

Defining endemism levels for biodiversity conservation: tree species in the Atlantic Forest hotspot

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Abstract

2 Endemic species are essential for setting conservation priorities. Yet, quantifying
endemism remains challenging because endemism concepts can be too strict (*i.e.* pure
4 endemism) or too subjective (*i.e.* near endemism). We use a data-driven approach to
objectively estimate the proportion of records outside a given the target area (*i.e.*
6 endemism level) that optimizes the separation of near-endemics from non-endemic
species. Based on millions of herbarium records for the Atlantic Forest tree flora, we
8 report an updated checklist containing 5062 species and compare how species-specific
endemism levels match species-specific endemism accepted by taxonomists. We show
10 that an endemism level of 90% delimits well near endemism, which in the Atlantic
Forest revealed an overall tree endemism ratio of 45%. The diversity of pure and near
12 endemics and of endemics and overall species was congruent in space, reinforcing that
pure and near endemic species can be combined to quantify endemism to set
14 conservation priorities.

Key-words: biodiversity hotspot, endemism centres, endemism ratio, near endemism,
16 occasional species, plant conservation, species richness

18 1 INTRODUCTION

In times of increasing threats to biodiversity and limited resources for its conservation,
20 prioritizing actions is essential. One common practice in biodiversity conservation is to
target areas with many flagship species, such as species threatened with extinction (*i.e.*
22 threatened species) or those exclusive to a given region or habitat (*i.e.* endemic species).
These flagship species are important for conservation because they have a greater
24 extinction risk than other species (Brooks et al., 2006; Myers et al., 2000; Peterson &
Watson, 1998). In addition, patterns of total and endemic species richness can be

26 congruent (Kier et al., 2009; Letters et al., 2002; Storch, Keil, & Jetz, 2012), so the
protection of high-endemism areas could also safeguard the remaining biodiversity.
28 However, there have been more efforts to delimit threatened species than endemic ones.
Threatened species are grouped by clearly-defined categories enclosed by objective
30 criteria (IUCN, 2018), while species often are classified simply as being endemic or not.

There are proposals to divide endemics species based on spatial scale (*e.g.*
32 narrow, regional and continental endemics), evolutionary history (*e.g.* neo and paleo
endemics) or habitat specificity (*e.g.* edaphic endemics; Ferreira & Boldrini, 2011;
34 Kruckeberg & Rabinowitz, 1985; Peterson & Watson, 1998). These proposals, however,
implicitly assume that all individuals of a species are confined to a given region or
36 habitat, also known as true or pure endemism (Tyler, 1996). If one record is found
outside the target region, the species is to be (re)classified as non-endemic. Since pure
38 endemism is rather strict, the term near-endemism has been used to describe species
with few records outside the target region (Carbutt & Edwards, 2006; Matthews, van
40 Wyk, & Bredenkamp, 1993; Noroozi et al., 2018; Platts et al., 2011). Near-endemics are
the result rare dispersal events, temporary establishment in different habitats or the
42 existence small satellite populations (Matthews, van Wyk, & Bredenkamp, 1993;
Perera, Ratnayake-Perera, & Procheş, 2011).

44 The differentiation between pure and near endemics is challenging, because it
may not be stable in time: near endemics can become pure endemics if habitat loss is
46 higher outside than inside the target region (Carbutt & Edwards, 2006). Conversely,
pure endemics may become near endemics with the accumulation of knowledge on their
48 geographical distribution (Werneck et al., 2011). This is particularly true for
geographically-restricted species, which often have scarce occurrence data.
50 Furthermore, pure endemics may be classified as near endemics due to species

misidentifications (Carbutt & Edwards, 2006) or by a questionable delimitation of the
52 target region (Platts et al., 2011). In practice, conservation aims at protecting as many
individuals as possible for a given species (IUCN, 2018). So, the differentiation
54 between pure and near endemism may have little impact to plan conservation actions.
Therefore, the practical question is: how to distinguish both groups of endemic species
56 from non-endemic species? Defining pure endemism is straightforward, but separating
near-endemics from non-endemic species can be quite subjective.

