

1 **Particle-associated and free-living bacterial communities in an oligotrophic sea**
2 **are affected by different environmental and anthropogenic factors**

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4 Dalit Roth Rosenberg¹, Markus Haber¹, Joshua Goldford², Maya Lalar³, Dikla Aharonovich¹, Ashraf Al-
5 Ashhab^{1,4}, Yoav Lehahn⁵, Michael D. Krom^{1,6}, Daniel Segrè², Laura Steindler¹ and Daniel Sher¹

6 ¹Department of Marine Biology, Leon H. Charney School of Marine Sciences, University of Haifa, Israel; ²

7 Bioinformatics Program, Boston University, USA; ³ Bioinformatics Support Unit, University of Haifa,

8 Israel; ⁴Microbial Metagenomics Division, Dead Sea and Arava Science Center, Masada, Israel;

9 ⁵Department of Maritime Geosciences, Leon H. Charney School of Marine Sciences, University of Haifa,

10 Israel. ⁶Morris Kahn Marine Research Station, Department of Marine Biology, Leon H. Charney School of

11 Marine Science, University of Haifa, Haifa, 3498838, Israel.

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13 Corresponding author: Daniel Sher, dsher@univ.haifa.ac.il

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15 **Originality – Significance Statement (need to take from summary)**

16

17 **Summary**

18 In the oceans and seas, environmental conditions change over multiple temporal and spatial scales.

19 Here, we ask what factors affect the bacterial community structure across time, depth and size fraction

20 during six seasonal cruises (two years) in the ultra-oligotrophic Eastern Mediterranean Sea. The bacterial
21 community varied most between size fractions (free-living vs particle-associated), followed by depth and
22 finally season. The free-living (FL) community was taxonomically richer and more stable than the
23 particle-associated (PA) one, which was characterized by recurrent “blooms” of heterotrophic bacteria
24 such as *Alteromonas* and *Ralstonia*. The heterotrophic FL and PA communities were also correlated with
25 different environmental parameters: depth and phytoplankton correlated with the FL population,
26 whereas PA bacteria were correlated primarily with season. A significant part of the variability in
27 community structure could not, however, be explained by the measured environmental parameters. The
28 metabolic potential of the PA community, predicted from 16S amplicon data, was enriched in pathways
29 associated with the degradation and utilization of biological macromolecules, as well as plastics, other
30 petroleum products and herbicides. The FL community was enriched in pathways for the metabolism of
31 inositol phosphate, a potential phosphorus source, and of polycyclic aromatic hydrocarbons.

32 **Originality – Significance Statement**

33 Marine microbial populations are complex and dynamic, and the environmental drivers of the structure
34 and function of these communities are mostly unclear. Specifically, marine microbial communities
35 change over time, over depth and between particle-associated and free-living size fractions, yet the
36 relative importance of each of these axes of variability is unclear. Our results highlight fundamentally
37 different population dynamics between free-living and particle-associated marine bacteria: free living
38 populations were more similar between seasons, whereas particle-associated populations were highly
39 variable and exhibited “blooms” of specific clades of heterotrophic bacteria. We also suggest that the
40 environmental conditions often measured as part of oceanographic cruises are not enough to explain
41 most of the variability in microbial population structure. We speculate that organismal interactions and

42 the presence of anthropogenic pollution may be also be important yet under-sampled drivers of
43 oligotrophic marine microbial communities.

44

45 **Keywords:** Eastern Mediterranean, 16S amplicon sequencing, Particle-associated, Free-living,
46 Seasonality, anthropogenic pollution, phytoplankton

47

48 **Running title:** Drivers of bacterial populations across scales

49

50 **Introduction**

51 To a microorganism, the marine environment presents a rich and ever-changing tapestry of conditions
52 and potential niches. On scales of microns to millimeters, which can be traversed by motile bacteria
53 within minutes, various types of particles and gels provide carbon- and nutrient-rich hotspots which
54 differ from the bulk surrounding seawater in their chemistry and physics (Azam and Malfatti 2007).
55 Other environmental parameters such as light intensity, temperature and the concentrations of
56 dissolved inorganic and organic nutrients change markedly with depth over scales of tens to hundreds of
57 meters, particularly when the water column is stratified (Karl 2007). Depending on the structure of the
58 water column and the vertical movement of macro-organisms and sinking particles, bacteria can
59 traverse such distances (both up and down) on scales of hours to days (e.g. (Grossart et al. 2010, Mestre
60 et al. 2018)). Finally, changes in environmental conditions can occur over spatial scales of hundreds of
61 kilometers or temporal scales of weeks to months. For example, seasonal changes in weather or
62 geographic changes in environmental conditions are driven by global climate patterns (Giovannoni and

63 Vergin 2012). Seasons affect the physical structure of the water column and, through mixing of water
64 masses, control the injection of inorganic nutrients into the photic zone. Indeed, microbial populations
65 in the oceans differ over multiple spatial and temporal scales (Fuhrman 2009, Martin-Platero et al. 2018,
66 Salazar et al. 2019). While our understanding of the processes that shape marine microbial communities
67 over these scales has been steadily increasing (Karl et al. 2002), many questions remain unanswered. For
68 example, changes in microbial population structure have been documented between different size
69 fractions (e.g. (Mestre et al. 2017)), along a depth gradient (DeLong et al. 2006) and over seasonal cycles
70 (Ward et al. 2017), yet few studies have addressed these spatio-temporal axes of variability together.
71 Which of these environmental factors has a more pronounced effect on microbial populations? Are
72 changes in heterotrophic microbial populations driven primarily by variation in “a-biotic” conditions, or
73 do other factors affect population structure, for example biotic interactions with co-occurring
74 phytoplankton? Can we predict which genetic traits or metabolic functions underlie changes in
75 community structure along these gradients of variability?

76 To begin addressing some of these questions, we characterized the environmental conditions, the
77 phytoplankton community, and the bacterial community structure across multiple spatio-temporal
78 scales in the Levantine Basin of the Eastern Mediterranean Sea (EMS). Despite being an inland sea,
79 surrounded by ~480 million people (Bleu 2009), the open waters of the Mediterranean, and particularly
80 the EMS, are oligotrophic to ultra-oligotrophic (Fig. 1A, (Berman et al. 1985)). Nutrient concentrations in
81 the photic zone are typically very low (close to the level of detection), as are chlorophyll and particulate
82 carbon, and the photic zone may extend quite deep, with a dominant Deep Chlorophyll Maximum
83 (DCM) often observed as deep as ~140m (Berman et al. 1985, Krom et al. 2005). Phytoplankton in the
84 EMS are often phosphorus-limited or nitrogen and phosphorus co-limited (Krom et al. 2005, Thingstad
85 et al. 2005). Pico-cyanobacteria such as *Prochlorococcus* and *Synechococcus* are the numerically
86 dominant phytoplankton, although photosynthetic pico-eukaryotes are also common and potentially

87 dominant in terms of production or biomass (Man-Aharonovich et al. 2010). The deep waters of the EMS
88 are much warmer than those of many oceanic regions, with the temperature never dropping below
89 $\sim 12^{\circ}\text{C}$. The deep waters are mixed into Levantine Intermediate waters and these combined waters exit
90 the Eastern Mediterranean (EMS) at the Straits of Sicily, eventually exiting the Mediterranean through
91 the straits of Gibraltar and affecting large parts of the Atlantic Ocean. Recent studies have shown that
92 several water masses of the EMS are warming and becoming more saline at a rate significantly higher
93 than the average global trend predicted by the IPCC (Ozer et al. 2017). Thus, while the EMS exhibits
94 conditions reminiscent of those in the major oceanic gyres (Powley et al. 2017), it is a potentially useful
95 “natural laboratory” to understand the effect of environmental conditions, including climate change, on
96 microbial processes.

