

# Is evolution in response to extreme events good for population persistence?

Kelsey Lyberger<sup>1,\*</sup>

Matthew Osmond<sup>1</sup>

Sebastian Schreiber<sup>1</sup>

1. University of California Davis, California 95616

\* Corresponding author; e-mail: [klyberger@ucdavis.edu](mailto:klyberger@ucdavis.edu).

*Manuscript elements:* Figures 1-5, online appendices A, B, C, and D (including supplementary figures S1-S4). All figures are to print in color.

*Keywords:* Extinction, evolutionary rescue, pulse disturbance, maladaptation.

*Manuscript type:* Article.

## Abstract

Climate change is predicted to increase the severity of environmental perturbations, including  
3 storms and droughts, which act as strong selective agents. These extreme events are often of  
finite duration (pulse disturbances). Hence, while evolution during an extreme event may be  
adaptive, the resulting phenotypic changes may become maladaptive when the event ends. Us-  
6 ing individual-based models and analytic approximations that fuse quantitative genetics and  
demography, we explore how heritability and phenotypic variance affect population size and ex-  
tinction risk in finite populations under an extreme event of fixed duration. Since more evolution  
9 leads to greater maladaptation and slower population recovery following an extreme event, slow-  
ing population recovery, greater heritability can increase extinction risk when the extreme event  
is short, as in random environments with low autocorrelation. Alternatively, when an extreme  
12 event is sufficiently long, heritability often helps a population persist, as under a press perturba-  
tion. We also find that when events are severe, the buffering effect of phenotypic variance can  
outweigh the increased load it causes. Our results highlight the importance of the length and  
15 severity of a disturbance when assessing the role of evolution on population recovery; the rapid  
adaptive evolution observed during extreme events may be bad for persistence.

## Introduction

18 Globally, humans are causing substantial environmental perturbations, and these perturbations  
are likely to become more severe in the future. In particular, climate change is projected to  
lead to more extreme weather events, including droughts and major storms (Ummenhofer and  
21 Meehl, 2017). With more severe events comes the potential for dramatic demographic and genetic  
consequences.

In the process of causing mass mortality, extreme events can act as catalysts of evolutionary  
24 change. In fact, there are many examples of rapid evolution in response to extreme events  
(reviewed in Grant et al., 2017). Famously, Bumpus (1899) documented phenotypic differences  
in house sparrows that survived a strong winter storm. More recently, Donihue et al. (2018)  
27 measured lizards before and after a series of hurricanes and found evidence for selection on  
body size, relative limb length, and toepad size. Another example is a study of the annual  
plant *Brassica rapa* in response to summer drought, in which, post-drought seeds flowered earlier  
30 when planted alongside pre-drought seeds (Franks et al., 2007). Finally, Grant and Grant (2014)  
not only documented shifts in beak depth of Darwin's ground finches in response to drought,  
but also the reversal of that evolution and population recovery in subsequent years. We have  
33 many fewer examples like this latter case, where the recovery from an extreme event is recorded.  
Hence exploring what factors influence recovery patterns is currently best done with theory.

Extreme events such as storms, hurricanes, and droughts are pulse disturbances, defined as  
36 a transient or short-term change in the environment (Bender et al., 1984). In the ecological lit-  
erature, pulse disturbances lie at the crossroads of two other forms of environmental change:  
press perturbations, a permanent or long-term change in the environment (Ives and Carpen-  
39 ter, 2007; Kéfi et al., 2019; Yodzis, 1988), and fluctuating environments such as a sequence of  
pulse disturbances of varying durations and intensities (Lande, 1993; Ozgul et al., 2012). De-  
spite their transient nature, pulse disturbances can substantially impact ecological systems from  
42 large, long-term transients in ecological dynamics to permanent shifts in ecological states (Fox

and Gurevitch, 2000; Hastings et al., 2018; Holling, 1973, 1996; Holt, 2008; Ives and Carpenter, 2007). While the most extreme form of a permanent shift is species extinction, we know of no  
45 studies that examine the effects of pulse disturbances of varying length on extinction risk (see, however, Figure 1 in Holt, 2008). In contrast, there have been many studies examining the effect of repeated disturbances on extinction risk (Cuddington and Yodzis, 1999; Mangel and Tier, 1994;  
48 Petchey et al., 1997). For example, Wilson and MacArthur (1967) showed that the mean time to extinction plateaus as a function of the initial population size. This plateau, which allows one to define a minimum viable population size, is lost when populations experience repeated large  
51 disturbances; there is a continual gradual increase in the mean time to extinction with increasing initial population sizes (Mangel and Tier, 1994). In these studies of repeated disturbances, extinction occurs eventually. Consequently, the focus is on when the extinction event occurs, not on  
54 whether it occurs after a particular disturbance. However, understanding short-term extinction risk after a single disturbance is critical for conservation and management.

Previous work on evolution in changing environments can provide intuition for how evolu-  
57 tion might affect extinction risk during or after a pulse disturbance. One focus of the evolutionary rescue literature has been on understanding the consequences of phenotypic change in the context of a sudden, long-term or permanent environmental shift (a press perturbation). These  
60 studies, some of which account for demographic stochasticity, underline the importance of genetic variance for increasing the probability of rescue (Gomulkiewicz and Holt, 1995, reviewed in Alexander et al., 2014; Bell, 2017). That is, populations that are able to adapt rapidly to the  
63 new environment have a higher chance of persisting. Similarly, studies of adaptation in fluctuating environments suggest that if an environment is predictable, such as the case of positively autocorrelated fluctuations, genetic variation reduces lag load (Charlesworth, 1993; Lande and  
66 Shannon, 1996; Chevin, 2013). This reduction in the lag load leads to higher population per-capita growth rates and, consequently, is expected to reduce extinction risk. Whereas, when the environment is unpredictable, genetic variance typically increases the lag load. These studies cal-  
69 culated lag load and per-capita growth rates in a number of environmental contexts including a

press perturbation, randomly fluctuating environments, and cyclic environments. However, they did not account for demographic stochasticity, the ultimate cause of extinction. Furthermore, we explore the impact of phenotypic variance on the probability of persistence, which has not been emphasized as much in previous studies.

To understand extinction risk during and following a pulse disturbance, we introduce an individual-based model that fuses population demography with quantitative genetics. Using a mixture of computational and analytical methods, we examine how phenotypic variation and the heritability of this variation influences population growth, lag load, and extinction risk during and following a pulse perturbation. Moreover, we examine how the magnitude and direction of these effects depend on the duration and intensity of the pulse perturbation.

## Model

We use an individual-based model that combines the infinitesimal-model of an evolving quantitative trait with density-dependent demography. To gain insights beyond simulating the model, we derive analytical approximations of the probability of extinction using a mixture of deterministic recursion equations and branching process theory (Harris, 1964). We assume discrete, non-overlapping generations. The life cycle starts with viability selection. In each generation  $t$ , we impose stabilizing selection around some optimal trait value  $\theta_t$ , which is set by the environment in that generation, by making the probability of survival

$$s_t(z) = \exp \left[ \frac{-(\theta_t - z)^2}{2\omega^2} \right], \quad (1)$$

a Gaussian function of phenotype,  $z$ , with a strength of selection proportional to  $1/\omega^2$ .

Following viability selection, survivors are randomly drawn with replacement to form mating pairs. Each pair then produces a Poisson number of offspring with mean  $2\lambda$ . The population lives in a habitat that supports at most  $K$  individuals. Hence, if more than  $K$  offspring are produced,  $K$  are randomly chosen without replacement. The genetics of the population follows the infinitesimal model in which breeding values are determined by many loci of small effect

(Fisher, 1918; Turelli, 2017). Under this model, an offspring's breeding value is a draw from a normal distribution centered on the mean of its parents' breeding values and with segregation  
96 variance  $V_0$  (which we assume is a constant). Its phenotype,  $z$ , is this breeding value,  $g$ , plus a random environmental component,  $e$ , which is a draw from a normal distribution with mean 0 and variance  $V_e$ . We ignore dominance and epistasis, thus the phenotypic variance in generation  $t$   
99 is the additive genetic variance plus the environmental variance,  $V_{p,t} = V_{g,t} + V_e$ . At equilibrium,  $\widehat{V}_p = \widehat{V}_g + V_e$ .