58 Here we establish objective limits to separate near-endemic from non-endemic
species for conservation purposes. We also separate widespread species from occasional
60 species, *i.e.*, species common in other regions but sporadic in the target region (Barlow
et al., 2010). We use a data-driven approach to estimate what ratio of occurrences inside
62 a target region could be used to separate species into pure-endemics, near-endemics,
widespread and occasional species. We perform this evaluation for over 5000 tree
64 species from the Atlantic Forest, a biodiversity hotspot with abundant knowledge on the
taxonomy and distribution of its flora. Using millions of carefully curated occurrences
66 from over 500 collections around the world, we evaluate which ratio of occurrences
inside the Atlantic Forest match species-specific endemism accepted by taxonomic
68 experts. Finally, we delimit centres of diversity for pure- endemics, near-endemics and
occasional species and discuss the implications for the conservation of this biodiversity
70 hotspot.

72 **2 METHODS**

2.1 Retrieval and validation of occurrence data

74 The Atlantic Forest covers Argentina, Brazil and Paraguay, so we searched for
occurrences using a list of tree names for South America (see Supporting Information

76 for more details), compiled from different sources (Grandtner & Chevrette, 2013; Lima
et al., 2015; Oliveira-Filho, 2010; ter Steege et al., 2016; Zappi et al., 2015; Zuloaga,
78 Morrone, & Belgrano, 2008). We carefully inspected the list of names to avoid the
inclusion of exotic and non-arborescent species. Arborescent species, hereafter ‘trees’,
80 were defined as species with free-standing stems that often exceed 5 cm of diameter at
breast height (1.3 m) or 4 m in total height, including arborescent palms, cactus, tree
82 ferns, and woody bamboos.

The list of South American tree names was used to download occurrence data
84 from *speciesLink* (www.splink.org.br), JABOT (<http://jabot.jbrj.gov.br>, Silva et al.,
2017), ‘Portal de Datos de Biodiversidad Argentina’ (<https://datos.sndb.mincyt.gob.ar>)
86 and the Global Biodiversity Information Facility (GBIF.org, 2019). We excluded all
occurrences described as being cultivated or exotic. We checked names for typos,
88 orthographical variants and synonyms in the Brazilian Flora 2020 (BF-2020) project
(Filardi et al., 2018; Zappi et al., 2015). Decisions for unresolved names were made by
90 consulting Tropicos (www.tropicos.org) or the World Checklist of Selected Plant
Families (<http://wcsp.science.kew.org>).

92 Using the 3.11 million records retrieved (Appendix S1), we conducted a detailed
data cleaning and validation procedure (see Supporting Information for details). We
94 standardized the notation of different fields (*e.g.* locality description, collector and
identifier names, collection and identification dates), which were then used to (i) search
96 for duplicate specimens among herbaria; (ii) validate the geographical coordinates at
country, state and/or county levels and (iii) to assess the confidence level of the
98 identification of each specimen (*i.e.* ‘validated’ and ‘probably validated’ - Appendix
S2). Moreover, (iv) we cross-validated information of duplicate specimens across
100 herbaria to obtain missing or more precise coordinates and/or valid identifications.

Finally, (iv) we removed specimens too distant from their core distributions (i.e. spatial
102 outliers).

104 **2.2 Endemism levels**

We obtained the endemism classification of each species from the BF-2020 (Filardi et
106 al., 2018), the best reference currently available for the Atlantic Forest flora. A species
was classified as ‘endemic’ if the BF-2020 field ‘phytogeographic domain’ contained
108 only the term ‘Atlantic Rainforest’. Correspondingly, a species was classified as
‘occasional’ if this field did not include this term. Species with no information on the
110 ‘phytogeographic domain’ were omitted from this analysis.

We calculated an empirical level of endemism based on the position of species
112 records in respect to the Atlantic Forest limits (IBGE, 2012; Olson & Dinerstein, 2002).
Each record was assigned as being inside, outside or in the transition of the Atlantic
114 Forest to other domains. Records in the transition were those falling inside the Atlantic
Forest limits, but in counties with less than 90% of its area inside the Atlantic Forest or
116 vice-versa. Because of the variable precision of the specimen’s coordinates and of the
arbitrariness of the domain delimitation at the scale of our reference map (1:5,000,000),
118 records in the transition received half the weight other records to calculated species
endemism levels:

$$120 \quad 100 \times (O_{in} + O_{ti}/2) / (O_{in} + O_{ti}/2 + O_{out} + O_{to}/2),$$

where, O_{in} , O_{ti} , O_{out} and O_{to} are the number of specimens inside, inside in the transition,
122 outside and outside in the transition to the Atlantic Forest, respectively. This endemism
level is actually a weighted proportion of occurrences inside the Atlantic Forest by the
124 total of valid occurrences found, varying from 0 (no occurrences) to 100% (all
occurrences inside the Atlantic Forest).