97 To obtain a dynamic view of how environmental parameters at multiple scales affect the microbial
98 population structure, we present and interpret physical, chemical and biological information from six
99 cruises to a pelagic station in the EMS. Station n-1200 is ~ 25 nautical miles from shore, where the water
100 column depth is 1200m. The cruises were performed during three seasons (fall, spring and summer) of
101 two consecutive years. Samples were collected from five discrete depths (surface, $\frac{1}{2}$ DCM, DCM, 200m
102 and 500m) and separated into different size fractions, $>11\mu\text{m}$ and $5-0.22\mu\text{m}$, representing particle-
103 associated and free-living microbes, respectively. These data allow us to describe and interpret the
104 natural changes in microbial populations in the EMS over different scales: (i) particle associated vs. free-
105 living, (ii) over depth, and (iii) over seasons.

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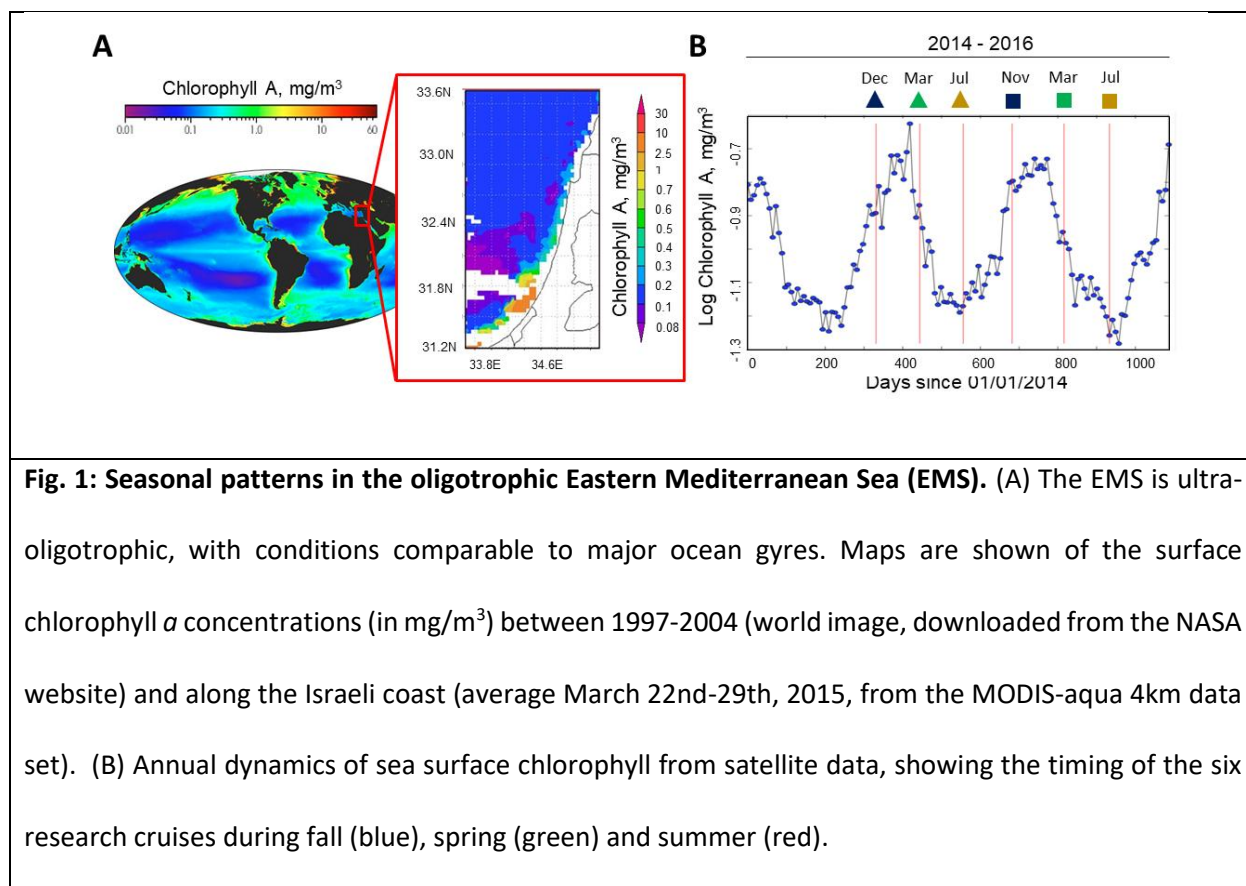
107 **Results and discussion**

108 **Physical, chemical and biological properties of the EMS during the cruises**

109 During 2014-2016, a clear seasonal cycle in sea surface Chlorophyll was observed in the EMS, with
110 maxima in winter and early spring and minima in summer, consistent with previous *in-situ*
111 measurements (Fig. 1B, (Azov 1986, Raveh et al. 2015)). Our sampling cruises were timed to coincide
112 with the increase in surface chlorophyll during fall, the subsequent decrease phase during spring, and
113 the stable, ultra-oligotrophic summer conditions. The changes in surface chlorophyll were associated
114 with changes in the structure of the water column, which was highly stratified during summer, with the
115 depth of mixing increasing during fall and peaking during early spring (Supporting Information Fig. S1).
116 Macronutrient profiles show that the surface waters, down to ~200m, were poor in nitrate+nitrite and
117 phosphorus with the lowest values in July compared to fall and spring (Supporting Information Fig. S2). A
118 deep chlorophyll maximum (DCM) was observed in all cruises apart from the November 2015 one, and
119 was generally shallower during fall and spring (70-90m) and deeper during the summer (90-140m,
120 Supporting Information Fig. S1).

