (Ruiz-Sanchez *et al.*, 2012) and our results from

nuclear data (Sun *et al.*, 2020). The diversification of *C. agrestis* has increased since the Pliocene

(Fig. 3), which is likely related with this period of intense uplift of the HM during the late Miocene

to late Pliocene (Sun *et al.*, 2011; Favre *et al.*, 2015). The environmental and climatic fluctuation

triggered by mountain uplifts can always led to divergent selection and species adaptation

associated with dramatic ecological niches changes (Ren *et al.*, 2017). Orogenic activities not only

provided numerous new habitats favoring for diversification, but also promoted population

divergence due to the hindered gene flow across a continuous altitudinal difference. Our

biogeographic analysis revealed that the demographic evolution of *C. agrestis* has not simply been

influenced by the orogeny at a particular time in history, but is more complex, involving repeated

episodes of dispersal and subsequent vicariance events over multiple timescales. Given the EBSP

analysis showing a relatively stable population size in demographic history of *C. agrestis*,

geographical events induced environmental and climatic fluctuation that promotes population

divergence, while likely had a small effect on demographic changes. Coupled with the evidence of

significant IBD and IBR patterns, we speculate natural selection driven by extraordinary

environmental heterogeneity and geomorphological dynamics may have profound impact in the

diversification of *C. agrestis*. In addition, during our field work we found population sizes of *C.*

agrestis in the Himalaya-Hengduan Mountains to be much smaller than other sites. Previous

studies have indicated a significant negative effect of small population size on adaptive evolution

(Nevado *et al.*, 2019). Hence, we speculate that natural selection may play double-edged roles in the

evolutionary history of any given species. On one hand, heterogeneous environments could provide

more new ecological niches conducive for speciation, on the other hand, it may put the populations

with weak adaptive ability to new environment in danger of extinction.

Reproductive isolation induced by geographical distance is likely to be another important

factor driving the population divergence of *C. agrestis*. The significant population structure from

AMOVA analysis, along with high level of F_{ST} values, and small population numbers discovered in

the field commonly reflect there is a potential for reproductive isolation among different *C. agrestis*

populations, especially in the western clade. For populations with small size, deleterious mutations

and compensatory adaptive mutations can be quickly and independently fixed, causing reproductive

isolation coupled with high level of genetic differentiation (Nielsen, 2005). However, *C. agrestis*

exhibits low morphological diversity and no sign of expanding population demographic history. We

speculate that strong environmental selective pressures are constraining morphological diversity

while mutations are leading to observable population differentiation. Nonetheless, cryptic

speciation cannot be ruled out, which needs further crossing experiments or/and testing for

ecological niche divergence/overlap. Furthermore, significant IBR pattern suggest that genetic

structure would be better explained by habitat isolation between populations than just by

geographical distance alone. Therefore, habitat changes may have large effect on genetic structure

for *C. agrestis*, indicating its sensitivity to environmental niche.

Studying local adaptation contributes to understanding the ability of populations to sustain or

adapt to rapid climate change (Jia *et al.*, 2019). We associated environmental variables with

genetic structure to demonstrate the impact of environmental heterogeneity on population

divergence. GF and RDA analyses indicated that covariation between genetic structure and

environmental variables have played a role in promoting the diversification of *C. agrestis*.

Specifically, both analyses found a significant association between genetic variation and

temperature variables (Fig. 4; bio03, AET, bio04, bio07), suggesting temperature is potentially an

important driver of genetic variation within *C. agrestis*, as detected in previous studies of

angiosperm trees (Ahrens *et al.*, 2019; Jia *et al.*, 2019; Jiang *et al.*, 2019). In fact, there is growing

evidence that high-mountain environments experience more rapid changes in temperature than

environments at lower elevations (Pepin et al., 2015; Palazzi, Mortarini, Terzago, & von

Hardenberg, 2019). Despite that, little attention has been given to understand the mechanisms of

adaptive response to climate changes in high altitude areas. *C. agrestis* is habitually confined to

grow under trees or shrubs, indicating it is sensitive to temperature. Moreover, the vertical

distribution of *C. agrestis* is extensively wide ranging from 2,100 to 5,000 meters above sea level

(Fu & Bartholomew, 2001), exhibiting a considerable differentiation in temperature. Therefore, it

is reasonable that temperature variables play a vital role in promoting genetic divergence within *C.*

agrestis. Furthermore, given the ongoing climate warming, temperature is likely still to be a key

driver for adaptation of *C. agrestis* in future.

