

1 **Genomic Prediction with Genotype by Environment Interaction Analysis for Kernel Zinc**
2 **Concentration in Tropical Maize Germplasm**

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Genomic prediction for kernel zinc

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ABSTRACT

Zinc (Zn) deficiency is a major risk factor for human health, affecting about 30% of the world's population. To study the potential of genomic selection (GS) for maize with increased Zn concentration, an association panel and two doubled haploid (DH) populations were evaluated in three environments. Three genomic prediction models, M (M1: Environment + Line, M2: Environment + Line + Genomic, and M3: Environment + Line + Genomic + Genomic x Environment) incorporating main effects (lines and genomic) and the interaction between genomic and environment (G x E) were assessed to estimate the prediction ability (r_{MP}) for each model. Two distinct cross-validation (CV) schemes simulating two genomic prediction breeding scenarios were used. CV1 predicts the performance of newly developed lines, whereas CV2 predicts the performance of lines tested in sparse multi-location trials. Predictions for Zn in CV1 ranged from -0.01 to 0.56 for DH1, 0.04 to 0.50 for DH2 and -0.001 to 0.47 for the association panel. For CV2, r_{MP} values ranged from 0.67 to 0.71 for DH1, 0.40 to 0.56 for DH2 and 0.64 to 0.72 for the association panel. The genomic prediction model which included G x E had the highest average r_{MP} for both CV1 (0.39 and 0.44) and CV2 (0.71 and 0.51) for the association panel and DH2 population, respectively. These results suggest that GS has potential to accelerate breeding for enhanced kernel Zn concentration by facilitating selection of superior genotypes.

70

INTRODUCTION

71 Malnutrition arising from zinc (Zn) deficiency is a major risk factor for human health
72 affecting nearly 30% of the world's population (Bouis and Saltzman 2017; Gannon *et al.* 2017).
73 The problem is more prevalent in low-and middle income countries (LMICs), and is highly
74 attributed to lack of access to a balanced diet, reliance on cereal-based diets and ignorance of good

75 nutritional practices (Welch and Graham 2004). Several approaches, such as food fortification,
76 diversification and supplementation have been tried to reduce Zn deficiency. However, in LMICs,
77 these methods have not been entirely successful (Misra *et al.* 2004; Stein 2010).

78 Breeding maize for increased Zn concentration may offer some relief. The Zn-enriched
79 varieties can be widely accessible, will not require continued investment once developed, and they
80 remain after the initial successful investment and research (Govindan 2011). Recently, maize
81 varieties with 15-36% more Zn were released in Guatemala and Colombia (Listman 2019).
82 Nevertheless, increased breeding efforts are required to develop more Zn-enriched varieties for a
83 diverse range of environments and management practices. Progress toward developing those
84 varieties has mainly relied upon conventional plant breeding approach that is labor-intensive and
85 time-consuming. However, with the recent advances in genomics, new methods for plant breeding
86 such as genomic selection (GS) can be used to identify genotypes with enhanced Zn concentration
87 more efficiently and rapidly.

88 In a GS breeding scheme, genome-wide DNA markers are used to predict which
89 individuals in a breeding population are most valuable as parents of the next generation (cycle)
90 of offspring (Meuwissen *et al.* 2001; de los Campos *et al.* 2009; Pérez-Rodríguez *et al.* 2012).
91 Kernel Zn concentration is determined at the end of a plant's life cycle, so GS can enable
92 selection of promising genotypes earlier in the life cycle. This reduces the time and cost of
93 phenotypic evaluation and may increase the genetic gain per unit time and cost (Heslot *et al.*
94 2015; Manickavelu *et al.* 2017; Arojju *et al.* 2019).

95 The utility and effectiveness of GS has been examined for many different crop species,
96 marker densities, traits and statistical models and varying levels of prediction accuracy have been
97 achieved (de los Campos *et al.* 2009, 2013; Crossa *et al.* 2010, 2013, 2014; Jarquín *et al.* 2014;

98 Pérez-Rodríguez *et al.* 2015; Zhang *et al.* 2015; Velu *et al.* 2016). Although the number of
99 markers needed for accurate prediction of genotypic values depends on the extent of linkage
100 disequilibrium between markers and QTL (Meuwissen *et al.* 2001), a higher marker density can
101 improve the proportion of genetic variation explained by markers and thus result in higher
102 prediction accuracy (Albrecht *et al.* 2011; Zhao *et al.* 2012; Combs and Bernardo 2013; Liu *et al.*
103 2018). Importantly, higher prediction accuracies have been obtained when genotypes of a
104 population are closely related than when genetically unrelated (Pszczola *et al.* 2012; Combs and
105 Bernardo 2013; Spindel and McCouch 2016).

106 Initially, GS models and methods were developed for single-environment analyses and
107 they did not consider correlated environmental structures due to genotype by environment (G x E)
108 interactions (Crossa *et al.* 2014). The differential response of genotypes in different environments
109 is a major challenge for breeders and can affect heritability and genotype ranking over
110 environments (Monteverde *et al.* 2018). Multi-environment analysis can model G x E using genetic
111 and residual covariance functions (Burgueño *et al.* 2012), markers and environmental covariates
112 (Jarquín *et al.* 2014), or marker by environment (M x E) interactions (Lopez-Cruz *et al.* 2015).
113 This approach to GS can successfully be used for biofortification breeding of maize because multi-
114 environment testing is routinely used in the development and release of varieties.

115 Modelling covariance matrices to account for G x E allows the use of information from
116 correlated environments (Burgueño *et al.* 2012). Mixed models that allow the incorporation of a
117 genetic covariance matrix calculated from marker data, rather than assuming independence among
118 genotypes improves the estimation of genetic effects (VanRaden 2008). The benefit of using
119 genetic covariance matrices in G x E mixed models is that the model relates genotypes across
120 locations even when the lines are not present in all locations (Monteverde *et al.* 2018). GS models

121 capable of accounting for multi-environment data have extensively been studied in different crops
122 (Zhang *et al.* 2015; Cuevas *et al.* 2016, 2017; Velu *et al.* 2016; Jarquín *et al.* 2017; Sukumaran *et*
123 *al.* 2017a; Monteverde *et al.* 2018; Roorkiwal *et al.* 2018). In those studies, incorporating G x E
124 demonstrated a substantial increase in prediction accuracy relative to single-environment analyses.