121 While clear seasonal patterns were observed in the surface chlorophyll, more complicated patterns
122 were observed in the composition and relative abundance of different phytoplankton groups within the
123 photic zone. Based on flow cytometry counts, during most of the year, *Prochlorococcus* were the
124 numerically most abundant phytoplankton, particularly deeper in the water column ($\frac{1}{2}$ DCM and DCM,
125 Supporting Information Fig. S3). *Synechococcus* were usually more abundant at the surface (10m and
126 occasionally $\frac{1}{2}$ DCM). Consistent with the flow cytometry counts, divinyl-chlorophyll *a* (a pigment which
127 is unique to *Prochlorococcus*) peaked at the DCM, yet never comprised more than ~12.5% of the total
128 chlorophyll concentration, suggesting that other phytoplankton may be more dominant in terms of
129 biomass (Supporting Information Fig. S4). Indeed, 19'-hex-fucoxanthin ("19-hex"), a pigment
130 characteristic of prymnesiophytes (Jeffrey et al. 2005), was relatively abundant at the DCM, particularly
131 during the first year. While 19'- but – fucoxanthin ("but") and fucoxanthin ("fuco") were also observed,
132 we cannot conclude that diatoms were present at high abundances, as these pigments are also found in

133 prymnesiophytes. Peridinin, which is characteristic of dinoflagellates, was observed mainly in the fall of
134 both years and the spring of the second year, at the $\frac{1}{2}$ DCM and DCM depths. These results are
135 consistent with previous studies suggesting that pico-eukaryotes, and particularly coccolithophores
136 (which are prymnesiophytes), are an important part of the phytoplankton community, which is dynamic
137 across time and depth (Man-Aharonovich et al. 2010). These results also suggest that non-seasonal
138 (potentially aperiodic) dynamics may influence phytoplankton dynamics (Karl et al. 2002).



139

140 **Overview of the changes in bacterial population structure across size fraction, depth and** 141 **season**

142 As shown in Figure 2 and Supporting Information Fig. S5, clear differences were observed between the
143 particle-associated bacterial communities (PA, $>11\mu\text{m}$) and the free-living ones (FL, $5\text{-}0.22\mu\text{m}$). Several

144 samples representing the intermediate size fraction (11-5 μ m) were also analyzed, and these revealed a
145 population structure somewhat similar to both PA and FL fractions, in agreement with previous studies
146 suggesting a continuous shift in population structure across different size fractions ((Mestre et al. 2017,
147 Mestre et al. 2017), Figure 2A, B). The FL and PA populations both varied with depth, although the
148 changes with depth in the PA community were less pronounced (see details below). No clear clustering
149 of the populations by season was observed in the NMDS ordination (Figure 2B). Indeed, the Bray Curtis
150 dissimilarity was highest between FL and PA communities, followed by the difference with depth, with
151 the lowest variability observed between seasons (Figure 2C). Thus, in the open EMS, size fraction and
152 depth are likely the most important drivers of microbial community composition.

153 The differences between FL and PA communities were lower at the surface, increasing with depth (Fig.
154 2, Supporting Information Fig. S6A). This is consistent with a mechanism whereby many particles
155 originate in the surface layer, either formed autochthonously or added as atmospherically derived
156 particles (e.g. dust (Herut et al. 2002)). The particles are colonized primarily by surface layer bacteria,
157 with the exchange between FL and PA populations becoming less common with depth (Mestre et al.
158 2018). At a coarse grained phylogenetic resolution, the FL population was dominated primarily by 16S
159 sequences belonging to α -proteobacteria from the SAR11 clade (*Pelagibacter*), which comprised up to
160 ~40% of the sequences in the FL fractions, mostly in the surface waters, (Supporting Information Fig. S5).
161 Other abundant heterotrophic lineages were SAR86 (γ -proteobacteria, up to 15% in surface waters
162 during summer and spring), SAR 406 (Marinimicrobia, up to 8% in deeper waters) and SAR 202
163 (Chloroflexi, primarily intermediate depths, up to 12%). Conversely, the PA communities were rich in γ -
164 proteobacteria and Planctomycetes, as well as occasionally β -proteobacteria, δ -proteobacteria or
165 Verrucomicrobia, phyla previously described as associated with marine particles from the Atlantic
166 (Milici et al. 2016), Baltic (Rieck et al. 2015) and Mediterranean Seas (Mestre et al. 2017, Mestre et al.
167 2017). Notably, while pico-cyanobacteria such as *Prochlorococcus* and *Synechococcus* were primarily

168 observed in the FL fraction, they also comprised as much as ~6% of the 16S reads from the PA
169 communities, and up to ~21% of the communities in the size range of 11-5 μ m (Supporting Information
170 Fig. S5, Supplementary excel file). *Prochlorococcus* and *Synechococcus* are primarily free-living cells,
171 however, previous studies have suggested that pico-cyanobacteria do contribute to particulate sinking
172 fluxes, although the magnitude of this process is unclear, and may depend on the specific oceanographic
173 conditions (Richardson and Jackson 2007, Lomas and Moran 2012, De Martini et al. 2018). The presence
174 of *Prochlorococcus* and *Synechococcus* cells in the PA fractions suggests a potential involvement of these
175 clades in sinking fluxes in the EMS.

176 Within the FL fraction, the population structure clearly changed with depth, with the main differences
177 seen between the communities from the photic zone (here defined as down to 200m) and the
178 mesopelagic (200m and 500m, Fig. 2A, Supporting Information Fig. S5, S6B). The dissimilarity among
179 sampling depths was higher during stratified seasons than during spring, consistent with a partial
180 homogenization of the water column by winter/spring mixing (t-test, $p < 0.001$, Supporting Information
181 Fig. S6B). As expected from other studies, cyanobacteria and Bacetroidetes were relatively more
182 abundant in the photic zone, whereas δ -proteobacteria, Marinimicrobia and Chloroflexi were more
183 associated with deeper samples (Haro-Moreno et al. 2018, Mende et al. 2019). The change of the PA
184 community with sampling depth was weaker, observed in the NMDS analysis (Figure 2) but not in the
185 one-dimensional clustering (Supporting Information Fig. S5), and no clear differences with sampling
186 depth were observed (Supporting Information Fig. S6C). Instead, the PA community seemed to partition
187 into samples dominated by γ -proteobacteria (primarily *Alteromonas*), and those where
188 Planktomycetota, Bacteroidetes or β -Proteobacteria were more dominant (Supplementary Figure 5 and
189 see below).

190 In contrast to the clear effects of particle association and depth on microbial community structure, no
191 obvious partitioning by season could be observed in the NMDS or clustering (Figure 2B, Supporting
192 Information Fig. S5). Nevertheless, seasonal differences were observed when the same size fraction and
193 depth were compared (Supporting Information Fig. S6B, C, D). The seasonal differences were
194 significantly larger in the PA compared to the FL populations, suggesting that qualitatively different
195 particles may be found at different seasons. Surprisingly, within the FL population, seasonal variability
196 was lower at the surface and $\frac{1}{2}$ DCM compared with the deeper samples. We initially expected seasonal
197 differences to be highest at the surface, reflecting major changes in sea surface temperature, which has
198 been suggested to be major driver of microbial community structure and function (Sunagawa et al.
199 2015). The higher inter-seasonal differences detected at intermediate depths (100-200m) suggest that
200 factors other than temperature are causing seasonal changes in the populations at these depths. These
201 may include the quality and intensity of light supporting variable deep phytoplankton populations, or
202 the availability of nutrients mixed up from the nutricline (Karl et al. 2002).

