A new on-line tool for the calculation of bioaccumulation factors of active substances within living organisms: MOSAIC$_{\text{Bioacc}}$.

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Abstract

Today, no effective tools are proposed in ecotoxicology to diagnose and predict the accumulation and effects of chemical substances on living organisms, accounting for exposure situations that are known to be complex (routes of exposure, metabolization processes, cocktail effects, etc.). Among elaborated tools, toxicokinetic/toxicodynamic (TKTD) models are now strongly recommended to describe the exposure time-dependency on individual life-history traits from experimental data collected through standard toxicity tests. In particular, the TK part is used to relate the exposure medium concentration to the internal concentration within organisms, considering various processes such as accumulation, depuration, metabolization and excretion (ADME). Regarding plant protection products in marketing authorization applications, regulation No 283/2013 (EU) defines the data requirements for active substances with a bioaccumulation test on fish following OECD guideline 305, which consists in an accumulation phase followed by a depuration phase. The concentration of the substance within fishes is followed over time during both phases thus allowing to describe the exposure time-dependency on individual (fit of TK model) and bioaccumulation factors. On a regulation point of view, these factors are decisive criteria for further estimating the concentration of active

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substances present in food items of vertebrates and more particularly of piscivorous birds and mammals. This paper presents a new ready-to-use tool providing on-line the estimation of the bioaccumulation factors associated with their uncertainty to facilitate the daily work of regulators, but also of any user, by benefiting of a freely available and user-friendly interface avoiding to invest into underlying methods. This tool, MOSAIC$_{\text{Bioacc}}$, is available from the following adress: https://mosaic.univ-lyon1.fr/bioacc, and can be used as a turn-key decision making tool in the framework of environmental risk assessment.

***Keywords**: Active substances, Bayesian inference, Bioaccumulation factors, Environmental Risk Assessment, R software, Toxicokinetic models, Uncertainty.*

1. Introduction

Nowadays, ecotoxicology has to provide relevant and effective methods and tools in a predictive (prior risk assessment) and diagnostic (posterior identification of effects and their remediation) perspective. The aims of these methods and tools are to quantify the exposure routes of potentially hazardous substances to assess dose-response relationships and define critical effect concentrations, helping in understanding the underlying biological and ecological mechanisms, and developing indicators for environmental quality. In this context, toxicokinetic/toxicodynamic (TKTD) models are now strongly recommended to describe the exposure time-dependency on individual life-history traits from experimental data collected through standard toxicity tests (Ockleford et al., 2018). In particular, the TK part is used to relate the exposure medium concentration to the internal concentration within organisms, considering various processes such as accumulation, depuration, metabolization and excretion (ADME). Regarding plant protection products in marketing authorization applications, regulation No 283/2013 (EU) (European Commission, 2013) defines the data requirements for active substances with a bioaccumulation test on fish following OECD guideline 305 (OECD, 2012), which consists in an accumulation phase followed by a depuration phase. During the accumulation phase, individuals are exposed to the substance under interest at a range of concentrations, chosen according to the mode of action of the substance. Then, organisms are transferred to a clear medium for the depuration phase. The concentration of the substance within fishes is followed over time during both phases thus allowing to define the bioconcentration factor (BCF), the biomagnification factor (BMF) or the bioaccumulation factor (BAF) according to the exposure route (via water, food or sediment respectively). On a regulation point of view, such factors are decisive criteria for further estimating the concentration of active substances present in

food items of vertebrates and more particularly of piscivorous birds and mammals. A test with fish according to OECD TG 305 is also preferred where experimental information on bioaccumulation is needed for PBT/vPvB assessment under REACH regulation (European Commission, 2006; ECHA, 2017). In most cases, preference is also given to experimentally derived BCF values to be used for secondary poisoning assessment under Biocidal Products Regulation (European Commission, 2012).

