1	Comparison of the <i>in vitro</i> activity of novel and established nitrification inhibitors
2	applied in agriculture: challenging the effectiveness of the currently available
3	compounds
4	
5	Evangelia S. Papadopoulou ^{1*} , Eleftheria Bachtsevani ¹ , Eleni Lampronikou ¹ , Eleni
6	Adamou ¹ , Afroditi Katsaouni ¹ , Cécile Thion ² , Sotirios Vasileiadis ¹ , Urania Menkissoglu-
7	Spiroudi ³ , Graeme W. Nicol ² , Dimitrios G. Karpouzas ¹ .
8	
9	¹ University of Thessaly, Department of Biochemistry and Biotechnology, Laboratory of
10	Plant and Environmental Biotechnology, Viopolis, 41500 Larissa, Greece
11	² Laboratoire Ampère, École Centrale de Lyon, University of Lyon, 69134 Ecully, France
12	³ Aristotle University of Thessaloniki, School of Agriculture, Forestry and Environment,
13	Faculty of Agriculture, Pesticide Science Laboratory, 54124 Thessaloniki, Greece
14	
15	* Corresponding author
16	Dr. Evangelia S. Papadopoulou
17	Tel. +30-2410-565232, Fax. +30-2410-565290, Email. evapapadopoulou@bio.uth.gr
18	Abbreviations ¹

¹NIs: nitrification inhibitors, EQ: ethoxyquin; QI: 2,6-dihydro-2,2,4-trimethyl-6-quinone imine; EQNL: 2,4-dimethyl-6-ethoxyquinoline; DCD: dicyandiamide, NP: nitrapyrin,

ABSTRACT Nitrification inhibitors (NIs) applied to soil reduce nitrogen fertilizer losses 19 20 from agricultural ecosystems. Currently available NIs appear to selectively inhibit 21 ammonia-oxidizing bacteria (AOB), while their impact on other groups of nitrifiers is 22 limited. Ethoxyquin (EQ), a preservative shown to inhibit ammonia-oxidizers (AO) in soil, 23 is rapidly transformed to 2,6-dihydro-2,2,4-trimethyl-6-quinone imine (QI) and 2,4-24 dimethyl-6-ethoxy-quinoline (EQNL). We compared the inhibitory potential of EQ and its 25 derivatives in vitro with other established NIs that have been applied in an agricultural setting (dicyandiamide (DCD), nitrapyrin (NP), 3,4-dimethylpyrazole phosphate (DMPP)) 26 27 by evaluating their impact on the activity and growth of five soil-derived strains (two AOB (Nitrosomonas europaea, Nitrosospira multiformis), two ammonia-oxidizing archaea 28 (AOA) ("Candidatus Nitrosocosmicus franklandus", "Candidatus Nitrosotalea sinensis"), 29 and one nitrite-oxidizing bacterium (NOB) (Nitrobacter sp.)). NIs degradation was also 30 31 determined. AOA were more sensitive than AOB or NOB to EQ and its derivatives. Despite 32 its transient character, QI was primarily responsible for AO inhibition by EQ, and the most 33 potent NI against AOA. For AOB, QI was more potent than DCD but less than nitrapyrin 34 and DMPP. AOA and NOB showed higher tolerance to the persistent compounds DCD and DMPP. Our findings benchmark the activity range of known and novel NIs with 35 practical implications for their use, and the development of novel NIs with broad or 36 37 complementary activity against all AO.

DMPP: 3,4-dimethylpyrazole phosphate, AOB: ammonia-oxidizing bacteria, AOA: ammonia-oxidizing archaea, AO: ammonia-oxidizers, NOB: nitrite-oxidizing bacteria, comammox: complete ammonia-oxidizing bacteria; AMO: ammonia monooxygenase

38

KEYWORDS nitrification inhibitors, ethoxyquin, quinone imine, ammonia-oxidizing
bacteria, ammonia-oxidizing archaea, nitrite-oxidizing bacteria, *in vitro* assays

41

42 Modern agricultural systems depend heavily on large inputs of synthetic N fertilizers to maintain crop productivity and alleviate food crisis for the growing global population (1). 43 However, approximately 70% of the annual global input of 100 Tg N fertilizer is lost from 44 45 agricultural ecosystems due to nitrification and subsequent denitrification processes leading to groundwater and atmospheric pollution through nitrate leaching and nitrogen 46 oxides (NxO) emissions, respectively (2). To reduce N losses and improve nitrogen use 47 efficiency, nitrification inhibitors (NIs) are routinely incorporated into N-stabilized 48 49 fertilizers to reduce the activities of nitrifying prokaryotes and increase N retention time in 50 soil (3, 4).

Hundreds of compounds have been identified as potential NIs (5), but only three of them have gained importance for practical use on a global scale: 2-chloro-6-(trichloromethyl) pyridine (nitrapyrin) (NP) (6), dicyandiamide (DCD) (7), and 3,4dimethylpyrazole phosphate (DMPP) (8). All three are known to act on ammonia monooxygenase (AMO), a key enzyme in the first and rate-limiting step of nitrification (9). In particular, NP is believed to act either as copper chelator or "suicide" inhibitor (10, 11), while the mode of action of DCD and DMPP is not yet fully established.

58 When NIs were first developed for widespread use, nitrification was considered a 59 two-step process carried out by ammonia- (AOB) and nitrite-oxidizing bacteria (NOB).

AOB oxidize ammonia to hydroxylamine (NH₂OH) via AMO, which is further oxidized through to nitrite (NO₂⁻). NOB subsequently transform NO₂⁻ to nitrate (NO₃⁻) via nitrite oxidoreductase (NXR) (12, 13). However, over the last 15 years, other groups contributing to nitrification have been discovered including ammonia-oxidizing archaea (AOA) (14, 15), and 'comammox' *Nitrospira* that are able to perform complete oxidation of ammonia to nitrate within an individual cell (16, 17).

These breakthroughs in the microbiology and biochemistry of nitrification were not 66 accompanied by complementary advances on NI research with respect to their spectrum of 67 activity. Most studies have focused on AOB (18-21), and only recently the activity of NIs 68 69 on AOA was explored (22, 23), while their activity on other groups of nitrifiers (including NOB and comammox bacteria) is not known. Current knowledge of the activity of NIs has 70 been derived from soil microcosm studies where AOB appear to be functionally dominant 71 72 (24 - 29). The only available in vitro study assessing the comparative activity of NIs on soil-73 derived AOB and AOA isolates, revealed selective inhibitory activity of DCD and NP 74 against AOB and AOA, respectively (23).

The variation in sensitivity toward different types of NIs, combined with the contribution of AOA, NOB and comammox bacteria to nitrification in distinct ecological niches (30, 31) suggests a suboptimal efficiency of the currently available NIs, and stresses the need for the discovery of novel NIs with a broader range of activity against all microorganisms contributing to nitrification. The use of *in vitro* inhibition assays with a diverse range of soil-derived strains is therefore a necessary benchmarking step to define the exact spectrum of activity of novel and known NIs.

82 In previous soil microcosm studies we showed that ethoxyquin (EO) (1,2-dihydro-6-ethoxy-2,2,4-trimethylquinoline), an antioxidant used as preservative in fruit-packaging 83 plants, and its derivative 2,6-dihydro-2,2,4-trimethyl-6-quinone imine (QI), strongly 84 85 inhibited the activity of AOB and AOA (32). EQ in soil is rapidly transformed to QI and 86 2,4-dimethyl-6-ethoxyquinoline (EQNL), with the former being the major metabolite (33). 87 The capacity of EQ to be rapidly transformed in soil to possibly potent NIs has particular interest at the application level, considering that the spectrum and the duration of inhibition 88 are desirable attributes of NIs in practice. 89

90 We aimed to determine, at the *in vitro* level, the inhibitory potency of EQ and its derivatives on representative soil strains of AOB and AOA, in comparison with three 91 92 widely used NIs (NP, DCD, and DMPP), whose full spectrum of activity against different ammonia-oxidizers (AO) has not yet been fully established. We expanded our in vitro 93 94 inhibition assays to NOB, whose response to NIs is largely unknown, to gain insights on 95 the impact of NIs on microbial groups functionally associated with ammonia oxidation. 96 Specifically, we determined the inhibitory effects on the growth and activity of two AOB 97 strains (Nitrosomonas europaea ATCC25978, Nitrosospira multiformis ATCC 25196), 98 two AOA strains ("Candidatus Nitrosocosmicus franklandus" C13 and "Candidatus 99 Nitrosotalea sinensis" ND2) and one NOB strain (*Nitrobacter* sp. NHB1) using q-PCR to 100 measure the abundance of amoA or nxrB functional genes, and the production or 101 consumption of nitrite, for AO or NOB, respectively. To establish possible correlations 102 between NI presence and inhibition, their degradation and transformation (for EQ) were also determined. 103

104

105 **RESULTS**

106 The impact of NIs on the activity and growth of AOB isolates. The effects of six NIs (EQ, QI, EQNL, DCD, NP and DMPP) on the activity and growth of the two AOB 107 108 isolates were initially tested over a range of concentration levels. EQ strongly inhibited 109 ammonia oxidation in liquid cultures of *N. europaea* (Fig. 1a) and *N. multiformis* (Fig. 2a) only at the concentration of 460 μ M. QI completely inhibited ammonia oxidation by N. 110 111 europaea (Fig. 1b) and N. multiformis (Fig. 2b) at concentrations $> 270 \mu$ M and $> 135 \mu$ M. 112 respectively. In contrast, EQNL only temporarily inhibited N. europaea activity at the highest concentration tested, 500 μ M (Fig. 1c), while at the same concentration level N. 113 multiformis activity was fully inhibited (Fig. 2c). 114

115 For the established NIs, DCD inhibited the activity of both AOB strains at 116 concentrations of 250 µM and 500 µM (Fig. 1d), with a mere recovery of N. multiformis at 250 µM (Fig 2d). NP completely inhibited the activity of both N. europaea and N. 117 118 *multiformis* at concentrations of \geq 5 µM (Fig. 1e and 2e). Similarly, DMPP reduced nitrite production by both AOB isolates at concentrations $\geq 1 \mu M$, with a late recovery observed 119 120 only for N. europaea at 1 μ M (Fig. 1f and 2f), and complete inhibition observed at concentrations $\geq 10 \,\mu$ M. The inhibition of AOB growth as determined by measurements of 121 122 *amoA* gene abundance were in agreement with the NO_2^- production rates (Fig. S1 and S2) 123 with the exception of a low level inhibition $(22.9 \pm 3.9 \%)$ of *N. multiformis* growth at 0.1 µM DMPP compared to the control, despite no difference in nitrite production rates (Fig. 124 125 S2f).