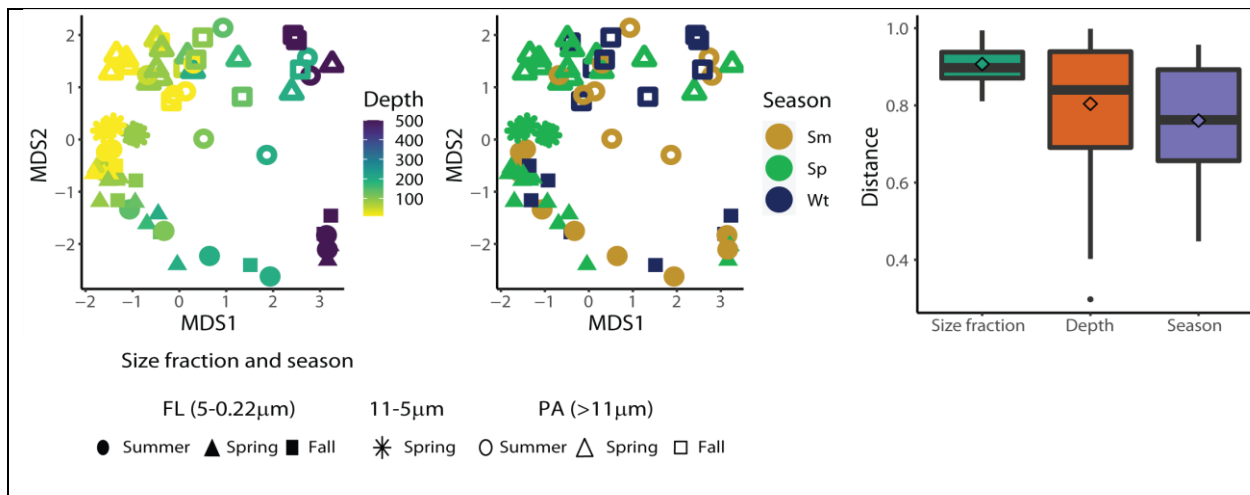


Figure 2: Effects of size fraction, depth and season on bacterial population structure. (A, B)

Nonparametric Multi-Dimensional Scaling (NMDS) plot of the bacterial populations, colored by depth (A) and by season (B). Full shapes are FL, open shapes are PA, with the two groups clearly separated

along NMDS axis 2. Both FL and PA communities change over depth along NMDS axis 2, but do not cluster by season. NMDS stress = 0.11. (C) Bray-Curtis dissimilarity is highest with size fraction, followed by depth and season. The dissimilarity was calculated within groups, e.g. comparisons of depth were performed within the same cruise and size fraction. The means of the size fractions are all statistically different (t-test, $p < 0.001$, (Fagerland 2012)).

203

204 **Particle-associated and free-living communities have different strain-level (ESV) dynamics**

205 A detailed inspection of the community structure (Supporting Information Fig. S5, Supporting excel file)
206 suggests that some PA communities are dominated by a limited number of ESVs (exact sequence
207 variants, which represent distinct 16S phylotypes). Indeed, the FL population was more diverse
208 (Shannon and inverse Simpson indices) and more even than the PA ones (Figure 3A). These results are
209 consistent with studies from major ocean gyres and the offshore western Mediterranean (Ghiglione et
210 al. 2007, Mestre et al. 2018), whereas studies from the North Sea, the Baltic Sea and coastal locations in
211 the western Mediterranean suggest that PA communities are more diverse than FL ones (Bižić-Ionescu
212 et al. 2015, Rieck et al. 2015, Mestre et al. 2017). It is possible that the productivity (trophic status) of
213 the community determines whether FL or PA communities are more diverse, perhaps because more
214 productive regions tend to produce larger particles or larger phytoplankton with potentially more micro-
215 niches. The alpha diversity indices of the FL populations, but not the PA ones, increased with depth
216 throughout the photic zone (Supporting Information Fig. S7). This again suggests that, despite the
217 variability in surface temperature, intermediate depths may provide more niches for FL bacteria than
218 the surface waters.

219 Consistent with the higher α -diversity of the FL populations, the most abundant ESVs comprised at most
220 ~8% of the 16S reads of the FL population, compared with up to 21% of the reads for in the PA

221 community (Fig. 3B). While these differences could, in principle, be due to more 16S genes per genome
222 in the PA bacteria compared with the FL ones, we also observed many more ESVs (each with at least 100
223 reads across our dataset) corresponding to the abundant FL clades compared with the PA ones. For
224 example, there were 192, 71 and 60 different ESVs belonging to the abundant FL clades SAR11, SAR86
225 and SAR406, respectively, compared to 9, 3 and 2 ESVs belonging to the *Alteromonas*, *Oleibacter* and
226 *Ralstonia* clades that dominated specific PA samples.

227 The FL and PA communities also differed qualitatively in the temporal dynamics of the most abundant
228 ESVs. The FL communities were relatively similar across cruises, changing primarily with depth. For
229 example, the ten most relatively abundant SAR11 ESVs were all observed in at least one depth from
230 each cruise, albeit with differing temporal and depth patterns (Fig. 3C). Similarly, the five most
231 fractionally abundant cyanobacterial ESVs were also observed in all of the cruises, and their dynamics
232 over time and depth were consistent with the flow-cytometry counts (Supporting Information Fig. S3)
233 and with ecotype-level dynamics observed in other oligotrophic oceans (e.g. (Malmstrom et al. 2010)).
234 In contrast, many of the most fractionally abundant ESVs in the PA populations were highly dominant in
235 samples from one or two cruises, and rare or almost absent in others. For example, four ESVs, all closely
236 related to *Alteromonas macleodii* (a γ -proteobacterium), were very common in the spring and summer
237 of 2015, together comprising up to 65% of the sequencing reads. These same ESVs were found at much
238 lower relative abundances at other sampling times. Similarly, *Ralstonia* (a β -proteobacterium)
239 dominated intermediate depths in summer and fall of 2015, and *Oleibacter* (also a γ -proteobacterium)
240 dominated during the spring and summer of 2016. In many cases, the same ESVs dominated replicate
241 samples from the same time or samples from different depths of the same cruise (Supporting Excel file),
242 suggesting that these “heterotroph blooms” are not limited to individual particles but rather represent a
243 general feature of the PA community at a specific time and depth. A more detailed description of the

244 dynamics of SAR11, *Prochlorococcus*, *Alteromonas* and *Ralstonia* is presented in the supplementary

245 information.