In the past decades, many types of methods have been proposed to estimate parameters of simple TK models, making it possible to deduce bioaccumulation factors (especially BCF) values depending on the kinetics of accumulation and depuration observed in dedicated experiments (OECD, 2012). Nowadays, TK model parameters are rarely estimated simultaneously, considering both their uncertainties with correlation. However, it is crucial to properly estimate them for a better risk assessment and still remains a challenge today. Besides, regulators are sceptical about the diversity and complexity of these methods and tools, while thinking about their standardization at the same time (Tan et al., 2020). Ratier et al. (2019) recently proposed a full revisit of the approach based on a unified inference method to estimate parameters for both accumulation and depuration phases simultaneously. This innovent inference process makes it possible to incorporate the TK part into complete TKTD model.

This paper presents a new ready-to-use tool, $MOSAIC_{Bioacc}$, leading to the estimation of bioaccumulation factors associated with their uncertainty to facilitate the daily work of regulators, but also of any user, benefiting of a freely available and user-friendly interface avoiding to invest into underlying methods. This web app clearly takes advantage of the new conceptual framework developed

by Ratier et al. (2019) in order to estimate parameters of a one-compartment TK model including several exposure routes (water, food, sediment), accounting for metabolization (three metabolites) of the chemical substance and possibly for growth dilution. The web app will only requires to upload standard experimental data, then some clicks provide all results. In particular, BCF/BMF/BAF are delivered as probability distributions summarized by a median and a 95% credible intervals. In addition, a collection of goodness-of-fit criteria is provided to ensure the robustness of the results. At last, any output can be downloaded for further inclusion into any home-made document, while a full report can be directly downloaded in different convenient formats.

2. Description

MOSAIC$_{\text{Bioacc}}$ was developed as a Shiny environment, available at https://mosaic.univ-lyon1.fr/bioacc and hosted at the Rhône-Alpes Bioinformatics Center PRABI (2016). A user guide and an explanatory video are available at the beginning of the application to assist the user.

2.1. Data uploading

When using MOSAIC$_{\text{Bioacc}}$, the first step is to upload input data (**Fig. 1-a**). MOSAIC$_{\text{Bioacc}}$ expects to receive bioassay data as a .txt file or a .csv file (comma, semicolon or tabular separator) with a specific format. Each line of the table corresponds to a time point for a given replicate and a given exposure concentration of the contaminant. The table must contain four columns, with the exact following header names, the order of column not mattering (**Table 1**): 'time' (the time point of the measurement at the exposure concentration, in hours, minutes, days or weeks), 'exposure' (the concentration of the contaminant in the exposure medium, in $\mu g.mL^{-1}$), 'replicate' (a number or a string that is

unique for each replicate, dimensionless), and 'concentration' (the concentration of the contaminant within the organism, in $\mu g.g^{-1}$).

Table 1: Example of a data set ready to be used with a TK model.

time	concentration	exposure	replicate
0	0.000	0.0044	2
3	0.225	0.0044	2
7	0.355	0.0044	2
14	0.553	0.0044	2
21	0.658	0.0044	2
28	0.785	0.0044	2

As shown on **Fig. 1-a**, the user of MOSAIC$_{\text{Bioacc}}$ can either upload its own data set with a click on 'Browse' (by taking care about the format specification of its file) or try the platform with an example data set. When the upload is complete, the user has to manually select the corresponding separator, the appropriate time unit and the duration of the accumulation phase. The application first provides an overview of the raw data which allows the user to check if the data were correctly entered. The user can also visualize a plot of the data (**Fig. 1-b**). In case the user uploads a file with several exposure concentration, it has to select the one for which it wants to see the results. Indeed, only one file at a time can be analysed by the web app. It should be noted that for each new choice of exposure concentration, the duration of the accumulation phase is reset and framed in orange to invite the user to update it before to proceed to new calculations.

Some example files are provided in MOSAIC$_{\text{Bioacc}}$ to illustrate it and help

Figure 1: Data uploading and user information to enter from the MOSAIC$_{Bioacc}$ homepage.

the users. MOSAIC$_\text{Bioacc}$ is illustrated in this paper with a typical data set of a toxicokinetic bioassay where concentrations were collected for rainbow trout (*Oncorhynchus mykiss*) exposed to contaminated water at a highly hydrophobic chemical (logK$_\text{ow}$ = 9.06) at an exposure concentration of 0.0044 $\mu g.mL^{-1}$ over 49 days, with one replicate at each time point. Then fishes are transferred in a clean medium for 98 days (Crookes & Brooke, 2011). The data set can be downloaded directly from the application.

