The phylogenetic and global distribution of extracellular bioplastic degrading genes

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Abstract

Polyhydroxyalkanoates (PHAs) are a family of microbially-made polyesters that have been commercialized as biodegradable plastics. PHA production rates are predicted to increase rapidly as global concerns around environmental plastic contamination and limited fossil fuel resources have increased the importance of bio-based plastic alternatives. PHAs are meant to quickly degrade in the environment, but this degradation is reliant on microbially-secreted PHA depolymerases, whose taxonomic and environmental distribution have not been well-defined. As a result, the impact of increased PHA production and disposal on global environments is unknown. Here we used 3,842 metagenomes to analyze the distribution of PHA depolymerase genes in microbial communities from diverse aquatic, terrestrial and waste management systems. Our results indicate that extracellular PHA depolymerases are globally widespread but unevenly distributed, with certain environments showing little to no evidence for this activity. In tandem, we screened 5,290 metagenome-assembled genomes to describe the phylogenetic distribution of this trait, which is substantially broader compared to current cultured representatives. We identified members of the Proteobacteria and Bacteroidetes as key lineages with PHA biodegradation potential and predict this activity in members of the Actinobacteria, the Candidate phylum Rokubacteria, Firmicutes, Planctomycetes and Spirochaetes.

Importance: Environmental concerns alongside legislation banning single-use petroleum-based plastics are expected to promote the production of bio-based plastics, including PHAs. PHAs represent a novel and emerging waste stream. If PHA disposal follows the precedent set by conventional plastics, a significant portion will be littered into the environment, or improperly discarded into landfills instead of composting facilities. Traditionally, the identification of

bioplastic degrading enzymes and organisms has relied on culture-dependent assays. As a result, the PHA degradation capabilities of the "unculturable" fraction of microorganisms remain largely unexplored. Here, we leverage large amounts of environmental sequence data to assess which environments harbor PHA-degrading organisms and to determine the taxonomic affiliations of bioplastic degraders. Our analyses inform our understanding of the biodegradation potential in the environment, with implications for the impact of bioplastic pollution. We identify enzymes and organisms that may be suitable for future bioremediation, chemical processing or biotechnological applications.

Introduction

Widespread environmental impacts associated with liberal plastic use have led to a demand for sustainable biodegradable alternatives, including polyhydroxyalkanoates (PHAs) (1). Conventional plastics rely on fossil resources and are extremely recalcitrant to degradation, causing them to accumulate in the environment (2). Annual plastic production now exceeds 380 million tons (3) and accounts for approximately 6% of global fossil feedstocks (4). Plastics often serve only a short service lifespan before being discarded (1). It is estimated that, as of 2015, humans had generated 6,300 million metric tons of plastic waste (3). According to a recent global assessment, only 21% of plastic waste has been incinerated or recycled; the remaining 79% has been discarded in landfills or littered into the environment (3). Plastics in the environment can harm organisms through ingestion or entanglement causing ecological damage (5). The effects of plastic pollution in the environment is a complex problem and the extent of these effects remains largely unknown (6).

New legislation aimed at mitigating the environmental impacts of plastics alongside growing public concern has generated a growing need for new biodegradable alternatives (7–9). PHAs have garnered substantial academic and industrial interest, and commercially available PHAs have been developed, for reviews see (10–12). PHAs are considered environmentally favorable to conventional plastics because they can be made from renewable carbon resources and have been shown to degrade in many different environments (13–20). 2.11 million tons of bioplastics were produced in 2018 (8). PHAs currently represent 1.4% of the global bioplastic market with production rates expected to more than double in the next five years (8).

PHAs are a family of polyesters that are synthesized naturally by many types of bacteria and archaea as an intracellular energy and carbon storage compound (13, 21). There are over 150 different types of PHA monomers which can be combined in different ways to yield polymers with a variety of chemical and structural properties (22). PHAs are broadly classified as either short-chain length (PHAscl) or medium chain length (PHAmcl) depending on the number of carbon atoms found in the monomer. PHAscl have 3-5 carbon atoms per monomer and PHAmcl have 6-15 carbon atoms per monomer. PHAs occur in two different biophysical states, which impacts which enzymes can break them down (13). Intracellular PHAs are amorphous and are found in protein-associated granules sometimes referred to as carbonosomes (23); extracellular PHAs become denatured upon extraction and take on a semi-crystalline state (13).

PHA biodegradation in the environment is mediated by microorganisms that produce PHA depolymerases (24). PHA depolymerases are a diverse family of intracellular and extracellular carboxylesterases and are members of the alpha/beta-hydrolase fold family (25). To

93 date, approximately 35 PHA depolymerases have been isolated and biochemically characterized
94 (Supp Data 1 and references within), from organisms within the phyla Ascomycetes, Firmicutes,
95 Proteobacteria (most) and Actinobacteria. PHA depolymerases are classified based on
96 subcellular location (extracellular or intracellular), substrate specificity (mcl [EC 3.1.1.76] or scl
97 [EC 3.1.1.75]), and features of the catalytic domain (13, 25). Intracellular PHA (inPHA)
98 depolymerases mobilize amorphous endogenous PHA stores whereas extracellular PHA (ePHA)
99 depolymerases degrade exogenous semi-crystalline PHAs (24). ePHA depolymerases allow
100 microorganisms to scavenge PHAs that have been released into the environment from ruptured
101 cells (13) and are the main avenue for biodegradation of commercial PHAs. Here, we focused
102 our analyses on ePHA depolymerases because we were interested in enzymes that are secreted
103 into the environment and have the potential to degrade commercial PHAs.
104
105 All known ePHA depolymerases contain a catalytic triad composed of a serine
106 (embedded in a lipase box motif), histidine and aspartic acid (13, 25). Extracellular short chain
107 length (ePHAscl) depolymerases typically consist of a signal peptide, a catalytic domain, a linker
108 domain and a C-terminal substrate binding domain (25). There are two recognized ePHAscl
109 depolymerase subtypes (denoted type 1 [T1] and type 2 [T2]) which differ in the arrangement of
110 their catalytic domains (13). In ePHAsclT1 depolymerases the oxyanion hole is N-terminal to the
111 lipase box, whereas in ePHAsclT2 depolymerases the oxyanion hole is C-terminal to the lipase
112 box (25). Characterized extracellular medium-chain length (ePHAmcl) depolymerases do not
113 exhibit a C-terminal substrate binding domain; instead the N-terminal region is presumed to
114 mediate substrate binding (13). ePHAmcl depolymerases share no significant homology with
115 ePHAscl depolymerases except for the catalytic triad (25), which is non-specific as it is shared

116 by all carboxylesterases (27). However, ePHAsclT1 depolymerases do share significant
117 homology with some biochemically characterized intracellular PHAscl (inPHAscl)
118 depolymerases (28) and ePHAsclT2 depolymerases share significant similarity with a
119 periplasmic PHA depolymerase specific for amorphous forms of PHA (29).
120
121 Rates of plastic degradation are influenced by polymer properties (e.g., available surface
122 area, monomeric composition, presence of additives), the extant microbial population, and
123 abiotic factors (e.g., temperature, UV, pH) (30). PHAs are considered biodegradable, however
124 the rate of this process depends on the polymer properties and the environmental context (1).
125 Thus, the term "biodegradable" is ambiguous unless the conditions and rate of degradation are
126 reported (31). Standardized tests exist to quantify degradation under specified conditions, such as
127 industrial composts (see ASTM D5338-15) (32), but these tests cannot be applied to non-
128 engineered environmental systems. Field studies on biodegradability have shown conflicting
129 results, and there is a growing concern that biodegradable plastics are often not as biodegradable
130 as claimed (33).
131
132 Here, we have focused specifically on determining which environments harbour microbes
133 with the genetic potential to mediate PHA biodegradation. PHA biodegradation rates are strongly
134 influenced by the abundance and diversity of microorganisms encoding PHA depolymerases
135 (24). With PHA production levels expected to increase, a strong understanding of the distribution
136 of PHA-degrading enzymes and organisms is crucial for understanding the impact of PHAs on
137 the environment (24). Early culture-based studies have isolated PHAscl degrading
138 microorganisms from different environments including soil, compost, sewage and aquatic

139 environments, but indicated that PHAmcl degraders appear to be relatively rare (13, 24).
140 Reflecting these findings, relatively few ePHAmcl depolymerases have been characterized
141 compared to ePHAscl depolymerases (25, 33).

