1	Efficient assessment of nocturnal flying insect communities by
2	combining automatic light traps and DNA metabarcoding
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4	Vanessa A. Mata ^{1,*} , Sónia Ferreira ^{1,*} , Rebecca Campos ¹ , Luís P. da Silva ¹ , Joana
5	Veríssimo ^{1,2} , Martin F. V. Corley ¹ , Pedro Beja ^{1,3}
6	
7	* These authors contributed equally to the paper
8	
9	¹ CIBIO-InBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos,
10	Universidade do Porto, Campus de Vairão, Vila do Conde, Portugal
11	
12	² Departamento de Biologia, Faculdade de Ciências, Universidade do Porto, Porto, Portugal
13	
14	³ CIBIO-InBIO, Centro de Investigação em Biodiversidade e Recursos Genéticos, Instituto
15	Superior de Agronomia, Universidade de Lisboa, Lisboa, Portugal
16	
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18 Summary

Increasing evidence for global insect declines is prompting a renewed interest in the survey of
 whole insect communities. DNA metabarcoding can contribute to assessing diverse insect
 communities over a range of spatial and temporal scales, but efforts are still needed to optimise
 and standardise procedures, from field sampling, through laboratory analysis, to bioinformatic
 processing.

Here we describe and test a methodological pipeline for surveying nocturnal flying insects,
 combining a customised automatic light trap and DNA metabarcoding. We optimised laboratory
 procedures and then tested the methodological pipeline using 12 field samples collected in
 northern Portugal in 2017. We focused on Lepidoptera to compare metabarcoding results with
 those from morphological identification, using three types of bulks produced from each sample
 (individuals, legs and the unsorted mixture).

3. The customised trap was highly efficient at collecting nocturnal flying insects, allowing a small 30 team to operate several traps per night, and a fast field processing of samples for subsequent 31 32 metabarcoding with low contamination risks. Morphological processing yielded 871 identifiable 33 individuals of 102 Lepidoptera species. Metabarcoding detected a total of 528 taxa, most of 34 which were Lepidoptera (31.1%), Diptera (26.1%) and Coleoptera (14.7%). There was a 35 reasonably high matching in community composition between morphology and metabarcoding 36 when considering the 'individuals' and 'legs' bulk samples, with few errors mostly associated 37 with morphological misidentification of small microlepidoptera. Regarding the 'mixture' bulk 38 sample, metabarcoding identified nearly four times more Lepidoptera species than 39 morphological examination.

40 4. Our study provides a methodological metabarcoding pipeline that can be used in standardised 41 surveys of nocturnal flying insects, showing that it can overcome limitations and potential 42 shortcomings of traditional methods based on morphological identification. Our approach 43 efficiently collects highly diverse taxonomic groups such as nocturnal Lepidoptera that are poorly 44 represented when using Malaise traps and other widely used field methods. To enhance the 45 potential of this pipeline in ecological studies, efforts are needed to test its effectiveness and 46 potential biases across habitat types and to extend the DNA barcode databases for important 47 groups such as Diptera.

48

49 Introduction

- 50 Recent studies have shown precipitous declines in insect populations, which can have far-reaching
- 51 consequences for ecosystem functioning and thus to human lives and livelihoods (Basset and
- 52 Lamarre 2019; Wagner 2020). One of the striking features of this apparent decline is that it seems to
- 53 be affecting entire insect communities, rather than a few species of conservation concern or
- 54 particularly sensitive species groups (Hallamann et al. 2017; Bell et al. 2020; Wagner 2020). Because
- of this, there is an urgent need to monitor whole insect communities and to understand the main
- 56 drivers of community change over a range of spatial and temporal scales (Sánchez-Bayo and
- 57 Wyckhuys 2019). This task is challenging due to taxonomic impediment (sensu, e.g., Ebach et al.
- 58 2011), which makes it hard to describe the huge diversity of insect communities with conventional
- 59 methods in a cost-effective way.

