Title:

DISPERSE: A trait database to assess the dispersal potential of aquatic macroinvertebrates

Authors:

Romain Sarremejane[1]*, Núria Cid[2,3]*, Thibault Datry[2], Rachel Stubbington[1], Maria Alp[2], Miguel Cañedo-Argüelles[3], Adolfo Cordero-Rivera[4], Zoltan Csabai[5,6], Cayetano Gutiérrez-Cánovas[7,8], Jani Heino[9], Maxence Forcellini[2], Andrés Millán[10], Amael Paillex[11], Petr Pařil[6], Marek Polášek[5], José Manuel Tierno de Figueroa[12], Philippe Usseglio-Polatera[13], Carmen Zamora-Muñoz[12] & Núria Bonada[3]

[1] School of Science and Technology, Nottingham Trent University, Nottingham NG11 8NS, UK

[2] INRAE, UR RiverLy, centre de Lyon-Villeurbanne, 5 rue de la Doua CS70077, 69626 Villeurbanne Cedex, France

[3] Grup de Recerca Freshwater Ecology, Hydrology and Management (FEHM), Departament de Biologia Evolutiva, Ecologia i Ciències Ambientals, Facultat de Biologia, Institut de Recerca de la Biodiversitat (IRBio), Universitat de Barcelona (UB), Diagonal 643, 08028 Barcelona, Catalonia, Spain

[4] ECOEVO Lab, E.E. Forestal, Univesidade de Vigo, Campus A Xunqueira, 36005 Pontevedra, Spain

[5] Department of Hydrobiology, University of Pécs, Ifjúság útja 6, H7624 Pécs, Hungary

[6] Department of Botany and Zoology, Faculty of Science, Masaryk University, Kotlářská 2, 61137 Brno, Czech Republic

[7] Centre of Molecular and Environmental Biology (CBMA), Department of Biology,

University of Minho, Braga, Portugal

[8] Institute of Science and Innovation for Bio-Sustainability (IB-S), University of Minho,

Braga, Portugal

[9] Finnish Environment Institute, Freshwater Centre, Paavo Havaksen Tie 3, FI-90570 Oulu,

Finland

[10] Department of Ecology and Hydrology, Biology Faculty, Murcia University. Campus de

Espinardo, 30100. Murcia, Spain

[11] Department of Aquatic Ecology, Eawag, Swiss Federal Institute of Aquatic Sciences,

Überlandstrasse 133, CH-8600 Dübendorf, Switzerland

[12] Departamento de Zoología, Facultad de Ciencias, Universidad de Granada, Avenida Fuente

Nueva, s/n, 18071 Granada, Spain

[13] Université de Lorraine, CNRS, UMR 7360, LIEC, Laboratoire Interdisciplinaire des

Environnements Continentaux, F-57070 Metz, France

*These authors contributed equally to this work.

BIOSKETCH

Many of the authors are members of the European Cooperation in Science and Technology

(COST) Action CA15113 Science and Management of Intermittent Rivers and Ephemeral

Streams (SMIRES, www.smires.eu), which aims to improve our understanding of intermittent

rivers and ephemeral streams (IRES) and translate this into science-based, sustainable

management of river networks. One of the main goals of SMIRES is to facilitate sharing of

data and experience, bringing together researchers from many disciplines and stakeholders

and creating synergies through networking. The dataset provided and analyzed herein reflects

networking and collaboration within the SMIRES working group dedicated to Community

Ecology and Biomonitoring.

Abstract

Motivation: Dispersal is an essential process in population and community dynamics but is

difficult to measure in the field. In freshwater systems, relevant information on the dispersal

of many taxa remains scattered or unpublished, and biological traits related to organisms'

morphology, life history and behaviour offer useful dispersal proxies. We compiled

information on selected dispersal-related biological traits of European aquatic

macroinvertebrates in a unique source: the DISPERSE database. Information within

DISPERSE can be used to address fundamental questions in metapopulation ecology,

metacommunity ecology, macroecology, and eco-evolutionary research. From an applied

perspective, the consideration of dispersal proxies can also improve predictions of ecological

responses to global change, and contribute to more effective biomonitoring and conservation

management strategies.

Main types of variable contained: Selected macroinvertebrate dispersal-related biological

traits: maximum body size, adult female wing length, adult wing pair type, adult life span,

number of reproductive cycles per year, lifelong fecundity, dispersal mode (i.e. active,

passive, aquatic, aerial), and propensity to drift.

Spatial location: Europe.

Time period: Taxa considered are extant with recent records. Most data were collected in the

20th century and in the period 2000-2019.