126 The comparison between the reference BF-2020 classification and the empirical
classification of species endemism was based on thresholds values varying from 0 to
128 100%, in intervals of 1% (*i.e.* 0, 1, ..., 99, 100%). If a given species had an observed
endemism level equal or higher than the threshold, it was classified as ‘endemic’. For
130 each threshold value, we calculated the number of mismatches between the two
classifications (*i.e.* species classified as ‘endemic’ in the BF-2020 and ‘not endemic’
132 from the observed endemism level or vice-versa). The same procedure was used to
calculate the number of mismatches for occasional species. We then plotted the number
134 of mismatches against all thresholds and estimated the optimum threshold that
minimizes the number of mismatches between classifications. Optimum thresholds were
136 estimated using piecewise regression, allowing up to five segments (*i.e.* four breaking
points). Thus, we provided the breaking point of each curve (and its 95% confidence
138 interval). We compared the results using only taxonomically ‘validated’ and using both
taxonomically ‘validated’ and ‘probably validated’ records.

140

2.3 Centres of diversity

142 We used the optimum threshold values obtained above to classify species into pure
endemics, near endemics and occasional species and to delimit their centres of diversity
144 (Laffan & Crisp, 2003). We plotted the valid occurrences of each group of species
against a 50×50 km grid covering the Atlantic Forest and surrounding domains. Next,
146 we obtained different diversity metrics for each group of species per grid cell. We
selected two metrics with best performance to describe our data (Figures S1 and S2):
148 corrected weighted endemism (WE) and rarefied/extrapolated richness (S_{RE}). The WE is
the species richness weighted by the inverse of the number of cells where the species is
150 present, divided by cell richness (Crisp et al., 2001). The S_{RE} is the rarefied/extrapolated

richness (depending on the observed number of occurrences per cell) for a common
152 number of 100 occurrences, calculated based on the species frequencies per cell (Chao
et al., 2014). We also obtained the sample coverage estimate (Chao & Jost, 2012), used
154 here as a proxy of sample completeness. We evaluated the relationship of the diversity
of endemic and occasional species with overall species diversity using spatial regression
156 models (*i.e.* linear regression with spatially correlated errors - Pinheiro & Bates, 2000).
Centres of diversity were delimited using ordinary kriging and only the grid cells
158 meeting some minimum criteria of sampling coverage (see Supporting Information).
We used the 80% quantile of predicted diversity distributions to delimit the centres of
160 endemism.

162 **3 RESULTS**

3.1 Number of species found for the Atlantic Forest

164 After the removal of duplicates, spatial outliers and the geographical and taxonomic
validation, we retained 593,920 valid records (disregarding records with ‘probably
166 validated’ taxonomy). We found 252,911 valid records being collected inside the
Atlantic Forest limits, which contained a total of 5062 arborescent species (4057 species
168 excluding tall shrubs; Appendix S3). Most species-rich families were Myrtaceae (681),
Fabaceae (658 species), Rubiaceae (328), Melastomataceae (290) and Lauraceae (222).
170 If we consider the valid occurrences in the transitions of the Atlantic Forest to other
domains, we could add 294 species as probably occurring in the Atlantic Forest
172 (Appendix S4). Another 3148 names were retrieved but were finally excluded from the
list for different reasons (*e.g.* synonyms, typos, orthographical variants, etc; Appendix
174 S5).