60 The advent of next-generation DNA sequencing coupled with metabarcoding approaches is

- 61 revolutionising the study of diverse insect communities by overcoming taxonomic impediment,
- allowing the cost-effective processing of hundreds to thousands of complex mixed community
- 63 samples at high taxonomic resolution (Douglas et al. 2012; Barsoum et al. 2019; Gueuning et al.
- 64 2019; Piper et al. 2019). Typically, metabarcoding studies of communities of insects and other
- 65 invertebrates involve the collection of field samples using a variety of methods, and then DNA from
- 66 multi-species samples is extracted, amplified using PCR and sequenced using a next-generation
- 67 sequencing platform (e.g. Braukmann et al. 2019; Marquina et al. 2019; Zenker 2020). Sequencing
- 68 data is used to produce a list of taxa recorded at each site, using bioinformatic pipelines and
- 69 reference libraries of DNA barcodes (Marquina et al. 2019; Zenker 2020). Applications of this general
- approach are increasing, particularly in the case of freshwater communities (e.g., Elbrecht et al.
- 71 2017; Bush et al. 2020), where efforts are underway to develop standardised metabarcoding
- approaches to be used in official monitoring programs such as the European Water Framework
- 73 Directive (Hering et al. 2018). The use of metabarcoding in studies of terrestrial insects has lagged
- behind that of freshwater communities, but the technique has already been tested, for instance, in
- the monitoring of wild bees (Gueuning et al. 2019), invasive species (Piper et al. 2019) and dung
- insects for ecotoxicological assessments (Blanckenhorn et al. 2016), among many others. Despite
- these advances, however, considerable efforts are still needed to develop, optimise and standardise
- 78 efficient methods to collect and process insect samples for DNA metabarcoding studies and
- 79 monitoring programs, as results are conditional on methodological alternatives adopted, including
- 80 DNA extraction, primer sets and bioinformatic pipelines (Brandon-Mong et al. 2016; Braukman et al.
- 81 2019; Elbrecht et al. 2019).

82 To the best of our knowledge, DNA metabarcoding has yet to be used for describing communities of 83 nocturnal insects, possibly due to the bias of ecologists towards studying daytime phenomena 84 (Gaston 2019). This is regrettable, because nocturnal insects encompass about half of all insect 85 species, and they can be negatively impacted by factors that do not operate during the day, such as 86 light pollution (Owens et al. 2020). Furthermore, they are key components of natural and 87 anthropogenic ecosystems, playing significant roles as, for instance, pollinators (Macgregor et al. 2015), crop pests (Aizpurua et al. 2018) and food resources for bats and other species (Sierro et al. 88 89 2001; Mata et al. 2016; Aizpurua et al. 2018). Although nocturnal insects are often captured in 90 passive traps aimed at sampling whole insect communities, these are often biased against some 91 taxonomic groups. For instance, Malaise traps in combination with DNA barcoding or metabarcoding 92 are increasingly used to survey insect communities worldwide (de Waard et al. 2019), but they are 93 mainly effective at collecting Diptera and Hymenoptera (Matthews & Matthews 1971), and thus can 94 underrepresent major nocturnal species groups such as moths (Lepidoptera). To overcome these 95 limitations, studying nocturnal insect communities requires targeted sampling devices, which in the 96 case of flying species normally involve light trapping combined with flight interception (Young 2005; 97 Häuser and Riede 2015). The variety of light traps available is very large, ranging from commercial to 98 customised models, and from models that are operated manually to automatic models with triggers 99 that switch them on and off at specific times (Young 2005; Häuser and Riede 2015). Typically, a light 100 trap can collect hundreds to thousands of individuals in a single night, and so DNA metabarcoding 101 could help speed up and increase the taxonomic resolution of sample processing (Zenker et al. 2020). 102 However, efforts are still needed to develop and optimise traps that can be efficiently combined with 103 metabarcoding in large scale field surveys. Methods that can avoid the need of hand-picking 104 individual specimens, thus reducing effort and contamination risks, and the use of chemical 105 compounds to kill or otherwise retain insects within traps, which would make the subsequent steps 106 of DNA extraction and amplification more difficult (Dillon et al. 1996; Ballare 2019), would be 107 particularly useful.

In this study, we describe a methodological pipeline to study nocturnal flying insects, which combines an automatic light trap device and DNA metabarcoding. Specifically, the study aims to: (i) describe the light trap and its operation; (ii) determine its capacity in sampling a high diversity of insects in short periods of time; and (iii) test whether the molecular procedures yield estimates of species richness and composition comparable to those obtained using conventional morphological identification. Overall, the study shows the value of our new approach to facilitate the sampling of highly diverse communities of nocturnal flying insects in a short time.

