- 1 Divergence in photoperiod responses of a classical biological control agent *Neogalerucella*
- 2 *calmariensis* (Coleoptera: Chrysomelidae) across a climatic and latitudinal gradient
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6 Abstract

7 A key knowledge gap in classical biological control is to what extent insect agents evolve 8 to novel environments. The introduction of biological control agents to new photoperiod regimes 9 and climates may disrupt the coordination of diapause timing that evolved to the growing season 10 length in the native range. We tested whether populations of *Neogalerucella calmariensis* (L.) 11 have evolved in response to the potential mismatch of their diapause timing since their 12 intentional introduction to the United States from Germany in the 1990s. Populations collected 13 from 39.4° to 48.8° latitude in the western USA were reared in growth chambers to isolate the 14 effects of photoperiod on diapause induction and development time. For all populations, shorter 15 day lengths increased the proportion of beetles that entered diapause instead of reproducing. The 16 critical photoperiods, or the day length at which half of a population diapauses, differed 17 significantly among the sampled populations, generally decreasing at lower latitudes. The 18 latitudinal trend reflects changes in growing season length, which determines the number of 19 generations possible, and in local day lengths at the time beetles are sensitive to this cue. 20 Development times were similar across populations, with one exception, and did not vary with 21 photoperiod. These results show that there was sufficient genetic variation from the two German 22 source populations to evolve different photoperiod responses across a range of environmental 23 conditions. This study adds to the examples of rapid evolution of seasonal adaptations in introduced insects. 24

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26 Keywords: Classical biological control, critical photoperiod, diapause, rapid evolution, voltinism

27 Introduction

28 A key concern in classical biological control is whether insect agents evolve to novel 29 environmental conditions and how this should inform introduction strategies in the new range 30 (McEvoy et al. 2012, Roderick et al. 2012). Evolution may be expected in traits regulating insect 31 lifecycles as local adaptations to the growing season length and environmental cues in the native 32 range are disrupted (Saikkonen et al. 2012, Grevstad and Coop 2015). Multivoltine insects, 33 which could potentially fit more than one generation in a year, have evolved genetic-based 34 responses to environmental cues that determine when populations switch developmental 35 pathways from reproduction to diapause in advance of winter (Danilevskii 1961, Tauber et al. 36 1986, Denlinger et al. 2017). Day length, which varies seasonally and across latitudes, is the 37 principal cue that insects in temperate zones use, in conjunction with other cues from maternal 38 effects, weather, or host plants, to initiate diapause development (Tauber et al. 1986, Bradshaw 39 and Holzapfel 2007). The response typically takes the form of a switch from reproductive to 40 diapause states when a sensitive life stage is exposed to a day length that is shorter than a 41 threshold value. This is defined as the critical day length or critical photoperiod, used 42 interchangeably in the literature to describe an individual's threshold or the median threshold in a 43 population.

The ubiquity of latitudinal clines in the critical photoperiod that induces diapause in native and long-established insects (reviewed by Masaki 1961, Danilevskii 1965, Tauber et al. 1986) suggests that similar selective pressures will drive geographic divergence in these seasonal adaptations in introduced insects. For two main reasons, the critical photoperiod is expected to evolve to be longer at higher latitudes: (1) summer day lengths used as cues are longer at higher latitudes and (2) shorter growing seasons at higher latitudes mean that insects need to induce

diapause earlier in the summer when days are longer to prepare for an earlier winter onset. While
day length for a given date is a strict function of latitude, growing season length is more locally
variable depending on its climate.

53 In this study, we experimentally compared the diapause response and development rates 54 in seven populations of *Neogalerucella calmariensis* (L.) (formerly *Galerucella calmariensis* 55 L.), a leaf beetle introduced across North America as a biological control agent for the wetland 56 weed purple loosestrife (Lythrum salicaria L.). Our study is preceded by a phenology model 57 applied to this insect (Grevstad and Coop 2015) that demonstrates how introductions to new 58 photoperiod and climate regimes may disrupt the coordination of diapause timing that evolved to 59 the growing season length and photoperiod regime at the population's origin. Newly introduced 60 insects can experience mismatches between the number of generations attempted based on 61 photoperiod-induced diapause and the *potential* number of generations that could be possible 62 with a well-adapted photoperiodic response (Bean et al. 2007, Grevstad and Coop 2015). 63 Attempting reproduction too late in the year could lead to high mortality as life stages unable to 64 diapause confront winter or senescing host plants (Van Dyck et al. 2015, Kerr et al. 2020). 65 Entering diapause too early in the growing season can lead to greater overwintering mortality 66 because of the added duration of diapause over the summer (Bean et al. 2007). Other adaptations 67 might include faster development to fit full generations within the available growing season or 68 accelerated development in response to shortened day lengths toward the end of the season 69 (McEvoy et al. 2012, Lindestad et al. 2019). The mismatches between season length and thermal 70 time requirements to complete the final generation set up selective pressure for the photoperiod 71 response, development time, and voltinism to adapt to novel climates and day lengths at different 72 latitudes.

73 Biological control introductions draw from limited source populations in the agent's 74 native range so that prerelease host range tests accurately assess the risk of nontarget attacks 75 (Hinz et al. 2014). Constrained genetic variation may limit the success of agents released across 76 a wide range of environmental conditions (Roderick et al. 2012, Wright and Bennett 2018). 77 While often assumed that biological control agents undergo a period of adaptation following 78 introduction, documented examples are limited to a few cases (Wright and Bennett 2018). They 79 include changes in springtime phenology in the ragwort flea beetle Longitarsus jacobaeae (Szűcs 80 et al. 2012), divergence in developmental rates between alpine and valley populations of the 81 cinnabar moth Tyria jacobaeae (McEvoy et al. 2012), and a shortening of the critical 82 photoperiod that cues diapause in the tamarisk leaf beetle *Diorhabda carinulata* (Bean et al. 83 2012). In the case of the D. carinulata, the evolution in photoperiod response allowed 84 populations to increase voltinism and extend geographic range boundaries southward into 85 regions where the original photoperiod response initially prevented establishment (Bean et al. 86 2007, 2012). 87 For *N. calmariensis*, we predicted that photoperiod responses had evolved to correspond 88 with season length and day length cues across a latitudinal gradient, based on the potential for 89 mismatched diapause timing modeled in Grevstad and Coop (2015). In this study, we reared 90 beetles in controlled environmental chambers set with different artificial day lengths to test for 91 genetic differences in photoperiod thresholds for diapause induction and photoperiod-dependent

93 diverged across populations in a manner consistent with local adaptation.