Based on the inhibition assays, EC₅₀ values for each strain x compound combination
were calculated. The two AOB showed equivalent EC₅₀ values for the different NIs tested

128 with the sole exception of EO derivatives. Higher EC_{50} values were observed for N. europaea (199.8±16.0 and 543.4±111.5 µM for QI and EQNL, respectively) compared to 129 N. multiformis (65.1±6.0 and 360.5±105.7 µM) (Table 1). DMPP and NP were the most 130 131 potent inhibitors of N. europaea (EC₅₀ = 2.1 ± 0.7 and 2.1 ± 0.4 µM, respectively), followed by EQ, QI, and DCD which did not differ in their ability to inhibit N. europaea, with EQNL 132 133 being the weakest inhibitor (EC₅₀ = $181.4 \pm 23.3 \mu$ M) (Table 1). For *N. multiformis*, DMPP, NP, and OI were equally effective and the most potent inhibitors, with EC_{50} values of 134 $0.6\pm0.1, 0.8\pm0.3, \text{ and } 65.1\pm0.6 \,\mu\text{M}$, respectively, followed by EQ (EC₅₀ = 214.8±39.6 μM), 135 136 DCD (EC₅₀ = 248.7 \pm 7.4 μ M) and EQNL (EC₅₀ = 360.5 \pm 105.7 μ M) (Table 1).

137 The impact of NIs on the activity and growth of AOA isolates. We further tested the inhibitory effects of NIs on the activity and growth of two soil-derived AOA isolates. 138 The activity of "Ca. N. franklandus" (Fig. 3a) and "Ca. N. sinensis" (Fig. 4a) was 139 140 significantly reduced by EQ at concentrations $\geq 4.6 \ \mu M$ and $\geq 0.46 \ \mu M$, respectively. However, complete recovery of "Ca. N. sinensis" activity was observed at 0.46 µM (Fig. 141 142 4a). QI significantly reduced the activity of AOA at all concentrations, with a gradual recovery observed for only "Ca. N. franklandus" at the lowest concentration level (0.27 143 144 uM) (Fig. 3b and 4b). EONL suppressed ammonia oxidation by "Ca. N. franklandus" (Fig. 145 3c) and "*Ca*. N. sinensis" (Fig. 4c) at concentrations $\geq 125 \,\mu\text{M}$ and $\geq 25 \,\mu\text{M}$, respectively, although recovery was observed. The impact of EQNL on the growth of "Ca. N. 146 147 franklandus" was not fully consistent with the activity measurements, and no significant 148 differences among the different concentrations were observed at the end of the incubation period (Fig. S3). 149

150	Regarding the other NIs, DCD halted the activity of "Ca. N. franklandus" (Fig. 3d)
151	and " <i>Ca.</i> N. sinensis" (Fig. 4d) at concentrations $\geq 1000 \ \mu M$ and $\geq 500 \ \mu M$ respectively,
152	with a complete recovery observed only for "Ca. N. franklandus" at 1000 μ M. However,
153	based on growth measurements, DCD significantly reduced "Ca. N. franklandus" growth
154	even at 500 μ M (Fig. S3). NP significantly reduced the activity of "Ca. N. franklandus"
155	(Fig. 3e) and " <i>Ca</i> . N. sinensis" (Fig. 4e) at concentrations $\ge 1 \mu M$ and $\ge 5 \mu M$, respectively,
156	although recovery was observed at these concentrations at the end of the incubation.
157	Finally, DMPP inhibited the activity of both AOA isolates at concentrations \ge 500 μ M (Fig.
158	3f and 4f). However, complete recovery was observed for "Ca. N. franklandus" at 500,
159	1000, and 2500 μ M (Fig. 3f), and for "Ca. N. sinensis" only at 500 μ M (Fig. 4f). In certain
160	cases, the impact of DMPP on nitrite production was not concomitant with growth patterns.
161	Hence, DMPP concentrations \geq 500 μ M induced a persistent reduction in <i>amoA</i> gene
162	abundance of "Ca. N. franklandus" (Fig. S3), while "Ca. N. sinensis" showed complete
163	recovery of its growth at all tested concentrations at the end of the incubation period (Fig.
164	S4).

When EC_{50} values were calculated "*Ca*. N. franklandus" differed in its sensitivity 165 to DCD and DMPP compared to "Ca. N. sinensis", with the former being more tolerant to 166 both DCD (EC₅₀ of 1568.5±237.1 µM compared to 477.8±56.6 µM for "Ca. N. sinensis"), 167 and DMPP (EC₅₀ of 1773.7±359.5 µM compared to 359.5.5±43.1 µM for "Ca. N. 168 sinensis") (Table 1). DCD and DMPP were also the weakest AOA inhibitors from those 169 170 tested (Table 1). In contrast, EQ and its oxidative derivatives as well as NP were equally 171 effective inhibitors of both AOA isolates, with QI scoring the lowest EC_{50} values (0.7±0.4 and 0.3±0.0 µM for "Ca. N. franklandus" and "Ca. N. sinensis", respectively). 172

173	The impact of NIs on the activity and growth of Nitrobacter sp. NHB1. Nitrite
174	oxidation by Nitrobacter sp. NHB1 was completely inhibited by EQ, EQNL, and DMPP
175	only at the highest tested concentrations of 460 $\mu M,$ 500 $\mu M,$ and 25000 $\mu M,$ respectively,
176	and DCD was not inhibitory even at the highest concentration tested (100000 μ M) (Fig. 5).
177	In contrast, QI and NP significantly suppressed Nitrobacter sp. NHB1 activity at
178	concentrations \geq 135 μM and \geq 100 $\mu M,$ respectively, though recovery was observed for QI
179	at 135 μM and 270 μM (Fig. 5). We observed variations in the growth inhibition patterns
180	of the tested NIs (assessed as a lack of increase in nxrB gene abundance). EQ induced a
181	significant reduction in the growth of Nitrobacter sp. NHB1 at all concentration levels at
182	the end of the incubation period. QI strongly inhibited growth at concentrations $\geq 135 \ \mu M$,
183	while EQNL did not affect bacterial growth even at the highest concentration level (500
184	$\mu M)$ (Fig. S5). On the contrary, DCD at 100000 $\mu M,$ and DMPP at concentrations ≥ 25000
185	μ M significantly suppressed <i>Nitrobacter</i> sp. NHB1 growth (Fig. S5). When EC ₅₀ values
186	were calculated EQ, QI, EQNL and NP were equally suppressive towards Nitrobacter sp.
187	NHB1, while DMPP and DCD showed no appreciable inhibition (Table 1).

Degradation of NIs in the in vitro assays. We further determined the degradation 188 of all tested NIs during the *in vitro* assays to establish whether there is a relationship 189 between the length of exposure and inhibition. The degradation of NIs in most cases was 190 best described by the single first order (SFO) kinetic model ($x^2 \le 15$, $r^2 \ge 0.75$), with the 191 exception of NP which in certain treatments followed a biphasic degradation pattern (Table 192 193 S1). EQ was rapidly transformed to QI and EQNL. QI was detected at concentrations below its EC₅₀ values for AOB/NOB but above its EC₅₀ values for AOA at the onset of EQ-194 induced inhibition, (Tables 1 and 2). However, the maximum concentrations of QI formed 195

196 in these cultures were equivalent or higher than the EC_{50s} for N. europaea and N. multiformis (AOB), respectively, while still lower than that for Nitrobacter sp. NHB1 197 (NOB). In contrast, EQNL was always formed at much lower concentrations than the 198 199 estimated EC_{50s} for AO and NOB with the exception of its maximum recorded concentration in the liquid culture of "Ca. N. sinensis" amended with 460 µM EQ (Table 200 201 1 and 2, Fig. S6). The degradation half-life (DT_{50}) for the sum of EQ+QI+EQNL ranged from 2.1 to 60.1 days in the cultures of Nitrobacter sp. NHB1 and N. multiformis 202 203 respectively when amended with 460 μ M of EQ (Table S1). QI showed limited persistence 204 and a weak dose-dependent degradation pattern with $DT_{50} = 0.05 - 1.52$ days at the lowest concentration level (2.7 μ M) and 2.23 - 5.65 days at the highest concentration level (540 205 µM). In contrast, EQNL persisted in the liquid cultures throughout the experiment 206 (extrapolated DT_{50} values >1000 days) (Table S1). 207

All other NIs did not show a clear dose-dependent degradation pattern. DMPP and DCD were rather persistent with their DT_{50} values varying from 4.34 to >1000 days and 45.9 days to >1000 days respectively. In contrast NP was rapidly degraded in most treatments with DT_{50} values ranging from 0.12 to 12.5 days (Table S1).

