# A general approach for analysis of physiologically structured population models: the R package 'PSPManalysis'

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# Summary

- How environmental conditions affect the life history of individual organisms and how these effects translate into dynamics of population and communities on ecological and evolutionary time scales is a central question in many eco-evolutionary studies.
- 2. Physiologically structured population models (PSPMs) offer a theoretical approach to address such questions as they are built upon a function-based model of the life history, which explicitly describes how life history depends on individual traits as well as on environmental factors. PSPMs furthermore explicitly account for population feedback on these environmental factors, which translates into density-dependent effects on the life history. PSPMs can thus capture life histories in quite some detail but lead to population-level formulations in terms of partial differential equations that are generally hard to analyse.
- 3. Here I present a general methodology and a R software package for computing how the ecological steady states of PSPMs depend on model parameters and to detect bifurcation points in the computed curves where dynamics change drastically. The package makes specifying the population model unnecessary and only requires a relatively straightforward implementation of the life history functions as input. It furthermore allows for analysing the evolutionary dynamics and evolutionary singular states of the PSPMs based on Adaptive Dynamics theory.
- 4. Given the central role of the individual life history in many studies, there is substantial scope for using the presented methodology in fields as diverse as ecology, ecotoxicology, conservation biology and evolutionary biology, where it has already been applied to problems like the evolution of cannibalism, niche shifts and metamorphosis.

# Keywords

Steady state computations, bifurcation analysis, life history, demography, environmental feedback, evolutionary analysis, adaptive dynamics

# Introduction

The life history of individual organisms plays a central role in ecology and evolution, determining the demography of populations and thereby their persistence and existence. Together with the interactions with other species it shapes the dynamics of interacting populations and communities and through mutation and selection it leads to evolutionary change in species traits. Methodologies to assess how life history characteristics translate into consequences at the population level, such as population growth rate, are hence a core part of many ecological and evolutionary studies.

Modelling approaches that describe population and community dynamics explicitly on the basis of in-7 dividual life histories are referred to as structured-population models (Tuljapurkar & Caswell, 1997). 8 Matrix models (Caswell, 2001) are the most common type of structured-population models. Matrix 9 models describe the population dynamics in discrete time, as do integral projection models (IPMs, Ell-10 ner, Childs, & Rees, 2016). Structured-population models describing dynamics in continuous time 11 include stage- (Nisbet & Gurney, 1983) and size-structured population models (Sinko & Streifer, 1969; 12 de Roos, Metz, Evers, & Leipoldt, 1990), which both belong to the more general class of physiolog-13 ically structured population models (PSPMs, Metz & Diekmann, 1986; de Roos, 1997). Although all 14 structured-population models explicitly include a modeled representation of individual life history, they 15 differ in the way they account for this life history (de Roos, 2020). Matrix models and IPMs are primar-16 ily data driven. IPMs, for example, are formulated using (non-)linear relationships that result from fitting 17 observational data on the scaling of somatic growth rate and fecundity with individual body size (Rees, 18 Childs, & Ellner, 2014). In contrast, PSPMs (Metz & Diekmann, 1986; de Roos, 1997) are formulated 19 using a function-based model of the individual life history, which also accounts for the effect of envi-20 ronmental variables, such as food availability and predator density, on this life history (de Roos, 2020). 21 For example, individual foraging, growth and reproduction are in many PSPMs described by a dynamic 22 energy budget model for individual energetics (Kooijman, 2010). In turn, PSPMs account for how these 23 environmental variables are impacted by the population as a whole. PSPMs hence capture with more 24 mechanistic detail how individual-level processes, like energetics, together with the interactions of the 25 individual with its environment shape the life history and how the feedback of the entire population on 26 this environment has a density-dependent impact on that life history. PSPMs are therefore especially 27

suited to analyse how particular mechanisms or aspects of the life history or ecology of an individual
would affect the population and community dynamics.

The downside of the increased mechanistic detail of PSPMs is their mathematical tractability (de Roos, 30 2020). Where linear algebra offers a rich tool set to analyse matrix models and IPMs, simple PSPMs 31 are formulated in terms of the more daunting partial differential equations (Metz & Diekmann, 1986; 32 de Roos, 1997). Fortunately, though, a collection of numerical techniques is now available that allows 33 for analysing the ecological and evolutionary dynamics of even fairly complicated PSPMs (Hin & de 34 Roos, 2019a; ten Brink, de Roos, & Dieckmann, 2019; Chaparro Pedraza & de Roos, 2020). The aim 35 of this paper is to provide an introduction to these techniques and to the R package 'PSPManalysis' 36 implementing them. This toolbox includes techniques for the demographic and steady-state analysis 37 (Diekmann, Gyllenberg, & Metz, 2003; de Roos, 2008) of PSPMs, which also allow for the analysis 38 of evolutionary dynamics of PSPMs, based on the framework of 'Adapative Dynamics' (Dieckmann 39 & Law, 1996; Metz, Geritz, Meszéna, Jacobs, & van Heerwaarden, 1996; Geritz, Kisdi, Meszéna, & 40 Metz, 1998). In addition, the 'PSPManalysis' package includes the 'Escalator Boxcar Train' (de Roos, 41 Diekmann, & Metz, 1992), a numerical integration technique specifically developed for PSPMs. 42

To use the 'PSPManalysis' package it is not necessary to bother with the population-level representation 43 of the model in terms of partial differential equations or the like. The user can concentrate on the life 44 history and the ecology of the individual organisms. The necessary ingredients of the model specification 45 are conceptually straightforward as they include (i) the individual state variables (traits), such as age or 46 body size, that determine the life history; (ii) the environmental variables, such as food availability or 47 predation pressure, that impinge on and shape the life history; (iii) the rates of development, reproduction 48 and mortality dependent on these individual state and environmental variables; (iv) how an individual 49 impacts its environment; and (v) the conditions determining that the environment of the individuals is in 50 steady-state. The routines implemented in the 'PSPManalysis' package take these individual life history 51 ingredients as input and numerically analyse their population-level consequences, both in ecological and 52 evolutionary time. 53

# 54 Materials and methods

#### **55** The individual life history model

To illustrate the analysis of ecological steady states and the evolutionary analysis of PSPMs with the 56 'PSPManalysis' package I use the life-history model described in Chaparro Pedraza and de Roos (2020) 57 as an example. This model is loosely based on the life history of salmon with individuals starting their 58 life in a safe nursery habitat, in which they are protected from predation but suffer from competition for 59 resources. At some point during their life the individuals switch to a more risky growing habitat, where 60 competition for resources is absent, but individuals are exposed to predation mortality. All equations 61 occurring in the model are presented in Table 1, while Table S1 in the supporting information lists all 62 parameters with their default values. 63