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development rates. We show that, within 27 years post-introduction, these two traits have

94 Materials and Methods

95 <u>Study system</u>

96 Neogalerucella calmariensis was first introduced in 1992 across the United States and 97 Canada as a classical biological control agent for purple loosestrife (L. salicaria), a noxious 98 invasive weed common in wetlands (Hight et al. 1995). It established populations at sites with 99 different climate regimes and photoperiod exposure (36-53°N) compared to its two source 100 populations in Germany (50 and 54°N) (Hight et al. 1995, Corrigan et al. 2013). Neogalerucella 101 calmariensis was introduced alongside its congener Neogalerucella pusilla (Duftschmidt), which 102 coexists in the same habitat, or even on the same plant, without apparent differences in life 103 history or ecological niche (Blossey 1995). We focus on N. calmariensis in this study, although 104 N. pusilla was also present at some of the sites. We identified N. calmariensis in the field and 105 later in the lab by their larger size, dark stripes on their elytra, and orange-brown coloration 106 (Manguin et al. 1993).

107 Neogalerucella calmariensis has a facultative multivoltine lifecycle in North America. 108 Adults overwinter and emerge as reproductive regardless of photoperiod exposure in spring 109 (Blossey 1995, Bartelt et al. 2008). Based on degree-day (base 10°C) requirements, oviposition 110 from overwintered adults occurs around 100 accumulated degree-days and subsequent 111 development from egg to adult to oviposition requires around 523 degree-days (McAvoy and 112 Kok 2004, Grevstad and Coop 2015). Newly eclosed (teneral) adults from subsequent 113 generations are photoperiod sensitive and will enter diapause in response to short days (Velarde 114 et al. 2002) and non-reproductive adults will start oviposition at a delay if exposed to longer days 115 (Grevstad 1999). The baseline critical photoperiod (=critical day length) in beetles introduced 116 from Germany was not measured experimentally. However, soon after release, it was estimated

117 as <15.2 day hours by Bartelt et al. (2008) and between 15-15.5 day hours by Grevstad (1999) 118 based on conditions inducing oviposition during rearing. Individual variation in diapause 119 induction at the same day length exposure shows genetic variation in the critical photoperiod 120 within these source populations (Velarde et al. 2002, Bartelt et al. 2008). As in other insects, 121 additional factors may modulate the photoperiod-based diapause response (Tauber et al. 1986). 122 Source populations in Germany were reported as primarily univoltine, with some 123 oviposition from the summer generation before entering diapause (Blossey 1995). In the United 124 States, 1 to 3 generations have been observed depending on location. The lifecycle model of 125 Grevstad and Coop (2015) predicts that up to 4 generations are possible for N. calmariensis in 126 the warmest parts of the new range based on available degree-days. However, its short-day 127 diapause response would restrict it to fewer generations. Voltinism in the field, which may refer 128 to either the completed or attempted (and not completed) number of generations, can be 129 predicted with the assumptions that critical photoperiods approximate the responses to day length 130 cues in natural settings and that day length provides the principal environmental cue. If day 131 length was longer than the critical photoperiod during the stage at which N. calmariensis are 132 sensitive, then most of the population would be reproductive. 133 We studied seven beetle populations that established in the states of California, Oregon, 134 and Washington after their introduction in the 1990s (Figure 1). Environmental conditions at 135 these sites vary with elevation (17-1,006 meters above sea level) and with latitude (39.41°) 136 48.76°) (Table 1). For each location, we summarized the average annual season length in degree-137 days over 25 years (1994-2018) using Daymet daily gridded interpolations of weather station 138 observations (Thornton et al. 1997, 2017). Grevstad and Coop's (2015) degree-day model was

139 used to predict average phenology of the F1 teneral adult eclosion, which varies with latitude and

140 the day of year (Forsythe et al. 1995). We selected sites with established populations of beetles 141 without recent transfers to our knowledge. The original releases in Washington used beetles from 142 both northern and southern German sources, while those in Oregon only used beetles from the 143 southern German source (Hight et al. 1995). California populations were introduced, often 144 multiple times, from established populations in Oregon and Washington (M. Pitcairn personal 145 communication) and could not establish south of 39° latitude (Pitcairn 2018). We summarize the 146 known release history and translocations in Table 1, but other exchanges between populations 147 that would alter their genetic variation and evolution are possible. In 2014, voltinism in the field 148 was determined by the presence or absence of larvae in the second and third (where appropriate) 149 generation after first confirming beetles were abundant early in the season. These surveys were 150 timed to coincide with the predicted peak larval abundance rather than adults, because of the 151 greater possibility for ambiguity about their generation of origin.

152 Experimental setup

153 We carried out two similar experiments (in 2014 and 2019) to compare the diapause 154 response to photoperiod among N. calmariensis populations and one experiment (in 2018) to 155 compare their rates of development from egg to adult. All experiments were performed in SG3-156 22 controlled environmental chambers (Hoffman Manufacturing, Inc.) equipped with six 32-watt 157 5000K tube fluorescent lights on the walls of each chamber. We reared populations in this 158 controlled environment to attribute phenotypic variation to genetic differences between 159 populations (Merilä and Hendry 2014). Beetles were collected from source populations in 2014 160 and 2018 for our lab colonies. At least one generation of rearing in a common environment 161 preceded each experiment to control for maternal effects.

162	We grew L. salicaria host plants from seeds collected in Portland, Oregon (45.471°,
163	-122.657°). Seedlings were transplanted to square 2.2-L plastic nursery pots with LA4PC potting
164	soil (Sun Gro Horticulture H, Agawam, MA) and 5 grams of 14-14-14 slow-release fertilizer.
165	Plants were maintained in standing water in a greenhouse with day/night temperatures of
166	approximately 27/21°C and supplemental grow lights that turned on to extend the day length to
167	16 hours. We trimmed L. salicaria frequently to prevent flowering and to encourage growth of
168	new shoots from axillary buds. When using cut leaves to feed N. calmariensis, we used new
169	growth including shoot apical meristems and the first few sets of leaves.
170	The experiments differed in the method details as described below.
171	Photoperiod-based diapause: 2014 experiment
172	Neogalerucella calmariensis eggs and early instar larvae were collected from established
173	field populations in Palermo, California and Sutherlin, Oregon in early May and adults were
174	collected near Ephrata and Bellingham, Washington in late May. We reared them in a
175	greenhouse inside 13.5 x 13.5 x 24-inch mesh cages on potted L. salicaria until the next
176	generation's adults emerged. After the new adults started mating, we obtained same-age cohorts
177	of eggs by placing 3 adult pairs on each of 12 caged L. salicaria (15-20 cm tall) for each
178	population. This resulted in roughly 90 eggs per plant. When larvae started to hatch, whole plants
179	were placed into growth chamber treatments (3 plants x 4 populations x 4 photoperiod
180	treatments) and beetles developed to adults. The photoperiod treatments in chambers were
181	14.75/9.25, 15.25/8.75, 15.75/8.25, and 16.25/7.75 hours of light/dark (L/D) per day. To provide
182	additional light for plant growth, two 24-watt 6500K tube fluorescent tube lights were suspended
183	above the plants as a supplement to the built-in lights on the walls of the chamber. These were
184	set to go on 15 minutes after and off 15 minutes before the built-in lights to emulate twilight.