Prior to experiencing an extreme event, the populations in the individual-based simulations  
102 start with a 100-generation burn-in from an initial state where all  $N = K$  individuals have breeding value  $\theta = 0$  and the optimal trait value  $\theta_t$  equals 0 throughout this period. The 100 generation burn-in is sufficiently long to ensure the model reaches a quasi-stationary state (Supplementary  
105 Figure S1). To model the extreme event of length  $\tau$  after the burn-in period from generation -100 to 0, we increase the optimum trait value by  $\Delta\theta$  and revert it back to its original value after  $\tau$  generations. For example, in a single-generation event, the optimum trait value changes before  
108 selection in generation 1 and then reverts back before selection in generation 2 (Figure 1). Unless otherwise stated, we use the parameter values  $\omega^2 = 1$ ,  $\widehat{V}_p = 1$ ,  $\Delta\theta = 3$ ,  $\lambda = 2$ , and  $K = 500$ . These  $\omega^2$  and  $\widehat{V}_p$  values represent strong selection and large phenotypic variance relative to those  
111 estimated in Turelli (1984), which we use to show the qualitative effect of variance load. Reducing the strength of selection (or decreasing the phenotypic variance) does not otherwise change our qualitative results. The values for  $\Delta\theta$  scale the strength of selection. For this set of parameter  
114 values, the optimum shift ( $\Delta\theta = 3$ ) corresponds to three standard deviations beyond the mean of the trait distribution, and consequently, roughly 99.5 percent of the trait distribution will initially be smaller than the optimum and we expect roughly 80% of the population to die in the  
117 first generation. We have chosen a high growth rate,  $\lambda$ , to reduce extinction from demographic stochasticity in the absence of disturbance. We have chosen a large enough starting population size and carrying capacity,  $K = 500$ , to make approximations reasonable (e.g. normal distribution  
120 of traits).

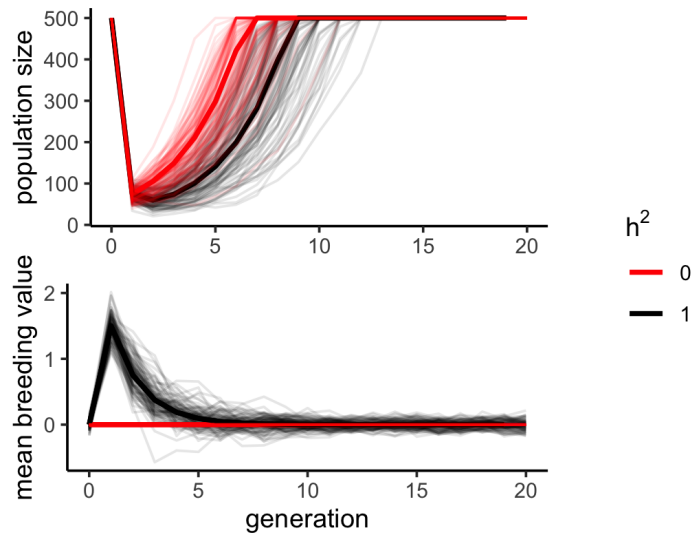


Figure 1: Population size over time for populations with  $h^2 = 0$  (red) and  $h^2 = 1$  (black) after a single-generation extreme event of size  $\Delta\theta = 3$ . Phenotypic variance is the same for both populations ( $\widehat{V}_p = 1$ ). Faded lines are 100 simulations and solid lines are the model predictions using Equations (2) and (4). Parameters:  $\omega = 1$ ,  $\lambda = 2$ . Red:  $V_0 = 0$ ,  $V_e = 1$ , Black:  $V_0 = 3/4$ ,  $V_e = 0$ .

## Approximations

### *Approximating the evolutionary and population size dynamics*

123 In Appendix A (see the supplementary Mathematica file for more details), we derive determin-  
 istic approximations for the dynamics of the mean breeding value  $\bar{g}_t$ , genetic variance  $V_{g,t}$ , and  
 population size  $N_t$ . Briefly, if we assume the distribution of breeding values remains normally  
 126 distributed, then we know the whole phenotypic distribution by tracking the mean and variance  
 in the breeding values. Given the mean and variance in a given generation, we can then calculate  
 the mean and variance in the next generation,

$$\bar{g}_{t+1} = \bar{g}_t \left(1 - \frac{V_{g,t}}{V_t}\right) + \theta_t \frac{V_{g,t}}{V_t} \quad (2)$$

129

$$V_{g,t+1} = \frac{V_{g,t}V_s}{V_t} \frac{1}{2} + V_0, \quad (3)$$

where  $V_t = V_{g,t} + V_s$ ,  $V_s = \omega^2 + V_e$  is the inverse of the effective strength of selection, and  $V_0$  is the variance in breeding values among siblings. We can also calculate the population size in the  
 132 next generation

$$N_{t+1} = \min(N_t \bar{s}_t \lambda, K), \quad (4)$$

where the mean survival probability,  $\bar{s}_t$ , is calculated by integrating Equation (1) over the distribution of phenotypes in the population,

$$\bar{s}_t = \sqrt{\omega^2/V_t} \exp[-(\theta_t - \bar{g}_t)^2/(2V_t)]. \quad (5)$$

135 Regardless of the trait or environmental dynamics, the genetic variance approaches an equilibrium  $\hat{V}_g = (2V_0 - V_s + \sqrt{4V_0^2 + 12V_0V_s + V_s^2})/4$  which increases with segregation variance and decreases with the strength of selection. In a constant environment,  $\theta_t = \theta$  for all  $t$ , the mean  
 138 breeding value approaches the optimum,  $\hat{g} = \theta$ , and, provided  $\lambda > 1$ ,  $N_0 \bar{s}_0$  is large enough, and  $\hat{V}_p = \hat{V}_g + V_e$  is small enough, the population size reaches carrying capacity,  $\hat{N} = K$ . Starting from this equilibrium, we can then approximate the response of the population to a shift in the  
 141 optimum using Equations (2)-(4).

### *Approximating Extinction Risk*

We next approximate the probability of extinction using branching processes (Harris, 1964). The  
 144 probability generating function for the number of offspring produced by an individual with survival probability  $s$  is

$$f(x, s) = 1 - s + s \exp[-(1-x)\lambda]. \quad (6)$$

The probability of no offspring is  $f(0, s)$ . Further, if  $s_1, \dots, s_{N_t}$  are the survival probabilities of the  
 147  $N_t$  individuals in generation  $t$ , then the probability of extinction in generation  $t$  is  $\prod_{i=1}^{N_t} f(0, s_i)$ . Here we approximate this by assuming all individuals in generation  $t$  have the average probability of survival,  $\bar{s}_t$ , which is a reasonable approximation when the strength of selection is weak



150 relative to the phenotypic variance. Defining  $f_t(x) = f(x, \bar{s}_t)$ , the probability of extinction in  
generation  $t$  is then simply  $f_t(0)^{N_t}$ . Assuming that the effects of density-dependence are negli-  
gible from generation  $t$  to generation  $T > t$ , we can approximate the probability of extinction by  
153 the end of generation  $T$  as  $(f_t \circ f_{t+1} \circ \dots \circ f_T(0))^{N_t}$ , where  $\circ$  denotes function composition (Harris,  
1964).

We take  $t = 1$  to be the first generation of the extreme event and assume the population  
156 begins at carrying capacity. For an extreme event of duration  $\tau$ , we define

$$P_{\text{extinct}}(\tau, T) = (f_1 \circ f_2 \circ \dots \circ f_T(0))^K \quad (7)$$

as our approximation for the probability of extinction by generation  $T$  since an extreme event of  
length  $\tau$  began. To calculate the  $\bar{s}_t$  in Equation (7) we assume  $V_{g,t} = \hat{V}_g$  and use Equation (2) to  
159 get  $\bar{g}_t$ , which together give  $\bar{s}_t$  (Equation (1)).

## Results

### *Demographic recovery*

162 We first explore extreme events lasting a single generation. To characterize the impact of pheno-  
typic variance and heritability on population size, we compare the demographic response of pop-  
ulations with low or high phenotypic variance,  $\hat{V}_p$ , across a range of heritabilities,  $h^2 = \hat{V}_g / \hat{V}_p$ .  
165 During the event, heritability has no effect on population size (compare black and red curves  
in 2A). In contrast, we see that phenotypic variation can have a large effect. A population with  
high phenotypic variance (thick gray curve) has a smaller population size than one with low  
168 phenotypic variance (thin gray curve) immediately following a low severity extreme event, but  
a higher population size following more severe events. We also see this effect in the generation  
after the event (Figure 2B,C). This pattern stems from the dual role of phenotypic variance, in  
171 that it both increases variance load and contributes individuals with extreme traits who are then  
able to survive an extreme event. High phenotypic variance therefore reduces both mean fitness