Individuals are in the model characterised by their length  $\ell$  and their population, to which I will refer 64 to as the 'consumer' population, is hence length-structured. Migration to the growing habitat and mat-65 uration occur on reaching threshold body sizes, at length  $\ell = \ell_s$  and  $\ell = \ell_m$ , respectively. Feeding on 66 resources, growth in body size and reproduction are in the model described by the dynamic energy bud-67 get (DEB) model developed by Jager, Martin, and Zimmer (2013). In the nursery habitat the consumers 68 compete for a shared resource X at a rate that is proportional to their squared length and to the scaled 69 functional response value X/(K+X) (Table 1). In the growth habitat competition is assumed negligi-70 ble and individuals have *ad-libitum* food. The DEB model predicts individuals under constant resource 71 densities to grow in length following a vonBertalanffy growth curve with ultimate body length equal to 72  $\ell_{inf}X/(K+X)$  and  $\ell_{inf}$  in the nursery and growth habitat, respectively (Table 1). Reproduction only 73 occurs after individuals have migrated to the growth habitat since  $\ell_s < \ell_m$ . Following the DEB model 74 adult fecundity is proportional to squared individual length. 75

<sup>76</sup> Consumer mortality in the nursery habitat is assumed constant, while in the growth habitat mortality is <sup>77</sup> negatively size-dependent, proportional to  $\ell^{-d}$  (Table 1). In Chaparro Pedraza and de Roos (2020) size-<sup>78</sup> dependent mortality in the growth habitat is assumed non-dynamic. Here I assume it to be proportional <sup>79</sup> to the density of a dynamic, unstructured predator population that forages on consumers of different <sup>80</sup> length following a linear functional response and that experiences a constant mortality rate  $\mu_p$ . In the <sup>81</sup> model this predator population is represented with its scaled density, which incorporates the (constant) <sup>82</sup> conversion efficiency between ingested biomass of consumers and the predator's numerical response, <sup>83</sup> indicated with *B* (Table 1). The contribution of individual consumers to predator intake equals the <sup>84</sup> product of their vulnerability to predation and their biomass  $\ell^3$ . Finally, following Chaparro Pedraza <sup>85</sup> and de Roos (2020) turn-over of the resource in nursery habitat in the absence of consumers is described <sup>86</sup> by a semi-chemostat growth equation.

This tritrophic interaction between a resource and a size-structured consumer population in a nursery 87 habitat, which goes through a habitat shift during its life history and subsequently supports a specialist 88 predator population in the growth habitat, is fully determined by 3 life history functions, describing 89 development, reproduction and mortality ( $\gamma(\ell, X), \beta(\ell)$ , and  $\mu(\ell, P)$  in Table 1, respectively), 2 functions 90 describing the impact of differently sized consumer individuals on their environment through foraging on 91 the resource and contribution to predator food intake ( $\alpha(\ell, X)$  and  $\varepsilon(\ell)$ , respectively), and by 2 functions 92 that determine the dynamics of the shared resource in the nursery habitat and the predator in the growth 93 habitat (g(R) and h(B), respectively; Table 1).94

#### 95 General methodology

In the context of PSPMs Metz and Diekmann (1986) introduced the fundamental distinction between the 96 individual and its environment with accompanying state concepts. The crucial aspect of this distinction 97 is that the individual life history is fully determined by the state of the individual in combination with the 98 state of its environment. Given a constant environment all individuals are therefore independent, which 99 implies that in PSPMs all forms of density dependence operate through the environment. As a further 100 consequence, the individual life history functions are the only necessary ingredients for the computation 101 of population equilibrium states. I will shortly present the general methodology for computation of 102 equilibria in PSPMs here using the tritrophic model introduced above, but it should be stressed that it is 103 straightforward to generalise the methodology to far more complex PSPMs (see Diekmann et al., 2003, 104 for a detailed discussion). 105

Given a constant, equilibrium resource density  $\tilde{X}$  in the nursery habitat the individual length in equilibrium  $\tilde{\ell}(a, \tilde{X})$  as a function of age *a* and resource density  $\tilde{X}$  is the solution of the ordinary differential

equation (ODE):

$$\frac{d\tilde{\ell}}{da} = \gamma \left( \tilde{\ell}(a, \tilde{X}), \tilde{X} \right) \tag{1}$$

with initial condition  $\tilde{\ell}(0,\tilde{X}) = \ell_0$ . Similarly, denoting the equilibrium predator density in the growth habitat as  $\tilde{P}$ , the probability for an individual to survive up to age a, which I indicate with  $\mathscr{F}(a,\tilde{X},\tilde{P})$ , is the solution of the ODE:

$$\frac{d\mathscr{F}}{da} = -\mu\left(\tilde{\ell}(a,\tilde{X}),\tilde{P}\right)\mathscr{F}\left(a,\tilde{X},\tilde{P}\right)$$
(2)

with initial condition  $\mathscr{F}(0, \tilde{X}, \tilde{P}) = 1$ . Notice that survival depends on both resource and predator density in equilibrium, as the resource density determines how quickly individual consumers grow and hence how long they experience low mortality in the nursery habitat and the predator density influences their survival in the growth habitat.

The expected reproduction rate by a consumer individual at a particular age equals its fecundity times the probability it survives up to that age. Accumulating these reproductive contributions by integration over all possible ages that individuals can reach results in the following expression for the expected number of offspring produced by a single consumer individual throughout its lifetime, indicated with  $R_0$ , as a function of resource density  $\tilde{X}$  and predator density  $\tilde{P}$ :

$$R_0\left(\tilde{X},\tilde{P}\right) = \int_0^\infty \beta\left(\tilde{\ell}(a,\tilde{X})\right) \mathscr{F}\left(a,\tilde{X},\tilde{P}\right) da \tag{3}$$

Obviously, the equilibrium state of the size-structured consumer population is determined by the condition  $R_0(\tilde{X}, \tilde{P}) = 1$ , implying that every newborn consumer is expected to just replace itself during life.

For the resource density in the nursery habitat to be in equilibrium the resource turn-over should balance total resource consumption by all consumers in the nursery habitat. The latter equals the product of the amount of resources that an individual consumer is expected to consume during its life time and the consumer population birth rate, which I will indicate with  $\tilde{b}$ . The expected lifetime consumption by an individual is an integral similar to the expression above for  $R_0$  but involving the foraging rate  $\alpha(\ell, X)$ as opposed to the fecundity  $\beta(\ell)$ . The steady-state condition for the resource is hence given by the

condition:

$$G\left(\tilde{b},\tilde{X},\tilde{P}\right) = g(\tilde{X}) - \tilde{b} \int_0^\infty \alpha\left(\tilde{\ell}(a,\tilde{X}),\tilde{X}\right) \mathscr{F}\left(a,\tilde{X},\tilde{P}\right) da = 0 \tag{4}$$

in which the integral represents the expected lifetime resource intake by a single consumer.