142 Most PHA biodegradation research has relied heavily on culture-based approaches or
143 black-box burial degradation trials (30). Culture-based research has developed the foundational
144 understanding of the enzymes and organisms mediating PHA biodegradation reviewed by (13).
145 However, a global summary of the environmental distribution of PHA depolymerases has not
146 been reported. Many microbial lineages are currently unamenable to culturing and are only
147 recognized from culture-independent analyses (35). With the advent of high-throughput
148 sequencing and metagenomics, there has been a proliferation of publicly available DNA
149 sequence data from microbial communities inhabiting diverse global environments (36, 37),
150 enabling broad surveys of trait distributions, see (38, 39). Here, we employed a culture-
151 independent approach using publicly available environmental sequence data to characterize the
152 phylogenetic and environmental distribution of PHA-degrading enzymes. We screened 5,290
153 metagenomes assembled genomes (MAGs) and 3,842 metagenomes sequenced from globally
154 distributed natural environments and waste-management systems. Biochemically characterized
155 ePHA depolymerases were used as biomarkers to assess the potential for PHA degradation
156 activity in different environments and across the bacterial tree of life.

157 **Methods**

158 *Identification of ePHA depolymerases*

159 A reference set of biochemically validated ePHA depolymerases were gathered from the
160 literature (Supp Data 1) and used as queries in a BLAST search (threshold e< $1x10^{-10}$) against

3,842 metagenomes (Supp Data 2) and 5,290 MAGs (Supp Data 3) that were publicly available on the Integrated Microbial Genomes (IMG) system (https://img.jgi.doe.gov) (37). Associated IMG metadata was used to classify each dataset by environment. Datasets screened were publicly available, but permissions were requested for PIs associated with the data. Many metagenomes came from datasets associated with publications, and we acknowledge the work of the scientific community in generating such a rich database (40–55). A series of curation steps were taken to increase PHA depolymerase prediction accuracy and to reduce or eliminate non-functional homologs. Hits identified in metagenomes and MAGs were curated separately. The results of the curation process applied to MAG hits (smaller, more tractable dataset) were used to develop a computationally feasible curation protocol for metagenome hits.

ePHA depolymerase hit curation:

Unique ePHA depolymerase blast hits (threshold < $1x10^{-10}$) were gathered and aligned with the corresponding reference sequences (7 ePHAmcl, 14 ePHAsclT1, and 7 ePHAsclT2) (Supp Data 4). All alignments were performed using MUSCLE v. 3.8.3 (56) as implemented in Geneious v. 10.2.6 (https://www.geneious.com) with default parameters. Hits were first sorted and divided into groups of similarly sized sequences before alignments were performed. This served to reduce the computational load of performing a single alignment with many dissimilar sequences and facilitated identification of areas of shared conservation with reference enzymes. An alignment-based curation was performed to identify hits that shared key catalytic residues with reference sequences (serine embedded in a lipase box motif [GxSxG], aspartic acid, and histidine, as well as the oxyanionic histidine for the ePHAscl depolymerases). To prevent removal of true PHA depolymerases containing insertions or deletions, higher scoring blast hits

(threshold e< 1x10-30) were retained for further analyses in addition to those retained from the alignment-based curation. Genes that were highly gapped in the alignment or which were truncated within the catalytic domain were removed. Next, placement within a maximum likelihood (ML) tree was used to distinguish PHA depolymerases from other types of closely related enzymes. ML trees were constructed using RAxML-HPC2 (57) with automatic bootstopping on the CIPRES webserver (58; Supp. Data 5). ModelTest-NG (59) was implemented for model selection assuming a uniform rate of evolution. PROTGAMMA WAG, VT+F, and WAG+F models were used for ePHAmcl, ePHAsclT1 and ePHAsclT2 hits respectively.

ML-based curation was performed using tree topology and branch length to assess relatedness to known PHA depolymerases. Representative MAG-encoded sequences for each major clade occurring in different regions of the ML tree were used as queries in a blast search against (1) the uniprot / swissprot database (https://www.uniprot.org/) (60); (2) NCBI's refseq database (https://www.ncbi.nlm.nih.gov/refseq/) (61); and (3) the pdb database (www.wwpdb.org) (62). Top blast hits were retrieved and used to predict clade annotations. Many MAG-encoded genes could not be confidently assigned to a PHA depolymerase subtype based on tree placement alone, due to overlapping phylogenetic signals between (1) ePHAsclT1 depolymerases and inPHAscl depolymerases and (2) ePHAsclT2 depolymerases and a periplasmic PHA depolymerase.

Ribosomal protein tree

206 A concatenated ribosomal protein tree was constructed to phylogenetically place MAGs with
207 predicted bioplastic degrading genes. 16 ribosomal proteins (rpL2, L3, L4, L5, L6, L14, L15,
208 L16, L18, L22, L24, S3, S8, S10, S17 and S19) were gathered from 143 MAGs encoding
209 predicted PHA depolymerases. Ribosomal proteins were independently aligned to a previously
210 described bacterial reference set (35). Alignments were masked to remove columns comprised of
211 95%+ gaps and trimmed to remove ambiguously aligned positions at the N and C termini. The
212 16 ribosomal protein alignments were concatenated to form a final alignment composed of 2,491
213 positions and 1,941 taxa. In tandem, MAG quality was estimated using checkM v1.0.13 (63).
214 Thirteen poor quality MAGs (>10% redundant and/or <50% complete) were excluded from the
215 final alignment. Fifteen MAGs were also excluded from the final alignment because their
216 ribosomal protein complement resulted in <50% of the expected number of ungapped residues in
217 the final alignment. A total of 115 MAGs encoding predicted PHA depolymerases were included
218 in the final alignment. Phylogenetic placement was inferred using FastTree (64) as implemented
219 in Geneious v.10.2.6. (https://www.geneious.com) and the tree was visualized using iTOL
220 (https://itol.embl.de/) (65).
221
222 *Environmental ePHA depolymerase hit curation*
223 Hits with an e-value <$1x10^{-25}$ for PHAmcl depolymerases (1,196) and <$1x10^{-30}$ for
224 ePHAscl (T1: 54,986, T2: 14,042) depolymerases were gathered and aligned to reference
225 sequences as for the MAG-derived genes. E-value thresholds were set at least five orders of
226 magnitude lower than the lowest quality hit that had ultimately been retained during the MAG hit
227 curation, to ensure a broad search. ePHAscl depolymerase hits missing either the lipase box
228 motif (GxSxG) and the oxyanion hole (HGC) motif were removed before alignments were

229 performed. Alignments and phylogenetic tree placement-based curation were performed as for
230 the MAG-derived sequences, with the following modifications to accommodate large dataset
231 sizes. High quality blast hits, defined as $<1x10^{-40}$ for ePHAmcl and $1<10^{-50}$ for ePHAscl datasets,
232 were retained for inclusion in the tree curation steps. CD-HIT (66) was used to cluster
233 ePHAsclT1 hits to 90% identity (word size 5) to reduce computational load of alignments and
234 tree-based curation. After tree-based curation, near-full-length sequences occurring within the
235 same clustered set for retained predicted ePHAsclT1 depolymerases were identified using a
236 custom script (Supp Data 6) and included in the final ePHAsclT1 sequence set.
237
238 Final sets of curated predicted ePHA depolymerases were independently aligned to the
239 relevant reference enzymes using MAFFT v.7 (67) (ePHAsclT1) and MUSCLE v. 3.8.3 (56)
240 (ePHAsclT2 and ePHAmcl). MAFFT was used for the final predicted ePHAclT1 alignment due
241 to large dataset size. Final protein trees were independently constructed for predicted ePHAmcl,
242 ePHAsclT1 and ePHAsclT2 depolymerases using FastTree (68) due to dataset size. Final trees
243 were visualized and annotated using ITOL (https://itol.embl.de/) (65).
244
245 **Results**
246 **Environmental distribution of PHA depolymerases**
247 To characterize the environmental capacity for PHA biodegradation, 3,842 publicly
248 available assembled and annotated metagenomes from diverse aquatic (n=1,982), terrestrial
249 (n=1,596), wastewater (n=230) and solid waste management systems (n=34) were screened for
250 ePHA depolymerases (Supp Data 2). Metagenomes sequenced from all seven continents were
251 included in our dataset but were unequally represented, with certain regions having more