125 Kernel Zn has been investigated in several quantitative trait loci (QTL) analyses in maize
126 and each study has reported that Zn concentration is under the control of several loci. The
127 phenotypic variation explained by those loci ranges from 5.9 to 48.8% (Zhou *et al.* 2010; Qin *et*
128 *al.* 2012; Šimić *et al.* 2012; Baxter *et al.* 2013; Jin *et al.* 2013; Zhang *et al.* 2017a; Hindu *et al.*
129 2018). A Meta-QTL analysis across several of those studies identified regions on chromosome 2
130 that might be important for kernel Zn concentration (Jin *et al.* 2013). Additionally, genomic
131 regions associated with Zn concentration were recently reported in a genome-wide association
132 study of maize inbreds adapted to the tropics (Hindu *et al.* 2018). Whereas some of the regions
133 were novel, four of the twenty identified were located in already reported QTL intervals. Taken
134 together, the QTLs may be used in a breeding program through marker-assisted selection (MAS)
135 or GS.

136 A wide array of maize genetic studies has reported considerable effects of G x E
137 interactions for kernel Zn concentration (Oikeh *et al.* 2003, 2004; Long *et al.* 2004; Chakraborti *et*
138 *al.* 2009; Prasanna *et al.* 2011; Agrawal *et al.* 2012; Guleria *et al.* 2013). However, genotypes with
139 high-Zn concentration have been identified in both tropical and temperate germplasm (Ahmadi *et*
140 *al.* 1993; Bänziger and Long 2000; Brkic *et al.* 2004; Menkir 2008; Chakraborti *et al.* 2011;
141 Prasanna *et al.* 2011; Hindu *et al.* 2018). Additionally, evaluation procedures for kernel Zn are
142 labor-intensive, expensive and time-consuming (Palacios-Rojas 2018). To the best of our
143 knowledge, no study has examined the predictive ability of GS methods that incorporate G x E for

144 Zn concentration in maize. Within the framework of the reaction norm model (Jarquín *et al.* 2014),
145 the potential of GS for Zn using maize inbreds adapted to tropical environments were assessed.
146 The objectives of this study were; (i) to evaluate the prediction ability for Zn using an association
147 mapping panel and two bi-parental populations evaluated in three tropical environments, (ii) to
148 assess and compare the predictive ability of different GS models, and (iii) to examine the effects
149 of incorporating G x E on prediction accuracy for Zn.

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MATERIALS AND METHODS

Zinc association mapping (ZAM) panel

154 The ZAM panel consists of 923 inbreds from maize breeding programs of the
155 International Maize and Wheat Improvement Center (CIMMYT). The panel represents wide
156 genetic diversity for kernel Zn concentration (Hindu *et al.* 2018).

157

Bi-parental DH populations

159 From the ZAM panel, four inbreds with contrasting Zn concentration were selected and
160 used to form two bi-parental (doubled haploid [DH]) populations (Table 1). DH1 was derived
161 from the F1 generation of a mating between CML503, a high-Zn inbred (31.21 µg/g) with
162 CLWN201, a low-Zn inbred (22.62 µg/g). DH2 was derived from the F1 generation of a mating
163 between CML465, another high-Zn inbred (31.55 µg/g) with CML451, a moderate-Zn inbred
164 (27.88 µg/g). DH1 and DH2 were comprised of 112 and 143 inbreds, respectively.

165

166 **Experimental design and phenotypic evaluation**

167 **Zinc association mapping (ZAM) panel**

168 The ZAM panel was grown at CIMMYT research stations in Mexico, during the months
169 of June through September and November through March at Agua Fria in 2012 and 2013, and
170 Celaya in 2012. Plot sizes and the experimental designs (Hindu *et al.* 2018).

171

172 **Bi-parental DH populations**

173 The DH populations were grown at CIMMYT research stations in Mexico; Celaya in
174 2014 and Tlaltizapan (18°41'N, 99° 07' W; 962.5 m asl) in 2015 and 2017. In 2014 and 2015,
175 both populations were evaluated in single-replication trials (Hindu *et al.* 2018). In 2017, a
176 randomized complete block design (RCBD) with two replications was used. The rows were 2.5
177 m long and 75 cm apart and each genotype was grown in a single row plot. All plots were
178 managed according to the recommended agronomic practices for each environment.

179 From the ZAM panel and each DH population, four to six plants in each plot were self-
180 pollinated, hand-harvested at physiological maturity, hand-shelled and dried to a moisture
181 content of 12.5%. The bulked kernels from each plot are considered a representative sample and
182 were used in subsequent Zn analyses as described (Hindu *et al.* 2018).

183

184 **Genotypic data**

185 Genomic DNA was extracted from leaf tissues of all inbred lines (ZAM panel and DH
186 populations) using the standard CIMMYT laboratory protocol (CIMMYT, 2005). The samples
187 were genotyped using the genotyping by sequencing (GBS) method at the Institute for Genomic
188 Diversity, Cornell University, USA (Elshire *et al.* 2011; Crossa *et al.* 2013). The restriction

189 enzyme ApeK1 was used to digest DNA, GBS libraries were constructed in 96-plex and
190 sequenced on a single lane of Illumina HiSeq2000 flow cell (Elshire *et al.* 2011). To increase the
191 genome coverage and read depth for SNP discovery, raw read data from the sequencing samples
192 were analyzed together with an additional ~30, 000 global maize collections (Zhang *et al.* 2015).

193 SNP identification was performed using TASSEL 5.0 GBS Discovery Pipeline with B73
194 (RefGen_v2) as the reference genome (Elshire *et al.* 2011; Glaubitz *et al.* 2014). The source code
195 and the TASSEL GBS discovery pipeline are available at <https://www.maizegenetics.net> and the
196 SourceForge Tassel project <https://sourceforge.net/projects/tassel>. For each inbred, the pipeline
197 yielded 955, 690 SNPs which were distributed on the 10 maize chromosomes. After filtering
198 using a minor allele frequency of 0.05 and removing SNPs with more than 10% missing data,
199 181,889 (ZAM panel) and 170, 798 (bi-parental) SNPs were used for genomic prediction.

