A validation scale to determine the readiness of

2 environmental DNA assays for routine species monitoring

- 3 Running title: eDNA assay validation scale
- 5 Bettina Thalinger^{1,2,3}*, Kristy Deiner^{4,5}, Lynsey R. Harper^{6,7}, Helen C. Rees⁸, Rosetta C.
- 6 Blackman^{9,10}, Daniela Sint^{3,11}, Michael Traugott^{3,11}, Caren S. Goldberg¹² & Kat Bruce¹³
- ¹ Centre for Biodiversity Genomics, University of Guelph, 50 Stone Road E, N1G 2W1, Guelph,
- 9 Ontario, Canada

1

4

- Department of Integrative Biology, College of Biological Sciences, University of Guelph, 50
- 11 Stone Road E, N1G2W1, Guelph, Ontario, Canada
- ³ Department of Zoology, University of Innsbruck, Technikerstr. 25, 6020, Innsbruck, Austria
- ⁴ Natural History Museum, Cromwell Road, London, SW7 5BD, UK
- ⁵ Institute of Biogeochemistry and Pollutant Dynamics, ETH Zurich, Universitätstr. 16, 8092
- 15 Zurich, Switzerland
- ⁶ Illinois Natural History Survey, Prairie Research Institute, University of Illinois at Urbana-
- 17 Champaign, Champaign, IL, USA
- ⁷ School of Biological and Environmental Sciences, Liverpool John Moores University, Liverpool,
- 19 L3 3AF, UK
- ⁸ RSK ADAS Ltd, School of Veterinary Medicine and Science, The University of Nottingham,
- 21 Sutton Bonington Campus, Loughborough, LE12 5RD, UK
- ⁹ Eawag: Swiss Federal Institute of Aquatic Science and Technology, Department of Aquatic
- 23 Ecology, Überlandstr. 133, 8600 Dübendorf, Switzerland
- ¹⁰ Department of Evolutionary Biology and Environmental Studies, University of Zurich,
- Winterthurerstr. 190, CH-8057 Zurich, Switzerland

¹¹ Sinsoma GmbH, Lannes 6, 6176 Völs, Austria 26 ¹² School of the Environment, Washington State University, Pullman, WA 99164, USA 27 ¹³ Nature Metrics Ltd, CABI Site, Bakeham Lane, Egham, TW20 9TY, UK 28 29 *Corresponding author: 30 Bettina Thalinger, bettina.thalinger@gmail.com 31 32 Centre for Biodiversity Genomics, University of Guelph, 50 Stone Road E, N1G 2W1, Guelph, 33 Ontario, Canada; phone: +1 519-824-4120 Ext. 53800 34 Key words: digital PCR, eDNA, endpoint PCR, https://edna-validation.com, quantitative PCR, 35

species-specific, species detection

36

37

Abstract

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59

60

61

62

63

64

65

The use of environmental DNA (eDNA) analysis for species monitoring requires rigorous validation from field sampling to the analysis of PCR-based results - for meaningful application and interpretation. Assays targeting eDNA released by individual species are typically validated with no predefined criteria to answer specific research questions in one ecosystem. Hence, the general applicability of assays as well as associated uncertainties and limitations, often remain undetermined. The absence of clear guidelines for assay validation prevents targeted eDNA assays from being incorporated into species monitoring and policy; thus, their establishment is essential for realizing the potential of eDNA-based surveys. We describe the measures and tests necessary for successful validation of targeted eDNA assays and the associated pitfalls to form the basis of guidelines. A list of 122 variables was compiled, consolidated into 14 thematic blocks, (e.g. "in silico analysis"), and arranged on a 5-level validation scale from "incomplete" to "operational" with defined minimum validation criteria for each level. These variables were evaluated for 546 published singlespecies assays. The resulting dataset was used to provide an overview of current validation practices and test the applicability of the validation scale for future assay rating. Of the 122 variables, 20% to 76% were reported; the majority (30%) of investigated assays were classified as Level 1 (incomplete), and 15% did not achieve this first level. These assays were characterised by minimal in silico and in vitro testing, but their share in annually published eDNA assays has declined since 2014. The meta-analysis demonstrates the suitability of the 5-level validation scale for assessing targeted eDNA assays. It is a user-friendly tool to evaluate previously published assays for future research and routine monitoring, while also enabling the appropriate interpretation of results. Finally, it provides guidance on validation and reporting standards for newly developed assays.

- **Key words:** digital PCR, eDNA, endpoint PCR, https://edna-validation.com/, quantitative PCR,
- species-specific, species detection

1. Introduction

Determining the occurrence of species is essential for ecology and requires sensitive and accurate detection methods. Within the last decade, species detection from environmental DNA (eDNA; i.e. detection of extra-organismal DNA released by organisms into their environment) has shown great potential for routine species surveys (Rees, Maddison, Middleditch, Patmore, & Gough, 2014; Goldberg et al., 2016; Deiner et al., 2017; Langlois, Allison, Bergman, To, & Helbing, 2020; Sepulveda, Nelson, Jerde, & Luikart, 2020). The interest in molecular species detection has fuelled the development of over 500 assays, reviewed herein, that utilise PCR to amplify DNA or RNA extracted from environmental samples. Generally, "targeted" eDNA assays must be specific to the species of interest and possess high sensitivity to allow detection at low densities, low DNA concentrations, and across spatiotemporal scales (Goldberg et al., 2016; MacDonald & Sarre, 2017).

A targeted eDNA assay encompasses the entire workflow used to detect a species' DNA from an environmental sample, inclusive of field sampling through to the interpretation of PCR-

from an environmental sample, inclusive of field sampling through to the interpretation of PCR-based results; it does not just consist of the primers and probes. Thus, adherence to workflows will determine the success or failure of an eDNA assay because methodological choices influence performance and sensitivity (e.g. Doi, Takahara, et al., 2015; Tsuji, Takahara, Doi, Shibata, & Yamanaka, 2019). In practice, assays are often validated within a specific system to answer a set question about the target species. Hence, applications beyond this initial development are hampered by the poor understanding of remaining uncertainties, such as the potential for false positives resulting from non-target amplification or contamination, or false negatives resulting from low sensitivity, sample degradation, low DNA yield protocols or inhibition (Goldberg et al., 2016; Lacoursière-Roussel & Deiner, 2019). In a management

context, false positives and false negatives may lead to misuse of resources (e.g. funds and personnel) for issues such as rare species protection and invasive species control. Both scenarios foster inaccurate interpretation of results, fuelling arguments against the routine use of eDNA detection for species monitoring (Jerde, 2019).

92

93

94

95

96

97

98

99

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

116

117

Aside from a few well-validated eDNA assays already incorporated into routine monitoring. the application of published assays is a minefield for end-users to navigate. We illustrate this through two examples. The assay for great crested newt (Triturus cristatus (Laurenti, 1768), a legally protected species in the UK [Natural England, 2015]) was one of the first eDNA assays validated in both laboratory and field trials against conventional tools, demonstrating its potential for routine monitoring (Thomsen et al., 2012; Rees, Bishop, et al., 2014). After successful validation, a national eDNA-based citizen science monitoring scheme was tested and showed that large-scale eDNA sampling can enable distribution modelling (Biggs et al., 2015). These initial studies paved the way for eDNA-based T. cristatus detection to inform new policies aimed at providing landscape-level species protection (L. R. Harper, Buxton, et al., 2019). Studies have since investigated optimal methods of eDNA capture, relative abundance and detection probability estimation, and the influence of seasonality as well as biotic and abiotic factors on T. cristatus eDNA detection and quantification (Buxton, Groombridge, & Griffiths, 2017; Buxton, Groombridge, Zakaria, & Griffiths, 2017; Buxton, Groombridge, & Griffiths, 2018a, 2018b). Due to these combined efforts, the assay has undergone exemplary validation and has been operational for management since 2014.