252 sampled locations and sample types than others (Supp Fig 1). Biochemically validated ePHA
253 depolymerase genes were used as biomarkers for biodegradation potential (Supp Data 1).
254 Homology-based searches using biomarker enzymes followed by rigorous curation of predicted
255 enzymes resulted in a global distribution for PHA depolymerases (Fig 1a). Metagenomes from
256 North America were the most heavily represented in our dataset and nearly every environmental
257 subtype showed some evidence for PHA degradation activity within this region (Fig 1b). A total
258 of 13,869 PHA depolymerases were predicted in 1,295 metagenomes, sampled from aquatic
259 (417/1,982 = 21.0%), solid waste (23/34 =67.6 %), terrestrial (764/1,596 =47.8%) and
260 wastewater (91/230 = 39.5%) environments (Fig 2).
261
262 ePHAmcl depolymerization potential was sparsely represented across environments.
263 Only 222 ePHAmcl depolymerases were predicted out of the 13,869 ePHA identified. ePHAmcl
264 depolymerases were identified from 153 metagenomes from aquatic (41/1,982 = 2.1%),
265 terrestrial (94/1,596 = 5.9%), and wastewater (18/230 = 7.8%) environments (Fig 2). Predicted
266 ePHAmcl depolymerases were placed in a ML tree anchored with biochemically characterized
267 ePHAmcl enzymes (Supp Fig 2a). ePHAmcl depolymerases recently isolated from *Streptomyces*
268 species share little sequence homology with other ePHAmcl biomarker enzymes and form a
269 novel subgroup (69). Based on the ML tree, 14 predicted ePHAmcl depolymerases placed with
270 the three ePHAmcl depolymerases isolated from *Streptomyces* (69, 70) while the other 208
271 predicted enzymes formed a second clade with ePHAmcl biomarker enzymes isolated from
272 species of *Bdellovibrio* (34) and *Pseudomonas* (71, 72). Based on tree placement, only a handful
273 of predicted enzymes were closely related to biomarker enzymes; the majority of predicted

274 enzymes formed more distantly related clades. If functional, these enzymes significantly expand
275 the known sequence diversity of ePHAmcl depolymerases.
276
277 In contrast to ePHAmcl depolymerases, ePHAscl depolymerases were more prevalent
278 across environments. A total of 10,245 ePHAsclT1 depolymerases were predicted from 1,136
279 metagenomes, from aquatic (338/1,982 =17.1%), solid waste (13/34 = 38.2 %), terrestrial (713
280 /1,596 = 44.7%), and wastewater (72/230 = 31.3%) environments (Fig 2). For ePHAsclT2
281 depolymerases, a total of 3,402 enzymes were predicted from 908 metagenomes from aquatic
282 (264/1,982 = 13.3%), solid waste (22/34 = 64.7%), terrestrial (540/1,596 = 33.8%), and
283 wastewater (82/230 = 35.6%) environments (Fig 2). Predicted ePHAscl T1 and T2
284 depolymerases were placed in protein trees anchored with biomarker enzymes (Supp Fig 2b and
285 2c). For both predicted ePHAscl T1 and T2 depolymerases, considerable sequence diversity was
286 seen relative to biomarker enzymes, with large clades of putative PHA depolymerases containing
287 no biomarker enzymes. As for ePHAmcl enzymes, this indicates the known diversity of ePHAscl
288 enzymes may be a marked underrepresentation of the true protein family.
289
290 *Aquatic Environments*
291 Plastic contamination is ubiquitous in aquatic environments (6, 73). Plastic debris, either
292 macroplastics or microplastics (typically <0.5mm) can harm aquatic organisms through
293 entanglement and ingestion (74). Microplastics are of particular concern because they can be
294 ingested by organisms and serve as a vector for chemical contaminants (74). As bioplastics
295 become more popular, it is expected that they will become a larger component of aquatic
296 contamination. PHAs have been shown to degrade in aquatic environments, however rates can

297 vary widely depending on polymer properties, temperature, and microbial population, and can be
298 quite slow (20, 75, 76). For example, in freshwater ponds, samples of PHAscl showed 7% mass
299 loss after six months (20). In marine water, PHAscl bottles showed only minor disintegration
300 after a year (76). Relatively little research has characterized the degradation of PHAmcl in
301 aquatic environments, and PHAmcl degraders are typically scarce relative to PHAscl degrading
302 organisms (69, 77).
303
304 Here, we screened 1,982 aquatic metagenomes for evidence of microbial PHA
305 biodegradation potential. Our analyses included 946 freshwater, 452 marine, 87 non-marine
306 saline and alkaline, 114 sediment and 383 thermal spring metagenomes. Less than half of the
307 aquatic metagenomes we screened contained any type of predicted ePHA depolymerases (Fig 2).
308 The highest proportion of aquatic metagenomes with predicted PHA depolymerization potential
309 were from sediments (55/114 = 48.2%) followed by non-marine saline and alkaline (31/87=
310 35.6%) and marine environments (140/452=31.0%). Only 19.2% of freshwater metagenomes
311 (182/946) and 2.3% of thermal spring metagenomes (9/383) contained predicted ePHA
312 depolymerases.
313
314 Very few aquatic metagenomes showed evidence for ePHAmcl depolymerization
315 potential. Only 1.9% (18/946) freshwater, 3.5% (16/452) marine, 2.3% (2/87) non-marine saline
316 and alkaline, 3.5% (4/114) sediment, and 0.3% (1/383) thermal spring metagenomes had
317 predicted ePHAmcl depolymerases. In contrast to ePHAmcl, all aquatic environmental subtypes
318 had predicted ePHAscl depolymerization potential to some degree (ePHAsclT1 and ePHAsclT2
319 depolymerases). 43.9% (50/114) of sediment, 24.1% (21/87) of non-marine saline and alkaline,

20.4% (92/452) of marine, and 17.5% (166/946) of freshwater metagenomes contained predicted ePHAsclT1 depolymerases. Very few thermal spring metagenomes (9/383 = 2.34%) had these predicted depolymerases. For ePHAsclT2 depolymerases, the highest proportion of metagenomes encoding this activity was sediments (45/114 = 39.5%), followed by 21.7% (98/452) marine, 20.7% (18/87) non-marine saline and alkaline environments, and only 0.5% (2/383) thermal spring metagenomes. Thermal spring metagenomes had the lowest PHA degradation potential across all aquatic environmental subtypes.

In aquatic environments, plastics provide a surface for microbial colonization and biofilm formation (78). Few studies have characterized how microbial assemblages change and respond to bioplastic waste, including PHAs. Pinnell & Turner (2017) found that PHA-associated biofilms in benthic coastal marine environments were distinct from non-biodegradable polyethylene terephthalate controls and were dominated by sulfate-reducing microorganisms (79). These results indicate that PHA waste may alter microbial communities and have unintentional impacts on biogeochemical activities in this environment through the stimulation of sulfate reducers. Our analyses indicate that the majority of aquatic environments are not equipped with an abundance of PHA degrading organisms. This may suggest that bioplastics (particularly PHAmcl products) may persist longer than anticipated in these environments, or their presence may slowly enrich for bioplastic-degrading organisms, altering the local microbial community structure.