Major taxa and level of measurement: Aquatic macroinvertebrates from riverine and

lacustrine ecosystems, including Mollusca, Annelida, Platyhelminthes, and Arthropoda such

as Crustacea and Insecta. For Insecta, aquatic stages and the aerial (i.e. flying) phases of

adults were considered separately. Genus-level taxonomic resolution used, except for some

Annelida and Diptera, which are coded at the sub-class, family or sub-family level. In total,

the database includes 39 trait categories grouped into nine dispersal-related traits for 480 taxa.

Software format: The data file is in Excel workbook.

KEYWORDS

Aquatic invertebrates, freshwater biodiversity, functional traits, insects, metacommunities,

lakes, ponds, proxies, rivers, streams

1. INTRODUCTION

Dispersal is a fundamental ecological process that affects the organization of biological diversity at different temporal and spatial scales (Bohonak & Jenkins, 2003; Clobert, Baguette, Benton, & Bullock, 2012). Dispersal strongly influences metapopulation and metacommunity dynamics through the movement of individuals and species, respectively (Heino et al., 2015). A better understanding of dispersal processes can inform biodiversity management practices (Barton et al., 2015; Heino et al., 2017). However, dispersal itself is difficult to measure, particularly for small organisms, including most invertebrates. Typically, dispersal is measured for single species (e.g. Macneale, Peckarsky, & Likens, 2005; Troast, Suhling, Jinguji, Sahlén, & Ware, 2016) or combinations of few species within one taxonomic group (Arribas et al., 2012; French & McCauley, 2019; Lancaster & Downes, 2017) using methods based on mark and recapture, stable isotopes, or population genetics (Heino et al., 2017; Lancaster & Downes, 2013). Such methods can directly assess dispersal events but are expensive, time-consuming, and thus impractical for studies conducted at the community level or at large spatial scales. In this context, taxon-specific biological traits represent a cost-effective alternative that may serve as proxies for dispersal (Heino et al., 2017; Outomuro & Johansson, 2019; Stevens et al., 2013). These traits interact with landscape structure to determine patterns of effective dispersal (Brown & Swan, 2010; Tonkin et al., 2018).

Aquatic macroinvertebrates inhabiting riverine and lacustrine ecosystems include taxa with diverse dispersal modes and abilities. Moreover, in species with complex life cycles, such as some insects, this diversity is enhanced by distinct differences among life stages. For example, juveniles of many aquatic insects disperse actively or passively in water whereas adults fly over long distances (Wikelski et al., 2006). In all cases, dispersal is affected by multiple traits relating to the morphology (Lancaster & Downes, 2013), life history and behaviour (Clobert et al., 2012) of multiple life stages.

We compiled and harmonised information on dispersal-related traits of aquatic macroinvertebrates across Europe, including both aquatic and aerial (i.e. flying) stages. Although some information on dispersal-related macroinvertebrate traits such as body size, reproduction, locomotion and dispersal mode is available in online databases for European (https://www.freshwaterecology.info/; Schmidt-Kloiber & Hering, 2015; Serra, Cobo, Graça, Dolédec, & Feio, 2016; Tachet, Richoux, Bournaud, & Usseglio-Polatera, 2010) and North American taxa (Vieira et al., 2006), other relevant information is scattered across published literature and unpublished data. Using the expertise of 19 European collaborators, we built a comprehensive database containing nine dispersal-related traits divided into 39 trait categories for 480 taxa. Dispersal-related traits were selected and their trait categories (Table 1) were fuzzy-coded (Chevenet, Dolédec, & Chessel, 1994), following a structure comparable to existing databases (Schmera, Podani, Heino, Erös, & Poff, 2015). Our aim was to provide a single resource facilitating the inclusion of dispersal in ecological, biogeographical and environmental change research, and providing a basis for a global dispersal database.

2. METHODS

Dispersal-related trait selection criteria

We defined dispersal as the unidirectional movement of individuals from one location to another (Bohonak & Jenkins, 2003), assuming that the population-level dispersal rate depends on both the number of dispersers/propagules and dispersers' ability to move across the landscape (Lancaster & Downes, 2017; Lancaster, Downes, & Arnold, 2011).

