

1 What makes *Hemidactylus* invasions successful? A case study on the island of Curaçao.

2

3 April D. Lamb^{*1,2}, Catherine A. Lippi³, Gregory J. Watkins-Colwell⁴, Andrew Jones⁵, Dan Warren⁶,
4 Teresa L. Iglesias⁷, Matt Brandley⁸, Connor Neagle^{1,9} and Alex Dornburg¹

5

6 ¹ North Carolina Museum of Natural Sciences, Raleigh, NC 27601

7 ² North Carolina State University, Department of Applied Ecology, Raleigh, NC 27695

8 ³ Quantitative Disease Ecology and Conservation (QDEC) Lab Group, Department of Geography,
9 University of Florida, Gainesville, FL 32611

10 ⁴ Division of Vertebrate Zoology, Yale Peabody Museum of Natural History, New Haven, CT 06520
11 USA

12 ⁵ Department of Ecology and Evolutionary Biology, Yale University, New Haven, CT 06520 USA

13 ⁶ Senckenberg Biodiversität und Klima - Forschungszentrum, Frankfurt am Main 60325, Germany

14 ⁷ Okinawa Institute of Science and Technology Graduate University, Okinawa Prefecture 904-0495,
15 Japan

16 ⁸ School of Biological Sciences, University of Sydney, NSW 2006, Australia

17 ^{1,9} North Carolina State University, Department of Forestry and Environmental Resources, Raleigh, NC
18 27695

19

20

21 *Corresponding author: adlamb@ncsu.edu. ORCID: 0000-0002-1794-2323

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

41

Authors' Contributions

AD, AJ, AL, DW, GWC, MCB, and TI conceived of the project. AD, AJ, DW, MCB, and TI conducted fieldwork. CL analyzed stomach contents. AD and MCB collected morphometric data. AL prepared isotopic samples. AD, AJ, AL, DW, and CN performed analyses. AD, AL, and CN wrote the initial draft of the manuscript. All authors contributed to subsequent revisions.

42
43

44 **Abstract**

45 *Hemidactylus* spp. (House geckos) rank among the most successful invasive reptile species worldwide.
46 *Hemidactylus mabouia* in particular has become ubiquitous across tropical urban settings in the Western
47 Hemisphere. *H. mabouia*'s ability to thrive in close proximity to humans has led to the rapid displacement
48 of native geckos in urban areas, however the mechanisms driving this displacement remain understudied.
49 Here we combine data from nitrogen and carbon stable isotopes, stomach contents, and morphometric
50 analyses of traits associated with feeding and locomotion to test alternate hypotheses of displacement
51 between *H. mabouia* and a native gecko, *Phyllodactylus martini*, on the island of Curaçao. Consistent
52 with expectations of direct food resource competition, we demonstrate substantial overlap of invertebrate
53 prey resources between the species. Additionally, we found strong evidence from both diet content and
54 stable isotope analyses that *H. mabouia* acts as a vertebrate predator, preying upon *P. martini* as well as
55 other native and non-native reptiles. Finally, we show that *H. mabouia* possesses several morphological
56 advantages, including larger sizes in feeding-associated traits and limb proportions that could offer a
57 propulsive locomotor advantage on vertical surfaces. Together, these findings suggest the successful
58 establishment of *H. mabouia* likely involves a combination of both exploitative interspecific competition
59 and predation. Given the ubiquity of *H. mabouia*, illuminating the role of this species as both a competitor
60 and a predator casts new concerns on the ecological and demographic impacts of this widespread urban
61 invader.

62

63 **Keywords:** invasive species, urbanization, vertebrate biodiversity loss, food web, trophic ecology

64

65 **Introduction**

66 Since the onset of the industrial revolution, the impact of invasive species on endemic fauna and
67 flora has been a central topic in the management and conservation of biodiversity worldwide (Paini et al.
68 2016; Young et al. 2017; Shechonge et al. 2019). This concern reflects dramatic losses in global
69 biodiversity and an increasing shift towards widespread homogenization of the planet's biota (McKinney
70 and Lockwood 1999; McKinney 2006; Trentanovi et al. 2013). These trends are especially acute in
71 urbanizing landscapes, which have repeatedly been shown to support higher numbers of non-native,
72 human-commensal species (Useni Sikuzani et al. 2018), such as cats (Buzan 2017; Bateman and Fleming
73 2012), rats (Bateman and Fleming 2012; Buzan 2017), and house sparrows (González-Oreja et al. 2018).
74 Following establishment, successful non-native species have been found to restructure resident
75 community assemblages by directly or indirectly altering top-down processes (e.g. predation, (Willson
76 2017; Pedersen et al. 2018)), bottom-up processes (e.g. resource availability (Yam et al. 2016)), or both
77 (i.e. "middle-out" effects, (Weber and Brown 2009)) at the expense of native taxa. In the most extreme
78 cases this can result in the extirpation or extinction of native species (Wiles et al. 2003; Toussaint et al.
79 2016; Liu et al. 2017). However, investigations into the impact and distribution of introduced species
80 have been largely restricted to species that are easily visible in the landscape (Beasley et al. 2018), are a
81 direct nuisance to humans (Bithas et al. 2018), or displace commercially important or game species
82 (Galanidi et al. 2018; Hill et al. 2004). While not misguided, this bias has left a critical gap in our
83 knowledge regarding the potential impacts of less readily observable, but equally common, non-native
84 human-commensal taxa (Morais and Reichard 2018).

85 Despite the prevalence of invasive reptiles around the world (Kraus 2015), most attention has
86 been devoted to the loss of biodiversity following the spread of a few larger bodied species such as
87 Burmese pythons (Smith et al. 2016; Willson 2017), green iguanas (Falcón et al. 2013; Burgos-Rodríguez
88 et al. 2016), and brown tree snakes (Wiles et al. 2003; Rodda and Savidge 2007; Richmond et al. 2015).
89 However, numerous smaller and more clandestine reptiles have also become globally pervasive (Kraus
90 2015; Capinha et al. 2017; Lapiedra et al. 2017). These invasions, while common, often go unnoticed
91 until native reptiles begin to disappear from the landscape (Kraus 2015). Such cryptic losses in
92 biodiversity are a hallmark of introduction of *Hemidactylus* spp. (House Geckos), a group commonly
93 associated with urbanized and developing areas. Over the past century, *Hemidactylus* spp. have become
94 an established feature of tropical and subtropical landscapes around the world (Carranza and Arnold
95 2006). Following establishment, these geckos have been repeatedly linked to local extirpation and even
96 extinction of native lizards (Petren and Case 1996; Cole et al. 2005; Hoskin 2011). One species in
97 particular, *Hemidactylus mabouia* (Tropical House Gecko), is perhaps the most pervasive and formidable
98 gecko to invade the Western Hemisphere (Weterings and Vetter 2018).

99 Native to Africa, *Hemidactylus mabouia* is now common throughout the Americas and Caribbean
100 (Carranza and Arnold 2006). Recent studies have linked the successful establishment of this species in
101 urban and suburban environments to its ability to capitalize on the aggregation of insects around human
102 light sources (Hughes et al. 2015). Restricting the access of native geckos to these clustered food
103 resources is thought to represent a competitive advantage for *H. mabouia* that promotes high densities of
104 individuals (van Buurt 2004; Short and Petren 2011; Williams et al. 2016). As *H. mabouia* adult males
105 are noted for being particularly aggressive (Short and Petren 2011), the ability of this species to
106 aggressively restrict access to spatially clustered food resources suggests interference competition,
107 whereby high densities of aggressive competitors fuel the displacement of native gecko species. However,
108 alternate hypotheses remain untested.

109 In addition to their impact as competitors, two aspects of *H. mabouia* invasions that have received
110 particularly little attention are locomotor morphology and role of *H. mabouia* as potential predators. The
111 feeding mode of *H. mabouia* combines ambush tactics (Vitt 1983) with active pursuit of nearby prey
112 (Dornburg et al. 2016). Such a foraging mode could have selected for limb proportions that offer a
113 mechanical advantage on sheer vertical surfaces (Zaaf and Van Damme 2001). Further, it is possible that
114 the generally robust body plan of *Hemidactylus* spp. facilitates the capture of larger prey not accessible to
115 other similarly sized geckos, although this hypothesis has not been tested. An alternative, but not
116 mutually exclusive, explanation for the success of *H. mabouia* comes from isolated natural history reports
117 of *H. mabouia* preying on other species of geckos (Dornburg et al. 2011, 2016) as well as cannibalizing
118 conspecifics (Bonfiglio et al. 2006). Both morphological advantages and predation have been invoked as
119 major drivers of displacement in the wake of invasions by the closely related *Hemidactylus frenatus* in the
120 Pacific (Petren and Case 1998; Petren and Case 1996; Bolger and Case 1992; Case et al. 1994; Short and
121 Petren 2012). However, both the role of morphological advantages and predation in driving the decline of
122 native gecko populations in the wake of an *H. mabouia* invasion remain unclear.

123 In this study we assess whether there is evidence for trait driven advantages or predation in the
124 invasion of *H. mabouia* in the Lesser Antillean island of Curaçao. We specifically focus on competition
125 between *H. mabouia* and the native *Phyllodactylus martini* (Dutch Leaf Tailed Gecko), as *P. martini*
126 declines have historically been linked to the invasion of *H. mabouia* (van Buurt 2004; van Buurt 2006;
127 Hughes et al. 2015; Dornburg et al. 2016). First we integrate analyses of nitrogen and carbon stable
128 isotopes with direct examination of stomach contents to test for levels of prey overlap and isotopic trophic

129 signatures consistent with hypotheses of resource competition or predation. We additionally collected
130 morphometric measurements from traits associated with feeding and locomotion to test the hypothesis
131 that *H. mabouia* possess trait advantages over its hypothesized native competitor.

