Weak support for the abundant niche-centre hypothesis in North American birds

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10 Abstract

¹¹ Species may be more abundant in the centre of their geographic range or climatic ¹² niche (the abundant-centre hypothesis). Recently, Osorio-Olvera *et al.* (2020) re-

¹³ ported strong support for the niche abundant-centre relationship. We demonstrate

14 here that methodological decisions strongly affected perceived abundant-centre

¹⁵ support. Upon re-analysis, we show that abundant-centre relationships are quite
¹⁶ rare.

The spatial distribution of abundance has long fascinated ecologists who searched 17 for general rules governing where species occur and the density at which they are 18 found (McGill et al., 2007; Sagarin & Gaines, 2002). Particularly controversial 19 rules are the abundant-centre and abundant-niche centre hypotheses, which pre-20 dict abundance to decrease gradually from the centre to the margins of species 21 geographic ranges and ecological niches respectively (Brown, 1984; Pironon et al., 22 2017)). Both theories have received mixed empirical support (Martínez-Meyer 23 et al., 2013; Sagarin & Gaines, 2002; Dallas et al., 2017) and limited theoretical 24 development (Osorio-Olvera et al., 2019; Holt, 2019). Moreover, recent analyses 25 highlighted that tests of these hypotheses were sensitive to the quality of the input 26 data and the methodological approach considered (Santini et al., 2019). 27

Osorio-Olvera et al. (2020) analyze data from the North American Breeding Bird 28 Survey (BBS) to test for a negative correlation between species abundance and the 29 distance to their climatic niche centroid. Counter to recent findings questioning its 30 generalizability (Sagarin & Gaines, 2002; Dallas et al., 2017; Santini et al., 2019), 31 the authors find general support for the hypothesis and propose that the distance 32 to species climatic niche centroid (quantified using minimum volume ellipsoids) 33 could represent a reliable and simple new metric to predict the current and future 34 distribution of species abundance. However, we show that, by (i) considering the 35 full species environmental niche, (ii) reporting non-significant relationships and 36 (iii) selecting models with the best fit only, overall conclusions may differ greatly. 37

To estimate the niche, Osorio-Olvera and colleagues contrast bird occurrence records with 10,000 background points sampled randomly all over the Americas including Canada and South America, i.e., a methodological approach that the authors point in the discussion as a potential reason why previous studies failed to support the abundant-niche centre relationship (Santini *et al.*, 2019). However,

many of the species considered in the study also occur in other continents (e.g. 43 Ardea alba, Corvus corax), and some only share a very small portion of the range 44 in the study area (e.g. *Thalasseus maximus*, *Aramus quarauna*). We calculated 45 geographic and climatic niche overlap of the BBS data with the BirdLife Inter-46 national data (BirdLife International, 2017), demonstrating a clear influence on 47 the estimation of the geographic range and climatic niche boundaries, as well as 48 their centroids (Figure 1). This subsequently affects the abundant-niche centre re-49 lationship, as discussed in Soberón et al. (2018). Oddly, many of the bird species 50 whose geographic ranges are underestimated [mean (sd) of geographic range over-51 lap = 55(27)% and whose niches have been underestimated [mean (sd) of niche 52 overlap = 53(23)%] also exhibit significant negative abundant-centre relationships, 53 putatively supporting the hypothesis. 54

The strongest support for the abundant-niche centre relationships comes from Osorio-Olvera *et al.* (2020) estimating the species niche as a minimum volume ellipsoid (MVE) by considering more than 4000 combinations of climatic variables, including all 19 commonly-used bioclimatic variables together with the first 15 PCA components of a PCA based on the same bioclimatic variables. The authors use every possible combination of two and three niche axes to estimate the niche. We identify two main issues associated with this procedure.

First, the authors report results only for models showing significant abundantniche centre relationships, omitting non-significant correlations (Figure 2a). This issue is not only present in the fit MVE models, but also in the 2 and 3 feature models using convex hulls or MVEs, which makes Figure 3 of Osorio-Olvera et al. (2020) quite a biased view of the abundant-niche centre relationship. In fact, by including non-significant correlations, the mean abundant-niche centre relationship across all model sets becomes quite weak ($\bar{\rho} + -sd = -0.08 + -0.01$),

and many species exhibit significantly positive abundant-niche centre relationships 69 (Figure 2b). Including these non-significant results is important, in our view, and 70 strongly influences the resulting perceived support for the abundant-centre pat-71 tern (Figure 2b). Our re-analysis suggests that only between 37% and 45% of 72 species have negative abundant-centre relationships, regardless of approach used 73 (see https://figshare.com/s/8fadf780810e73d44623), while the majority of the es-74 timated relationships are either positive or non-significant. Interestingly, this low 75 empirical support is consistent with previous findings for the geographic interpre-76 tation of the hypothesis (Pironon *et al.*, 2017; Sagarin & Gaines, 2002). 77