185 Chambers were maintained at 23°C constant temperature for day and night. Adults were 186 removed from plants as they emerged and grouped in 150mm-diameter petri dishes in the 187 chambers with cut leaves for the pre-oviposition or pre-diapause development time. 188 We assessed the developmental status (reproductive versus diapausing) in all adult 189 females (556 individuals total) 10-14 days after their emergence, which was sufficient time for 190 completion of pre-oviposition or pre-diapause development. Adults were placed individually into 191 50mm-diameter petri dishes with fresh leaves and moist filter paper for 48 hours. We assigned 192 beetles as reproductive if oviposition occurred. Indeterminate behavior, such as feeding with no 193 eggs, led to an extra 24-hour observation before classification. Diapause beetles had no feeding 194 and were often hiding and inactive underneath the filter paper. The proportion of females that 195 were reproductive for each treatment combination (pooled across plants) was the dependent 196 variable for our analysis. 197 Development time: 2018 experiment 198 To determine if development rates differed among populations of *N. calmariensis*, we 199 carried out a separate experiment using the same environmental chambers. We intended this 200 experiment to measure both diapause response and development timing. However, a high 201 diapause rate across all photoperiods (at 21°C) meant that we were unable to obtain and compare diapause response curves. Therefore, we present it to compare development times between 202 203 populations and test whether shorter photoperiods accelerate development. 204 We collected 50-100 adult beetles from two sites in Washington (Bellingham and 205 Yakima Training Center) and two sites in Oregon (Rickreall and Sutherlin) between April 26 and 206 May 8, 2018. Beetles from the two sites in California (McArthur and Palermo) were collected a 207 month later. After rearing them for a generation in a greenhouse, we placed 30-35 beetles on

208 individual, caged potted L. salicaria for 24 hours to establish same-age cohorts of eggs. After 209 removing the adults, the plants with affixed eggs were placed in a constant 21°C growth chamber 210 at 16.75/7.75 L/D hours. Shortly before hatching, eggs from the same cohort were distributed to 211 150mm-diameter petri dishes with fresh leaves lined with moist filter paper at a density of 30 to 212 50 eggs per dish. Each dish was randomly assigned to photoperiod treatments of 14.75/9.25, 213 15.75/8.25, 16.25/7.75, and 16.75/7.25 L/D hours (3 cohorts x 6 populations x 4 photoperiod 214 treatments). We exchanged fresh leaves as needed and added 10 grams of potting soil substrate 215 for pupation to each dish when the beetles reached the third instar. Adults were removed and 216 counted every day as they emerged. In total, we tracked the development of 1808 beetles. In our 217 statistical analysis, the duration from oviposition to adult eclosion for each individual was the 218 dependent variable compared across populations.

219 Photoperiod-based diapause: 2019 experiment

220 Diapausing beetles from the two shortest day length treatments in the 2018 experiment 221 were stored outdoors in mesh cages with leaf litter from November 2018 through April 2019 in 222 Corvallis, Oregon. We only used beetles from these two photoperiod treatments, which all went 223 into diapause, so that we were not artificially selecting beetles based on their photoperiod 224 response. These beetles emerged reproductive in the spring and their offspring entered this 225 experiment. Experimental methods differed from the diapause experiment carried out in 2014 in 226 that larvae were reared in 150mm-diameter petri dishes (approximately 30 individuals per dish), 227 no supplemental lights were added to the chambers (used only built-in lights), and both males 228 and females were assessed for diapause. We used the same rearing methods as the 2018 229 development time experiment, but with a higher constant temperature of 23°C and an extended 230 range of photoperiod treatments compared to the 2014 and 2018 experiments. We assigned

231	rearing dishes from each population to photoperiod treatments at 14.5/9.5, 15.0/9.0, 15.5/8.5,
232	16.0/8.0, 16.5/7.5, and 17.0/7.0 L/D hours (2 dishes x 6 populations x 6 photoperiod treatments).
233	As they emerged each day, we removed and re-pooled teneral adults into group trays with fresh
234	leaves for the pre-oviposition or pre-diapause development period. In the diapause behavior test
235	(described earlier), 1,782 beetles were scored as reproductive female, reproductive male, or
236	diapause (sex not determined) according to their oviposition, feeding, and activity. The
237	proportion reproductive for each treatment combination pooled across rearing dishes was the
238	dependent variable for our analysis.
239	Statistical analysis
240	We performed all analysis and visualization with R 3.6.2 (R Core Team 2019). First, each
241	year's diapause data (2014 and 2019) were analyzed separately for beetle reproductive rates with
242	generalized linear mixed effects models with a binomial error distribution and logit link using the
243	<i>lme4</i> package (Bates et al. 2015). The fixed effects of population and photoperiod treatment
244	(scaled continuous covariate) and their interaction predicted the proportion reproductive in each
245	treatment combination, accounting for the total number of beetles in each group. As these groups
246	contained beetles from the same population reared in the same growth chamber, they were
247	assigned random intercepts to account for non-independence and allow for overdispersion in the
248	residual variance (Bolker 2015). We removed the interaction term if non-significant ($p > 0.05$)
249	by a Chi-square test. Population differences in responses (estimated marginal means) were
250	compared using the emmeans package with post-hoc pairwise comparisons while adjusting for
251	multiple comparisons using Tukey's method (Lenth 2020). Populations' critical photoperiods
252	(=critical day lengths), the hours of light exposure at which half of the beetles are predicted to

253	diapause, were derived with 95% confidence intervals using inverse prediction from the
254	generalized linear mixed models (Venables and Ripley 2002).

255 Second, we tested for a latitudinal gradient in the diapause response by analyzing the 256 2014 and 2019 experiments in the same model. We changed the predictor variables in the 257 generalized linear mixed models to shift the focus from populations' critical photoperiods to the 258 latitudinal cline in the trait. The fixed effects were photoperiod treatment, latitude (scaled), year 259 of experiment, and all interactions. We removed interaction terms if non-significant (p > 0.05) by 260 a Chi-square test. The two experiments differed in methods such that the results from their 261 separate models (see above) had substantial differences in the slope of the diapause response to 262 photoperiod treatments. By including year and its interactions with other variables in this model, 263 we account for the potential different responses to photoperiod treatment across the latitudinal 264 gradient that might be artifacts from the experiments' methods. In addition to the random 265 intercept for treatment combination to control for the non-independence of beetles reared in the 266 same chamber, we included a random intercept for population because we are attempting to 267 make inferences about the larger, unsampled set of beetle populations along the latitudinal 268 gradient rather than just these seven populations (Bolker 2015). The predicted change in critical 269 photoperiods across the latitudinal gradient was derived from this model with 95% confidence 270 intervals using simulations from the model parameters and their uncertainty (Population 271 Prediction Intervals in Bolker (2008)).