132

133 **Materials/Methods**

134 *Fieldwork and Data Acquisition*

135 *Hemidactylus mabouia* (n=90) and *Phyllodactylus martini* (n=71) specimens were collected at six
136 sites across Curaçao between July 2009 and September 2011: Lagun, Westpunt, CARMABI, Shete Boca,
137 Saint Anna Bay, and Willemstadt (Figure 1; Supplemental materials). Habitat type and species occupancy
138 vary across sampling locations. For example, both species co-occur in Lagun and Westpunt. At these sites
139 we restricted our sampling to suburban areas near natural habitats to maximize the potential of both
140 species co-occurring as *P. martini* has been found to be absent far from edge habitats in the presence of
141 *H. mabouia* (Hughes et al. 2015). In contrast, Shete Boca is a natural area in which *H. mabouia* are
142 absent, while Saint Anna Bay and Willemstadt are urban areas in which *P. martini* are absent. This
143 sampling design allowed us to capture a greater degree of diet breadth of each species across the island.
144 Across sites, sample locations included walls, rocks, outcrops, trees, thatch roofs, open ground, and
145 shrubbery. At no point during sampling did we document individuals of both species occupying the same
146 structure (e.g., same wall or tree), and individuals were collected opportunistically at each site. Prior to
147 preservation, muscle biopsies were taken from each individual and dehydrated for analysis of stable
148 isotopes. Additionally, leaf samples from each locality and temporal sampling event were collected and
149 dehydrated for use as baselines in isotopic analyses. Specimens then were fixed in 10% formalin and later
150 transferred to 70% ethanol and deposited in the Yale Peabody Museum of Natural History (supplemental
151 materials).

152 Specimens collected in 2011 (n=59) had their stomach contents preserved in 10% formalin and
153 dissected, with contents identified and enumerated under a dissecting MVX10 microscope (Olympus
154 Corp.; <http://www.olympus-lifescience.com/>). Prey items were identified to the taxonomic groupings
155 similar to those in other studies of Caribbean lizards (e.g., (Perry 1996)): Arachnida (scorpiones),
156 Arachnida (Araneae), Blattaria (Blattodea), Chilopoda, Coleoptera, Diptera, Ephemeroptera, Hemiptera,
157 Hymenoptera, Isopoda, Lepidoptera, Orthoptera, and “other”. Any vertebrate remains encountered were
158 additionally identified to the highest taxonomic resolution possible, and we additionally identified any
159 parasites encountered in the stomach. As formalin and alcohol preservation can have heterogeneous
160 effects on the volume of invertebrate organisms (Donald and Paterson 1977), enumeration of diet contents
161 was restricted to % frequency.

162 We further collected measurement data on 79 *Hemidactylus mabouia* for 10 morphological traits
163 associated with feeding and locomotion: snout-vent length (SVL), postorbital width, temporalis width,
164 head length, jaw length, head height, humerus length, radius length, femur length and tibia length. All
165 measurements were taken to the nearest 0.01 mm using digital calipers (Fowler Promax). Both stomach
166 content and morphological data were integrated with the dataset of Dornburg et al. (2016) who previously
167 measured *Phyllodactylus martini* specimens for the same morphological traits (n=34; Zenodo DOI:
168 10.5281/zenodo.61569) and prey items (n=69; Zenodo DOI: 10.5281/zenodo.61569).

169

170 *Stable isotopic analysis of trophic ecology*

171 Leg muscle biopsies from 21 individual *Hemidactylus mabouia* and 17 *Phyllodactylus martini*
172 legs as well as 8 plant stems and leaf baseline samples were used in nitrogen ($\delta^{15}\text{N}$) and carbon ($\delta^{13}\text{C}$)

173 stable isotope analysis. Skin was removed from each muscle biopsy, and individual muscle and plant
174 baseline samples were dehydrated at 40°C degrees for 48 hours. Following dehydration, samples were
175 powdered using a bead beater (MP FastPrep24 Hyland Scientific). From each sample, 1.5 mg of powder
176 was loaded into 3x5 mm tins. Samples were analyzed at the University of California Davis Stable Isotope
177 Facility using an isotope ratio mass spectrometer. As nitrogen enrichment can vary over spatial or
178 temporal periods, quantification of trophic position for each individual was standardized using primary
179 producer baseline samples from plant leaves and stems collected at each locality (Vidal and Sabat 2010;
180 Roches et al. 2016). To account for $\delta^{15}\text{N}$ values not reflecting primary producer level values (Marshall et
181 al. 2007), baseline samples were compared across sites with aberrant samples (i.e., primary producer
182 $\delta^{15}\text{N} >$ consumer $\delta^{15}\text{N}$) removed. Nitrogen values were standardized following Post (2002), in
183 subtracting the mean $\delta^{15}\text{N}$ of the primary producers from $\delta^{15}\text{N}$ of each individual lizard and assuming
184 fractionation of 3.4% per trophic level (Post 2002). $\delta^{15}\text{N}$ values for each species were visualized using
185 violin plots which allow for simultaneous inspection of quartiles and the underlying probability
186 distribution through integration of a rotated kernel density plot with a boxplot (Hintze and Nelson 1998).
187 We tested for differences between the mean $\delta^{15}\text{N}$ values of *H. mabouia* and *P. martini* using a Welch's t-
188 test and additionally used Levene's test to assess whether there was a significant difference in $\delta^{15}\text{N}$
189 variance between species. A significant positive difference in $\delta^{15}\text{N}$ between *H. mabouia* and *P. martini*
190 would be consistent with the hypothesis that individual *H. mabouia* are vertebrate predators. To test for
191 potential differences in $\delta^{13}\text{C}$, we used the same statistical approaches as those used in the analysis of
192 $\delta^{15}\text{N}$, assuming carbon fractionation to be 0% (Post 2002). In this case, non-significant differences in
193 $\delta^{13}\text{C}$ would support the expectation that these species forage in similar habitats. All analyses were
194 conducted in R, v. 3.4.3 (R Development Core Team 2018).

195

196 *Stomach Content Analysis*

197 For each species, relative frequency of each prey item was calculated based on the total number
198 of prey items encountered across all individuals of that species. Differences in stomach contents between
199 species were visualized using a non-metric multidimensional scaling (NMDS) analysis in the vegan
200 software package (Oksanen et al. 2007; Oksanen 2011). Stress values were quantified to test if NMDS
201 ordination represents a viable indicator of species dissimilarity, with stress values less than 0.1 indicating
202 good representation of the dissimilarities (Clarke 1993). An analysis of similarity (ANOSIM; (Clarke
203 1993); (Chapman and Underwood 1999)) was used to additionally test for significant differences between
204 species, using Manhattan distances and 999 permutations in the vegan software package (Oksanen et al.
205 2007; Oksanen 2011). Differences in mean ranks were quantified using the R statistic, with values close
206 to zero indicating high similarity and values close to one indicating high dissimilarity (Chapman and
207 Underwood 1999). As *P. martini* has been found to consume different prey in rural versus suburban edge
208 habitats (Dornburg et al. 2016), NMDS and ANOSIM analyses were repeated with analyses restricted to
209 individual geckos collected in areas where both species co-occur. This allowed us to test whether pooling
210 across habitat types potentially masked differences or overlap in prey items. Additionally, differences in
211 parasite load between species were compared using a Welch's t-test.

212

213 *Comparisons of morphology*

214 We compared absolute differences in log snout-vent length (SVL) between species using an
215 ANOVA and created raincloud plots (Allen et al. 2019) to visualize differences. These plots combine
216 classic boxplots with violin raw data plots to simultaneously visualize data, the difference in size

217 quartiles, and a kernel density estimate of the probability density of the SVL data. We conducted a
218 principal components analysis (PCA) to visualize the overall morphospace occupied by both species. In
219 geckos, size has been shown to covary with our target morphological measurements (Dornburg et al.
220 2016). As such, we first regressed all the measurements per species against SVL (supplemental materials)
221 and used the residual values of individual traits regressed against SVL as data for the PCA. To assess if
222 differences in morphospace occupancy were mostly driven by uneven sample sizes, we randomly sampled
223 equal numbers of both gecko species from our data 200 times in intervals of 5 between 10 and 55. For
224 each of these 2000 datasets, we conducted a PCA and computed the mean and quantiles (25% & 75%) of
225 the ratio of *H. mabouia* to *P. martini* morphospace.

226 While morphospace visualization is advantageous for assessing the overall overlap of phenotypic
227 variation, it is possible that allometric slopes are identical between species and simply have different
228 intercepts (i.e., at a given body size a focal trait in one species is larger in one species than the other). To
229 further scrutinize our data, we used an analysis of covariance (ANCOVA) to test for differences in each
230 morphological trait between species. For each analysis, we kept log transformed SVL as the covariate and
231 treated each log transformed morphometric measurement (e.g., jaw length, limb length, etc) as the
232 response. This approach allowed us to test the potential correlation for each measured trait and SVL as
233 well as the possibility of significant differences between species that take trait covariation with SVL into
234 account. We repeated analyses with non-significant interactions removed, as inclusion or omission of
235 non-significant interactions can potentially impact ANCOVA analyses.

236 In many lizard species, including geckos, head size is a sexually dimorphic trait with males often
237 having larger heads relative to females (Kratochvíl et al. 2002; Scharf and Mieri 2013; Iturriaga and
238 Marrero 2013). As such, we used an ANCOVA to assess whether morphological differences for each trait
239 were potentially masked when pooling sexes by species. For all analyses, we again kept log transformed
240 SVL as the covariate and treated each log transformed morphometric measurement as the response.
241 Finally, we assessed potential differences in total limb lengths (humerus length + radius length; femur
242 length + tibia length) between species and sexes using log transformed limb length as the response and
243 log transformed SVL as the covariate in an ANCOVA. This additional analysis facilitated additional
244 comparisons of expectations of gecko locomotion as studies often discuss differences in total limb
245 lengths.