2.2. Model and parameters

The organisms are considered as single compartments for which a first-order kinetic bioaccumulation model accounting for several exposure and elimination routes can be expressed in a generic way as follows (Eqs. (1) to (4)):

$$
\begin{cases}
\frac{dC_p(t)}{dt} = U - (E + M)C_p(t) \quad (1) \\
\frac{dC_{m_\ell}(t)}{dt} = k_{m_\ell}C_p(t) - k_{e_m\ell}C_{m_\ell}(t) \quad (2)
\end{cases}
\quad for \quad 0 \leq t \leq t_c
$$

$$
\begin{cases}
\frac{dC_p(t)}{dt} = -(E + M)C_p(t) \quad (3) \\
\frac{dC_{m_\ell}(t)}{dt} = k_{m_\ell}C_p(t) - k_{e_m\ell}C_{m_\ell}(t) \quad (4)
\end{cases}
\quad for \quad t > t_c
$$

where $C_p(t)$ is internal concentration of the parent compound at time t ($\mu g.g^{-1}$), $C_{m,\ell}(t)$ internal concentration of metabolite ℓ ($\forall \ell = 1 \ldots L$ with L total number of metabolites) at time t ($\mu g.g^{-1}$), U sum of all uptake terms, E sum of all elimination terms for the parent compound, M sum of all elimination terms for metabolite, $k_{m,\ell}$ metabolization rate of metabolite ℓ ($time^{-1}$), $k_{e,m\ell}$ elimination rate of metabolite ℓ ($time^{-1}$). A summary of the meaning of symbols and letters is provided in **Table 2**.

The previous dynamical system can explicitly be solved when exposure concentration is assumed to be constant over time (Annex 1).

Table 2: Summary of the meaning of symbols and letters of the TK model.

Symbol	Meaning
I	total number of exposure sources
J	total number of elimination processes
L	total number of metabolites
i	index of exposure sources, $i = 1 \ldots I$
j	index of elimination processes, $j = 1 \ldots J$
ℓ	index of metabolites, $\ell = 1 \ldots L$
t	time (expressed in time unit)
c_i	exposure concentration of route i (in $\mu g.mL^{-1}$)
$C_p(t)$	internal concentration of the parent compound at time t (in $\mu g.g^{-1}$)
$C_{m_\ell}(t)$	internal concentration of metabolite ℓ (in $\mu g.g^{-1}$)
k_{u_i}	uptake rate of exposure source i (expressed in $time^{-1}$)
k_{e_j}	elimination rate of elimination process j (expressed in $time^{-1}$)
$k_{e_{m\ell}}$	elimination rate of metabolite ℓ (expressed in $time^{-1}$)
k_{m_ℓ}	metabolization rate of metabolite ℓ (expressed in $time^{-1}$)
t_c	duration of the accumulation phase (expressed per time unit)
$\mathcal{N}$	Normal law
$C_{obs,p}(t)$	internal measured concentration of the parent compound at time t (in $\mu g.g^{-1}$)

Symbol	Meaning
$C_{obs,m_\ell}(t)$	internal measured concentration of metabolite ℓ (in $\mu g.g^{-1}$)
σ	standard deviation for the internal concentration of the parent compound (in $\mu g.g^{-1}$)
σ_{met_ℓ}	standard deviation for the internal concentration of metabolite ℓ (in $\mu g.g^{-1}$)
Letter	
$U = \sum_{i=1}^{I} k_{u_i} c_i$	sum of all uptake terms
$E = \sum_{j=1}^{J} k_{e_j}$	sum of all elimination terms for the parent compound
$M = \sum_{\ell=1}^{L} k_{m_\ell}$	sum of all elimination terms for metabolite ℓ

Within the MOSAIC$_{\text{Bioacc}}$ web app, we choose to only propose three routes

of exposure ($I = 3$, *i.e.*, water, food and sediment), three processes of elimination

($J = 3$, *i.e.*, excretion, growth dilution and metabolization) and a maximum

of three metabolites ($L = 3$) which directly come from the parent compound.