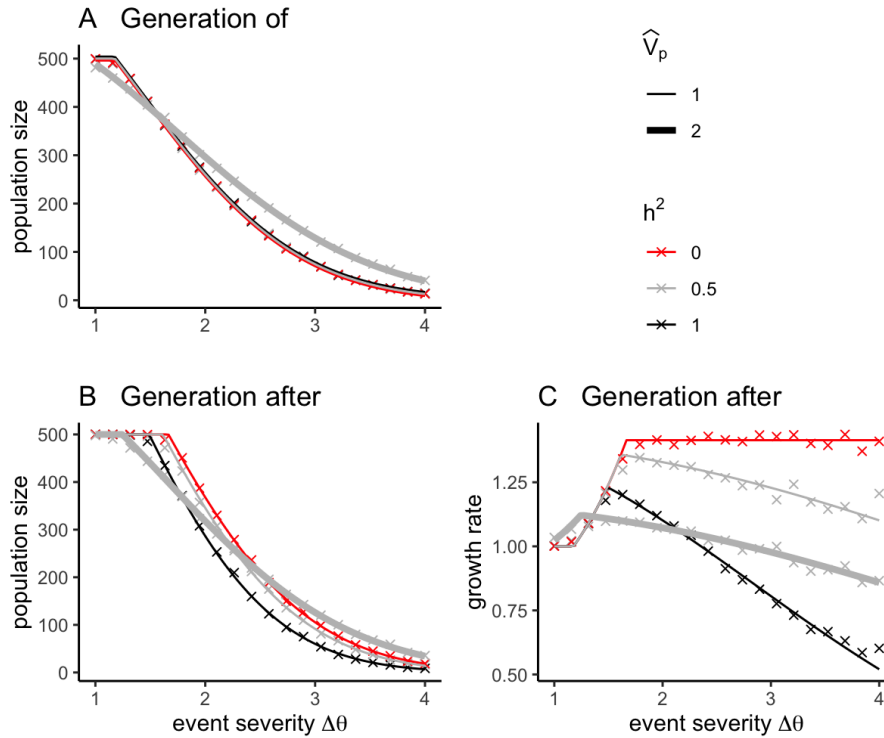


Figure 2: Population size response during the generation of a single-generation extreme event (A), the generation after the event (B), and the growth rate  $(\frac{N_t}{N_{t-1}})$  calculated as the population size in the generation after the event divided by the population size in the generation of (C) shown over a range of event severities  $\Delta\theta$ . Expectations using Equation (4) as curves and simulation results (mean of 100 replicates) as crosses. Parameters:  $\omega = 1, \lambda = 2$ . Red:  $V_0 = 0, V_e = 1$ . Thick gray:  $V_0 = 2/3, V_e = 1$ . Thin gray:  $V_0 = 5/16, V_e = 1/2$ . Black:  $V_0 = 3/4, V_e = 0$ .

within a generation and the variance in fitness across generations – a form of short-term bet-  
 174 hedging which can increase the geometric mean of fitness in the generations during and after  
 the the disturbance event. The positive effect of bet hedging is seen when the event is severe and  
 variation means more individuals on the tail of the distribution will survive the event.

177 While heritability has no effect on survival during the event, it has a strong effect on pop-  
 ulation recovery in subsequent generations. In particular, heritability dampens the growth rate  
 $(\frac{N_t}{N_{t-1}})$  in subsequent generations (Figure 2C) as evolution in the generation of the event induces

180 future maladaptation. This explains why in the generation after the event increasing segregation  
variance increases population size (thick gray curve crosses red near  $\Delta\theta = 3$  in panel B) at a  
higher severity than the point which increasing environmental variance becomes beneficial (thick  
183 gray curve crosses black near  $\Delta\theta = 2$  in panel B). The maladaptation induced by heritability con-  
tinues past the generation after the event, generally slowing population recovery (Figure 1). In  
conclusion, phenotypic variance can be beneficial for population growth under single-generation  
186 severe events, but heritability is generally deleterious.

### *Extinction Risk*

When a single-generation extreme event is severe enough, increasing phenotypic variation lowers  
189 extinction risk both during and after the event (compare thick and thin gray curves in Figure  
3A,C). The biological intuition behind this pattern is the same as in Figure 2A, where increased  
variance means more individuals survive the extreme event. However, at such large population  
192 sizes the extinction risk is essentially zero during a mild event. In other words, while having too  
much variance leads to considerable reduction in population size when events are mild, it is very  
unlikely to lead to extinction unless there is extremely high phenotypic variance or if carrying  
195 capacity is very low. In the former case load will cause extinction in the absence of extreme  
events (Supplementary Figure S2).

Next, we compare populations with the same phenotypic variance but different heritabilities,  
198 to control for the effect of variance (i.e., variance load and bet hedging) and isolate the effect of  
evolution (compare black and red in Figure 3). When the extreme event lasts only one generation  
(Figure 3A,C), heritability increases extinction risk in the generation following moderately severe  
201 extreme events. There is little effect when events are very mild or incredibly strong. Whereas,  
when an extreme event lasts two generations (Figure 3B,D), heritability reduces the risk of ex-  
tinction in the generation following a moderately severe extreme event.

204 Finally, we explored how extinction risk varies across time for one- to four-generation mod-  
erately ( $\Delta\theta = 3.5$ ) extreme events across a range of heritabilities. For single generation events,

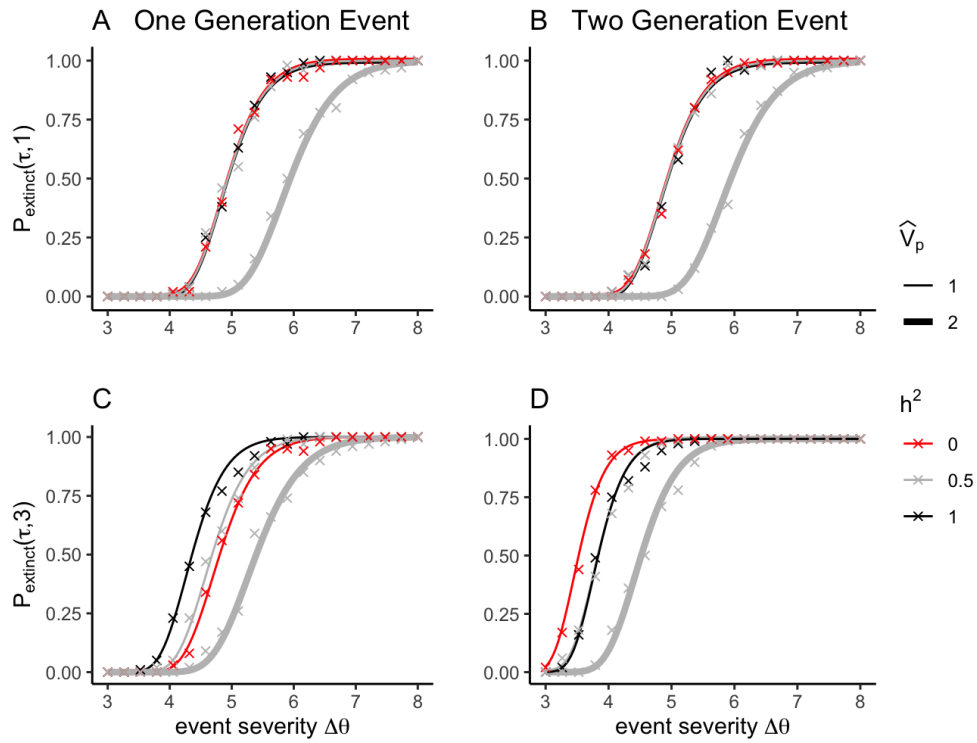


Figure 3: Extinction risk across increasingly severe events in the first generation of an extreme event, (A,B) and two generations later, (C,D). In A and C, the extreme event persists for a single generation, and in B and D, the extreme event persists for two generations. Expectations using Equation (7) as curves and simulation results (mean of 100 replicates) as crosses. Parameters:  $\omega = 1$ ,  $\lambda = 2$ . Red:  $V_0 = 0$ ,  $V_e = 1$ . Thick gray:  $V_0 = 2/3$ ,  $V_e = 1$ . Thin gray:  $V_0 = 5/16$ ,  $V_e = 1/2$ . Black:  $V_0 = 3/4$ ,  $V_e = 0$ .

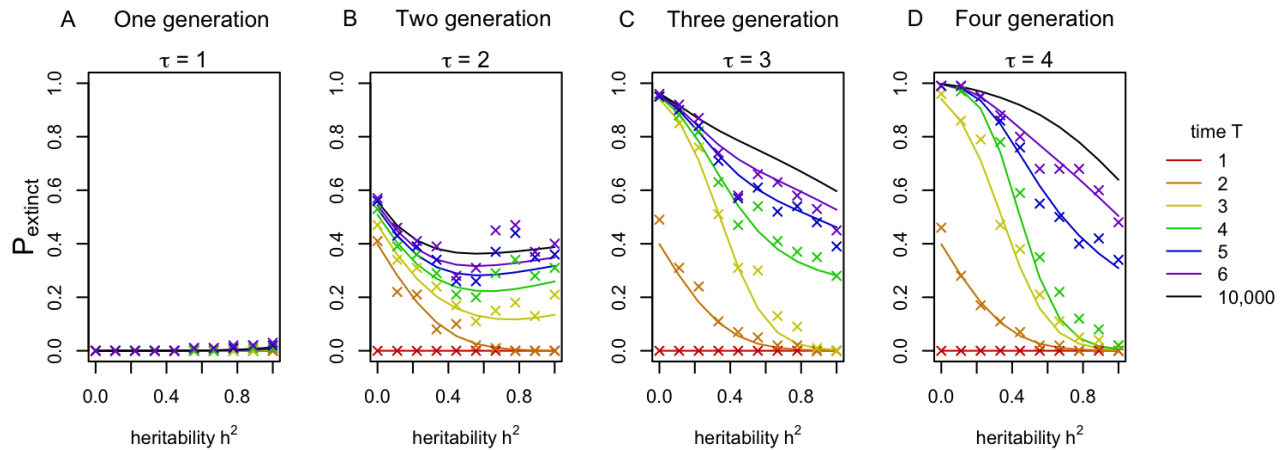


Figure 4: Extinction risk through time  $T$  across a range of heritability for extreme events lasting 1, 2, 3, or 4 generations. Time starts the generation the event began. Parameters:  $\hat{V}_p = 1$ ,  $\omega = 1$ ,  $\lambda = 2$ ,  $\Delta\theta = 3.5$ . Expectations using Equation (7) as curves and simulation results (mean of 100 replicates) as crosses.

long-term extinction risk (10,000 generations) increases with heritability (Figure 4A), for the reasons above. However, for two generation events, long-term extinction risk is lowest at intermediate heritabilities (Figure 4B). And for three and four generation events, long-term extinction risk decreases with heritability (Figure 4C,D). These patterns hold for milder ( $\Delta\theta = 2.5$ ) and more severe ( $\Delta\theta = 4.5$ ) extreme events (Supplementary Figures S3-S4).