115

116 Materials and Methods

117 Study design

118 We conducted field testing of the customised light trap in July 2017, within a protected area in northeastern Portugal (Parque Natural Regional do Vale do Tua; 41.33 N, 7.35 W). We collected a total of 119 120 12 arthropod bulk samples, by setting 6 light traps in each of two nights in two areas of cork oak 121 woodlands, which were expected to yield high insect diversity. The light traps were spread out so that they were not visible from one trap to the other. Evaluation of metabarcoding results used the 122 123 12 field samples, involving comparisons of species richness and composition estimated through 124 molecular procedures versus conventional morphotaxonomy. We focused on species of moths 125 (Lepidoptera) because this is a taxonomically and functionally highly diverse and relatively wellknown group in the country (Corley 2015), there was a highly experienced taxonomist (MFVC) 126 127 available to undertake the field identifications, and a comprehensive library of DNA barcodes was already available for a large proportion of moth species occurring in the region. Comparisons 128 129 between morphological and molecular results involved three different approaches to produce the 130 bulk samples, each representing a particular study design and objectives: (i) 'individuals' – bulk 131 sample for each site produced using one individual per species present in the field sample, which can 132 be used for studies targeting a single taxonomic group, and thus where metabarcoding of the entire 133 bulk sample is unnecessary and may eventually introduce biases; (ii) 'legs' – bulk sample similar to (i), 134 but including only one leg from each individual instead of the entire specimen, which may be used when the individuals need to be preserved for other analysis or PCR inhibitors are likely to be present 135 in digestive tract or other tissues; and (iii) 'mixture' - bulk sample retaining all specimens collected, 136 137 without any sorting and irrespective of taxa. In our case, the approach (iii) excluded the individuals 138 retained for (i) and (ii), and thus did not consider the rarest species (i.e., represented by a single individual in the field sample). 139

140 Light trap design and operation

We adapted a light trap device to allow daily deployment and automatic operation of many traps by a small field team (Figure 1). Each light trap is equipped with one-meter long IP65 3528 UV LED light strip of 395-405 nm, containing 60 LEDs of 0.08 W each, folded in three sections within a 35 cm long transparent plastic tube with 3 cm diameter, and powered by a 12 Ah 12 V lead battery. UV light was used because it is more efficient than other light sources at attracting Lepidoptera (Young 2005), which was the main focus of our study. The strip is connected to a solar light sensor that automatically activates the circuit at sunset and shuts it down at sunrise, enabling the placement of

several traps throughout daytime while ensuring an equal functioning time for each of them

149 throughout the night, avoiding bias regarding flight time activity and unnecessary draining of the 150 battery during the day. The plastic tube is installed in the centre of 3-4 acrylic plates (50 cm x 16.6 cm 151 x 2.5 mm), to intercept flying insects, and on top of a transparent funnel made of rigid 0.75 mm PVC 152 film with a 4 cm diameter opening that leads into a 30 L bucket. Inside the bucket there is a breathable fabric bag containing cardboard egg boxes for the insects to rest and hide. Flying insects 153 attracted to the light collide with the acrylic plates and eventually fall through the tunnel and get 154 155 trapped inside the bucket bag. Each trap is visited in early morning by the field team, usually within 156 one hour after dawn.

157 Sample processing and DNA extraction

158 We designed the sample processing and DNA extraction approaches to speed up processing time, reduce contamination risks and maximise DNA recovery from bulk samples. During the early-morning 159 160 visit to each trap, the bag was removed from the bucket, sealed with a rubber band and frozen for at least three hours at -20 °C to immobilize the collected specimens. Each frozen sample was then 161 thawed at room temperature. All specimens were inspected and morphological identification was 162 163 attempted by a specialist (MFVC). No specimen was dissected to fully validate species identification, 164 as the objective was to compare regular ecological studies based on simple visual identification of live moths with metabarcoding approaches. One specimen of each species identified per sample was 165 166 transferred into a 50 mL falcon tube to constitute the "individual" bulk sample. From these selected 167 specimens, a leg was removed from each and transferred into a second falcon to make up the "leg" 168 bulk sample. Finally, all the remaining specimens (both identified and unidentified) were transferred 169 to a third falcon to constitute the 'mixture' bulk sample. When very large insects (>5 cm body length) 170 were collected, only legs were retained in the bulks to avoid biases in DNA amplification (i.e., 171 overrepresentation of species with high biomass, and masking of rare and low biomass species). The 172 36 falcon tubes (three per sample) were filled with 96% ethanol and stored at room temperature 173 until further processing. Maximum care was taken when handling specimens to avoid cross-174 contamination among samples, namely by thoroughly cleaning all materials used (tweezers and 175 spatulas). Prior to DNA extraction, the ethanol was filtered out and the bulks were dried to constant 176 weight in an incubator at 56 ºC for about 2 days. Afterwards, each 'individual' and 'mixture' bulk 177 sample was homogenized into a fine powder using the Bullet Blender 50-DX homogenizer (Next 178 Advance, USA), with 4 glass beads of 8 mm diameter during 15 min. DNA from 'leg' bulk samples was 179 extracted without further processing. DNA extraction was done using the E.Z.N.A.® Tissue DNA Kit, 180 following an adapted protocol (Supplementary Methods). We performed up to 3 DNA extractions per 181 sample (replicates), each one using 80 to 100 mg of the homogenized insect powder to increase