200

201 **Phenotypic data analysis**

202 For the ZAM panel, broad-sense heritability (H^2) across environments was estimated as:

203

$$H^2 = \frac{\sigma_G^2}{\sigma_G^2 + \sigma_{GE}^2/l + \sigma_e^2/lr}$$

204 where σ_G^2 is the variance due to genotype, σ_{GE}^2 is variance due to genotype x environment, σ_e^2 is
205 the error variance, l is the number of environments and r is the number of replications using
206 multi-environment trial analysis with R (META-R) (Alvarado *et al.* 2016). For the DH
207 populations, variance components based on the genomic relationship matrix were computed
208 using BGLR package as implemented in GBLUP (Pérez and de los Campos 2014). An estimate
209 of narrow-sense heritability (\hat{h}^2) for each DH population was calculated as:

210
$$\hat{h}^2 = \frac{\hat{\sigma}_g^2}{\hat{\sigma}_g^2 + \hat{\sigma}_e^2}$$

211 where $\hat{\sigma}_g^2$ is an estimate of the additive genetic variance and $\hat{\sigma}_e^2$ is an estimate of the residual
212 variance.

213 Correlation coefficients between Zn and environments, descriptive statistics and
214 phenotypic data distribution using boxplots were generated in R (core Team 2018). Line means
215 (genotypic values) for the ZAM panel were estimated as Best Linear Unbiased Estimators
216 (BLUEs) with a random effect for replications nested within each environment. Raw data
217 (values) were used for the DH populations.

218

219 **Statistical models**

220 Genomic models used in this study were based on the reaction norm model which models
221 the markers (genomic) by environment interaction (Jarquín *et al.* 2014). This model is an
222 extension of the Genomic Best Linear Unbiased Predictor (GBLUP) random effect model, where
223 the main effects of lines (genotypes), genomic, environments and their interactions are modelled
224 using covariance structures that are functions of marker genotypes and environmental covariates.
225 In this study, environment is the combination of site and year (site-by-year). A brief description
226 of the models is given below.

227

228 **M0. Phenotypic baseline model**

229 The phenotypes y_{ij} are modelled as:

230
$$y_{ij} = \mu + E_i + L_j + EL_{ij} + e_{ij},$$

231 this linear model represents the response of the j^{th} ($j=1, \dots, J$) genotype/line tested in the i^{th}
232 ($i=1, \dots, I$) environment $[\{y\}_{ij}]$ as the sum of an overall mean μ plus random environmental main
233 effect $[E_i \stackrel{iid}{\sim} N(0, \sigma_E^2)]$, the random genotype main effect $[L_j \stackrel{iid}{\sim} N(0, \sigma_L^2)]$, the random interaction
234 between the j^{th} genotype and the i^{th} environment $[EL_{ij} \stackrel{iid}{\sim} N(0, \sigma_{EL}^2)]$ and a random error term
235 $[e_{ij} \stackrel{iid}{\sim} N(0, \sigma_e^2)]$. From this linear model, $N(.,.)$ denotes a normal random variable, *iid* stands for
236 independent and identically distributed responses and σ_E^2 , σ_L^2 , σ_{EL}^2 , σ_e^2 are the variances for
237 environment, genotype, genotype by environment and residual error, respectively. The baseline
238 model does not allow borrowing of information among genotypes because the genotypes were
239 treated as independent outcomes. Thus, models used in this study were derived from the baseline
240 model by subtracting terms or modifying assumptions and/or incorporating genomics/marker
241 information.

242

243 **M1. Environment + Line**

244 This model is obtained by retaining the first three components from the baseline model
245 (overall mean, random environment main effect and random line main effect) while their
246 underlying assumptions remain unchanged.

$$247 \quad y_{ij} = \mu + E_i + L_j + e_{ij}. \quad [1]$$

248 Here environments were considered as site-by-year combinations.

249

250 **M2. Environment + Line + Genomic**

251 Another representation of the random main effect of line L_j in the previous model is
252 considering a linear combination between markers and their correspondent marker effects, $g_j =$
253 $\sum_{m=1}^p x_{jm} b_m$, such that

$$254 \quad y_{ij} = \mu + E_i + L_j + g_j + e_{ij} \quad [2]$$

255 where $b_m \stackrel{iid}{\sim} N(0, \sigma_b^2)$ represents the random effect of the m^{th} ($m=1, \dots, p$) marker, x_{jm} is the
256 genotype of the j^{th} line at the m^{th} marker and σ_b^2 its correspondent variance component.

257 Therefore, $\mathbf{g} = (g_1, \dots, g_J)'$, is the vector of genetic effects, and follows a normal density with
258 mean zero, and a co-variance matrix $Cov(\mathbf{g}) = \mathbf{G}\sigma_g^2$ with $\mathbf{G} = \frac{\mathbf{XX}'}{p}$ being the genomic relationship
259 matrix (Lopez-Cruz *et al.* 2015) that describes genetic similarities among pairs of individuals. In
260 this model, the line effect L_j is retained to account for imperfect information and model mis-
261 specification because of potential imperfect linkage disequilibrium between markers and
262 quantitative trait loci (QTLs).

263

264 **M3. Environment + Line + Genomic + Genomic \times Environment**

265 This model accounts for the effects of lines L_j , of markers (genomic) g_j , of environments (E_i)
266 and the interaction between markers (genomic) and the environment (Eg_{ij}). The model includes
267 the interaction between markers (genomics) and the environment via co-variance structure
268 (Jarquín *et al.* 2014). The model is as follows:

$$269 \quad y_{ij} = \mu + E_i + L_j + g_j + Eg_{ij} + e_{ij} \quad [3]$$

270 Where Eg_{ij} is the interaction between the genetic value of the i^{th} genotype in the j^{th} environment
271 and $\mathbf{Eg} = \{Eg_{ij}\} \sim N(\mathbf{0}, (\mathbf{Z}_g \mathbf{G} \mathbf{Z}_g') \# (\mathbf{Z}_E \mathbf{Z}_E') \sigma_{Eg}^2)$, where \mathbf{Z}_g and \mathbf{Z}_E are the correspondent
272 incidence matrices for the effects of genetic values of genotypes and environments, respectively,

273 σ_{Eg}^2 is the variance component of Eg and # denotes the Hadamard product (element-to-element
274 product) between two matrices.