While Equation (7) gives a good approximation of extinction risk, the function itself is too complex to give us intuition. Next, by writing down the geometric mean fitness of a population, we reproduce the general trends in long-term extinction risk, but with added clarity for how maladaptation contributes to these outcomes.

### *Contribution of Lag Load*

To better understand how evolution affects the probability of extinction, we approximate the geometric mean fitness of a population under the assumption that the genetic variance remains

at the equilibrium value,  $\widehat{V}_g$ , as expected based on Equation (3). If the extreme event lasts  $\tau$  generations, then the geometric mean of fitness after  $T > \tau$  generations is

$$\begin{aligned} \overline{W}(\tau, T) &= \left( \prod_{t=1}^T \lambda \bar{s}_t \right)^{1/T} \\ &= \lambda \sqrt{\omega^2 / \widehat{V}} \exp \left[ -\frac{1}{2\widehat{V}T} \sum_{t=1}^T (\theta_t - \bar{g}_t)^2 \right]. \end{aligned} \quad (8)$$

From Equation (8) we see that geometric mean fitness depends on the cumulative lag load,  $\sum_{t=1}^T (\theta_t - \bar{g}_t)^2 / (2\widehat{V})$ .

Using Equation (2), we show in Appendix C that the cumulative lag load over  $T > \tau$  generations for an event of length  $\tau$  is

$$\begin{aligned} L(\tau, T) &= \frac{1}{2\widehat{V}} \sum_{t=1}^T (\theta_t - \bar{g}_t)^2 \\ &= \frac{\Delta\theta^2 [1 - (1-v)^\tau] [2 - (1-v)^{2(T-\tau)} + (1-v)^{2T-\tau}]}{2\widehat{V} (2-v)v}, \end{aligned} \quad (9)$$

where  $v = \widehat{V}_g / \widehat{V}$  is a measure of evolvability (see Equation (2) and, e.g., equation 1 in Charlesworth 1993).

Taking the limit as time,  $T$ , goes to infinity, the cumulative lag load is

$$\begin{aligned} L_\infty(\tau) &= \lim_{T \rightarrow \infty} L(\tau, T) \\ &= \frac{\Delta\theta^2 [1 - (1-v)^\tau]}{\widehat{V} (2-v)v}. \end{aligned} \quad (10)$$

Generally, increasing the event length or increasing the event severity increases the cumulative lag load.

As  $v = h^2 \widehat{V}_p / (\widehat{V}_p + \omega^2)$ , we can use Equation (10) to determine how heritability affects the cumulative lag load in the long term (Figure 5), holding  $\widehat{V}_p$  and  $\omega^2$  (and thus the variance load) constant. When the extreme event only lasts one generation ( $\tau = 1$ ), the cumulative lag load equals  $\frac{\Delta\theta^2}{2(\widehat{V}_p + \omega^2) - h^2 \widehat{V}_p}$ . Hence, increasing heritability while holding  $\widehat{V}_p$  constant increases the cumulative lag load (solid purple curve in Figure 5), a trend consistent with the extinction probabilities for  $\tau = 1$  (Figure 4A). Alternatively, when the extreme event lasts two generations ( $\tau = 2$ ),

the cumulative lag load equals  $\frac{\Delta\theta^2}{\widehat{V}_p + \omega^2}$  and is therefore independent of heritability when  $\widehat{V}_p$  and  $\omega^2$  are held constant (solid pink curve in Figure 5). Finally, when the extreme event lasts for  
 237 more than two generations, the cumulative lag load is a decreasing function of heritability (yellow and green curves in Figure 5), a trend consistent with extinction probabilities decreasing with heritability when  $\tau \geq 3$  (Figure 4C,D).

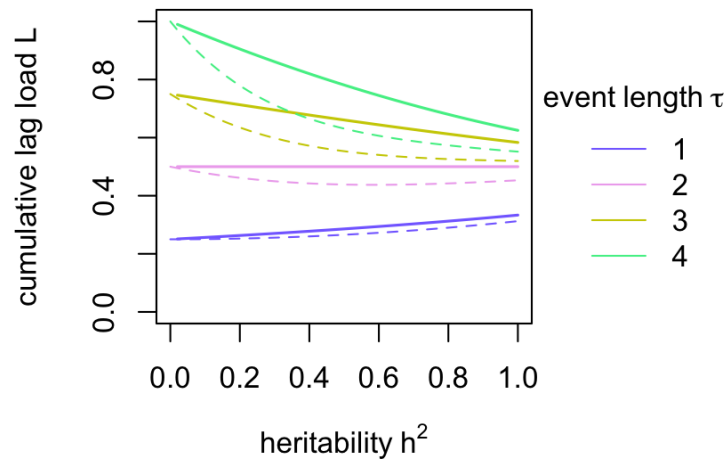


Figure 5: Cumulative lag load as a function of heritability. Dashed curves are Equation (9) with  $T = \tau + 1$  and solid are Equation (10). Colors correspond to the length of the extreme event. Parameters:  $\widehat{V}_p = 1$ ,  $\omega = 1$ ,  $\Delta\theta = 1$ .

240 Taking the limit as both the time and event length go to infinity in Equation (9) and assuming weak selection ( $V_s \rightarrow \infty$ ) we recover the cumulative lag load following a sudden non-reversing shift in the environment,  $\frac{\Delta\theta^2}{4V_g}$  (equation 10 in Chevin, 2013). This is roughly half of Equation (10),  
 243 since the environment never reverts back. Equation (9) can therefore be seen as a generalization of Chevin's result to arbitrary time and event length under arbitrarily strong selection (provided our approximations in Appendix A hold).

246

## Discussion

Although it has been long recognized that evolution may affect a population's response to a changing environment, previous studies have primarily focused on understanding this effect over  
249 the long term in the context of a single non-reversing environmental shift (a press disturbance) or  
a constantly fluctuating environment. Here, we are concerned with the short-term effect (finite  $T$ )  
of a pulse disturbance on population growth and extinction risk. By allowing pulses to be of any  
252 duration, this allows us to connect our results with those of both the press disturbance (large  $\tau$ )  
literature and the constantly changing environment literature while providing insights into the  
transient effects (small  $T$ ) following a disturbance. Our results provide two general conclusions  
255 about the effect of trait variation and its heritability on population growth and extinction risk  
during and following a pulse disturbance. First, trait variance, whether it is heritable or not,  
is a double-edged sword: adding a variance load due to stabilizing selection, yet providing  
258 individuals with more extreme traits who can survive large shifts in the environment. Second,  
while variance can be useful in the generation of a severe event, if heritable it slows demographic  
recovery and can therefore increase extinction risk in the generations after the event.

261

### *Phenotypic Variance*

Phenotypic variance, whether heritable or not, can be beneficial or deleterious. A simultaneous  
reduction in the mean and variance in fitness during the generations immediately prior, during,  
264 and immediately after an extreme event can increase the geometric mean of fitness during this  
time frame (Figure 2). This increase occurs when disturbances are sufficiently severe, in which  
case phenotypic variation can serve as a kind of short-term bet-hedging strategy. In addition to  
267 its effect on the geometric mean across multiple generations, variation in survival rates due to  
phenotypic variation, in and of itself, reduces variation in the total number of offspring produced  
by the population (Kendall and Fox, 2002) and, thereby, lowers extinction risk (Lloyd-Smith et al.,  
270 2005). This effect contributes to phenotypic variation reducing extinction risk.