182 species detection. In one sample (SL6) only one 'individual' and two 'mixture' replicates were

183 possible due to low biomass.

184 Metabarcoding library prep

DNA extracts were diluted 1:100 (after a small test comparing PCR amplification success at different 185 186 dilutions) and amplified using the BF2-BR2 primer pair (Elbrecht and Leese 2017) in three 187 independent reactions. PCR reactions comprised 5 µl of Qiagen Multiplex Master Mix, 0.3 µl of each 188 10 nM primer, 3.4 μ l of H₂O and 1 μ l of diluted DNA. Cycling conditions consisted of initial denaturing 189 at 95 °C for 15 min, followed by 35 cycles of denaturing at 95 °C for 30 s, annealing at 45 °C for 30 s 190 and extension at 72 °C for 30 s, with final elongation at 60 °C for 10 min. The PCR products were 191 tested in 2% agarose gel to check for the amplification success. All PCR products were diluted 1:4 192 with water and further subjected to a second PCR reaction in order to incorporate 7 bp long 193 identification tags and Illumina P5 and P7 adaptors. PCR reactions were similar to that of first PCR, 194 except that 5 uL of Phusion Hot Start Flex 2X Master Mix (NEB) was used, as well as 0.5 uL of each 10 195 nM indexing primer. Cycling conditions consisted of initial denaturing at 98 °C for 2 min, followed by 196 8 cycles of denaturing at 98 °C for 5 s, annealing at 55 °C for 20 s and extension at 72 °C for 20 s, with 197 a final elongation at 72 °C for 1 min. PCR products were purified using Agencourt AMPure XP beads 198 (Beckman Coulter) in a 1:0.8 ratio, quantified using Nanodrop and diluted to 15 nM. Purified and 199 normalized PCR products were further pooled into a single library and quantified using qPCR (KAPA 200 Library Quant Kit gPCR Mix, Bio-Rad iCycler). The final library was diluted to 4 nM and sequenced in 201 an Illumina HiSeq 2500 Platform using a 2x250bp RapidRun kit for an average coverage of 160,000 202 paired reads per PCR product.

203 Bioinformatic pipeline

General sequence processing was carried out using OBITools (Boyer et al. 2016), along with VSEARCH 204 205 (Rognes et al. 2016) and LULU (Frøslev et al. 2017) for denoising. First, paired-end reads were aligned 206 using the command 'illuminapairedend' and discarded if overlapping quality was less than 40. 207 Second, reads were assigned to samples and primer sequences were removed using 'ngsfilter', 208 allowing a total of 4 mismatches to the expected primer sequence. Finally, reads were collapsed into 209 haplotypes using the 'obiuniq' command and singletons (haplotypes with only one read per sample) 210 were removed. The remaining haplotypes left per sample were all joined in a single file and again 211 dereplicated. The '--cluster unoise' VSEARCH command was then used to denoise the dataset by 212 removing spurious sequences resultant from PCR and sequencing errors, followed by the command '-213 -uchime3_denovo' to remove potential chimeric sequences. Afterwards, the remaining sequences were clustered at a 99% similarity threshold and the original reads were mapped back to the 214

remaining haplotypes. Finally, we used LULU to remove genetically similar co-occurring haplotypes,

this way highly reducing the number of mitochondrial nuclear copies present in the final dataset thattend to artificially increase the number of molecular units and taxa.