275

276 **Prediction accuracy assessment using cross-validation**

277 Two distinct cross-validation schemes that mimic prediction problems that breeders may face
278 when performing genomic prediction were used (Burgueño *et al.* 2012). One random cross-
279 validation (CV1) evaluates the prediction ability of models when a set of lines have not been
280 evaluated in any environment (prediction of newly developed lines). In CV1, predictions are
281 entirely based on phenotypic records of genetically related lines. The second cross-validation
282 (CV2) is related to incomplete field trials also known as sparse testing, in which some lines are
283 observed in some environments but not in others. In CV2, the goal is to predict the performance
284 of lines in environments where they have not yet been observed. Thus, information from related
285 lines and the correlated environments is used, and prediction assessment can benefit from
286 borrowing information between lines within an environment, between lines across environments
287 and among correlated environments.

288 In CV1 and CV2, a fivefold cross-validation scheme was used to generate the training and
289 validation sets to assess the prediction ability for Zn within the ZAM panel and each DH
290 population. The data were randomly divided into five subsets, with 80% of the lines assigned to
291 the training set and 20% assigned to the validation set. Four subsets were combined to form the
292 training set, and the remaining subset was used as the validation set. Permutation of five subsets
293 taken one at a time led to five training and validation data sets. The procedure was repeated 20
294 times and a total of 100 runs were performed in each population. The average value of the
295 correlations between the phenotype and the genomic estimated breeding values (GEBVs) from

296 100 runs was calculated for the ZAM panel, and each DH population for Zn in each environment
297 and was defined as the prediction ability (r_{MP}).

298

299 **Data availability**

300 All models were fitted in R (core Team 2018) using the BGLR package (Pérez and de los
301 Campos 2014). All phenotypic and genomic data can be downloaded from the link:

302 <http://hdl.handle.net/11529/10548331>

303

304

RESULTS

305 **Descriptive statistics**

306 Mean values of kernel Zn concentration were estimated for each environment and across
307 environments (Tables 2A and 2B). For the ZAM panel, kernel Zn ranged from 14.76 to 39.80
308 $\mu\text{g/g}$ in Celaya 2012, 15.16 to 42.52 $\mu\text{g/g}$ and 17.05 to 46.52 $\mu\text{g/g}$ in Agua Fria 2012 and 2013,
309 respectively (Figure 1). The highest mean (29.53 $\mu\text{g/g}$) for Zn was observed in Agua Fria 2013.
310 DH1 had Zn values ranging from 16.00 to 48.00 $\mu\text{g/g}$ in Celaya 2012, 16.00 to 35.00 $\mu\text{g/g}$ in
311 Tlaltizapan 2015 and 15.50 to 39.00 $\mu\text{g/g}$ in Tlaltizapan 2017, while the respective values for DH
312 2 were 17.70 to 43.14 $\mu\text{g/g}$, 15.60 to 37.80 $\mu\text{g/g}$ and 14.70 to 37.60 $\mu\text{g/g}$ (Figures 2A and 2B).
313 The highest means for Zn were observed in Celaya 2014 (25.38 $\mu\text{g/g}$) and 2017 (27.96 $\mu\text{g/g}$) for
314 DH1 and DH2, respectively (Table 2B). Across environments, heritability (H^2/\widehat{h}^2) estimates
315 were 0.85, 0.83 and 0.76 for the ZAM panel, DH1 and DH2, respectively (Tables 2A and 2B).
316 There were significant positive correlations between environments for Zn (Table 3), accounting
317 for the moderate to high heritability estimates.

318 Principal component analysis for the ZAM panel suggested presence of a relatively
319 diverse set of lines, and 452 principal components (PCs) were needed to explain 80% of the
320 genotypes' variance (Figures 3A and 3B). The first two principal components explained 3.85%
321 of the total variance. For the DH populations first two eigenvectors separated them two groups
322 (DH1 and DH2) and 56 principal components were needed to explain 80% of the genotypes'
323 variance (Figures 3C and 3D). The first two principal components explained 27.50% of the total
324 variation for the DH populations.

325

326 **Prediction ability in different populations**

327 Cross-validated r_{MP} values for kernel Zn were estimated for the ZAM panel and DH
328 populations (Tables 4, 5 and 6). The average r_{MP} values in CV1 were consistently lower than
329 those in CV2, suggesting the importance of using information from correlated environments
330 when predicting performance of inbred lines. The mean r_{MP} values in CV1 and CV2 for the
331 ZAM panel were 0.39 and 0.71, respectively (Table 4). For the DH populations, average r_{MP}
332 values were 0.53 for DH1-CV1, 0.44 for DH2-CV1 (Table 5), 0.70 for DH1-CV2 and 0.51 for
333 DH2-CV2 (Table 6).

334 In the ZAM panel, the highest values in CV1 (0.47) and CV2 (0.72) were obtained in
335 Celaya and Agua Fria 2012 (Table 4). For the bi-parental populations, both under CV1 and CV2,
336 higher r_{MP} values were observed for DH1 compared to DH2. The highest values in CV1 (0.56)
337 and CV2 (0.71) were observed in Tlaltizapan 2017 and 2015, all for DH1 (Tables 5 and 6). The
338 consistently higher r_{MP} values in CV1 and CV2 of DH1 could be attributed to the higher (0.58 to
339 0.62) correlation values between environments (Table 3).

340

341 Prediction ability of different models

342 Comparing the r_{MP} values obtained from each model, M1 had the lowest (-0.001, -0.03
343 and 0.04) accuracies in CV1 for the ZAM panel and DH populations (Tables 4 and 5). Those
344 values were improved in CV2 because the predictions benefited from previous records (collected
345 from other environments) of lines whose Zn values were being predicted. When M1 was
346 expanded to M2 by adding the main effects of markers, the r_{MP} values at each environment and
347 across environments were increased. For example, in CV1, M2, >100-fold increase in r_{MP} values
348 were observed for the ZAM panel and DH populations, and in CV2, M2, average r_{MP} values
349 increased by 2.98%, 2.94% and 11.11% for the ZAM panel, DH1 and DH2, respectively (Tables
350 4, 5 and 6).