Conversely, no assays have been successfully applied to routine monitoring of invasive American crayfish. For example, there is a lack of consensus on a single assay for the signal crayfish (*Pacifastacus leniusculus* (Dana, 1852)). Larson et al. (2017) developed and tested an assay against conventional trapping, but five other assays have also been proposed with differing degrees of validation and divergent *in silico* and *in vitro* approaches (Agersnap et al., 2017; Dunn, Priestley, Herraiz, Arnold, & Savolainen, 2017; K. J. Harper, Anucha, Turnbull,

Bean, & Leaver, 2018; Mauvisseau et al., 2018; Robinson, Uren Webster, Cable, James, & Consuegra, 2018). The *P. leniusculus* assays were developed using a variety of strategies for eDNA sampling, capture, extraction, qPCR of different genetic markers, and applied across genetically diverse populations within the species range. Due to this substantial methodological variability, direct comparisons between results obtained from these assays are impossible. Therefore, the *P. leniusculus* assays represent a minefield for end-users, despite the need for accurate and sensitive tools to enable actionable species management of this invasive species in Europe, Japan, and California (USA).

These case studies exemplify how consensus and dissent in assay validation can influence the implementation of eDNA analysis for species monitoring. Developing guidelines to determine the suitability of eDNA assays for end-users will therefore ensure that ecological insights or management decisions are based on robust molecular analyses with quantifiable uncertainties and clear inference limits (Goldberg et al., 2016; MacDonald & Sarre, 2017; Nicholson et al., 2020; Sepulveda et al., 2020). Here, we describe the general validation process for targeted PCR-based methods and examine the extent of assay validation and reporting in the eDNA literature. We present an eDNA assay validation scale, which establishes criteria to enable the classification of assays based on their accuracy and sensitivity for single-species detection. To demonstrate the utility of the scale, we performed a meta-analysis of targeted eDNA assays published in 327 papers as of 11 April 2019 (546 assays). By placing an eDNA assay on the validation scale, end-users can determine the recommended scenarios for application and improve assay performance with further validation.

2. Criteria and principles of validation

2.1. General requirements for an eDNA laboratory

All laboratory activities are subject to error. In order to have confidence in results, quality standards and good practices are required in diagnostic laboratory environments (e.g. World Health Organization, 2011; Halling, Schrijver, & Persons, 2012). Although few eDNA-processing laboratories will employ 'ancient DNA' practices (e.g. full body suit and positive air pressure with HEPA filtered inflow), all laboratories conducting eDNA analysis should utilise a unidirectional workflow where pre-PCR steps are performed in separate laboratories dedicated to low DNA quality and quantity (Goldberg et al., 2016). Completely standardised laboratory environments are rare and the use of proficiency tests (as conducted by UK laboratories participating in *T. cristatus* monitoring) can help end-users understand the quality of results obtained among different laboratories. Even results obtained from an extensively validated assay can be questionable when they are not produced within a suitable laboratory environment (Goldberg et al., 2016).

2.2. Reporting standards for in silico, in vitro and in situ validation of assays

Targeted eDNA assay validation is a multi-step process. It can be divided into *in silico* validation (i.e. computer-based tests for primer specificity), *in vitro* validation (i.e. laboratory tests with reference tissue samples) and *in situ* validation (i.e. field tests with eDNA samples) (see Goldberg et al., 2016; MacDonald & Sarre, 2017; Langlois et al., 2020; So, Fong, Lam, & Dudgeon, 2020). Understanding the utility of an assay requires both knowledge of the context in which it has been designed, and a broader understanding of how it was developed. Here, we give a brief overview of what key steps comprise assay validation, with a focus on troubleshooting steps that may be necessary when applying previously published assays to new locations or with modified laboratory practices (Table 1).

The first step is *in silico* assay validation, the goal of which is to determine assay specificity based on known sequence diversity. Sequence diversity has three categories, including sequences from: (i) closely related and co-occurring species, (ii) closely related but geographically distinct species, and (iii) distantly related but co-occurring species (i.e. sequences that could co-amplify and produce false positive results for a target species). By checking primer specificity from available sequences, the geographic area of applicability for an assay can be maximised through identifying and removing potential issues of co-amplification. Typically, public or custom databases are used for performing *in silico* amplification (e.g. ecoPCR [Boyer et al., 2016]; primerBlast [Ye et al., 2012]; PrimerTREE [Cannon et al., 2016]; PrimerMiner [Elbrecht & Leese, 2017]). While reference sequence libraries are often far from complete and many of the factors influencing successful PCR amplification cannot be simulated, *in silico* testing provides a first impression of primer performance and should be conducted.

The essential components of *in vitro* assay validation are optimisation, specificity, and sensitivity. Tests with varying PCR chemistry, reaction volume, primer/probe concentration, cycling conditions and technical replication will ensure optimal, standardised and error-free target DNA amplification (Bustin et al., 2009; Wilcox, Carim, McKelvey, Young, & Schwartz, 2015; Goldberg et al., 2016). The assay must then be tested against closely related and co-occurring non-target taxa to ensure specificity, which is not automatically guaranteed after successful *in silico* testing (Goldberg et al., 2016; So et al., 2020). Ideally, tissue-derived DNA samples from multiple individuals spanning a defined geographic area are tested to ensure the assay is robust to genetic variants of target and non-target species. Amplicons should be Sanger sequenced to confirm species identity (Goldberg et al., 2016), although short fragments (<100 bp) have limited sequence length available for species determination (Meusnier et al., 2008). Next, the Limit of Detection (LOD) must be determined to assess assay sensitivity, and the Limit of Quantification (LOQ) determined, if the measurement of eDNA quantity is desired. Generally, these values are obtained using a dilution series of quantified DNA amplicons or

synthesized gene fragments (e.g. IDT gBlocks™ Gene Fragment) based on public or *de novo* reference sequences (Langlois et al., 2020). The LOD and LOQ have various definitions in the eDNA literature, but were recently standardised by Klymus et al. (2019), where LOD is the lowest standard concentration at which 95% of technical replicates amplify and LOQ is the lowest standard concentration for which the coefficient of variation (CV) value is <35%. Unfortunately, the existence of past definitions requires the final LOD and LOQ to be reported as well as the definition used. We note that these metrics apply directly to the assay as developed and assume no interference during PCR from the rest of the species' genome (i.e. if a gBlock is used), other genetic material, or inhibitory compounds.

Finally, the assay must be validated *in situ* by surveying sites with and without the target species (Goldberg et al., 2016). It must be tested against conventional tools for presence/absence detection and tests for estimation of relative abundance/biomass are advisable. Assays are deemed successful if eDNA and conventional detections concur at occupied sites and no eDNA detections are observed at definitively unoccupied sites. Sanger sequencing of eDNA amplicons can provide additional evidence but cannot distinguish sample contamination from true detections (Goldberg et al., 2016). Besides screening for the target species, negative eDNA samples (or all eDNA samples if quantification is necessary) should be tested for inhibition. This requires an Internal Positive Control (IPC) assay for synthetic DNA (e.g. ThermoFisher) or an assay for non-target species using exogenous or endogenous DNA (Goldberg et al., 2016; Veldhoen et al., 2016; Doi et al., 2017; Furlan & Gleeson, 2017).