176 **3.2 Endemism levels**

We found evidence of pure endemism (i.e. endemism level= 100%) for 1548 tree
178 species (31%; Appendix S6). We found that 90.2% of records inside the Atlantic Forest
(95% Confidence Interval, CI: 89.3–91.2%) was the threshold that best matched the
180 endemism accepted by taxonomy experts (Figure 1a). The curve of mismatches between
observed and reference classifications decreases until it reaches a minimum and then it
182 increases again, meaning that more or less restrictive thresholds increase the number of
mismatches. The 90.2% threshold in the Atlantic Forest added 733 near endemic
184 species (15%). Together, pure and near endemics lead to an overall endemism ratio of
45.1% for the Atlantic Forest arborescent flora (Figure 1b) and 1.01 endemic
186 arborescent species per 100 km² of remaining forest (i.e. 2261.2 km²; Fundación Vida
Silvestre Argentina & WWF, 2017). Some families had high average endemism level,
188 namely Monimiaceae (94%), Symplocaceae (85%), Poaceae: Bambusoideae (84.2%),
Araliaceae (84%), Myrtaceae (80%), Cyatheaceae (80%), Proteaceae (79%),
190 Aquifoliaceae (77%) and Lauraceae (76%). Conversely, we found that 8.7% (95% CI:
8.2–9.3%) was the best threshold for separating occasional from more common Atlantic
192 Forest species (Figure 1a), leading to a total of 646 occasional species (13%). The
remaining 42% of the species were classified as widespread (Appendix S6). Results
194 using only occurrences with taxonomy flagged as ‘validated’ were similar (pure
endemism: 32%; near endemism: 15%; occasional species: 14%, widespread species:
196 39% - Figure S3, Appendix S6).

198 **3.3 Centres of diversity**

The diversity of endemic species was strongly correlated with the overall species
200 diversity in the Atlantic Forest (Figure 2). There was also a strong and positive

correlation between the number of pure and near endemic species (Figure S4), meaning
202 that the centres of diversity of pure and near endemics are highly congruent in space.
The combination of pure and near endemics resulted in the same pattern of high
204 diversity in the rainforests along the coast (Figure 3), corresponding to the Serra do Mar
and Bahia Coastal Forests ecoregions (Olson & Dinerstein, 2002). Occasional species in
206 the Atlantic Forest were really rare in the colder Araucaria region. Most of the
distribution of occasional species was concentrated in the Brazilian Cerrado, but also in
208 the Amazon and slightly less in the Caatinga domain. General patterns were fairly
similar when using other diversity measures (Figures S5-S7).

210

4 DISCUSSION

212 Near endemism has been used to assess endemism levels of regional floras and faunas.
However, such assessments often use loose (Carbutt & Edwards, 2006; Platts et al.,
214 2011) or arbitrary definitions (Noroozi et al., 2018; Perera et al., 2011) of near
endemics. Here, we used a data-driven approach to find that 90% of the occurrences
216 inside a target region can be used to tell apart endemics from non-endemic trees. This is
the same limit used to detect plant endemism in the Mediterranean Basin hotspot
218 (Médail & Baumel, 2018), suggesting that a 90% limit could be used in assessment of
plant endemism of other species-rich regions. This limit has another important
220 implication: the average endemism concept adopted by taxonomic experts for Atlantic
Forest trees implicitly includes the concept of near endemism. Indeed, the overall
222 endemism ratio found here for pure and near endemics combined (45%) is within the
range of 40-50% endemism level previously reported for the Atlantic Forest flora
224 (Myers et al., 2000; Stehmann et al., 2009; Zappi et al., 2015). Thus, we propose that
pure and near endemics can be used together to objectively delimit endemism or as two

226 categories of endemism, similarly to what already exists for the categories of species
threat (IUCN, 2018).