Lastly, the predator population in the growth habitat is in steady state when its numerical response *B* equals its per-capita mortality rate  $\mu_p$ . Given the scaling of the predator population density such that its numerical response equals its functional response, the quantity *B* equals the product of the consumer population birth rate  $\tilde{b}$  and the expected amount of biomass that a consumer individual during its lifetime contributes to the per-capita food intake rate of the predator. The latter is given by an integral similar to the expression for  $R_0$  in equation (3) but involving the function  $\varepsilon(\ell)$  as opposed to the fecundity  $\beta(\ell)$ . The steady-state condition for the predator is given by the condition:

$$H\left(\tilde{b},\tilde{X},\tilde{P}\right) = \tilde{b}\int_{0}^{\infty} \varepsilon\left(\tilde{\ell}(a,\tilde{X})\right)\mathscr{F}\left(a,\tilde{X},\tilde{P}\right)da - \mu_{p} = 0$$
(5)

The integral in the above expression represents the expected lifetime contribution by a consumer to the food intake of a single predator.

Even though the ODE (1) for the growth in length is in the current model sufficiently simple to allow for 116 an explicit expression for the length at age  $\ell(a, \tilde{X})$  in equilibrium, analytical evaluation of the integrals 117 in the expressions for  $R_0(\tilde{X}, \tilde{P}), G(\tilde{b}, \tilde{X}, \tilde{P})$  and  $H(\tilde{b}, \tilde{X}, \tilde{P})$  is not possible because of the dependence of 118 consumer mortality on length. Hence, steady states of the PSPM can only be computed by solving the 119 equilibrium conditions (eqs. (3), (4) and (5)) numerically and iteratively for the unknown variables  $\tilde{b}$ , 120  $\tilde{X}$  and  $\tilde{P}$ . Solving such a system of (non-linear) equations can be achieved by standard methods, such 121 as the Newton-Raphson method (Press, Flannery, Teukolsky, & Vetterling, 1988), but for the fact that 122 it is impossible to derive explicit expressions for the integrals in the functions  $R_0(\tilde{X}, \tilde{P}), G(\tilde{b}, \tilde{X}, \tilde{P})$  and 123  $H(\tilde{b}, \tilde{X}, \tilde{P})$  even in a model as simple as the one discussed here. Definitely, the same holds for more 124 complex PSPMs as well. 125

The key idea to address this issue, originally proposed by Kirkilionis et al. (2001), is to consider the integrals occurring in the equilibrium conditions as a function of the upper limit of the integration and

define the following functions:

$$\begin{split} R_0(a,\tilde{X},\tilde{P}) &= \int_0^a \beta\left(\tilde{\ell}(\zeta,\tilde{X})\right) \mathscr{F}(\zeta,\tilde{X},\tilde{P}) \, d\zeta \\ I_1(a,\tilde{X},\tilde{P}) &= \int_0^a \alpha\left(\tilde{\ell}(\zeta,\tilde{X}),\tilde{X}\right) \mathscr{F}(\zeta,\tilde{X},\tilde{P}) \, d\zeta \\ I_2(a,\tilde{X},\tilde{P}) &= \int_0^a \varepsilon\left(\tilde{\ell}(\zeta,\tilde{X})\right) \mathscr{F}(\zeta,\tilde{X},\tilde{P}) \, d\zeta \end{split}$$

The value of these integrals can then be computed by numerically integrating the system of ODEs:

$$\begin{cases}
\frac{d\ell}{da} = \gamma(\tilde{\ell}(a,\tilde{X}),\tilde{X}) & \tilde{\ell}(0,\tilde{X}) = \ell_{0} \\
\frac{d\mathscr{F}}{da} = -\mu(\tilde{\ell}(a,\tilde{X}),\tilde{P})\mathscr{F}(a,\tilde{X},\tilde{P}) & \mathscr{F}(0,\tilde{X},\tilde{P}) = 1 \\
\frac{dR_{0}}{da} = \beta(\tilde{\ell}(a,\tilde{X}))\mathscr{F}(a,\tilde{X},\tilde{P}) & R_{0}(0,\tilde{X},\tilde{P}) = 0 \\
\frac{dI_{1}}{da} = \alpha(\tilde{\ell}(a,\tilde{X}))\mathscr{F}(a,\tilde{X},\tilde{P}) & I_{1}(0,\tilde{X},\tilde{P}) = 0 \\
\frac{dI_{2}}{da} = \varepsilon(\tilde{\ell}(a,\tilde{X}))\mathscr{F}(a,\tilde{X},\tilde{P}) & I_{2}(0,\tilde{X},\tilde{P}) = 0
\end{cases}$$
(6)

for the length  $\tilde{\ell}(a,\tilde{X})$  at age *a*, survival  $\mathscr{F}(a,\tilde{X},\tilde{P})$ , expected cumulative reproduction  $R_0(a,\tilde{X},\tilde{P})$ , expected cumulative resource ingestion  $I_1(0,\tilde{X},\tilde{P})$  and expected biomass contribution to the predator food intake  $I_2(0,\tilde{X},\tilde{P})$  of a consumer individual up to age *a* (The ODEs for these last 3 quantities are derived by differentiating their integral expressions with respect to *a*). Using these quantities the steady-state conditions of the PSPM can be expressed as:

$$\begin{cases} R_0(\infty, \tilde{X}, \tilde{P}) = 1\\ \tilde{b}I_1(\infty, \tilde{X}, \tilde{P}) = g\left(\tilde{X}\right)\\ \tilde{b}I_2(\infty, \tilde{X}, \tilde{P}) = \mu_p \end{cases}$$
(7)

The Newton-Raphson method can be used to solve this system of equations iteratively for the unknown variables  $\tilde{b}$ ,  $\tilde{X}$  and  $\tilde{P}$ , but for every evaluation of these equations the ODEs (6) have to be integrated numerically. This integration in theory has to proceed until infinite age but in practice integration is stopped when the probability to survive  $\mathscr{F}(a, \tilde{X}, \tilde{P})$  has dropped below some very low value (e.g. 1.0.

130 10<sup>-9</sup>).