Terrestrial Environments

342 Soil microbial communities have been extremely well-sampled globally relative to other
343 environments. From our screen, 51.5% (740/1,437) of soil metagenomes show ePHA
344 depolymerization potential (Fig 2). Other terrestrial environments with evidence for this activity
345 were deep subsurface environments (21/143= 14.7%) and rock-dwelling communities (3/3 =
346 100.0%). Loam (0/3), peat (0/5) and volcanic (0/5) microbial communities showed no evidence
347 for ePHA depolymerization potential, although this may be an artifact of small sample sizes for
348 these environments.
349
350 Rock-dwelling metagenomes were not predicted to contain any ePHAmcl depolymerases,
351 whereas 0.7% (1/143) of deep subsurface and 6.5% (93/1,437) of soil metagenomes did encode
352 one or more predicted ePHAmcl depolymerases. 13.3% (19/143) deep subsurface, 100% (3/3)
353 rock-dwelling, and 48.1% (691/1,437) of soil metagenomes screened positive for ePHAsclT1
354 depolymerases. ePHAsclT2 depolymerases were predicted in 5.6% (8/143) of deep subsurface,
355 66.7% (2/3) of rock-dwelling environments, and 36.9% (530/1437) of soil metagenomes.
356
357 Recently, culture-independent approaches have revealed the diversity and heterogeneity
358 of soil microbial communities. Soil can have distinct microbial communities across micrometers
359 or millimeters in space (80). PHAscl biodegradation in soil has been well-studied and many PHA
360 degrading organisms, mostly polyhydroxybutyrate (PHB) degraders (the most common type of
361 PHA monomer), have been isolated from these environments (14, 17, 30, 81). Again, less work
362 has characterized PHAmcl degradation in soils relative to PHAscl products. However, in tropical
363 soils, PHAmcl has been showed to degrade much more slowly compared to PHB films (77).
364

365 *Solid Waste Environments*

366 Over half of metagenomes screened from solid waste environment subtypes contained
367 predicted ePHAscl depolymerases, with 66.7% (20/30) of compost metagenomes and 75.0%
368 (3/4) of landfill metagenomes showing ePHA depolymerization potential. When parsed by
369 ePHAscl depolymerase subtype, 33.3% (10/30) of compost and 75.0% (3/4) of landfill samples
370 had predicted ePHAsclT1 depolymerases and 63.3% (19/30) of compost and 75.0% (3/4) of
371 landfill metagenomes had predicted ePHAsclT2 depolymerases. Notably, our analyses did not
372 predict any ePHAmcl depolymerases from solid waste environments, which may have
373 implications for bioplastic waste-management.

374 Composting is considered the ideal way to handle certain types of biodegradable plastic
375 waste because it allows for energy and material recovery (31). Furthermore, since plastics are
376 broken down, they are removed as a potential physical threat to organisms in the environment.
377 Composting expands the end-of-life options available for bioplastics and is particularly
378 beneficial when mechanical recycling is not an option (31). However, composting requires
379 existing infrastructure and collection systems and access to these varies based on locale (82).
380 Commercial PHA products are certified as compostable under various standards, including the
381 European standard (EN 13432) which stipulates that under industrial composting conditions 90%
382 of the plastic must be converted to CO_2 after six months (31). It was concerning that a third of
383 composting samples showed no predicted PHA depolymerases. This may reflect a lack of
384 bioplastics in the current waste streams, where low abundance degraders are below current
385 detection limits, or may indicate a future requirement for targeted amendment of bioplastic
386 degrading organisms to composting facilities to ensure degradation occurs at mandated rates.
387

388 A large fraction of plastic waste is disposed of in landfills. Despite landfills being a likely
389 site of disposal, very little characterization of the biodegradation of biodegradable plastics in
390 landfills has taken place (82). Very few plastics are recycled due to low consumer compliance,
391 lack of infrastructure, and contamination of recycling streams (3). Bioplastics will likely follow
392 this precedent, although because they can be composted, a larger proportion may be diverted
393 away from landfills. Landfills are not an ideal solution for bioplastic waste; plastics in landfills
394 represent a lost opportunity as their stored energy cannot be recovered and they take up valuable
395 space (1, 83). Landfills are complex heterogenous environments and landfill microbial
396 communities are not well-characterized (84). The landfill metagenomes included in our analyses
397 originated from a Canadian site and showed the potential for PHAscl degradation, but not
398 PHAmcl degradation. It is thus possible that PHAmcl polymers will behave similarly to
399 conventional plastics and remain relatively inert over a long timeframe. This poses little threat to
400 the environment provided landfills are well-maintained and do not disperse wastes into the
401 surrounding environments, but this cannot be assumed for 100% of landfilled bioplastics. Our
402 identification of predicted ePHAscl depolymerases in landfill metagenomes coupled with data
403 from earlier studies indicates that ePHAscl plastics biodegrade in landfills if amenable abiotic
404 conditions (temperature, pH, moisture) are maintained. One concern is that anoxic degradation of
405 PHAs generates greenhouse gases in the form of CO_2 and CH_4 (24). However, most modern
406 landfills have infrastructure to capture methane as biogas, generating energy and reducing
407 greenhouse gas emissions, which would mitigate harmful by-products of landfill PHA
408 degradation (85).
409
410 *Wastewater Environments*

411 Within wastewater environments, the majority of anaerobic digestors (51/54 = 94.4%)
412 and nutrient removal communities (27/35=77.1%) were predicted to conduct ePHA degradation,
413 whereas only a small fraction of activated sludge communities showed evidence for this activity
414 (13/141=9.2%).
415 ePHAmcl depolymerases were predicted in 7.8% (11/141) of activated sludge, 18.5%
416 (10/54) of anaerobic digestors, and 57.1% (20/35) of nutrient removal systems. A high
417 proportion of anaerobic digestors (38/54=70.4%) and nutrient removal systems (21/35 = 60.0%)
418 had predicted ePHAsclT1 depolymerases whereas only 9.2% (13/141) of activated sludge
419 environments had predicted enzymes. ePHAsclT2 depolymerases were also well-represented in
420 anaerobic digestors (47/54 = 87.0%) and nutrient removal systems (27/35=77.1%), but only
421 occurred in 8/141, or 5.7% of activated sludge metagenomes.
422
423 Interestingly, wastewater can be used as a carbon source for PHA production , and
424 appears to be a promising avenue for biopolymer production (86). Given the potential presence
425 of PHA degraders in the majority of wastewater systems, there is the possibility of a circular
426 system, where wastewater acts as both feedstock and waste processing components of the
427 bioplastic lifecycle.
428
429 **Phylogenetic distribution of predicted PHA depolymerases**
430 We screened 5,290 MAGs available on IMG for ePHA depolymerases, comprising 4,927
431 bacterial and 363 archaeal genomes. Screened MAGs were reconstructed from metagenomes
432 originating from engineered (443), environmental (4,659), host-associated (186), and
433 unclassified (2) systems (Supp Data 3). We identified 143 MAGs (2.7%) encoding a total of 231

predicted extracellular PHA depolymerases. CheckM was used to assess MAG quality, with 130 MAGs >50% percent complete and <10% redundant, within which 122 MAGs were >70% complete and <10% redundant (Supp Data 7). All MAGs encoding predicted ePHA depolymerases were from the bacterial domain. A number of distantly related genes were identified in archaeal MAGs but these did not pass our curation steps. These archaeal sequences may represent true PHA depolymerases that are distantly related from reference enzymes and thus may be interesting targets for future characterization (Supp Data 8).

A concatenated ribosomal protein tree was created to phylogenetically place the 115 predicted ePHA degrading MAGs that were >50% complete and <10% redundant and had an appropriate complement of ribosomal proteins (see methods) (Fig 3). From this, 6 pairs of curated MAGs had identical concatenated ribosomal protein sequences. These MAG pairs often originated from related samples (e.g., time series samples of the same environment), indicating they may represent the same bacterial populations (Supp Data 7). From the curated set of 115 MAGs, we identified putative PHA depolymerases from members of the Proteobacteria (92), Bacteroidetes (14), Actinobacteria (3), Firmicutes (2), Candidate phylum Rokubacteria (2), Spirochaetes (1), and Planctomycetes (1). Notably, we did not detect PHA biodegradation potential in any MAGs associated with the Candidate Phyla Radiation/Patescibacteria. Prior to our screen, bacterial PHA depolymerases were reported from the Proteobacteria, the Actinobacteria and the Firmicutes, as well as one fungal phylum, the Ascomycetes. Our screen extends the phylogenetic distribution of this activity to four additional bacterial phyla, doubling its taxonomic diversity. The distribution of fungal PHA depolymerases was not investigated in

our screen, but is an important avenue for future characterization to holistically characterize PHA depolymerization potential.