Hence, the maximum number of parameters is 18 when the model is complete,

and the user will have at the end a total of 56 possible models depending on the

combination it chooses (**Annex 2**). According to the experimental design, the

user is invited to select the corresponding parameters (**Fig. 1-c**). We assume

a Gaussian probability distribution for the contaminant and its metabolites

concentration within the organism as follow (Eqs. (5) and (6)):

$$C_{obs,p}(t) \sim \mathcal{N}(C_p(t), \sigma^2) \quad (5)$$

$$C_{obs,m_\ell}(t) \sim \mathcal{N}(C_{m_\ell}(t), \sigma_{met_\ell}^2) \quad (6)$$

167   where variables and parameters are explained in **Table 2**.

168 In this paper, the results will be illustrated with a data set considering only

169 the water exposure route (k_u) with the corresponding elimination rate (k_e).

170 Thus the corresponding first-order kinetic bioaccumulation model can simply be

171 expressed as follows as given by Eqs. (7) and (8):

$$\begin{cases} \frac{dC(t)}{dt} = k_u \times C_w - k_e \times C(t) & \text{for } 0 \le t \le t_c \quad (7) \\ \frac{dC(t)}{dt} = -k_e \times C(t) & \text{for } t > t_c \quad (8) \end{cases}$$

By default, the basic parameters of the model are selected, *i.e.*, water for the

exposure medium and excretion for the depuration process (k_u and k_e). If the

file has correctly been uploaded and the exposure concentration has been chosen

for the analysis, a reminder of the latter is displayed on the screen and the

button for launching the calculations is unlocked. When the user clicks on the

'Calculate and Display' button (**Fig. 1-c**), the calculations are launched with a

notification informing it of the progress. When calculations are finished, results

are displayed, as some plots and tables by default, and a button is available to

display all the goodness-of-fit criteria.

2.3. Calculations

2.3.1. Definition of prior distributions

In MOSAIC$_{\text{Bioacc}}$, the choice of priors is hidden to the user. As the application

has to be the most generic as possible, priors were assumed following a non-

informative (-5, 5) log10-uniform distribution for all uptake and elimination rate

constants, and a non-informative (0.001, 0.001) Gamma prior for all standards

deviations, as usually done in the field of bioaccumulation studies (Lambert et

al., 2005; Ratier et al., 2019; Richards & Chaloupka 2009).

 2.3.2. Bayesian inference

Model computations that underline MOSAIC$_{\text{Bioacc}}$ calculations are performed

with JAGS (Plummer, 2019) and R software (version 3.6.3, R Core Team 2020)

via the `rjags` package (Plummer, 2019). Models are fitted to bioaccumulation

data using Bayesian inference via Markov Chain Monte Carlo (MCMC) sampling.

For each model chosen by the user, running starts with a short sampling on

three chains (5,000 iterations after a burn-in phase of 10,000 iterations) using

the Raftery and Lewis method (Raftery & Lewis, 1992) to set the necessary

thinning and number of iterations to reach a precise and accurate estimation of

each model parameter. Thanks to `rjags`, model parameters are estimated with

the computation of the joint posterior distribution from the likelihood of the

observed data combined with prior distributions on parameters. All details on

this approach can be found in the original research paper (Ratier et al., 2019)

but also in many other papers in the field of ecotoxicology (Billoir et al., 2011;

Ockleford et al., 2018).

2.3.3. Bioaccumulation factors

In MOSAIC$_{\text{Bioacc}}$, bioaccumulation factors are calculated and provided as a

first result. This paper illustrates the case of an organism exposed to contam-

inated water, providing both the kinetic bioconcentration factor (BCF_k) and

the BCF at steady state (BCF_{ss}), which are given by the application. These

factors have exact mathematical expressions given by Equations (9) and (10):

$$BCF_k = \frac{k_u}{k_e} \quad (9)$$

$$BCF_{ss} = \frac{C_p(t_c)}{C_w} \quad (10)$$

211 where $C_p(t_c)$ is the internal parent compound concentration (in $\mu g.g^{-1}$) at

Table 3: Example of BCF (a) and parameter (b) medians (50% quantile) with their 95% credibility intervals (2.5% - 97.5% quantiles). The hyphen stands for dimensionless parameters.

