

1 **Exposure to a mixture of BMAA and MCLR synergistically modulates behavior in**
2 **larval zebrafish while exacerbating molecular changes related to**
3 **neurodegeneration.**

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24 **Abstract**

25 Exposure to toxins produced by cyanobacteria (i.e., cyanotoxins) is an emerging health
26 concern due to their increased occurrence and previous associations with
27 neurodegenerative disease including amyotrophic lateral sclerosis (ALS). The objective
28 of this study was to evaluate the neurotoxic effects of a mixture of two co-occurring
29 cyanotoxins, β -methylamino-L-alanine (BMAA) and microcystin leucine and arginine
30 (MCLR), using the larval zebrafish model. We combined high-throughput behavior-
31 based toxicity assays with discovery proteomic techniques to identify behavioral and
32 molecular changes following 6 days of exposure. While neither toxin caused mortality,
33 morphological defects, or altered general locomotor behavior in zebrafish larvae, both
34 toxins increased acoustic startle sensitivity in a dose-dependent manner by at least 40%
35 ($p < 0.0001$). Furthermore, startle sensitivity was enhanced by an additional 40% in
36 larvae exposed to the BMAA/MCLR mixture relative to those exposed to the individual
37 toxins. Supporting these behavioral results, our proteomic analysis revealed a 4-fold
38 increase in the number of differentially expressed proteins (DEPs) in the mixture-
39 exposed group. Additionally, prediction analysis reveals activation and/or inhibition of 8
40 enriched canonical pathways (enrichment p -value < 0.01 ; z -score $\geq |2|$), including ILK, Rho
41 Family GTPase, RhoGDI, and calcium signaling pathways, which have been implicated
42 in neurodegeneration. We also found that expression of TDP-43, of which cytoplasmic
43 aggregates are a hallmark of ALS pathology, was significantly upregulated by 5.7-fold
44 following BMAA/MCLR mixture exposure. Together, our results emphasize the
45 importance of including mixtures of cyanotoxins when investigating the link between
46 environmental cyanotoxins and neurodegeneration as we reveal that BMAA and MCLR

47 interact *in vivo* to enhance neurotoxicity.

48

49 **1. Introduction**

50 Amyotrophic lateral sclerosis (ALS) is the most common neurodegenerative
51 disease of midlife and is rapidly fatal with a median survival period of three years from
52 symptom onset (Brown and Al-Chalabi 2017). It is defined by a progressive loss of both
53 upper and lower motor neurons, resulting in muscle spasticity, weakness, and atrophy
54 (Swinnen and Robberecht 2014). Approximately 10% of ALS cases are classified as
55 familial due to the inheritance of single gene mutations (Renton et al. 2014). For the
56 remaining 90% of cases the disease etiology is unknown and likely stems from a
57 complex interplay between genetic and environmental factors. (Ingre et al. 2015; Jones
58 2009). Although the contribution of environmental factors to sporadic ALS (sALS) is
59 difficult to assess as the search space is infinite, several studies have associated ALS
60 incidence with exposure to heavy metals, pesticides, and electromagnetic fields
61 (reviewed in (Bozzoni 2016)). In addition, there is also strong evidence that exposure to
62 cyanotoxins is a major risk factor for sALS (Bradley and Mash 2009).

63 The link between beta-methylamino-L-alanine (BMAA), a toxin produced by a
64 diverse taxa of cyanobacteria (Cox et al. 2005) and sALS was first observed on the
65 island of Guam in the 1950s (Kurland and Mulder 1955). The indigenous population of
66 Guam succumbed to an ALS/parkinsonism-dementia (PD) neurodegenerative complex
67 with a 100-fold greater incidence than the rest of the world (Bradley and Mash 2009).
68 The elevated rates of ALS/PD in Guam were attributed to BMAA exposure as the
69 indigenous population consumed flour made from BMAA-containing cycad seeds as

70 well as flying foxes in which BMAA was biomagnified up to 10,000-fold greater than in
71 free living bacteria (3,556 $\mu\text{g/g}$ BMAA) (Banack and Cox 2003). Since then, numerous
72 studies have implicated BMAA in sALS cases outside of Guam, including clusters of
73 ALS along the French Mediterranean coast, New Hampshire, and Maryland (Caller et
74 al. 2009; Field et al. 2013; Masseret et al. 2013). These epidemiological studies provide
75 evidence that exposure to BMAA is associated with neurodegeneration. Epidemiological
76 findings are further supported by laboratory studies in which BMAA was found to cause
77 neurotoxic effects consistent with neurodegenerative disease (Beri et al. 2017; Karlsson
78 et al. 2017). Furthermore, neonatal BMAA exposure in rats has been shown to produce
79 motor defects in rats (Scott et al. 2017), indicating that exposure during neural
80 developmental may enhance sALS risk. However, a major limitation for these and many
81 other BMAA studies is that BMAA is just one of many toxic metabolites produced by
82 cyanobacteria, some of which have been reported to co-occur with BMAA around the
83 world (Banack et al. 2015; Sabart et al. 2015). Thus, to obtain a more thorough
84 understanding of the risk posed by exposure to cyanotoxic blooms, it is essential to
85 investigate the toxicity of other cyanotoxins with BMAA.