Advanced *in situ* validation may investigate the influence of biotic (e.g. abundance, biomass, life stages, microbial activity) and abiotic (e.g. temperature, pH, ultraviolet light, salinity) factors influencing eDNA origin, state, fate, and transport (Barnes & Turner, 2016; Lacoursière-Roussel & Deiner, 2019; Wang et al., 2021). Assays that account for spatial (e.g. shoreline versus offshore) and temporal (e.g. summer versus winter) variation in eDNA distribution and abundance due to the ecology of a species can be implemented with greater confidence (de

Souza, Godwin, Renshaw, & Larson, 2016; Lawson Handley et al., 2019). Occupancy modelling using eDNA data is desirable as it accounts for detection probability while estimating site occupancy, even if all field samples from a site return negative. Hierarchical models that incorporate eDNA occupancy and detection probabilities at site, sample, and technical replicate levels are most accurate and can be implemented in software such as R (e.g. package "eDNAoccupancy" [Dorazio & Erickson, 2018]) or PRESENCE (MacKenzie et al., 2002). However, model assumptions regarding false positives should be carefully considered.

3. Types and trade-offs of targeted eDNA detection methods

Amid the processing chain (i.e. sampling to data analysis) for a targeted eDNA assay, PCR warrants extra consideration as the technological spectrum and potential for variation is enormous. Previous publications have typically defined an assay as the primers (and probe) required for DNA amplification and associated visualisation (e.g. agarose gel electrophoresis, qPCR instrumentation). However, differences between the multiple detection instruments and chemistry used in combination with species-specific primer (and probe) sets can fundamentally change the sensitivity and specificity of targeted eDNA assays. Table 2 provides an overview of amplification types and their associated trade-offs.

Many assays - especially those published in earlier years - use endpoint PCR. However, most assays to date employ real-time quantitative (q)PCR allowing for greater sensitivity and quantitative data. More recent publications have used digital (d)PCR for absolute quantification. Alternatively, LAMP and CRISPR have been shown to be suitable for eDNA applications, decreasing the requirements of in-field testing equipment (M. R. Williams et al., 2017; M. A. Williams et al., 2019). A few publications use alternative methods such as PCR combined with restriction fragment length polymorphism (RFLP). All amplification types (Table 2) enable distribution and occupancy modelling, provided enough biological and technical replication is

employed (Hunter et al., 2015; Goldberg, Strickler, & Fremier, 2018; Wilcox et al., 2018). However, endpoint PCR in combination with agarose gels is the only type where a detection limit cannot be set objectively (low sensitivity) and which does not provide estimates of DNA copy number (Doi, Uchii, et al., 2015; Yamanaka & Minamoto, 2016; Hunter et al., 2017; Thalinger, Wolf, Traugott, & Wanzenböck, 2019). Depending on the management context and study size, the optimal detection instrument can vary, albeit technological advances and required accuracy will continue to shift the focus away from endpoint PCR.

4. Evaluating the current status of assay validation

247

248

249

250

251

252

253

254

255

256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

271

272

To understand current validation practices for targeted eDNA assays, we generated an extensive list of variables deemed important for assay validation by 35 experts in the field of targeted eDNA detection whom convened at a DNAqua-Net EU COST Action (Leese et al., 2016) workshop held on 26-27 March 2018 at the University of Innsbruck, Austria. The variables in this list consist of 18 categorical variables (e.g. species identity, target gene, sample type) and 104 binomial variables directly associated with the eDNA processing chain, from primer design to interpretation of field study results (SI1).

A comprehensive literature database for targeted eDNA assays was built in three steps. we included all listed on the 'eDNA First, papers assays' web page (https://labs.wsu.edu/edna/edna-assays/) as of 10 April 2019. Second, we conducted a Web of Science literature search on 11 April 2019, including the search terms "environmental DNA" and "eDNA" but excluding terms associated with microbial organisms and metabarcoding (SI2A). Third, the resulting 660 Web of Science entries were manually checked for suitability (i.e. macrobial target organism, targeted eDNA detection intended) leading to a combined database of 327 papers. For each of the assays contained in these papers, the 122 variables were recorded in a checklist by one of the authors. Before data entry, all authors validated the same

four papers (Deiner & Altermatt, 2014; Rees, Bishop, et al., 2014; Thalinger et al., 2016; L. R. Harper, Griffiths, et al., 2019) to ensure recorder standardisation. Validation efforts were classified as 1 for "tests done, or parameter reported", 0 for "variable not reported, or no testing done", and NA in cases where the respective variable did not match the assay type (e.g. filter type when samples were precipitated). When an assay was used in multiple papers, all validation efforts were summarised in one entry and the literature database was extended with the papers reporting primer sequences or other methodological aspects. As the type of amplification is important for assay validation, primer pairs used on multiple detection platforms were given separate database entries per amplification type. However, because most assays developed were presented in one publication, we did not account for slight variations in other aspects of the workflow (e.g. different extraction method, different filter type). After recording the values for each eDNA assay using the validation checklist (SI1), each author scored the assay intuitively based on a preliminary version of the validation scale (see section 5). The resulting database of 122 variables for each assay was the basis for all further analyses using R (R Core Team, 2020) and associated packages (SI2B).

Altogether, 546 assays from 327 papers were assessed. Of these assays, 227 were designed to detect fish species and 74 were designed to detect amphibian species; hence, it is unsurprising that ~80% of assays utilised water sampling. Fourteen percent of the assays were tested on tissue only and few assays were optimised for other sample types such as aerosol, sediment, snow or soil. More than 80% of assays were reported in only one paper, and most were designed for qPCR (~60%) or endpoint PCR (~35%) platforms. The cytochrome c oxidase subunit I (*COI*) gene was the most popular (>40%) genetic marker, followed by the cytochrome b (*cytb*) gene (~23%) (SI3).

5. The 5-level assay validation scheme

300

301

302

303

304

305

306

307

308

309

310

311

312

313

314

315

316

317

318

319

320

321

322

323

324

To enable standardised assay validation and reporting in the future, we assigned the assayspecific variables to 14 thematic blocks such as "in silico analysis", "PCR", or "extensive field testing of environmental samples" (Table 3, Fig. 1). Each of the variable blocks contains between three and 26 variables (SI1). Some of the variable blocks summarize basic practices (e.g. "in silico analysis") while others describe advanced assay validation (e.g. "detection probability estimation from statistical modelling"). To simplify the reporting of an assay's validation level in the future, the blocks were placed on a five-level scale enabling the categorisation of assays from Level 1 (incomplete) to Level 5 (operational), and the interpretation of associated field study results (Fig. 1). The scale is additive, which means for an assay to be placed at Level 3, it must fulfil the reporting requirements of Levels 1, 2, and 3. Level 1 (incomplete) summarizes assays for which basic in silico analysis, target tissue testing and general information regarding PCR were reported. Level 2 (partial) - assays were characterized by comprehensive reporting of PCR conditions and in vitro testing on closely related non-target species. For assays placed at Level 3 (essential), the target organism was successfully detected from an environmental sample and the specifics of DNA extraction and concentration of eDNA from the environmental sample (i.e. filtration, precipitation) were reported. The LOD, extensive field testing and in vitro testing on co-occurring non-target species were the variable blocks specifically associated with Level 4 (substantial). At Level 5 (operational), the assay has been subjected to comprehensive specificity testing, detection probability estimates from statistical modelling and investigations of ecological and physical factors potentially influencing eDNA in the environment (Table 3, Fig. 1). The placement of assays on this 5-level scale is not straightforward, since each of the 14 variable blocks contains variables associated with either rudimentary or substantial validation and reporting. For instance, the thematic block "concentration of eDNA from environmental

sample" contains the variable "volume/weight of environmental sample", which was reported for

almost all assays, but also contains "pressure used for filtration", which was rarely reported and/or measured. Therefore, a minimum criterion was introduced for each variable block functioning as proof of validation and ensuring standardized placement of assays on the validation scale. For example, "detection from an environmental sample" was used as evidence that some validation had been undertaken in the block "detection obtained from environmental samples" (Table 3). To reach a level on the validation scale, an assay must fulfil all minimum reporting criteria for that level and any preceding levels.