The methodology discussed above is sufficiently general that it can be applied to a wide range of PSPMs, 131 including those with finitely many individual and environmental state variables and with individuals that 132 are born with finitely many different states at birth (Diekmann et al., 2003). The R package 'PSP-133 Manalysis' uses this methodology to compute steady states of PSPMs but also implements pseudo-134 arclength continuation techniques to compute steady state curves as a function of 1 or 2 model param-135 eters (Kuznetsov, 1998, Chapter 10). The results section illustrates how to use this curve continuation 136 approach for model analysis. While computing such curves the 'PSPManalysis' package furthermore 137 detects certain bifurcation points, which are points along a curve where the nature of the computed equi-138 librium undergoes a qualitative change. As illustrated in the results section, such a qualitative change 139 could refer to whether or not a particular equilibrium state can or can not be invaded by a population. 140 For the detection of these bifurcation points the 'PSPManalysis' package again uses the techniques and 141 tests presented in Kuznetsov (1998, Chapter 10). 142

Diekmann et al. (2003) discuss that the approach to compute steady states of PSPMs can also be used to analyse evolutionary dynamics using the theory of Adaptive Dynamics (Dieckmann & Law, 1996; Metz et al., 1996; Geritz et al., 1998). Adaptive dynamics theory explicitly relates evolution by natural selection to population dynamics by considering whether rare mutant phenotypes can invade and take over a resident population. The invasion fitness of such rare mutants is determined by their population growth rate under the environmental conditions imposed by the population with resident phenotype. Because the quantity  $R_0(\infty, \tilde{X}, \tilde{P}) - 1$  has the same sign as this mutant invasion fitness, it can be used as fitness proxy (Diekmann et al., 2003; Durinx, Metz, & Meszéna, 2007). Therefore, the sign of the selection gradient is determined by the derivative of  $R_0(\infty, \tilde{X}, \tilde{P})$  with respect to a model parameter representing a life history trait. I will focus below on the length to switch to the growth habitat  $\ell_s$ . Endpoints of evolution in  $\ell_s$ , also referred to as *evolutionarily singular strategies* or ESSs then satisfy the condition

$$\frac{dR_0(\infty,\tilde{X},\tilde{P})}{d\ell_s} = 0$$

which implies that the invasion fitness reaches a maximum or minimum and the selection gradient vanishes for the given set of environmental conditions  $\tilde{X}$  and  $\tilde{P}$ . The characteristics of the ESS can be determined on the basis of second-order derivatives of  $R_0(\infty, \tilde{X}, \tilde{P})$  with respect to the life history parameter as explained in detail by Geritz et al. (1998). Furthermore, the fitness gradient  $dR_0(\infty, \tilde{X}, \tilde{P})/d\ell_s$ also determines the rate at which the life history trait  $\ell_s$  changes over evolutionary time following the 'canonical equation of Adaptive Dynamics' derived by Dieckmann and Law (1996).

Detailed introductions to the theory of adaptive dynamics are found in Dieckmann and Law (1996), Metz 149 et al. (1996), Geritz et al. (1998), Durinx et al. (2007), and Lion (2018). The 'PSPManalysis' package 150 implements the techniques and conditions from adaptive dynamics discussed in these publications to 151 locate ESSs and to identify their properties; for example, by classifying them as a 'continuously stable 152 strategy', which refers to an ESS to which the life history trait evolves and is furthermore evolutionary 153 stable, or as 'branching point', where evolutionary branching or diversification can occur (see Geritz et 154 al., 1998, for details). For this purpose the 'PSPManalysis' package computes the first and second-order 155 derivatives of  $R_0(\infty, \tilde{X}, \tilde{P})$  with respect to a life history parameter through numerical differentiation. 156

#### 157 Model implementation for 'PSPManalysis'

The life history model presented in Table 1 has to be implemented in an R script to be analysed with the 'PSPManalysis' package. The implementation requires the specification of 3 vectors and 4 functions (see Table 2). The vectors define the model dimensions, the environmental state variables and the model parameters, whereas the functions define the right-hand side of the ODEs (6), their starting values and the boundaries of the different stages in the life history as well as the last two of the conditions (7) that determine the steady states of the environmental variables. In the Supporting Information this implementation of the example model in R is discussed in more detail.

Although the life-history model to be analysed can be specified in R, the entire 'PSPManalysis' package is written in the C language. The package is designed in such a way that implementation of the life history model in C is also possible and in fact preferable. The implementation of the example model used in this paper in the C language is included in the Supporting Information to this paper as a C header file (with extension . h). The implementation in C has a similar, self-explanatory layout as the model implementation in R shown in Table 2 and adaptation of the implementation for another PSPM should only require basic knowledge of the C language. Specifying the model definition in R may be more

straightforward, but comes at a substantial computational cost: Using the model implementation in C
rather than the R implementation decreases the computational time for the results presented in the next
section with a factor 40-50.

### 175 **Results**

#### 176 Bifurcation analysis of ecological steady states

The main purpose of the 'PSPM analysis' package is to compute the steady states of a PSPM as a func-177 tion of one of the model parameters, which I from here on refer to as the bifurcation parameter. This not 178 only requires the 'PSPManalysis' package implementing the methodology, but also a strategy to execute 179 the computations in a comprehensive manner. Figure 1 shows the equilibrium bifurcation curves of the 180 example model as a function of  $X_{max}$ . The figure itself was produced using basic plotting commands 181 in R on the basis of data computed with the main function called PSPMequi provided by the 'PSPM-182 analysis' package. A total of 3 computational steps were needed to generate the data for Figure 1 with 183 the 'PSPManalysis' package. Here I discuss these computational steps in broad terms, focussing on the 184 computational strategy rather than on the relevant R commands, which are discussed in more detail in 185 the supporting information. 186