218 All haplotypes retained after the bioinformatic processing, were identified to the lowest possible 219 taxonomic level, considering both moths (Lepidoptera) and other arthropod groups. The taxonomic 220 assignment of each haplotype to a taxon was done with the support of a neighbour-joining 221 phylogenetic tree based on an alignment of the haplotypes sorted by their read count. This allowed 222 to visually define clusters of haplotypes that corresponded to the same taxon, as well as to identify 223 chimeric sequences and PCR errors that remained in the dataset. Taxa were identified by comparing 224 the representative haplotypes of each cluster against online databases (BOLD and NCBI), as well as 225 unpublished sequences of arthropods collected in northern Portugal (InBIO Barcoding Initiative; e.g., 226 Ferreira et al. 2020). Species level identifications were usually made for similarity values above 98.5% 227 (da Silva et al. 2019), except in rare cases where no other species of the genus are known to exist, 228 and thus genetic divergence reflects local haplotype diversity. Whenever a haplotype matched 229 several species, genus, or families at similar identity levels, we selected the most inclusive taxonomic 230 rank. For example, if a haplotype matched with 95% similarity two species of different genus 231 belonging to the same family, we identified it only to family level. Those that best matched the same 232 taxa were collapsed into a single taxon, as we assumed that they belonged to the same OTU. 233 Haplotypes whose identification was only possible up to family, order or class level were clustered 234 according to their similarity into distinct taxa (e.g., Noctuidae 1, Noctuidae 2, and so on). All 235 identifications were checked for plausibility, considering the species geographical ranges and 236 seasonal flight periods. Implausible identifications (e.g., species from other biogeographic regions not 237 known to occur in the Iberian Peninsula), were reviewed and moved if needed to a higher taxonomic 238 rank (e.g., from species to genus level).

239 Statistical analysis

240 We compared both Lepidoptera richness and composition between morphological identification and 241 metabarcoding approaches for each type of bulk sample. For Lepidoptera richness we applied a 242 generalized linear mixed model (GLMM) with a gaussian distribution using the function 'Imer', and 243 with light trap as random variable. Statistical significance (p < 0.05) of the model was tested with a 244 likelihood ratio test using the function 'anova'. To compare the species composition between the 245 different treatments we first calculated a Jaccard distance matrix based on presence/absence records of Lepidoptera taxa in each sample, and then performed a permutational multivariate 246 247 analysis of variance (PerMANOVA) using the function 'adonis'. All statistical analyses were done in R

- version 3.5.2 (R Core Team 2018) using packages Ime4 (Bates et al. 2015) and vegan (Oksanen et al.
- 249 **2018**).
- 250 **Results**

251 Trap operation and morphological identification

252 All traps operated as planned, showing their effectiveness at collecting a large number of nocturnal 253 flying insects. Sorting of the bulk samples to extract Lepidoptera for subsequent comparisons with 254 metabarcoding yielded a total of 871 identifiable individuals (average ± SD of 72.6 ± 23.0 individuals 255 per sample), which were assigned to a minimum of 102 species (26.8 ± 6.7 species per sample) 256 (Supplementary Table S1). Some specimens could not be identified visually as wing scales were 257 largely absent and coloration patterns were no longer observable. A total of seven species accounted 258 for 52.4% of the individuals identified, while 41 species were represented by a single individual. The 259 individuals identified were then used to produce the bulk samples for comparison with 260 metabarcoding. All 102 species were considered in the 'individuals' and 'legs' bulk samples, while the 261 'mixture' bulk samples included only 40 recognizable species $(11.9 \pm 3.6 \text{ species per sample})$ after 262 removing the individuals to produce the former two bulks. However, in the 'mixture' bulk there were 263 also the unidentifiable individuals, which were too damaged or otherwise lacked diagnostic features

264 required for species level identification.

265 Metabarcoding results

276

Considering the overall results for the three bulk types ('individuals', 'legs' and 'mixture'), sequencing
of libraries generated a total of 47,764,194 paired reads, with an average (±SD) of 151,632 ± 81,181
per PCR product. After bioinformatic filtering, we retained a total of 14,484,254 paired reads, with an

- 269 average of 46,573 ± 30,371 per PCR product.
- In the mixture bulk sample, we detected a total of 528 arthropod taxa (124.5 ± 33.3 taxa per sample), of which 61.2% were identified to species, 10.6% to genus, and 28.2% to family or higher taxonomic ranks. Most taxa detected belonged to Lepidoptera (31.1%), Diptera (26.1%) and 42.8% to 14 other orders (Figure 2). In the case of Lepidoptera, which were the main focus of this study, we detected a total of 189 taxa (55.2 ± 13.7 taxa per sample), of which 111 taxa (26.8 ± 8.1 taxa per sample) were represented in the 'individuals' bulk, 106 taxa (26.3 ± 6.7 taxa per sample) in the 'legs' bulk, and 164

taxa (45.8 ± 12.7 taxa per sample) in the 'mixture' bulk sample (Supplementary Table S1).