Third, we tested if development time was affected by population and photoperiod exposure with a generalized linear mixed effects model with a gamma error distribution and log link using the *lme4* package (Bates et al. 2015). The fixed effects of population and photoperiod treatment (scaled continuous covariate) and their interaction predicted the duration from

276	oviposition to adult eclosion in days for each beetle. Rearing dishes, the experimental unit, with
277	beetles from the same population and oviposition day, were assigned random intercepts to
278	account for their non-independence. We removed the interaction term if non-significant ($p >$
279	0.05) by a Chi-square test. Population differences in responses (estimated marginal means) were
280	compared using the emmeans package with post-hoc pairwise comparisons while adjusting for
281	multiple comparisons using Tukey's method (Lenth 2020).
282	Results
283	Introduced populations have established at sites with average season lengths from 1067-
284	2769 cumulative growing degree-days (base 10°C) and two months difference in the predicted
285	eclosion of F1 adults that would be sensitive to the photoperiod cues that induce diapause.
286	Surveys confirmed voltinism of 1-3 generations in the field in 2014 (Table 1).
287	Photoperiod-based diapause
288	All populations exhibited a short-day diapause response where shorter photoperiod
289	treatments induced diapause and longer photoperiod treatments resulted in reproduction (Table 2,
290	Figure 2, Figure 3). Populations differed in their critical photoperiods by up to 0.7 hours in the
291	2014 experiment and 2.3 hours in the 2019 experiment (Table 2). There is no interaction between
292	population and photoperiod response in either experiment, meaning that the slopes of the
293	photoperiod response curves do not significantly differ across populations (Table 3).
294	Populations had a lower proportion reproductive, regardless of photoperiod, at higher
295	latitudes (β = -0.918, SE = 0.393, P = 0.0196, Table 4). The 2019 experiment had marginally
296	higher proportions reproductive ($\beta = 0.747$, SE = 0.414, $P = 0.0709$, Table 4) and a significant
297	interaction with the photoperiod response ($\beta = -2.97$, SE = 0.643, $P = <0.0001$, Table 4). This
298	interaction is visualized by the steeper slopes of the photoperiod response in 2014 (Figure 2)

17.4) minutes per 5° latitude in the 2014 experiment and 51.0 (CI 45.0-56.9) minutes per 5°
latitude in the 2019 experiment (Figure 4).
However, the six populations tested in 2019 do not have a monotonic increase in critical
photoperiod with latitude (Figure 4, Table 2). For example, the two central populations
(Rickreall and Sutherlin, Oregon) have the longest (16.8, CI 16.3-17.3) and shortest (14.5, CI
14.0-15.1) critical photoperiods, respectively (3.87 difference, SE = 0.551 , $P < 0.001$ on logit
scale). Populations from McArthur, California and Yakima Training Center, Washington have
similar critical photoperiods in spite of their separation of 5.7° latitude (0.00679 difference, SE =
0.502, <i>P</i> > 0.999).
Development time
Development time averaged 33.5 days and did not change with photoperiod treatments
(Table 3, Figure 5). Populations had similar development times with the exception that the
Rickreall, Oregon population required an additional day of development compared to the
Sutherlin, Oregon population (0.0419 difference on log scale, SE = 0.0127, $P = 0.0122$) and the
Yakima Training Center, Washington population (0.0364 difference on log scale, SE = 0.0127, P
= 0.0471). There was no interaction between population and photoperiod for development time
(Table 3).
Discussion

Nearly three decades after their introduction, populations of the introduced leaf beetle
 Neogalerucella calmariensis have genetically diverged in their diapause response to photoperiod
 across 9.4° of latitude in the western USA. Despite a genetic bottleneck at the time of their

introduction from Germany in the 1990's, the measured critical photoperiods, where 50% of the 322 323 population enters diapause, range from 14.5 to 16.8 hours of daylight, greatly extending the 324 estimated 15 to 15.5 hour critical photoperiod in the original population. The measured critical 325 photoperiods roughly correlate with latitude, with longer critical photoperiods found in the north 326 where summer days are longer and where conditions favorable for reproduction tend to end 327 earlier in the year (Figure 4). Critical photoperiods were shorter in the south where summer day 328 lengths are shorter and conditions usually remain favorable later into the autumn. 329 The evolved geographic variation resembles the photoperiod response clines commonly 330 seen in native or long established insects. Typical latitudinal gradients for critical photoperiods in 331 locally-adapted insects in temperate zones range from 60-90 minutes per 5° latitude in classic 332 experiments (Danilevskii 1961) or 48 minutes per 5° latitude in a recent meta-analysis 333 (Joschinski and Bonte 2020). For the studied N. calmariensis populations, we estimated critical 334 photoperiod clines of 15.7 minutes and 51.0 minutes per 5° latitude in 2014 and 2019 335 experiments, respectively (Figure 4). The lower rates of change with latitude found in this study 336 may reflect incomplete adaptation in this introduced insect, or may be a result of the particular 337 locations sampled (see below). If we assume that the latitudinal clines in native insects represent 338 adaptations to the local seasonal environment, then we can say that *N. calmariensis* has evolved 339 in an adaptive direction. However, reciprocal transplant experiments would be needed to confirm 340 that these changes in photoperiod responses are local adaptations that increase fitness (Kawecki 341 and Ebert 2004, Merilä and Hendry 2014, Tsai et al. 2020). 342 We found that development rates remain similar among populations with the exception of 343 a slower development rate, by one day, in the Rickreall, Oregon population (Figure 5). The small

344 difference, in comparison to the wider variation found in critical photoperiods, suggests that

345 photoperiod response plays a relatively larger role in adapting to local seasonal regimes 346 (Bradshaw and Holzapfel 2007). It is intriguing that this population with slower development 347 also has a longer critical photoperiod than its neighboring populations. Field observations from 348 recent years confirm a single generation at this site, whereas the Sutherlin, Oregon population 349 just 1.6° latitude to the south and unsampled populations within 0.5° latitude to the north have 2 350 generations (Table 1, F. Grevstad personal observations). It is unclear why this population should 351 be limited to one generation given sufficient degree days for two generations and plants that 352 remain green until early fall. Relaxation of selection for faster development may be a result of 353 the predominantly univoltine lifecycle at this site. Other insects have evolved nonlinear 354 "stepped" or "sawtooth" clines in developmental traits across latitudes in tandem with locally-355 adapted photoperiod responses, which has the added effect of limiting attempted generations in 356 places where season length is only marginally supportive of increased voltinism (Roff 1980, 357 Tauber and Tauber 1981, Levy et al. 2015).