Key to the computation of steady state curves as a function of model parameters is a good starting point. 187 For the example model we can start the computations at the trivial steady state  $(\tilde{b}, \tilde{X}, \tilde{P}) = (0, X_{max}, 0)$ , 188 which will be the only steady state for maximum resource densities close to 0, because consumers then 189 do not encounter sufficient resources to reach the length at maturation (note the maximum length equal 190 to  $\ell_{inf}X/(K+X)$ ). As a first step to generating the data for Figure 1 this resource-only equilibrium with 191 zero density of both the length-structured consumer and predator was computed for increasing values of 192 the bifurcation parameter  $X_{max}$ , starting from  $X_{max} = 0.1$ . The result is the curve section with increasing 193 equilibrium densities  $\tilde{X}$  at low values of  $X_{max}$  in Figure 1. This computational step would be superfluous 194 as the value of this steady state is known exactly as  $(\tilde{b}, \tilde{X}, \tilde{P}) = (0, X_{max}, 0)$ , but for the fact that the 195 function PSPMequi in the 'PSPManalysis' package with which this computational step is carried out, 196 locates a bifurcation point along the curve. It labels this bifurcation point with the string "BP #0", 197

indicating that this bifurcation point represents a branching point (also called transcritical bifurcation 198 point (Kuznetsov, 1998)) for the structured population with index 0 (the consumer population; because 199 the 'PSPManalysis' package is written in C, indices conform to the C convention, in which the first 200 element of a vector has index 0 rather than index 1 such as in R). From the output of the function 201 PSPMequi it can be inferred that for  $X_{max}$  values to the left of this branching point consumers can not 202 establish themselves in the computed, resource-only equilibrium, because their lifetime reproductive 203 output  $R_0$  is below 1, whereas they can for values to the right of it. These steady states to the left and 204 right of the bifurcation point are therefore stable and unstable (saddle points), respectively. 205

Step 2 of the computations uses the bifurcation point located in the first computational step as starting 206 point to compute the curve of consumer-resource steady states as a function of  $X_{max}$ . The resulting curve 207 corresponds to the part of the equilibrium curves shown in Figure 1 with constant resource density, 208 linearly increasing densities of consumer biomass in the nursery and growth habitat and zero density for 209 the unstructured predator. In this curve the PSPMequi function detects a branching point (or transcritical 210 bifurcation point) for the environment variable with index 1 (the unstructured predator population), 211 which it labels as "BPE #1" (see Figure 1). The consumer-resource steady states to the right of this 212 branching point can be invaded by the unstructured predator population, as indicated by the positive 213 per-capita growth rate that the function PSPMequi produces as output. Predators can hence increase in 214 abundance from low densities for higher  $X_{max}$ -values. In contrast, predators have a negative per-capita 215 growth rate and can thus not invade the consumer-resource steady state for  $X_{max}$ -values to the left of 216 the branching point labeled "BPE #1" in Figure 1, which may therefore represent a stable equilibrium. 217 Whether or not these steady states are indeed stable or, alternatively, cycles in consumer and resource 218 abundance occur can only be investigated using numerical studies of the dynamics, because appropriate 219 test functions that detect transitions from stable equilibrium states to limit cycles (occurring at so-called 220 Hopf bifurcation points (Kuznetsov, 1998)) are currently not available for PSPMs. 221

The last step of the analysis uses the detected branching point for the unstructured predator population to start a computation of the steady states with positive predator density as a function of  $X_{max}$ . The result of this computation is a folded curve of steady state values which extends to a minimum just below  $X_{max} = 4$ . This point at which the equilibrium curve doubles back on itself is another bifurcation

point called limit point (or saddle-node bifurcation point Kuznetsov, 1998) and labeled by the func-226 tion PSPMequi with "LP" (see Figure 1). Ecologically, this minimum value of  $X_{max}$  represents the 227 persistence boundary of the unstructured predator population whereas the branching point detected in 228 step 2 and labeled "BPE #1" represents the predator's invasion boundary. In between the persistence 229 and invasion boundary two steady states are possible: a consumer-resource steady state that can not be 230 invaded by the predator and a predator-consumer-resource equilibrium. On the basis of the general bi-231 furcation theory presented in Kuznetsov (1998) it can be inferred that the part of the predator-consumer-232 resource equilibrium curve between the two bifurcation points represents unstable equilibrium states 233 (saddle points). 234

Because they represent the invasion and persistence boundary of the predator, it may be ecologically 235 relevant to assess how the location of the bifurcation points labeled "BPE #1" and "LP" depends on 236 other model parameters. The PSPMequi function therefore also allows for the computation of these 237 bifurcation points as a function of a second model parameter. Figure 2 shows the location of these two 238 bifurcation points as a function of the maximum resource density  $X_{max}$  and the predator mortality rate 239  $\mu_p$ . For values in the region of parameter space between the two lines in Figure 2 two steady states 240 occur that are potentially stable, a tritrophic steady steady state with predators and a consumer-resource 241 steady state that predators can not invade. This region of bistability is larger at higher rates of predator 242 mortality. 243

#### 244 Evolutionary analysis

A particular strength of the 'PSPManalysis' package is that it can also be used to calculate evolutionary 245 singular strategies (ESSs) using the framework of Adaptive Dynamics (Dieckmann & Law, 1996; Metz 246 et al., 1996; Geritz et al., 1998). The package therefore allows studying questions about the evolution of 247 life history traits in a context with population and community feedback on the environment in which the 248 evolutionary process takes place. Figure 3 provides as an example the equilibrium bifurcation curves of 249 the example model as a function of the individual length at the habitat switch  $\ell_s$ . Because it influences 250 the extent of resource competition and predator vulnerability an individual experiences throughout life, 251 this trait can be expected to be under strong selection. The data shown in Figure 3 have been computed 252

as before with the function PSPMequi, which can also produce as output the value of the selection 253 gradient, that is, the derivative  $dR_0(\infty, \tilde{X}, \tilde{P})/d\ell_s$  of the lifetime reproductive output  $R_0$  with respect to 254 the life-history parameter  $\ell_s$  (see bottom panel of Figure 3). In the absence of the predator selection 255 for a smaller size at habitat shift occurs, but in the presence of the predator the function PSPMequi 256 detects an evolutionary singular state (ESS), which it labels as "CSS #0" on the basis of the second-257 order derivatives of  $R_0$  with respect to  $\ell_s$ . The label indicates that the ESS is convergent stable, such 258 that the value of  $\ell_s$  will evolve toward the  $\ell_s$  value of this CSS, while after fixation mutants with slightly 259 different values of  $\ell_s$  will not be able to invade (see Geritz et al., 1998, for further details about the ESS 260 classification). 261

Once an evolutionary singular strategy has been detected the function PSPMequi can be used to construct a pairwise invasibility plot (PIP; Van Tienderen & de Jong, 1986; Geritz et al., 1998), that is, a graph of the sign of a mutant's invasion fitness as a function of the life-history trait value of both the resident and the mutant (Figure 4, left panel). Starting from the CSS detected in Figure 3 the function PSPMequi was used to compute a boundary with zero mutant fitness as a function of the resident and mutant life-history trait value. This boundary corresponds to the curve separating regions with positive and negative mutant fitness in the PIP shown in Figure 4 (left panel).