86 BMAA at low concentrations ($\sim 10 \mu\text{M}$) in combination with other non-cyanotoxic
87 neurotoxins has been found to potentiate neuronal damage *in vitro* (Lobner et al. 2006).
88 More recently, our laboratory demonstrated that co-exposure to BMAA and its isomers
89 (i.e., AEG and 2,4DAB) at low concentrations ($\sim 166 \mu\text{M}$) produces a synergistic
90 interaction *in vitro*, perturbing regulation of various canonical pathways, bioprocesses,
91 and upstream regulators involved in neurodegenerative processes (Martin et al. 2019).
92 Recent studies have shown that microcystin leucine-arginine (MCLR) is also a potent

93 neurotoxin (Tzima et al. 2017; Wu et al. 2016), and like BMAA, MCLR can
94 bioaccumulate in tissues (Wang et al. 2008; Zhao et al. 2015). Microcystins are the
95 most abundant cyanotoxins in the environment and have been shown to co-occur with
96 BMAA (Banack et al. 2015; Jungblut et al. 2018; Metcalf et al. 2012). Therefore, co-
97 exposure to BMAA and MCLR is of increasing toxicological significance.

98 To identify potential neurotoxic effects *in vivo*, we exposed larval zebrafish to
99 BMAA and/or MCLR and assessed neural function with a set of behavioral assays using
100 a high-throughput testing platform. While neither BMAA nor MCLR caused changes in
101 locomotion, both toxins increased acoustic startle sensitivity in a dose-dependent
102 manner. Furthermore, a mixture of BMAA and MCLR enhanced toxicity in the startle
103 assay. Finally, we examined the protein profile of larval zebrafish exposed to the
104 BMAA/MCLR mixture and identified molecular signatures consistent with
105 neurodegeneration, including upregulation of the ALS-associated protein TDP-43
106 (Mackenzie et al. 2010). Together, our data highlight the importance of studying toxic
107 mixtures and reveal novel mechanisms that may link cyanotoxin exposure to sALS.

108 **2. Materials and Methods**

109 2.1. Chemicals

110 Synthetic BMAA standards were obtained from Sigma Aldrich (St. Louis, MO, USA),
111 and purified MCLR (purity > 95%) was obtained from Enzo Life Sciences, Inc.
112 (Farmingdale, NY, USA). Water, acetonitrile, methanol, acetic acid, and formic acid
113 were all Optima LC–MS grade solvents purchased from Fisher Scientific (Tewksbury,
114 MA, USA). A stock solution of BMAA at 10 mg.mL⁻¹ and MCLR at 1 mg.mL⁻¹ was used

115 for all dilutions. All BMAA dilutions were prepared in HPLC grade water while MCLR
116 dilutions were prepared in DMSO.

117 2.2. Zebrafish husbandry and exposures

118 All animal use and procedures were approved by the North Carolina State
119 University IACUC. Zebrafish (*Danio rerio*) embryos from multiple crosses of wild-type
120 tupfel longfin (TLF) strain adults were collected and placed into Petri dishes containing
121 E3 medium, and unfertilized eggs were removed as described previously (Burgess and
122 Granato 2007). Embryos from all clutches were mixed and randomly sorted into 24 well
123 plates (8-10 animals per well) containing 1 mL of E3 per well.

124 At 6 hpf, all E3 was removed and replaced with vehicle (HPLC-grade water), 100,
125 500, or 1000 μ M BMAA in E3, vehicle (DMSO), 1, 2.5, 5, or 10 μ M MCLR in E3, or 100
126 μ M BMAA plus 1 μ M MCLR in E3. All treatments were performed in triplicate and were
127 repeated in each of 3 separate experiments. Embryos were incubated at 29°C on a
128 14h:10h light-dark cycle, and 100% of the media was exchanged for fresh solutions
129 daily. Embryos/ larvae were exposed to treatments until 6 days post fertilization (6 dpf).

130 2.3 Behavior assays and analysis

131 All 6 dpf larvae were thoroughly screened for developmental defects, and those
132 with uninflated swim bladder, edema, or other morphological defects were removed
133 from analysis. Screened larvae were adapted to the testing lighting and temperature
134 conditions for 30 minutes prior to testing. Behavior testing was done as previously
135 described (Burgess and Granato 2007; Marsden et al. 2018). Briefly, 6 dpf larvae were
136 transferred to individual 9 mm round wells on a 36-well laser-cut acrylic testing grid.
137 Larvae acclimated for 5 min and then spontaneous locomotor activity was recorded for

138 18.5 min at 640 x 640 px resolution at 50 frames per sec (fps) using a Photron mini UX-
139 50 high-speed camera. The same set of larvae were then presented with a total of 60
140 acoustic stimuli, 10 at each of 6 intensities (13.6, 25.7, 29.2, 35.5, 39.6, and 53.6 dB),
141 with a 20s interstimulus interval (ISI). Startle responses were recorded at 1000 fps.
142 Stimuli were delivered by an acoustic-vibrational shaker (Bruel and Kjaer) to which the
143 testing grid was directly mounted. All stimuli were calibrated with a PCB Piezotronics
144 accelerometer (#355B04) and signal conditioner (#482A21), and voltage outputs were
145 converted to dB using the formula $\text{dB} = 20 \log V$. Analysis of recorded behaviors was
146 done using FLOTE software as described previously (Burgess and Granato 2007;
147 Marsden et al. 2018). Short-latency C-bends (SLCs) and long-latency C-bends (LLCs)
148 were determined by defined kinematic parameters. A startle sensitivity index was
149 calculated for individual larvae by calculating the area under the curve of startle
150 frequency versus stimulus intensity using Prism 8 software (GraphPad). Statistical
151 analyses were performed using JMP pro 14 from SAS institute, Cary, NC. Data were
152 analyzed for effects between the groups (comparison of means), using Tukey-Kramer
153 HSD, Alpha 0.05. Violin plots were generated using Prism 8.