277 Comparisons between morphological identification and metabarcoding

278 Considering the Lepidoptera, comparisons between morphological and metabarcoding results were 279 largely similar for the 'individuals' and 'legs' bulks, both in terms of species richness per sample (χ^2 = 280 1.3838, df = 2, p = 0.5006) and species composition (pseudo-F = 0.8399, df = 2, $R^2 = 0.0484$, p = 0.7373). The overall number of species detected was slightly higher for metabarcoding than for 281 282 morphology, both for 'individuals' (111 vs. 102) and 'legs' (106 vs. 102) bulks. The mean percentage 283 of species detected by both methods per sample was similar for the 'individuals' (64.6%±6.2%, 54.8-284 75.0%) and 'legs' (67.4%±6.2%, 59.4-77.4%) bulks, with similar values also for the mean percentage 285 of species detected by morphology but not by metabarcoding ('individuals': 18.3%±5.0%, 11.4-29.0%; 286 'legs': 17.3%±3.6%, 11.8-21.9%), and vice versa ('individuals': 17.1%±5.0%, 11.1-28.1%; 'legs': 287 15.3%±3.2%, 9.7-20.0%). When considering only families normally assigned to macrolepidoptera 288 (e.g., Erebidae, Geometridae, Noctuidae), the percentage of species shared between methods 289 increased for both the 'individuals' (72.3%±10.5%, 54.5-86.4%) and 'legs' (74.9%±10.4%, 60.0-90.5%) 290 bulks, while in microlepidoptera it reduced in both cases ('individuals': 55.4%±14.9%, 31.3-85.7%; 291 'legs': 58.7%±16.0%, 38.9-100.0%). Many of the mismatches observed were thus caused by 292 microlepidoptera species, mainly due to species of the same genus identified differently by 293 morphological and molecular methods, to taxa that could be identified unambiguously to species 294 level by one method but not the other, and to more species of the same genus being detected by 295 one of the methods. For instance, while morphology detected Apatema mediopallidum, 296 metabarcoding detected only two unidentified Apatema species that did not match the DNA barcode 297 of A. mediopallidum currently available in the databases. Likewise, morphological examination 298 detected Scopula submutata and Scythris parafuscoaenea, while metabarcoding detected Scopula 299 marginepunctata and Scythris dissimilella, respectively. Examples of differences in identification 300 resolution included Eurodachtha canigella detected by morphology and Eurodachtha 301 siculella/canigella by metabarcoding, while morphology detected an unidentified Depressaria while 302 metabarcoding detected *Depressaria adustatella*. Finally, common mismatches occurred with species belonging to the genus *Idaea* and the family Brachodidae, with morphology recording *Idaea* 303 304 belemiata, Brachodes gaditana and Agonopterix scopariella, and metabarcoding detecting I. 305 obsoletaria instead of belemiata, Brachodes gaditana plus B. funebris, and A. atomella instead of A. 306 scopariella.

307 In comparisons involving the 'mixture' bulk sample, morphology and metabarcoding recovered a

- 308 significantly different community of moths. Not only the average species richness identified per
- 309 sample was different (χ^2 = 111.46, df = 1, p < 0.0001), but also the species composition (pseudo-F =
- 6.0464, df = 1, R² = 0.2156, p = 0.001). The percentage of species detected by both methods was only
- 311 20.9%±5.5% (11.3-28.1%), with 74.6%±6.7% (64.3-83.0%) of species detected through metabarcoding

but not morphology, while only 4.6%±2.1% (1.9-9.5%) of species were detected by morphology but
not by metabarcoding. As for the other two bulks, the percentage of species detected by both
methods was higher for macrolepidoptera (34.6%±14.6%, 10.0-58.3%) vs. microlepidoptera
(12.8%±7.5%, 5.0-31.6%).