Based on this classification, the results (detected vs. not detected) obtained from eDNA assays become directly interpretable: when Level 1 or Level 2 assays are applied to environmental samples without any further validation steps, it is impossible to tell whether the target species is present or absent independent of the PCR result. Amplifications with a Level 3 assay can be interpreted as "target is likely present"; however, non-amplifications are inconclusive. When Level 4 and Level 5 assays do not lead to amplification, the target is likely absent. Positive PCR results at Level 4 and 5 mean that the target species is almost certainly present (Fig. 1).

For a quantitative analysis of reporting practices, we calculated a total scoring percentage and a block scoring percentage for each assay. The total scoring percentage was defined as the proportion of the 104 binary variables which were reported. For each of the 14 blocks, the block scoring percentage was calculated by dividing the number of variables tested/reported by the complete set of variables associated with the block. Both calculations included only variables relevant to the applied methods (e.g. for assays using filtration, precipitation variables were omitted; see Box 1 for example assays).

Of the 546 assays analysed, the majority (30%) were classed as Level 1. Of the remainder, 15% (N = 83) did not fulfil the minimum criteria necessary to reach Level 1, and no assay reached Level 5 (Fig. 1 and Fig. 2A). Newer assays published after 2016 were more likely to reach Level 4, and the percentage of assays failing to reach Level 1 gradually declined since

352

353

354

355

356

357

358

359

360

361

362

363

364

365

366

367

368

369

370

371

372

373

374

375

376

2014 (Fig. 2B). Generally, the total scoring percentage for all variables in a level increased from Level 0 to Level 4, but its variation was non-uniform with most outliers observed at Level 2 (Fig. 2C). Assays not reaching Level 1 exhibited scoring percentages between ~20% and ~55%, clearly showing the difference between incomplete and partially validated assays, which did not achieve higher levels due to one or several missing validation step(s). Generally, variables associated with lower levels on the validation scale were more likely to be reported or tested. Nevertheless, some variables (e.g. "haplotypes of target tissue" or "pressure used for filtration"), were addressed by fewer than 10% of assays (Fig. 2D).

The rigour of the minimum criteria was evaluated by tallying the not achieved/reported cases. Specifically, 62 assays did not reach Level 1 because the targeted species sequence(s) used in primer design were not reported, which is a prerequisite for this level. Eight assays did not report the primer sequence. The lack of target detection from an environmental sample and omission of filter type or precipitation chemicals were the most restrictive criteria and were not fulfilled by 80 and 43 assays respectively, most of which were exclusively used for tissue tests (Fig. 2E). For assays ranked at Levels 2 and 3, there was agreement between the intuitive assay rating provided by the recorder and that assigned by the objective criteria. For assays placed at Level 1 following objective criteria, authors tended to be more liberal and rated them one or two levels higher (SI4). Finally, a classification tree analysis (De'Ath & Fabricius, 2000) was carried out to identify common characteristics of assays placed at each level of the validation scale (SI5). Most assays failing to reach Level 1 showed distinctly low levels of in silico validation. This was also true for most Level 1 assays (N = 103), albeit these showed higher levels of target tissue validation (Fig. 3). On the other end of the spectrum, Level 4 assays were characterised by substantial testing or reporting for in vitro testing, field sample processing, LOD determination, PCR, and field testing.

6. Conclusions and outlook

377

378

379

380

381

382

383

384

385

386

387

388

389

390

391

392

393

394

395

396

397

398

399

400

401

The validation scale and reporting standards developed here are an inclusive set of guidelines for targeted eDNA assays. They can facilitate communication between the scientific community, commercial providers and government agencies, and provide guidelines regarding the application of previously published assays or the development and publication of new assays (Sepulveda et al., 2020). One needs to acknowledge that a strict standardisation of eDNA assays will not be possible due to applications for manifold taxa in diverse ecosystems combined with technological advances. By checking which of the 122 validation variables have been addressed, it is possible to identify both available and missing information needed to successfully develop or reuse an assay. The level of validation required for successful routine species monitoring will also differ on a case-by-case basis. For example, the *T. cristatus* assay was not placed on Level 5 of the validation scale (in part due to lack of reporting) but was still approved as an official survey method in the UK. As a general recommendation, authors should report as much information as possible on the conducted validation steps, either in the main text or in the supplementary material. Additionally, developers should consult existing guidelines for best practices along the validation workflow prior to assay design and fieldwork (Bustin et al., 2009; Goldberg et al., 2016; MacDonald & Sarre, 2017; Klymus et al., 2019). To make the validation process accessible, we provide a checklist in SI1. Furthermore, a website (https://edna-validation.com/) was created to summarise the cornerstones of the validation process and the validation scale. This website will function as a curated repository for newly developed assays and authors are encouraged to enter all 122 variables and the minimum criteria to rank assays and calculate their scoring percentages on the validation scale. The website will also serve as a living document when improvements in technology and/or our understanding of eDNA in the environment advance.

The 5-level validation scale designed here provides an overview of the capabilities and uncertainties of targeted eDNA assays. However, the binary data entry system cannot replace a close check of previous publications as it does not always allow a qualitative assessment. Details for validation variables are often spread across different sections in a publication or ambiguously displayed. Thus, the checklist can be used as standard reporting guidelines for targeted eDNA assays. It should be emphasised that for specific research questions and associated publications, minimal validation efforts may be sufficient. Nevertheless, thorough validation is needed to reduce uncertainties and overcome the limitations associated with eDNA-based species monitoring. Furthermore, it is important that practitioners consider how an assay can be modified (e.g. using different PCR reagents) and whether this changes its validation level.

The successful application of targeted eDNA assays for routine species detection and monitoring largely depends on the scientific community and the industry providing eDNA services. Laboratories participating in ring tests such as that proposed for metabarcoding (Blackman et al., 2019) can facilitate consensus on analysis standards. For now, assay developers must respond to queries and help troubleshoot reproducibility issues. Such engagement will facilitate the application of targeted eDNA assays by other users and outside their original geographic scope or academic context. Finally, it is necessary for both the scientific community and the commercial laboratories to communicate realistic applications and limitations to end-users, as often an assay is not bad *per se*, but simply unsuitable.