	2.5%	50%	97.5%	Units
(a)				
BCF_k	232	276	336	-
BCF_{ss}	147	232	315	-
(b)				
k_u	7.49	10.62	15.35	d^{-1}
k_e	0.02365	0.03873	0.06105	d^{-1}
σ	0.1249	0.1676	0.2453	$\mu g.g^{-1}$

²¹² the end of the accumulation phase (at $t = t_c$, in time) and C_w is the contaminant

²¹³ concentration in the medium at steady-state ($\mu g.mL^{-1}$).

²¹⁴ *2.4. Results*

²¹⁵ *2.4.1. Bioconcentration factors*

²¹⁶   When the user clicks on 'Calculate and Display' button, some results are

²¹⁷ provided by default. First, the BCF_k is given as a probability distribution

²¹⁸ (**Fig. 2-a**) and summarized with its median and its 95% uncertainty limits (95%

²¹⁹ credible interval, **Table 3-a**). If the user asks for the BCF_{ss}, its probability

²²⁰ distribution is also given (**Fig. 2-b**) as well as its uncertainty limits (**Table**

²²¹ **3-a**).

²²² *2.4.2. Predictions*

²²³   Then, the fitted curve superimposed to the observations are provided (**Fig.**

²²⁴ **2-c**). From the joint posterior distribution, MOSAIC$_{\text{Bioacc}}$ also provides the

²²⁵ marginal posterior distributions for each parameter, which are also summarized

²²⁶ in a table (**Table 3-b**). All estimated parameters are provided as medians (for

²²⁷ point estimates) and 2.5 and 97.5% quantiles (for 95% credible intervals) of

²²⁸ marginal posterior distributions. The credible intervals are crucial to quantify

Figure 2: Probability distributions of BCF_k (a) and BCF_{ss} (b). The middle-dotted line represents the median value, while left and right dotted lines are the 2.5 and 97.5% quantiles. (c) Measured (black dots) and predicted contaminant concentrations in the organism (µg.g^{-1}), where the median is represented by the orange plain line and the uncertainty band by the gray zone, delimited by the 2.5% and 97.5% quantiles in orange dotted lines.

 the uncertainty on parameter estimates.

2.4.3. Goodness-of-fit

²³¹ At the request of the user, several goodness-of-fit criteria are given in a

²³² prioritized order we chose based on their relevance and the case to interpret

²³³ them. The fitting quality of the model can be the first check using a posterior

²³⁴ predictive check plot: the idea is to compare each observed value to its prediction

²³⁵ from the fitted model at the corresponding concentration associated with its 95%

²³⁶ credible interval. If the fit is correct, we expect to see 95% of the observed values

²³⁷ falling within the credible intervals of their predictions. As shown on **Fig. 3-a**,

²³⁸ the observed values are read on the x-axis, while the y-axis reports the point

²³⁹ estimates predicted by the fitted model at the corresponding concentrations, as

²⁴⁰ well as their 95% credible intervals.

²⁴¹ The robustness of the inference process can be secondly checked using the com-

²⁴² parison of prior and posterior distributions. The overall expectation is to get a

²⁴³ narrower posterior distribution compared to the prior one, what reflects that

²⁴⁴ data contributed enough to precisely estimate parameters as illustrated in **Fig.**

²⁴⁵ **3-b**.

²⁴⁶ MOSAIC$_{\text{Bioacc}}$ also provides a correlation plot (Fig. S1), which allows the user

²⁴⁷ to visualize the correlation between parameters by projecting the joint posterior

²⁴⁸ distribution in a plot matrix in planes of parameter pairs. This plot also gives

²⁴⁹ marginal posterior distribution of each model parameter and Pearson correlation

²⁵⁰ coefficients. Correlations between parameters are important to consider in par-

²⁵¹ ticular when they are high (namely, greater than 0.75) what would mean that

²⁵² one parameter estimate could considerably influence the other.