351 The multi-environment model (M3), which includes the interaction between markers
352 (genomic) and the environment (Eg_{ij}) gave higher prediction accuracy than single-environment
353 models (M1 and M2). In CV1, mean r_{MP} values increased from 0.37 (M2) to 0.39 (M3) for the
354 ZAM panel and from 0.43 (M2) to 0.44 for DH2 (Tables 4 and 5). Similar trends were observed
355 in CV2 for the ZAM panel and DH2 (Tables 4 and 6). However, in both CV1 and CV2 of DH1,
356 incorporating Eg_{ij} did not improve r_{MP} values for Zn (Tables 5 and 6). For CV1, M3, r_{MP} values
357 for Zn in individual environments ranged from 0.34 to 0.47 for the ZAM panel (Table 4), 0.51 to
358 0.55 for DH1 and 0.35 to 0.50 for DH2 (Table 5). For CV2, M3, those values ranged from 0.69
359 to 0.72 for the ZAM panel, 0.68 to 0.70 for DH1 and 0.43 to 0.56 for DH2 (Tables 4, 5 and 6).

360

361

DISCUSSION

362 Overall, moderate to high prediction ability values for kernel Zn were observed for the
363 ZAM panel and DH populations. This could be attributed to the heritabilities observed for kernel
364 Zn (Tables 2A and 2B). Similar observations were reported for Zn concentration in wheat (Velu
365 *et al.* 2016; Manickavelu *et al.* 2017). High quality predictions with high accuracy for GS
366 programs are expected for traits with moderate to higher heritability estimates (Combs and
367 Bernardo 2013; Lian *et al.* 2014; Muranty *et al.* 2015; Saint Pierre *et al.* 2016; Manickavelu *et*
368 *al.* 2017; Zhang *et al.* 2017b, 2019; Arojju *et al.* 2019). Consistent with a study on Zn and iron
369 (Fe) concentration in spring wheat, the prediction accuracies in this study are sufficient to
370 discard at least 50% of the inbreds with low-Zn concentration (Velu *et al.* 2016).

371 Data from both bi-parental populations and diverse collection of inbreds have been used
372 for GS and cross-validation (CV) experiments have shown that prediction accuracies could also
373 be affected by the relatedness between training and prediction sets (Habier *et al.* 2007; de Roos
374 *et al.* 2009; Asoro *et al.* 2011; Daetwyler *et al.* 2013; Cericola *et al.* 2017; Crossa *et al.* 2017). In
375 this study, average predicted accuracies were higher for CV1 of the bi-parental populations (0.53
376 for DH1 and 0.44 for DH2) compared to the ZAM panel (0.39). Higher predicted values in CV1
377 of the DH populations could be attributed to the closer relationship between DH lines in the
378 training and prediction sets, maximum linkage disequilibrium (LD) between a marker and a
379 QTL, and controlled population structure (Bernardo and Yu 2007; Albrecht *et al.* 2011; Zhang *et*
380 *al.* 2015). In collections of diverse inbreds, prediction accuracy may depend on the ancestral
381 relationships between the lines. So, in experiments using such collections of lines, prediction
382 accuracies have been more variable than accuracies achieved using bi-parental populations
383 (Spindel and McCouch 2016).

384 Cross-validation (CV) schemes are used in genomic prediction to estimate the accuracy
385 with which predictions for different traits and environments can be made (Burgueño *et al.* 2012;
386 Zhang *et al.* 2015; Saint Pierre *et al.* 2016; Velu *et al.* 2016; Sukumaran *et al.* 2017a, 2017b;
387 Monteverde *et al.* 2018; Roorkiwal *et al.* 2018). In this study, two CV schemes (CV1- predicting
388 the performance of newly developed lines, and CV2- predicting the performance of lines that
389 have been evaluated in some environments, but not in others) were used. The utility of these
390 schemes indicated that prediction values for newly developed lines (CV1) were generally lower
391 (0.39 for the ZAM panel, 0.53 for DH1 and 0.44 for DH2) than the values for lines which have
392 been evaluated in different but correlated environments (CV2; 0.71, 0.70 and 0.51 for the ZAM
393 panel, DH1 and DH2, respectively). Such observations indicate the importance of using
394 information from correlated environments when predicting the performance of inbred lines.
395 However, selection of new lines without field testing, as simulated in CV1 allows shortening of
396 the generation interval (cycle time) by replacing the time-intensive phenotypic evaluation for Zn
397 with genomic-estimated breeding values. But, the quality of prediction accuracy may be lower
398 such that the annual rate of genetic progress in a GS program is compromised (Burgueño *et al.*
399 2012). So, the ultimate decision of how a breeding scheme should be structured could depend on
400 the compromise between the desired prediction accuracy and the generation interval (Burgueño
401 *et al.* 2012).

402 Genotype by environment interaction is an important factor affecting kernel Zn
403 concentration in maize and genomic prediction models that incorporate G x E may enhance the
404 potential of GS for biofortification breeding. For different crop species and traits, genomic
405 prediction models which incorporated G x E achieved higher prediction accuracies in both CV1
406 and CV2 schemes relative to models which did not include G x E (Burgueño *et al.* 2012; Guo *et*

407 *al.* 2013; Jarquín *et al.* 2014; Lopez-Cruz *et al.* 2015; Zhang *et al.* 2015; Monteverde *et al.*
408 2018). In this study, the impact of modeling G x E variance structures for multi-environment
409 trials was investigated and results indicated that the average predicted values from M3 (G x E
410 model) were higher (0.39 and 0.44 for CV1 and 0.71 and 0.51 for CV2) than the values from M2
411 (non-G x E; 0.37 and 0.43 for CV1-M2, 0.69 and 0.50 for CV2-M2) for the ZAM panel and
412 DH2. These findings agree with those reported on Zn concentration in wheat (Velu *et al.* 2016),
413 providing evidence that incorporating G x E in GS models can enhance their power and
414 suitability for improving maize for kernel Zn concentration. Conversely, the average predicted
415 values for CV1 and CV2 of DH1 were higher in M2 (0.53 and 0.70) than in M3 (0.53 and 0.69).
416 Except for differences in population size (112 lines vs 143 lines), this was unexpected since DH1
417 and DH2 were grown in the same environments.