Prior studies of evolutionary rescue have emphasized the beneficial aspect of genetic variance, but not non-heritable phenotypic variance, in rescuing a population from an abrupt shift in environment. For example, for constantly fluctuating environments, [Charlesworth \(1993\)](#) found that higher genetic variance reduces lag load when environmental fluctuations are large and, thereby, increases the long-term geometric mean of fitness. Similarly, studies of a sudden or gradual directional environmental shift found that high genetic variance at the time of the environmental shift promotes rescue ([Alexander et al., 2014](#); [Barfield and Holt, 2016](#); [Bell and Collins, 2008](#); [Gomulkiewicz and Holt, 1995](#)). Here, by teasing out the effects of heritability and phenotypic variance, we emphasize the costs and benefits of each.

### *Heritability*

Contrary to evolutionary rescue for populations experiencing a press-perturbation ([Barfield and Holt, 2016](#); [Gomulkiewicz and Holt, 1995](#)), we find that heritability increases extinction risk when pulse perturbations only last a single generation. We can gain some intuition for why this is by considering the limiting cases of traits not evolving versus tracking the optimal trait perfectly with a one generation lag. When the population is adapted to the original environment, but does not evolve in response to the extreme event, it experiences a reduction in fitness for the duration  $\tau$  of the extreme event. In contrast, when selection tracks the optimal trait with a one generation lag, the population experiences a reduction in fitness only in the first and last generation of the extreme event. Hence, when the extreme event lasts one generation, extinction risk is higher for the evolving populations and when the extreme event lasts more than two generations, extinction risk is higher for the non-evolving populations. A similar understanding can be gained by adapting a classic population genetic model of allele frequency change with time-varying selection ([Dempster, 1955](#); [Felsenstein, 1976](#), see Appendix D).

In general, the trends in short-term extinction risk are parallel to the lag load predictions (Figures 4 and 5). However, they differ in two ways. When the extreme event lasts exactly two generations, the non-evolving population experiences the reduction in fitness in successive

297 generations while the evolving population experiences this reduction in alternate generations.  
Hence, the evolving population is slightly less likely to go extinct (see Appendix B). Also, when a  
population exhibits an intermediate amount of tracking of the optimum, the variance in survival  
300 from year to year is reduced and therefore can lower the overall extinction probability. This effect  
is especially apparent in the result of two year events (Figure 4B).

While previous studies continuously varying environments have focused on large popula-  
303 tions in the long term, calculating lag load and growth rates when rare, they provide intuition  
for our results on short-term extinction risk after a one-time event. A single-generation extreme  
event functions most like a negatively autocorrelated fluctuating or randomly fluctuating envi-  
306 ronment, in that a strong genetic response to selection in one generation is likely maladaptive  
in the next generation (Benaim and Schreiber, 2019; Charlesworth, 1993; Chevin, 2013; Lande  
and Shannon, 1996). However, an extreme event lasting three or more generations acts like a  
309 positively autocorrelated environment in that the environment is more predictable and hence  
evolvability is favored. Cyclic oscillations with a high amplitude and long period would also fall  
into this category.

### 312 *Future Challenges and Directions*

Our models include a number of simplifications to both evolutionary and demographic pro-  
cesses. First, we do not model the erosion of genetic variance with decreasing population size,  
315 which is expected due to greater genetic drift in smaller populations (Barfield and Holt, 2016;  
Lande and Barrowclough, 1987). Furthermore, we have limited our analysis to truly quantitative  
genetic traits (i.e. infinitely many small-effect alleles) where adaptation is not mutation-limited  
318 and evolution is easily reversed. Different genetic architectures, such as a few loci of large ef-  
fect, likely will respond differently (Barghi et al., 2020). Second, in our model, the phenotypic  
variation due to environmental variation is random, which ignores the potential for phenotypic  
321 plasticity. Phenotypic plasticity has been shown to have variable effects on evolution and ex-  
tinction risk that depend on the nature of environmental change (Kopp and Matuszewski, 2014;

Lande, 2015). Third, we are only tracking a single trait, whereas extreme events likely select on  
324 many correlated traits. As genetic covariance can change the outcome of selection, further work  
is needed to explore the effects of multiple correlated traits. Fourth, we used the simplest possible  
model for density-dependence, the ceiling model, as used in previous evolutionary rescue  
327 studies (e.g., Bürger and Lynch, 1995). For other models of compensating density-dependence,  
such as the Beverton-Holt model (Beverton and Holt, 1957), we expect similar results. However,  
over-compensatory density-dependence, as seen in the Ricker model (Ricker, 1954), can result  
330 in oscillatory population-dynamics for which the timing of the extreme event relative to the  
oscillations may play a subtle role.

Our results call for the need of more empirical studies assessing trait and fitness changes after  
333 an extreme event has ended. The many case studies of evolution in response to extreme events  
focus on the adaptive nature of species responses in the short-term (e.g. Campbell-Staton et al.,  
2017; Coleman et al., 2020). What these studies often fail to mention is that this evolution can  
336 be maladaptive in the longer term. When the environment returns to normal, populations with  
shifted trait means could be worse off. To explore this effect, future empirical studies could be  
extended to track changes in trait values and population size over several generations following  
339 extreme events. For example, lizards can be tracked for several generations following a hurricane  
(Donihue et al., 2018) or a cold snap (Campbell-Staton et al., 2017). We highlight the Darwin's  
finch example as one such study to do this (Grant and Grant, 2002), where finch traits and  
342 selection gradients were found to fluctuate in response to extreme events that lasted less than  
a generation. Larger beak sizes were selected for immediately following a drought due to the  
change in the seed composition, but in later years these beak sizes were maladaptive. However,  
345 because mean survival is higher in an average wet year regardless of beak size, the effect of this  
maladaptation on population recovery is not readily apparent. Ideally, this recovery could be  
compared to one causing similar mortality but less evolution.

348 An important next step will be to understand evolution and extinction risk under repeated  
extreme events. Extreme events, or catastrophic events, can be characterized by causing abrupt,

infrequent, and large reductions in biomass or population size. Hence, prior work on adap-  
351 tation and persistence using autoregressive processes to model environmental fluctuations (e.g.,  
[Benaïm and Schreiber, 2019](#); [Charlesworth, 1993](#); [Chevin, 2013](#); [Lande and Shannon, 1996](#)), do not  
accurately capture the nature of extreme events such as those presented in the continuous-time  
354 ecological models of [Mangel and Tier \(1994\)](#). We hope future studies exploring the impact of dis-  
turbance regime on evolution and extinction risk will benefit from the detailed understanding,  
like that provided here, of an evolving population's response to a single extreme event.

## 357 Appendix A Dynamics of the breeding value distribution and population size

Let the trait value of an individual be the sum of a genetic component (breeding value) and  
360 an environmental component,  $z = g + e$ . Assume we start with a population, in generation  $t$ ,  
that has a normal distribution of breeding values,  $p_g(g, t)$ , with mean  $\bar{g}_t$  and variance  $V_{g,t}$ . And  
assume each environmental component is independently chosen from a normal distribution,  
363  $p_e(e)$ , with mean 0 and variance  $V_e$ . The joint distribution of  $g$  and  $e$ ,  $p_{g,e}(g, e, t)$ , is then initially  
multivariate normal with mean  $(\bar{g}_t, 0)$ , variances  $V_{g,t}$  and  $V_e$ , and no covariance.

Let the probability of survival for an individual with trait value  $z$  in generation  $t$  be

$$s(z, t) = \exp \left[ -\frac{(\theta_t - z)^2}{2\omega^2} \right], \quad (\text{A1})$$

366 where  $\theta_t$  is the optimum trait value in generation  $t$  and  $1/\omega^2$  is the strength of selection. The  
joint distribution of  $g$  and  $e$  following viability selection is

$$p'_{g,e}(g, e, t) = \frac{s(z, t)p_{g,e}(g, e, t)}{\bar{s}_t}, \quad (\text{A2})$$

where

$$\begin{aligned} \bar{s}_t &= \int_{-\infty}^{\infty} \int_{-\infty}^{\infty} s(z, t)p_{g,e}(g, e, t) dgde \\ &= \sqrt{\frac{\omega^2}{V_t}} \exp \left[ -\frac{(\theta_t - \bar{g})^2}{2V_t} \right], \end{aligned} \quad (\text{A3})$$

369 is the expected fraction of the population that survives in generation  $t$  (i.e., the population mean  
survival probability), with  $V_t = V_{g,t} + V_s$  and  $V_s = \omega^2 + V_e$  the inverse of the effective strength of  
selection. Integrating over environmental effects then gives the distribution of breeding values  
372 amongst the survivors

$$p'_g(g, t) = \int_{-\infty}^{\infty} p'_{g,e}(g, e, t) de, \quad (\text{A4})$$

which is normal with mean  $\bar{g}_t(1 - V_{g,t}/V_t) + \theta_t V_{g,t}/V_t$  and variance  $V_{g,t}(1 - V_{g,t}/V_t)$ . The mean  
breeding value is thus shifted towards  $\theta_t$  with a weight of  $V_{g,t}/V_t$  and the genetic variance has  
375 been reduced by this fraction.