154 2.4 Proteomics analysis

155 *Sample preparation and LC MS/MS*

156 Details of sample preparation, protein extraction and digestion via filter aided
157 sample preparation (FASP) can be found in Supplemental Methods. Details regarding
158 the LC-MS/MS data collection are also provided in the Supplemental Methods. Raw
159 data files obtained in this experiment have been made available on the Chorus LC-MS
160 data repository and can be assessed under the project ID#1679.

161 *Proteomics Data Analysis*

162 Details for the label free quantitation (LFQ) have been previously described here
163 (Martin et al. 2019). In brief, LFQ was performed in MaxQuant (version.1.5.60), and
164 data were searched against the *Danio rerio* Swiss Prot protein database (# protein
165 sequences = 56 281, accessed 03/22/2019). Comparison of LFQ intensities across the
166 whole set of measurements was investigated using Perseus software (version 1.5.1.6),
167 where calculation of statistical significance was determined using two-way Student-t test
168 and FPR ($p \leq 0.05$).

169 *Pathway Analysis*

170 Ingenuity Pathway Analysis (IPA) software was used to identify the function,
171 specific processes, and enriched pathways of the differentially expressed proteins using
172 the “Core Analysis” function. Only significantly differentially expressed proteins ($p \leq$
173 0.05) were submitted to IPA. We used an empirical background protein database to
174 evaluate the significance of pathway enrichment. The database was created by using all
175 of the proteins that were detected in our samples (Bereman et al. 2018; Khatri and
176 Drăghici 2005).

177 **3. Results**

178 An overview of the experimental design is illustrated in **Figure 1**. In brief, we first
179 conducted a dose-response study to determine the no observed adverse effect levels
180 (NOAELs) to be implemented in subsequent mixture analyses. Zebrafish larvae were
181 exposed to increasing concentrations of BMAA or MCLR from 6 hours post-fertilization
182 (hpf) to 6 days post-fertilization (dpf). At 6 dpf, neurotoxicity was evaluated via two
183 behavioral assays: spontaneous movement and acoustic startle response assays.

184 Based on these data, a mixture was created using BMAA and MCLR at their respective
185 NOAELs in which zebrafish larvae were exposed as before, followed by behavior
186 analysis to identify potential interactions between BMAA and MCLR. Finally, to
187 investigate perturbed molecular pathways associated with cyanotoxic mixture exposure,
188 the mixture-exposed group and their respective controls were subjected to shotgun
189 proteomics.

190 *3.1. BMAA and MCLR Dose Response Study: identification of NOAELs*

191 To determine if exposure to environmentally relevant concentrations of BMAA or
192 MCLR cause neurotoxicity in wild-type zebrafish, we treated TLF strain embryos from 6
193 hpf to 6 dpf with increasing concentrations of BMAA (100, 500, and 1000 μ M) and
194 MCLR (1, 2.5, 5, and 10 μ M). We did not observe increased mortality, morbidity, or any
195 overt developmental phenotypes in any of the exposed groups of larvae. First, we
196 examined the effect of BMAA and MCLR on general locomotion (**Figure 2**) using a
197 custom built, high-throughput behavior platform and unbiased, automated FLOTE
198 tracking and analysis software (Burgess and Granato 2007). In order to investigate if
199 various concentrations of BMAA and/or MCLR could alter spontaneous movement, 6
200 dpf larvae were adapted to the testing conditions for 30 min, transferred to a multi-well
201 grid mounted below a high-speed camera, habituated for 5 additional minutes, and then
202 their spontaneous movements were recorded for 18.5 min. We detected no significant
203 differences in total distance travelled for zebrafish larvae treated with either BMAA or
204 MCLR compared to their respective vehicle controls (**Figure 2A**). Average speed was
205 also unchanged in all groups, except for larvae treated with 1000 μ M BMAA, whose
206 speed was significantly reduced (**Figure 2B**). We also examined the frequency of

207 turning and swimming behaviors, as defined by specific kinematic parameters (Hao et al. 2012). There were no significant differences in the ratio of turns to swims in BMAA or
208 al. 2012). There were no significant differences in the ratio of turns to swims in BMAA or
209 MCLR treated larvae (**Figure 2C**). The overall frequency of these movements was also
210 unchanged, except for a slight increase in turn frequency in 10 μ M MCLR-treated larvae
211 (**Supplemental Figure 1**). These data indicate that developmental exposure to BMAA
212 or MCLR does not substantially affect general locomotor activity in larval zebrafish.