Data availability

The literature database, the validation checklist compiled from it, the R code, and seven datasets derived from the checklist and used for analyses are available on Figshare at

428

429

430

431

432

433

434

435

436

437

438

439

440

441

442

443

444

445

446

447

448

449

450

451

approval for publication.

https://doi.org/10.6084/m9.figshare.12184860.v1 The literature database, the validation scale, and the full checklist are also available at https://edna-validation.com/. **Acknowledgements** This article is based upon work from COST Action DNAqua-Net (CA15219), supported by the COST (European Cooperation in Science and Technology) programme, and RCB was also funded by SNF grant nr. 31003A 173074. We thank the workshop participants who convened on 26-27 March 2018 at the University of Innsbruck, Austria, where initial brainstorming for this manuscript occurred. We acknowledge additional guidance and input from Micaela Hellström, and are grateful to Cathryn Abbott for the improved design of Fig.1. We thank Nature Metrics Ltd. for support in hosting the independent website formed to operationalise the validation scale for future use. **Conflict of interest** KB is the co-founder and CEO of Nature Metrics Ltd., a for profit company dedicated to the analysis of environmental DNA. HCR manages environmental DNA services for RSK ADAS Ltd., a for profit environmental consultancy. DS and MT are co-founders of Sinsoma GmbH., a for profit company dedicated to the analysis of environmental DNA. **Author's contributions** KB conceived the original idea, which was developed by all co-authors. BT compiled the literature database and BT, KD, LRH, HCR, RCB, DS, MT, and KB collected the data. BT analysed the data and led the writing of the manuscript for which first drafts of sections were provided by LH, RCB and KD. All authors contributed critically to the drafts and gave final

454

455

456

457

458

459

460

461

462

463

464

465

466

467

468

469

470

471

472

473

474

475

476

477

478

References Agersnap, S., Larsen, W. B., Knudsen, S. W., Strand, D., Thomsen, P. F., Hesselsøe, M., ... Møller, P. R. (2017). Monitoring of noble, signal and narrow-clawed crayfish using environmental DNA from freshwater samples. PLOS ONE, 12(6), e0179261. doi:10.1371/journal.pone.0179261 Barnes, M. A., & Turner, C. R. (2016). The ecology of environmental DNA and implications for conservation genetics. Conservation Genetics. Springer Netherlands. doi:10.1007/s10592-015-0775-4 Biggs, J., Ewald, N., Valentini, A., Gaboriaud, C., Dejean, T., Griffiths, R. A., ... Dunn, F. (2015). Using eDNA to develop a national citizen science-based monitoring programme for the great crested newt (Triturus cristatus). Biological Conservation, 183, 19–28. doi:10.1016/j.biocon.2014.11.029 Blackman, R., Mächler, E., Altermatt, F., Arnold, A., Beja, P., Boets, P., ... Deiner, K. (2019). Advancing the use of molecular methods for routine freshwater macroinvertebrate biomonitoring – the need for calibration experiments. Metabarcoding and Metagenomics, 3, e34735. doi:10.3897/mbmg.3.34735 Boyer, F., Mercier, C., Bonin, A., Le Bras, Y., Taberlet, P., & Coissac, E. (2016). obitools: A unix-inspired software package for DNA metabarcoding. Molecular Ecology Resources, 16(1), 176–182. doi:10.1111/1755-0998.12428 Bustin, S. A., Benes, V., Garson, J. A., Hellemans, J., Huggett, J., Kubista, M., ... Wittwer, C. T. (2009). The MIQE guidelines: Minimum information for publication of quantitative real-time PCR experiments. Clinical Chemistry, 55(4), 611–622. doi:10.1373/clinchem.2008.112797 Buxton, A. S., Groombridge, J., & Griffiths, R. (2018a). Comparison of Two Citizen Scientist Methods for Collecting Pond Water Samples for Environmental DNA Studies. Citizen Science: Theory and Practice, 3(2). doi:10.5334/cstp.151

480

481

482

483

484

485

486

487

488

489

490

491

492

493

494

495

496

497

498

499

500

501

502

503

504

Buxton, A. S., Groombridge, J. J., & Griffiths, R. A. (2017). Is the detection of aquatic environmental DNA influenced by substrate type? PLOS ONE, 12(8), e0183371. doi:10.1371/journal.pone.0183371 Buxton, A. S., Groombridge, J. J., & Griffiths, R. A. (2018b). Seasonal variation in environmental DNA detection in sediment and water samples. PLOS ONE, 13(1). doi:10.1371/journal.pone.0191737 Buxton, A. S., Groombridge, J. J., Zakaria, N. B., & Griffiths, R. A. (2017, April 10). Seasonal variation in environmental DNA in relation to population size and environmental factors. Scientific Reports. Nature Publishing Group. doi:10.1038/srep46294 Cannon, M. V., Hester, J., Shalkhauser, A., Chan, E. R., Loque, K., Small, S. T., & Serre, D. (2016). In silico assessment of primers for eDNA studies using PrimerTree and application to characterize the biodiversity surrounding the Cuyahoga River. Scientific Reports, 6. doi:10.1038/srep22908 De'Ath, G., & Fabricius, K. E. (2000). Classification and regression trees: A powerful yet simple technique for ecological data analysis. Ecology, 81(11), 3178-3192. doi:10.1890/0012-9658(2000)081[3178:CARTAP]2.0.CO;2 de Souza, L. S., Godwin, J. C., Renshaw, M. A., & Larson, E. (2016). Environmental DNA (eDNA) detection probability is influenced by seasonal activity of organisms. PLOS ONE, 11(10), e0165273. doi:10.1371/journal.pone.0165273 Deiner, K., & Altermatt, F. (2014). Transport distance of invertebrate environmental DNA in a natural river. PLOS ONE, 9(2). doi:10.1371/journal.pone.0088786 Deiner, K., Bik, H. M., Mächler, E., Seymour, M., Lacoursière-Roussel, A., Altermatt, F., ... Bernatchez, L. (2017). Environmental DNA metabarcoding: Transforming how we survey animal and plant communities. Molecular Ecology, 26(21), 5872-5895. doi:10.1111/mec.14350 Doi, H., Katano, I., Sakata, Y., Souma, R., Kosuge, T., Nagano, M., ... Tojo, K. (2017).

505 Detection of an endangered aquatic heteropteran using environmental DNA in a wetland 506 ecosystem. Royal Society Open Science, 4(7). doi:10.1098/rsos.170568 Doi, H., Takahara, T., Minamoto, T., Matsuhashi, S., Uchii, K., & Yamanaka, H. (2015). Droplet 507 508 digital polymerase chain reaction (PCR) outperforms real-time PCR in the detection of 509 environmental DNA from an invasive fish species. Environmental Science and Technology. 510 49(9), 5601–5608. doi:10.1021/acs.est.5b00253 511 Doi, H., Uchii, K., Takahara, T., Matsuhashi, S., Yamanaka, H., & Minamoto, T. (2015). Use of droplet digital PCR for estimation of fish abundance and biomass in environmental DNA 512 513 surveys. PLOS ONE, 10(3), 1–11. doi:10.1371/journal.pone.0122763 514 Dorazio, R. M., & Erickson, R. A. (2018), ednaoccupancy: An r package for multiscale 515 occupancy modelling of environmental DNA data. Molecular Ecology Resources, 18(2), 516 368-380. doi:10.1111/1755-0998.12735 517 Dunn, N., Priestley, V., Herraiz, A., Arnold, R., & Savolainen, V. (2017). Behavior and season 518 affect crayfish detection and density inference using environmental DNA. Ecology and 519 Evolution, 7(19), 7777–7785. doi:10.1002/ece3.3316 520 Elbrecht, V., & Leese, F. (2017). PrimerMiner: an R package for development and in silico 521 validation of DNA metabarcoding primers. Methods in Ecology and Evolution, 8(5), 622-626. doi:10.1111/2041-210X.12687 522 523 Furlan, E. M., & Gleeson, D. (2017). Improving reliability in environmental DNA detection 524 surveys through enhanced quality control. Marine and Freshwater Research, 68(2), 388. doi:10.1071/MF15349 525 526 Goldberg, C. S., Strickler, K. M., & Fremier, A. K. (2018). Degradation and dispersion limit 527 environmental DNA detection of rare amphibians in wetlands: Increasing efficacy of 528 sampling designs. Science of the Total Environment, 633, 695–703. 529 doi:10.1016/j.scitotenv.2018.02.295 530 Goldberg, C. S., Turner, C. R., Deiner, K., Klymus, K. E., Thomsen, P. F., Murphy, M. A., ...