358 Measured critical photoperiods are useful in combination with developmental degree-day 359 requirements for predicting voltinism at a given location over time (Beck and Apple 1961, Tobin 360 et al. 2008, Kerr et al. 2020). The relationship of a population's mean critical photoperiod to 361 modeled voltinism for a given climate and latitude has been covered in detail in Grevstad and 362 Coop (2015). In brief, a degree-day phenology model predicts when the photoperiod-sensitive 363 stage emerges, and if the day length on that date for the latitude is longer than the critical 364 photoperiod then the population is reproductive and attempts another generation. Voltinism can 365 shift (sometimes dramatically) when an insect is moved to a new climate or latitude and with annual variation in phenology of the sensitive stage. However, laboratory-measured critical 366 367 photoperiods may differ that those used in the field, where multiple cues interact and conditions

368 vary from rearing experiments. For example, compared to the abrupt transition between light and 369 dark in the lab, we do not know what proportion of the twilight period N. calmariensis perceives 370 as part of the day length for comparison against the critical photoperiod. Where this has been 371 studied in other insects, the portion of twilight sensed as part of the photoperiod appears be quite 372 variable among insect species (Menzel 1979) and may even vary asymmetrically between dawn 373 and dusk (Takeda and Masaki 1979). Other laboratory conditions, such as constant temperatures, 374 may also change the estimated critical photoperiods. Our 2018 experiment anecdotally had much 375 lower reproductive rates (and presumably higher critical photoperiods) after rearing at 2°C cooler 376 temperatures. In other beetles, rearing at constant temperatures increased the measured critical 377 photoperiod by 15 minutes compared to a fluctuating regime with the same mean (Bean et al. 378 2007). Now that our study has demonstrated genetic differences in a simplified experimental 379 setting, field validation would help gauge how diapause initiation changes in the presence of 380 twilight, temperature fluctuations, density-dependence, and host plant defenses.

381 The estimated critical photoperiods vary substantially between experiments and about the 382 latitudinal trend. We cannot rule out that the photoperiod response evolved over a few years 383 (Bean et al. 2012, Urbanski et al. 2012) or even changes annually with strong fluctuating 384 selection (Bell 2010). However, methodological differences likely caused variation between 385 years, such as the higher precision of critical photoperiods estimated in 2014. The use of whole 386 potted plants for rearing in the 2014 experiment could have provided beetles with additional cues 387 from the host plant to induce a stronger diapause response (Izzo et al. 2014). The inclusion of 388 males in the sample in 2019, but not 2014, may have led to increased standard deviation in the 389 critical photoperiod (Table 2) and the marginally significant higher proportion of reproduction 390 beetles (Table 4). There are also climatic and phenological reasons why populations' critical

391 photoperiods vary from the latitudinal trend. In the western United States, local climates are 392 influenced by complex topography relative to the eastern United States. The sawtooth pattern in 393 critical photoperiods estimated in 2019 (Figure 4) correspond to differences in average degree-394 days and sensitive stage phenology at the sampled sites (Table 1). Local climate determines the 395 number of generations that are possible in a location and the evolved critical photoperiod may 396 vary over short distances at similar latitudes (Lindestad et al. 2019). Insect phenology models 397 that incorporate demography along with photoperiod-based diapause can estimate annual fitness 398 for a particular critical photoperiod (Kerr et al. 2020), and extending these models across years 399 could answer whether a critical photoperiod has reached an optimum for a population's 400 environment (similar to trait models in Kivelä et al. (2013)).

401 The photoperiod response in *N. calmariensis* appears to be quite variable among 402 individuals, with a mix of reproductive and diapausing beetles in the same treatment groups 403 (Figures 2 and 3). While less individual variation, and a steeper slope in the photoperiod 404 response curve, may show stronger selection on the critical photoperiod (Tauber et al. 1986), 405 some insect species show substantial individual variation with flatter slopes (Joschinski and 406 Bonte 2020). In the field, this variability can translate into partial generations toward the end of 407 the season, where a portion of adults within the same generation enters diapause while the rest go 408 on to reproduce. Two years of observations suggest that this commonly occurs in colonies reared 409 outdoors in Corvallis, OR (T. Wepprich personal observation). The extent to which this occurs in 410 natural populations of *N. calmariensis* would be worth further study. Higher individual variation, 411 or a weaker diapause response to photoperiod cues, may facilitate insect invasions to new 412 locations (Reznik et al. 2015) and may be a trait to screen in potential biological control agents.

413 Based on this study, there was sufficient genetic variation from two source populations to 414 evolve divergent photoperiod responses across a range of environmental conditions within 27 415 years. The evolution of *N. calmariensis*, as well as other biological control agents (Bean et al. 416 2012, McEvoy et al. 2012, Szűcs et al. 2012), suggests the importance of season length as a 417 selective force over ecological timescales. Screening for developmental traits or environmental 418 factors that predict successful establishment and population growth of biological control agents 419 would advance our ability to manage planned introductions (Zalucki and van Klinken 2006, 420 Abram and Moffat 2018). More targeted matching of observed photoperiod responses in the 421 native range combined with lifecycle simulations of growing season lengths in the introduced 422 range may reduce the length of time or number of introductions needed to establish robust 423 populations (Grevstad and Coop 2015, Pitcairn 2018). Beyond the field of biological control, 424 evolution to changing seasonality is the predominant genetic response to rapid climate change in 425 natural populations (Bradshaw and Holzapfel 2006) and photoperiodism will likely determine 426 range shifts across diverse taxa (Saikkonen et al. 2012). Quantifying changing photoperiod 427 responses in introduced populations provides examples of how quickly and to what extent 428 evolution can track anthropogenic warming, with the changes in both season length and timing 429 of environmental cues it brings.

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438 **References**

- 439 Abram, P. K., and C. E. Moffat. 2018. Rethinking biological control programs as planned
- 440 invasions. Curr. Opin. Insect Sci. 27: 9–15.
- 441 Bartelt, R. J., A. A. Cossé, B. W. Zilkowski, R. N. Wiedenmann, and S. Raghu. 2008. Early
- 442 summer pheromone biology of *Galerucella calmariensis* and relationship to dispersal and
- 443 colonization. Biol. Control. 46: 409–416.
- Bates, D., M. Mächler, B. Bolker, and S. Walker. 2015. Fitting linear mixed-effects models
 using lme4. J. Stat. Softw. 67: 1–48.
- 446 Bean, D. W., P. Dalin, and T. L. Dudley. 2012. Evolution of critical day length for diapause
- 447 induction enables range expansion of *Diorhabda carinulata*, a biological control agent
 448 against tamarisk (*Tamarix* spp.). Evol. Appl. 5: 511–523.
- 449 Bean, D. W., T. L. Dudley, and J. C. Keller. 2007. Seasonal timing of diapause induction
- 450 limits the effective range of *Diorhabda elongata deserticola* (Coleoptera:
- 451 Chrysomelidae) as a biological control agent for Tamarisk (*Tamarix* spp.). Environ.
- 452 Entomol. 36: 15–25.
- 453 Beck, S. D., and J. W. Apple. 1961. Effects of temperature and photoperiod on voltinism of
- 454 geographical populations of the European corn borer, *Pyrausta nubilalis*. J. Econ.
- 455 Entomol. 54: 550–558.
- 456 **Bell, G. 2010**. Fluctuating selection: the perpetual renewal of adaptation in variable
- 457 environments. Philos. Trans. R. Soc. B Biol. Sci. 365: 87–97.
- 458 **Blossey, B. 1995**. Coexistence of two leaf-beetles in the same fundamental niche, distribution,
- 459 adult phenology and oviposition. Oikos. 74: 225–234.
- 460 Bolker, B. M. 2008. Ecological models and data in R. Princeton University Press.