213 We then we examined sensorimotor function using an acoustic startle assay
214 consisting of 60 total stimuli, 10 at each of 6 intensities with a 20 sec inter-stimulus
215 interval (**Figure 3**). In response to an acoustic stimulus, zebrafish larvae perform one of
216 two types of high-velocity startle behaviors: Short-latency C-bends (SLCs), which rely
217 on the Mauthner neurons (Burgess and Granato 2007), or Long-latency C-bends
218 (LLCs), which are independent of the Mauthner cells but require a set of prepontine
219 neurons (Marquart et al. 2019). To investigate if BMAA or MCLR alters startle
220 performance, we measured SLC and LLC frequency across the 60-stimulus assay
221 (**Figure 3**). **Figure 3A** highlights both the SLC and LLC response frequency disparities
222 between zebrafish larvae exposed to 1000 μ M BMAA and vehicle control. 1000 μ M
223 BMAA increases SLC responses while decreasing LLC responses, indicating that
224 BMAA shifts the behavioral response bias toward SLCs.

225 To quantify SLC and LLC sensitivity, we calculated the area under the startle
226 frequency curves in **Figure 3A** for each individual larva to create a startle sensitivity
227 index (Marsden et al. 2018). Both BMAA and MCLR increased SLC sensitivity in a
228 dose-dependent manner (**Figure 3B**). LLC responses decreased in a dose-dependent
229 manner in both BMAA and MCLR-treated larvae, supporting an overall shift in response

230 bias (**Figure 3C**). These data indicate that environmentally relevant concentrations of
231 BMAA and MCLR enhance activity of the SLC circuit. In addition, these startle
232 sensitivity data reveal NOAELs for BMAA (100 μ M) and MCLR (1 μ M), with NOAEL was
233 defined as the highest non-statistically significant dose tested.

234 *3.2. BMAA and MCLR Mixture Study: Interaction Amongst Cyanotoxins at a Behavioral* 235 *Level*

236 We next aimed to assess whether BMAA and MCLR interact *in vivo* by
237 measuring the effects of a mixture of BMAA and MCLR at their respective NOAELs in
238 larval zebrafish. We exposed wild-type zebrafish embryos from 6 hpf to 6 dpf to 4
239 different treatment conditions: 1) vehicle controls, 2) 100 μ M BMAA, 3) 1 μ M MCLR,
240 and 4) 100 μ M BMAA + 1 μ M MCLR. As before, no overt developmental or
241 morphological defects were observed in any exposed larvae. We then measured
242 general locomotor activity and sensorimotor function using the same assays described
243 above. In this cohort of animals, 100 μ M BMAA very slightly decreased total distance
244 traveled over 18.5 min (**Figure 4A**), and all three treatment groups showed a minor
245 reduction in average speed during the assay (**Figure 4B**). No differences were
246 observed in turning or swimming behaviors (**Figure 4C**). These data reinforce the
247 results from our dose-response study that BMAA and MCLR do not substantially alter
248 locomotor activity.

249 We next measured startle frequency in the same 4 groups of larvae. **Figure 4D**
250 highlights both the SLC and LLC response frequency disparities between zebrafish
251 larvae exposed to BMAA/MCLR mixture solution (101 μ M) and controls. Neither 100 μ M
252 BMAA nor 1 μ M MCLR altered startle behavior, as both SLC (**Figure 4E**) and LLC

253 sensitivity indices (**Figure 4F**) were unchanged. The 101 μM BMAA/MCLR mixture,
254 however, significantly enhanced SLC sensitivity (**Figure 4E**) while leaving LLC
255 sensitivity unchanged (**Figure 4F**), in contrast to the effect of BMAA alone (**Figure 3**).
256 These data demonstrate not only that BMAA and MCLR interact *in vivo* to enhance SLC
257 circuit activity, but because of the different effects of the mixture and the individual
258 toxins on LLC responses (**Figure 3C vs. Figure 4F**), they also suggest that different
259 cellular and/or molecular mechanisms are impacted by the mixture.

260

261 3.3 Global Proteomics Study: Interaction amongst cyanotoxins at a molecular level

262 To explore the molecular underpinnings associated with these behavioral
263 phenotypes, we performed shotgun proteomics on larval zebrafish exposed to 100 μM
264 BMAA and 1 μM MCLR alone and in combination. After behavioral testing, we carefully
265 collected and flash-froze the treated zebrafish larvae, followed by protein extraction and
266 digestion. Proteomes of larvae for each treatment condition were analyzed by LC-
267 MS/MS, and approximately 3100 proteins were identified in each sample.

268 Differentially expressed proteins (DEPs) were determined by comparing the
269 mean abundance within treatment to the control group for each protein using a two-way
270 Student-t-test ($P < 0.05$) (Tyanova et al. 2016). DEPs in all treatments can be found in
271 **Supplemental Tables 2-4**. Volcano plots were used to visualize statistically significant
272 differences in protein abundance across treatments in comparison to controls
273 (**Supplemental Figure 2**). Notably, the BMAA/MCLR mixture induced the greatest
274 molecular perturbation, with 259 DEPs compared to 79 for BMAA and 112 for MCLR,
275 representing a 2.5-fold increase for the mixture-exposed group (**Figure 5A**). Although

276 minimal overlap in DEPs between treatments was observed, we were intrigued by the
277 nine proteins that were significantly differentially expressed in all three treatment groups
278 (**Figure 5B; Supplemental table 1**). Out of these nine proteins, four proteins were
279 mapped in the enrichment analysis (**Table 1**), and their general cellular functions
280 include roles in cellular assembly, organization, and development. Interestingly,
281 exposure to the BMAA/MCLR mixture also enhanced the abundance of these four
282 DEPs by at least 2.5-fold relative to the individual cyanotoxins. These results reflect an
283 enhanced toxicity after BMAA/MCLR mixture exposure *in vivo*.