246 Prior work has suggested large hind limbs compensate for large heads in the locomotion of
247 *Hemidactylus* spp. geckos (Cameron et al. 2013). As such, we examined scaling relationships between
248 head size and hindlimb length for both species by constructing a set of generalized linear models (GLMs).
249 We built models using sex, species, SVL, and head size as explanatory variables, with one set of models
250 using head length to quantify head size and another set using post-orbital width. All models except the
251 intercept-only null models contained an interaction term between SVL and the head size term, so that the
252 effects of head size on limb length would be controlled for overall body size. Additional candidate models
253 included (1) sex, (2) species identity, and (3) sex, species identity, and an interaction term between
254 species identity and head size. Model fit was evaluated with the Akaike information criterion with a
255 correction for small sample size (AICc). This method of model selection identifies models that predict the
256 data well while penalizing overparameterization (Burnham and Anderson 2004).

257

258 **Results**

259 *Differences in feeding ecology*

260 Analysis of $\delta^{15}\text{N}$ revealed a significant (Welch's t-test: $p < 0.004$; $t = 3.123$ $df = 34.272$) shift
261 towards a higher mean trophic level in *H. mabouia* versus *P. martini* (Figure 2A). In contrast, analysis of
262 $\delta^{13}\text{C}$ isotopes revealed no significant (Welch's t-test: $p < 0.401$; $t = 0.857$ $df = 21.812$) shift in mean $\delta^{13}\text{C}$
263 between *H. mabouia* versus *P. martini* (Figure 2B), supporting the expectation that these two species
264 overlap in major foraging habitat type. Levene's tests did not support a significant increase in $\delta^{15}\text{N}$
265 variance within *H. mabouia* versus *P. martini* ($F = 0.480$; $p = 0.493$), though a single *P. martini* outlier
266 point in our analysis depicted a carbon signature consistent with marine prey resource use, suggesting the
267 possibility that some individuals may opportunistically forage close to the shoreline. Regardless, the
268 difference in variance of $\delta^{13}\text{C}$ values between species was found to be non-significant ($F = 0.585$;
269 $p = 0.449$), even after removing this potential outlier point ($F = 7.624$; $p = 0.155$).

270 Our analyses of individual stomach contents revealed *H. mabouia* to generally have fewer prey
271 items per stomach than *P. martini* (Welch's t-test: $p < 0.001$; $t = 3.31$ $df = 84.74$). Across 59 specimens of
272 *H. mabouia*, we found 0 to 3 prey items per individual, which spanned a wide range of invertebrates
273 (Table 1 & Figure 3). Additionally, three individual *H. mabouia* each contained a single vertebrate prey
274 item. These prey items were identified as *Gonatodes antillensis*, *Phyllodactylus martini*, and
275 *Ramphotyphlops braminus*. Comparing the invertebrate prey found in *H. mabouia* to *P. martini* revealed
276 the two species to consume similar prey items with differences in the overall percentages of prey items
277 consumed. We find both species to generally consume the same major invertebrate prey groups but at
278 different frequencies: Arachnida (*H. mabouia* = 14%; *P. martini* = 19%); Insecta (*H. mabouia* = 50%; *P.*
279 *martini* = 58%); Isopoda (*H. mabouia* = 28%; *P. martini* = 9%). Further, the species varied with regard to
280 individual prey items and frequency within these major groupings (Figure 3 & Table 1) and ANOSIM
281 results supported significant differences between groups ($R = 0.213$, $p = 0.001$). Visualizations of diet data
282 based on NMDS analyses of invertebrate prey items for the two species support a large degree of overlap
283 in diet with *H. mabouia* utilizing the same resources as *P. martini*, but with *H. mabouia* also utilizing
284 more resources not exploited by *P. martini* (Figure 4A). Repeating analyses for just individuals residing
285 in areas of co-occurrence again supported significant differences between species in an ANOSIM analysis
286 when vertebrates were included as a prey category ($R = 0.020$, $p = 0.040$). All instances of vertebrate
287 predation by *H. mabouia* were found in areas where the two species overlap (supplemental materials).
288 Restricting an ANOSIM analyses to just invertebrate prey items supported no significant differences in
289 diet between the two species ($R = 0.018$, $p = 0.070$). Visualizations of both the raw (supplemental materials)
290 and NMDS analyses of the invertebrate prey item diet data for the two species areas further depicted a
291 large degree of overlap (Figure 4B). In addition to prey contents, parasitism infestations by nematodes
292 were significantly different between the two species (Welch's t-test: $p < 0.001$; $t = -3.768$ $df = 71$),
293 suggesting higher parasite pressure within *P. martini* (Table 1).

294 295 *Differences in morphology*

296 We found a significant overall size difference between *H. mabouia* versus *P. martini* ($F = 10.61$; p
297 $= 0.00143$), with *H. mabouia* generally being larger (Figure 5A). Three axes of a principal components
298 analysis (PCA) of morphological traits collectively capture 64.1% of the measured variation (PC1:
299 34.84%; PC2: 16.90%; PC3: 12.37%). PC1 largely captures differences in limb lengths (~39% total
300 hindlimb, 18% total front limb) and variation in the postorbital width (~24%). In contrast, PC2 mostly
301 captures variation in cranial measurements with over 70% of the loadings belonging to a combination of
302 head length (~29%), jaw length (~17%), temporalis width (~13%), and postorbital width (~13%). PC3
303 largely captured further variation in cranial morphology (Supplemental materials). Visualization of these

304 PC axes revealed a high degree of overlap between species, with *H. mabouia* occupying more
305 morphospace overall. Between PC1 and PC2 (Figure 5B) the total morphospace occupancy based on the
306 convex hull area [CHA] of *H. mabouia* was 64% larger (*H. mabouia* CHA = 18.370; *P. martini* =
307 11.140). Similarly, between PC1 & PC3 (*H. mabouia* CHA = 22.724; *P. martini* = 5.898; Figure 5C) and
308 PC2 & PC3 (*H. mabouia* CHA = 16.532; *P. martini* = 5.474; Figure 5D) the CHAs of *H. mabouia* were
309 larger. Results of our dataset resampling analyses support that these differences were not due to sample
310 size differences alone (Supplemental materials). SVL was significantly correlated with all measured
311 morphological traits (Table 2; Supplemental materials) and ANCOVA results further support significant
312 differences between residual trait variation after accounting for SVL scaling between species for all traits
313 (Table 2; Supplemental materials). The only exception to this general trend of a significant relationship
314 between species identity and trait was head height ($F= 3.232$; $p= 0.075$). These results were consistent
315 whether non-significant interactions were included in the analysis or not (Supplemental materials). Tests
316 for sexual dimorphism for no evidence for trait differences between male and female *P. martini*. In
317 contrast, head width was significantly different between male and female *H. mabouia*, suggesting *H.*
318 *mabouia* males have wider heads than females (Supplemental materials).

319 GLM analyses of the relationship between head size and hind limb length reveal largely
320 concordant patterns regardless of which metric (head length or post-orbital width) is used to quantify head
321 size (Table 3; supplemental materials). For both measurements, the top model (lowest AICc score) was
322 the one containing a different intercept of the relationship between head size and limb length for the two
323 species, but without a difference in slope (i.e., no interaction between species identity and the head
324 size/SVL relationship). These top models also include no effect of sex on the relationship between head
325 size and limb length, but in both cases the model that did include sex was also within or nearly within the
326 set of credible models (deltaAIC of 1.52 for head length, and deltaAIC of 2.2 for post-orbital width).

327

328 Discussion

329 *Hemidactylus mabouia* ranks among the most pervasive invasive lizard species in the neotropics
330 (Rödder et al. 2008; Weterings and Vetter 2018). This species has repeatedly been hypothesized to
331 represent a superior competitor that restricts access to food resources (Rocha et al. 2011; Hughes et al.
332 2015; Williams et al. 2016) and thereby promotes the extirpation of native, as well as non-native, geckos
333 (van Buurt 2004; van Buurt 2006). Our study provides support for this hypothesis, showing that on
334 Curaçao, *H. mabouia* not only competes with the native gecko *Phyllodactylus martini* for prey resources
335 but preys upon this and other vertebrate species. Notably, both stable isotopic and stomach contents
336 demonstrate that *H. mabouia* will readily consume vertebrate prey items that include *P. martini*,
337 *Gonatodes antillensis* (the Venezuelan Coastal Clawed Gecko), and the non-native blind snake
338 *Ramphotyphlops braminus*. Additionally, we demonstrate larger sizes in feeding associated traits and
339 limb lengths that may be advantageous for *H. mabouia* during the rapid forward propulsive locomotion
340 associated with ambush predation. Given the ubiquity of *H. mabouia* throughout the neotropics, our
341 results provide a new perspective for understanding the complexity of *Hemidactylus* spp. invasions,
342 suggesting their potential impact to be vastly underestimated.