532

533

534

535

536

537

538

539

540

541

542

543

544

545

546

547

548

549

550

551

552

553

554

555

556

Taberlet, P. (2016). Critical considerations for the application of environmental DNA methods to detect aquatic species. Methods in Ecology and Evolution, 7(11), 1299–1307. doi:10.1111/2041-210X.12595 Halling, K. C., Schrijver, I., & Persons, D. L. (2012). Test verification and validation for molecular diagnostic assays. Archives of Pathology and Laboratory Medicine, 136(1), 11-13. doi:10.5858/arpa.2011-0212-ED Harper, K. J., Anucha, N. P., Turnbull, J. F., Bean, C. W., & Leaver, M. J. (2018). Searching for a signal: Environmental DNA (eDNA) for the detection of invasive signal crayfish. Pacifastacus Ieniusculus (Dana, 1852). Management of Biological Invasions, 9(2), 137-148. doi:10.3391/mbi.2018.9.2.07 Harper, L. R., Buxton, A. S., Rees, H. C., Bruce, K., Brys, R., Halfmaerten, D., ... Hänfling, B. (2019). Prospects and challenges of environmental DNA (eDNA) monitoring in freshwater ponds. Hydrobiologia, 826(1), 25-41. doi:10.1007/s10750-018-3750-5 Harper, L. R., Griffiths, N. P., Lawson Handley, L., Sayer, C. D., Read, D. S., Harper, K. J., ... Hänfling, B. (2019). Development and application of environmental DNA surveillance for the threatened crucian carp (Carassius carassius). Freshwater Biology, 64(1), 93–107. doi:10.1111/fwb.13197 Hunter, M. E., Dorazio, R. M., Butterfield, J. S. S., Meigs-Friend, G., Nico, L. G., & Ferrante, J. A. (2017). Detection limits of quantitative and digital PCR assays and their influence in presence-absence surveys of environmental DNA. Molecular Ecology Resources, 17(2), 221-229. doi:10.1111/1755-0998.12619 Hunter, M. E., Oyler-McCance, S. J., Dorazio, R. M., Fike, J. A., Smith, B. J., Hunter, C. T., ... Hart, K. M. (2015). Environmental DNA (eDNA) sampling improves occurrence and detection estimates of invasive Burmese pythons. PLOS ONE, 10(4), e0121655. doi:10.1371/journal.pone.0121655 Jerde, C. L. (2019). Can we manage fisheries with the inherent uncertainty from eDNA? Journal

558

559

560

561

562

563

564

565

566

567

568

569

570

571

572

573

574

575

576

577

578

579

580

581

582

of Fish Biology, ifb.14218. doi:10.1111/ifb.14218 Klymus, K. E., Merkes, C. M., Allison, M. J., Goldberg, C. S., Helbing, C. C., Hunter, M. E., ... Richter, C. A. (2019). Reporting the limits of detection and quantification for environmental DNA assays. Environmental DNA, edn3.29. doi:10.1002/edn3.29 Lacoursière Roussel, A., & Deiner, K. (2019). Environmental DNA is not the tool by itself. Journal of Fish Biology, jfb.14177. doi:10.1111/jfb.14177 Langlois, V. S., Allison, M. J., Bergman, L. C., To, T. A., & Helbing, C. C. (2020). The need for robust qPCR □ based eDNA detection assays in environmental monitoring and species inventories. Environmental DNA, edn3.164. doi:10.1002/edn3.164 Larson, E. R., Renshaw, M. A., Gantz, C. A., Umek, J., Chandra, S., Lodge, D. M., & Egan, S. P. (2017). Environmental DNA (eDNA) detects the invasive crayfishes Orconectes rusticus and Pacifastacus leniusculus in large lakes of North America. Hydrobiologia, 800(1), 173-185. doi:10.1007/s10750-017-3210-7 Lawson Handley, L., Read, D. S., Winfield, I. J., Kimbell, H., Johnson, H., Li, J., ... Hänfling, B. (2019). Temporal and spatial variation in distribution of fish environmental DNA in England's largest lake. Environmental DNA, 1(1), 26–39. doi:10.1002/edn3.5 Leese, F., Altermatt, F., Bouchez, A., Ekrem, T., Hering, D., Meissner, K., ... Zimmermann, J. (2016). DNAqua-Net: Developing new genetic tools for bioassessment and monitoring of aquatic ecosystems in Europe. Research Ideas and Outcomes, 2, e11321. doi:10.3897/rio.2.e11321 MacDonald, A. J., & Sarre, S. D. (2017). A framework for developing and validating taxonspecific primers for specimen identification from environmental DNA. Molecular Ecology Resources, 17(4), 708-720. doi:10.1111/1755-0998.12618 MacKenzie, D. I., Nichols, J. D., Lachman, G. B., Droege, S., Royle, A., & Langtimm, C. A. (2002). Estimating site occupancy rates when detection probabilities are less than one. Ecology, 83(8), 2248-2255.

584

585

586

587

588

589

590

591

592

593

594

595

596

597

598

599

600

601

602

603

604

605

606

607

608

Mauvisseau, Q., Coignet, A., Delaunay, C., Pinet, F., Bouchon, D., & Souty-Grosset, C. (2018). Environmental DNA as an efficient tool for detecting invasive crayfishes in freshwater ponds. Hydrobiologia, 805(1), 163-175. doi:10.1007/s10750-017-3288-y Meusnier, I., Singer, G. A. C., Landry, J. F., Hickey, D. A., Hebert, P. D. N., & Hajibabaei, M. (2008). A universal DNA mini-barcode for biodiversity analysis. BMC Genomics, 9. doi:10.1186/1471-2164-9-214 Natural England. (2015). Great crested newts: protection and licences. Retrieved 7 April 2020, from https://www.gov.uk/guidance/great-crested-newts-protection-surveys-and-licences Nicholson, A., McIsaac, D., MacDonald, C., Gec, P., Mason, B. E., Rein, W., ... Hanner, R. H. (2020). An analysis of metadata reporting in freshwater environmental DNA research calls for the development of best practice guidelines. Environmental DNA. doi:10.1002/edn3.81 R Core Team. (2020). R: A Language and Environment for Statistical Computing. Vienna, Austria: R Foundation for Statistical Computing, Retrieved from https://www.r-project.org/ Rees, H. C., Bishop, K., Middleditch, D. J., Patmore, J. R. M., Maddison, B. C., & Gough, K. C. (2014). The application of eDNA for monitoring of the Great Crested Newt in the UK. Ecology and Evolution, 4(21), 4023–4032. doi:10.1002/ece3.1272 Rees, H. C., Maddison, B. C., Middleditch, D. J., Patmore, J. R. M., & Gough, K. C. (2014). The detection of aquatic animal species using environmental DNA - a review of eDNA as a survey tool in ecology. Journal of Applied Ecology, 51(5), 1450-1459. doi:10.1111/1365-2664.12306 Robinson, C. V., Uren Webster, T. M., Cable, J., James, J., & Consuegra, S. (2018). Simultaneous detection of invasive signal crayfish, endangered white-clawed crayfish and the crayfish plague pathogen using environmental DNA. Biological Conservation, 222, 241–252. doi:10.1016/j.biocon.2018.04.009 Sepulveda, A. J., Nelson, N. M., Jerde, C. L., & Luikart, G. (2020). Are Environmental DNA Methods Ready for Aquatic Invasive Species Management? Trends in Ecology &