We selected nine dispersal-related morphological, behavioural and life-history traits. Selected morphological traits were maximum body size, female wing length and wing pair type, the latter two being relevant for flying adult stages of insects. Body size influences organisms' dispersal, especially for active dispersers (Jenkins et al., 2007), with larger animals more

capable of active dispersal over longer distances (e.g. flying adult odonates). In addition, for

aquatic stages, body size can be indicative of drift: larger animals drift less (Wilzbach &

Cummins, 1989). For flying adults, female wing length was selected because females connect

and sustain populations through oviposition, representing adult insects' flying and

recolonisation capacity (Graham, Storey, & Smith, 2017): females with larger wings are likely

to oviposit farther from their source population (Arribas et al., 2012; Hoffsten, 2004; Rundle,

Bilton, & Foggo, 2007). Selected life-history traits were adult life span, life-cycle duration,

annual number of reproductive cycles and lifelong fecundity. Adult life span and life-cycle

duration respectively reflect the adult (i.e. reproductive) and total life duration, with longer-

lived animals typically having more dispersal opportunities (Stevens et al., 2013). The annual

number of reproductive cycles and lifelong fecundity were chosen to assess indirect dispersal

capacity, i.e. potential propagule production, with multiple reproductive cycles and abundant

eggs increasing the expected number of dispersal events. Dispersal behaviour was represented

by strategies describing a taxon's predominant dispersal mode (passive, active, aquatic,

aerial), and by propensity to drift, which indicates the frequency of flow-mediated passive

downstream.

Fuzzy-coding approach and taxonomic resolution

Traits were mostly coded at genus level, but some Diptera and Annelida were coded at family

or sub-family level because of their complex taxonomy, identification difficulties and the

scarcity of reliable information about their traits. Traits were coded using a fuzzy approach in

which a value given to each trait category indicates if the taxon has weak (1), moderate (2) or

strong (3) affinity with the category (e.g. Chevenet et al., 1994). This method can incorporate

intra-taxonomic variability when trait profiles differ among e.g. species within a genus or

within one species in different environments (Bonada & Dolédec, 2018).

Data acquisition, compilation and quality control

Trait information was sourced primarily from the literature by Google Scholar searches using

keywords including trait names, synonyms and taxon names, and by searching in existing

databases (i.e. https://www.freshwaterecology.info/; Vieira et al., 2006). Altogether, >300

peer-reviewed articles and book chapters were consulted (Table 1). When published

information was lacking, traits were coded based on authors' expert knowledge and direct

measurements (Table 1). When no European studies were available, we considered

information from North America (e.g. Vieira et al., 2006), if experts evaluated traits as

comparable across regions (e.g. non-labile traits such as wing pair type do not vary

geographically , whereas life-cycle durations may vary across regions). Most traits were

coded using multiple sources representing multiple species within a genus. When only one

study was available, we supplemented this information with expert knowledge to ensure that

trait codes represented the potential variability in the taxon (Table 1).

TABLE 1   Selected dispersal-related aquatic macroinvertebrate traits and the relative

contribution of different sources of information (i.e. literature, expert knowledge, direct

measurement) used to build the database.

Trait	Categories	Taxa completed (%)	Source of information (%)		
			Literature	Expert	Measured
Maximum body size (cm)	< 0.25 ≥ 0.25 – 0.5 ≥ 0.5 – 1 ≥ 1 – 2 ≥ 2 – 4 ≥ 4 – 8 ≥ 8	99	100		
Female wing length (insects only) (mm)	< 5 ≥ 5 – 10 ≥ 10 – 15 ≥ 15 – 20	95	57	12	31

Trait	Categories	Taxa completed (%)	Source of information (%)		
			Literature	Expert	Measured
	≥ 20 – 30				
	≥ 30 – 40				
	≥ 40 – 50				
	≥ 50				
Wing pair type (insects only)	1 pair + halters 1 pair + elytra or hemielytra 1 pair + small hind wings 2 similar-sized pairs	100	100		
Life cycle duration	≤ 1 year > 1 year	98	100		
Adult life span	< 1 week ≥ 1 week – 1 month ≥ 1 month – 1 year ≥ 1 year	79	65	35	
Lifelong fecundity	< 100 eggs ≥ 100 – 1000 eggs ≥ 1000 – 3000 eggs ≥ 3000 eggs	75	77	23	
Potential number of reproductive cycles per year	< 1 1 >1	98	100		
Dispersal strategy	Aquatic active Aquatic passive Aerial active Aerial passive	98	100		
Propensity to drift	Rare/catastrophic Occasional Frequent	80	90	10	

3. RESULTS: DESCRIPTION OF THE DATABASE

Sixty-one percent of taxa have complete trait information, with 1-2 traits being incomplete for

most remaining taxa (Table 1, Fig. 1). Most of the trait data (88%) originated from published

literature and the rest was coded based on expert knowledge and direct measurements of

Coleoptera and Heteroptera female wing length (Table 1).