The vast majority of the predicted PHA degrading organisms were Proteobacteria, with 13 Alphaproteobacterial, 51 Betaproteobacterial, 8 Deltaproteobacterial, 18 Gammaproteobacterial and two Oligoflexia MAGs identified. Alphaproteobacterial MAGs with ePHA depolymerases were from marine, freshwater, and soil environments, with members of the families Rhizobiales (7), Sphingomonadales (4), and Rhodobacterales (1), and two unclassified Alphaproteobacteria. Many Betaproteobacterial PHA degraders have been identified through culture-based studies, particularly from soils (13). As a result, many of the biomarker enzymes used in our screen were from this class (Supp Data 1 and references within). The Betaproteobacterial MAGs with predicted PHA degradative capacity were predominantly from the Burkholderiales (31), but also from the Rhodocyclales (2) and unclassified lineages (13). The Burkholderiales *Paucimonas lemoignei* is a model organism for PHA degradation and encodes at least five PHA depolymerases (87). It is not clear whether PHA depolymerases from this class were so abundant in our screen due to a detection bias based on the closely related biomarker enzymes in our reference set or because this family is genuinely enriched for this function compared to other lineages. Eight Deltaproteobacteria with PHA depolymerization potential were identified, including six Myxococcales, a Desulfobacterales and an unclassified bacterium. A total of 18 Gammaproteobacteria including species from the Alteromonadales (6), Chromatiales (1), Methylococcales (1), Oceanospirillales (3), Pseudomonadales (1), Xanthomonadales (2), and four unclassified organisms were predicted to encode ePHA depolymerases.

Beyond the Proteobacteria, 13 MAGs from the Bacteroidetes encoded predicted ePHAsclT1 depolymerase(s). All encoded a single predicted ePHAsclT1 except for two unclassified bacteria from soil and groundwater which each encoded two predicted ePHAsclT1 enzymes that shared high-level of sequence identities. Bacteroidetes with predicted ePHA degradation capacities included an unclassified Cryomorphaceae (Flavobacteria) and an unclassified Saprospiraceae (Saprospiria) from a wastewater nutrient removal system, and nine unclassified organisms from groundwater. The three Actinobacteria MAGs were predicted to encode ePHAsclT1 depolymerases. Two were unclassified members of the Pseudonocardiales from soil. One of these MAGs encoded a single ePHAsclT1 depolymerase whereas the other encoded two. The third Actinobacteria MAG was from an organism from the Actinomycetales. Two Firmicutes from the Clostridiales sequenced from an anerobic bioreactor were predicted to have PHAmcl depolymerization potential. One unclassified Spirochaetes bacterium from groundwater was predicted to encode three closely related ePHAsclT1 depolymerases. Two MAGs from the Candidate phylum Rokubacteria from groundwater and soil had predicted ePHAsclT1 depolymerases. One nutrient removal wastewater system-derived Planctomycetes MAG from the Phycisphaerales encoded a predicted ePHAsclT1. For a detailed overview of the PHA depolymerase complement for each MAG, please see Supp Data 7.

MAGs encoding multiple predicted PHA depolymerases

Of the 115 high-quality MAGs with PHA biodegradation potential, 67 were predicted to encode one PHA depolymerase, 24 were predicted to encode two, 23 were predicted to encode three, and one, a Burkholderiales, was predicted to encode nine PHA depolymerases (Supp Data 7).

The Burkholderiales MAG encoding nine putative PHA depolymerases originated from soil and was 98% complete and 5% redundant. Predicted depolymerases included two closely related ePHAsclT1, four ePHAsclT2 and three closely related ePHAmcl depolymerases.

MAGs from the Proteobacteria (22) and Spirochaetes (1) encoded three predicted ePHA depolymerases. One Alpha-, 18 Beta-, one Delta- and two Gammaproteobacteria all encoded three predicted PHA depolymerases. An Alphaproteobacterial MAG (100% complete, 1% redundant) that placed near the Bradyrhizobiaceae of the Rhizobiales encoded three closely related ePHAsclT1 enzymes. From the Betaproteobacterial MAGs, eight closely related *Candidatus* Accumulibacter organisms encoded two ePHAsclT2 enzymes and one ePHAsclT1 enzyme (all >84% complete, <5% redundant). PHA cycling is a core phenotype to the *Candidatus* Accumulibacter lineage, though most focus has been on PHA synthesis to date (49). Nine Burkholderiales encoded three ePHAscl enzymes and one Rhodocyclales (98% complete and 0% redundant) encoded one predicted ePHAsclT1 and two predicted ePHAsclT2 enzymes. For the Gammaproteobacteria, two near-identical *Halomonas* MAGs of the Oceanospirillales (both 99% complete, 1% redundant) encoded two ePHAsclT1 and one ePHAsclT2. The Spirochaetes MAG (90% complete, 0% redundant) had three predicted ePHAsclT1depolymerases that clustered closely within the ML tree.

Distribution of PHA depolymerase subtypes

ePHAmcl depolymerases: We predicted 10 ePHAmcl depolymerase genes from 8 MAGs (Fig 4a). A soil Burkholderiales bacterium from the Betaproteobacteria encoded 3 predicted enzymes. The remaining putative ePHAmcl depolymerases were found as singletons within MAGs, including a Pseudomonadales (Gammaproteobacteria) from soil, two Myxococcales

(Deltaprotebacteria) from a wastewater bioreactor and a marine environment, two *Sandaracinus* spp. (Deltaprotebacteria) from a marine environment, and a *Clostridium* (Firmicutes) from an anaerobic bioreactor (Fig. 4a). To our knowledge, ePHAmcl depolymerases have not previously been identified in the Firmicutes. Several *Clostridia* species are human pathogens and obligate anaerobes (88) which may have hindered recognition of PHA degrading activity in cultivation-based studies, which have largely been conducted aerobically. Based on ML analyses, the putative *Clostridium* ePHAmcl depolymerases were more closely related to the *Streptomyces* ePHAmcl depolymerase subgroup, whereas the other predicted PHAmcl enzymes were more closely related to those isolated from *Pseudomonas* and *Bdellovibrio* (Fig 4a).

ePHAsclT1 depolymerases: We predicted 156 ePHAsclT1 depolymerases in 120 unique MAGs from the Actinobacteria (7), Bacteroidetes (17), Candidate phylum Rokubacteria (2), Planctomycetes (1), Proteobacteria (125), Spirochaetes (3) and Tenericutes (1) (Fig 4b). ePHAsclT1 depolymerases are the most well-characterized type of PHA depolymerase (13, 24). Biochemically validated enzymes in our reference set originated from species of *Pseudomonas* (Gammaproteobacteria), *Ralstonia* (Betaprotebacteria), *Paucimonas* (Betaprotebacteria), *Alcaligenes* (Betaprotebacteria), and *Bacillus* (Firmicutes) (Supp Data 1 and references within). Our screen extends the distribution of ePHAsclT1 enzymes to a further six phyla. An ePHAsclT1 depolymerase was also identified from a Tenericute MAG, however due to MAG quality (49% complete, 0% redundant), it was excluded from the final concatenated ribosomal protein tree.

ePHAsclT2 depolymerases*:* We predicted 65 ePHAsclT2 depolymerases in 48 MAGs from members of the Alpha, Beta and Gammaproteobacteria (Fig 4c). Most ePHAsclT2 biochemically