212

213 **DISCUSSION**

We determined *in vitro* the inhibition potency of EQ and its oxidation derivatives, QI and EQNL, on the growth and activity of a range of soil nitrifiers and compared them to the NIs most widely used in agricultural practice (DCD, NP, and DMPP). Five isolates, representative of diverse and globally distributed lineages of soil AO (34, 35) and NOB (36), were tested. *N. europaea* and *N. multiformis* are representatives of AOB clusters 3

219 and 7, respectively (37), and are commonly used as model soil AOB, with cluster 3 sequences being often the dominant lineage in agricultural and grassland soil ecosystems 220 221 (38, 39, 34). The two soil AOA isolates represent contrasting ecological niches, with "Ca. 222 N. franklandus (40) and "Ca. N. sinensis" (41, 42) being representatives of widely 223 distributed neutrophilic and acidophilic AOA lineages, respectively. *Nitrobacter* sp. NHB1 224 (43) was chosen as a representative of one of the two dominant NOB lineages found in soil (36), with *Nitrobacter* strains typically having greater nitrite oxidation activity compared 225 to those from the genus Nitrospira (44, 45). 226

227 We first determined the inhibitory potential of EQ to AO and NOB. As a powerful 228 antioxidant, EQ is prone to oxidation, producing a range of transformation products 229 depending on the interacting matrix (e.g. animals, plants, and soils) (46). In this study, EO was rapidly transformed to OI and EONL, with the former being the major but least 230 231 persistent transformation product, while the latter being a minor but more persistent 232 product, in line with previous studies in soil (32, 33). We observed a different inhibition 233 potential for AOB and AOA. All three compounds were more potent inhibitors of the two 234 AOA isolates compared to AOB isolates, with *N. multiformis* being more sensitive than *N*. 235 *europaea*. EQ was characterized by EC_{50} values lower than those of EQNL, but equal or 236 higher (only in case of *N. multiformis*) than those of QI. Given the transformation pattern of EQ in the microbial cultures, we presume that its calculated EC_{50} values also include the 237 activity of its two oxidative derivatives, QI and EQNL. Considering that in the cultures 238 239 inhibited by EQ (i) EQNL was formed at concentrations substantially lower than those expected to result in an inhibitory effect on the AO tested, and (ii) QI was formed at 240 241 concentrations equal or higher than those expected to induce an inhibitory effect on the AO

242 tested, we suggest that OI is the main determinant of the persistent inhibitory effect of EO on AO and NOB. This is consistent with our previous soil studies where EQ, but mostly 243 QI when applied alone, induced a significant inhibition of potential nitrification and 244 245 transcription of both bacterial and archaeal *amoA* genes (32). Unlike our previous soil studies where QI showed equivalent inhibitory effects against AOB and AOA, we observed 246 247 a difference in the inhibition potential of QI between these two AO groups. This discrepancy could be attributed to the concentrations of OI applied or formed in soil 248 samples (up to 86.1 umol Kg⁻¹ dwt soil) which most probably reached the inhibition 249 250 threshold levels for both AO groups. Further studies under a range of conditions known to affect the activity of AO in soil (e.g. soil pH, NH_4^+ amendment) or the performance of NIs 251 (e.g. rate of application, temperature, N source) will determine the potency of these 252 compounds as broad range or AOA-specific NIs. 253

254 To establish the full potential of EQ or its derivatives as new potent NIs, we compared them with the *in vitro* inhibitory activity of three established NIs. In contrast to 255 256 EQ and its derivatives, DCD exhibited higher in vitro toxicity towards AOB isolates 257 compared to AOA, in line with most previous soil studies (47-49). In our study, DCD 258 inhibited both AOA isolates at concentrations \geq 500 µM. Similar *in vitro* tests with soil 259 enrichment cultures of AOA Nitrososphaera sp. JG1 and Nitrosarchaeum koreense MY1, showed strong inhibition by DCD at 500 μ M (50, 51), while a pure isolate of "Ca. 260 Nitrosotalea devanatera" ND1 was inhibited at concentrations >1000 μ M (22). 261 262 Furthermore, Shen et al., (23) reported an EC₅₀ of 940.6 (±85.3) µM for Nitrososphaera viennensis EN76. It is interesting to note that "Ca. N. sinensis" and "Ca. N. franklandus" 263 264 have different sensitivities to DCD compared to their phylogenetically associated strains

265 "Ca. Nitrosotalea devanatera" ND1 and Nitrososphaera viennensis EN76, respectively. Despite being closely related, the two Nitrosotalea isolates exhibit different physiologies 266 which might explain the observed differences. Although "*Ca*. N. sinensis" (μ max = 0.025 267 h^{-1}) grows approximately twice as fast as "*Ca*. N. devanaterra (ND1)" (µmax = 0.011 h^{-1}), 268 their cell yields are similar (4-4.5 cells μ M⁻¹NH₃), while the specific cell activity of ND1 269 $(0.072 \text{ fmol NO}_2^{-}\text{cell}^{-1} \text{ h}^{-1})$ is slightly higher than this of "*Ca*. N. sinensis" (0.065 fmol 270 NO_2 cell⁻¹ h⁻¹) (40). On the other hand, the two *Nitrososphaera* isolates are characterized 271 by similar growth rates (μ max = 0.024 h⁻¹). In contrast to our results, a 3x lower EC₅₀ was 272 273 reported by Shen *et al.* (23) for *N. multiformis* (EC₅₀= 80.28 (\pm 6.20) µM vs. 248.7 (\pm 7.4) uM in our study). However, the persistence of DCD in the cultures was not determined. 274

275 Similar to DCD, the more recently discovered DMPP showed higher inhibition potency to AOB compared to AOA. For the AOA, "Ca. N. franklandus" was less sensitive 276 277 to DMPP than "Ca. N. sinensis", a difference potentially attributed to the higher specific cell activity of "Ca. N. franklandus" (2.02 vs. 0.065 fmol NO₂⁻cell⁻¹ h⁻¹ for "Ca. N. 278 279 sinensis") (40). A differential activity of DMPP towards AOA and AOB has been observed previously in soil studies (28, 52-54), although the bioactivity of DMPP in soil, unlike in 280 281 vitro tests, is influenced by various edaphic, environmental and microbial factors (55). Our 282 study provides the first data on the *in vitro* range activity of DMPP against soil AO.

NP was the only tested NI that showed an equivalent and strong inhibitory effect towards both AOB and AOA isolates, suppressing their activity at concentrations $\ge 1 \mu M$. In accordance with our results, previous studies had found that NP inhibited the activity of *Nitrosomonas, Nitrosospira,* and *Nitrosolobus* strains at concentrations $\ge 0.86 \mu M$ (18). Comparative tests with various terrestrial and marine AOA (*Nitrososphaera* sp. JG1,

288 Nitrosarchaeum koreense MY1, Nitrosopumilus maritimus SCM1, and Nitrosopumilus cobalaminigenes HCA1), and AOB (including N. europaea and N. multiformis), showed 289 an inhibitory effect of NP at 10 µM (56, 51, 57). In line with our findings for "Ca. N. 290 291 sinensis" (EC₅₀ = 6.7 μ M), Lehtovirta-Morley *et al.*, (22) reported that NP halted the activity and growth of the phylogenetically closely related "Ca. N. devanaterra" ND1 at 292 293 concentrations $\geq 10 \ \mu$ M. In contrast, Shen *et al.*, (23) reported significantly higher EC₅₀ values for NP (> 173 μ M and 118.1 μ M for N. multiformis and N. viennensis, respectively). 294 295 This could be due to the different approach used by Shen et al. (23) for adding NP into the 296 cultures. This involved adding non-dissolved NP in the culture medium to achieve concentrations in the range of 40-173 μ M, with the highest level corresponding to the upper 297 limit of NP water solubility at 20°C (40 mg L⁻¹) entailing a risk for precipitation of the 298 299 active compound during incubation at 28°C (N. multiformis) or 37°C (N. viennensis). Soil studies with NP, although limited, are influenced by the relative functional dominance of 300 301 one group of AO over another and the applied concentration of the inhibitor. Cui et al., (27) reported preferential inhibition of AOB by NP applied at 1.3 µmol Kg⁻¹. Conversely, 302 303 Lehtovirta-Morley *et al.*, (22) found AOA to be sensitive to NP at concentrations $\geq 10 \mu$ mol Kg⁻¹ in an acidic soil where "Ca. N. devanaterra" ND1 was dominant. 304

A comparative analysis of the inhibitory range of the tested NIs highlights the serious practical implications of our findings (Fig. 6). DMPP and NP were equally effective and the most potent NIs against AOB, followed by QI, (EQ), and DCD, with QI being more active than DCD against *N. multiformis*. On the other hand, QI (EQ) and NP were equally effective and the most active NIs against AOA, while DCD and DMPP, were not inhibitory to AOA at the concentrations tested (Fig. 6). These findings suggest that NP is the only

311 commercial NI capable of effectively inhibiting both AOB and AOA, and hence the most 312 effective currently available NI, although it is not currently registered for use in Europe. 30% of the World's soils have a pH <5.5 and European agricultural soils have a mean pH 313 314 of 5.8 (58). These results therefore have practical implications for low pH soils where 315 ammonia oxidation may be dominated by AOA (41). Slight differences in the inhibition 316 thresholds between AOA and AOB may largely affect agricultural practice, since AOA are expected to contribute to nitrogen fertilizer loss in conditions where AOB would be 317 inhibited (22). On the other hand, universal inhibitory effects on both AOB and AOA (and 318 319 comammox bacteria) suggest that nitrification inhibition will not be compromised by 320 functional redundancy. Alternatively, the use of mixtures of NIs exhibiting complementary activity against different AO groups or targeting different parts of the ammonia oxidation 321 pathway could be equally efficient with broad range NIs. In this regard, the potential use 322 323 of EQ as a novel NI is promising, considering its unique feature to be transformed in soil 324 to QI, which is a highly potent inhibitor of AOA, and has a satisfactory inhibitory effect on 325 AOB, comparable with that of established NIs such as DCD.