FIGURE 1 Total number of taxa and % of the nine traits completed in each insect order

and invertebrate phylum, sub-phylum, class or sub-class. "Other" includes Hydrozoa,

Hymenoptera, Megaloptera and Porifera, for which only one taxon is in the database.

Using insects as an example, we performed a fuzzy correspondence analysis (FCA,

Chevenet et al., 1994) to visualise and compare differences in trait composition among taxa

(Figure 2). Insect groups were clearly divided based on dispersal-related traits, with 32% of

the variation explained by the first two FCA axes. Wing pair type and lifelong fecundity had

the highest correlation ratios (0.87 and 0.63, respectively) with axis A1, reflecting their clear

arrangement along the axis. Wing length (0.73) and maximum size (0.55) best correlated with

axis A2 (Figures 2 and S1). For example, Coleoptera typically produce few eggs and have

intermediate body sizes and wing lengths, Odonata produce an intermediate number of eggs

and have long wings, and Ephemeroptera produce many eggs and have short wings (Figure

2).

FIGURE 2   Variability in the dispersal-related trait composition of all insect taxa with

complete trait profile along fuzzy correspondence analysis axes A1 and A2. Dots indicate taxa

and lines converge to the centroid of each taxonomic group/order to depict within-group

dispersion.

4. DISCUSSION AND FUTURE PERSPECTIVES

DISPERSE is the first publicly available database describing the dispersal traits of aquatic macroinvertebrates including information on both aquatic and aerial (i.e. flying) life stages. It will promote incorporation of dispersal proxies into fundamental and applied population and community ecology in aquatic ecosystems (Heino et al., 2017). In particular, metacommunity ecology may benefit from the use of dispersal traits (Jacobson & Peres-Neto, 2010; Tonkin et al., 2018), by enabling classification of taxa according to their dispersal potential in more detail than is currently possible. Such classification, used in combination with, for example, spatial distance measurements (Datry, Moya, Zubieta, & Oberdorff, 2016; Sarremejane et al., 2017), may improve understanding of the effects of regional dispersal processes on community assembly and biodiversity patterns. Improved knowledge of taxon-specific dispersal abilities may also inform the design of more effective management practices. For example, recognising dispersal abilities in biomonitoring methods could inform enhanced catchment-scale management strategies that support ecosystems adapting to global change (Cid et al., under review; Datry, Bonada, & Heino, 2016). DISPERSE could also inform conservation strategies by establishing different priorities depending on organisms' dispersal capacities and spatial connectivity (Hermoso, Cattarino, Kennard, Watts, & Linke, 2015). This is especially relevant for highly dynamic freshwater ecosystems, such as temporary rivers and ponds, which repeatedly lose connectivity due to drying, and in which dispersal facilitates populations and community re-establishment after rewetting (Datry, Bonada, et al., 2016).

DISPERSE could also improve species distribution models (SDMs), in which dispersal is rarely considered due to a lack of data (Stevens et al., 2013), limiting the accuracy of model predictions (Thuiller et al., 2013). Recent trait-based approaches have integrated dispersal into SDMs, and DISPERSE could increase model accuracy (Cooper & Soberón,

2018; Willis et al., 2015). Including dispersal in SDMs is especially relevant to assessments of

biodiversity loss and species vulnerability to climate change (Bush & Hoskins, 2017;

Markovic et al., 2014; Willis et al., 2015). DISPERSE could also advance understanding of

eco-evolutionary relationships and biogeographical phenomena. In the context of evolution,

groups with lower dispersal abilities should be genetically and taxonomically richer due to

long-term isolation (Bohonak, 1999; Dijkstra, Monaghan, & Pauls, 2014). From a

biogeographical perspective, regions affected by glaciations should have species with greater

dispersal abilities, enabling postglacial recolonisation (Múrria et al., 2017).

By capturing different dispersal-related biological traits, DIPERSE provides

information on organisms' potential ability to move between localities as well as on

recruitment and reproduction (Tonkin et al., 2018). As an indirect measure of potential but not

actual dispersal, traits also facilitate comparison of taxa with different strategies, which could

inform studies conducted at large spatial scales, independent of taxonomy (Statzner & Bêche,

2010).

The diverse sources used herein complement existing trait databases by providing new

information, most of which would otherwise not be accessible to the scientific community.

DISPERSE offers a good coverage of macroinvertebrate groups at the genus level, which is

generally considered as adequate to capture biodiversity dynamics (Cañedo-Argüelles et al.,

2015; Datry, Melo, et al., 2016; Swan & Brown, 2017; Sarremejane et al., 2017). The

database currently represents a Europe-wide resource, which can be updated and expanded as

new information becomes available, to include more taxa and traits from across and beyond

Europe. DISPERSE lays the foundations for a global trait dispersal database on aquatic taxa,

as others have suggested (Maasri, 2019). As freshwater biodiversity declines at unprecedented

rates (Reid et al., 2019; Strayer & Dudgeon, 2010), collecting, harmonising and sharing

dispersal-related data on freshwater organisms will underpin evidence-informed initiatives
that seek to protect it.