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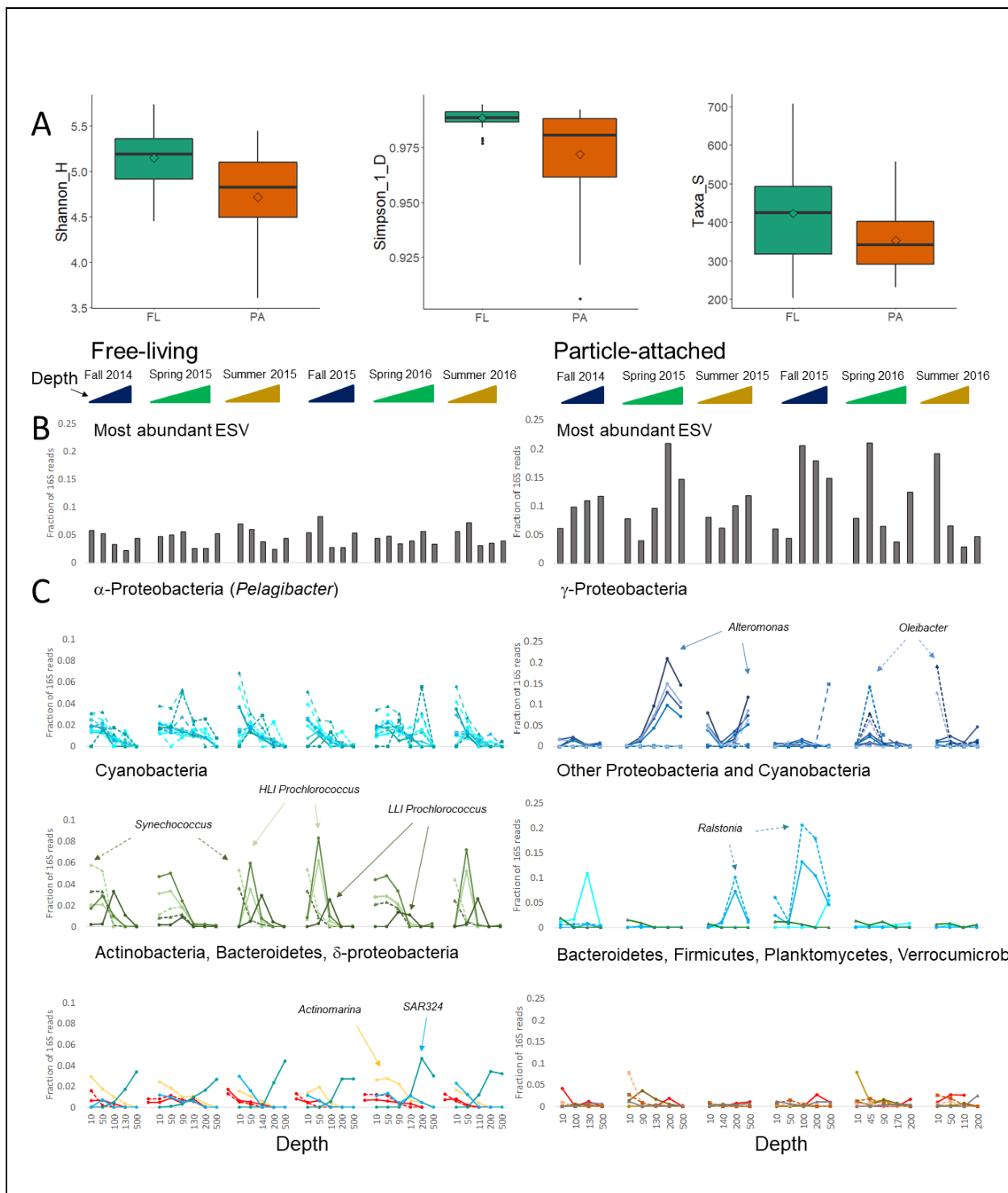


Figure 3: Diversity and ESV dynamics differ between FL and PA communities. In panels B and C, each individual graph shows a depth profile of a separate cruise. (A) FL populations are more diverse than PA ones. Boxes show medians, 25th and 75th quartiles; whiskers show 1.5 times the interquartile range,

and outliers are plotted individually. Diamonds show means. All differences are statistically significant (t-test, $p < 0.001$ for comparisons of Inverse Simpson and Shannon indices and $p = 0.002$ for comparison of the number of taxa). (B) Relative abundance of the most common ESV in each sample (cruise, depth and size-fraction). (C) Relative abundance of the 20 most common ESVs across each of the FL and PA populations. For depths with duplicate samples, the mean of the duplicates is shown. ESVs are colored based on their phylogeny using the same color code as in Supporting Information Fig. 5, and specific ESVs mentioned in the text are highlighted by arrows. For detailed information please see Supporting Information Excel file.

247

248 **Different environmental factors affect the free-living and particle-associated heterotrophic**

249 **populations**

250 The increase in FL bacterial diversity at intermediate depths, and the “heterotroph blooms” observed in
251 the PA community, prompted us to ask whether the heterotrophic bacterial community structure might
252 be determined, at least in part, by interactions with specific phytoplankton groups. Phytoplankton can
253 affect heterotrophic bacteria by providing particulate niches and sources of organic material, as well as
254 through direct signaling (reviewed by (Amin et al. 2012, Buchan et al. 2014, Cirri and Pohnert 2019,
255 Durham et al. 2019)). To answer this question, we sought for statistical correlations between the
256 heterotrophic bacterial community structure and three matrices of conditions corresponding to those
257 associated with seasonality, depth and phytoplankton community structure, with the latter defined by
258 the ratios of the concentrations of photosynthetic pigments (Supporting Information Figure S4,
259 Supporting excel file). As shown in Fig. 4, the matrix of depth-related parameters explained the largest
260 amount of variability in the FL populations (16%), with water depth and the concentration of $\text{NO}_3 + \text{NO}_2$
261 being significant explanatory variables. The matrices related to season and pigments alone explained

262 little variation, however, the combination of the matrices of pigments and depth had significant
263 explanatory power. Of the four photosynthetic pigments significantly correlating with the FL
264 heterotrophic population structure, two are associated with cyanobacteria, and specifically
265 *Prochlorococcus* (DVChl-a, Chl-b), and one with prymnesiophytes (19'-hex). In contrast to the FL
266 heterotrophic community, season-related factors, and specifically the depth of the mixed layer (MLD),
267 were the only statistically significant determinants of PA heterotrophic communities, with a small but
268 significant interaction with phytoplankton community. Thus, the main factors driving the heterotrophic
269 FL and PA community structures were fundamentally different. In total, only 30-43% of the variation in
270 heterotrophic population structure could be explained by correlations with seasonality, depth and/or
271 phytoplankton community structure, suggesting that other environmental factors, which were not
272 measured in this study, are important drivers of heterotrophic bacterial communities.