We next assume that the breeding value is determined by a large number of small effect loci, such that the distribution of breeding values amongst siblings,  $p_{g,\text{sibs}}(g|g_{\text{mid}})$ , is normal with a mean equal to the midpoint of the parental breeding values,  $g_{\text{mid}}$ , and a variance,  $V_0$ , that does not depend on the parental genotypes or trait values (i.e., the infinitesimal model; [Barton et al., 2017](#); [Fisher, 1918](#)). The distribution of breeding values among the offspring is then

$$p_g(g, t+1) = \int_{-\infty}^{\infty} \int_{-\infty}^{\infty} p'_g(g_m, t) p'_g(g_p, t) p_{g,\text{sibs}}(g|(g_m + g_p)/2) dg_m dg_p, \quad (\text{A5})$$

which is normal with mean

$$\bar{g}_{t+1} = \bar{g}_t \left(1 - \frac{V_{g,t}}{V_t}\right) + \theta_t \frac{V_{g,t}}{V_t} \quad (\text{A6})$$

and variance

$$V_{g,t+1} = \frac{V_{g,t} V_s}{V_t} \frac{1}{2} + V_0. \quad (\text{A7})$$

That is, the mean breeding value remains constant through reproduction while the variance before reproduction is first halved (due to essentially "blending inheritance" between the parents) and then increased by segregation,  $V_0$ .

So we see that given the initial distribution of breeding values is normal, with Gaussian selection the breeding value distribution remains normal, allowing us to track the entire distribution of breeding values (and therefore phenotypes) across generations by keeping track of only its mean and variance. The variance dynamics are independent of the environment ( $\theta_t$ ) and the breeding values; solving Equation (A7) gives the genetic variance in generation any  $t$ . This expression is rather complicated (see Mathematica file), however it reaches an equilibrium

$$\widehat{V}_g = \frac{2V_0 - V_s + \sqrt{4V_0^2 + 12V_0V_s + V_s^2}}{4}. \quad (\text{A8})$$

Holding genetic variance constant at its equilibrium (which is reasonable given the variance is not expected to change with the environment or breeding values), in a constant environment,  $\theta_t = \theta$ , the mean breeding value in any generation  $t$  is found by solving Equation (A6),

$$\bar{g}_t = \theta - (\theta - \bar{g}_0) \left(1 - \frac{\widehat{V}_g}{V_s + \widehat{V}_g}\right)^t, \quad (\text{A9})$$

implying a geometric approach to  $\widehat{g} = \theta$  that becomes faster with  $\widehat{V}_g / (V_s + \widehat{V}_g)$ .

396 We assume each individual that survives viability selection produces  $\lambda$  offspring, and that  
if more than  $K$  offspring are produced then  $K$  of these are randomly chosen to start the next  
generation. If the population size in generation  $t$  was  $N_t$  then the population size in generation  
399  $t + 1$  is expected to be

$$N_{t+1} = \min(N_t \bar{s}_t \lambda, K). \quad (\text{A10})$$

## Appendix B Extinction Risk in Single and Two Generation Events

In this Appendix, we examine the effect of long-term extinction risk when populations are either  
402 not evolving or are perfectly tracking, with a one-generation lag behind the optimal trait value.  
Let  $s_o$  and  $s_m$  be the survivorship of individuals with the optimal trait or the maladaptive trait.  
The offspring probability generating functions for these individuals are  $f_o(x) = f(x, s_o)$  and  
405  $f_m(x) = f(x, s_m)$ , respectively, where  $f(x, s) = 1 - s + s \exp(\lambda(1 - x))$ . Let  $x_o^*$  and  $x_m^*$  be the  
asymptotic extinction probability for the lineage of a single individual if it always exhibits the  
optimal trait and if it always is maladapted, respectively. Namely,  $x_o^*$  and  $x_m^*$  are the smallest  
408 fixed points of  $f_o$  and  $f_m$ , respectively, on the interval  $0 \leq x \leq 1$ .

If a disturbance event lasts  $\tau \geq 1$  generations, then the eventual extinction probability of the  
lineage of a non-evolving individual equals

$$e_m := \lim_{T \rightarrow \infty} f_m^\tau(f_o^{T-\tau}(0)) = f_m^\tau(\lim_{T \rightarrow \infty} f_o^{T-\tau}(0)) = f_m^\tau(x_o^*).$$

411 While the eventual extinction probability of the lineage of an individual with a one-generation  
lagged tracking of the optimal trait equals

$$e_o := \lim_{T \rightarrow \infty} f_m(f_o^{\tau-1}(f_m(f_o^{T-\tau-1}(0))) = f_m(f_o^{\tau-1}(f_m(\lim_{T \rightarrow \infty} f_o^{T-\tau-1}(0))) = f_m(f_o^{\tau-1}(f_m(x_o^*))).$$

As  $s_o > s_m$ , we have  $f_o(x) < f_m(x)$  for all  $0 \leq x < 1$ , and  $x_o^* < x_m^*$ . Furthermore,  $f_i(x)$  are  
414 strictly increasing functions of  $x$ ,  $f_i(x) > x$  for  $x < x_i^*$ , and  $f_i(x) < x$  for  $x > x_i^*$  for  $i = o, m$ . Now  
suppose  $\tau = 1$ . Then  $e_m = f_m(x_o^*)$  and  $e_o = f_m(f_m(x_o^*))$ . As  $x_o^* < x_m^*$ ,  $f_m(x_o^*) > x_o^*$ . As  $f_m$  is an

increasing function, it follows that  $e_o = f_m(f_m(x_o^*)) > f_m(x_o^*) = e_m$ . Now suppose  $\tau = 2$ . Then  
 417  $e_m = f_m^2(x_o^*)$  and  $e_o = f_m(f_o(f_m(x_o^*)))$ . As  $f_m(x_o^*) > x_o^*$ ,  $f_o(f_m(x_o^*)) < f_m(x_o^*)$ . As  $f_m$  is increasing,  
 it follows that  $e_o = f_m(f_o(f_m(x_o^*))) < f_m(f_m(x_o^*)) = e_m$ .

## Appendix C Cumulative lag load

420 Here we show how to derive Equation (9). Our goal is to develop a formula for the cumulative  
 squared displacement,  $C(\tau, T) = \sum_{t=1}^T (\theta_t - \bar{g}_t)^2$ , given event length  $\tau$ . First note that Equation  
 (2) implies that, with constant genetic variance  $\widehat{V}_g$ , the mean trait displacement in the next gener-  
 423 ation is  $g_{t+1} - \theta_{t+1} = (g_t - \theta_t)(1 - v)$ , where  $v = \widehat{V}_g / \widehat{V}$  is a measure of evolvability. Thus, if the  
 optimum is fixed at some arbitrary value for  $\tau$  generations then the displacement in generation  $t$ ,  
 $d_t = g_t - \theta_t$ , is  $d_t = d_0(1 - v)^t$  and the cumulative squared displacement over those  $\tau$  generations  
 426 is  $d_0^2 \sum_{t=0}^{\tau-1} (1 - v)^{2t}$ . If the optimum then reverts to its original value for a further  $T - \tau > 0$  gener-  
 ations then the initial displacement is  $d_0(1 - v)^\tau - d_0$  and the cumulative squared displacement  
 over this period is  $d_0^2 [(1 - v)^\tau - 1]^2 \sum_{t=\tau}^{T-1} (1 - v)^{2(t-\tau)}$ . Combining these two sums we get

$$\begin{aligned} C(\tau, T) &\equiv \sum_{t=1}^T (\theta_t - \bar{g}_t)^2 \\ &= d_0^2 \sum_{t=0}^{\tau-1} (1 - v)^{2t} + d_0^2 [(1 - v)^\tau - 1]^2 \sum_{t=\tau}^{T-1} (1 - v)^{2(t-\tau)}. \end{aligned}$$

429 Multiplying by  $\frac{1-v}{2\widehat{V}_g}$ , evaluating the sums, and setting the initial displacement as  $d_0 = \Delta\theta$  gives  
 Equation (9) in the main text.

## Appendix D Adapting a Population Genetic Model

432 To gain a better understanding of why cumulative lag load depends on event length, we adapt  
 previous population genetic models of temporally variable selection. Consider a haploid case  
 with the ratio of the initial frequencies of two alleles being  $\frac{q_0}{1-q_0}$ . The ratio of the frequencies of  
 435 the alleles at time  $t + 1$ ,  $\frac{q_{t+1}}{1-q_{t+1}}$ , equals the product of the selection coefficients  $s_t$  from  $t = 0$  to



time  $t = T$ , multiplied by the ratio of the initial frequencies (Dempster, 1955; Felsenstein, 1976).

$$\frac{q_{t+1}}{1 - q_{t+1}} = \prod_{t=0}^T (1 + s_t) \frac{q_t}{1 - q_t} \quad (\text{B1})$$

Here, rather than allele frequency change, we consider the product of fitnesses over one extreme  
438 event of length  $\tau$ , where  $s$  is the selection coefficient corresponding to the change in the optimum during the event. When a population starting at the optimum perfectly tracks the extreme event, the product of fitnesses is

$$(1 - s)(1)^{\tau-1}(1 - s)(1)^{T-\tau} = (1 - s)^2 * 1^{T-2} \quad (\text{B2})$$

441 where fitness is reduced by  $s$  initially when the environment shifts to a new optimum, and then again when the environment returns to the original optimum. On the other hand, when a population starting at the optimum does not track the extreme event, the product of fitnesses is

$$(1 - s)(1 - s)^{\tau-1}(1)(1)^{T-\tau} = (1 - s)^\tau (1)^{T-\tau} \quad (\text{B3})$$

444 In the case of a two generation event, the product of fitnesses is  $(1 - s)^2$  regardless of whether it is a perfectly tracking population or a population that does not track the event. In events longer than two generations, perfectly tracking the environment is better.