5. ACKNOWLEDGMENTS

The study was supported by the COST Action CA15113 (SMIRES, Science and Management of Intermittent Rivers and Ephemeral Streams, www.smires.eu), funded by COST (European Cooperation in Science and Technology). NC was supported by the French research program Make Our Planet Great Again (MOPGA). MC was supported by the MECODISPER project (CTM2017-89295-P) funded by the Spanish Ministerio de Economía, Industria y Competitividad (MINECO) - Agencia Estatal de Investigación (AEI) and cofunded by the European Regional Development Fund (ERDF). ACR acknowledges funding by MINECO-AEI-ERDF (CGL2014-53140-P). CG-C is supported by the European Regional Development Fund (COMPETE2020 and PT2020) and the Portuguese Foundation for Science and Technology (FCT), through the CBMA strategic program UID/BIA/04050/2019 (POCI-01-0145-FEDER-007569) and the STREAMECO project (Biodiversity and ecosystem functioning under climate change: from the gene to the stream, PTDC/CTA-AMB/31245/2017). PP and MP were supported by project Grant Agency of the Czech Republic (project no. 20-17305S). ZC was supported by the projects no. EFOP-3.6.1.-16-2016-00004, 20765-3/2018/FEKUTSTRAT and TUDFO/47138/2019-ITM. PP and MP were supported by the Czech Science Foundation (P505-20-17305S).

6. DATA ACCESSIBILITY

DISPERSE is accessible on the Intermittent River Biodiversity Analysis and Synthesis webpage http://irbas.inrae.fr/ for download (as Excel files).

The database is publicly available under a Creative Commons Attribution 4.0 International copyright (CC BY 4.0) and should be appropriately referenced by citing this paper.

7. REFERENCES

Arribas, P., Velasco, J., Abellán, P., Sánchez-Fernández, D., Andújar, C., Calosi, P., … Bilton, D. T. (2012). Dispersal ability rather than ecological tolerance drives differences in range size between lentic and lotic water beetles (Coleoptera: Hydrophilidae). *Journal of Biogeography*, *39*, 984–994. https://doi.org/10.1111/j.1365-2699.2011.02641.x

Barton, P. S., Lentini, P. E., Alacs, E., Bau, S., Buckley, Y. M., Burns, E. L., … Smith, A. L. (2015). Guidelines for using movement science to inform biodiversity policy. *Environmental Management*, *56*, 791–801. https://doi.org/10.1007/s00267-015-0570-5

Bohonak, A. J. (1999). Dispersal, gene flow, and population structure. *The Quarterly Review of Biology*, *74*, 21–45. http://www.jstor.org/stable/2664376

Bohonak, A. J., & Jenkins, D. G. (2003). Ecological and evolutionary significance of dispersal by freshwater invertebrates. *Ecology Letters*, *6*, 783–796. https://doi.org/10.1046/j.1461-0248.2003.00486.x

Bonada, N., & Dolédec, S. (2018). Does the Tachet trait database report voltinism variability of aquatic insects between Mediterranean and Scandinavian regions? *Aquatic Sciences*, *80*, 1–11. https://doi.org/10.1007/s00027-017-0554-z

Brown, B. L., & Swan, C. M. (2010). Dendritic network structure constrains metacommunity properties in riverine ecosystems. *Journal of Animal Ecology*, *79*, 571–580. https://doi.org/10.1111/j.1365-2656.2010.01668.x

Bush, A., & Hoskins, A. J. (2017). Does dispersal capacity matter for freshwater biodiversity

291 under climate change? *Freshwater Biology*, *62*, 382–396.

292 https://doi.org/10.1111/fwb.12874

293 Cañedo-Argüelles, M., Boersma, K. S., Bogan, M. T., Olden, J. D., Phillipsen, I., Schriever,

294 T. A., & Lytle, D. A. (2015). Dispersal strength determines meta-community structure in

295 a dendritic riverine network. *Journal of Biogeography*, *42*, 778–790.

296 https://doi.org/10.1111/jbi.12457

297 Chevenet, F., Dolédec, S., & Chessel, D. (1994). A fuzzy coding approach for the analysis of

298 long-term ecological data. *Freshwater Biology*, *31*, 295–309.