- 461 Bolker, B. M. 2015. Linear and generalized linear mixed models, pp. 309–333. In Fox, G.A.,
- 462 Negrete-Yankelevich, S., Sosa, V.J. (eds.), Ecol. Stat. Oxford University Press.
- 463 Bradshaw, W. E., and C. M. Holzapfel. 2006. Evolutionary response to rapid climate change.
- 464 Science. 312: 1477–1478.
- 465 Bradshaw, W. E., and C. M. Holzapfel. 2007. Evolution of animal photoperiodism. Annu. Rev.
- 466 Ecol. Evol. Syst. 38: 1–25.
- 467 Corrigan, J., D. R. Gillespie, R. De Clerck-Floate, and P. G. Mason. 2013. Lythrum salicaria
- 468 L., purple loosestrife (Lythraceae). Biol. Control Programme Can. 2001–2012. 363.
- 469 Danilevskii, A. S. 1961. Photoperiodism and seasonal development of insects. (English version
- 470 published in 1965) Oliver and Boyd, London, United Kingdom.
- 471 Denlinger, D. L., D. A. Hahn, C. Merlin, C. M. Holzapfel, and W. E. Bradshaw. 2017.
- 472 Keeping time without a spine: what can the insect clock teach us about seasonal
- 473 adaptation? Philos. Trans. R. Soc. B Biol. Sci. 372: 20160257.
- 474 Forsythe, W. C., E. J. Rykiel, R. S. Stahl, H. Wu, and R. M. Schoolfield. 1995. A model
- 475 comparison for daylength as a function of latitude and day of year. Ecol. Model. 80: 87–
 476 95.
- 477 Grevstad, F. S. 1999. Experimental invasions using biological control introductions: the
- 478 influence of release size on the chance of population establishment. Biol. Invasions. 1:
- 479 313–323.

480 Grevstad, F. S., and L. B. Coop. 2015. The consequences of photoperiodism for organisms in 481 new climates. Ecol. Appl. 25: 1506–1517.

482 Hight, S. D., B. Blossey, J. Laing, and R. Declerck-Floate. 1995. Establishment of insect

- 483 biological control agents from Europe against *Lythrum salicaria* in North America.
- 484 Environ. Entomol. 24: 967–977.

485 Hinz, H. L., M. Schwarzländer, A. Gassmann, and R. S. Bourchier. 2014. Successes we may

- 486 not have had: A retrospective analysis of selected weed biological control agents in the
- 487 United States. Invasive Plant Sci. Manag. 7: 565–579.
- 488 Izzo, V. M., J. Armstrong, D. Hawthorne, and Y. Chen. 2014. Time of the season: the effect
- 489 of host photoperiodism on diapause induction in an insect herbivore, *Leptinotarsa*
- 490 *decemlineata*. Ecol. Entomol. 39: 75–82.
- Joschinski, J., and D. Bonte. 2020. Diapause is not selected as a bet-hedging strategy in insects:
 a meta-analysis of reaction norm shapes. bioRxiv. 752881.
- 493 Kahle, D., and H. Wickham. 2013. ggmap: Spatial visualization with ggplot2. R J. 5: 144–161.
- Kawecki, T. J., and D. Ebert. 2004. Conceptual issues in local adaptation. Ecol. Lett. 7: 1225–
 1241.
- 496 Kerr, N. Z., T. Wepprich, F. S. Grevstad, E. B. Dopman, F. S. Chew, and E. E. Crone.
- 497 **2020**. Developmental trap or demographic bonanza? Opposing consequences of earlier
- 498 phenology in a changing climate for a multivoltine butterfly. Glob. Change Biol. 26:
- 499 2014–2027.
- 500 Kivelä, S. M., P. Välimäki, and K. Gotthard. 2013. Seasonality maintains alternative life501 history phenotypes. Evolution. 67: 3145–3160.
- 502 Lenth, R. 2020. emmeans: Estimated marginal means, aka least-squares means. R package
 503 version 1.4.4.

504 Levy, R. C., G. M. Kozak, C. B. Wadsworth, B. S. Coates, and E. B. Dopman. 2015.

- 505 Explaining the sawtooth: latitudinal periodicity in a circadian gene correlates with shifts
- 506 in generation number. J. Evol. Biol. 28: 40–53.
- 507 Lindestad, O., C. W. Wheat, S. Nylin, and K. Gotthard. 2019. Local adaptation of
- 508 photoperiodic plasticity maintains life cycle variation within latitudes in a butterfly.
- 509 Ecology. 100: e02550.
- 510 Manguin, S., R. White, B. Blossey, and S. D. Hight. 1993. Genetics, taxonomy, and ecology of
- 511 certain species of *Galerucella* (Coleoptera: Chrysomelidae). Ann. Entomol. Soc. Am. 86:
 512 397–410.
- 513 Masaki, S. 1961. Geographic variation of diapause in insects. Bull. Fac. Agric. Hirosaki Univ. 7:
 514 66–98.
- 515 McAvoy, T. J., and L. T. Kok. 2004. Temperature dependent development and survival of two
- 516 sympatric species, *Galerucella calmariensis* and *G. pusilla*, on purple loosestrife.
- 517 BioControl. 49: 467–480.
- 518 McEvoy, P. B., K. M. Higgs, E. M. Coombs, E. Karaçetin, and L. Ann Starcevich. 2012.
- 519 Evolving while invading: rapid adaptive evolution in juvenile development time for a
- biological control organism colonizing a high-elevation environment. Evol. Appl. 5: 524–
 536.
- 522 Menzel, R. 1979. Spectral Sensitivity and Color Vision in Invertebrates, pp. 503–580. In
- 523 Autrum, H., Bennett, M.F., Diehn, B., Hamdorf, K., Heisenberg, M., Järvilehto, M.,
- 524 Kunze, P., Menzel, R., Miller, W.H., Snyder, A.W., Stavenga, D.G., Yoshida, M.,
- 525 Autrum, H. (eds.), Comp. Physiol. Evol. Vis. Invertebr. Invertebr. Photoreceptors,
- 526 Handbook of Sensory Physiology. Springer, Berlin, Heidelberg.