316 **Discussion**

317 Our study provides an effective pipeline for large scale sampling of nocturnal flying insects, from field 318 sampling using a customised trap design, through lab processing of samples, to bioinformatic analysis 319 of sequencing data. The trap we customized proved to be successful at collecting a large number of 320 nocturnal insects, from which bulk samples for subsequent metabarcoding processing can be easily 321 obtained with minimal risks of contamination. We found that metabarcoding and morphological identification provide similar results when comparing bulk samples with relatively few individuals of 322 323 known species, but that metabarcoding detects far more species when dealing with complex 324 mixtures (real samples) that include many individuals that cannot be easily identified through 325 morphological examination (e.g., due to specimen's age, small size and scale damage, this last one 326 probably accentuated by the transport to a freezer in a bag). Overall, we suggest that our 327 methodological pipeline can be widely applied in ecological studies, contributing to improve our 328 understanding of the trends and drivers of nocturnal insect communities and complement studies 329 based on Malaise traps that are not so efficient in capturing Lepidoptera and other insects like 330 Coleoptera, Mantodea, Neuroptera, Orthoptera and Hemiptera (Matthews & Matthews 1971).

331 The trap customized in our study shares many similarities to others used elsewhere (Zenke et al. 332 2020), including traps that are available commercially. However, it has the advantage of being 333 relatively cheap and easy to produce, which makes it suitable for large scale field studies where 334 resources are limited, and a large number of sites need to be sampled. Furthermore, the trap was 335 adapted to retain insects without the need to use any chemical product to kill the individuals, which likely facilitates the subsequent steps of DNA extraction and amplification (Dillon et al. 1996; Ballare 336 337 2019). Finally, the use of a breathable fabric bag makes it easier to extract insects from the trap and 338 to undertake the initial processing steps in the field with minimal handling of samples, thereby 339 reducing the risks of contamination across samples. Despite these advantages, however, it is 340 necessary to bear in mind some limitations and potential shortcomings, which are similar to those 341 already described for light traps in general. For instance, light traps only sample part of the nocturnal 342 flying insect community, as many species are not attracted to light and/or have no nocturnal activity and thus cannot be represented in the samples (Young 2005). Also, the area effectively sampled by a 343 344 light trap (i.e., trapping radius) may be conditional on vegetation density affecting penetration of the

345 light source (Bowden 1982), with traps set in areas with higher visibility (e.g., grasslands) possibly 346 attracting individuals from farther away than those set in more cluttered habitats (e.g., forests and 347 shrublands). Therefore, when designing a study based on light traps such as ours, researchers need 348 to be aware of these and other potential constrains, using best practices developed from previous 349 studies to control or correct eventual sampling biases (Bowden 1982, Young 2005).

350 The comparisons of species composition between morphological identification and metabarcoding 351 revealed a reasonable matching when using the 'individuals' and 'legs' bulk samples. Nevertheless, 352 about 35% of the species occurrences were non-congruent between methods, likely due to 353 limitations of species identification using morphology, metabarcoding, or both. A significant part of 354 the mismatches was probably due to the difficulties of morphological identification of small 355 microlepidoptera species, as species occurrences detected with both methods raised markedly from 356 micro- to macrolepidoptera. Because of these difficulties, morphological examination sometimes 357 assigned an individual to a species while metabarcoding recorded another similar species from the 358 same genus. In a few cases, mismatches involved morphology assigning several similar individuals to 359 the same species, while metabarcoding revealed that a few of these belonged to a second or even a 360 third species. These results point out the limitations of using morphological examination as the 361 benchmark for metabarcoding validation, even in cases such as ours where the fauna studied was reasonably well known (Corley 2015), and a very experienced specialist was involved in the study (M. 362 363 F. V. Corley). This is in line with the results of recent efforts to DNA barcode the Portuguese moth 364 fauna, which has revealed several new species, some of which had been previously misidentified 365 (e.g., Corley & Ferreira 2019, Corley et al. 2019). In a few cases, mismatches were associated with 366 limitations of metabarcoding, with the markers used being unable to discriminate between closely 367 related species, while morphology readily provided species-level identifications. Metabarcoding also 368 failed to identify species for which there were no DNA barcodes available at the time of the study, 369 which is a general problem affecting metabarcoding studies. Finally, metabarcoding possibly 370 produced a few false positives, such as Agrotis bigramma, Leucochlaena oditis and Luperina 371 Testacea, as these species are known to fly only in late summer and early autumn while our samples 372 were collected in July. All these species were represented by just a few reads, and possibly resulted 373 from the assignment of reads to incorrect samples (cross-talk) (e.g., Edgar 2018), as our samples 374 were run together with light trap samples collected in September. Overall, therefore, these 375 comparisons suggest that metabarcoding provided an accurate description of the communities 376 represented in the 'individuals' and 'legs' bulk samples, with most errors resulting from problems of 377 unresolved taxonomy or incomplete databases, though there might have been errors associated with 378 metabarcoding itself.