326 In addition to the practical implications, the considerable differences in the range 327 of inhibitory activities exhibited by the tested NIs might indicate differences in their mode 328 of action not considered before. For example, DCD, DMPP and NP, all considered as Cuchelators, showed variable activity towards AOA, which, unlike AOB, rely on copper-329 330 containing proteins for electron transfer (59). NP on the other hand was previously 331 proposed to serve also as an alternative AMO substrate, generating products (6chloropicolinic acid) that irreversibly deactivate ammonia oxidation (10). This dual 332 333 inhibitory mechanism of NP might offer an explanation for its more universal inhibitory

334 activity towards both AOA and AOB. EO and its derivatives possess high-antioxidative capacity acting as free radical scavengers (46). As EQ and QI showed similar inhibitory 335 effects to other NO-scavengers (e.g. PTIO) (57), their effectivity against AOA may be due 336 337 to a similar mode of action. Alternatively, as QI is a strong antioxidant, it could be involved 338 in oxidative stress-related cell disruption particularly in AOA, with AOB being capable of 339 coping with oxidative stress using catalases which are largely absent in AOA (60). Further studies should define the inhibition mechanism of EQ on AO and clarify the corresponding 340 mechanisms of the other NIs which remain unknown. 341

342 Given the effect of NIs on AO, they will also have an indirect inhibitory effect on 343 NOB, the functional partners of AO. We demonstrated that DMPP and DCD, despite their strong impact on AOB, had no inhibitory effect on *Nitrobacter* sp. NHB1, in contrast to 344 NP, EQ and its derivatives which were more active against *Nitrobacter* sp. NHB1. Previous 345 346 soil studies also demonstrated that DCD was not suppressive to Nitrospira- and *Nitrobacter*-like bacteria at levels up to 150 µmol Kg⁻¹ soil (61, 62, 49). To date there are 347 348 no data, either *in vitro* or in soil, regarding the impact of DMPP or EQ and its derivatives 349 on NOB, while NP applied at rates up to 50 μ M did not inhibit the nitrite-oxidizing activity 350 of the widely distributed Nitrobacter agilis (20). We provide the first evidence for the 351 toxicity of NIs on a *Nitrobacter* sp. Further studies extended to other NOB, including the widely distributed and diverse *Nitrospira*-like bacteria, would determine the full inhibitory 352 potential of NIs on soil NOB. 353

In parallel to activity and growth measurements, we determined the degradation and transformation of the tested NIs to identify potential links between the duration of exposure (persistence) and the effects observed. The total residues of EQ showed limited persistence 357 in the AOA and NOB cultures ($DT_{50} = 2.4-8.7$ days), and low to moderate persistence in the AOB cultures ($DT_{50} = 8.7-60.1$ days), a difference most likely attributed to abiotic 358 factors such as medium pH (acidic for AOA and NOB vs. alkaline for AOB) rather than an 359 360 enzymatic transformation, considering the autotrophic lifestyle of the tested isolates (60) 361 and the recalcitrance of EQ under aerobic and anaerobic conditions (63). However, a direct 362 interaction of these compounds with the tested organisms cannot not be excluded (10). The three commercial NIs showed remarkably different stability in the liquid cultures. DCD 363 showed moderate to high persistence ($DT_{508} = 44.5$ to >1000 days), whereas NP degraded 364 365 rapidly ($DT_{50} = 0.12 \cdot 12.5$ days). DMPP showed a high persistence in all liquid cultures, with the exception of Nitrobacter cultures where a great variation in the persistence of 366 DMPP was observed. Considering that *Nitrobacter* sp. NHB1 and AOA were cultured in 367 media of similar content and pH, the above variation was possibly induced by interaction 368 369 between the NI and the Nitrobacter sp. strain. Overall, we did not observe any clear 370 correlations between NIs persistence and inhibition potency.

371

372 **Conclusions.** We determined *in vitro* the inhibition potential of novel (EQ and its derivatives) and all NIs currently used in the agricultural practice on representative soil-373 374 derived AOA, AOB, and NOB. EQ, and primarily its major transformation product QI, 375 showed high potency against AOA, in contrast to DCD and DMPP, the only NIs currently 376 registered for use in Europe and which are inhibitory only to AOB. In contrast, NP showed 377 an inhibitory activity against all groups tested. The activity of those NIs on comammox 378 bacteria are still unknown due to the lack of soil-derived isolates, and their characterization will be required to provide a complete understanding of their inhibition potency on all AO. 379

380 Our study (i) offers benchmarking knowledge of the activity range of known and potentially new NIs to soil AO and Nitrobacter NOB, (ii) introduces the novel potential 381 NI EQ, which possesses desirable characteristics including transformation into another 382 383 potent NI, QI, which has high potency against AOA in contrast to other registered NIs in 384 Europe, and (iii) demonstrates the different sensitivity of AOA and AOB to NIs, providing 385 potentially novel strategies relying on new broad-range NIs, or more likely, using mixtures of NIs which possess complementary activity against different nitrifier groups. Further 386 elucidation of EQ and QI inhibitory mechanisms, and verification of their efficacy under 387 388 diverse soil conditions affecting both the activity of AO and the performance of NIs, might 389 lead to the development of a novel NI for more efficient N conservation in agricultural 390 soils.

391

392 MATERIAL AND METHODS

Microbial strains, growth conditions and chemicals. Five soil-derived nitrifying isolates were used in the *in vitro* assays: two AOB (*N. europaea*, *N. multiformis*), two AOA ("*Ca.* N. franklandus", "*Ca.* N. sinensis"), and one NOB (*Nitrobacter* sp. NHB1). All strains were grown aerobically in the dark without shaking. Details on the cultivation media and incubation temperatures used are given in the Supplemental Material.

Analytical standards of DCD (99 % purity), NP (\geq 98 %), and EQ (95%) were purchased from Sigma-Aldrich (Germany), while the DMPP (99.1%) analytical standard was provided by BASF Hellas. The oxidation derivatives of EQ, QI and EQNL were synthesized as described by Thorisson *et al.* (64). The chemical structures of all studied compounds are shown in Fig. S7.

403 Liquid culture assays. The activity of all NIs was determined in liquid batch cultures over a broad range of concentrations to establish relevant inhibition thresholds 404 (EC₅₀ values) per strain and compound. For each, triplicate strain x NI x concentration 405 406 replicates were established in 100-mL Duran bottles containing 50 mL of growth medium 407 and inoculated with a 1 or 2% (v/v) transfer from exponentially growing cultures of AOB 408 or AOA/NOB, respectively. EQ, QI, EQNL and NP were added to the cultures as filter sterilized DMSO solutions due to their low water solubility (<60 mg L⁻¹ at 20°C) and the 409 410 final concentration of DMSO in all cultures was 0.1% (v/v). DCD and DMPP were 411 dissolved in sterile dH₂O before addition of 25 μ l (0.5% v/v) in the different cultures. All NIs were added to batch cultures at the beginning of the exponential growth phase. For all 412 assays, cultures were established in triplicate with the same inoculum without NI 413 amendment. Upon inoculation all liquid batch cultures were sampled at regular time 414 415 intervals to determine the effect of NIs on the activity and growth of AO by measuring 416 changes in nitrite concentrations, and amoA or nxrB gene abundance for AO and NOB 417 populations, respectively.

418 Nitrite measurements and gene abundance quantification. Nitrite concentrations were determined colorimetrically at 540 nm in a 96-well plate format assay 419 420 by diazotizing and coupling with Griess reagent (65). *amoA* and *nxrB* gene abundance was determined in a Biorad CFX Real-Time PCR system. DNA was extracted from a cell pellet 421 422 obtained from 2-ml aliquots of the microbial cultures using the tissue DNA extraction kit 423 (Macherey-Nagel, Germany). The amoA genes of AOB and AOA was amplified with primers amoA-1F/amoA-2R (66) and Arch-amoAF/Arch-amoAR (67), respectively as 424 425 described by Rousidou et al., (68). The nxrB gene of Nitrobacter was quantified with

primers nxrB-1F and nxrB-1R (69) using the following thermal cycling conditions: 95°C for 3 min, followed by 40 cycles of 95°C for 30 seconds, 57°C for 20 seconds, 72°C for 30 seconds, with a final dissociation curve analysis. The abundance of *amoA* and *nxrB* genes were determined via external standard curves as described by Rousidou *et al.*, (68). qPCR amplification efficiencies ranged from 80.3% to 109.4%, with r^2 values ≥ 0.98 .

Nitrification Inhibitors extraction. EQ, QI, EQNL, and NP residues were extracted from liquid media by mixing 0.3 mL liquid culture with 0.7 mL of acetonitrile. Residues of DCD and DMMP were extracted by mixing 0.1 mL liquid culture with 0.9 mL of ddH₂O water and methanol, respectively. The derived mixtures were vortexed for 30 s and stored at -20° C until analysis. Recovery tests at three concentration levels (in the range of the tested concentrations) showed recoveries of >80% for all compounds studied.