- Merilä, J., and A. P. Hendry. 2014. Climate change, adaptation, and phenotypic plasticity: the
 problem and the evidence. Evol. Appl. 7: 1–14.
- 529 Pitcairn, M. J. 2018. Weed biological control in California, USA: review of the past and
- 530 prospects for the future. BioControl. 63: 349–359.
- **R Core Team**. 2019. R: A language and environment for statistical computing. R Foundation for
 Statistical Computing, Vienna, Austria.
- 533 Reznik, S. Ya., M. Yu. Dolgovskaya, A. N. Ovchinnikov, and N. A. Belyakova. 2015. Weak
- 534 photoperiodic response facilitates the biological invasion of the harlequin ladybird
- 535 *Harmonia axyridis* (Pallas) (Coleoptera: Coccinellidae). J. Appl. Entomol. 139: 241–249.

536 Roderick, G. K., R. Hufbauer, and M. Navajas. 2012. Evolution and biological control. Evol.
537 Appl. 5: 419–423.

- **Roff, D. 1980**. Optimizing development time in a seasonal environment: The "ups and downs"
 of clinal variation. Oecologia. 45: 202–208.
- 540 Saikkonen, K., K. Taulavuori, T. Hyvönen, P. E. Gundel, C. E. Hamilton, I. Vänninen, A.
- 541 Nissinen, and M. Helander. 2012. Climate change-driven species' range shifts filtered
- 542 by photoperiodism. Nat. Clim. Change. 2: 239–242.
- 543 Szűcs, M., U. Schaffner, W. J. Price, and M. Schwarzländer. 2012. Post-introduction
- 544 evolution in the biological control agent *Longitarsus jacobaeae* (Coleoptera:
- 545 Chrysomelidae). Evol. Appl. 5: 858–868.
- 546 Takeda, M., and S. Masaki. 1979. Asymmetric perception of twilight affecting diapause
- 547 induction by the fall webworm, *Hyphantria Cunea*. Entomol. Exp. Appl. 25: 317–327.
- 548 **Tauber, C. A., and M. J. Tauber**. **1981**. Insect Seasonal Cycles: Genetics and Evolution. Annu.

549 Rev. Ecol. Syst. 1: 281–308.

550	Tauber, M. J., C. A. Tauber, and S. Masaki. 1986. Seasonal adaptations of insects. Oxfor	rd
551	University Press.	

552	Thornton, P. E., S. W. Running, and M. A. White. 1997. Generating surfaces of daily
553	meteorological variables over large regions of complex terrain. J. Hydrol. 190: 214–251.
554	Thornton, P. E., M. M. Thornton, B. W. Mayer, Y. Wei, R. Devarakonda, R. S. Vose, and
555	R. B. Cook . 2017 . Daymet: Daily surface weather data on a 1-km grid for North
556	America, version 3. ORNL Distributed Active Archive Center.
557	Tobin, P. C., S. Nagarkatti, G. Loeb, and M. C. Saunders. 2008. Historical and projected
558	interactions between climate change and insect voltinism in a multivoltine species. Glob.
559	Change Biol. 14: 951–957.
560	Tsai, HY., D. R. Rubenstein, YM. Fan, TN. Yuan, BF. Chen, Y. Tang, IC. Chen, and
561	SF. Shen. 2020. Locally-adapted reproductive photoperiodism determines population
562	vulnerability to climate change in burying beetles. Nat. Commun. 11: 1398.
563	Urbanski, J., M. Mogi, D. O'Donnell, M. DeCotiis, T. Toma, and P. Armbruster. 2012.
564	Rapid adaptive evolution of photoperiodic response during invasion and range expansion
565	across a climatic gradient. Am. Nat. 179: 490–500.
566	Van Dyck, H., D. Bonte, R. Puls, K. Gotthard, and D. Maes. 2015. The lost generation
567	hypothesis: could climate change drive ectotherms into a developmental trap? Oikos.
568	124: 54–61.
569	Velarde, R. A., R. N. Wiedenmann, and D. J. Voegtlin. 2002. Influence of photoperiod on the
570	overwintering induction of Galerucella calmariensis L. BioControl. 47: 587-601.

- 571 Venables, W. N., and B. D. Ripley. 2002. Modern applied statistics with S, Fourth. ed.
- 572 Springer, New York.

573 Wright, M. G., and G. M. Bennett. 2018. Evolution of biological control agents following

- 574 introduction to new environments. BioControl. 63: 105–116.
- 575 Zalucki, M. P., and R. D. van Klinken. 2006. Predicting population dynamics of weed
- 576 biological control agents: science or gazing into crystal balls? Aust. J. Entomol. 45: 331–
- 577 344.
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580 Tables

581 Table 1: Description of sites in the western USA with introduced *Neogalerucella calmariensis* populations.

Site	Elevation (meters above sea level)	Max day length (hours) ^a	Avg annual degree- days (SD) ^b	Average F1 adult emergence prediction ^c	Population history (year introduced) ^d	Voltinism in the field ^e
Bellingham, Washington (48.76°, -122.48°)	17	16.3	1067 (122)	July 20	Introduced from Ephrata, WA (unknown year in 1990s)	1 generation
Ephrata, Washington (47.16°, -119.66°)	350	16.1	1677 (127)	June 24	Original release from north and south German sources (1992-93)	3 generations
Yakima Training Center, Washington (46.76°, -119.99°)	204	16.0	1809 (131)	June 19	Introduced from Ephrata, WA (unknown year in 1990s).	Not observed, assumed similar to nearby Ephrata, WA
Rickreall, Oregon (44.98°, -123.27°)	64	15.8	1366 (114)	July 9	Original release from south German source (1992)	1 generation
Sutherlin, Oregon (43.39°, -123.32°)	153	15.6	1601 (143)	June 27	Original release from south German source (1992)	2 generations
McArthur, California (41.10°, -121.41°)	1006	15.3	1559 (122)	July 1	Introduced from Rickreall, OR and vicinity (1998, 2000) & Ephrata, WA (2002)	2 generations
Palermo, California (39.41°, -121.58°)	35	15.1	2769 (175)	May 13	Introduced from Rickreall, OR and vicinity (1998-2001) & McArthur, CA (2006). Release from Ephrata, WA 10km northwest of this site (2004-05) could have dispersed here.	3 generations

582 ^{*a*}Maximum day length includes twilight until the sun is 1.5° below the horizon (using method of Forsythe et al. 1995).

 b Degree-days are calculated with the triangle method with 10°C/30°C thresholds over 25 years (1994-2018) from Daymet daily gridded interpolations of weather station observations.

585 Predicted sensitive stage emergence is based on 25 years (1994-2018) of results from a degree-day lifecycle model (Grevstad and Coop 2015).

⁴History is from Hight et al. (1995) for original releases in Washington and Oregon and from M. Pitcairn (personal communication) for transfers to California

587 ^eF. Grevstad and M. Pitcairn surveyed sites in 2014 to coincide with the expected timing of larval emergence in later generations.