284 To further analyze the identified DEPs across treatments, we performed
285 enrichment analysis to identify significantly perturbed pathways. **Supplemental table 5**
286 lists all canonical pathways found to be significantly perturbed ($z\text{-score} \geq |2|$) along with
287 their associated z-scores. Exposure to the BMAA/MCLR mixture enhanced the
288 predicted activation/inhibition of eight canonical pathways ($z\text{-score} \geq |2|$) compared to
289 BMAA (zero) and MCLR (two), which supports the observation of a synergistic
290 interaction *in vivo* (**Figure 5C**). RhoGDI signaling ($z\text{ score} = -2.121$) and calcium
291 signaling ($z\text{ score} = -2.449$) were inhibited, while signaling by Rho Family GTPases ($z\text{-}$
292 $\text{score} = 2.121$) and ILK ($z\text{-score} = 2.121$) were activated. Although exposure to MCLR at 1
293 μM did not cause behavioral modulation, two canonical pathways were significantly
294 affected: 1) inhibition of RhoGDI signaling ($z\text{-score} = -2$), and 2) activation of signaling by
295 Rho family GTPases ($z\text{-score} = 2$) (**Figure 5C**). All four canonical pathways impacted by
296 mixture exposure are broadly associated with neurotoxic processes related to
297 reorganization of the actin cytoskeleton. Moreover, these pathway analysis results are
298 reinforced by our protein interaction network analysis, in which we found significant

299 differential regulation of key proteins associated with skeletal/muscular disorder and
300 cellular assembly/organization (**Supplemental Table 6**). Within these networks, key
301 neuronal proteins, including cell division cycle 42 (CDC-42; Enrichment p-value=0.0049;
302 Log₂ fold-change=1.8525), glutamate dehydrogenase 1 (GLUD1; Enrichment p-
303 value=0.0299; Log₂ fold-change= 2.6413), and the ALS-associated TDP-43 (TARDBP;
304 Enrichment p-value=0.0299; Log₂ fold-change= 2.5508) were significantly upregulated.
305 Because TDP-43 is significantly associated with ALS disease pathology, we looked at
306 expression of TDPBP and TDPBPL in all treatment groups. TDPBP was increased by
307 MCLR exposure, but not BMAA exposure, and this increase was further enhanced by
308 exposure to the BMAA/MCLR mixture (**Table 2**). Together, these data indicate that low
309 concentrations of BMAA and MCLR in combination impact neurodegenerative
310 processes in larval zebrafish.

311

312 **4. Discussion**

313 Since the 1950s, BMAA has been investigated for its potential to contribute to
314 neurodegenerative diseases, including amyotrophic lateral sclerosis (ALS) (Reed et al.
315 1966). However, BMAA is only one of thousands of toxic metabolites produced by
316 cyanobacteria (Dolman et al. 2012). While an increasing number of studies have singly
317 addressed BMAA and its adverse effects, major knowledge gaps remain regarding the
318 neuropathological effects of combined exposure to a cocktail of cyanotoxins. A number
319 of studies have reported that cyanotoxins co-occur in natural environments (Banack et
320 al. 2015; Jungblut et al. 2018; Metcalf et al. 2008; Sabart et al. 2015), including BMAA
321 and the most abundant cyanotoxin, MCLR (McKindles et al. 2019). Although first

322 considered to be primarily a hepatotoxin, MCLR has recently been shown to have
323 neurotoxic effects both *in vitro* and *in vivo* (Li et al. 2012; Li et al. 2015; Wang et al.
324 2017; Wu et al. 2016). Thus, because BMAA and MCLR are ubiquitously present in the
325 environment, have been previously detected together, and are neurotoxic, our study
326 addresses the important question of whether they interact *in vivo* to enhance adverse
327 effects. Building on prior work demonstrating that cyanotoxins can interact *in vitro* (Main
328 et al. 2018; Martin et al. 2019), we show that 1) both BMAA and MCLR alter the
329 behavior of larval zebrafish (**Figure 3**), 2) a simple binary mixture of BMAA and MCLR
330 at low concentrations enhances behavioral neurotoxicity (**Figure 4**), and 3) BMAA and
331 MCLR synergistically alter molecular changes associated with neuromuscular
332 dysfunction (**Figure 5**).