418 The gains in prediction accuracies for the GS model that accounted for G x E were
419 dependent on the correlation between environments and CV method used. In this study, the
420 phenotypic correlations between environments were all positive (ranging from 0.58 to 0.62 for
421 DH1, 0.29 to 0.46 for DH2 and 0.61 to 0.66 for the ZAM panel). Such correlations can be
422 exploited using multi-environment models to derive predictions that use information from across
423 both the lines and environments (Burgueño *et al.* 2012). For instance, although the phenotypic
424 correlations between environments for DH2 were positive (0.29 to 0.46), the lowest average
425 prediction value (0.51) for CV2 was observed for this population. This was expected because
426 CV2 uses phenotypic information from genotypes which have already been tested; hence,
427 effectively exploiting the correlations between environments (Burgueño *et al.* 2012; Jarquín *et*
428 *al.* 2014; Crossa *et al.* 2015; Pérez-Rodríguez *et al.* 2015; Saint Pierre *et al.* 2016; Monteverde *et*
429 *al.* 2018). However, for CV1, the information between environments could only be accounted for

430 through the genomic relationship matrix (Monteverde *et al.* 2018). Hence, the gains in CV1 may
431 likely attribute to more accurate estimate of environment-specific marker effects (Guo *et al.*
432 2013). In contrast, when multiple environments are weakly correlated, prediction accuracies
433 from across environment analyses can be negatively affected relative to prediction accuracies
434 within environments (Bentley *et al.* 2014; Wang *et al.* 2014; Spindel and McCouch 2016). Thus,
435 before designing a GS experiment, identifying correlated environments where environments can
436 differ in terms of site, year or season in which data were collected is of great interest (Spindel
437 and McCouch 2016).

438 The ability to predict kernel Zn concentration using high-throughput SNP markers
439 including G x E interactions creates an opportunity for efficiently enhancing Zn concentration in
440 maize breeding programs. For instance, during early generations of a breeding program, GS can
441 be utilized to identify genotypes with favorable alleles when numbers of progenies and families
442 are large. This could potentially reduce the resource-intensive evaluation process and
443 advancement of false-positive progenies (Velu *et al.* 2016). Coupled with advances in
444 technologies for assessing Zn, plant scientists can more rapidly measure Zn concentration in
445 maize kernels using the energy dispersive x-ray fluorescence (XRF) assays (Guild *et al.* 2017).
446 Thus, with more validations and model refinements, GS can potentially accelerate the breeding
447 process to enhance Zn concentration in maize for a wider range of environments.

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CONCLUSION

450 The moderate to high prediction accuracies reported in this study shows that GS can be
451 used in maize breeding to improve kernel Zn concentration. Assuming two possible seasons of
452 Zn evaluation per year, the predicted genetic gains can be estimated from prediction accuracies

453 and genetic variances of the training populations. The genetic variances for the ZAM panel, DH1
454 and DH2 were 12.38, 12.20 and 14.88, and prediction accuracies were 0.71, 0.70 and 0.51,
455 respectively. If the inbreds in each predicted population are ranked based on their predicted Zn
456 values and the top 10% selected, then their expected average Zn values can be estimated from
457 the proportion of inbreds selected, their respective training population genetic variances,
458 prediction accuracies and the time interval for evaluating the lines. With reference to this, the
459 expected average values of Zn are approximately 31 $\mu\text{g/g}$ for the ZAM panel, 30 $\mu\text{g/g}$ for DH1
460 and 27 $\mu\text{g/g}$ for DH2. These averages are higher than the averages of the respective training
461 populations (~ 27 $\mu\text{g/g}$ for the ZAM panel, ~ 25 $\mu\text{g/g}$ for DH1 and ~ 26 $\mu\text{g/g}$ for DH2) suggesting
462 that the prediction accuracies achieved are sufficient to select at least 10% of the predicted
463 inbreds with higher Zn concentration.

464 The prediction accuracies were of lower quality when genomic predictions were
465 conducted across populations. When the ZAM panel was used as the training population,
466 prediction accuracies for DH1, DH2 and DH1+DH2 were 0.15, -0.10 and 0.09, respectively.
467 When DH1 and DH2 were used as a training and prediction set for each other, prediction
468 accuracies were 0.08 and 0.16 (Unpublished data). These prediction accuracies are considerably
469 lower than those reported in this study and the differences may be attributed to: (i) weak genetic
470 relationships between the training and prediction population sets and (ii) different methods of
471 analysis because the prediction accuracies reported in this study were partly achieved by
472 modelling the random-effects environment structure to account for G x E while for the
473 unpublished data, the random-effects environment structure of G x E was not included.

474 This study also showed that higher prediction accuracies can be achieved when some of
475 the lines are predicted using previous information about them collected from correlated

476 environments. The multi-environment model (M3) which included the interaction between
477 markers, and the environment gave higher prediction accuracy both in CV1 and CV2 for the
478 association panel and DH2 compared with the models which only included main effects (M1 and
479 M2) indicating the importance of accounting for G x E in genomic prediction.

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TABLES

707 **Table 1 Pedigree and average concentration of kernel Zn ($\mu\text{g/g}$) concentration for the parents of**
708 **the DH populations**

DH population	Pedigree	Parent1	Parent2	Zn ($\mu\text{g/g}$)	
				Parent1	Parent2
DH1	CML503/CLWN201	CML503	CLWN201	31.21	22.62
DH2	CML 465/CML451	CML465	CML451	31.55	27.88

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729 **Table 2 Descriptive statistics for kernel Zn concentration in (A) the ZAM panel and (B) DH**
 730 **populations grown in each environment, variance components and broad-and narrow sense**
 731 **heritabilities.**

732 **A**

Population	Population size	Location	Mean ± se (µg/g)	σ_G^2 ^a	σ_{GE}^2 ^a	H^2
ZAM panel	923	Agua Fria 2012	26.15 ± 0.15	12.04	2.42	0.85
		Celaya 2012	25.06 ± 0.14			
		Agua Fria 2013	29.53 ± 0.16			
		Across	26.94 ± 0.10			

733 **B**

Population	Population size	Location	Mean ± se (µg/g)	\hat{h}^2
DH1	112	Celaya 2014	25.38 ± 0.48	0.83
		Tlaltizapan 2015	24.01 ± 0.38	
		Tlaltizapan 2017	24.53 ± 0.37	
		Across	24.65 ± 0.26	
DH2	143	Celaya 2014	27.96 ± 0.39	0.76
		Tlaltizapan 2015	24.08 ± 0.33	
		Tlaltizapan 2017	24.64 ± 0.37	
		Across	25.59 ± 0.22	

734 Broad-sense heritability H^2 of Zn in each environment and across environments

735 Narrow-sense heritability \hat{h}^2 of Zn across environments

736 ^avariance due to genotypes σ_G^2 and the interaction between genotypes and the environment σ_{GE}^2 significant

737 at P<0.001

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741 **Table 3 Phenotypic correlation between environments for kernel Zn**

	DH1	DH 2	ZAM Panel
^a Env1 vs Env2	0.62	0.46	0.63
^a Env1 vs Env3	0.58	0.29	0.66
^a Env2 vs Env3	0.62	0.45	0.61

742 Phenotypic correlation coefficients were significant at $\alpha = 0.001$

743 ^aDH populations; Env 1, Env2 and Env 3=Celaya,2014, Tlaltizapan, 2017 and Tlaltizapan 2017,
744 respectively.