447

## Literature Cited

- Alexander, H. K., G. Martin, O. Y. Martin, and S. Bonhoeffer. 2014. Evolutionary rescue: linking theory for conservation and medicine. *Evol Appl* 7:1161–79.
- 450 Barfield, M., and R. D. Holt. 2016. Evolutionary rescue in novel environments: towards improving predictability. *Evolutionary Ecology Research* 17:771–786.
- Barghi, N., J. Hermisson, and C. Schlötterer. 2020. Polygenic adaptation: a unifying framework  
453 to understand positive selection. *Nature Reviews Genetics* pages 1–13.
- Barton, N. H., A. M. Etheridge, and A. Véber. 2017. The infinitesimal model: Definition, derivation, and implications. *Theoretical population biology* 118:50–73.
- 456 Bell, G. 2017. Evolutionary rescue. *Annual Review of Ecology, Evolution, and Systematics* 48:605–627.
- Bell, G., and S. Collins. 2008. Adaptation, extinction and global change. *Evolutionary Applications* 1:3–16.  
459
- Benaïm, M., and S. Schreiber. 2019. Persistence and extinction for stochastic ecological models with internal and external variables. *Journal of Mathematical Biology* 79:393–431.
- 462 Bender, E. A., T. J. Case, and M. E. Gilpin. 1984. Perturbation experiments in community ecology: theory and practice. *Ecology* 65:1–13.
- Beverton, R. J. H., and S. J. Holt. 1957. On the dynamics of exploited fish populations. *Marine Fisheries, Great Britain Ministry of Agriculture, Fisheries and Food* 19.
- 465 Bumpus, H. C. 1899. *The Elimination of the Unfit as Illustrated by the Introduced Sparrow, Passer Domesticus:(a Fourth Contribution to the Study of Variation)*. Gin.
- 468 Bürger, R., and M. Lynch. 1995. Evolution and extinction in a changing environment: a quantitative-genetic analysis. *Evolution* 49:151–163.

- Campbell-Staton, S. C., Z. A. Cheviron, N. Rochette, J. Catchen, J. B. Losos, and S. V. Edwards.  
471 2017. Winter storms drive rapid phenotypic, regulatory, and genomic shifts in the green anole  
lizard. *Science* 357:495–498.
- Charlesworth, B. 1993. Directional selection and the evolution of sex and recombination. *Genetics*  
474 *Research* 61:205–224.
- Chevin, L. M. 2013. Genetic constraints on adaptation to a changing environment. *Evolution*  
67:708–21.
- 477 Coleman, M. A., A. J. Minne, S. Vrancken, and T. Wernberg. 2020. Genetic tropicalisation following  
a marine heatwave. *Scientific reports* 10:1–11.
- Cuddington, K., and P. Yodzis. 1999. Black noise and population persistence. *Proceedings of the*  
480 *Royal Society of London. Series B: Biological Sciences* 266:969–973.
- Dempster, E. R. 1955. Maintenance of genetic heterogeneity. *Cold Spring Harbor Symp* 20:25–32.
- Donihue, C. M., A. Herrel, A. C. Fabre, A. Kamath, A. J. Geneva, T. W. Schoener, J. J. Kolbe, and  
483 J. B. Losos. 2018. Hurricane-induced selection on the morphology of an island lizard. *Nature*  
560:88–91.
- Felsenstein, J. 1976. The theoretical population genetics of variable selection and migration.  
486 *Annual review of genetics* 10:253–280.
- Fisher, R. A. 1918. The correlation between relatives on the supposition of mendelian inheritance.  
*Transactions of the Royal Society of Edinburgh* 52:399–433.
- 489 Fox, G. A., and J. Gurevitch. 2000. Population numbers count: tools for near-term demographic  
analysis. *The American Naturalist* 156:242–256.
- Franks, S. J., S. Sim, and A. E. Weis. 2007. Rapid evolution of flowering time by an annual plant  
492 in response to a climate fluctuation. *Proc Natl Acad Sci* 104:1278–82.

- Gomulkiewicz, R., and R. D. Holt. 1995. When does evolution by natural selection prevent extinction? *Evolution* 49:201–207.
- 495 Grant, P. R., and B. R. Grant. 2002. Unpredictable evolution in a 30-year study of darwin’s finches. *Science* 296:707–711.
- . 2014. 40 years of evolution: Darwin’s finches on Daphne Major Island. Princeton University Press.
- 498
- Grant, P. R., B. R. Grant, R. B. Huey, M. T. J. Johnson, A. H. Knoll, and J. Schmitt. 2017. Evolution caused by extreme events. *Philos Trans R Soc Lond B Biol Sci* 372.
- 501 Harris, T. E. 1964. The theory of branching process. Rand Corporation.
- Hastings, A., K. C. Abbott, K. Cuddington, T. Francis, G. Gellner, Y.-C. Lai, A. Morozov, S. Petrovskii, K. Scranton, and M. L. Zeeman. 2018. Transient phenomena in ecology. *Science* 361.
- 504 Holling, C. S. 1973. Resilience and stability of ecological systems. *Annual review of ecology and systematics* 4:1–23.
- . 1996. Engineering resilience versus ecological resilience. *Engineering within ecological*
- 507 *constraints* 31:32.
- Holt, R. D. 2008. Theoretical perspectives on resource pulses. *Ecology* 89:671–681.
- Ives, A. R., and S. R. Carpenter. 2007. Stability and diversity of ecosystems. *Science* 317:58–62.
- 510 Kéfi, S., V. Domínguez-García, I. Donohue, C. Fontaine, E. Thébault, and V. Dakos. 2019. Advancing our understanding of ecological stability. *Ecology Letters* 22:1349–1356.
- Kendall, B. E., and G. A. Fox. 2002. Variation among individuals and reduced demographic
- 513 *stochasticity*. *Conservation Biology* 16:109–116.
- Kopp, M., and S. Matuszewski. 2014. Rapid evolution of quantitative traits: theoretical perspectives. *Evol Appl* 7:169–91.

- 516 Lande, R. 1993. Risks of population extinction from demographic and environmental stochasticity  
and random catastrophes. *The American Naturalist* 142:911–927.
- . 2015. Evolution of phenotypic plasticity in colonizing species. *Molecular ecology* 24:2038–  
519 2045.
- Lande, R., and G. Barrowclough. 1987. Effective population size, genetic variation, and their use  
in population. *Viable populations for conservation* pages 87–123.
- 522 Lande, R., and S. Shannon. 1996. The role of genetic variation in adaptation and population  
persistence in a changing environment. *Evolution* 50:434–437.
- Lloyd-Smith, J. O., S. J. Schreiber, P. E. Kopp, and W. M. Getz. 2005. Superspreading and the  
525 effect of individual variation on disease emergence. *Nature* 438:355–359.
- Mangel, M., and C. Tier. 1994. Four facts every conservation biologists should know about  
persistence. *Ecology* 75:607–614.
- 528 Ozgul, A., T. Coulson, A. Reynolds, T. C. Cameron, and T. G. Benton. 2012. Population responses  
to perturbations: the importance of trait-based analysis illustrated through a microcosm ex-  
periment. *The American Naturalist* 179:582–594.
- 531 Petchey, O., A. Gonzalez, and H. Wilson. 1997. Effects on population persistence: the interaction  
between environmental noise colour, intraspecific competition and space. *Proceedings of the  
Royal Society of London. Series B: Biological Sciences* 264:1841–1847.
- 534 Ricker, W. E. 1954. Stock and recruitment. *Journal of the Fisheries Board of Canada* 11:559–623.
- Turelli, M. 1984. Heritable genetic variation via mutation-selection balance: Lerch’s zeta meets  
the abdominal bristle. *Theoretical population biology* 25:138–193.
- 537 ———. 2017. Commentary: Fisher’s infinitesimal model: a story for the ages. *Theoretical popu-  
lation biology* 118:46–49.

Ummerhofer, C. C., and G. A. Meehl. 2017. Extreme weather and climate events with ecological  
540 relevance: a review. *Philosophical Transactions of the Royal Society B: Biological Sciences*  
372:20160135.

Wilson, E. O., and R. H. MacArthur. 1967. *The theory of island biogeography*. Princeton Univer-  
543 sity Press.

Yodzis, P. 1988. The indeterminacy of ecological interactions as perceived through perturbation  
experiments. *Ecology* 69:508–515.