Finally, the 'PSPManalysis' package also includes a function PSPMevodyn, which can be used to simulate the dynamics of evolving life-history parameters over evolutionary time (Figure 4, right panel). These evolutionary dynamics are described by the canonical equation of Adaptive Dynamics (Dieckmann & Law, 1996). The trajectory of  $\ell_s$  over evolutionary time shown in Figure 4 (right panel) confirms that the individual length at the habitat switch evolves to the value of the convergent stable ESS shown in Figure 3, but since the function PSPMevodyn can not simulate combined mutant and resident dynamics, it is not possible to verify whether or not evolutionary branching is possible at this ESS.

The results shown in Figure 3 and 4 were on purpose computed with a low value of  $X_{max} = 0.5$ , as the default value of  $X_{max} = 5.0$  results in a more complicated, and hence also more intriguing evolutionary outcome of the selection process in  $\ell_s$  (Figure 5). For this value of  $X_{max}$  the range of  $\ell_s$  over which both a consumer-resource equilibrium and a predator-consumer-resource equilibrium occur is more extensive and the evolutionary singular state (classified by the PSPMequi function as an evolutionary repellor and

labeled "ERP #0") now occurs on the part of the curve representing saddle-node steady states of the 281 predator, consumer and resource. This state is hence ecologically unstable and thus unreachable. As 282 before smaller values of  $\ell_s$  are selected for in the absence of the predator, as this reduces competition 283 for resources among consumers. In the presence of the predator, however, larger values of  $\ell_s$  confer 284 a selective advantage. As a consequence, with predators present evolution will result in larger values 285 of  $\ell_s$  until the limit point is reached around  $\ell_s = 22.5$ , at which point the predator goes extinct and the 286 community collapses to a consumer-resource equilibrium. After predator extinction, the direction of 287 evolution reverses and smaller values of  $\ell_s$  are selected for until the value of this life-history parameter 288 reaches the bifurcation point labeled "BPE #1" in Figure 5, below which the predator can once again 289 invade the consumer-resource equilibrium. From Figure 5 it can hence be inferred that over evolutionary 290 time cycling will occur between a consumer-resource steady state and a steady state with predator, driven 291 by the selection process in  $\ell_s$ . Simulating the evolutionary dynamics with the PSPMevodyn function 292 following the canonical equation of Adaptive Dynamics would, however, not reveal such evolutionary 293 cycling as this function can not switch between the two ecological steady states. 294

## 295 Discussion

This paper outlines a general methodology for the analysis of PSPMs, presents a strategy how to apply 296 this methodology to ecological and evolutionary questions and introduces a R package that implements 297 the numerical tools required by this methodology. This methodology has been used to gain insight about 298 how individual development affects the ecological dynamics of size-structured populations and commu-299 nities (de Roos & Persson, 2013). More recently, it was applied to a variety of evolutionary problems, 300 ranging from the evolution of ontogenetic niche shifts (ten Brink & de Roos, 2017), metamorphosis 301 (ten Brink et al., 2019), cannibalism (Hin & de Roos, 2019a), ontogenetic size-scaling (Hin & de Roos, 302 2019b) and the timing of habitat shifts (Chaparro Pedraza & de Roos, 2020). Given the importance 303 of the individual life history for many eco-evolutionary questions and the importance of environmental 304 feedback on this life history (Lion, 2018), however, there is scope to apply the methodology to a wide 305 range of eco-evolutionary problems. 306

307 Physiological structured population models are key to the presented methodology. In contrast to for

example integral projection models (Ellner et al., 2016) PSPMs are not directly based on observational 308 life history data. Such data, however, are often collected under at most a few different sets of environ-309 mental or density-dependent conditions and hence offer limited information about how the individual 310 life history changes under environmental feedback. In contrast, PSPMs are based on a functional model 311 of the individual life history that mechanistically accounts for the interplay between the life history and 312 the environment (de Roos, 2020). PSPMs therefore often address the question how the individual life 313 history is shaped by both individual traits and environmental, density-dependent impacts. The current 314 paper shows that to address such questions it suffices to specify only the model of the individual life 315 history model dependent on individual traits and environmental variables. Formulating a population-316 level model is not required as the translation to the population level is sufficiently generic that it can be 317 abstracted into a software approach. 318

The capabilities of the 'PSPM analysis' package are more extensive than highlighted in this paper. Next 319 to the routines for the bifurcation analysis of ecological steady states, the package includes routines 320 for the demographic analysis of PSPMs (following de Roos, 2008), the simulation of the ecological 321 dynamics of structured populations (de Roos et al., 1992), the computation of evolutionary singular 322 states as a function of parameters and the computation of the individual life history under different 323 environmental conditions. All procedures use the same specification of the individual life history, such 324 as the one shown in Table 2, which moreover has a generic structure that is readily adapted to many 325 different life history models. Other methods for the bifurcation analysis of PSPMs do exist (Breda, 326 Diekmann, Gyllenberg, Scarabel, & Vermiglio, 2015; Gyllenberg, Scarabel, & Vermiglio, 2018) and 327 can handle more complex bifurcations, such as Hopf bifurcation points, that can not be detected and 328 analysed with the 'PSPManalysis' package, but these methods always have to be specifically tailored 329 to the particular PSPM. The generic and easy-to-adapt implementation of the life history model is the 330 key aspect of the 'PSPManalysis' package that makes it useful for a wide range of eco-evolutionary 331 problems. 332

Theoretical studies in ecology and evolutionary biology often aim for deriving analytical insight, as such insight is considered to apply more generally to a wide range of systems. The complexity of models that can be investigated analytically is, however, severely limited. For example, conditions determining

the equilibrium of a size-structured consumer-resource system and its stability can be derived analyt-336 ically only if mortality is size-independent and an explicit expression for the size-at-age relationship 337 is available (de Roos et al., 1990). The derived conditions are themselves, however, not transparent 338 and nevertheless have to be solved numerically. To facilitate mathematical analysis structured mod-339 els are then often simplified, for example by replacing size-dependent life history functions with far 340 easier age-dependent ones. Size-dependent and age-dependent life histories (and hence size-structured 341 and age-structured population models) are however fundamentally different from each other, if growth 342 in body size and hence the progression through ontogeny depends on environmental factors like food 343 availability (de Roos & Persson, 2013). Often the choice is hence between a simple model for which 344 analytical results are possible or a model with a more faithful representation of the complexity of real 345 life histories that can only be investigated numerically. In this context the power of numerical bifur-346 cation analysis (Kuznetsov, 1998) has received too little attention. It offers a more powerful approach 347 than brute-force numerical simulations of dynamics, as it provides more comprehensive insight about 348 the different types of stable dynamic patterns that can occur for given combinations of model parame-349 ters. The very essence of bifurcation theory furthermore guarantees that these dynamic patterns occur 350 over at least a range of parameter values, lending the results some measure of generality. Methodology 351 and software for the numerical bifurcation analysis of models in terms of ODEs have been available for 352 a while (Dhooge, Govaerts, & Kuznetsov, 2003), the 'PSPManalysis' package is intended to provide 353 some of the same capabilities for the general class of PSPMs and thereby facilitate investigating ques-354 tions about the relationship between complex individual life histories and the dynamics of populations 355 and communities on both ecological and evolutionary time scales. 356