745 ^aZAM panel; Env 1, Env2 and Env 3= Agua Fria, 2012, Celaya, 2012 and Agua Fria 2013, respectively.

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747 **Table 4 Correlations (mean \pm SD) between observed and genomic estimated breeding values for**
748 **kernel Zn in the three environments for three GBLUP models for cross-validations CV1 and CV2 of**
749 **the ZAM panel**

		Prediction accuracy in CV1		
Population	Environment	M1 ^a	M2	M3
	Agua Fria, 2012	-0.01 \pm 0.04	0.33 \pm 0.01	0.34 \pm 0.02
ZAM panel (923)	Celaya, 2012	0.004 \pm 0.04	0.43 \pm 0.01	0.47 \pm 0.01
	Agua Fria, 2013	-0.001 \pm 0.03	0.34 \pm 0.01	0.35 \pm 0.01
	Average	-0.001 \pm 0.03	0.37 \pm 0.01	0.39 \pm 0.01
		Prediction accuracy in CV2		
Population	Environment	^a M1	M2	M3
	Agua Fria, 2012	0.71 \pm 0.00	0.71 \pm 0.00	0.72 \pm 0.00
ZAM panel (923)	Celaya, 2012	0.64 \pm 0.00	0.68 \pm 0.00	0.72 \pm 0.00
	Agua Fria, 2013	0.67 \pm 0.00	0.67 \pm 0.00	0.69 \pm 0.01
	Average	0.67 \pm 0.00	0.69 \pm 0.00	0.71 \pm 0.00

750 ^aModels: M1= Environment +Line; M2 = Environment + Line + Genomic; M3 = Environment + Line +

751 Genomic + Genomic \times Environment

752 **Table 5 Correlations (mean ± SD) between observed and genomic estimated breeding values for**
 753 **Zn in the three environments for three GBLUP models for cross-validation CV1 of DH populations**

Population	Environment	Prediction accuracy in CV1		
		M1 ^a	M2	M3
DH1	Celaya, 2014	-0.05 ± 0.10	0.52 ± 0.04	0.51 ± 0.04
	Tlaltizapan, 2015	-0.02 ± 0.12	0.52 ± 0.05	0.51 ± 0.05
	Tlaltizapan, 2017	-0.01 ± 0.10	0.56 ± 0.05	0.55 ± 0.05
	Average	-0.03 ± 0.10	0.53 ± 0.04	0.52 ± 0.04
DH2	Celaya, 2014	0.05 ± 0.08	0.47 ± 0.03	0.50 ± 0.04
	Tlaltizapan, 2015	0.03 ± 0.08	0.45 ± 0.03	0.45 ± 0.03
	Tlaltizapan, 2017	0.04 ± 0.08	0.35 ± 0.03	0.35 ± 0.04
	Average	0.04 ± 0.06	0.43 ± 0.03	0.44 ± 0.02

754 ^aModels: M1= Environment +Line; M2 = Environment + Line + Genomic; M3 = Environment + Line +
 755 Genomic + Genomic × Environment

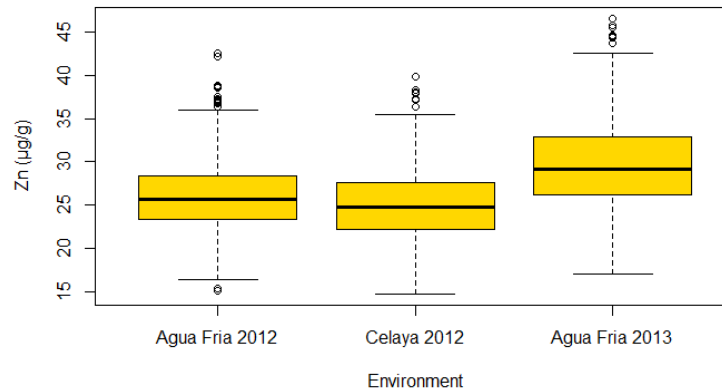
756 **Table 6 Correlations (mean ± SD) between observed and genomic estimated breeding values for**
 757 **Zn in the three environments for three GBLUP models for cross-validation CV2 of DH populations**

Population	Environment	Prediction accuracy in CV2		
		M1 ^a	M2	M3
DH1	Celaya, 2014	0.67 ± 0.02	0.68 ± 0.02	0.68 ± 0.03
	Tlaltizapan, 2015	0.70 ± 0.02	0.71 ± 0.02	0.70 ± 0.02
	Tlaltizapan, 2017	0.67 ± 0.02	0.70 ± 0.02	0.69 ± 0.02
	Average	0.68 ± 0.01	0.70 ± 0.01	0.69 ± 0.01
DH2	Celaya, 2014	0.46 ± 0.016	0.53 ± 0.02	0.56 ± 0.02
	Tlaltizapan, 2015	0.50 ± 0.020	0.55 ± 0.02	0.55 ± 0.02
	Tlaltizapan, 2017	0.40 ± 0.023	0.43 ± 0.02	0.43 ± 0.02
	Average	0.45 ± 0.02	0.50 ± 0.01	0.51 ± 0.01

758 ^aModels: M1= Environment +Line; M2 = Environment + Line + Genomic; M3 = Environment + Line +
759 Genomic + Genomic × Environment