228 The Atlantic Forest is arguably the tropical forest with the largest botanical
knowledge available, with ca. 680 thousand unique specimens of tree species or 42 per
230 100 km² – average collection density in the Amazon forest is below 10 per 100 km² (ter
Steege et al., 2016). Nevertheless, here we added 714 new valid occurrences of tree
232 species for this biodiversity hotspot, an increase of 21% to the 3343 trees previously
reported by the Brazilian Flora (Zappi et al., 2015). As expected, about 47% of these
234 new records corresponded to occasional species. These species corresponded to 13% of
the total richness of the Atlantic Forest tree flora, confirming that occasional species,
236 despite of their rarity, make an important contribution to overall biodiversity patterns
(Barlow et al., 2010; ter Steege et al., 2019). More importantly, 53% of the new records
238 corresponded to widespread and endemic species. An increase of 16% in the total
richness was also observed for the Espírito Santo state flora compared to the reported in
240 the Brazilian Flora (Dutra, Alves-Araújo, & Carrijo, 2015). These results highlight how
data-driven approaches combined with careful validation steps can help to refine the
242 knowledge of local and regional floras. The Brazilian Flora is permanently being
improved and it already is of utmost importance for the understanding of the Brazilian
244 flora (Zappi et al., 2015), the richest in the world (Ulloa et al., 2017). Here, we provide
products that can be readily integrated into the Brazilian Flora project (*e.g.* more refined
246 endemism filters) and a workflow to perform similar assessments in other regions or for
other groups of organisms.

248 The delimitation of centres of endemic diversity provided here (Figure 3,
Appendix S7) has direct implications for conservation planning. For instance, they can
250 assist the identification of Important Plant Areas (IPA - www.plantlifeipa.org), provided

by the Target 5 of the Global Strategy for Plant Conservation (www.cbd.int/gspc).

252 Although the delimitation of IPAs predicts the use of endemic species, their definition is
mainly based on the presence of threatened species. Moreover, since many of the
254 endemic species identified here are probably also threatened, the Brazilian Alliance for
Extinction Zero (www.biodiversitas.org.br/baze) could incorporate the concept of
256 endemism proposed here to select plant trigger-species and to design conservation.
Considering that defining threatened and endemic species have the same constraints
258 related to data availability and to the time and spatial scale considered (Ferreira &
Boldrini, 2011), the detection of endemics is more straightforward than threatened
260 species, which could speed up the decision-making process for conservation. Therefore,
the objective detection of endemic species proposed here (Appendix S6) could help to
262 bridge the scarcity of conservation policies based on endemic species.

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270 are given in Appendix S1.

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404 **AUTHOR CONTRIBUTIONS SUPPORTING INFORMATION**

RAFL developed the idea of this study, conducted data compilation, analysis and
406 drafted the paper. HTS assisted with herbarium data editing and study design. RAFL
Geographical and taxonomical validation of the records were supported by MFS and
408 VCS, respectively. All authors contributed to the interpretation and discussion of the
results and writing the final manuscript.

410

SUPPORTING INFORMATION

412 **Supporting Methods and References**

Supplementary Figures

414

LIST OF APPENDICES

416 (Note: Full Appendices will be provided after the publication of the manuscript)

418 **Appendix S1:** List of collections and data providers used for data compilation.

The numbers of records retrieved per collection correspond to overall sum of records
420 before data validation, thus including both valid and invalid records.

422 **Appendix S2:** List of names of taxonomists per family used for taxonomical validation.

The ‘tdwg.name’ represents the taxonomist name following the standard notation of the
424 Biodiversity Information Standards (<https://www.tdwg.org>), which includes different
variants of notation found for the same taxonomist name.

426

Appendix S3: Updated, taxonomically vetted checklist of the Atlantic Forest tree flora.

428 For each name included in the checklist we provide the life form, the status of the name
in respect to the Brazilian Flora 2020 project, the number of records found inside the
430 Atlantic Forest (both ‘validated’ and ‘probably validated’ taxonomy) and a list of up to
30 vouchers (only specimens with ‘validated’ taxonomy), giving priority to type
432 specimens. We also indicate which species were regarded as being taxa of low
taxonomic complexity (TBC) or taxa commonly cultivated outside its original range.

434

Appendix S4: List of species with probable occurrence in the Atlantic Forest.

436 We present all names with valid records found only in the transition of the Atlantic
Forest to other domains and those names cited in the Brazilian Flora 2020 project as
438 being an Atlantic Forest species, but for which we did not find any valid records. Again,
we present for each name the life form, the number of records found and a list of up to
440 30 vouchers.

442 **Appendix S5:** List of names excluded from the final Atlantic Forest checklist.

For each name on the list we provide the life form and the reason why the name was
444 excluded. For synonyms, orthographical variants, common typos we also provide the
corresponding valid name used in this study.