548 validated enzymes originated from members of the Burkholderiales family (Betaproteobacteria).
549 Additional reference enzymes came from *Streptomyces* (Actinobacteria) and a *Penicillium*
550 (Ascomycota), neither of which had related sequences in our environmentally-derived
551 ePHAsclT2 depolymerases (Supp Data 1 and references within).
552
553 *Summary of environmental distribution data*
554 Screening of metagenomic sequence data from diverse aquatic, terrestrial and waste-
555 management environments revealed that, while PHA-degrading bacteria are globally distributed,
556 some environments show very little evidence for this activity. In particular, less than a third of
557 the thermal springs, freshwater and marine environments we screened showed evidence for
558 degradative capacity. This patchy distribution of PHA depolymerases may have impacts for
559 pollution mitigation. If biodegradable plastics do not degrade in the environment at an
560 appreciable rate, they have the potential to cause ecological damage analogous to conventional
561 plastics. Alternatively, the presence of bioplastic waste may select for low-abundance
562 community members with PHA degradation capacity and significantly alter microbial
563 communities, which may have downstream consequences on global biogeochemical cycling.
564
565 Notably, PHAmcl depolymerases were poorly represented in the global dataset,
566 indicating degradation of bioplastics with medium chain monomers may not be robust in most
567 environments. There are very few reports of isolated PHAmcl degraders (69). PHAmcls are not
568 easily emulsified into microbiological media, which may have hindered their detection in clear
569 zone assays (13). This is an unfortunate gap because the more flexible and less brittle properties
570 of PHAmcl make them more suitable than PHAscl for certain applications (89). Our screen

571 confirms the sparse environmental distribution of mcl PHA degrading organisms, with less than
572 5% of environmental samples showing evidence for PHAmcl degradation capacity. This suggests
573 that commercial PHAs may be substantially more recalcitrant to degradation than current
574 forecasts.
575
576 *Summary of phylogenetic distribution data*
577 We predicted PHA depolymerization activity for organisms from seven different phylum-
578 level lineages, where existing reference enzymes derive from only three bacterial phyla. The
579 majority of PHA depolymerases were identified from Proteobacteria, and specifically from the
580 Betaproteobacteria. The majority of MAGs encoded one or two ePHA depolymerases, with a
581 similarly low proportion of PHAmcl enzymes as were predicted from the environmental
582 metagenomes. Our ePHAmcl reference set originated from only two Proteobacterial classes and
583 the Actinobacteria. However, our analyses expanded the predicted distribution of PHAmcl
584 depolymerases to the Firmicutes as well as an additional class of Proteobacteria. One organism, a
585 Burkholderiales, encoded a complement of nine ePHA depolymerases, suggesting it may be an
586 active, efficient, and/or highly versatile degrader.
587
588 **Conclusions**
589 This work presents the first summary of the global environmental and phylogenetic
590 distribution of ePHA depolymerases using cultivation-independent approaches. Our analyses
591 present a carefully curated set of predicted PHA depolymerases which provide a conservative
592 estimate of the existing sequence diversity and environmental distribution of these protein
593 families. Our screen predicts that many environments are not well-suited for PHA degradation,

which is a concern as this class of bioplastics increases in production. To fully determine bioplastic fate in global environments, a combination of metagenomic predictions from public datasets, burial degradation trials with functional profiling, and biochemical validation of predicted enzymes (particularly in enzymes distantly related to biomarker enzymes) is needed. This comprehensive examination would help establish the true diversity of functional PHA depolymerases and allow for more accurate activity predictions in the future.

Our analyses provide insights into the bioplastic degradation potential of microbial communities found around the world in unique and relevant environments. Understanding the limitations for bioplastic degradation in global environments is critical to ensure that the proper waste-management solutions are developed and enforced. Biodegradable plastics that meet compost certifications degrade under specified conditions that may not be present in a natural environment, and therefore rates of biodegradation may be highly variable. It is important that consumers do not assume that biodegradable plastics can be littered into the environment without consequence. Polymeric properties such as monomeric composition, surface area, and crystallinity strongly influence PHA biodegradability, but the underlying microbial community and the enzymes they secrete is critical for mediating bioplastic depolymerization (1, 13). The key bioplastic degraders occurring in the environment may aid in the development of future bioremediation options for these plastics and may include enzymes and organisms suitable for use as waste-management tools in the processing, degradation or chemical recycling of these plastics.

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Figures

Figure 1: 13,869 PHA depolymerases were predicted in 1,295 globally distributed metagenomes from aquatic (417), terrestrial (764), solid waste (23) and wastewater (91) environments. **A)** World map showing geocoordinates of samples with metagenomes containing one or more predicted PHA depolymerase, colored by environmental subtype. **B)** Detailed view of North America, which had many samples predicted to contain PHA depolymerases.

Figure 2: Environmental distribution of 13,869 PHA depolymerases (222 ePHAmcl, 10,245 ePHAsclT1 and 3,402 ePHAsclT2 depolymerases) identified by screening 3,842 metagenomes. Graph depicts the percentage of metagenomes that contained one or more predicted PHA depolymerase.

Figure 3: The predicted phylogenetic distribution of PHA degrading bacteria. Two concatenated ribosomal protein trees inferred using FastTree show the placement of 115 high-quality MAGs

641 encoding predicted PHA depolymerases within the bacterial domain. **A)** All predicted PHA
642 degrading organisms place within select phyla in a shared region (purple) of bacterial diversity,
643 excluding the Candidate Phylum Radiation/Patescibacteria. **B)** Detailed phylogeny of the purple
644 subset of bacteria from (A) containing all MAGs encoding predicted ePHA depolymerases.
645 Phyla containing members with predicted PHA depolymerases are indicated by the coloured
646 inner ring, and the number of predicted scl and mcl PHA depolymerases in each MAG are
647 indicated by the size of the circles in the outer rings.
648
649 **Figure 4:** Environmental origin and phylogenetic relationships of the 231 PHA depolymerases
650 predicted from 143 MAGs. Outer rings are colored by environmental origin of predicted PHA
651 depolymerases. Inner rings are colored by phylum of the MAG encoding the predicted PHA
652 depolymerase. ML trees showing **A)** 10 (8 unique) ePHAmcl depolymerases **B)** 156 (133
653 unique) ePHAscl T1 depolymerases and **C)** 65 (56 unique) ePHAscl T2 depolymerases.
654
655 **Supplementary Figures**
656
657 **Supplementary Figure 1:** World map showing geocoordinates of sampling locations for the
658 3,842 metagenomes screened in this analysis, colored by environmental subtype. Several
659 samples share geocoordinates and represent samples taken at different depths, samples taken
660 within close proximity to each other, or at different time points.
661
662 **Supplementary Figure 2:** Environmental origin and phylogenetic relationships of the 13,869
663 predicted PHA depolymerases identified in 1,295 metagenomes. Outer colored ring indicates the

environmental origin of the metagenome encoding each predicted PHA depolymerase. Biomarker enzymes are indicated with black arrows. Protein trees were inferred with FastTree and show the sequence diversity of A) 222 predicted ePHAmcl depolymerases B) 10,245 predicted ePHAsclT1 depolymerases and C) 3,402 predicted ePHAsclT2 depolymerases, each with biochemically confirmed biomarker proteins included as references.