343

344 *On the competitive advantages of Hemidactylus*

345 Prior work has suggested that *H. mabouia* directly competes with *Phyllodactylus martini* for food
346 resources (Hughes et al. 2015), suggesting resource competition is a major driver of *P. martini*'s
347 displacement. Our analyses are consistent with the expectations of a competitive exploitation hypothesis,

348 demonstrating substantial overlap of major invertebrate prey categories between *H. mabouia* and *P.*
349 *martini* when the two species co-occur (Fig. 3; Supplemental materials). These prey categories largely
350 reflect common groups of invertebrates associated with human dwellings and artificial lighting in
351 Curaçao (Dornburg et al. 2016) and are consistent with studies of the diet of *H. mabouia* in other
352 urbanizing areas (Bonfiglio et al. 2006; Iturriaga and Marrero 2013; Drüke and Rödder 2017). In addition
353 to dietary overlap, a potential explanation for the dissimilarity in number of prey items found in *H.*
354 *mabouia* and *P. martini* may stem from the ambush prey capture tactics of *H. mabouia*. Fragments of
355 presumably larger prey items such as roaches, beetles, and spiders were often found in the stomachs of *H.*
356 *mabouia* in comparison with *P. martini*. Additionally, high numbers of isopods were found in some
357 individuals. This suggests that *H. mabouia* could be opportunistically feeding on larger prey as well prey
358 encountered in daytime refugia. The latter could also explain the finding of a blind snake within an
359 individual *H. mabouia*. While partially digested fragments of invertebrate body parts prohibit further
360 testing of whether *H. mabouia* is more effectively harnessing larger prey, this hypothesis raises several
361 possibilities of how the natural history of these species influences differential patterns of foraging and
362 prey capture.

363 There are different responses to bright lighting between these species with *Hemidactylus mabouia*
364 readily foraging directly at brightly lit artificial lights (Perry and Fisher 2006; Hughes et al. 2015). This
365 strategy reduces the energetic cost of finding prey as *H. mabouia* can harness the potential of artificial
366 lights as a lure for attracting large prey resources (Gaston et al. 2013) while simultaneously gaining a
367 potential thermal advantage (Perry et al. 2008). In contrast, *P. martini* avoids direct bright lights, and is
368 often found foraging along the more dimly lit periphery of buildings (Hughes et al. 2015). As such, *P.*
369 *martini* may have an ecological disadvantage to *H. mabouia*, as the former may need to spend more time
370 locating prey. Furthermore, this small change in prey foraging may put *P. martini* in contact with
371 arthropod vectors for nematodes not encountered by *H. mabouia*, as suggested by our observation of a
372 difference in parasite infestations between species. Given that lizards are often transport hosts for
373 mammalian parasites (Incedogan et al 2014; Dornburg et al 2019), including nematodes (Goldberg and
374 Bursey 2000), further testing of differences in parasite frequencies between *Hemidactylus* and its native
375 competitors represents an exciting direction additional research of high relevance to animal health.

376 In addition to having an advantage in light tolerance, our analyses of trait morphological variation
377 suggest that *H. mabouia* has a size advantage over *P. martini*, possessing overall larger size, as well as
378 larger heads, hind limbs, and other traits (Fig. 4 & Table 2). Increases in head height and head length are
379 associated with increases in bite force and more efficient prey capture in geckos (Cameron et al. 2013;
380 Massetti et al. 2017), as well as other lizard species (Verwaijen et al. 2002; Dufour et al. 2018).
381 Functionally, this advantage is thought to arise by the combination of increasing space to accommodate
382 increases in mandible adductor muscle sizes as well as changes in attachment angle that provide force
383 advantages (Herrel et al. 2001). Head sizes were larger in both male and female *H. mabouia* relative to *P.*
384 *martini*, with only head width (temporal width) significantly different between the sexes (Supplemental
385 materials). Males of the closely related *Hemidactylus turcicus* have also been found to have larger head
386 widths that are hypothesized to be the result of sexual selection (Iturriaga and Marrero 2013), and our
387 results suggest a similar pattern of dimorphism occurs in *H. mabouia*. However, larger heads also come at
388 a cost. Increased head sizes can negatively impact sprinting speed in lizards (Cameron et al. 2013), and
389 our additional finding of increased hind limb lengths in both sexes of *H. mabouia* may reflect the species
390 avoiding a fundamental locomotor trade-off (Table 3, Supplemental materials). A similar compensation

391 has been reported in *Hemidactylus frenatus* (Cameron et al. 2013) suggesting this is potentially a general
392 feature of *Hemidactylus* locomotor morphology.

393 *Hemidactylus mabouia* has been subjected to an unintentional experiment of introduction to
394 human mediated landscapes across the new world for centuries (Goeldi 1902; Van Buurt 2004; Carranza
395 and Arnold 2006). But, whether colonization of human structures has placed this species under selection
396 for changes in locomotor morphology remains unclear. Longer hind limbs in lizards are often correlated
397 with increased sprint speeds and forward propulsion in lizards (Bonine and Garland 1999; Cameron et al.
398 2013; Winchell et al 2018), thereby providing an advantage for an ambush predator such as *H. mabouia*
399 relying on a combination of ambush and pursuit to capture prey. Additionally, recent work placing front
400 limbs into the context of gecko locomotion models (Birn-Jeffery and Higham 2016; Zhuang and Higham
401 2016) provides strong evidence that locomotor function is decoupled between fore- and hind limbs. In
402 contrast to hind limbs, which act as primary axes of propulsion, front limbs are primarily used for braking
403 and downward locomotion (Birn-Jeffery and Higham 2016). Quantifications of limb morphology across
404 major lineages of geckos suggest shorter front limbs relative to hindlimbs to be a hallmark of gecko
405 locomotor morphology, with all species having between a 10 to 35% reduction in front limb proportions
406 (Hagey et al. 2017), a finding consistent with our analysis *H. mabouia* limb proportions (Supplemental
407 materials).

408 Primarily shorter front limbs could shorten the swing time, thereby aiding in maintaining speed
409 and stance in downward movements (Birn-Jeffery and Higham 2016). Our finding of shorter front to
410 hindlimbs in *H. mabouia* are consistent with expectations of selection for locomotion on steeply inclined
411 surfaces such as walls that is coupled with large hindlimbs for sprinting. But, the significant negative
412 scaling relationship between forelimb length and body size for *P. martini* also highlights the potential that
413 additional major differences in locomotor mode and performance between these species exist. Currently,
414 the foraging mode and activity patterns of *P. martini* remain little studied, as do those of *H. mabouia* in
415 their native range. As comparative studies of gecko functional locomotor morphology and performance
416 continue to illuminate the role of forelimbs in gecko locomotor morphology, future comparisons of
417 locomotor morphology and performance between and within these species offers a promising and exciting
418 research frontier.

419

420 *The role of predation in Hemidactylus invasions?*

421 Superiority in food resource competition has repeatedly been hypothesized as a major factor
422 facilitating the establishment of *H. mabouia* at the expense of native geckos (Petren and Case 1996;
423 Hoskin 2011; Hughes et al. 2015). Our stomach content analyses revealed a significant overlap of major
424 invertebrate prey resources, thereby supporting expectations of food resource competition (Table 1; Fig. 2
425 & 3). However, our study additionally provides direct evidence from stomach contents and indirect
426 evidence from the analysis of $\delta^{15}\text{N}$ isotopes that *Hemidactylus mabouia* acts as an opportunistic
427 vertebrate predator. These provide a broader context to previously reported single instances of predation
428 by *Hemidactylus mabouia* on *Phyllodactylus martini* (Dornburg et al. 2016) and *Gonatodes antillensis*
429 (Dornburg et al. 2011), as well as observations of cannibalism (Bonfiglio et al. 2006). We additionally
430 report the first instance of ophiophagy in *H. mabouia* (Table 1), suggesting that this species readily
431 consumes smaller vertebrate prey. This raises a question: How frequent are such predation events?

432 Our isotopic analyses provide some insights to this question, indicating numerous individual *H.*
433 *mabouia* are feeding at a trophic level higher than *P. martini*. As vertebrates represent the only consumers
434 of a higher trophic position in Curaçao, $\delta^{15}\text{N}$ values suggest that predation of vertebrates by *H. mabouia*

435 may not be rare events. Investigations of feeding ecology of the closely related *H. frenatus* have also
436 reported intraspecific juvenile predation (Hunsaker 1966; Bolger and Case 1992; Case et al. 1994),
437 suggesting *H. mabouia* to be similarly opportunistic. Juvenile *H. mabouia* are known to avoid predation
438 by larger adults by foraging low to the ground (Howard et al. 2001). It is likely that our results reflect a
439 signature of juvenile mortality of *P. martini* as juveniles of this species will readily forage across a range
440 of wall elevations including those that are occupied by *H. mabouia* (van Buurt 2004). In addition to
441 lowering recruitment, predation on juvenile *P. martini* could offer a competitive advantage for juvenile *H.*
442 *mabouia*. By reducing the density of interspecific competitors at the juvenile stage, more juvenile *H.*
443 *mabouia* would be able to transition to adult and more swiftly increase overall population sizes. As *H.*
444 *mabouia* readily achieves high carrying capacities that can exceed those of other *Hemidactylus* species
445 (Short and Petren 2011), this in turn could greatly increase the pressure of additional density dependent
446 effects on the persistence of native species.