Genomic signatures associated with local adaptation

Simulation studies showed that the GEA method, which incorporates environmental information

in the form of "environmental differentiation", has a much lower false positive rate than

genome-scan methods, and is robust to IBD pattern as well as a hierarchically structured scenario

(de Villemereuil & Gaggiotti, 2015). In the present study, we first employed PCA to calculate the

overall environmental differentiation and provide the first axis (PC1) to BAYESCENV as a

description of the environmental distance between populations. Considering that all environmental

pressures drive species divergence together, this approach could provide a realistically and

comprehensively environmental gradient for studying local adaptation.

In our analysis, BAYESCENV identified a set of genes that may be involved in a

continuously adaptive process. Some genes as pivotal kinases involved in signaling pathway

regulation, may induce abiotic stress responses for defending harsh climate in QTP. For example,

genes associated with abscisic acid (ABA) signaling pathway (*SRK2EA*) can regulate numerous

ABA responses, such as stomata closure, to response to high altitudes with extreme difference in

temperature (Sierla *et al.*, 2018). Considering the tiny leaves of *C. agrestis* and their sensitivity to

temperature, we speculate that stomatal regulation would be a vital process for maintaining

hydration and regulating temperature. As altitude increases, hypobaric hypoxia becomes the main

factor interfering with life activity. Genes related to oxidation-reduction processes involved in

cellular respiration, such as *NDUFS7* and *NDUFB3*, are of importance in facilitating adaptation of

C. agrestis to high altitude areas of hypoxia. In addition, some putatively adaptive genes (*CYP71*

and *CSI1*) are associated with physiological trait regulation of vegetative and reproductive organs

(Gu *et al.*, 2010). Thus, morphological adaptations may be an essential alternative for *C. agrestis*

to respond to environmental stress.

Despite limited available gene information in this study, it should be noted that genetic

patterns of adaptation to environmental heterogeneity are likely driven by a wide range of alleles

of small-effect for complex traits (Savolainen *et al.*, 2013). The physiological processes involved

in the adaptation of *C. agrestis* to alpine environments are most likely complex traits determined

by multiple genes with potentially adaptive mechanisms including regulation of expression and

epigenetic alteration (Gomulkiewicz *et al.*, 2010; Ahrens *et al.*, 2019). The genes identified as

putatively adaptive signals in *C. agrestis* are most likely a part of the broader process of adaptation

to climate and should constitute the first step in studying genetic mechanism of local adaptation in

alpine environments. Further work with characterizing whole genome could further illuminate

patterns of functional genes controlling the response to climate changes. Generally, these patterns

can serve as a proxy for understanding and monitoring the adaptive process in rapidly changing

environment (Ahrens *et al.*, 2019) and inform future management strategies.

Conclusions

In this study, we demonstrate two major advances. First, we provide new insights into the genetic

structure and evolutionary history of a relictual alpine herb surviving in heterogeneous environment,

which can inform further conservation. Our results shed light on vital roles of mountain uplift in

promoting population diversification and the dual effect of environmental heterogeneity on species

evolutionary history. Second, our analyses provide a roadmap to study evolution of local adaptation

of non-model species for which common gardens and genomic resources are limited. By conducting

multiple analyses, we associated environmental variables with genetic variation and explicitly

measured the relative influence of bioclimate on population divergence. In turn, the important

environmental predictors were implemented in identifying potential loci that may under divergent

selection to gain in-depth knowledge of genetic basis involved in local adaptation. More importantly,

our results improve our understanding of drivers and mechanisms of adaptive evolution, aiding

efforts on developing sound conservation programs in facing of rapid changing climate.

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Author contributions

HCW, HS, and YXS developed the idea and designed the experiment. JWZ, LSY, NL, HJZ, RG,

LJL, YHZ and TD collected the leaf materials. XZ and YXS performed the statistical analyses; XZ,

YXS, JBL, HS and HCW interpreted the results and wrote the manuscript. All authors read, edited

and approved the final manuscript. XZ and YXS contributed equally to this work.

Data accessibility

All newly sequenced plastomes were deposited in National Center for Biotechnology Information

(NCBI) with accession numbers xxx-xxx. RAD raw read data are stored at the NCBI Sequence

Read Archive in the Bioproject xxx. Genomic data sets and R scripts used in this study are

available in the DRYAD archives under accession xxx.