446

Appendix S6: Endemism levels for the Atlantic Forest tree flora and the corresponding

448 classification into pure endemic, near endemic, widespread and occasional species.

For each species name, we provide the number of valid records outside the Atlantic

450 Forest, outside but in the transition to the Atlantic Forest, inside the Atlantic Forest but

in the transition to other domains, and inside the Atlantic Forest. We present the

452 endemism levels and species classifications using only records with validated taxonomy

and using records with validated and probably validated taxonomy. Finally, we present

454 the endemism classification currently accepted in the Brazilian Flora 2020 in respect to

the Atlantic Forest.

456

Appendix S7: Shapefiles delimiting the centres of the endemic and occasional species

458 diversity in the Atlantic Forest for pure endemics, near endemics, pure + near endemics

and occasional species.

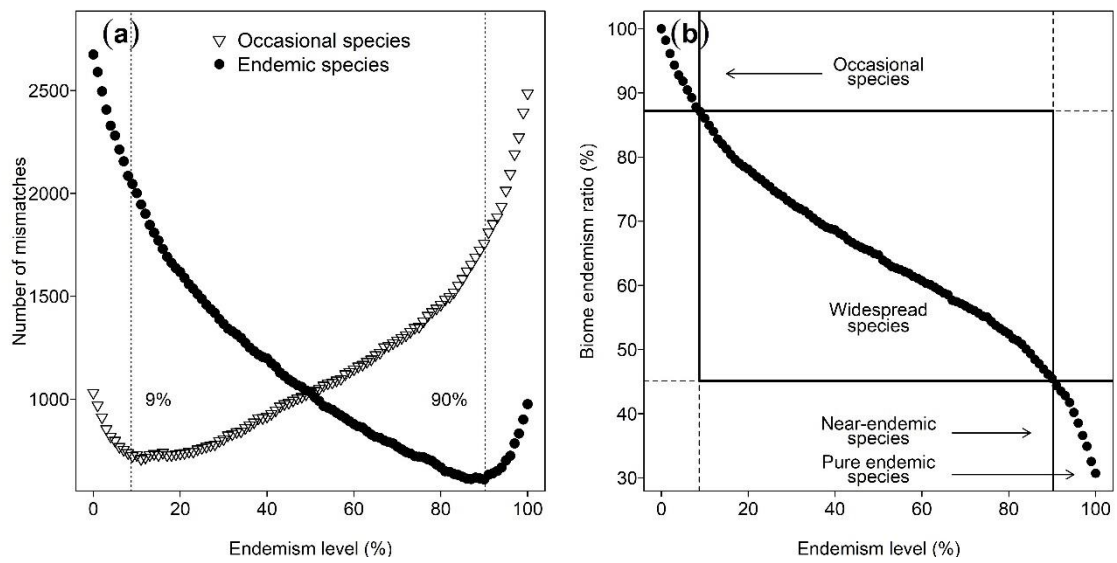
460 Each shapefile contains the isoclines corresponding to the 75%, 80%, 85%, 90% and

95% quantiles of the distribution of rarefied/extrapolated richness for 100 specimens,

462 predicted using ordinary kriging.

464

Figures



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472

Figure 1. Defining near endemic and occasional tree species using herbarium records for the Atlantic Forest biodiversity hotspot. For both endemic (black circles) and occasional species (triangles), we present (a) the optimum endemism levels (vertical dashed lines) estimated from the distribution of mismatches between the empirical and the Brazilian Flora 2020 classifications and (b) the overall endemism ratio of the Atlantic Forest in intervals of 1% (x -axis in both panels).