Supplementary Data Files

Supplementary Data 1: Supp_Data_1_PHA_depolymerases_references.xls

Supplementary Data 2: Supp_Data_2_PHA_depolymerase_reference_set_191118.fasta

Supplementary Data 3: Supp_Data_3_3842_IMG_metadata.xls

Supplementary Data 4: Supp_Data_4_5290_IMG_MAGs_metadata.xls

Supplementary Data 5: Supp_Data_5_final_trees_in_newick_format.txt

Supplementary Data 6: Supp_Data_6_cd_hit_uncluster_using_lengths.txt

Supplementary Data 7: Supp_Data_7_PHA_depolymerase_MAGs_CheckM_env.xls

Supplementary Data 8: Supp_Data_8_unique_archaeal_IMG_MAG_hits_e10.xls

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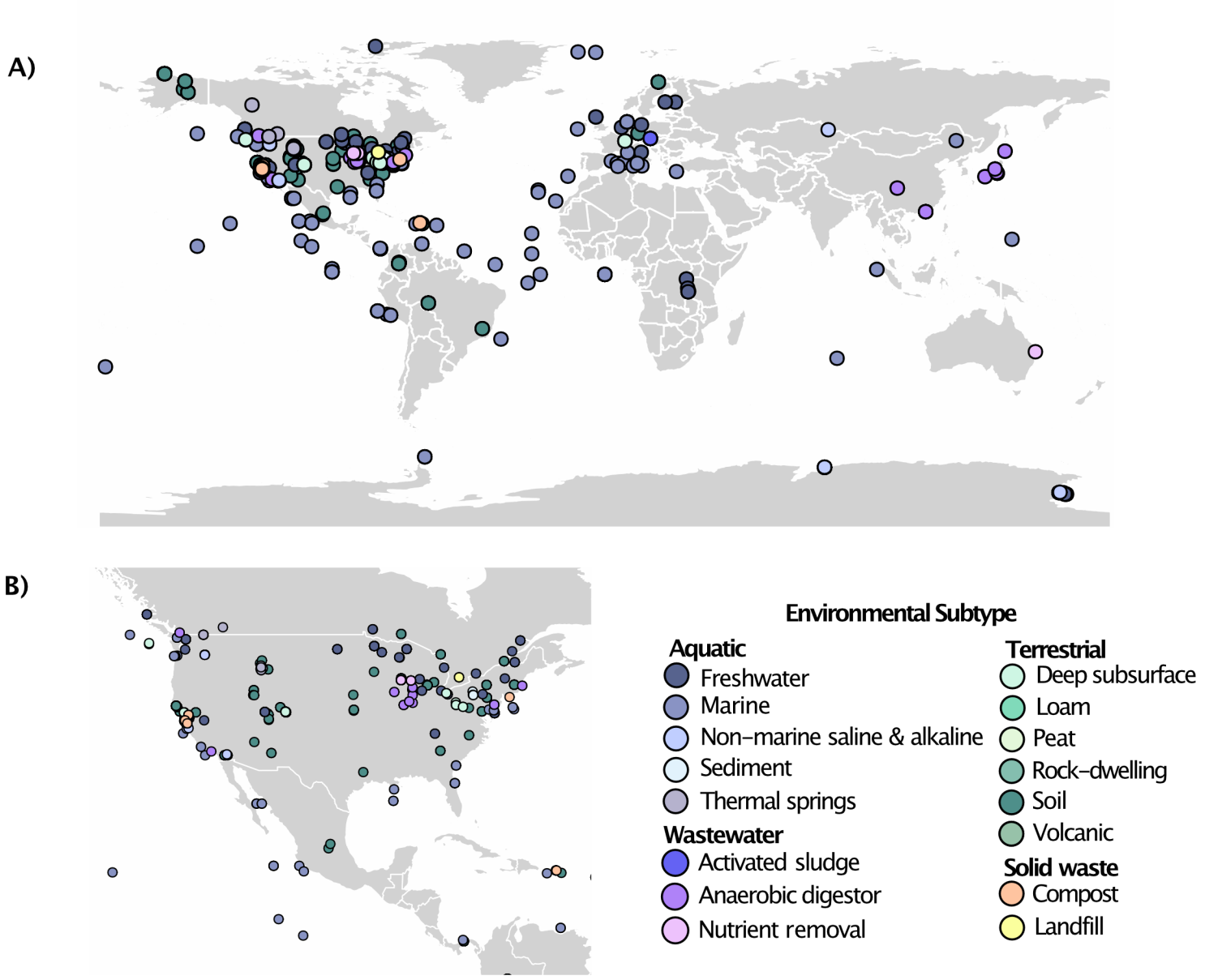

Fig 1

Metagenomes with predicted PHA depolymerase(s) (%)

Environment	PHAsclT1	PHAsclT2	PHAmcl	Total	Metagenomes (n)
Aquatic					
Freshwater	18%	11%	2%	19%	946
Marine	20%	22%	4%	31%	452
Saline and Alkaline	24%	21%	2%	36%	87
Sediment	44%	39%	4%	48%	114
Thermal springs	2%	1%	0%	2%	383
	17%	**13%**	**2%**	**21%**	**1,982**
Solid Waste					
Compost	33%	63%	0%	67%	30
Landfill	75%	75%	0%	75%	4
	38%	**65%**	**0%**	**68%**	**34**
Terrestrial					
Deep subsurface	13%	6%	1%	15%	143
Loam	0%	0%	0%	0%	3
Peat	0%	0%	0%	0%	5
Rock-dwelling	100%	67%	0%	100%	3
Soil	48%	37%	6%	51%	1,437
Volcanic	0%	0%	0%	0%	5
	45%	**34%**	**6%**	**48%**	**1,596**
Wastewater					
Activated Sludge	9%	6%	5%	9%	141
Anaerobic digestor	70%	87%	13%	94%	54
Nutrient removal	60%	77%	11%	77%	35
	31%	**36%**	**8%**	**40%**	**230**
Total	**30%**	**24%**	**4%**	**34%**	**3,842**