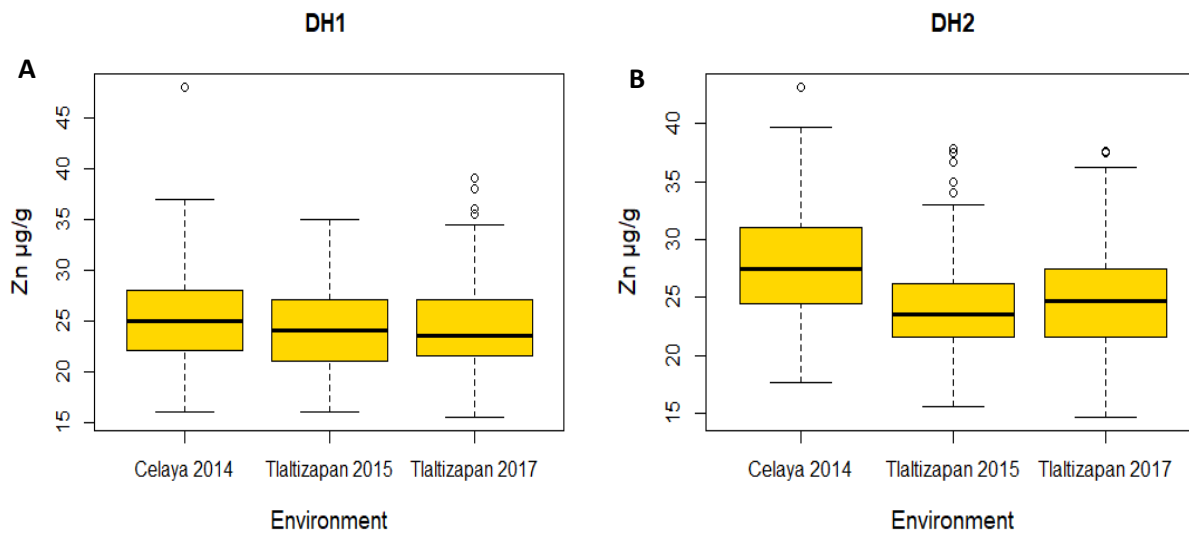
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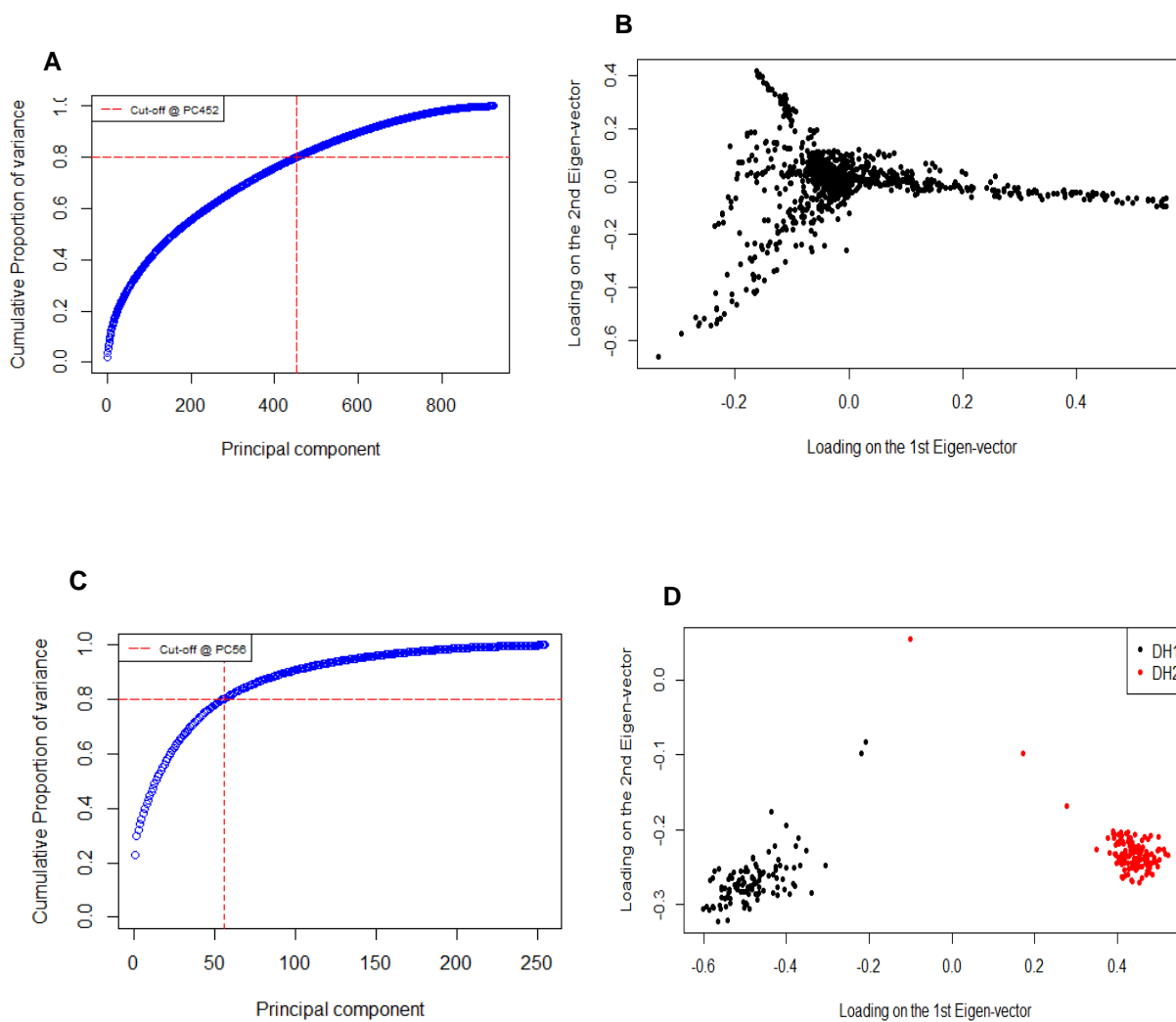
762 **Figure 1** Box plots for kernel Zn (µg/g) in the ZAM panel in three environments (Agua Fria, 2012, Celaya,
763 2012 and Agua Fria, 2013)



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765 **Figure 2** Box plots for kernel Zn (µg/g) for (A) DH1 and (B) DH2 in three environments (Celaya 2014,
766 Tlaltizapan, 2015 and Tlaltizapan, 2017)

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770 **Figure 3** Scree plots (*A and C*) and loadings of the first two eigenvectors (*B and D*) of the covariance
771 matrices derived from markers for the ZAM panel (*A and B*) and for the DH populations (*C and D*)

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