²⁵³ The convergence of the MCMC chains is checked with the Gelman-Rubin di-

²⁵⁴ agnostic (Gelman & Rubin, 1992) expressed via the potential scale reduction

²⁵⁵ factor (PSRF) which is expected to be close to 1.00 (Fig. S1). It also could be

Figure 3: Some goodness-of-fit criteria as given in MOSAIC Bioacc : (a) Posterior Predictive Check (PPC) where black dots symbolyze measured data and vertical segments represent 95% credible intervals (coloured in green if it contains the observed value, in red otherwise) ; (b) prior (gray) and posterior (orange) distributions of parameters for the simplest TK model.

visually verified by the MCMC traceplot, which shows the time series of the sampling process leading to the posterior distribution ; we expect to obtain the overlapping of the three chains (Fig. S1).

The Deviance Information Criterion (DIC) is also given. It is a penalized deviance statistics accounting for the number of parameters and it is used to compare to the same data set. Sub-models with lower DIC values will be preferred. So, the DIC should be used when comparing several models based on different choices of parameters from combinations at the beginning of the web app.

2.5. Downloads

All results can be downloaded from the bottom of the page and a report can be downloaded with all results. The user can also download the R code corresponding to all graphs and calculations from its data set. This guarantees the transparency and the reproducibility of MOSAIC$_{\mathrm{Bioacc}}$ results. This code can be used as a stepping stone to change default options and to perform further analyses directly in R. For example, the user can modify figures at its convenience, or make several analyses on several data sets at the same time.

3. Conclusions

The development of such a tool raises important methodological issues as for example: (1) the automatization of the parameter inference, in particular for the use of the Bayesian inference; (2) the default options to fix and the choices left to the discretion of the users; (3) the outputs to primarily provide, the goodness-of-fit to show, the downloads to propose, the kind of analysis report to render. We are convinced that $MOSAIC_{Bioacc}$ may be particularly useful to estimate parameters of TK models and to predict chemical concentrations bioaccumulated in organism from standard data. Besides its user-friendliness, this application is free of use and guarantees the privacy of the uploaded data as well as the transparency and the reproducibility of the results, together with a short response time. This tool could thus be used by risk assessors and decision makers in their daily work of assessing dossiers, *e.g.*, for the authorization of substances on the market (Ockleford et al., 2018), since all results provided by $MOSAIC_{Bioacc}$ account for uncertainty and correlations between parameters. It can also be used by any environmental scientists, ecotoxicologists and managers when OECD n°305 tests are performed and need to be analysed. Further developments are already in progress to make available $MOSAIC_{Bioacc}$ as a new module dedicated to TK modelling approaches in the MOSAIC platform (Charles, Veber & Delignette-Muller 2018). Above all, this application will be extended to provide a prediction section that could allow to better design TK experiments in order to collect more relevant data for the model fitting and the estimation of bioaccumulation factors.

4. Authors' contributions

GM contributed to the Shiny application development, from a basic script provided by CL. The whole TK model has been developed by AR and CL. All

authors equally contributed to make the script the most generic as possible, then allowing to facilitate the Shiny application development. AR and SC led the writing of the manuscript.

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6. Data accessibility

Directly within MOSAIC$_{\text{Bioacc}}$ at https://mosaic.univ-lyon1.fr/bioacc.

References

Billoir, E., Delhaye, H., Clément, B., Delignette-Muller, M. L., & Charles, S. (2011). Bayesian modelling of daphnid responses to time-varying cadmium exposure in laboratory aquatic microcosms. Ecotoxicology and Environmental Safety, 74, 693-702.

Charles, S., Veber, P., & Delignette-Muller, M.L. (2018). MOSAIC: A Web-Interface for Statistical Analyses in Ecotoxicology. Environmental Science and Pollution Research, 25(12), 11295–302. https://doi.org/10.1007/s11356-017-9809-4.

Crookes, M., & Brooke, N. (2011). Estimation of fish Bioconcentration Factor (BCF) from depuration data. Environmental Agency. Bristol, UK.
https://assets.publishing.service.gov.uk/government/uploads/system/uploads/
attachment_data/file/291527/scho0811buce-e-e.pdf.