## **357** Acknowledgements

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# **Jata accessiblity**

- <sup>361</sup> No experimental or empirical data are used or collected in this study. The 'PSPManalysis' package can
- <sup>362</sup> be installed from CRAN using the command:

```
363 install.packages("PSPManalysis")
```

<sup>364</sup> The most recent version can always be installed using the command:

```
365 devtools::install_bitbucket("amderoos/PSPManalysis", subdir = "R/",
366 build_vignettes = TRUE)
```

- <sup>367</sup> The Supplementary Information contains:
- 1. Table S1 with model parameters and their default values.
- 2. A discussion of the implementation of the life history model of Chaparro Pedraza and de Roos
- 370 (2020) in the R script called "Salmon.R" (also shown in Table 2).
- 371 3. A discussion of the R commands used to generate the data for the figures in this paper.
- 4. The R script called "EcoFigures.R" with the R code to generate Figures 1 and 2.
- 5. The R script called "EvoFigures.R" with the R code to generate Figures 3, 4 and 5.
- 6. The C header file called "Salmon.h" with the C implementation of the life history model of
- <sup>375</sup> Chaparro Pedraza and de Roos (2020).

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# 449 Tables

Table 1: Model functions of the individual life history model of Chaparro Pedraza and de Roos (2020).

Function	Description

Life history functions

$\gamma(\ell,X) = egin{cases} \xi\left(\ell_{inf}rac{X}{K+X}-\ell ight) \ \xi\left(\ell_{inf}-\ell ight) \end{cases}$	$\ell < \ell_s$ otherwise	Growth rate in length
$eta(\ell) = egin{cases} B_{max}\ell^2 \ 0 \end{bmatrix}$	$\ell > \ell_m$ otherwise	Fecundity
$\mu(\ell,P) = egin{cases} \mu_1 \ \mu_2 + \phi \ell^{-d} P \end{cases}$	$\ell < \ell_s$ otherwise	Mortality
Impacts on the environment $ \int \frac{X}{1-\frac{X}{2}}\ell^2 $	l < l	

$\alpha(\ell, X) = \begin{cases} I_{max} \overline{K + X}^{\ell^2} \\ 0 \end{cases}$	$\ell < \ell_s$ otherwise	Foraging on resource in nursery habitat
$oldsymbol{arepsilon}(\ell) = egin{cases} \phi \ell^{3-d} \ 0 \end{bmatrix}$	$\ell > \ell_s$ otherwise	Contribution to predator numerical re- sponse

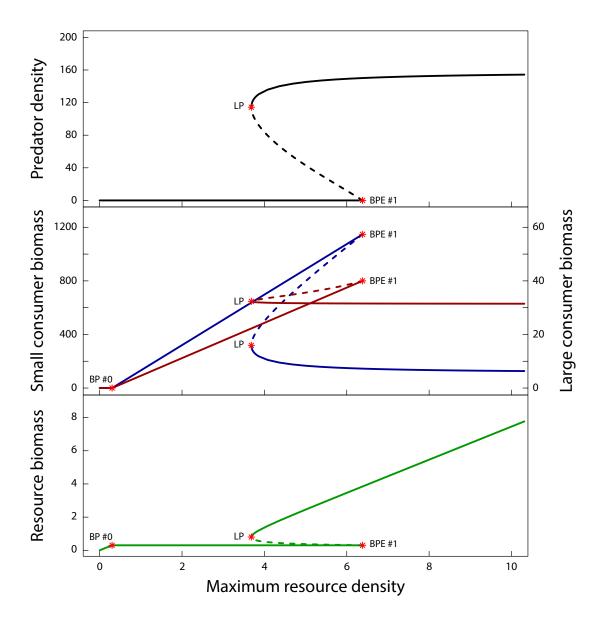
function of the numerical response B

Functions related to the environment	
$g(X) = \rho \left( X_{max} - X \right)$	Resource turn-over rate
$h(B) = B - \mu_p$	Predator per-capita growth rate as a function of the numerical response $R$

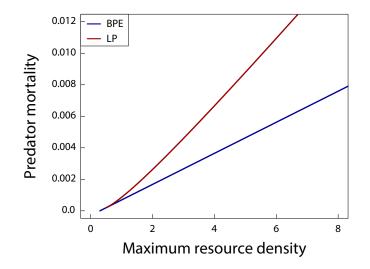
**Table 2**: The implementation in R of the individual life history model of Chaparro Pedraza and de Roos (2020) for the analysis with the 'PSPManalysis' package.

```
PSPMdimensions <- c(PopulationNr = 1, IStateDimension = 2,</pre>
                     LifeHistoryStages = 3, ImpactDimension = 5)
EnvironmentState <- c(X = "GENERALODE", P = "PERCAPITARATE")</pre>
DefaultParameters <- c(Rho = 0.01, Xmax = 5.0,
                       K = 1.0, Imax = 0.0025, Bmax = 0.002725,
                       L0 = 2.0, Ls = 20.0, Lm = 30.0, Linf = 115.0,
                       Xi = 0.00051, Mu1 = 0.002, Mu2 = 0.006,
                       D = 0.75, Phi = 0.001, Mup = 0.006)
StateAtBirth <- function(E, pars) {</pre>
  with(as.list(c(E, pars)),{
    c(Age = 0.0, Length = L0)
  })
}
LifeStageEndings <- function(lifestage, istate, birthstate, BirthStateNr, E, pars) {</pre>
 with(as.list(c(E, pars, istate)),{
   maturation = switch(lifestage, Length - Ls, Length - Lm, -1)
  })
}
LifeHistoryRates <- function(lifestage, istate, birthstate, BirthStateNr, E, pars) {
  with(as.list(c(E, pars, istate)),{
    list(
      development = c(1.0),
                       switch(lifestage, Xi*(Linf*X/(K+X) - Length),
                              Xi*(Linf - Length), Xi*(Linf - Length))),
      fecundity = switch(lifestage, 0, 0, Bmax*Length^2),
      mortality = switch(lifestage, Mu1, Mu2 + Phi*P*Length^(-D),
                         Mu2 + Phi*P*Length^(-D)),
      impact = switch(lifestage,
                      c(Imax*X/(K+X)*Length^2, 0, Length^3, 0, 0),
                       c(Imax*X/(K+X)*Length^2, Phi*Length^(3-D), 0, Length^3, 0),
                       c(Imax*X/(K+X)*Length^2, Phi*Length^(3-D), 0, 0, Length^3))
    )
  })
}
EnvEqui <- function(I, E, pars) {</pre>
 with(as.list(c(E, pars)),{
    c(Rho*(Xmax - X) - I[1], I[2] - Mup)
  })
}
```