447 Evidence for predation of smaller vertebrates by *H. mabouia* raises the concern that in addition to
448 displacing populations of native geckos, the presence of *H. mabouia* can negatively impact overall
449 population structure. Although demographic studies of geckos impacted by *H. mabouia* have been
450 limited, analyses of *Phyllodactylus tuberculosis* in Mexico have implicated the presence of *H. mabouia* in
451 severe contractions of effective population size and recent genetic bottlenecks (Blair et al. 2015). We
452 argue that further assessing the role of *H. mabouia* in juvenile survivorship represents an important, but
453 currently neglected aspect of this species invasion biology. These studies are of particular importance as
454 *H. mabouia* is increasingly being found in non-urban areas throughout its invaded range (Rocha et al.
455 2011), challenging the assumption that this invasion is limited to urbanizing areas. Fortunately for
456 Curaçao and other similar desert habitats, invasion into the native bush habitat may not be possible due to
457 mechanical properties of *H. mabouia* toe pad adhesion. *Hemidactylus* geckos all possess a basal toe pad
458 system that may not be capable of successfully gaining traction on the loose, and dusty, rocky soil of the
459 island (Russell and Delaugerre 2017). This hypothesis remains to be tested, but if it is supported, the
460 impact of predation and competition on native gecko population sizes in dusty, arid environments could
461 be mitigated by integrating continual corridors of native habitat into urban planning efforts. Such efforts
462 would yield ‘enemy-free’ space and thereby increase the probability of the long-term persistence of native
463 gecko species (Cole et al. 2005).

464

465 **Acknowledgements**

466 We thank Mark Vermeij and the entire staff of the Caribbean Research and Management of
467 Biodiversity foundation (CARMABI) for facilitating collection efforts and export permitting, Sunshine
468 and David at Curaçao Sunshine Getaways for enabling us to convert the second floor of their home into a
469 field station, and the Westpunt Syndicate for their continual field assistance. Specimens were collected
470 under the approved guidelines of IACUC (Institutional Animal Care and Use Committee) protocol 2012-
471 10681 (Ichthyology and Herpetology at the Yale Peabody Museum of Natural History). All data and
472 analysis scripts associated with this study have been archived on Zenodo (doi pending acceptance). We
473 also would like to thank E. Ferraro, Enie Hensel, and two anonymous referees for excellent feedback on
474 previous versions of this manuscript.

475

476 **Literature Cited**

477 Allen M, Poggiali D, Whitaker K, Marshall TR, Kievit, RA (2019) Raincloud plots: a multi-platform tool
478 for robust data visualization. Wellcome open research 4:63

- 479
480 González-Oreja JA, Zuria I, Carbo-Ramirez P, Charre GM (2018) Using variation
481 partitioning techniques to quantify the effects of invasive alien species on native urban bird
482 assemblages. *Biol Invasions* 20:2861–2874
483
- 484 Bateman PW, Fleming PA (2012) Big city life: carnivores in urban environments. *J Zool* 287:1–23
485
- 486 Birn-Jeffery AV, Higham TE (2016) Geckos decouple fore- and hind limb kinematics in response to
487 changes in incline. *Front Zool* 13:11
488
- 489 Blair C, Jiménez Arcos VH, de la Cruz FRM, Murphy RW (2015) Historical and contemporary
490 demography of leaf-toed geckos (Phyllodactylidae: *Phyllodactylus tuberculosus saxatilis*) in the
491 Mexican dry forest. *Conservation Genetics* 16:419–429
492
- 493 Bolger DT, Case TJ (1992) Intra- and interspecific interference behaviour among sexual and asexual
494 geckos. *Animal Behaviour* 44:21–30
495
- 496 Bonfiglio F, Balestrin RL, Cappellari LH (2006) Diet of *Hemidactylus mabouia* (Sauria, Gekkonidae) in
497 urban area of southern Brazil. *Biociências* 14:107–111
498
- 499 Bonine KE, Garland Jr T (1999) Sprint performance of phrynosomatid lizards, measured on a high-speed
500 treadmill, correlates with hindlimb length. *Journal of Zoology*, 248:255-265
501
- 502 Burgos-Rodríguez JA, Avilés-Rodríguez KJ, Kolbe JJ (2016) Effects of invasive Green Iguanas (*Iguana*
503 *iguana*) on seed germination and seed dispersal potential in southeastern Puerto Rico. *Biological*
504 *Invasions* 18:2775–2782
505
- 506 Burnham KP, Anderson DR (2004) Multimodel inference: understanding AIC and BIC in model
507 selection. *Sociological methods & research*, 33:261-304
508
- 509 Buzan E (2017) Changes in rodent communities as consequence of urbanization and inappropriate waste
510 management. *Appl Ecol Environ Res* 15:573–588
511
- 512 Cameron SF, Wynn ML, Wilson RS (2013) Sex-specific trade-offs and compensatory mechanisms: bite
513 force and sprint speed pose conflicting demands on the design of geckos (*Hemidactylus frenatus*).
514 *J Exp Biol* 216:3781–3789
515
- 516 Capinha C, Seebens H, Cassey P, García-Díaz P, Lenzner B, Mang T, Moser D, Pyšek P, Rödder D,
517 Scalera R, Winter M, Dullinger S, Essl F (2017) Diversity, biogeography and the global flows of
518 alien amphibians and reptiles. *Diversity and Distributions* 23:1313–1322
519
- 520 Carranza S, Arnold EN (2006) Systematics, biogeography, and evolution of *Hemidactylus* geckos
521 (Reptilia: Gekkonidae) elucidated using mitochondrial DNA sequences. *Mol Phylogenet Evol*
522 38:531–545

- 523
524 Case TJ, Bolger DT, Petren K (1994) Invasions and Competitive Displacement among House Geckos in
525 the Tropical Pacific. *Ecology* 75:464–477
526
527 Chapman MG, Underwood AJ (1999) Ecological patterns in multivariate assemblages: information and
528 interpretation of negative values in ANOSIM tests. *Mar Ecol Prog Ser* 180:257–265
529
530 Clarke KR (1993) Non-parametric multivariate analyses of changes in community structure. *Aust J Ecol*
531 18:117–143
532
533 Cole NC, Jones CG, Harris S (2005) The need for enemy-free space: The impact of an invasive gecko on
534 island endemics. *Biological Conservation* 125:467–474
535
536 Donald GL, Paterson CG (1977) Effect of preservation on wet weight biomass of chironomid larvae.
537 *Hydrobiologia* 53:75–80
538
539 Donihue CM, Herrel A, Fabre A-C, Kamath A, Geneva AJ, Schoener TW, Kolba JJ, Losos JB (2018)
540 Hurricane-induced selection on the morphology of an island lizard. *Nature* 560:88–91
541
542 Dornburg A, Lippi C, Federman S, et al (2016) Disentangling the Influence of Urbanization and Invasion
543 on Endemic Geckos in Tropical Biodiversity Hot Spots: A Case Study of *Phyllodactylus martini*
544 (Squamata: Phyllodactylidae) along an Urban Gradient in Curaçao. *Bulletin of the Peabody*
545 *Museum of Natural History* 57:147–164
546
547 Dornburg A, Warren DL, Iglesias T, Brandley MC (2011) Natural History Observations of the
548 Ichthyological and Herpetological Fauna on the Island of Curaçao (Netherlands). *Bulletin of the*
549 *Peabody Museum of Natural History* 52:181–186
550
551 Dornburg A, Lamb AD, Warren D, Watkins-Colwell GJ, Lewbart GA, Flowers J (2019). Are Geckos
552 Paratenic Hosts for Caribbean Island Acanthocephalans? Evidence from *Gonatodes antillensis*
553 and a Global Review of Squamate Reptiles Acting as Transport Hosts. *Bulletin of the Peabody*
554 *Museum of Natural History*, 60:55-79.
555
556 Drüke Y, Rödder D (2017) Feeding ecology of the invasive gecko species *Hemidactylus mabouia*
557 (Moreau de Jonnès, 1818) (Sauria: Gekkonidae) in São Sebastião (Brazil). *Bonn Zool Bull*
558 66:85–93
559
560 Dufour CMS, Losos JB, Herrel A (2018) Do differences in bite force and head morphology between a
561 native and an introduced species of anole influence the outcome of species interactions?
562 *Biological Journal of the Linnean Society* XX:1–10
563
564 Falcón W, Ackerman JD, Recart W, Daehler CC (2013) Biology and Impacts of Pacific Island Invasive
565 Species. 10. *Iguana iguana*, the Green Iguana (Squamata: Iguanidae). *Pac Sci* 67:157–186
566

- 567 Gaston KJ, Bennie J, Davies TW, Hopkins J (2013) The ecological impacts of nighttime light pollution: a
568 mechanistic appraisal. *Biol Rev Camb Philos Soc* 88:912–927
569
- 570 Goeldi EA (1902) Lagartos do Brazil. *Boletim do Museu Paraense* 3:499–560
571
- 572 Goldberg SR, Bursley CR (2000) Transport of helminths to Hawaii via the brown anole, *Anolis sagrei*
573 (Polychrotidae). *Journal of Parasitology* 86:750-756.
574
- 575 Hagey TJ, Harte S, Vickers M, et al (2017) There's more than one way to climb a tree: Limb length and
576 microhabitat use in lizards with toe pads. *PLoS One* 12:e0184641
577
- 578 Herrel A, De Grauw E, Lemos-Espinal JA (2001) Head shape and bite performance in xenosaurid lizards.
579 *J Exp Zool* 290:101–107
580
- 581 Hintze JL, Nelson RD (1998) Violin Plots: A Box Plot-Density Trace Synergism. *The American*
582 *Statistician* 52:181
583
- 584 Hoskin CJ (2011) The invasion and potential impact of the Asian House Gecko (*Hemidactylus frenatus*)
585 in Australia. *Austral Ecology* 36:240–251
586
- 587 Howard KG, Parmerlee JS, Powell R (2001) Natural history of the edificarian geckos *Hemidactylus*
588 *mabouia*, *Thecadactylus rapicauda*, and *Sphaerodactylus sputator* on Anguilla. *Caribb J Sci*
589 37:285–287
590
- 591 Hughes DF, Meshaka WE Jr, van Buurt G (2015) The superior colonizing gecko *Hemidactylus mabouia*
592 on Curaçao: conservation implications for the native gecko *Phyllodactylus martini*. *J Herpetol*
593 49:60–63
594
- 595 Hunsaker D (1966) Notes on the population expansion of the house gecko, *Hemidactylus frenatus*.
596 *Philipp J Sci* 95:121–122
597
- 598 Incedogan S, Yildirimhan HS, Bursley CR, 2014. Helminth parasites of the ocellated skink, *Chalcides*
599 *ocellatus* (Forsk., 1775) (Scincidae) from Turkey. *Comparative Parasitology*, 81:260-269
600
- 601 Iturriaga M, Marrero R (2013) Feeding ecology of the Tropical House Gecko *Hemidactylus mabouia*
602 (Sauria: Gekkonidae) during the dry season in Havana, Cuba. *Herpetol Notes* 6:11–17
603
- 604 Kratochvíl L, Frynta D (2002) Body size, male combat and the evolution of sexual dimorphism in
605 eublepharid geckos (Squamata: Eublepharidae). *Biological Journal of the Linnean Society*
606 76:303-314.
607
- 608 Kraus F (2015) Impacts from Invasive Reptiles and Amphibians. *Annu Rev Ecol Evol Syst* 46:75–97
609
- 610 Lapedra O, Chejanovski Z, Kolbe JJ (2017) Urbanization and biological invasion shape animal