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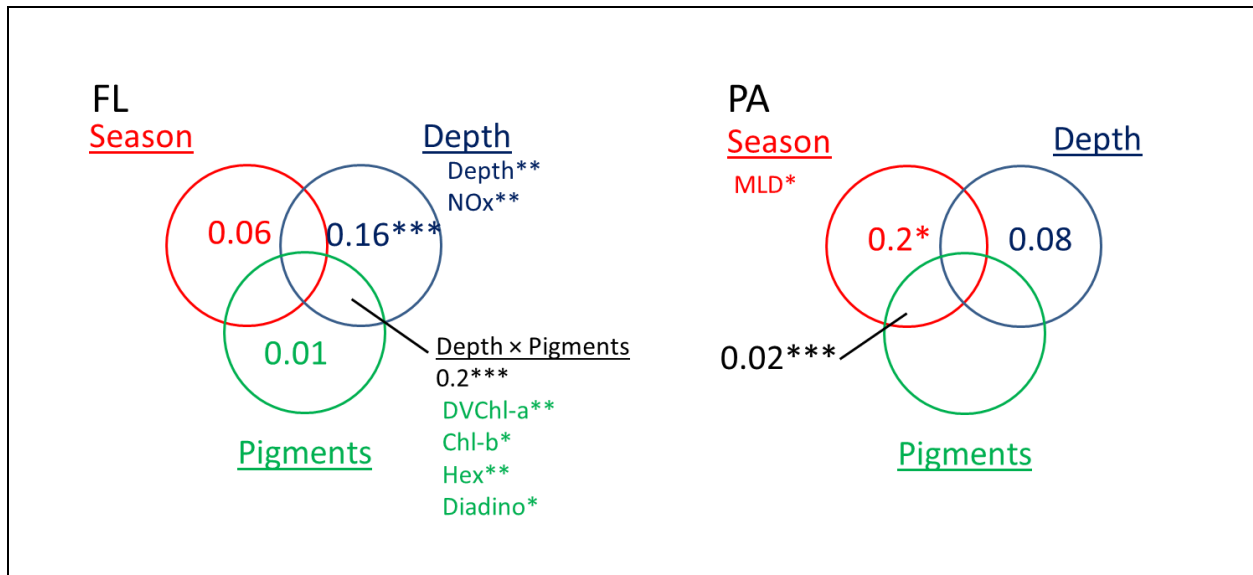


Figure 4: Different environmental factors are correlated with the structure of the PA and FL heterotrophic populations. The results of variation partitioning analysis are shown, with the numbers representing the fraction of the variation explained by each group of environmental parameters. When

no numbers are shown, this matrix had no explanatory power (zero or negative). Stars represent statistical significance: * - <0.05, ** - <0.01, *** - <0.001, as estimated using canonical correspondence analysis. When pigments are shown, it is their ratio to chlorophyll a that is significantly correlated with heterotrophic population structure.

274

275

276 **Predicting metabolic functions that might underlie changes between the FL and PA heterotrophic**

277 **community structure**

278 In this study, we measured multiple environmental parameters known to affect bacterial physiology in
279 the lab, and which could be expected to drive bacterial community structure in the oceans (Supporting
280 Information Figs S1-4). Given that >55% of the variability could not be explained by these factors (Fig. 4),
281 we used the 16S amplicon data itself to raise testable hypotheses as to additional environmental
282 conditions that might drive heterotrophic population structure. We inferred the metagenome of the
283 heterotrophic populations using PICRUSt (Langille et al. 2013), and searched for specific metabolic
284 pathways predicted to be differentially abundant between the heterotrophic PA and FL populations, as
285 the differences between these microbial populations were most clear (Fig. 2, 3). As shown in Fig. 5,
286 many pathways were enriched in the PA compared with the FL heterotrophic populations, consistent
287 with the generally larger genomes and richer metabolic capacities of particle-associated bacteria (Lauro
288 et al. 2009). As expected, many of the pathways enriched in the heterotrophic PA bacteria were related
289 to the degradation of macromolecules that comprise part of phytoplankton cells or zooplankton
290 carapaces (e.g. glycoaminoglycans, starch, lipids, steroids and amino acids and precursors such as
291 benzoate, Fig. 5). Many other enriched pathways, however, were for the degradation of anthropogenic
292 pollutants. These included pathways for the degradation of organic compounds used extensively in the

293 petroleum or plastic industries, such as caprolactam, styrene, xylene, bisphenol A, ethylbenzene and
294 toluene/nitrotoluene. Other pathways were related to the degradation of herbicides, pesticides or
295 compounds involved in their biosynthesis (e.g. fluorobenzoate, atrazine). Plastics (including micro-plastic
296 particles) are ubiquitous in many marine environments (e.g. (Kwon et al. 2015)). They may potentially be
297 important in the dominantly down-welling EMS, in a similar way to their concentration in mid-oceanic
298 gyres (Powley et al. 2017, Chen et al. 2018). Plastics provide surfaces which bacteria can colonize and
299 potentially utilize (Oberbeckmann and Labrenz 2020). Plastics also leach dissolved organic carbon, which
300 may stimulate bacterial growth (Romera-Castillo et al. 2018), but these leachates can also inhibit the
301 growth of key organisms such as pico-phytoplankton (Echeveste et al. 2010, Tetu et al. 2019). Finally,
302 plastics are also thought to adsorb many other contaminants such as pesticides, leading to a local
303 increase in the latter molecule's concentration (e.g. (Rios et al. 2010, Chen et al. 2018)). We speculate
304 that the presence of microplastics may select for PA bacteria able to colonize and utilize these carbon
305 sources, while being resistant to toxic leachates and adsorbed contaminants, thus affecting
306 heterotrophic PA population structure.

307 Only two pathways were enriched in the heterotrophic FL bacteria compared to the PA ones, consistent
308 with their typically smaller genomes, one for the degradation of polycyclic aromatic hydrocarbons
309 (PAHs) and the other for the degradation of inositol phosphates. PAHs may be produced naturally but
310 the anthropogenic input of PAHs is far larger than that from natural sources, and thus we hypothesize
311 that the predicted enrichment of pathways to degrade PAHs in heterotrophic FL bacteria is due to their
312 presence as a result of anthropogenic pollution (Ghosal et al. 2016). It is known that PAH's are present in
313 atmospheric aerosol particles deposited in the EMS particularly from air masses that have passed over
314 polluted areas of southern Europe (Iakovides et al. 2019). Inositol phosphates are a group of ubiquitous
315 signaling molecules in animals, a major form of P storage in plant seeds, and abundant forms of P in
316 some freshwater phytoplankton (Turner et al. 2002). They are found widely in the environment, yet are

317 relatively under-studied and represent a major gap in our understanding of the global P cycle (Turner et
318 al. 2002).