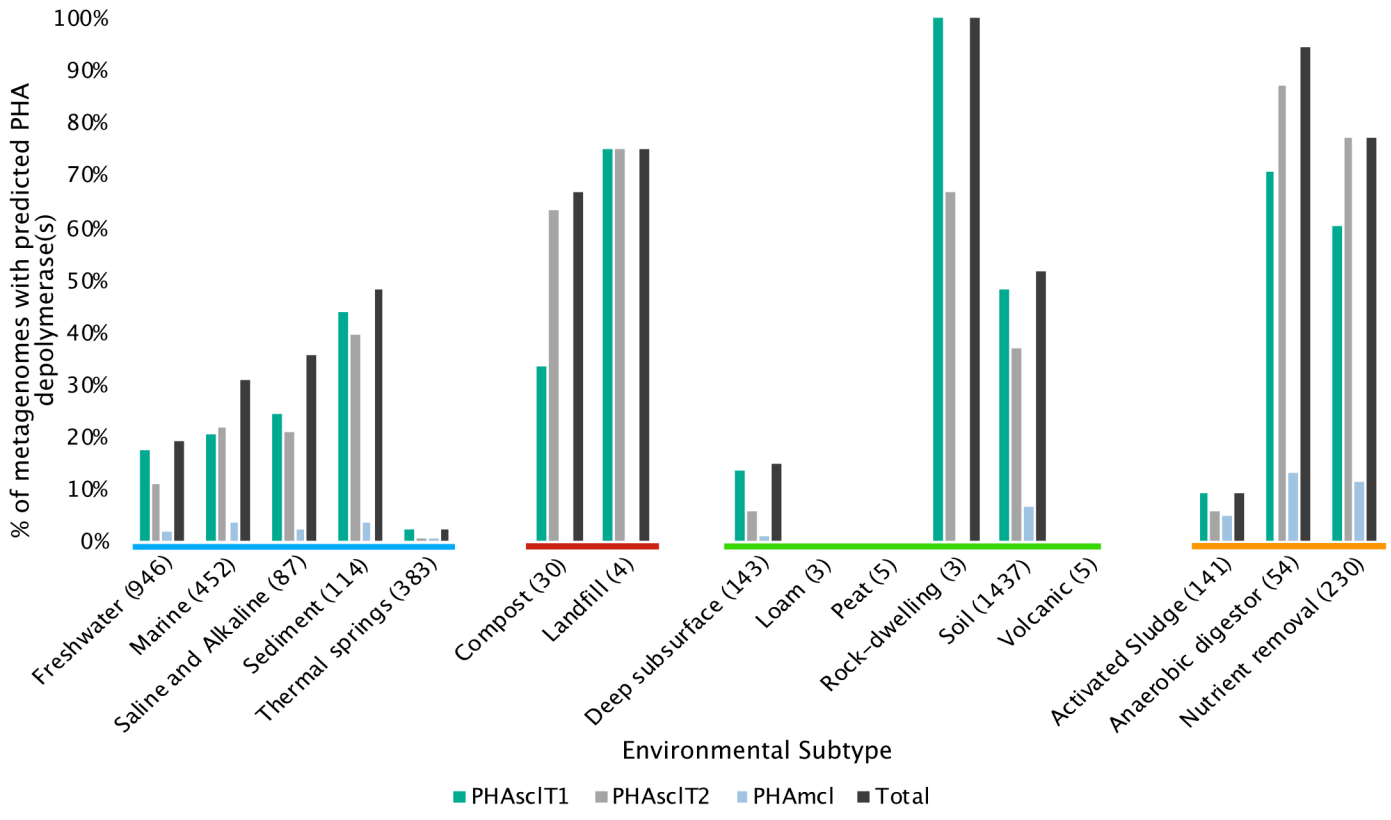

Fig 2

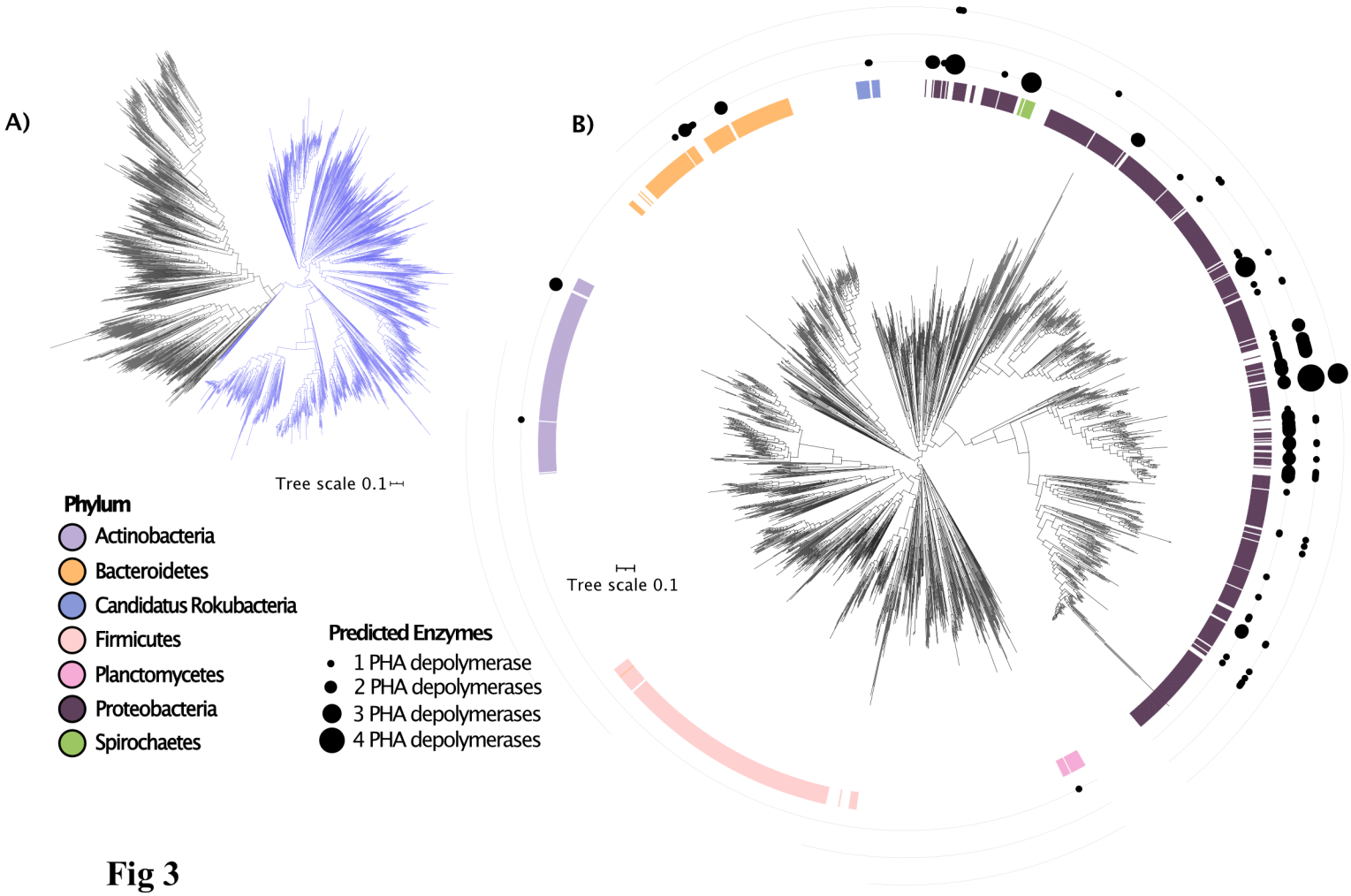

Fig 3

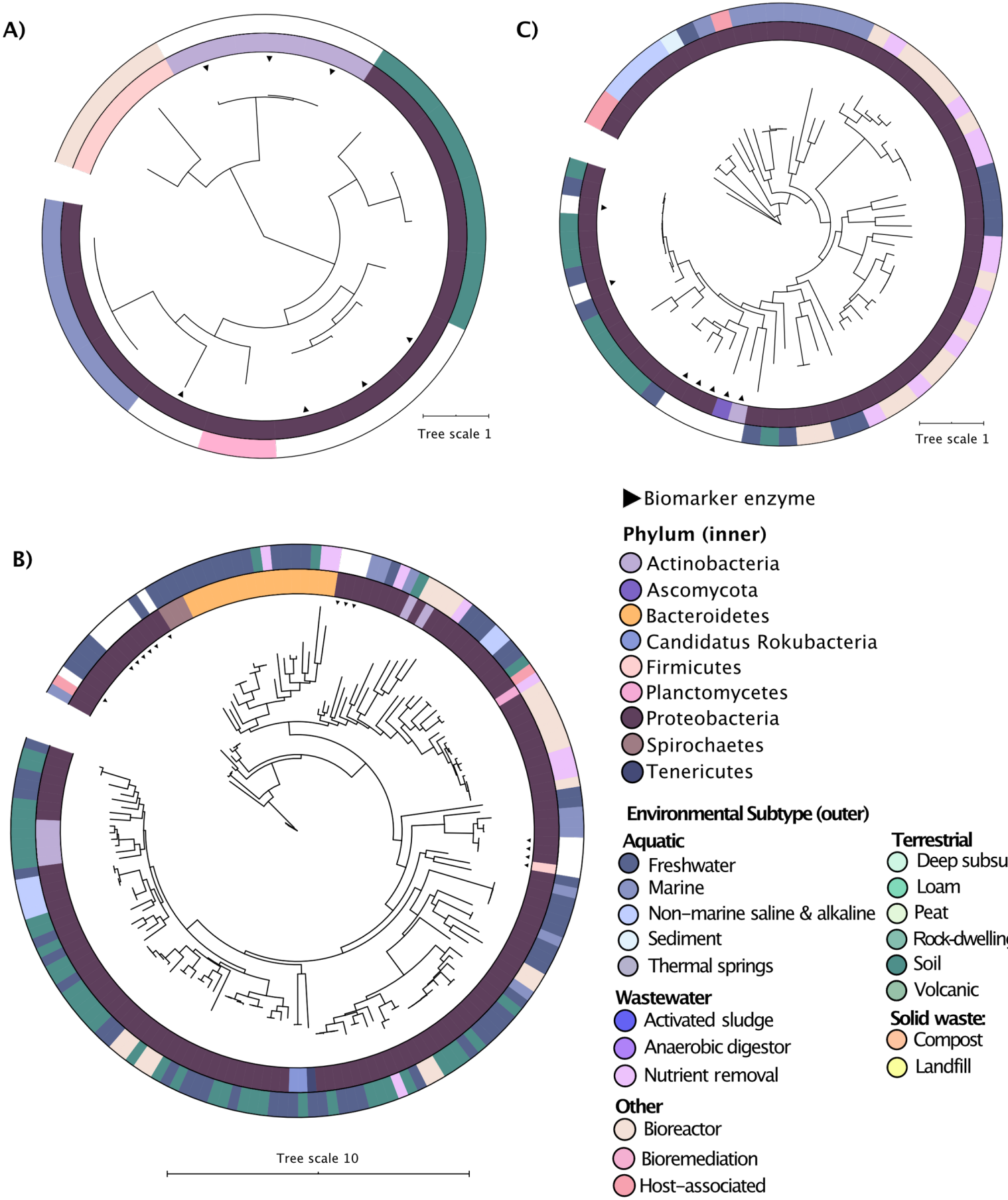

Fig 4