333 Larval zebrafish have emerged as a powerful vertebrate model for studying
334 neural development and behavioral circuits, as well as for translational toxicology (Tal et
335 al. 2020; Wolman and Granato 2012). Although studying larvae is less directly relevant
336 for studies of neurodegeneration, there is increasing evidence that developmental
337 exposures can lead to disease later in life (Heindel and Vandenberg 2015). Indeed,
338 neonatal exposure to BMAA has been found to cause motor defects and
339 neurodegeneration in adult rats (Scott and Downing 2019; Scott et al. 2017). Thus,
340 understanding the developmental impact of cyanotoxin exposure is critical for identifying
341 potential early indicators of degenerative pathology. Here, we show that larval zebrafish
342 behavior is modulated upon exposure to relatively low concentrations of both BMAA and
343 MCLR in a dose-dependent manner. Although only traces of cyanotoxins have been
344 found in large natural bodies of water (mean=41 $\mu\text{g}\cdot\text{L}^{-1}$) (Wiltsie et al. 2018), BMAA and

345 MCLR can be found at relatively high concentrations (from ~ 0.02 to 8 mg.kg^{-1}) in
346 freshwater fish, crustaceans, and other types of seafood (Lance et al. 2018; Sahin et al.
347 1996) due to bioaccumulation through the food web. Thus, our mixture paradigm is an
348 appropriate model of natural exposures.

349 Previous studies in larval zebrafish have indicated that BMAA may cause clonus-
350 like convulsions (Purdie et al. 2009) and pericardial edema and altered heart rate
351 (Frøyset et al. 2016; Purdie et al. 2009). We did not observe these effects, but this could
352 be due to differences in strain, embryo medium, exposure route, and analysis methods.
353 In contrast to our data showing no effect on locomotion in bright light conditions (**Figure**
354 **2**), MCLR has previously been shown to reduce activity in zebrafish larvae in a light-
355 dark assay (Tzima et al. 2017; Wu et al. 2016). This discrepancy could also arise from
356 strain and media differences, but in agreement with these studies, we did not observe
357 mortality or morphological defects in MCLR-exposed larvae. However, we detected
358 significant, dose-dependent changes in acoustic startle behavior upon exposure to both
359 BMAA and MCLR (**Figure 3**). These data reveal a need for greater standardization in
360 zebrafish rearing methods, and they also show that our acoustic startle assay using
361 high-speed cameras and kinematic analysis may be a broadly useful and highly
362 sensitive addition to standard behavioral neurotoxicity testing.

363 The increased frequency of Mauthner-cell dependent short-latency startles
364 (SLCs) in BMAA and MCLR-treated larvae indicates that the underlying sensorimotor
365 circuit is hyperexcitable. BMAA is known to directly agonize glutamatergic receptors
366 (Chiu et al. 2012; 2013), so the startle hypersensitivity in BMAA-treated fish could reflect
367 that startle circuit neurons fire more easily following acoustic stimuli. Alternatively,

368 hypersensitivity from exposure to these cyanotoxins could result from a reduction in
369 inhibitory control of the startle circuit. Interestingly, both of these mechanisms have
370 implications for ALS pathology, as excitotoxicity either from direct overstimulation of
371 excitatory pathways, or from a loss of inhibitory input have been implicated in motor
372 neuron death (Martin et al. 2012). Furthermore, our data show that BMAA and MCLR
373 interact to enhance startle sensitivity at their respective NOAELs (**Figure 4**). To the best
374 of our knowledge, only one previous study has examined the effects of exposure to
375 BMAA and MCLR as a mixture. Anxiety-like behavior, exploratory behavior, and general
376 locomotion were all found to be unchanged by acute exposure to a BMAA/MCLR
377 mixture in the adult C57BL/6 mouse model (Myhre et al. 2018). This could indicate that
378 the effects of the BMAA/MCLR mixture are limited to specific brain circuits, and/or that
379 these neurotoxins exert their effects more strongly during early developmental stages
380 (Karlsson et al. 2012; Scott et al. 2017). Future studies will examine the longer-term
381 effects of developmental exposure to BMAA and MCLR.

382 To understand how BMAA and MCLR drive neurotoxicity, we used a label-free
383 proteomics approach to identify the molecular pathways disrupted by BMAA/MCLR
384 exposure in larval zebrafish. Our proteomics data display a clear trend towards
385 enhanced toxicity in the mixture exposed group versus single exposures (**Figure 5A**),
386 further supporting the conclusion from our behavioral data that the two toxins interact *in*
387 *vivo*. Interestingly, DEPs displayed minimal overlap between treatments (**Figure 5B**),
388 suggesting they act through different modes of action. It is notable that the
389 BMAA/MCLR mixture impacted multiple critical cellular pathways, including signaling by
390 ILK, Rho Family GTPases, RhoGDI, and calcium (**Figure 5C**), which all impinge on