- 611 personalities. *Glob Chang Biol* 23:592–603
- 612
- 613 Liu C, He D, Chen Y, Olden JD (2017) Species invasions threaten the antiquity of China’s freshwater fish
- 614 fauna. *Diversity and Distributions* 23:556–566
- 615
- 616 Marshall JD, Brooks JR, Lajtha K (2007) Sources of variation in the stable isotopic composition of plants.
- 617 *Stable isotopes in ecology and environmental science* 2:22–60
- 618
- 619 Massetti F, Gomes V, Perera A, et al (2017) Morphological and functional implications of sexual size
- 620 dimorphism in the Moorish gecko, *Tarentola mauritanica*. *Biological Journal of the Linnean*
- 621 *Society* 122:197–209
- 622
- 623 McKinney ML (2006) Urbanization as a major cause of biotic homogenization. *Biological Conservation*
- 624 127:247–260
- 625
- 626 McKinney ML, Lockwood JL (1999) Biotic homogenization: a few winners replacing many losers in the
- 627 next mass extinction. *Trends Ecol Evol* 14:450–453
- 628
- 629 Mooney HA, Cleland EE (2001) The evolutionary impact of invasive species. *Proc Natl Acad Sci USA*
- 630 98:5446–5451
- 631
- 632 Morais P, Reichard M (2018) Cryptic invasions: A review. *Sci Total Environ* 613-614:1438–1448
- 633
- 634 Paini DR, Sheppard AW, Cook DC, et al (2016) Global threat to agriculture from invasive species. *Proc*
- 635 *Natl Acad Sci USA* 113:7575–7579
- 636
- 637 Pedersen T, Fuhrmann MM, Lindstrøm U, et al (2018) Effects of the invasive red king crab on food web
- 638 structure and ecosystem properties in an Atlantic fjord. *Mar Ecol Prog Ser* 596:13–31
- 639
- 640 Perry G (1996) The evolution of sexual dimorphism in the lizard *Anolis polylepsis* (Iguania): evidence
- 641 from intraspecific variation in foraging behavior and diet. *Can J Zool* 74:1238–1245
- 642
- 643 Perry G, Buchanan BW, Fisher RN, et al (2008) Effects of artificial night lighting on amphibians and
- 644 reptiles in urban environments. *Urban herpetology* 3:239–256
- 645
- 646 Perry G, Fisher RN (2006) Night lights and reptiles: observed and potential effects. In: *Ecological*
- 647 *consequences of artificial night lighting*. 169–191
- 648
- 649 Petren K, Case TJ (1998) Habitat structure determines competition intensity and invasion success in
- 650 gecko lizards. *Proc Natl Acad Sci U S A* 95:11739–11744
- 651
- 652 Petren K, Case TJ (1996) An Experimental Demonstration of Exploitation Competition in an Ongoing
- 653 Invasion. *Ecology* 77:118–132
- 654

- 655 Post DM (2002) Using stable isotopes to estimate trophic position: models, methods, and assumptions.
656 Ecology 83:703–718
657
- 658 R Core Team (2018) R: A language and environment for statistical computing. R Foundation for
659 Statistical Computing, Vienna, Austria
660
- 661 Richmond JQ, Wood DA, Stanford JW, Fisher RN (2015) Testing for multiple invasion routes and source
662 populations for the invasive brown treesnake (*Boiga irregularis*) on Guam: implications for pest
663 management. Biological Invasions 17:337–349
664
- 665 Rocha CFD, Anjos LA, Bergallo HG (2011) Conquering Brazil: the invasion by the exotic gekkonid
666 lizard *Hemidactylus mabouia* (Squamata) in Brazilian natural environments. Zoologia (Curitiba)
667 28:747–754
668
- 669 Roches SD, Des Roches S, Harmon LJ, Rosenblum EB (2016) Colonization of a novel depauperate
670 habitat leads to trophic niche shifts in three desert lizard species. Oikos 125:343–353
671
- 672 Rodda GH, Savidge JA (2007) Biology and Impacts of Pacific Island Invasive Species. 2. *Boiga*
673 *irregularis*, the Brown Tree Snake (Reptilia: Colubridae) 1. Pac Sci 61:307–325
674
- 675 Rödder D, Solé M, Böhme W (2008) Predicting the potential distributions of two alien invasive
676 Housegeckos (Gekkonidae: *Hemidactylus frenatus*, *Hemidactylus mabouia*). North-Western
677 Journal of Zoology 4:236–246
678
- 679 Russell AP, Delaugerre MJ (2017) Left in the dust: differential effectiveness of the two alternative
680 adhesive pad configurations in geckos (Reptilia: Gekkota). Journal of Zoology 301:61–68
681
- 682 Scharf I, Meiri S, (2013) Sexual dimorphism of heads and abdomens: different approaches to ‘being
683 large’ in female and male lizards. Biological Journal of the Linnean Society, 110:665-673.
684
- 685 Shechonge A, Ngatunga BP, Bradbeer SJ, et al (2019) Widespread colonisation of Tanzanian catchments
686 by introduced *Oreochromis tilapia* fishes: the legacy from decades of deliberate introduction.
687 Hydrobiologia 832:235–253
688
- 689 Short KH, Petren K (2011) Rapid species displacement during the invasion of Florida by the tropical
690 house gecko *Hemidactylus mabouia*. Biol Invasions 14:1177–1186
691
- 692 Smith BJ, Cherkiss MS, Hart KM, et al (2016) Betrayal: radio-tagged Burmese pythons reveal locations
693 of conspecifics in Everglades National Park. Biological Invasions 18:3239–3250
694
- 695 Toussaint A, Beauchard O, Oberdorff T, et al (2016) Worldwide freshwater fish homogenization is driven
696 by a few widespread non-native species. Biological Invasions 18:1295–1304
697
- 698 Trentanovi G, von der Lippe M, Sitzia T, et al (2013) Biotic homogenization at the community scale:

- 699 disentangling the roles of urbanization and plant invasion. *Diversity and Distributions* 19:738–
700 748
701
- 702 Useni Sikuzani Y, Sambiéni Kouagou R, Maréchal J, et al (2018) Changes in the Spatial Pattern and
703 Ecological Functionalities of Green Spaces in Lubumbashi (the Democratic Republic of Congo)
704 in Relation With the Degree of Urbanization. *Tropical Conservation Science*
705 11:1940082918771325
706
- 707 van Buurt G (2004) Field Guide to the Reptiles and Amphibians of Aruba, Curacao and Bonaire. Serpents
708 Tale
709
- 710 van Buurt G (2006) Conservation of amphibians and reptiles in Aruba, Curaçao and Bonaire. *Applied*
711 *Herpetology* 3:307–321
712
- 713 Verwajen D, Van Damme R, Herrel A (2002) Relationships between head size, bite force, prey handling
714 efficiency and diet in two sympatric lacertid lizards. *Functional Ecology* 16:842–850
715
- 716 Vidal MA, Sabat P (2010) Stable isotopes document mainland–island divergence in resource use without
717 concomitant physiological changes in the lizard *Liolaemus pictus*. *Comparative Biochemistry and*
718 *Physiology Part B: Biochemistry and Molecular Biology* 156:61–67
719
- 720 Vitt LJ (1983) Tail loss in lizards: the significance of foraging and predator escape modes. *Herpetologica*
721 39:151–162
722
- 723 Weber MJ, Brown ML (2009) Effects of Common Carp on Aquatic Ecosystems 80 Years after “Carp as a
724 Dominant”: Ecological Insights for Fisheries Management. *Rev Fish Sci* 17:524–537
725
- 726 Weterings R, Vetter KC (2018) Invasive house geckos (*Hemidactylus* spp.): their current, potential and
727 future distribution. *Curr Zool* 64:559–573
728
- 729 Wiles GJ, Bart J, Beck RE, Aguon CF (2003) Impacts of the Brown Tree Snake: Patterns of Decline and
730 Species Persistence in Guam’s Avifauna. *Conservation Biology* 17:1350–1360
731
- 732 Williams R, Pernetta AP, Horrocks JA (2016) Outcompeted by an invader? Interference and exploitative
733 competition between tropical house gecko (*Hemidactylus mabouia*) and Barbados leaf-toed gecko
734 (*Phyllodactylus pulcher*) for diurnal refuges in anthropogenic coastal habitats. *Integr Zool*
735 11:229–238
736
- 737 Willson JD (2017) Indirect effects of invasive Burmese pythons on ecosystems in southern Florida. *J*
738 *Appl Ecol* 54:1251–1258
739
- 740 Winchell KM, Maayan I, Fredette JR, Revell LJ (2018) Linking locomotor performance to morphological
741 shifts in urban lizards. *Proc Biol Sci* 285.: doi: 10.1098/rspb.2018.0229
742