610

611

612

613

614

615

616

617

618

619

620

621

622

623

624

625

626

627

628

629

630

631

632

633

634

Evolution, 0(0). doi:10.1016/j.tree.2020.03.011 So, K. Y. K., Fong, J. J., Lam, I. P. Y., & Dudgeon, D. (2020). Pitfalls during in silico prediction of primer specificity for eDNA surveillance. Ecosphere, 11(7). doi:10.1002/ecs2.3193 Thalinger, B., Oehm, J., Mayr, H., Obwexer, A., Zeisler, C., & Traugott, M. (2016). Molecular prey identification in Central European piscivores. Molecular Ecology Resources, 16(1), 123-137. doi:10.1111/1755-0998.12436 Thalinger, B., Wolf, E., Traugott, M., & Wanzenböck, J. (2019). Monitoring spawning migrations of potamodromous fish species via eDNA. Scientific Reports, 9(1). doi:10.1038/s41598-019-51398-0 Thomsen, P. F., Kielgast, J., Iversen, L. L., Wiuf, C., Rasmussen, M., Gilbert, M. T. P., ... Willersley, E. (2012). Monitoring endangered freshwater biodiversity using environmental DNA. Molecular Ecology, 21(11), 2565–2573. doi:10.1111/j.1365-294X.2011.05418.x Tsuji, S., Takahara, T., Doi, H., Shibata, N., & Yamanaka, H. (2019). The detection of aquatic macroorganisms using environmental DNA analysis—A review of methods for collection, extraction, and detection. Environmental DNA, 1(2), 99-108. doi:10.1002/edn3.21 Veldhoen, N., Hobbs, J., Ikonomou, G., Hii, M., Lesperance, M., & Helbing, C. C. (2016). Implementation of novel design features for qpcr-based eDNA assessment. PLoS ONE, 11(11), 1–23. doi:10.1371/journal.pone.0164907 Wang, S., Yan, Z., Hänfling, B., Zheng, X., Wang, P., Fan, J., & Li, J. (2021, February 10). Methodology of fish eDNA and its applications in ecology and environment. Science of the Total Environment. Elsevier B.V. doi:10.1016/j.scitotenv.2020.142622 Wilcox, T. M., Carim, K. J., McKelvey, K. S., Young, M. K., & Schwartz, M. K. (2015). The dual challenges of generality and specificity when developing environmental DNA markers for species and subspecies of Oncorhynchus. PLOS ONE, 10(11), e0142008. doi:10.1371/journal.pone.0142008 Wilcox, T. M., Young, M. K., McKelvey, K. S., Isaak, D. J., Horan, D. L., & Schwartz, M. K.

(2018). Fine □scale environmental DNA sampling reveals climate □ mediated interactions 635 636 between native and invasive trout species. Ecosphere, 9(11), e02500. 637 doi:10.1002/ecs2.2500 638 Williams, M. A., O'Grady, J., Ball, B., Carlsson, J., de Eyto, E., McGinnity, P., ... Parle-639 McDermott, A. (2019). The application of CRISPR-Cas for single species identification from 640 environmental DNA. Molecular Ecology Resources, 19(5), 1106-1114. doi:10.1111/1755-641 0998.13045 Williams, M. R., Stedtfeld, R. D., Engle, C., Salach, P., Fakher, U., Stedtfeld, T., ... Hashsham, 642 S. A. (2017). Isothermal amplification of environmental DNA (eDNA) for direct field-based 643 644 monitoring and laboratory confirmation of Dreissena sp. PLOS ONE, 12(10). 645 doi:10.1371/journal.pone.0186462 646 World Health Organization. (2011). Laboratory quality standards and their implementation. New Delhi: WHO Regional Office for the Western Pacific. 647 648 Yamanaka, H., & Minamoto, T. (2016). The use of environmental DNA of fishes as an efficient method of determining habitat connectivity. Ecological Indicators, 62, 147–153. 649 650 doi:10.1016/j.ecolind.2015.11.022 Ye, J., Coulouris, G., Zaretskaya, I., Cutcutache, I., Rozen, S., & Madden, T. L. (2012). Primer-651 652 BLAST: a tool to design target-specific primers for polymerase chain reaction. BMC 653 Bioinformatics, 13, 134. doi:10.1186/1471-2105-13-134 654

Tables and Figures

655

656

Table 1: A guideline for troubleshooting at different stages of the validation process.

General issues	eral issues In silico validation In vitro validation		In situ validation			
Applying assay to a new geographic location	Test non-target sequences available on public/custom databases from this location	Test tissue from non-target species present at this location, including species present at original location if populations are genetically diverse	Test assay at sites where new, potentially problematic non-target species are known to occur			
Sparse reference data	Where possible, Sanger sequence tissue DNA from target and non-target species to generate reference sequences	Extensively test assay on tissue available for target and non-target species, i.e. test extracts from multiple individuals	Extensively test assay at sites where target species co-occurs with non-target species lacking reference sequences, and at sites where only non-target species lacking reference sequences occur			
In vitro validation issues	Troubleshooting guidelines	<i>In situ</i> validation issues	Troubleshooting guidelines			
Poor amplification efficiency	Optimise reaction volume	Expected species	Survey sites with conventional tools and eDNA metabarcoding if possible			
	Optimise cycle number	presence/absence not confirmed	EDIAM III Etabarcouling II possible			
	Optimise primer-probe concentration	False positives	Discard samples corresponding to contaminated controls; consider lack of specificity; Sanger sequence results			
	Optimise annealing temperature using gradient PCR					
	Optimise technical replication		Resample with more stringent			
	Check pipetting accuracy		decontamination measures in place			
Poor specificity	Increase annealing temperature Add hydrolysis probe	False negatives	Check for inhibition and treat samples if inhibited; consider timing and spacing of			
	Perform melting curve analysis		sampling			
	Try other enzymes (e.g. EMM)		Increase technical replication Resample and increase biological replication and volume of water collected			
	Consider redesign					
Low sensitivity (see also Klymus et al., 2019)	Ensure enough replication of fresh standards used when establishing the Limit of Detection	Poor quality Sanger sequencing	Purify amplicons prior to sequencing			
	Use TE buffer and tRNA to make standard dilutions, not molecular grade water		Concentrate amplicons prior to sequencing			
	Use low retention tubes and pipette tips when preparing standards to prevent adsorption to plastic	Inhibition (determined by failed/skewed amplification of IPC or non-target assay)	Use a DNA extraction kit that includes an inhibitor removal step (e.g. mu-DNA, Qiagen QIAamp DNA Stool Mini Kit, Qiagen DNeasy PowerWater Kit, Qiagen DNeasy PowerSoil Kit)			
	If possible, switch to PCR platform with greater sensitivity		Apply an inhibitor removal kit (e.g. OneStep Inhibitor Removal Kit, Zymo Research Corp.) to samples			
	Consider redesign		Use PCR reagents designed to handle inhibition (e.g. TaqMan Environmental Master Mix, Bovine Serum Albumin)			

Table 2: Detection methods and their trade-offs used for targeted eDNA assays. Abbreviations are as follows: polymerase chain reaction (PCR), capillary electrophoresis (CE), quantitative (q), digital (d), intercalating dye (ID), loop-mediated isothermal amplification (LAMP).