ECHA. (2017). Guidance on Information Requirements and Chemical Safety Assessment. Chapter R.11: PBT/vPvB Assessment. European Chemicals Agency, Version 3.0 – June 2017, 1-158.

European Commission. (2006). COMMISSION REGULATION (EU) No 1907/2006 of the European Parliament and of the Council of 18 December 2006 concerning the Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH), establishing a European Chemicals Agency, amending Directive 1999/45/EC and repealing Council Regulation (EEC) No 793/93 and Commission Regulation (EC) No 1488/94 as well as Council Directive 76/769/EEC and Commission Directives 91/155/EEC, 93/67/EEC, 93/105/EC and 2000/21/EC. Official Journal of the European Union, L 396, 1–849.

European Commission. (2012). COMMISSION REGULATION (EU) No 528/2012 of the European Parliament and of the Council of 22 May 2012 concerning the making available on the market and use of biocidal products. Official Journal of the European

Union, L 167, 1-123.

European Commission. (2013). COMMISSION REGULATION (EU) No 283/2013 of 1 March 2013 Setting out the Data Requirements for Active Substances, in Accordance with Regulation (EC) No 1107/2009 of the European Parliament and of the Council Concerning the Placing of Plant Protection Produc. Official Journal of the European Union, 3, 1–84.

Gelman, A., & Rubin, D. (1992). Inference from iterative simulation using multiple sequences. Statistical Science, 7, 457-511.

https://www.jstor.org/stable/2246093.

Lambert, P. C., Sutton, A. J., Burton, P. R., Abrams,K. R. & Jones, D. R. (2005). How Vague Is Vague? A Simulation Study of the Impact of the Use of Vague Prior Distributions in MCMC Using WinBUGS. Statistics in Medicine, 24(15), 2401–28. https://doi.org/10.1002/sim.2112.

Ockleford, C., Adriaanse, P., Berny, P., Brock, T., Duquesne, S., Grilli, S., . . . Hernandez-Jerez, A.F. (2018). Scientific Opinion on the State of the Art of Toxicokinetic/Toxicodynamic (TKTD) Effect Models for Regulatory Risk Assessment of Pesticides for Aquatic Organisms. EFSA Journal, 16(8).

https://doi.org/10.2903/j.efsa.2018.5377.

OECD. (2012). Test No. 305: Bioaccumulation in Fish: Aqueous and Dietary Exposure. In OECD Guidelines for the Testing of Chemicals. OECD Editition, 3. Paris. https://doi.org/https://doi.org/10.1787/9789264185296-en.

Plummer, M. (2019). Rjags: Bayesian Graphical Models Using MCMC. R package version 4-10. https://cran.r-project.org/package=rjags.

R Core Team. (2020). R: A Language and Environment for Statistical Computing. Vienna, Austria: R Foundation for Statistical Computing. https://www.r-project.org/.

Raftery, A.E., & Lewis, S.M. (1992). Practical Markov Chain Monte Carlo: Comment: One Long Run with Diagnostics: Implementation Strategies for Markov Chain Monte Carlo. Statistical Science, 7, 493–97.

Ratier, A., Lopes, C., Labadie, P., Budzinski, H., Delorme, N., Quéau, H., . . . Babut, M. (2019). A Bayesian Framework for Estimating Parameters of a Generic Toxicokinetic Model for the Bioaccumulation of Organic Chemicals by Benthic Invertebrates: Proof of Concept with PCB153 and Two Freshwater Species. Ecotoxicology and Environmental Safety, 180. https://doi.org/10.1016/j.ecoenv.2019.04.080.

Richards, R. G., & Chaloupka, M. (2009). Temperature-Dependent Bioaccumulation of Copper in an Estuarine Oyster. Science of the Total Environment, 407(22), 5901–5906. https://doi.org/10.1016/j.scitotenv.2009.07.039.

Tan, Yu-mei, Chan, M., Chukwudebe, A., Domoradzki, J., Hack, C.E., Hinderliter, P., . . . Hirasawa, K. (2020). "PBPK Model Reporting Template for Chemical Risk Assessment Applications." Regulatory Toxicology and Pharmacology. https://doi.org/10.1016/j.yrtph.2020.104691.