Chromatographic analyses. High performance liquid chromatography (HPLC) 437 analyses were performed in a Shimadzu LC-20ADHPLC system equipped with an UV/VIS 438 439 PDA detector. A Shimadzu GVP-ODs (4.6 mm by 150mm, 5µm) pre-column, connected to a RP Shimadzu VP-ODs (4.6 mm x 150 mm, 5µm) column, was used for NI separation. 440 The injection volume was 20 µl. The flow rate of the mobile phase was set at 0.8 mL min⁻ 441 ¹ for DCD, and at 1 mL min⁻¹ for all other NIs. Column temperature was set at 40°C for 442 DCD and DMPP, and at 25°C for all the other NIs. Mixtures of acetonitrile and ammonia 443 444 (0.25% [vol/vol]) or ortho-phosphoric acid (0.1% [vol/vol]) were used at a ratio of 70:30 (vol/vol) for mobile phases in the analyses of EO, OI, EONL, and NP, respectively, and 445 detection was achieved at 225, 245, 230, and 269 nm, respectively. Similarly, 446 chromatographic separation of DCD and DMPP was achieved using ddH_2O (100%) and a 447

mixture of methanol and *ortho*-phosphoric acid (0.1% [vol/vol]) solution 50:50 by volume,
respectively. DCD and DMPP residues were detected at 218 nm and 225 nm, respectively.

Calculation of inhibition threshold levels (EC₅₀). In this study, EC₅₀ describes 450 451 the concentration of the inhibitor that reduces half of the activity (nitrite accumulation or 452 consumption) of AO or NOB, with dose-response modeling performed using normalized data whereby nitrite concentration values were divided by the mean value of the matching 453 454 control. Analyses were carried out using the dose response curves (drc) v3.0-1 package 455 (70) of the R software (71). A brief description of the tested models can be found in Ritz 456 et al., (72). An empirical modeling approach was initially used for selecting the best fitting 457 model according to tested goodness of fit indices (see Supplemental Material), followed by the choice of the four-parameter log logistic model as the best compromise among tested 458 models for comparing endpoint values. 459

Data analysis. Nitrite and qPCR data were subjected to one-way ANOVA, 460 461 followed by Tukey's post hoc test (P < 0.05). Variance between the EC₅₀ values of the different NIs for one strain and between different strains for a given NI was analyzed by 462 463 one-way ANOVA, and Duncan post hoc test (P < 0.05). The four kinetic models proposed by the FOCUS working group on pesticide degradation kinetics (73) (SFO and the biphasic 464 465 models hockey stick (HS), first order multi-compartment (FOMC), and double first order 466 in parallel (DFOP)) were used to calculate NI degradation kinetic parameters (DT₅₀, k_{deg}). Curve fitting was performed with the mkin v0.9.47.1 (74) package of the R v3.4.3 software 467 468 (71). More details on degradation kinetics are given in the Supplemental Material.

469

470 ACKNOWLEDGMENTS

471	Th	is work is part of the project "NITRIC – Looking up for Novel nITR ification Inhibitors:
472	Ne	w stories with old Compounds" which has received funding from the Hellenic
473	Fo	undation for Research and Innovation (HFRI) and the General Secretariat for Research
474	anc	l Technology (GSRT), under grant agreement No. 1229.
475		The authors declare no conflict of interest.
476		
477	RF	CFERENCES
478	1.	Fowler D, Coyle M, Skiba U, Sutton MA, Cape JN, Reis S, Sheppard LJ, Jenkins A,
479		Grizzetti B, Galloway JN, Vitousek P, Leach A, Bouwman AF, Butterbach-Bahl K,
480		Dentener F, Stevenson D, Amann M, Voss M. 2018. The global nitrogen cycle in the
481		twenty first century. Phil Trans R Soc B 368: 20130164.
482	2.	Raun WR, Johnson GV. 1999. Improving nitrogen use efficiency for cereal production.
483		Agronomy J 9: 357-363.
484	3.	Abbasi MK, Adams WA. 1998. Loss of nitrogen in compacted grassland soil by
485		simultaneous nitrification and denitrification. Plant Soil 200: 265–277.
486	4.	Moir JL, Cameron KC, Di HJ. 2007. Effects of the nitrification inhibitor
487		dicyandiamide on soil mineral N, pasture yield, nutrient uptake and pasture quality in
488		a grazed pasture system. Soil Use Manag 23: 111-120.
489	5.	McCarty GW. 1999. Modes of action of nitrification inhibitors. Biol Fertil Soils
490		29:1–9.
491	6.	Goring CAI. 1962. Control of nitrification of ammonium fertilizers and urea by 2-
492		chloro-6-(trichloromethyl)-pyridine. Soil Sci 93: 211–218.

493	7.	Solansky S. 1982. N-Stabilisator SKW-DIDIN verbessert die Stickstoffwirkung der
494		Gülle. Blickfeld 61:1–4.

Zerulla W, Barth T, Dressel J, Erhardt K, von Locquenghien KH, Pasda G, Rädle M,
 Wissemeier AH. 2001. 3,4-Dimethylpyrazole phosphate (DMPP) – A new nitrification
 inhibitor for agriculture and horticulture – An introduction. Biol Fertil Soils 34: 79–
 84.

- 499 9. Ruser R, Schulz R. 2015. The effect of nitrification inhibitors on the nitrous oxide
 500 (N₂O) release from agricultural soils a review. J Plant Nutr Soil Sci 178:171-188.
- Vannelli T, Hooper AB. 1992. Oxidation of nitrapyrin to 6- chloropicolinic acid by the
 ammonia-oxidizing bacterium *Nitrosomonas europaea*. Appl Environ Microbiol 58:
 2321-2325.
- 504 11. Subbarao GV, Ito O, Sahrawat KL, Berry WL, Nakahara K, Ishikawa T, Watanabe T,
 505 Suenaga K, Rondon M, Rao IM. 2006. Scope and Strategies for Regulation of
 506 Nitrification in Agricultural Systems Challenges and Opportunities. Crit Rev Plant
 507 Sci 25:303–335.
- 12. Coskun D, Britto DT, Shi W, Kronzucker, HJ. 2017. Nitrogen transformations in
 modern agriculture and the role of biological nitrification inhibition. Nat Plants
 3:17074.
- 511 13. Beeckman F, Motte H, Beeckman T. 2018. Nitrification in agricultural soils: impact,
 512 actors and mitigation. Curr Opin Biotechnol 50: 166–173.
- 513 14. Könneke M, Bernhard AE, de la Torre JR, Walker CB, Waterbury JB, Stahl DA. 2005.
- Isolation of an autotrophic ammonia- oxidizing marine archaeon. Nature 437:543-546.

515	15.	Tourna M, Stieglmeier M, Spang A, Könneke M, Schintlmeister A, Urich T, Engel M,
516		Schloter M, Wagner M, Richter A, Schleper C. 2011. Nitrososphaera viennensis, an
517		ammonia oxidizing archaeon from soil. Proc Natl Acad Sci U S A 108:8420-8425.
518	16.	Daims H, Lebedeva EV, Pjevac P, Han P, Herbold C, Albertsen M, Jehmlich N,
519		Palatinszky M, Vierheilig J, Bulaev A, Kirkegaard RH, von Bergen M, Rattei T, van
520		Kessel MAHJ, Speth DR, Albertsen M, Nielsen PH, Op den Camp HJM, Kartal B,
521	17.	van Kessel MAHJ, Speth DR, Albertsen M, Nielsen PH, Huub JM, Op den Camp
522		HJM, Kartal B, Jetten MSM, Lücker S. 2015. Complete nitrification by a single
523		microorganism. Nature 528:555-559.
524	18.	Belser LW, Schmidt EL. 1981. Inhibitory effect of nitrapyrin and the three genera of
525		ammonia-oxidizing nitrifiers. Appl Environ Microb 41: 819-821.
526	19.	Jones RD, Morita RY, Griffiths RP. 1984. Method for estimating chemolithotrophic
527		ammonium oxidation using carbon monoxide oxidation. Mar Ecol Prog Ser 17: 259-
528		269.
529	20.	Matsuba D, Takazaki H, Sato Y, Takahashi R, Tokuyama T, Wakabayashi K. 2003.
530		Susceptibility of Ammonia-Oxidizing Bacteria to Nitrification Inhibitors. Z
531		Naturforsch 58c: 282-287.
532	21.	Liu Y, Yang Y, Qin H-L, Zhu Y-J, Wei W-X. 2014. Differential responses of nitrifier
533		and denitrifier to dicyandiamide in short- and long-term intensive vegetable cultivation
534		soils. J Integr Agric 13: 1090-1098.
535	22.	Lehtovirta-Morley LE, Verhamme DT, Nicol GW, Prosser JI. 2013. Effect of
536		nitrification inhibitors on the growth and activity of Nitrosotalea devanaterra in
537		culture and soil. Soil Biol Biochem 62:129-133.

538	23. Shen T, Stieglmeier M, Dai J, Urich T, Schleper C. 2013. Responses of the terrestrial
539	ammonia-oxidizing archaeon Ca. Nitrososphaera viennensis and the ammonia-
540	oxidizing bacterium Nitrosospira multiformis to nitrification inhibitors. FEMS
541	Microbiol Lett 344:121-129.

24. Di HJ, Cameron KC, Shen JP, Winefield CS, O'Callaghan M, Bowatte S, He JZ. 2010. 542 543 Ammonia-oxidizing bacteria and archaea grow under contrasting soil nitrogen conditions. FEMS Microb Ecol 72: 386-394.

545 25. Di HJ, Cameron KC. 2011. Inhibition of ammonium oxidation by a liquid formulation

544

546 of 3,4-dimethylpyrazole phosphate (DMPP) compared with a dicyandiamide (DCD)

solution in six New Zealand grazed grassland soils. J Soil Sediment 11: 1032–1039. 547

26. Kleineidam K, Kôsmrlj K, Kublik S, Palmer I, Pfab H, Ruser R, Fiedler S, Schloter 548

M. 2011. Influence of the nitrification inhibitor 3,4-dimethylpyrazole phosphate 549

550 (DMPP) on ammonia-oxidizing bacteria and archaea in rhizosphere and bulk soil. 551 Chemosphere 84: 182–186.