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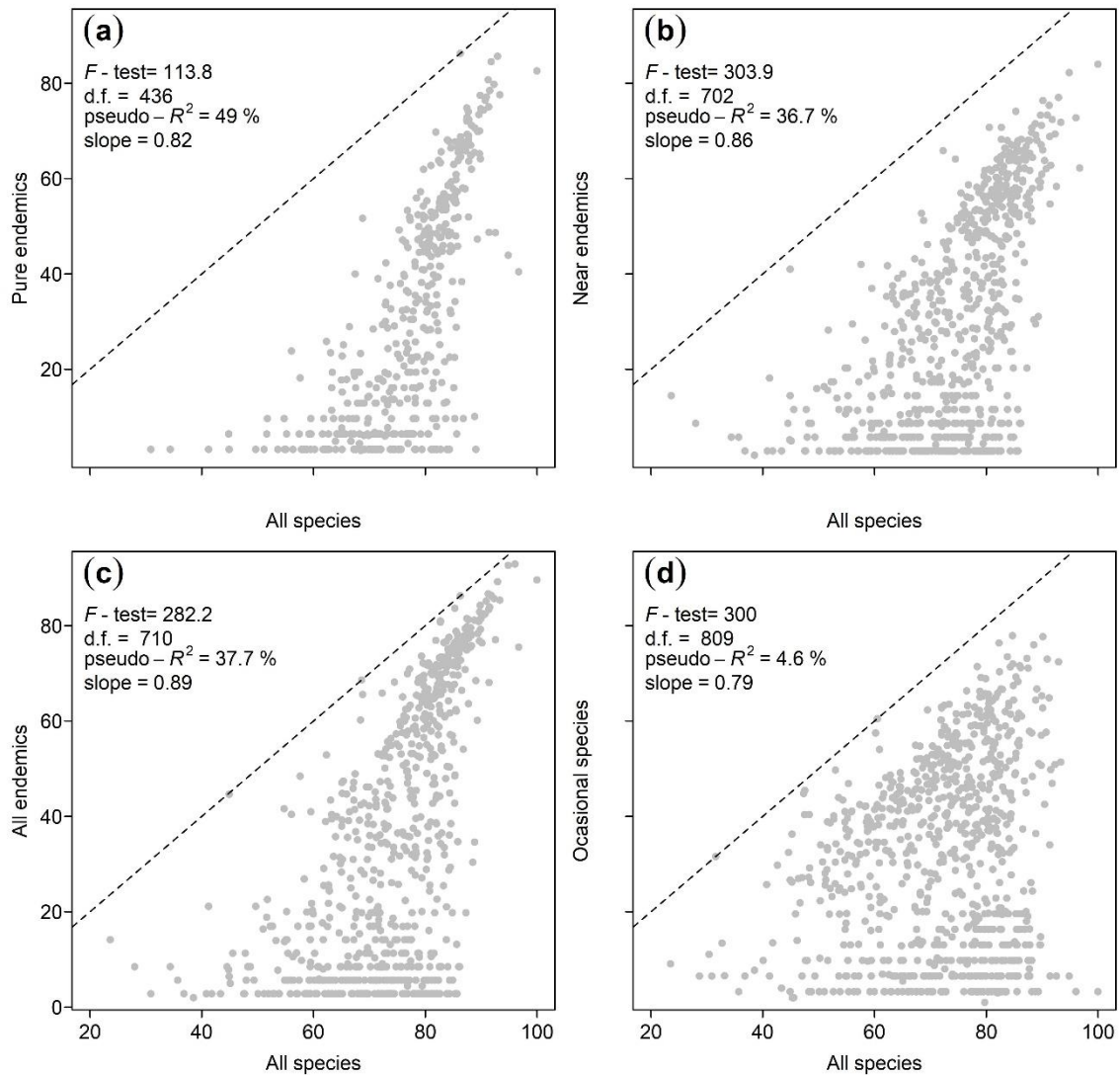
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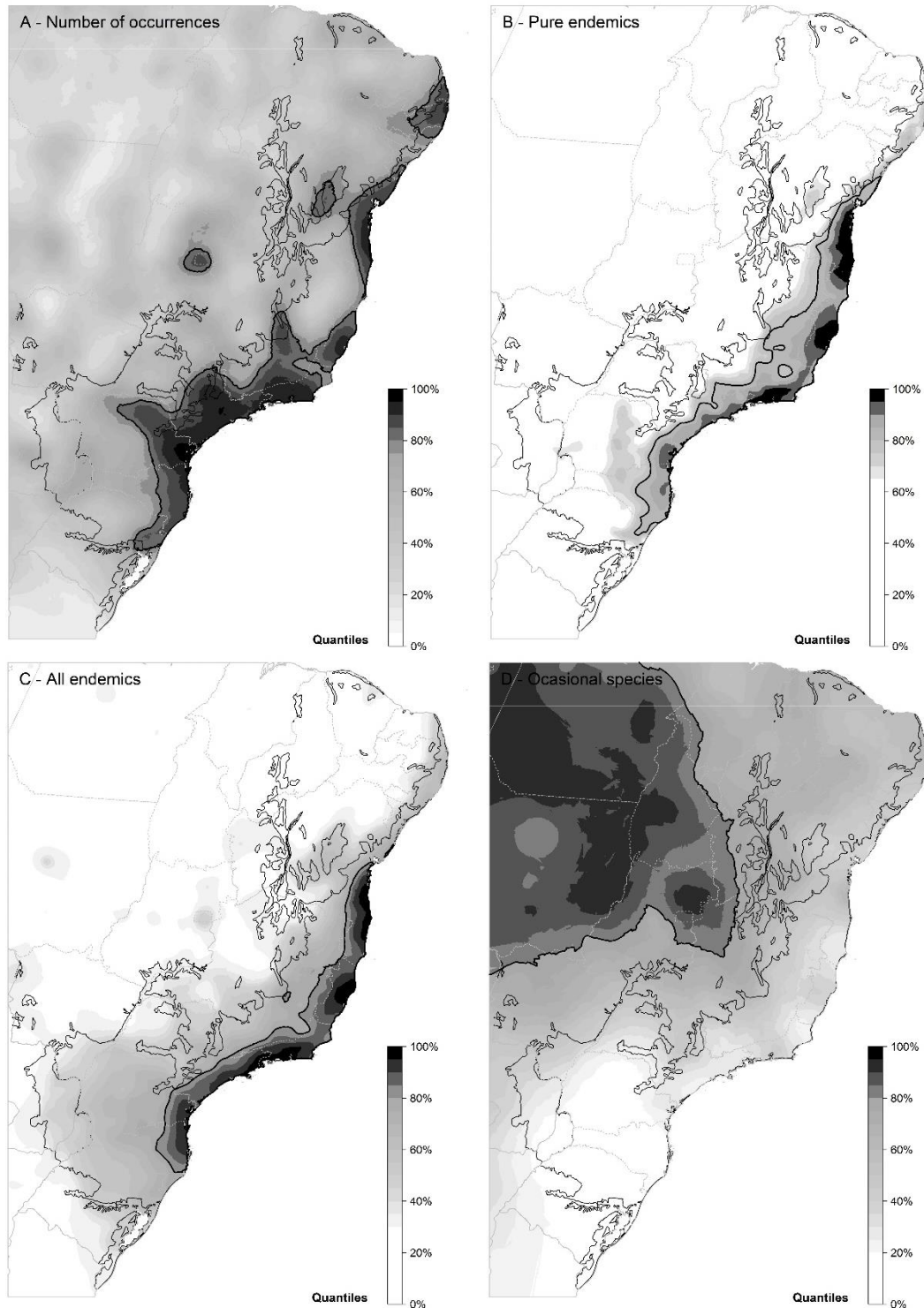
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Figure 2. Relationship between the number of rarefied/extrapolated richness per 50×50 km grid cell and the same diversity metric obtained for (a) pure endemics, (b) near endemics, (c) all endemics (pure + near endemics) and (d) occasional species. For each group of species, we present the summary statistics of each spatial regression model (top left; d.f.= degrees of freedom), including the predicted slope of the regression prediction. The spatial regression analysis was performed only for grid cells meeting some minimum criteria of sampling coverage (see Supplementary Methods). The dashed line represents the 1:1 line.

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502



504 **Figure 3.** The spatial distribution of (A) the number of occurrences retrieved for the species occurring in the Atlantic Forest, and the centres of diversity of (B) pure
506 endemics, (C) all endemics (pure + near) and (D) occasional species. Maps were
508 produced using ordinary kriging based on rarefied/extrapolated species richness
510 obtained for a common number of 100 records per grid cell. The colour scale represents
512 the 5% quantiles of the metrics distribution, from 0-5% (white) to 95-100% (black).
Bold black lines are the area containing the 80% higher richness values. The black line
marks the limits of the Atlantic Forest, while the solid and dashed grey lines mark the
limits of South American countries and of the Brazilian states, respectively.