379 Comparisons between morphology and metabarcoding involving the 'mixture' bulk sample provided 380 the lowest congruence levels in species occurrences. Errors were mainly due to metabarcoding 381 detecting a large number of species not recorded through morphology, while only a very few species 382 were detected through morphology but not metabarcoding. These results are likely a consequence of 383 the 'mixture' bulks retaining many small individuals from unidentified species, particularly 384 microlepidoptera, which contributed to increase taxonomic diversity. Furthermore, there were 385 probably body parts retained in the mixture that could not be detected, let alone identified, and that 386 further contributed to increase diversity. Finally, some species could have been detected through the 387 stomach contents of carnivorous arthropods (Sheppard et al. 2005; da Silva et al. 2019), and not 388 directly through individuals collected in the trap. Additionally, although eventual problems of lab and 389 field contamination inflating numbers of taxa could not be totally discarded, we believe that these 390 should have had little influence in the results, as this problem was not detected for the 'individuals' 391 and 'legs' bulks. Therefore, it is unlikely that the higher number of taxa recorded through 392 metabarcoding in 'mixture' bulks were false positives, except in the few cases involving cross talk 393 errors. Instead, the higher number of taxa likely reflected the much-increased sensitivity of 394 metabarcoding to estimate species diversity in relation to the conventional morphological approach. 395 This result also suggests that metabarcoding validation studies using only artificial mock communities 396 may be underestimating the true power of the technique to reveal the diversity of real communities, 397 and so the use in validation of actual field samples is highly recommended.

398 **Conclusions**

399 The customised light trap described in this study in combination with DNA metabarcoding offers a 400 relatively simple and cost-effective approach to describe communities of nocturnal flying insects in 401 large scale studies. Although like other approaches based on metabarcoding it cannot provide 402 information on species abundances (Elbrecht & Leese 2015; Piñol et al. 2015), it offers the ability to 403 process hundreds to thousands of samples at high taxonomic resolution in a relatively short time 404 frame, which can hardly be achieved through conventional morphological approaches (Ji et al. 2013). 405 A key advantage of our approach is that a large number of insect specimens can be collected at a 406 number of sites in a single night by a small field team (Zenke et al. 2020), which can be particularly 407 advantageous if the study is carried out in remote areas, while other sampling devices such as 408 Malaise traps are usually more labour intensive and require deployment in the field for much longer 409 periods (e.g., Matthews & Matthews 1971; Häuser & Riede 2015). Like any sampling device, 410 however, our method is biased towards some species groups such as flying Lepidoptera, Diptera and Coleoptera that are attracted to light, and thus a combination of sampling methods should be used in 411 412 studies requiring a comprehensive account of entire insect communities (e.g., Yang et al. 2014,

413 Marquina et al. 2019). To further enhance the value of our approach, it would be important to 414 increase the DNA barcode reference databases for hyper-diverse insect groups such as Diptera (e.g., 415 Ferreira et al 2020), to apply bioinformatic procedures correcting for cross talk problems causing 416 false positives in metabarcoding (e.g., Edgar 2018), to standardise procedures that may affect catchability of different insect groups such as for instance the light source and intensity, and the 417 weather and moonlight conditions (Young 2005), and to further assess potential errors and 418 419 limitations such as variations in effective catching distance across habitats with different vegetation 420 cluttering (e.g., forests with or without undergrowth; Young 2005). Overall, we suggest that the 421 methodological pipeline proposed in this study provides a convenient approach for the standardised 422 monitoring of spatio-temporal variations in the diversity and composition of complex nocturnal 423 insect communities, with the potential to help addressing pressing societal challenges such as global

424 insect declines.

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FIGURES



Figure 1. Customised UV light traps designed to collect night flying insects, showing a trap ready to operate in the field (A), a schematic representation of the assembly of all the pieces making up the complete trapping system (B), and the electric components of the trap and how to connect them (C).



579 **Figure 2.** Taxa richness detected through metabarcoding in the 'mixture' bulk samples. Grey bars

represent the total number of taxa detected per taxonomic order in the sum of all 12 samples.

581 Whiskers represent the average \pm sd number of taxa detected per sample.