27. Cui P, Fan F, Yin C, Li Z, Song A, Wan Y, Liang Y. 2013. Urea- and nitrapyrin-552 affected N2O emission is coupled mainly with ammonia oxidizing bacteria growth in 553 microcosms of three typical Chinese arable soils. Soil Biol Biochem 66: 214-221. 554

- 555 28. Florio A, Clark IM, Hirsch PR, Jhurreea D, Benedetti A. 2014. Effects of the 556 nitrification inhibitor 3,4-dimethylpyrazole phosphate (DMPP) on abundance and activity of ammonia oxidizers in soil. Biol Fertil Soils 50:795-807. 557
- 29. Yang W, Wang Y, Tago K, Tokuda S, Hayatsu M. 2017. Comparison of the Effects of 558 Phenylhydrazine Hydrochloride and Dicyandiamide on Ammonia-Oxidizing Bacteria 559 560 and Archaea in Andosols. Front Microbiol 8:2226

561	30. Prosser JI, Nicol GW. 2012. Archaeal and bacterial ammonia-oxidisers in soil: th
562	quest for niche specialisation and differentiation. Trends Microbiol 20: 523-531.

- 563 31. Kits KD, Sedlacek CJ, Lebedeva EV, Han P, Bulaev A, Pjevac P. Daebeler A,
- Romano S, Albertsen M, Stein LY, Daims H, Wagner M. 2017. Kinetic analysis of a

complete nitrifier reveals an oligotrophic lifestyle. Nature 549: 269-272.

- 32. Papadopoulou ES, Tsachidou B, Sułowicz S, Menkissoglu-Spiroudi U, Karpouzas
 DG. 2016. Land spreading of wastewaters from the fruit packaging industry and
 potential effects on soil microbes: effects of the antioxidant ethoxyquin and its
 metabolites on ammonia oxidizers. Appl Environ Microbiol 82: 747–755.
- 33. Karas P, Metsoviti A, Zisis V, Ehaliotis C, Omirou M, Papadopoulou E,
 Menkissoglou-Spiroudi U, Manta S, Komiotis D, Karpouzas DG.2015. Dissipation,
 metabolism and sorption of pesticides used in fruit-packaging plants: towards an
 optimized depuration of their pesticide-contaminated agro-industrial effluents. Sci
 Total Environ 530–531: 129–139.
- 575 34. Kozlowski JA, Stieglmeier M, Schleper C, Klotz MG, Stein LY. 2016. Pathways and
- key intermediates required for obligate aerobic ammonia-dependent chemolithotrophy
 in bacteria and *Thaumarchaeota*. ISME J 10: 1836–1845.
- 35. Alves RJE, Minh BQ, Urich T, von Haeseler A, Schleper C. 2018. Unifying the global
 phylogeny and environmental distribution of ammonia-oxidising archaea basedon
 amoA genes. Nat Commun 9:1517.
- 36. Daims H, Lücker S, Wagner M. 2016. A New Perspective on Microbes Formerly
 Known as Nitrite-Oxidizing Bacteria. Trends Microbiol 24: 699-712.

583	37	Purkhold U	Pommerening-Röser	А	Juretschko	S	Schmid MC	Koons	HP	Wagner

- 584 M. 2000. Phylogeny of All Recognized Species of Ammonia Oxidizers Based on
- 585 Comparative 16S rRNA and amoA Sequence Analysis: Implications for Molecular
- 586 Diversity Surveys. Appl Environ Microbiol 66: 5368–5382.
- 38. Kowalchuk GA, Stephen JR. 2001. Ammonia-Oxidizing Bacteria: A Model for
 Molecular Microbial Ecology. Annu Rev Microbiol 55:485–529.
- 39. Norton JM. 2011. Diversity and environmental distribution of ammonia-oxidizing
 bacteria.In Nitrification, Ward BB, Arp DJ, Klotz MG. ASM Press, Washington, DC,
 pp. 39-57.
- 40. Lehtovirta-Morley LE, Ross J, Hink L, Weber EB, Gubry-Rangin C, Thion C, Prosser
- JI, Nicol GW. 2016. Isolation of "*Candidatus* Nitrosocosmicus franklandus," a novel
 ureolytic soil archaeal ammonia oxidiser with tolerance to high ammonia
 concentration. FEMS Microbiol Ecol 92: fiw057.
- 596 41. Lehtovirta-Morley LE, Ge C, Ross J, Yao H, Nicol GW, Prosser JI. 2014.
- 597 Characterisation of terrestrial acidophilic archaeal ammonia oxidisers and their 598 inhibition and stimulation by organic compounds. FEMS Microbiol Ecol 89: 542–552.
- 42. Herbold CW, Lehtovirta-Morley LE, Jung MY, Jehmlich N, Hausmann B, Han P, Loy
- 600 A, Pester M, Sayavedra-Soto LA, Rhee S-K, Prosser JI, Nicol GW, Wagner M, Gubry-
- Rangin C. (2017) Ammonia-oxidising archaea living at low pH: insights from
 comparative genomics. Environ Microbiol 19: 4939–4952.
- 43. De Boer, W, Klein Gunnewiek PGA, Veenhuis LM, Bock E, Laanbroek HJ. 1991.
- Nitrification at Low pH by Aggregated Chemolithotrophic Bacteria. Appl Environ
 Microbiol 57: 3600-3604

606	44. Xia W, Zhang C, Zeng X, Feng Y, Weng J, Lin X, Zhu J, Xiong Z, Xu J, Cai Z, Jia Z.
607	2011. Autotrophic growth of nitrifying community in an agricultural soil. ISME J 5:
608	1226–1236.

609 45. Nowka B, Daims H, Spieck E. 2015. Comparison of Oxidation Kinetics of Nitrite-610 Oxidizing Bacteria: Nitrite Availability as a Key Factor in Niche Differentiation. Appl

Environ Microbiol 81: 745–753. 611

46. Błaszczyk AB, Augustyniak A, Skolimowski J. 2013. Ethoxyguin: an antioxidant used 612 613 in animal feed. Int J Food Sci 2013:585931.

47. O'Callaghan M, Gerard EM, Carter PE, Lardner R, Sarathchandra U, Burch G. Ghani 614

A, Bell N. 2010. Effect of the nitrification inhibitor dicyandiamide (DCD) on microbial 615 616 communities in a pasture soil amended with bovine urine. Soil Biol Biochem 42:1425-1436. 617

48. Di HJ, Cameron KC, Podolyan A, Robinson A. 2014. Effect of soil moisture status 618 619 and a nitrification inhibitor, dicyandiamide, on ammonia oxidizer and denitrifier

growth and nitrous oxide emissions in a grassland soil. Soil Biol Biochem 73: 59-68. 620

621 49. Fu Q, Clark IM, Zhu J, Hu H, Hirsch PR. 2018. The short-term effects of nitrification

inhibitors on the abundance and expression of ammonia and nitrite oxidizers in a long-622

term field experiment comparing land management. Biol Fertil Soils 54:163–172. 623

624 50. Jung M-Y, Park S-J, Min D, Kim J-S, Rijpstra WIC, Damste JSS, Kim G-J, Madsen

- EL, Rhee S-K. 2011. Enrichment and Characterization of an Autotrophic Ammonia-625
- Oxidizing Archaeon of Mesophilic Crenarchaeal Group I.1a from an Agricultural Soil. 626
- 627 Appl Environ Microbiol 77: 8635-8647.

628	51.	Kim J-G, Jung M-Y, Park S-J, Rijpstra WIC, Damsté JSS, Madsen EL, Min D, Kim J-
629		S, Kim G-J, Rhee S-K. 2012. Cultivation of a highly enriched ammonia-oxidizing
630		archaeon of thaumarchaeotal group I.1b from an agricultural soil. Environ Microbiol
631		14:1528–1543.
632	52.	Shi X, Hu H-W, Móller C, He J-Z, Chen D, Suter HC. 2016. Effects of the Nitrification
633		Inhibitor 3,4-Dimethylpyrazole Phosphate on Nitrification and Nitrifiers in Two
634		Contrasting Agricultural Soils. Appl Environ Microb 82: 5236-5248.
635	53.	Torralbo F, Menéndez S, Barrena I, Estavillo J.M, Marino D, González-Murua C.
636		2017. Dimethyl pyrazol-based nitrification inhibitors effect on nitrifying and
637		denitrifying bacteria to mitigate N2O emission. Sci Rep 7:13810.
638	54.	Zhang M, Wang W, Bai SH, Zhou X, Teng Y, Xu Z. 2018. Antagonistic effects of
639		nitrification inhibitor 3,4-dimethylpyrazole phosphate and fungicide iprodione on net
640		nitrification in an agricultural soil. Soil Biol Biochem 116:167-170.
641	55.	Vilas MP, Verburg K, Thorburn, PJ, Probert ME, Bonnett GD. 2019. A framework for
642		analysing nitrification inhibition: A case study on 3,4-dimethylpyrazole phosphate
643		(DMPP). Sci Total Environ 672 : 846–854.
644	56.	Jung M-Y, Park S-J, Min D, Kim J-S, Rijpstra WIC, Damste' JSS, Kim G-J, Madsen
645		EL, Rhee S-K. (2011). Enrichment and characterization of an autotrophic ammonia-
646		oxidizing archaeon of mesophilic crenarchaeal group I.1a from an agricultural soil.