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Table 1 Summary of statistics calculated for the 6120 variant positions. n, number of genotype; π, average nucleotide

diversity; H_e, average expected heterozygosity per locus; H_o, average observed heterozygosity per locus; the

inbreeding coefficients (F_{IS}). The statistics of six defined groups are bolded. E1, eastern group1; E2, eastern group2;

W1, western group1; W2, western group2; W3, western group3; W4, western group4

Group IDs	Population IDs	n	π	H_e	H_o	F_{IS}
E1	**-**	**75**	**0.0464**	**0.0461**	**0.0029**	**0.1952**
	SNJ	13	0.0207	0.0199	0.0038	0.0383
	TBS	17	0.0218	0.0212	0.0029	0.0532
	HZ2	9	0.0008	0.0007	0.0014	-0.0010
	HZ3	10	0.0013	0.0012	0.0022	-0.0017
	KMG	8	0.0015	0.0014	0.0027	-0.0023
	TSM	9	0.0214	0.0203	0.0027	0.0425
	WLG	4	0.0306	0.0287	0.0034	0.0621
	XLS	5	0.0223	0.0210	0.0026	0.0375
E2	**-**	**19**	**0.0384**	**0.0374**	**0.0034**	**0.1011**
	BLS	9	0.0216	0.0204	0.0035	0.0529
	SGN	10	0.0254	0.0241	0.0033	0.0573
W1	**-**	**7**	**0.0688**	**0.0639**	**0.0548**	**0.0384**
	MXG	4	0.0215	0.0189	0.0220	-0.0005
	YL	1	0.0016	0.0008	0.0016	-
	ZG	2	0.1633	0.1225	0.1470	0.0244
W2	**DDL**	**9**	**0.1646**	**0.1555**	**0.2650**	**-0.1806**
W3	**SJL**	**10**	**0.0206**	**0.0196**	**0.0134**	**0.0227**
W4	**-**	**19**	**0.0249**	**0.0242**	**0.0088**	**0.0510**
	YH	8	0.0150	0.0140	0.0057	0.0238
	GS	1	0.0255	0.0128	0.0255	-
	ZF	10	0.0136	0.0130	0.0095	0.0100

837 **Table 2** RDA results based on six important environmental variables identified by GF analysis.

Environmental variables	Constrained proportion	F	p
Bio03	0.1893	31.98	0.001**
AET	0.1712	28.31	0.001**
Bio04	0.1408	22.45	0.001**
Bio12	0.0902	13.59	0.001**
Bio07	0.0627	9.17	0.001**
Bio15	0.0203	2.84	0.001**

 Table 3 Annotation information of 16 candidate loci under selection that the E-value of blastx is less than 1E-5.

840   Detailed functional annotation is provided in Table S9.

Loci ID	Gene names	Description	Length	E-Value	similarity
2763	*UGT74E2*	UDP-glycosyltransferase 74E2-like	146	1.16E-19	82.08
8038	*SRK2E*	Serine/threonine-protein kinase TIO-like isoform X1	146	4.97E-07	81.56
8150	*CSC1*	Calcium permeable stress-gated cation channel 1-like	146	5.12E-20	92.71
19591	*At1g18270*	Ketose-bisphosphate aldolase class-II family protein	146	6.34E-06	87.07
83859	*NDUFS7*	NADH dehydrogenase subunit 7 (mitochondrion)	146	7.55E-26	93.96
126986	*RTL*	Retrotransposon gag protein	147	4.78E-14	84.5
146932	*RPPL1*	Putative disease resistance RPP13-like protein 1 isoform X1	146	2.78E-06	87.52
203056	*AHK5*	Histidine kinase 5	146	7.55E-14	84.61
206146	*ABCC2*	ABC transporter C family member 2-like	146	2.77E-15	89.16
218068	*CSI1*	Protein cellulose synthase interactive 1	146	3.74E-14	86.88
219291	*NDUFB3*	NADH dehydrogenase [ubiquinone] 1 beta subcomplex subunit 3-B-like	146	8.26E-20	82.13
243120	*NUDT8*	Nudix hydrolase 8	146	8.49E-14	81.72
275249	*N/A*	Glycosyl transferase (glycosyl transferase family 2)	146	3.38E-24	85.59
293398	*CYP71*	Peptidyl-prolyl cis-trans isomerase CYP71	146	7.81E-17	92.11
296907	*N/A*	Glucan endo-1,3-beta-glucosidase 13-like	146	1.45E-20	83.33
305334	*CSLC6*	Probable xyloglucan glycosyltransferase 6	146	7.31E-26	98.54

Figure legends

Fig. 1 Geographical distribution of haplotypes detected in polymorphic sites of Circaeaster

agrestis populations. Colors indicate genetic groups based on STRUCTURE and PoCA results

(see also Fig. 2), with the Eastern clade being made up of the orange and dark green circles,

with the Western clade consisting of the light green, light pink, purple, and blue circles. Circle

sizes reflect the number of individuals sampled.