Trade-offs	Endpoint PCR-gel	Endpoint PCR-CE	qPCR- ID	qPCR- probe	dPCR- ID	dPCR- probe	LAMP
Quantification	no	yes (limited precision)	yes	yes	yes	yes	yes, with real- time monitoring
LOD (sensitivity)	medium - high DNA concentration (depending on which intercalating dye)	low - medium DNA concentration (with a well-designed assay)	low DNA quantities	low DNA quantities	absolute DNA quantities	absolute DNA quantities	medium DNA concentration
Specificity	medium	medium	medium	high	medium	high	high
Cost reagents	low	medium	high	high	high	very high	medium
Cost equipment	low	medium	medium	medium	high	high	low - medium (depending on LAMP cycler)
Multiplex possible	yes, if different sizes are used	yes, if different sizes or fluorophores are used	difficult	yes, if different fluorophores are used	difficult	yes, if different fluorophores are used	yes
Time (PCR to data)	slow	slow	medium	medium	slow	slow	fast (15-30 min)
Assay type transferable	yes	yes	yes	yes	yes, to qPCR	yes, to qPCR	not easily
Effort to design assay	low - medium	low - medium	high	high	high	high	high
Specialised software	advisable	advisable	advisable	required	advisable	required	required
Sequencing confirmation	yes	yes	yes	yes	limited	limited	limited

- Table 3: The thematic variable blocks of the 5-level validation scale and their respective
- 661 minimum criteria.

Validation level	Variable blocks	Minimum criteria			
Level 1	in silico analysis	target species			
	target tissue testing	target tissue			
	target tissue PCR	primer (and probe) sequence			
Level 2	comprehensive reporting of PCR conditions	DNA extract volume in PCR			
	in vitro testing on closely related non-target species	any in vitro non-target testing			
Level 3	extraction method performed on eDNA samples	method of extraction			
	concentration of eDNA from environmental sample	filter type or precipitation chemicals			
	detection obtained from environmental samples	detection from an environmental sample (artificial or natural habitat)			
Level 4	Limit of Detection (LOD)	LOD determined			
	extensive field testing of environmental samples	multiple locations or multiple samples			
	in vitro testing on co-occurring non-target species	any advanced in vitro testing			
Level 5	comprehensive specificity testing	non-co-occurring/closely related species checked from in silico			
	detection probability estimation from statistical modelling	any effort made towards detection probability estimation			
	understanding ecological and physical factors influencing eDNA in the environment	any factor influencing eDNA in the environment tested			

Figure 1: An overview of the 5-level validation scale. For each of the levels (incomplete to operational), the main accomplishments in the validation process and an appropriate interpretation of results are provided.

Figure 2: The main outcomes of the meta-analysis based on 546 assays from 327 publications are presented in panels A to E. Assay classification is based on the minimum criteria presented in Figure 1. Level 0 codes for assays that did not reach Level 1 on the validation scale. The colour coding is consistent for all panels: Level 0 (grey), Level 1 (dark purple), Level 2 (blue), Level 3 (turquoise), Level 4 (green), and Level 5 (yellow). Panel A shows the distribution of assays across levels of the validation scale. Panel B displays the percentage of assays (*N* = 546) rated Level 0 to 4 that have been published each year since 2003. Panel C summarises variable reporting per assay level. Panel D shows the percentage of assays reporting a specific variable (colour-coded according to level). Panel E shows the minimum criteria necessary to reach each level of the validation scale, and the percentage of Level 0 to 4 assays that did not report these. All variable abbreviations are listed in SI1.

Figure 3: Classification tree analysis identifying the criteria distinguishing assays at different levels of the validation scale. The conditions along the branches show the criteria on which the dataset is split. Numbers in coloured leaves show the validation level of the assays in the respective leaf. Numbers below the leaves represent the number of assays per validation level, summarised inside an individual leaf. The displayed percentage is the proportion of assays summarised per leaf.

Box 1: Examples of assays rated at Levels 0 to 4. The vertical tile plot shows which of the minimum criteria the assay fulfils (yellow tiles), and the bar chart gives the scoring percentage (i.e. the proportion of variables that were tested or reported) for each of the variable blocks.

- Bars are coloured according to the score obtained from a block with dark purple coding for "no
- validation" and yellow coding for "comprehensive validation".

Figure 1

Level 1 incomplete

Level 2 partial

Level 3 essential

Level 4 substantial

Level 5 operational

assay designed

tested on target tissue

Level 1

assay optimized

tested on closely related non-target species

Level 2

assay tested on eDNA samples

positive detections obtained all sample processing steps reported in detail

Level 3

Limit of Detection (LOD) established

extensive field testing and in vitro testing on co-occurring non-target species

Level 4

+

detection probabilities estimated by statistical modelling

comprehensive specificity testing and investigating environmental influences

Interpretation of Results



Levels 1 and 2

impossible to tell if target is present or absent

Level 3

<u>not detected</u>: impossible to tell if target is present or absent

detected: target is likely present if

- field negative controls return negative
- eDNA-appropriate laboratory
- · positive detections are sequenced



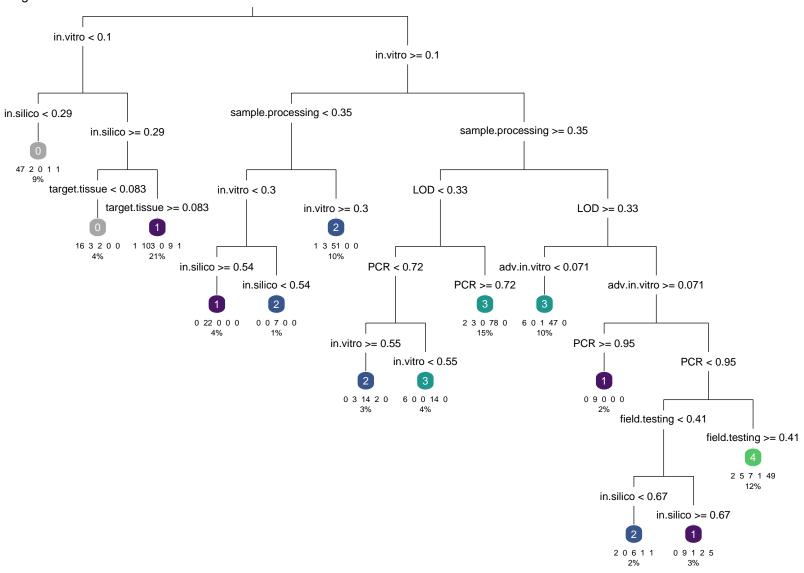
Levels 4 and 5

not detected: target likely absent, assuming appropriate timing and replication in sampling; Level 5 provides the probability of species presence despite negative results

detected: target very likely present

minimum criteria

Figure 3:



Box 1:

Level 4: Assay no. 165 detecting brook trout, *Salvelinus fontinalis*. Assay information was published in paper nos. 291, 12, 145, 295, 144, with latest publication in 2018. Despite low validation efforts for *in vitro* testing and field detection, the only minimum criterion that prevented this assay attaining Level 5 was "comprehensive specificity testing".

Level 3: Assay no. 473 detecting the crucian carp, *Carassius carassius*. Assay information was published in paper no. 119 in 2019. No advanced *in vitro* testing was conducted for this assay, but validation efforts were substantial for most thematic blocks, especially "*in silico* testing", "LOD

determination", and "comprehensive

specificity testing".

common eelgrass, Zostera marina. Assay information was published in paper no. 117 in 2018. Despite low validation efforts in many thematic blocks, only the minimum criterion of "sample processing" prevented this assay from reaching Level 4 on the validation scale

Level 2: Assay no. 270 detecting

Level 1: Assay no. 315 detecting the waterlouse, Asellus aquaticus. Assay information was published in paper no. 163 in 2015. With additional in vitro testing, this assay would have attained Level 3. A clear decrease in validation efforts is visible for variable blocks associated with higher levels of the validation scale.

Level 0: Assay no. 343 detecting Daphnia magna. Assay information was published in paper no. 243 in 2018. Despite some advanced validation, this assay does not fulfill the minimum criterion of "in silico testing" and "target tissue testing". Therefore, it could not be classed as Level 1 on the validation scale.