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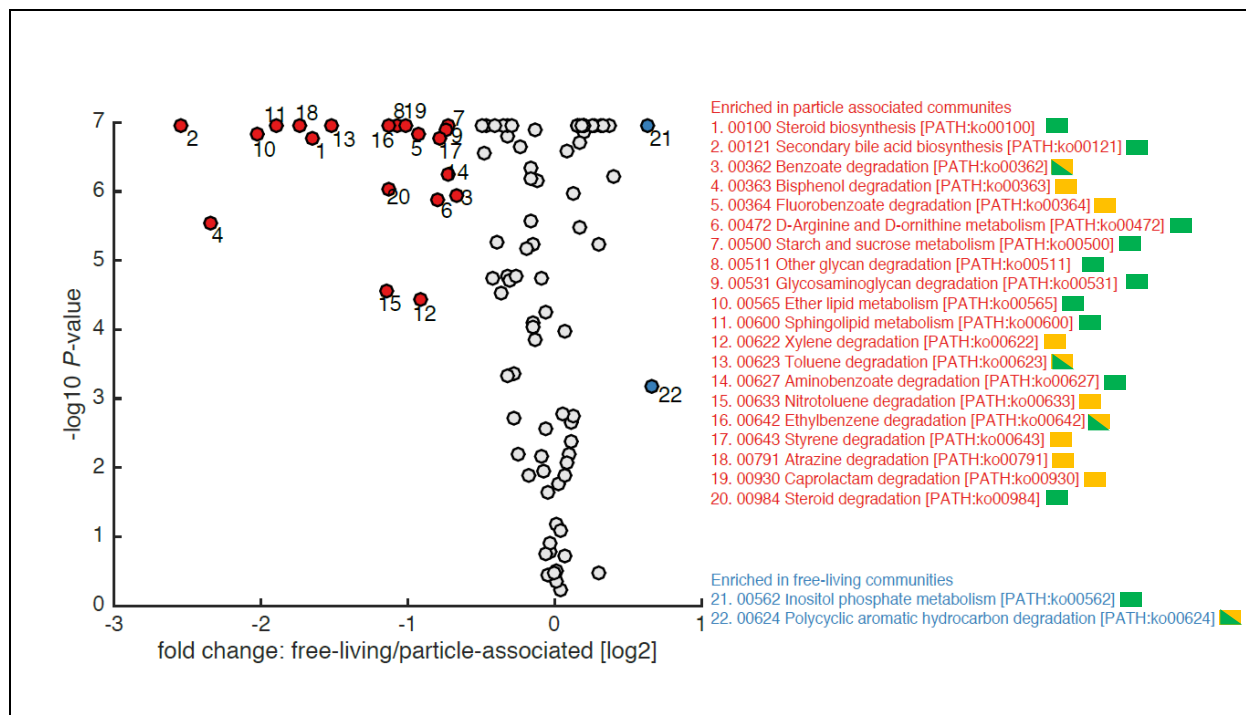


Figure 5: KEGG pathways predicted by PICRUSt to be enriched in the PA (red) and FL (blue) heterotrophic populations. Pathways are ordered by KEGG number and those related to the degradation of animal or plant-derived compounds and of anthropogenic pollutants are marked in green and orange, respectively. Benzoate and ethylbenzene are produced anthropogenically in large amounts but can also occur naturally (in natural petroleum or tar), and thus are marked in both green and orange. Similarly, toluene is used extensively as a thinner and in the petroleum industry, but can also be found in some resins, and is thus marked in both green and orange. The number of pathways enriched in the heterotrophic PA community related to the degradation of anthropogenic pollution, and the number of genes in these pathways, are both statistically higher than those in the full set of KEGG pathways analyzed (Fishers' Exact Test, $p=0.0081$ for number of pathways, $p<0.001$ for number

of genes in the KEGG pathways in PiCRUST). The enrichment for pathways involved in degrading anthropogenic pollutants is specific for the comparison of PA vs FL: Pathways co-varying with depth are not enriched in these functions (Fisher's Exact Test, $p=0.702$), and while there may be some seasonal signal to the enrichment of these pathways, it is not significant (Fisher's Exact Test $p=0.054$).

320

321 **Summary and outlook**

322 Understanding the myriad factors that determine the structure and function of marine microbial
323 communities requires two distinct view points, operating at different scales. At the one end,
324 oceanographic processes, driven by planetary phenomena such as the changing of the seasons,
325 determine the physical structure of the water column (e.g. temperature, stratification). These, in turn,
326 affect the chemical conditions such as the supply and availability of nutrients. At the other end, the
327 physiological traits that dictate where an organism can live operate at scales of nanometers or microns.
328 These include, for example, the affinity for nutrients, the temperature optimum of enzymes or the
329 ability of the organism to swim, perform chemotaxis and colonize particles. While clearly simplistic,
330 these ideas provide a framework with which to interpret and understand the observed patterns in
331 community structure in light of physical and biological drivers (Azam and Malfatti 2007, Karl 2007,
332 Giovannoni and Vergin 2012, Seymour et al. 2017). In this study, we aimed to bridge these two
333 viewpoints, describing how the microbial communities in the EMS change over different scales: between
334 PA and FL communities, over depth and across seasons. Our results suggest that, in the EMS, depth and
335 association with particles are stronger drivers of bacterial community structure when compared to
336 seasonality. Moreover, the PA community exhibited fundamentally different dynamics from the FL one,
337 being less diverse and exhibiting "heterotroph blooms" where different bacterial clades dominate at
338 specific times and depths. We note that our dataset does not have the temporal resolution to fully

339 resolve seasonal changes (e.g. (Ward et al. 2017)), in particular as we did not sample during the
340 maximum winter bloom.

341 In this study, we were able to explain less than 30% of the variability in the FL and PA community
342 structure using “a-biotic” environmental conditions (the depth and season matrices). This is consistent
343 with most other studies of bacterioplankton in marine environments (e.g. (Louca et al. 2016), but see
344 (Thompson et al. 2016)). This leads us to propose that organismal interactions may have a strong impact
345 on microbial population structure. In support of this hypothesis, 20% of the variability in the FL
346 heterotrophic population structure could be explained only when both the depth and phytoplankton
347 community structure matrices were included. The significant factors included pigments that are
348 associated with prymnesiophytes and *Prochlorococcus*. Previous laboratory studies have shown that
349 both organisms interact with co-occurring heterotrophic bacteria, and in turn are affected by them in
350 both positive and negative ways (e.g. (Sher et al. 2011, Segev et al. 2016, Ma et al. 2017, Barak-Gavish et
351 al. 2018)). The observation that the presence of these phytoplankton taxa correlates mostly with
352 heterotrophic FL population structure suggest that, in the EMS, the main interactions occur through the
353 production and utilization of dissolved organic matter rather than being contact-mediated interactions,
354 e.g. in the phycosphere.