Fig. 2 Population genetic structure in Circaeaster agrestis based on 6,120 SNPs from (a) PCoA

for all populations, east-clade and west-clade; and (b) STRUCTURE with K=2 (showing the

highest ΔK; Fig. S4a) and 6 (representing the divergence within east and west clades).

Fig. 3 Evolutionary history of Circaeaster agrestis. (a) Divergent time estimation and ancestral

area reconstructions of 66 haplotypes based on SNPs sequences. Mean divergence dates and 95%

HPDs for major nodes (1–5) are summarized in top left. Pie charts on each node indicate

marginal probabilities for each alternative ancestral area derived from S-DEC. Results are based

on a maximum area number of two. Inferred dispersal and vicariance events are indicated by

green and red circle respectively. (b) Five defined areas according to geographic distributions of

C. agrestis: A, Qinling Mountains; B, Sichuan-South of Gansu; C, North of Qinghai; D,

Hengduan Mountains; E, Himalayas. (c) Temporal changes in the effective population size (Ne)

changes inferred with Extended Bayesian skyline plots (EBSP) using BEAST2. Re,

reproductive number, δ, death rate.

Fig. 4 Topographic and ecological effects on genetic structure. Relationship of genetic distance

and (a) geographical distance, (b) resistance distance based on climatic niche suitability, as

tested by a Mantel test. (c) Conductance grid derived from the Isolation-by-Resistance (IBR)

model. Grid with warm colour indicates high conductance, cold colour indicates high resistance.

Fig. 5 (a) R2-weighted importance of environmental variables that explain genetic gradients

from gradient forest (GF) analysis. (b) Cumulative importance of allelic change along the first

six environmental gradients.

Fig. 6 Redundancy analysis showing the relationship between the independent climate

parameters and population structure. Individuals are colored points and colors represent six

groups (E1, E2, W1, W2, W3, W4). Small black points are SNPs. The percentage of explained

variance for each axis is provided.

Supporting information

Fig. S1 The number of 80% polymorphic loci shared across most samples (the r80 loci) as M=n

increases.

Fig. S2 The distribution of the number of SNPs per locus for a range of M=n values.

Fig. S3 (a) F_{ST} values within populations estimated by *hierfstat,* Group names of populations were

labeled in the top of boxes. (b) F_{ST} values within genetic groups estimated by *hierfstat*. (c) The

results of G-statistic test.

Fig. S4 The optimal K value identified using (a) delta-K method in STRUCTURE HARVESTER,

and (b) cross-entropy criterion in R package *LEA*.

Fig. S5 Divergence times of Ranunculales estimated by BEAST2 with a relaxed molecular clock

based on the combined protein-coding region sequences. A and B indicate fossil calibration points. C

and D indicate the origination and diversification of Circaeaster agrestis, respectively. Median ages

of nodes are shown with bars indicating the 95% highest posterior density intervals for each node.

Table S1 Summary of population information of Circaeaster agrestis, including the number of

individuals genotyped at each localities (n), geographic information,voucher specimens information,

and 20 bioclimate values (bio01-bio19 and AET) of each location. AET, the actual

evapotranspiration.

Table S2 Numbers of reads of RAD-seq for each individual of Circaeaster agrestis.

Table S3 Assemble and annotation information of newly sequenced plastomes of Circaeaster

agrestis populations.

Table S4 Summary of statistics calculated for the 22915 variant positions. n, number of genotype; π,

average nucleotide diversity; He, average expected heterozygosity per locus; Ho, average observed

heterozygosity per locus; the inbreeding coefficients (FIS). The statistics of six defined groups are

bolded. E1, eastern group1; E2, eastern group2; W1, western group1; W2, western group2; W3,

western group3; W4, western group4.

Table S5 Haplotypes information detected in 18 populations based on 6120 SNPs.

Table S6 Paired population-level FST estimated by hierfstat using Weir and Cockerham's method.

Table S7 The analysis of molecular variance (AMOVA) for SNP data among six genetic groups (E1,

E2, W1, W2, W3 and W4)

Table S8 Paired group-level FST estimated by hierfstat using Weir and Cockerham's method.

Table S9 Detailed annotation information of sixteen genes under potential divergent selection.

(a)

Node	Time	95% HPD
1	6.71	5.18-7.98
2	5.03	3.45-7.49
3	3.96	2.63-5.04
4	3.77	2.50-4.90
5	2.90	1.71-4.37

East clade

West clade

E1

E2

W4

W2

W1

W3

Legends

*

A   BE
AB   C
B   D
BC   DE
BD   E

○ Dispersal
○ Vicariance

Miocene   Pliocene   Pleistocene
7.0 6.0 5.0 4.0 3.0 2.0 1.0 0.0 Ma

(b)

(c)