299 https://doi.org/10.1007/978-3-319-71368-7_23

300 Cid, N., Bonada, N., Heino, J., Cañedo-Argüelles, M., Crabot, J., Sarremejane, R., … Datry,

301 T. (under review). A Metacommunity approach to improve biological assessments in

302 highly dynamic freshwater ecosystems.

303 Clobert, J., Baguette, M., Benton, T. G., & Bullock, J. M. (2012). *Dispersal ecology and*

304 *evolution*. Oxford University Press.

305 Cooper, J. C., & Soberón, J. (2018). Creating individual accessible area hypotheses improves

306 stacked species distribution model performance. *Global Ecology and Biogeography*, *27*,

307 156–165. https://doi.org/10.1111/geb.12678

308 Datry, T., Bonada, N., & Heino, J. (2016). Towards understanding the organisation of

309 metacommunities in highly dynamic ecological systems. *Oikos*, *125*, 149–159.

310 https://doi.org/10.1111/oik.02922

311 Datry, T., Melo, A. S., Moya, N., Zubieta, J., De la Barra, E., & Oberdorff, T. (2016).

312 Metacommunity patterns across three Neotropical catchments with varying

313 environmental harshness. *Freshwater Biology*, *61*, 277–292.

https://doi.org/10.1111/fwb.12702

Datry, T., Moya, N., Zubieta, J., & Oberdorff, T. (2016). Determinants of local and regional

communities in intermittent and perennial headwaters of the Bolivian Amazon.

Freshwater Biology, *61*, 1335–1349. https://doi.org/10.1111/fwb.12706

Dijkstra, K.-D. B., Monaghan, M. T., & Pauls, S. U. (2014). Freshwater biodiversity and

aquatic insect diversification. *Annual Review of Entomology*, *59*, 143–163.

https://doi.org/10.1146/annurev-ento-011613-161958

French, S. K., & McCauley, S. J. (2019). The movement responses of three libellulid

dragonfly species to open and closed landscape cover. *Insect Conservation and*

Diversity. https://doi.org/10.1111/icad.12355

Graham, E. S., Storey, R., & Smith, B. (2017). Dispersal distances of aquatic insects:

upstream crawling by benthic EPT larvae and flight of adult Trichoptera along valley

floors. *New Zealand Journal of Marine and Freshwater Research*, *51*, 146–164.

https://doi.org/10.1080/00288330.2016.1268175

Heino, J., Alahuhta, J., Ala-Hulkko, T., Antikainen, H., Bini, L. M., Bonada, N., … Soininen,

J. (2017). Integrating dispersal proxies in ecological and environmental research in the

freshwater realm. *Environmental Reviews*, *25*, 334–349. https://doi.org/10.1139/er-2016-

0110

Heino, J., Melo, A. S., Siqueira, T., Soininen, J., Valanko, S., & Bini, L. M. (2015).

Metacommunity organisation, spatial extent and dispersal in aquatic systems: Patterns,

processes and prospects. *Freshwater Biology*, *60*, 845–869.

https://doi.org/10.1111/fwb.12533

Hermoso, V., Cattarino, L., Kennard, M. J., Watts, M., & Linke, S. (2015). Catchment zoning

for freshwater conservation: Refining plans to enhance action on the ground. *Journal of*

Applied Ecology, *52*, 940–949. https://doi.org/10.1111/1365-2664.12454

Hoffsten, P. O. (2004). Site-occupancy in relation to flight-morphology in caddisflies.

Freshwater Biology, *49*, 810–817. https://doi.org/10.1111/j.1365-2427.2004.01229.x

Jacobson, B., & Peres-Neto, P. R. (2010). Quantifying and disentangling dispersal in

metacommunities: How close have we come? How far is there to go? *Landscape*

Ecology, *25*, 495–507. https://doi.org/10.1007/s10980-009-9442-9

Jenkins, D. G., Brescacin, C. R., Duxbury, C. V., Elliott, J. A., Evans, J. A., Grablow, K. R.,

… Williams, S. E. (2007). Does size matter for dispersal distance? *Global Ecology and*

Biogeography, *16*, 415–425. https://doi.org/10.1111/j.1466-8238.2007.00312.x

Lancaster, J., & Downes, B. J. (2013). Dispersal in the terrestrial environment. In *Aquatic*

Entomology. Oxford: Oxford University Press.

https://doi.org/10.1093/acprof:oso/9780199573219.003.0009

Lancaster, J., & Downes, B. J. (2017). Dispersal traits may reflect dispersal distances, but

dispersers may not connect populations demographically. *Oecologia*, *184*, 171–182.