647 Appl Environ Microbiol 77: 8635-8647.

57. Martens-Habbena W, Qin W, Horak RE, Urakawa H, Schauer AJ, Moffett JW, 648

Armbrust EV, Ingalls AE, Devol AH, Stahl DA. 2015. The production of nitric oxide 649

- by marine ammonia-oxidizing archaea and inhibition of archaeal ammonia oxidation 650
- by a nitric oxide scavenger. Environ Microbiol 17: 2261-74. 651

- 58. Fabian C, Reimann C, Fabian K, Birke M, Baritz R, Haslinger E, The GEMAS Project
- Team. 2014. GEMAS: Spatial distribution of the pH of European agricultural andgrazing land soil. Appl Geochem 48: 207-216.
- 55 59. Walker CB, de la Torre JR, Klotz MG, Urakawa H, Pinel N, Arp DJ, Brochier-
- Armanet C, Chain PS, Chan PP, Gollabgir A, Hemp J, Hügler M, Karr EA, Könneke
- 657 M, Shin M, Lawton TJ, Lowe T, Martens-Habbena W, Sayavedra-Soto LA, Lang
- D, Sievert SM, Rosenzweig AC, Manning G, Stahl DA. 2010. *Nitrosopumilus maritimus* genome reveals unique mechanisms for nitrification and autotrophy in
 globally distributed marine crenarchaea. Proc Natl Acad Sci USA 107: 8818-8823.
- 661 60. Kim J-G, Jung M-Y, Park S-J, Damsté JSS, Schouten S, Rijpstra WIC, Jung M-Y,
- Kim S-J, Gwak J-H, Hong H, Si O-J, Lee S, Madsen EL, Rhee S-K. 2016. Hydrogen
 peroxide detoxification is a key mechanism for growth of ammonia-oxidizing archaea.
 Proc Natl Acad Sci USA 113: 7888–7893.
- 665 61. Clough TJ, Ray JL, Buckthought LE, Calder J, Baird D, O'Callaghan, Sherlock RR,
- 666 Condron LM. 2009. The mitigation potential of hippuric acid on N2O emissions from
- urine patches: An in-situ determination of its effect. Soil Biol Biochem 41: 2222–2229.
- 668 62. Carneiro J, Cardenas ZLM, Hatch DJ, Trindade H, Scholefield D, Clegg CD, Hobbs
- P. 2010. Effect of the nitrification inhibitor dicyandiamide on microbial communities
 and N₂O from an arable soil fertilized with ammonium sulphate. Environ Chem Lett
 8: 237–246.
- 672 63. Shah AG, Pierson JA, Pavlostathis SG. 2005. Fate and effect of the antioxidant
 673 ethoxyquin on a mixed methanogenic culture. Water Res 39: 4251–4263.
- 674 64. Thorisson S, Gunstone FD, Hard R. 1992. Some oxidation products of ethoxyquin
- 675 including those found in autoxidising systems. Chem Phys Lipids 60:263–271.

- 676 65. Shinn MB. 1941. Colorimetric method for determination of nitrite. Ind Eng Chem 13:677 33–35.
- 678 66. Rotthauwe JH, Witzel KP, Liesack W. 1997. The ammonia monooxygenase structural
- gene *amoA* as a functional marker: molecular fine-scale analysis of natural ammonia-
- 680 oxidizing populations. Appl Environ Microbiol 63:4704–4712.
- 681 67. Francis CA, Roberts KJ, Beman JM, Santoro AE, Oakley BB. 2005. Ubiquity and
- diversity of ammonia-oxidizing archaea in water column sand sediments of the ocean.
- 683 Proc Natl Acad Sci U S A 102:14683–14688.
- 684 68. Rousidou C, Papadopoulou E, Kortsinidou M, Giannakou IO, Singh BK,
 685 Menkissoglu-Spiroudi U, Karpouzas DG. 2013. Bio-pesticides:harmful or harmless to
 686 ammonia oxidizing microorganisms? The case of a *Paecilomyces lilacinus*-based
 687 nematicide. Soil Biol Biochem 67:98–105.
- 688 69. Vanparys B, Spieck E, Heylena K, Wittebollec L, Geetsc J, Boonc N, De Vosa P.
 2007.The phylogeny of the genus *Nitrobacter* based on comparative rep-PCR, 16S
- rRNA and nitrite oxidoreductase gene sequence analysis. Syst Appl Microbiol 30:
 297–308.
- 692 70. Ritz C, and Streibig JC. 2005. Bioassay Analysis using R. Journal of Statistical693 Software 12.
- 694 71. R Core Team. 2017. R: A language and environment for statistical computing,
 695 reference index version 3.4.3. 2017. http://www.R-project.org.
- 696 72. Ritz C, Baty F, Streibig JC, Gerhard D. 2016. Dose-response analysis using R. PLoS
 697 ONE 10: e0146021.
- 698 73. FOCUS. 2006. Guidance Document on Estimating Persistence and Degradation
 699 Kinetics from Environmental Fate Studies on Pesticides in EU Registration. Report of

- the FOCUS Work Group on Degradation Kinetics, EC Document Reference
 Sanco/10058/2005 Version, 2.0, 2006, p. 434.
- 702 74. Ranke J. 2018. mkin: Kinetic Evaluation of Chemical Degradation Data. R package
 703 version 0.9.47.1. https://CRAN.R-project.org/package=mkin.
- 704
- 705 TABLES
- **Table 1.** Mean EC₅₀ values (μ M) of the tested nitrification inhibitors (NIs) on ammonia or nitrite oxidation activity of the nitrifying strains. Standard errors of the mean values are given in brackets. Upper case letters indicate significant differences (p<0.05) between microorganisms for each individual NI, and lower-case letters indicate significant differences (p<0.05) between NIs for each tested microorganism.
- **Table 2.** Mean concentrations \pm standard errors (μ M) of Quinone Imine (QI) and Ethoxyquinoline (EQNL) formed in the liquid cultures of the nitrifying isolates amended with Ethoxyquin (EQ) (i) at the onset of inhibition, (ii) at the time when maximum concentration levels were detected. The timepoint (days) at which each measurement was taken is given in brackets.
- 716

717 FIGURE LEGENDS

Fig. 1. The effect of different concentrations of Ethoxyquin (EQ) (a), Quinone Imine (QI)
(b), Ethoxyquinoline (EQNL) (c), Dicyandiamide (DCD) (d), Nitrapyrin (NP) (e), and
DMPP (f), on nitrite production by *Nitrosomonas europaea*. Error bars represent the

standard error of the mean of triplicate cultures. Arrows indicate the time point when thenitrification inhibitor (NI) was added.

Fig. 2. The effect of different concentrations of Ethoxyquin (EQ) (a), Quinone Imine (QI) (b), Ethoxyquinoline (EQNL) (c), Dicyandiamide (DCD) (d), Nitrapyrin (NP) (e), and DMPP (f) on nitrite production by *Nitrosospira multiformis*. Error bars represent the standard error of the mean of triplicate cultures. Arrows indicate the time point when the nitrification inhibitor (NI) was added.

- **Fig. 3.** The effect of different concentrations of Ethoxyquin (EQ) (a), Quinone Imine (QI)
- (b), Ethoxyquinoline (EQNL) (c), Dicyandiamide (DCD) (d), Nitrapyrin (NP) (e), and
- 730 DMPP (f) on nitrite production by "*Candidatus* Nitrosocosmicus franklandus". Error bars
- represent the standard error of the mean from triplicate cultures. Arrows indicate the time
- point when the nitrification inhibitor (NI) was added.
- **Fig. 4.** The effect of different concentrations of Ethoxyquin (EQ) (a), Quinone Imine (QI)
- (b), Ethoxyquinoline (EQNL) (c), Dicyandiamide (DCD) (d), Nitrapyrin (NP) (e), and
- 735 DMPP (f) on nitrite production by "*Candidatus* Nitrosotalea sinensis". Error bars represent
- the standard error of the mean of triplicate cultures. Arrows indicate the time point whenthe nitrification inhibitor (NI) was added.
- **Fig. 5.** The effect of different concentrations of Ethoxyquin (EQ) (a), Quinone Imine (QI)
- (b), Ethoxyquinoline (EQNL) (c), Dicyandiamide (DCD) (d), Nitrapyrin (NP) (e), and
- 740 DMPP (f) on nitrite transformation by Nitrobacter sp. Error bars represent the standard
- ror of the mean from triplicate cultures. Arrows indicate the time at which the nitrification
- 742 inhibitor (NI) was added.

Fig. 6. Heatmap representation of the qualitative impact of different concentrations of NIs
on the nitrifying activity of soil ammonia- and nitrite-oxidizing isolates. The level of
inhibition observed was classified in the following categories: "Inhibition" (shown in red);
"Inhibition and Recovery" (shown in orange); "No inhibition" (shown in green); "Not
tested" (shown in white).