		20	14 Photop	eriod respons	e		20	19 Photop	eriod response	9
Predictors	Beta	Std. error	Р	CP ^a (95% CI)	Population differences ^b	Beta	Std. error	Р	CP (95% CI)	Population differences
Photoperiod: light hours	4.05	0.67	<0.001	-	-	1.29	0.16	<0.001	-	-
Bellingham, Washington	-2.20	0.58	< 0.001	16.1 (15.9-16.4)	А	-1.36	0.40	< 0.001	16.6 (16.0-17.1)	А
Ephrata, Washington	-2.43	0.59	< 0.001	16.2 (16.0-16.4)	А	-	-	-	-	-
Yakima TC, Washington	-	-	-	-	-	-0.29	0.41	0.488	15.9 (15.5-16.4)	AB
Rickreall, Oregon	-	-	-	-	-	-1.76	0.41	< 0.001	16.8 (16.3-17.3)	А
Sutherlin, Oregon	0.40	0.52	0.443	15.6 (15.4-15.8)	В	1.84	0.40	< 0.001	14.5 (14.0-15.1)	С
McArthur, California	_	-	-	-	-	-0.34	0.37	0.365	15.9 (15.5-16.4)	AB
Palermo, California	0.86	0.49	0.080	15.5 (15.3-15.7)	В	0.95	0.38	0.013	15.1 (14.6-15.6)	BC

589 Table 2: Photoperiod responses of *Neogalerucella calmariensis* populations in two experiments at constant 23°C temperatures.

590 The main effects of photoperiod treatment and population on the proportion reproductive come from two generalized linear mixed models (logistic regression), one for each year's experiment. Figures 2 and 3 plot these model results.

⁴Critical photoperiod (CP), or the light hours at which 50% would enter diapause, was estimated for each population and year with 95%

593 confidence intervals by inverse predictions from the models.

^bPopulations with significant differences have different letters, which come from pairwise comparison tests with Tukey's method to adjust *P*

595 values.

596	Table 3: ANOVA for tests	of photoperiod and	population effects on	diapause proportion ar	d development time.
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	20)14 Diapa	use	20	19 Diapat	ise	2018 De	evelopmer	nt time
							Chi		
Fixed effects	Chi square	df	Р	Chi square	df	Р	square	df	Р
Photoperiod (light hours) ^a	36.6	1	< 0.001	82.7	1	< 0.001	0.0396	1	0.842
Population	25.8	3	< 0.001	68.7	5	< 0.001	14.8	5	0.0113
Photoperiod x Population	1.41	3	0.703	6.53	5	0.258	9.62	5	0.0869
Random effects	Groups	SD		Groups	SD		Groups	SD	
Intercept ^b	16	0.602		36	0.708		71	0.0135	

597 Type II sum of squares (Wald chi square tests) from generalized linear mixed models.

⁶598 ^aPhotoperiod is modeled as a continuous variable and had different growth chamber treatments in different experiments.

⁵⁹⁹ ^bRandom intercepts in were based on treatment group for diapause models and rearing dish for development time model. Standard deviation (SD)

600 shows between-group variation.

601

	Тур	e II ANOV	VA	Co	pefficients (reduced mod	lel)
Fixed effects	Chi square	df	Р	Beta	Std. error	Р
Intercept	-	-	-	-1.06	0.482	0.0282
Photoperiod (light hours) ^{a}	78.9	1	< 0.001	4.28	0.629	< 0.001
Latitude	5.70	1	0.017	-0.918	0.393	0.0196
Year indicator (2019)	4.09	1	0.043	0.747	0.414	0.0709
Photoperiod x Latitude	0.710	1	0.400	-	-	-
Photoperiod x Year	21.8	1	< 0.001	-2.97	0.643	< 0.001
Latitude x Year	1.40	1	0.326	-	-	-
Photoperiod x Latitude x Year	0.113	1	0.737	-	-	-
Random effects	Groups	SD		Groups	SD	
Population (intercept)	7	0.907		7	0.875	
Treatment group (intercept)	52	0.845		52	0.930	

Table 4: Test of latitudinal gradient in photoperiod effects on diapause proportion over two experiments.

604

Type II sum of squares (Wald chi square tests) from generalized linear mixed models predicting the proportion reproductive across both the 2014

606 and 2019 diapause experiments.

607 Figure captions

608 Figure 1: Neogalerucella calmariensis populations in the northwest USA in this study. Map

- tiles by Stamen Design (CC BY 3.0) with data by OpenStreetMap (ODbL) are plotted with the
- 610 ggmap R package (Kahle and Wickham 2013).

611 Figure 2: Photoperiod responses diverged between northern and southern populations in

612 **2014.** Each data point shows the proportion of female beetles that were reproductive for each

613 treatment combination, with size scaled by number of beetles, for four populations (panels) and

four photoperiod treatments (x-axis). Solid lines and 95% confidence intervals show generalized

615 linear mixed model predictions. The critical photoperiods, at which 50% choose to diapause, are616 labeled below the dotted lines.

617 **Figure 3: Photoperiod responses diverged between populations in 2019.** Each data point

618 shows the proportion of male and female beetles that were reproductive for each treatment

619 combination, with size scaled by number of beetles, for six populations (panels) and six

620 photoperiod treatments (x-axis). Solid lines and 95% confidence intervals show generalized

621 linear mixed model predictions. The critical photoperiods, at which 50% choose to diapause, are622 labeled below the dotted lines.

623 Figure 4: Latitudinal clines in critical photoperiod with estimated population traits. Lines

shows the modeled change in critical photoperiod with latitude, estimated for the 2014 (solid)

and 2019 (dashed) experiments. Confidence intervals (95%) for each line are shaded gray. Note

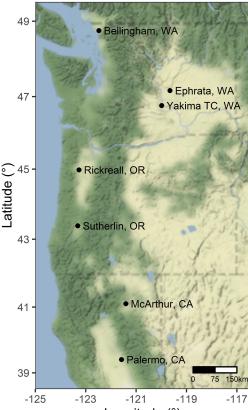
- 626 that the lines do not significantly differ in slope (i.e. no latitude x year interaction, Table 4).
- Triangles show the critical photoperiods and 95% confidence intervals estimated for populations

628 in separate models (corresponding to Figures 2 and 3). The x-axis values for the three sites with

629 repeat measures were offset by 0.1° to avoid overlap.

630 Figure 5: Main effects of population and photoperiod on development time at constant

- 631 **21°C.** Points represent daily observations of the proportion of the total number of beetles that
- 632 eclosed in each group. Development time is similar for most populations, but the Rickreall, OR
- 633 population took longer than the populations in Sutherlin, OR and Yakima, WA (A). Photoperiod
- 634 treatments had no effect on development time in the four growth chambers (B).



Longitude (°)