391 regulation of the actin cytoskeleton. For example, overexpression of proteins in the Rho
392 Family GTPase pathway such as CDC42 has specific effects on the actin filamentous
393 system (Nobes and Hall 1995). CDC42 has a well-established role in triggering the
394 formation/assembly of stress fibers mediated by Arp2/3-dependent actin nucleation
395 (Aspenström 2019). These data are consistent with prior work showing that loss-of-
396 function mutations in *cytoplasmic FMRP-interacting protein 2 (cyfip2)*, a key regulator of
397 Arp2/3-mediated actin polymerization, cause startle hypersensitivity in larval zebrafish
398 similar to that seen with BMAA/MCLR exposure (Marsden et al. 2018). In addition,
399 previous reports show that MCLR induces neurotoxicity by triggering reorganization of
400 actin cytoskeleton components (Li et al. 2012; Meng et al. 2011) by inhibiting
401 serine/threonine-specific protein phosphatases (PPs) 1 and 2A (Huynh-Delerme et al.
402 2005; MacKintosh et al. 1990). Here, we show here that MCLR in combination with
403 BMAA at low concentrations inhibits expression of these same protein phosphatases
404 associated with cytoskeletal organization (PP1CAB (Q7ZVR3), enrichment p-
405 value=0.0106; Log₂ fold-change=-2.476; PP2CA (F1Q6Z7), enrichment p-value=0.0110;
406 Log₂ fold-change=-4.03; **Supplemental table 3**). Together, our molecular proteomics
407 data support the idea that acoustic startle hypersensitivity may be an early indicator of
408 neuronal stress.

409 That our unbiased proteomic analysis also revealed an upregulation of TDP-43 in
410 BMAA/MCLR-exposed larvae (**Table 2, Supplemental Table 6**) is particularly striking.
411 Cytoplasmic TDP-43 inclusions are the key pathological hallmark in 98% of sALS
412 cases (Mackenzie et al. 2010). Although our results cannot verify the sub-cellular
413 localization of upregulated TDP-43, previous reports have shown that overexpression of

414 TDP-43 in the cytoplasm leads to depletion of nuclear TDP-43, which has detrimental
 415 effects in mice (Fratta et al. 2018; Wils et al. 2010). While MCLR—but not BMAA—
 416 exposure also increased TDP-43 expression, this increase was exacerbated by the
 417 mixture, indicating that exposure to multiple cyanotoxins may enhance sALS disease
 418 processes. While the molecular mechanisms that specifically drive cyanotoxin-mediated
 419 neurotoxicity are not fully understood, our data support a model in which cyanotoxin
 420 mixtures cause neural dysfunction through multiple disease-associated pathways.

421 Together, our data provide new evidence that cyanotoxins synergistically interact
 422 *in vivo* to cause changes not only at the molecular level but also at the whole-organism
 423 level, as demonstrated by altered behavioral performance. Future work will seek to link
 424 specific molecular pathways, behavior regulation, and neuronal dysfunction, with the
 425 goal of revealing novel therapeutic and/or diagnostic targets for intractable
 426 neurodegenerative diseases such as ALS.

427

	Protein IDs	Gene Names	P Value	Log₂ Fold Change
MCLR (1 μM)	Q4QRD2	myl4	0.0029	1.5834
	Q9I8V1	actc1b	0.0076	1.2424
	Q7T368	pdhb	0.0187	1.0509
	Q6P0V6	rpl8	0.0491	0.8516
BMAA (100μM)	Q6P0V6	rpl8	0.0023	-0.892
	Q4QRD2	myl4	0.0047	1.0115
	Q9I8V1	actc1b	0.033	-0.9068
	Q7T368	pdhb	0.0487	-0.9224
Mixture (101μM)	Q9I8V1	actc1b	0.0002	5.3546
	Q4QRD2	myl4	0.0135	2.7762
	Q6P0V6	rpl8	0.0215	1.6328
	Q7T368	pdhb	0.025	2.3007

428 **Table 1:** Shared DEPs amongst treatments.

429

	Protein IDs	Gene Names	P Value	Log₂ Fold Change
BMAA (100µM)	Q802C7	tardbp	0.3614	-1.3194
	Q6NYX2	tardbpl	0.9402	0.0674
MCLR (1µM)	Q802C7	tardbp	0.0245	1.6390
	Q6NYX2	tardbpl	0.7002	0.5127
Mixture (101µM)	Q802C7	tardbp	0.0159	2.5508
	Q6NYX2	tardbpl	0.0520	2.2426

430 **Table 2:** Impact of cyanotoxin exposure on TDP pathway proteins.

431 **Conflicts of interest**

432 The authors declare no conflict of interest.

433 **Acknowledgments**

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438 and technical support for all experiments.

439 **Abbreviations**

440 SLC, short latency c-startle; LLC, long latency c-startle; HPLC, high-performance liquid
441 chromatography; LC/MS, high-pressure liquid chromatography combined mass
442 spectrometry; IPA, ingenuity pathway analysis; ANOVA, analysis of variance; GO, gene
443 ontology; BMAA, β -methylamino-L-alanine; MCLR, microcystin leucine-arginine; DEPs,
444 differentially expressed proteins.

445 **Figure Legends**

446 **Fig.1. Experimental Design.** (A) Cyanotoxin exposure plan for zebrafish from 6 hpf to
447 6 dpf. (B) High throughput behavior testing apparatus: multi-well testing grid is mounted
448 on an acoustic shaker above an infrared (IR) LED array and below an IR-sensitive high-
449 speed camera. A white LED is mounted above the grid to simulate daylight conditions.
450 Videos are analyzed with automated tracking software (FLOTE). (C) Proteomics
451 workflow: zebrafish larvae from the mixture exposed groups were pooled for protein
452 extraction and tryptic digestion of extracted proteins into peptides. nLC-MS/MS label
453 free protein quantitation via MaxQuant statistical analysis via Perseus software and
454 enrichment analysis via ingenuity pathway analysis (IPA).