https://doi.org/10.1007/s00442-017-3856-x

Lancaster, J., Downes, B. J., & Arnold, A. (2011). Lasting effects of maternal behaviour on

the distribution of a dispersive stream insect. *Journal of Animal Ecology*, *80*, 1061–1069.

https://doi.org/10.1111/j.1365-2656.2011.01847.x

Maasri, A. (2019). A global and unified trait database for aquatic macroinvertebrates: The

missing piece in a global approach. *Frontiers in Environmental Science*, *7*, 1–3.

https://doi.org/10.3389/fmars.2019.00065

Macneale, K. H., Peckarsky, B. L., & Likens, G. E. (2005). Stable isotopes identify dispersal

patterns of stonefly populations living along stream corridors. *Freshwater Biology*, *50*,

1117–1130. https://doi.org/10.1111/j.1365-2427.2005.01387.x

Markovic, D., Carrizo, S., Freyhof, J., Cid, N., Lengyel, S., Scholz, M., … Darwall, W.

(2014). Europe's freshwater biodiversity under climate change: Distribution shifts and

conservation needs. *Diversity and Distributions*, *20*, 1097-1107.

https://doi.org/10.1111/ddi.12232

Múrria, C., Bonada, N., Vellend, M., Zamora-Muñoz, C., Alba-Tercedor, J., Sainz-Cantero,

C. E., … Vogler, A. P. (2017). Local environment rather than past climate determines

community composition of mountain stream macroinvertebrates across Europe.

Molecular Ecology, *26*, 6085–6099. https://doi.org/10.1111/mec.14346

Outomuro, D., & Johansson, F. (2019). Wing morphology and migration status, but not body

size, habitat or Rapoport's rule predict range size in North-American dragonflies

(Odonata: Libellulidae). *Ecography*, *42*, 309–320. https://doi.org/10.1111/ecog.03757

Reid, A. J., Carlson, A. K., Creed, I. F., Eliason, E. J., Gell, P. A., Johnson, P. T. J., … Cooke,

S. J. (2019). Emerging threats and persistent conservation challenges for freshwater

biodiversity. *Biological Reviews*, *94*, 849–873. https://doi.org/10.1111/brv.12480

Rundle, S. D., Bilton, D. T., & Foggo, A. (2007). By wind, wings or water: body size,

dispersal and range size in aquatic invertebrates. In A. G. Hildrew, D. G. Raffaelli, & R.

Edmonds-Brown (Eds.), *Body Size: The Structure and Function of Aquatic Ecosystems*

(pp. 186–209). Cambridge: Cambridge University Press. https://doi.org/DOI:

10.1017/CBO9780511611223.011

Sarremejane, R., Cañedo-Argüelles, M., Prat, N., Mykrä, H., Muotka, T., & Bonada, N.

(2017). Do metacommunities vary through time? Intermittent rivers as model systems.

Journal of Biogeography, *44*, 2752–2763. https://doi.org/10.1111/jbi.13077

Schmera, D., Podani, J., Heino, J., Erös, T., & Poff, N. L. R. (2015). A proposed unified

terminology of species traits in stream ecology. *Freshwater Science*, *34*, 823–830.

https://doi.org/10.1086/681623

Schmidt-Kloiber, A., & Hering, D. (2015). www.freshwaterecology.info – An online tool that

unifies, standardises and codifies more than 20,000 European freshwater organisms and

their ecological preferences. *Ecological Indicators*, *53*, 271–282.

https://doi.org/https://doi.org/10.1016/j.ecolind.2015.02.007

Serra, S. R. Q., Cobo, F., Graça, M. A. S., Dolédec, S., & Feio, M. J. (2016). Synthesising the

trait information of European Chironomidae (Insecta: Diptera): Towards a new database.

Ecological Indicators, *61*, 282–292. https://doi.org/10.1016/j.ecolind.2015.09.028

Statzner, B., & Bêche, L. a. (2010). Can biological invertebrate traits resolve effects of

multiple stressors on running water ecosystems? *Freshwater Biology*, *55*, 80–119.

https://doi.org/10.1111/j.1365-2427.2009.02369.x

Stevens, V. M., Trochet, A., Blanchet, S., Moulherat, S., Clobert, J., & Baguette, M. (2013).