Second, while the authors train an average of 1,852 models per species to cal-78 culate MVEs, they perform no form of model selection. This functionally treats 79 the poorest fit MVE and the best fit MVE as equivalent, provided the model pro-80 duced a significant abundant-centre relationship. When non-significant results are 81 included, and only best models are retained, the overall pattern changes substan-82 tially (Figure 2; but note the niches are still biased to data in North America). 83 When only the best fit models are considered, 115 out of 379 species had significant 84 abundant-niche centre relationships, with a mean correlation coefficient of -0.07. 85 Some of these best models had higher omission rates than what Osorio-Olvera 86 et al. (2020) considered. Removing these models reduces the number of species 87 down to 303 species, of which 94 had significantly negative abundant-niche centre 88 relationships, while 180 and 29 had non-significant or significantly positive rela-89 tionships, respectively (Figure 2c). A large part of the presentation of the results 90 in Osorio-Olvera *et al.* (2020) is dedicated to species that meet the expectations 91 and exhibit higher abundance in areas closer to the centre of their niche (Fig. 92 1 and 2 in Osorio-Olvera et al. (2020)), we note that a less biased overview of 93 their findings shed doubts on the putative support for the abundant niche-centre 94

95 hypothesis.

The study from Osorio-Olvera *et al.* (2020) highlights the timely need for disentangling the complex relationship between species ecological niche, geographic distribution and demographic performance (Holt, 2019; Bohner & Diez, 2020). Explaining the convergence and divergence of results of studies exploring occurrence and abundance patterns is key for improving our understanding of biodiversity and ability to predict its response to ongoing changes in the global environment.

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$_{142}$ Figures

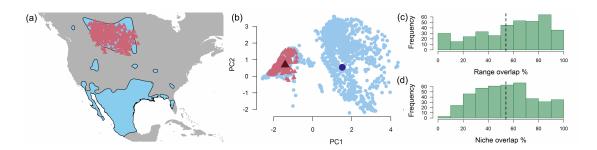


Figure 1: Mismatch in geographic and niche estimate between abundance data used in Osorio-Olvera *et al.* (2020) and the full resident and breeding range distribution of species, estimated IUCN range polygons. We acknowledge, as others previously have Soberón *et al.* (2018), that IUCN polygons may not capture a species actual range. **a**) Sprague's pipit (*Anthus spragueii*) IUCN geographic range (in blue) and sample data to estimate the niche (in red); **b**) First two PCA axes of all bioclimatic variables showing environmental values considered in the study (red triangles) and those estimated considering the cells in the IUCN range (blue dots). The darker and larger triangle and circle represent the estimated centroids of the two hypervolumes; **c**) Distribution of geographic range overlap between convex hulls drawn around abundance estimates and the IUCN ranges for all species considered in the study; **d**) Distribution of niche overlap between convex hulls drawn around abundance estimates and grid cells within the IUCN ranges for all species in the study. Niche overlap and niche centroids were estimated using the hypervolume package Blonder *et al.* (2015).

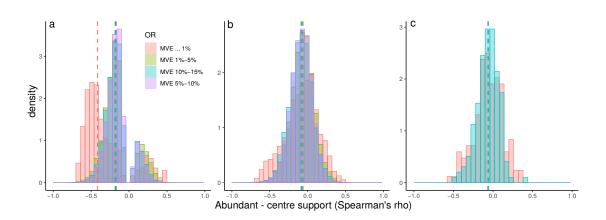


Figure 2: Variation in support for the abundant-niche centre hypothesis across North American birds as a function of analytical decisions involved in the formation of over 700,000 models of species Minimum Volume Ellipsoids (MVEs). **a**) the reproduced results from Osorio-Olvera *et al.* (2020) showing predominantly negative abundant-centre relationships, especially when omission rates were low; **b**) the effect of including non-significant abundant-niche centre relationships. **c**) only considering the best fit MVE model for a given species instead of using all fit models, while also considering that the best fit model could have a non-significant abundant-centre relationship.