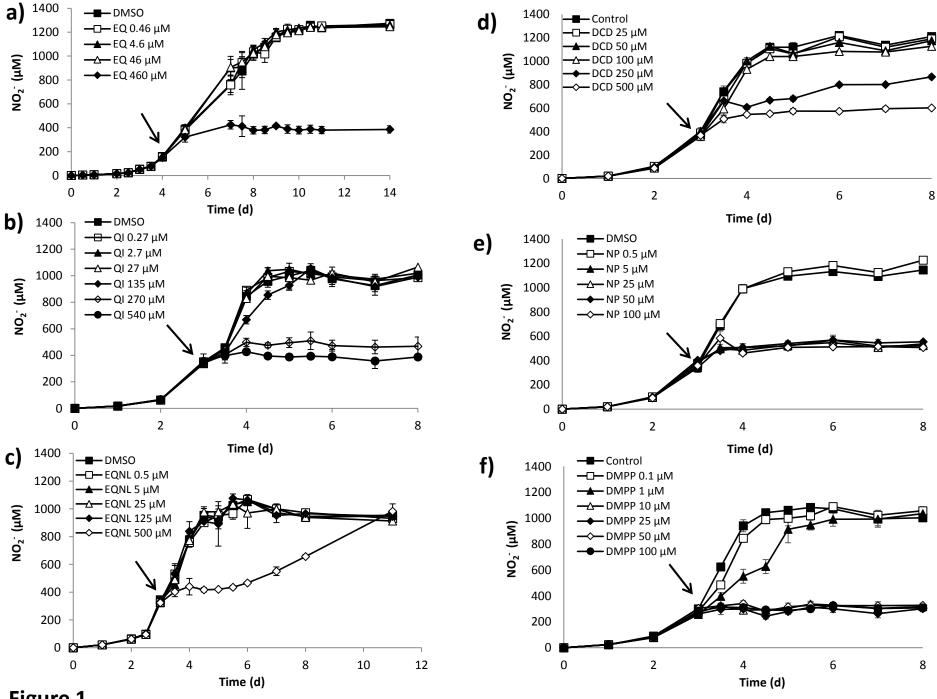
Table 1. Mean EC₅₀ values (μ M) of the tested nitrification inhibitors (NIs) to the ammonia or nitrite oxidation activity of the nitrifying strains. Standard errors of the mean values are given in brackets. Upper case letters indicate significant differences (p < 0.05) between microorganisms regarding each tested NI and lower-case letters indicate significant differences (p < 0.05) between NIs within each microorganism.

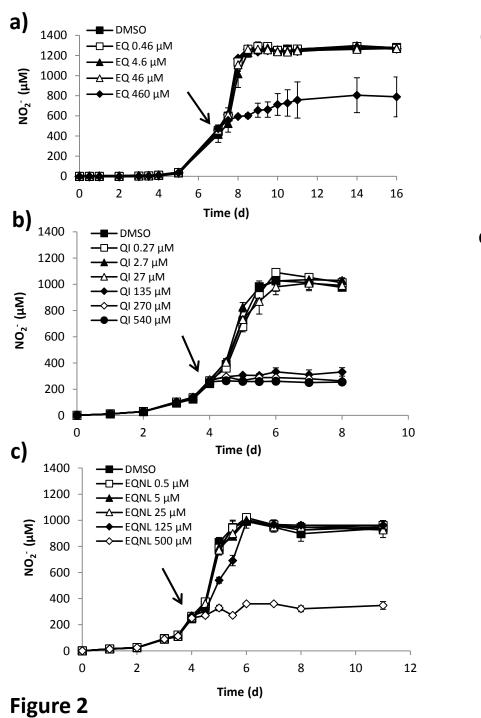
Microorganisms	Ethoxyquin	Quinone Imine	Ethoxyquinoline	DCD	Nitrapyrin	DMPP
	(EQ)	(QI)	(EQNL)		(NP)	
Nitrosomonas europaea	181.4 (23.3) bB	199.8 (10.8) bC	543.4 (111.5) cC	221.9 (29.0) bA	2.1 (0.4) aA	2.1 (0.7) aA
Nitrosospira multiformis	214.8 (39.6) bB	65.1 (6.0) aB	360.5 (105.7) cB	248.7 (7.4) bcA	0.8 (0.3) aA	0.6 (0.1) aA
Ca. Nitrosocosmicus franklandus	1.4 (0.3) aA	0.7 (0.4) aA	129.5 (25.0) aA	1568.5 (237.1) bB	1.0 (0.3) aA	1773.7 (439.9) bB
Ca. Nitrosotalea sinensis	1.0 (0.4) aA	0.3 (0.0) aA	26.6 (5.7) aA	477.8 (56.6) cA	6.7 (1.8) aA	359.5 (43.1) bA
Nitrobacter sp.	166.7 (53.5) aB	247.2 (65.7) aD	562.0 (38.5) aC	>100000*	167.8 (41.2) aB	12581.3 (1979.2) bC

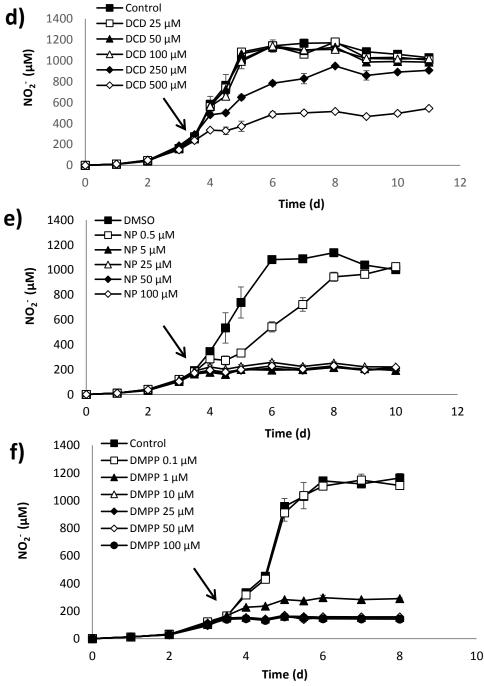
*Maximum tested concentration

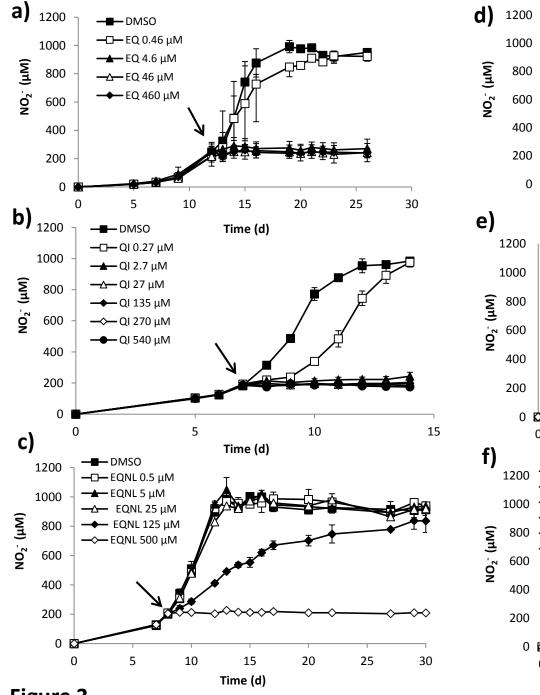
Table 2. Mean concentrations \pm standard errors (μ M) of Quinone Imine (QI) and Ethoxyquinoline (EQNL) formed in the liquid cultures of the nitrifying isolates amended with Ethoxyquin (EQ) (i) at the onset of inhibition, (ii) at the time when maximum concentration levels were detected. The timepoint (days) at which each measurement was taken is given in brackets.

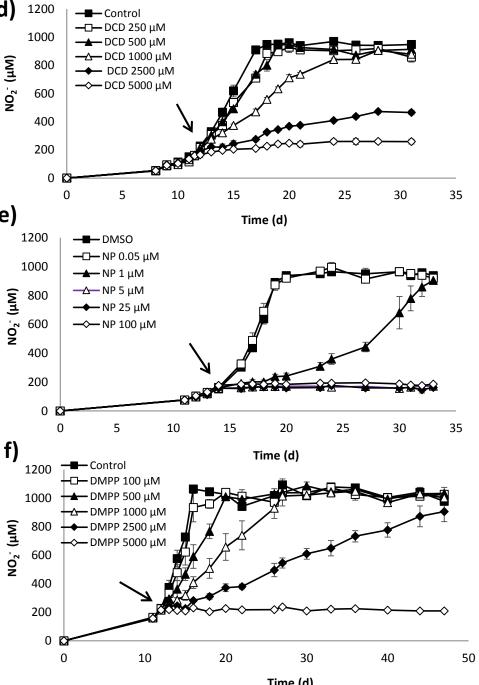
Ethoxyquin treatment		Concentration (µM)			
		Inhibition Onset		Maximum concentration formed	
		QI	EQNL	QI	EQNL
AOB	N. europaea- 460 μM	95.4 ±2.1 (7d)	3.42±0.2 (7d)	170.2±1.4 (14d)	3.42±0.2 (7d)
	N. multiformis- 460 µM	36.6±3.4 (8d)	8.98±0.9 (8d)	167.6±16.1 (16d)	10.3±0.8 (14d)
ΑΟΑ	" <i>Ca.</i> N. franklandus" – 460 μ M	74.9±0.9 (15d)	8.80±0.1 (15d)	96.3±1.5 (19d)	10.3±0.2 (22d)
	" <i>Ca</i> . N. franklandus" – 46 μM	4.97±0.7 (15d)	0.61±0.3 (15d)	6.75±0.1 (25d)	0.61±0.4 (22d)
	" <i>Ca.</i> N. franklandus" – 4.6 μ M	1.78±0.1 (15d)	0.0 (0.0)	1.83±0.1 (16d)	0.0 (0.0)
	" <i>Ca.</i> N. sinensis" – 460 μM	132.6±1.5 (7d)	16.9±0.4 (7d)	189.8±13.8 (10d)	27.4±1.1 (17d)
	" <i>Ca.</i> N. sinensis" – 46 μM	13.3±0.4 (7d)	2.3±0.1 (7d)	13.5±0.7 (6d)	3.4±1.0 (17d)
	" <i>Ca.</i> N. sinensis" – 4.6 μM	1.0±0.0 (7d)	0.20±0.0 (7d)	4.3±0.4 (5d)	0.4±0.1 (27d)
NOB	Nitrobacter sp 460 µM	50.8±13.1 (4d)	11.5±0.9 (4d)	114.5±1.8 (9d)	14.2± 0.6 (3d)











Time (d)