743 Winchell KM, Reynolds RG, Prado-Irwin SR, et al (2016) Phenotypic shifts in urban areas in the tropical
744 lizard *Anolis cristatellus*. *Evolution* 70:1009–1022
745

746 Yam RSW, Fan Y-T, Wang T-T (2016) Importance of Macrophyte Quality in Determining Life-History
747 Traits of the Apple Snails *Pomacea canaliculata*: Implications for Bottom-Up Management of an
748 Invasive Herbivorous Pest in Constructed Wetlands. *Int J Environ Res Public Health* 13: doi:
749 10.3390/ijerph13030248
750

751 Young HS, Parker IM, Gilbert GS, et al (2017) Introduced Species, Disease Ecology, and Biodiversity-
752 Disease Relationships. *Trends Ecol Evol* 32:41–54
753

754 Zaaf A, Van Damme R (2001) Limb proportions in climbing and ground-dwelling geckos (Lepidosauria,
755 Gekkonidae): a phylogenetically informed analysis. *Zoomorphology* 121:45–53
756

757 Zhuang MV, Higham TE (2016) Arboreal Day Geckos (*Phelsuma madagascariensis*) Differentially
758 Modulate Fore- and Hind Limb Kinematics in Response to Changes in Habitat Structure. *PLoS*
759 *One* 11:e0153520
760
761
762
763
764
765
766
767
768
769
770
771
772
773
774
775
776
777
778
779
780
781
782
783
784
785
786

787
788
789
790
791
792
793
794

Tables

Table 1 | Relative frequency of prey stomach content items across all sampled individual *Hemidactylus mabouia* (total items = 72) and *Phyllodactylus martini* (total items = 115). Bold values indicate sums of prey items within major categories (i.e., Insecta).

	<i>Hemidactylus mabouia</i>	<i>Phyllodactylus martini</i>
Prey	Relative Frequency	Relative Frequency
Arachnida	0.141	0.191
Scorpiones	---	0.052
Araneae	0.14	0.139
Chilopoda	---	0.009
Insecta	0.507	0.583
Blattaria	0.042	0.00
Coleoptera	0.127	0.235
Diptera	0.056	0.078
Ephemeroptera	0	0.009
Hemiptera	0.056	0.017
Hymenoptera	0.099	0.061
Lepidoptera	0.099	0.096
Orthoptera	0.028	0.087
Isopoda	0.282	0.087
Skin Shed	0.014	0.009
Vertebrata	0.042	---
<i>Gonatodes antillensis</i>	0.014	---
<i>Phyllodactylus martini</i>	0.014	---
<i>Ramphotyphlops braminus</i>	0.014	---
Other	0.014	0.128

795

796
797
798
799
800
801

Table 2 | ANCOVA results testing the effect of snout-vent length (SVL), species, and their interaction on measured morphological characters. Bolded values indicate significant effects. * stands for *P*-values ranging from 0.05 to 0.01, ** for *P*-values ranging from 0.01 to 0.001 and *** for *P*-values smaller than 0.001

Trait	log(SVL) (F/P)	Species (F/P)	Species:log(SVL) (F/P)
Post Orbit Width	121.82/***	5.98/*	0.07/0.78
Temporalis Width	1151.42/***	32.45/***	0.07/0.78
Head Length	1298.73/***	101.84/***	0.07/0.78
Jaw Length	526.77/***	86.16/***	0.18/0.66
Head Height	358.09/***	3.23/0.07	0.35/0.55
Humerus Length	163.53/***	43.52/***	8.30/**
Radius Length	464.95/***	54.76/***	0.09/0.76
Femur Length	307.27/***	59.16/***	2.19/0.14
Tibia Length	140.64/***	20.94/***	0.08/0.76

802
803
804
805
806
807
808
809
810
811
812
813
814
815
816
817
818
819
820
821
822
823
824
825

826
827
828
829
830
831
832
833
834
835
836
837
838
839
840
841
842
843
844
845
846
847
848
849
850
851

Figure Legends

Fig. 1 | Map of study area. Circles indicate sampling locations.

Fig. 2 | Violin plots of isotopic data. **A.** Estimated trophic position for *Phyllodactylus martini* (n=17) and *Hemidactylus mabouia* (n=21) using Nitrogen. **B.** Carbon. Raw nitrogen and carbon isotopic values were corrected using average baseline values across all sites.

Fig. 3 | Visualization of stomach content data. **A.** Data from analyzed *Phyllodactylus martini* (n=72). **B.** *Hemidactylus mabouia* (n=59). Columns in sphere correspond to the relative frequency of an individual's prey items. Colors correspond to matching prey categories in legend.

Fig. 4 | Non-metric multidimensional scaling analysis contrasting stomach contents. Data from analyzed *Phyllodactylus martini* (n=79) and *Hemidactylus mabouia* (n=57). Ellipses encompass the 95% confidence interval around the centroid of each species, and prey labels indicate location of prey categories within the diet space.

Fig. 5 | Analysis of morphometric traits. **A.** Raincloud plots visualizing SVL differences between *Phyllodactylus martini* (yellow) and *Hemidactylus mabouia* (brown), depicting the probability distribution through a rotated violin plot (top), box plot summary of quartiles (middle), and raw data (bottom) for each species. **B-D.** Principal components analysis showing overlap of morphological traits between species. Principal component scores are visualized for each axis and species with background shading representing fitted convex hulls of the morphospace occupied by each species.