546

## Supplementary Figures

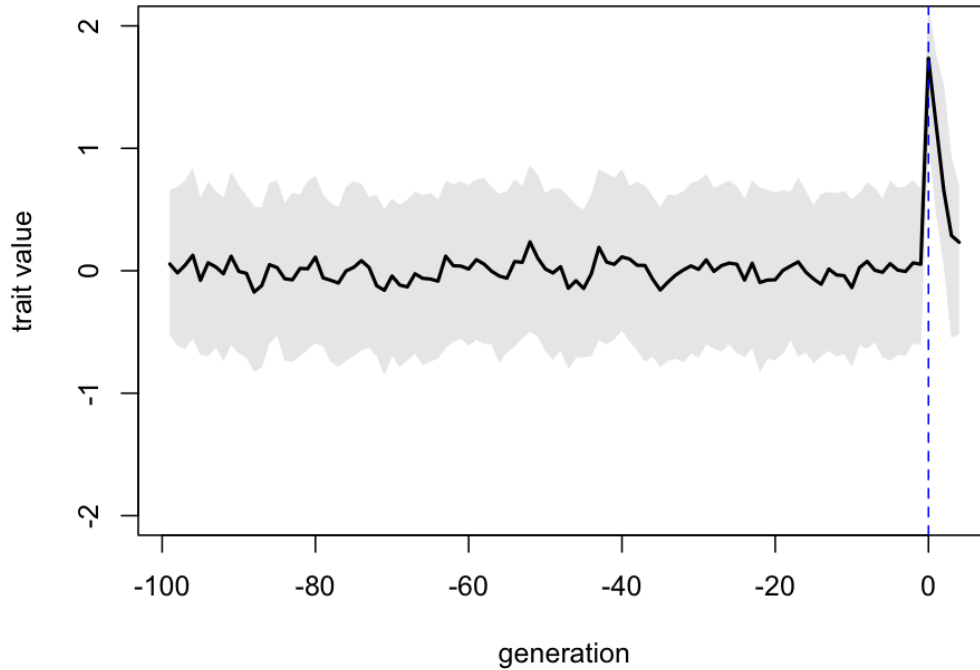


Figure S1: Rapid expansion and stabilization of phenotypic variance during the 100 generation burn-in with  $V_e = 0$ ,  $V_0 = 1$ . Black line is mean trait value and gray shaded region extends from minimum to maximum trait values. The dashed blue curve indicates a one generation extreme event.

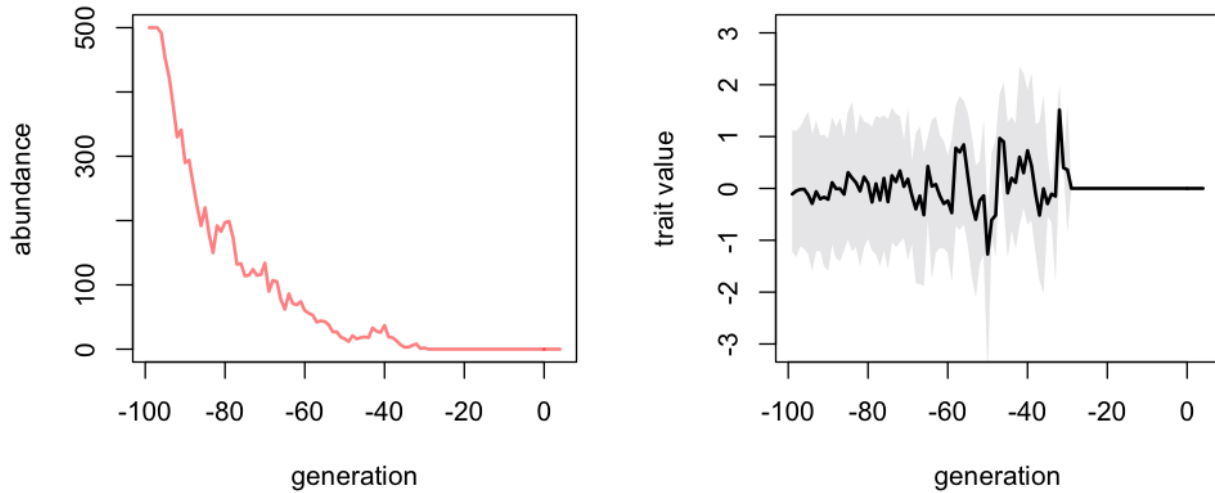


Figure S2: Extinction in a population with high variance load with  $V_0 = 3$ ,  $V_e = 0$ . Black line is mean trait value, grey shaded region extends from minimum to maximum trait values.

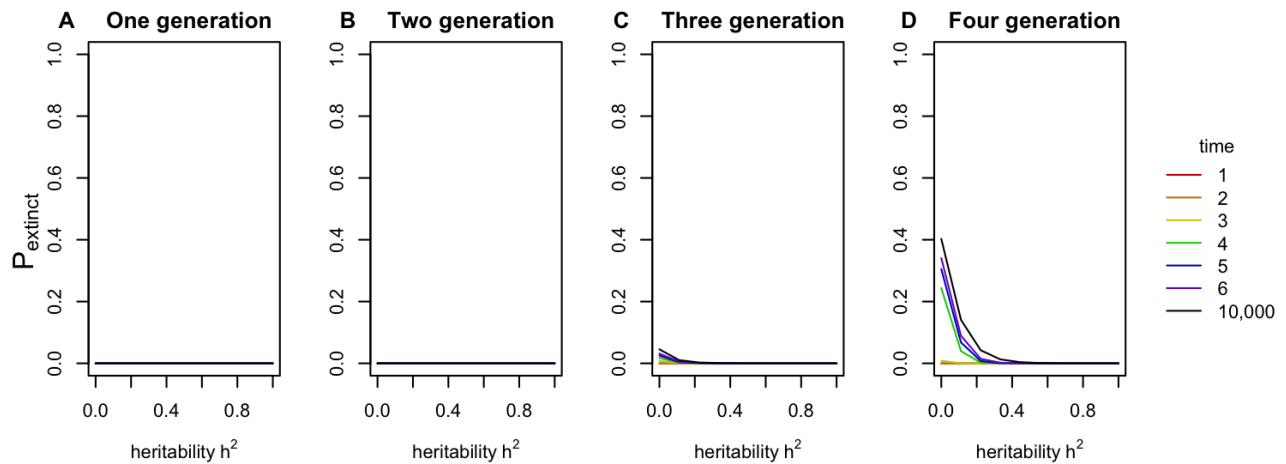


Figure S3: Extinction risk through time across a range of heritability for extreme events lasting 1, 2, 3, or 4 generations. Time starts the generation the event began.  $\hat{V}_p = 1$ ,  $\omega = 1$ ,  $\Delta\theta = 2.5$ .



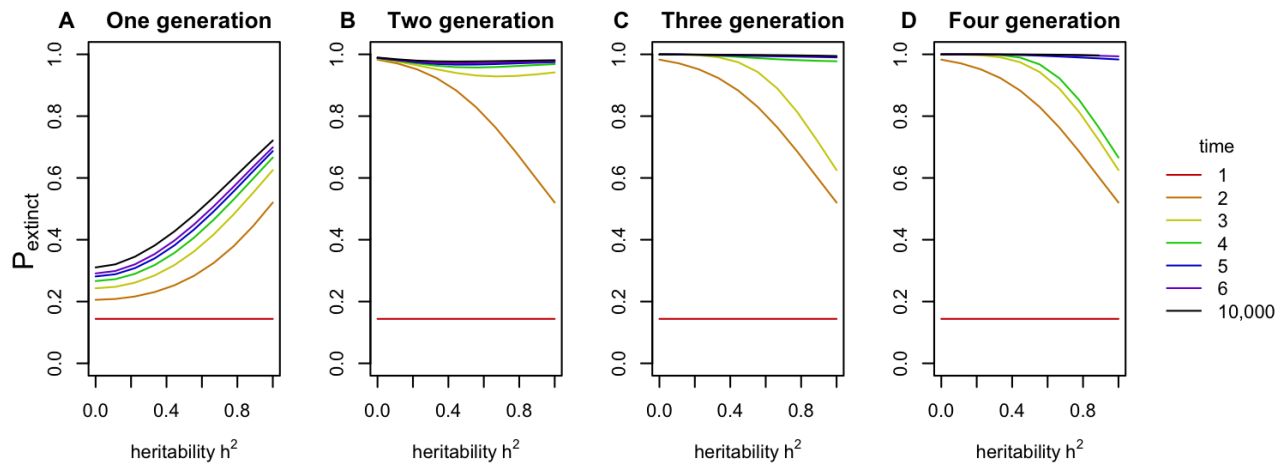


Figure S4: Extinction risk through time across a range of heritability for extreme events lasting 1, 2, 3, or 4 generations. Time starts the generation the event began.  $\widehat{V}_p = 1$ ,  $V_s = 1$ ,  $\Delta\theta = 4.5$ .