355 Aiming to raise testable hypotheses as to additional environmental factors that may impact the
356 heterotrophic bacterial population structure in the EMS, we identified a possible enrichment in the PA
357 community versus the FL one in pathways for the metabolism of anthropogenic pollutants, including
358 pesticides, petroleum products and plastics. These results are based on an inference of community
359 metagenomes, which may be biased by many factors, including unequal representation of different
360 heterotrophic bacterial lineages in the genomic databases and genomic variability between taxa with
361 closely related 16S sequences. We also currently do not have measurements of such pollutants across

362 time and space from the EMS, which are critical in order to test the potential impact of such molecules.
363 We note, however, that, possibly due to analytical reasons (costs, detection limits etc.), anthropogenic
364 pollutants are rarely incorporated into the oceanographic “world view”. Very few oceanographic cruises
365 or time-series observatories include measurements of such contaminants. As a result, global maps of
366 their distributions and dynamics (e.g. similar to those for nutrients, trace metals or organic carbon) are
367 rare or nonexistent. Nevertheless, anthropogenic pollutants can affect the physiology of marine
368 microorganisms, providing nourishment to some groups of organisms (Romera-Castillo et al. 2018) but
369 potentially poisoning others, including picocyanobacteria which are at the base of oligotrophic food
370 webs (Echeveste et al. 2010, Tetu et al. 2019). Our results suggest the need for better understanding of
371 anthropogenic pollutants as potential drivers of microbial communities in the oceans.

372 Finally, to what extent can the results and hypotheses from this study be generalized from the EMS to
373 other oligotrophic oceans? The EMS is unique among oligotrophic marine environments in being
374 strongly depleted in P, with N:P ratios of inorganic nutrients, dissolved and particulate organic matter
375 $\gg 16:1$, compared to ~ 16 in most other marine environments (Krom et al. 2005). This might explain the
376 enrichment in the heterotrophic FL communities in the pathway for utilization of inositol phosphates,
377 which are relatively abundant sources of organic P yet is not well characterized in the oceans (Turner et
378 al. 2002). Additionally, the proximity of the EMS to land might increase the input fluxes of anthropogenic
379 pollution, although such pollution is prevalent also in the most remote areas of the ocean (Rios et al.
380 2010, Chen et al. 2018). Given the relative accessibility of the EMS and its importance as a sea that
381 provides ecosystem services to millions of people, we anticipate that it will become a useful model
382 system to study the processes affecting oligotrophic microbial communities, and how these processes
383 change in a changing world.

384

385 **Brief Experimental Procedures**

386 **Sampling and analysis of nutrients, cell numbers and photosynthetic pigments:** Six one-day cruises
387 were performed over a two-year period to station n-1200 (32 27.36 N, 034 22.47 E) onboard the R/V
388 Mediterranean Explorer. Two cruises were in fall, (December 1st, 2014; November 18th, 2015), two in
389 spring (March 24th, 2015; March 30th, 2016) and two in summer (July 14th, 2015; July 25th, 2016).
390 Samples were collected using 8L Niskin bottles mounted on a rosette with a SeaBird CTD profiler (SBE
391 19plus V2). Water was collected from 5-6 depths, corresponding to surface waters (10m), one-half of
392 the observed DCM, the DCM, 200m and 500m (during the November 2015 cruise, when there was a
393 shallow chlorophyll maximum, samples were collected at 50m and 100m). In several cruises an
394 additional sample was collected from below the DCM and above 200m (termed the “Twilight zone”, or
395 TZ, Supporting Information Excel Table). Due to the amount of water needed for each analysis a
396 separate rosette cast was performed for each depth, resulting in samples being collected up to ~4 hours
397 apart. The mixed layer depth was defined as the depth where potential density differed from surface
398 values by $>0.125 \text{ kg m}^{-3}$ (Malmstrom et al. 2010). We note that, during the spring cruises, there was no
399 clear pycnocline and therefore the actual mixing likely extended significantly below the calculated MLD,
400 but the estimates of the effect of seasonality on community structure are robust to differences in the
401 MLD. Photosynthetic pigments were analyzed using a method adapted for UPLC from the LOV method
402 (Hooker et al. 2005). More information on the pigment analysis, as well as details of the nutrient and
403 flow cytometry analyses, can be found in the Supporting Information methods section.

404 **DNA sequencing and analysis:** 5-11.5L of seawater were filtered using a peristaltic pump onto three
405 filters maintained in-line: 47mm 11 and 5 μm nylon filters and 0.22 μm sterivex filters (Millipore).
406 Storage buffer (40 mM EDTA, 50 mM Tris pH 8.3, 0.75 M sucrose) was added to the samples which were
407 frozen on-board on dry ice and maintained at -80°C until analysis. DNA was extracted using a semi-

408 automated protocol that included manual mechanical and chemical cell lysis followed by automated
409 nucleic acid extraction with a QIAcube system (Haber et al. 2020). PCR amplification was performed
410 using the 16S primer set 515F-Y and 926R that targets the variable V4-5 region and is modified to
411 amplify common oligotrophic bacterial lineages such as SAR11 (Parada et al. 2016). Libraries were
412 sequenced using a MiSeq instrument, paired-end sequencing reads were merged, denoised, pre-
413 processed, and assigned to taxonomic identifiers using Dada2 (version 1.1.6) (Callahan et al. 2016). Exact
414 Sequence Variants (ESVs) were assigned taxonomy using the “classify.seqs” command in MOTHUR, the
415 SILVA database (version 128) and an 80% identity cutoff (Schloss et al. 2009). For Variation Partitioning
416 Analysis (VPA), we defined three matrices of conditions corresponding to those associated with
417 seasonality (cruise number, season, sea surface temperature and mixed layer depth), depth (depth,
418 temperature, NO_x concentration and salinity) and phytoplankton community structure (the ratios of
419 divinyl-chlorophyll a, chlorophyll b, 19'-hex fucoxanthin, fucoxanthin, peridinin and diadinoxanthin to
420 total chlorophyll a). VPA was performed using “VarPart”, followed by conditional Canonical
421 Correspondence Analysis (CCA) using the “cca” and “anova.cca” commands (all in the Vegan package).
422 Metagenome inference from denoised sequences were performed using PICRUSt (Langille et al. 2013),
423 as described previously (Goldford et al. 2018). Communities were normalized using the
424 normalize_otus.py function in PICRUSt, and the metagenomes were estimated using the
425 estimate_metagenome.py routine. The weighted NSTI values ranged between 0.07-0.25 (mean 0.16) for
426 the PA heterotrophic bacteria and 0.09-0.20 (average 0.15) for the FL heterotrophs, within the range of
427 other less-studied environments such as soil and mammalian metagenomes (Langille et al. 2013),
428 suggesting that the results are useful to raise testable hypotheses but should be interpreted with
429 caution. Due to the relatively large sample sizes (e.g. when comparing sample dissimilarities), Welch’s t-
430 tests were used in Microsoft Excel to compare means, as recommended by (Fagerland 2012). More

431 detailed information on the DNA extraction, sequencing and quality control can be found in the

432 Supporting Information methods section.

433 **Data availability:** An excel table with the full environmental dataset, the ESV tables with and without

434 cyanobacteria, and the dynamics of ESVs belonging to specific clades, are presented in the Supporting

435 Excel File. The oceanographic data were deposited in the BCO-DMO under project acronym HADFBA.

436 The sequencing reads were deposited in the NCBI SRA database under project number PRJNA548664.

437 These data include also a transect from the coast to station n-1200, described elsewhere, (Haber et al.

438 2020).

439

440

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453

454 **Conflict of interest**

455 The authors declare no conflict of interest.

456

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