Figure 1

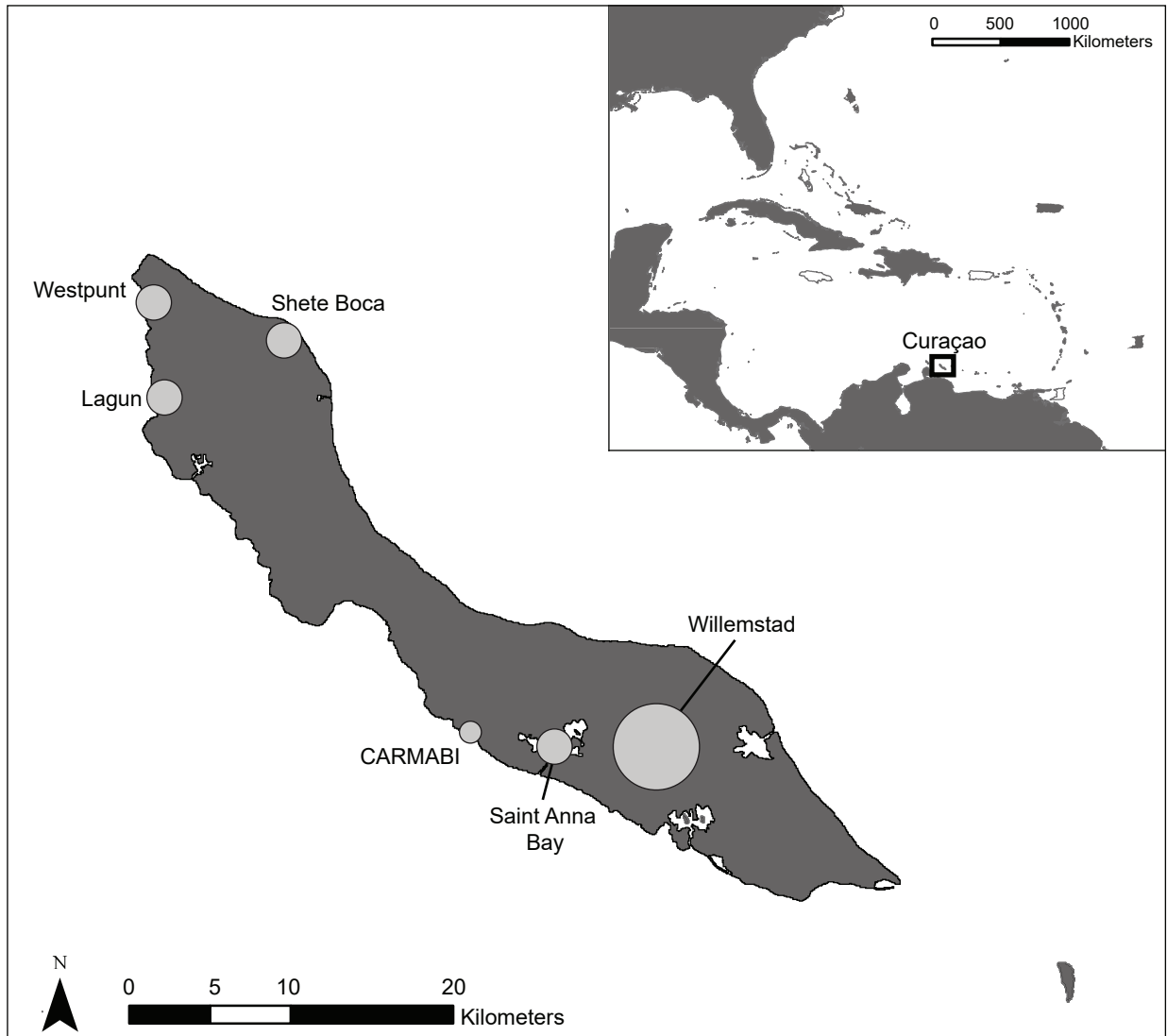


Figure 2

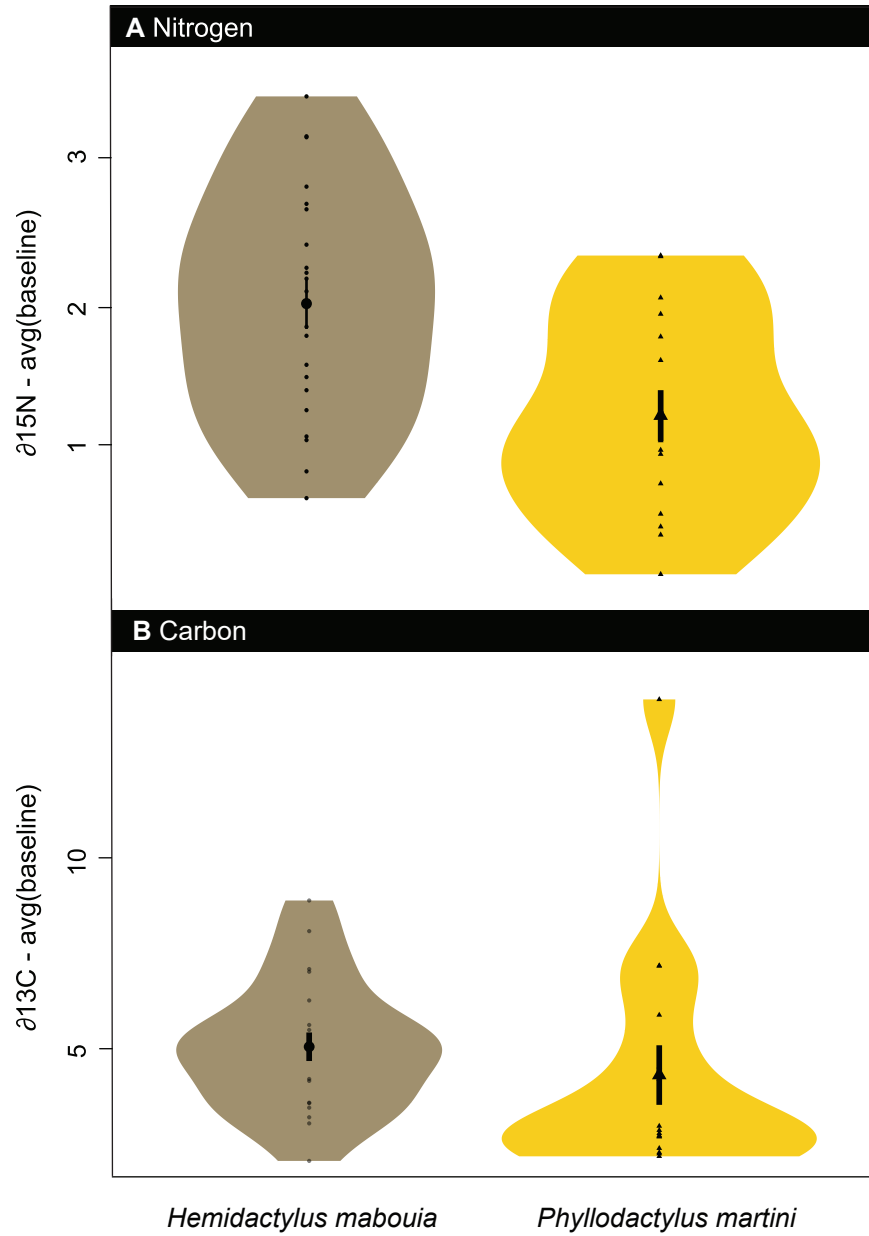


Figure 3

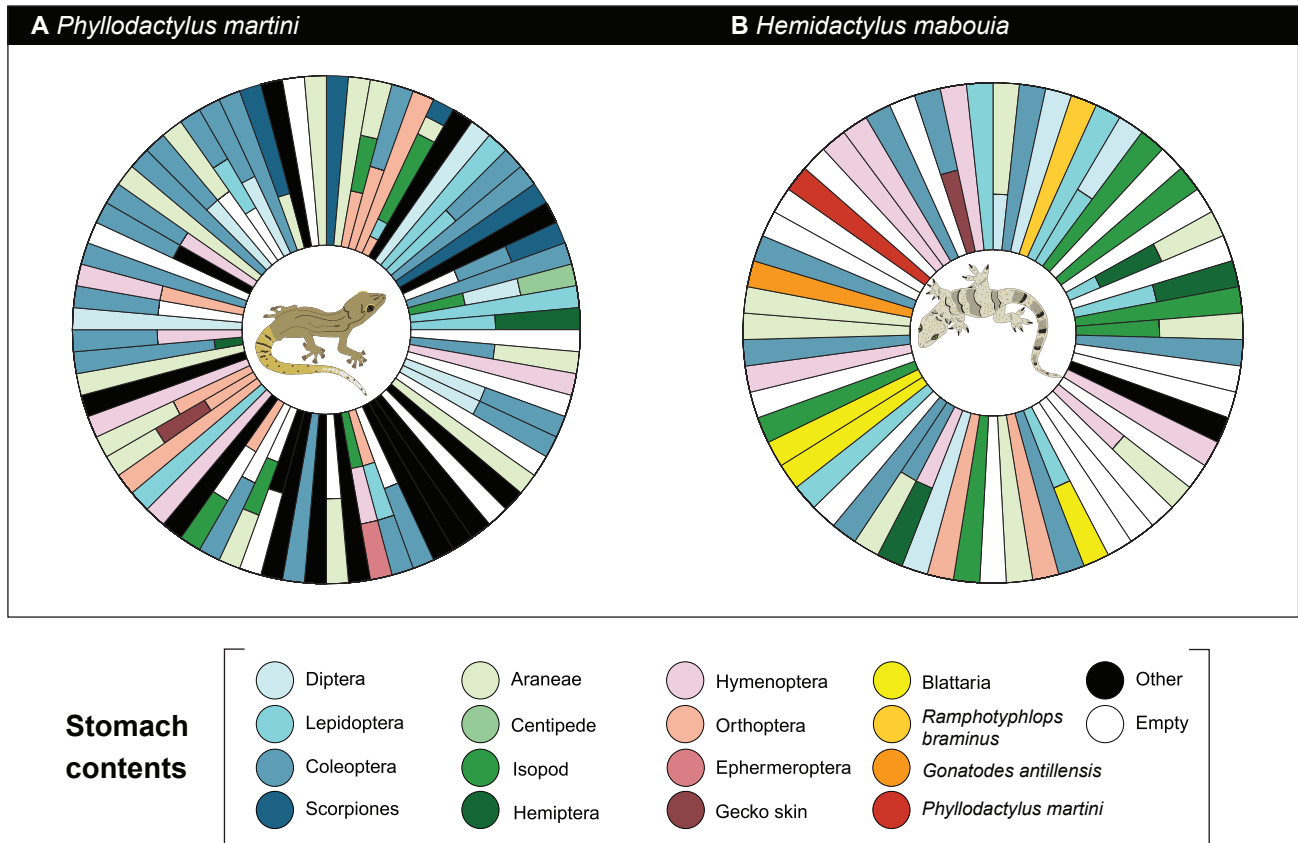


Figure 4

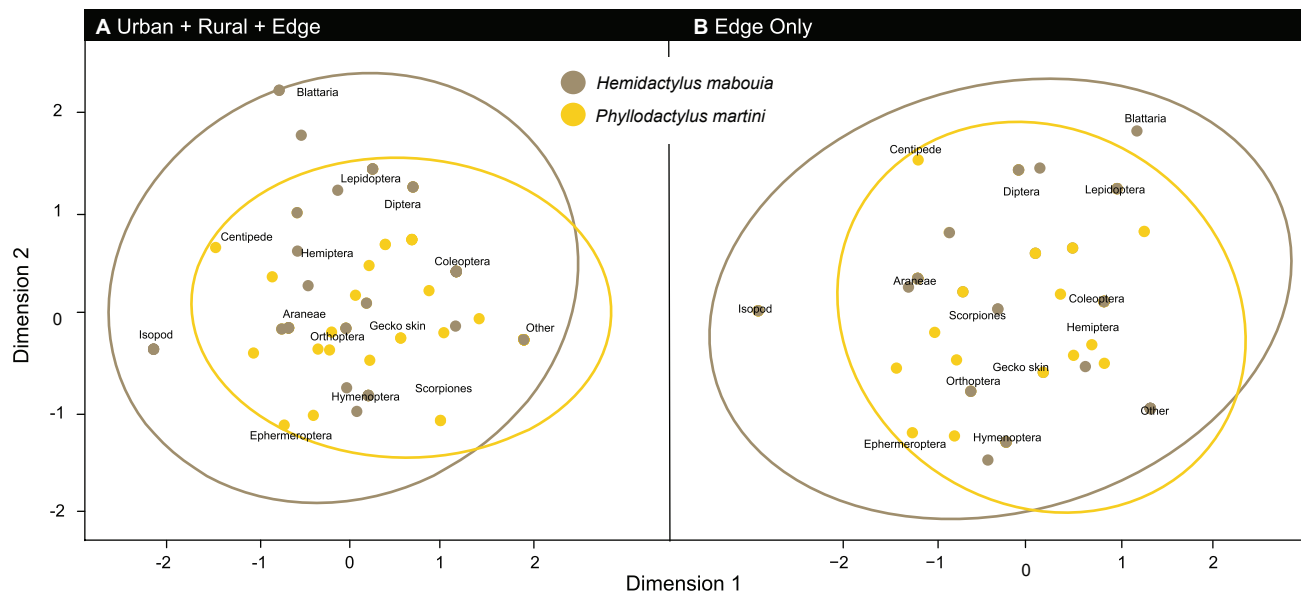


Figure 5