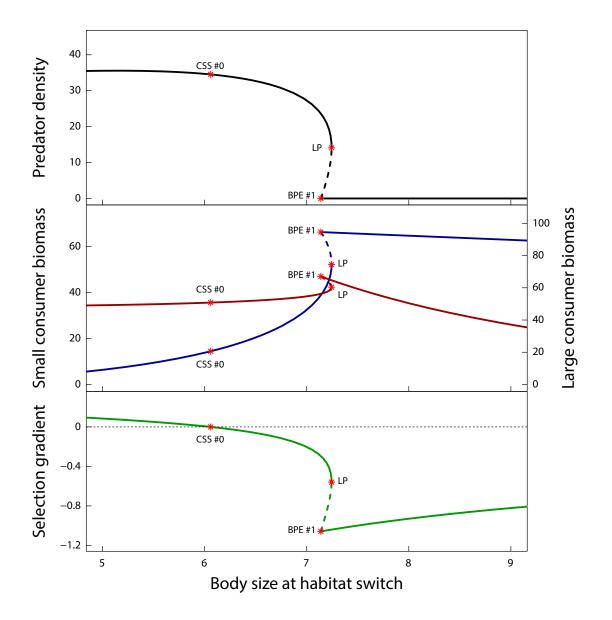
## 450 Figures



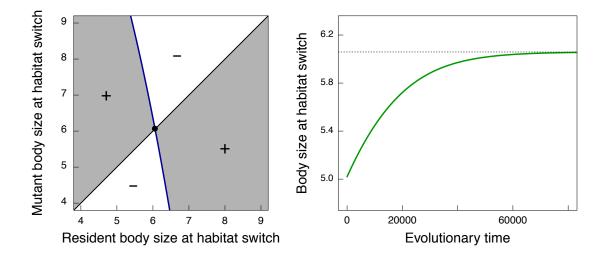
**Figure 1**: Steady state densities of the unstructured predator population (*top panel*), the length-structured consumer population in the nursery and growth habitat (*middle panel*) and the basic resource (*lower panel*) in the example model (see Table 1) as a function of the maximum resource density  $X_{max}$ . All other parameters have default values (Table S1 in the supporting information). See the main text for details about the bifurcation points labeled "BP #0" (branching point for structured population with index 0), "BPE #1" (branching point for the environmental variable with index 1) and "LP" (limit point). Solid lines represent possibly stable equilibria, dashed lines represent saddle points. The curve sections with unstable resource-only and consumer-resource steady states that can be invaded by the structured consumer and unstructured predator, respectively, have been omitted for clarity.



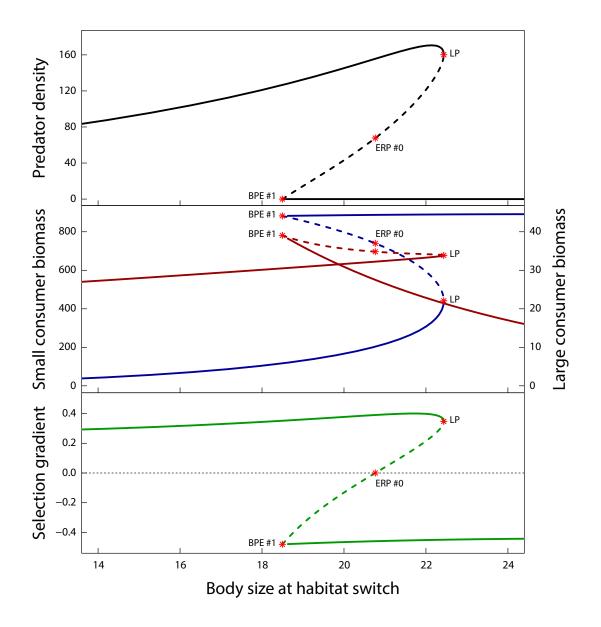
**Figure 2**: Location of the bifurcation points shown in Figure 1 as a function of the maximum resource density  $X_{max}$  and the predator mortality rate  $\mu_p$ . Only the location of the bifurcation points labeled "BPE #1" (branching point for the environmental variable with index 1) and "LP" (limit point) are shown, as the location of the bifurcation point labeled "BP #0" (branching point for structured population with index 0) is independent of the predator mortality rate  $\mu_p$ . All other parameters have their default values (Table S1 in the supporting information).



**Figure 3**: Steady state densities of the unstructured predator population (*top panel*) and the lengthstructured consumer population in the nursery and growth habitat (*middle panel*), as well as the selection gradient on the individual length at habitat switch  $\ell_s$  in the example model (see Table 1) as a function of this length at habitat switch  $\ell_s$ . Maximum resource density  $X_{max} = 0.5$ , all other parameters have their default values (Table S1 in the supporting information). See the main text for details about the bifurcation points labeled "BPE #1" (branching point for the environmental variable with index 1) and "LP" (limit point) and the evolutionary steady state labeled "CSS #0" (convergent stable evolutionary state for structured population with index 0). Solid lines represent possibly stable equilibria, dashed lines represent saddle points. The curve sections with unstable resource-only and consumer-resource steady states that can be invaded by the structured consumer and unstructured predator, respectively, have been omitted for clarity.



**Figure 4**: *Left:* Pairwise invasibility plot showing combinations of the resident and mutant value of the length at habitat switch  $\ell_s$  with positive and negative invasion fitness of the mutant. *Right:* Simulation of the dynamics of the evolving value of the length at habitat switch  $\ell_s$  over evolutionary time as predicted by the canonical equation of Adaptive Dynamics (Dieckmann & Law, 1996).



**Figure 5**: As Figure 3 but for  $X_{max} = 5.0$ . The point labeled "ERP #0" refers to an evolutionary repellor for the structured population with index 0 (see main text for details). Notice the opposing signs of the selection gradient in the consumer-resource and predator-consumer-resource equilibrium, which predicts that evolutionary cycling will occur for the body length at habitat switch  $\ell_s$  between the bifurcation points labeled "BPE #1" and "LP" resulting in repeated invasion and extinction of the predator population.