Dispersal syndromes and the use of life-histories to predict dispersal. *Evolutionary*

Applications, *6*, 630–642. https://doi.org/10.1111/eva.12049

Strayer, D. L., & Dudgeon, D. (2010). Freshwater biodiversity conservation: recent progress

and future challenges. *Journal of the North American Benthological Society*, *29*, 344–

358. https://doi.org/10.1899/08-171.1

Swan, C. M., & Brown, B. L. (2017). Metacommunity theory meets restoration: Isolation may

mediate how ecological communities respond to stream restoration: Isolation. *Ecological*

Applications, *27*, 2209–2219. https://doi.org/10.1002/eap.1602

Tachet, H., Richoux, P., Bournaud, M., & Usseglio-Polatera, P. (2010). *Invertébrés d'eau*

douce: systématique, biologie, écologie. (C. Éditions, Ed.) (3rd ed.). Paris.

Thuiller, W., Münkemüller, T., Lavergne, S., Mouillot, D., Mouquet, N., Schiffers, K., &

Gravel, D. (2013). A road map for integrating eco-evolutionary processes into

biodiversity models. *Ecology Letters*, *16*, 94–105. https://doi.org/10.1111/ele.12104

Tonkin, J. D., Altermatt, F., Finn, D. S., Heino, J., Olden, J. D., Pauls, S. U., & Lytle, D. A.

(2018). The role of dispersal in river network metacommunities: Patterns, processes, and

pathways. *Freshwater Biology*, *63*, 141–163. https://doi.org/10.1111/fwb.13037

Troast, D., Suhling, F., Jinguji, H., Sahlén, G., & Ware, J. (2016). A global population genetic

study of pantala flavescens. *PLoS ONE*, *11*, 1–13.

https://doi.org/10.1371/journal.pone.0148949

Vieira, N. K. M., Poff, N. L., Carlisle, D. M., Moulton II, S. R., Koski, M. L., & Kondratieff,

B. C. (2006). *A database of lotic invertebrate traits for North America. Data Series*.

https://doi.org/10.3133/ds187

Wikelski, M., Moskowitz, D., Adelman, J. S., Cochran, J., Wilcove, D. S., & May, M. L.

(2006). Simple rules guide dragonfly migration. *Biology Letters*, *2*, 325–329.

https://doi.org/10.1098/rsbl.2006.0487

Willis, S. G., Foden, W., Baker, D. J., Belle, E., Burgess, N. D., Carr, J. A., … Butchart, S. H.

M. (2015). Integrating climate change vulnerability assessments from species

distribution models and trait-based approaches. *Biological Conservation*, *190*, 167–178.

https://doi.org/10.1016/j.biocon.2015.05.001

Wilzbach, M. A., & Cummins, K. W. (1989). An assessment of short-term depletion of

stream macroinvertebrate benthos by drift. *Hydrobiologia*, *185*, 29–39.

https://doi.org/10.1007/BF00006065

Supporting information

a — Dispersal strategy

b — Propensity to drift

c — Fecundity

d — Life cycle duration

e — Adult life span

f — Maximum body size

g — Potential number of reproduction cycles

h — Wing length

i — Wing pair type

Figure S1: Trait category location in the Fuzzy Correspondence Analysis ordination space for each trait: (a) Dispersal strategy = **dis1**: aquatic active, **dis2**: aquatic passive, **dis3**: aerial active, **dis4**: aerial passive; (b) Propensity to drift = **drift1**: rare/catastrophic, **drift2**: occasional, **drift3**: frequent; (c) Fecundity = **egg1**: < 100 eggs, **egg2**: $\geq 100 - 1000$ eggs, **egg3**: $1000 - 3000$ eggs, **egg4**: ≥ 3000 eggs; (d) Life cycle duration = **cd1**: ≤ 1 year, **cd2**: > 1 year; (e) Adult life span = **life1**: < 1 week, **life2**: ≥ 1 week $- 1$ month, **life3**: ≥ 1 month $- 1$ year, **life4**: > 1 year; (f) Maximum body size (cm) = **s1**: < 0.25, **s2**: $\geq 0.25 - 0.5$, **s3**: $\geq 0.5 - 1$, **s4**: $\geq 1 - 2$; **s5**: $\geq 2 - 4$, **s6**: $\geq 4 - 8$; (g) Potential number of reproductive cycles per year = **cy1**: < 1, **cy2**: 1, **cy3**: > 1; (h) Female wing length (cm) = **fwl1**: < 5, **fwl2**: $\geq 5 - 10$, **fwl3**: $\geq 10 - 15$, **fwl4**: $\geq 15 - 20$, **fwl5**: $\geq 20 - 30$, **fwl6**: $\geq 30 - 40$, **fwl7**: $\geq 40 - 50$, **fwl8**: ≥ 50; (i) Wing pair type = **Wbn2**: 1 pair + halters, **Wbn3**: 1 pair + elytra or hemielytra, **Wbn4**: 1 pair + small hind wings, **Wbn5**: 2 similar-sized pairs.