455
456 **Fig. 2. BMAA and MCLR do not substantially alter general locomotor activity.** (A)
457 Violin plots depict the total distance travelled during the 18.5 min spontaneous
458 movement assay for each larva. (B) Average speed across the same assay. (C) The
459 ratio of turning movements to swimming movements performed during the spontaneous
460 movement assay. Levels not connected by the same letter are significantly different–
461 Tukey-Kramer HSD, Alpha 0.05.

462
463 **Fig. 3. BMAA and MCLR significantly increase acoustic startle sensitivity.** (A)
464 Startle frequency curves for short-latency C-bends (SLCs, left axis, sigmoidal curve fits)
465 and long-latency C-bends (LLCs, right axis, sigmoidal curve fits) in control larvae (black)
466 and treated larvae (1000 μ M BMAA; green). $n = 54$ larvae; mean \pm SEM. (B) and (C),
467 SLC and LLC sensitivity indices, calculated for each larva using the curves as in (A).

468 Levels not connected by the same letter are significantly different–Tukey-Kramer HSD,
469 Alpha 0.05.

470

471 **Fig. 4. BMAA and MCLR interact to enhance startle sensitivity.** (A) Violin plots
472 depict the distribution of the total distance travelled during the 18.5 min spontaneous
473 movement assay for each larva. (B) Violin plot of average speed. (C) Ratio of turns to
474 swims. (D) Startle frequency curves for SLCs (left axis) and LLCs (right axis) in control
475 larvae (0 μ M of cyanotoxin, black) and mixture-treated larvae (100 μ M BMAA plus 1 μ M
476 MCLR, purple). n = 54 siblings; mean \pm SEM. (E,F) SLC and LLC indices for each larva.
477 Levels not connected by the same letter are significantly different–Tukey-Kramer HSD,
478 Alpha 0.05.

479

480 **Fig. 5. BMAA/MCLR mixture increases protein dysregulation *in vivo*.** (A) Number of
481 differentially expressed proteins (DEPs) per treatment condition. (B) Venn diagram
482 showing overlap in DEPs between all treatment groups (red: BMAA protein group, the
483 blue: MCLR protein group, purple: BMAA plus MCLR mixture protein group). (C) Heat
484 map displaying the impacted canonical pathways from IPA functional analysis. The red
485 or blue colored rectangles in each column indicates the z-score activities. Red shading
486 indicates predicted activation and blue shading indicates predicted inhibition according
487 to the scale at right. This heat map represents the z-scores obtained from the
488 comparison between significantly expressed proteins ($P \leq 0.05$).

489

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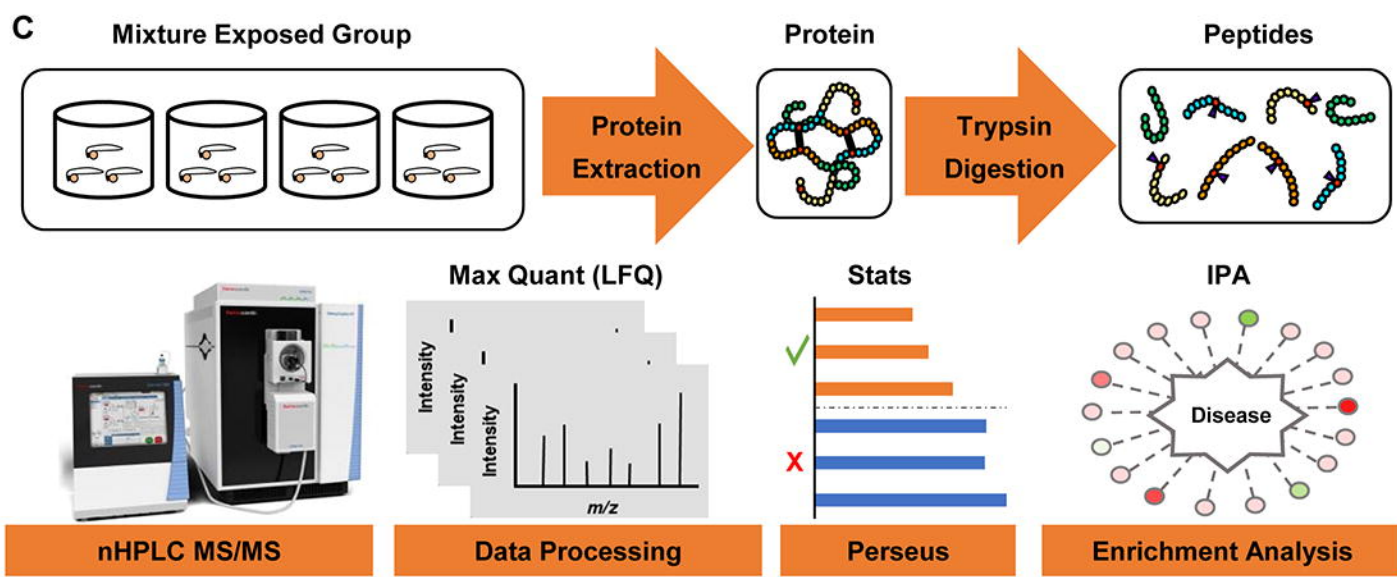
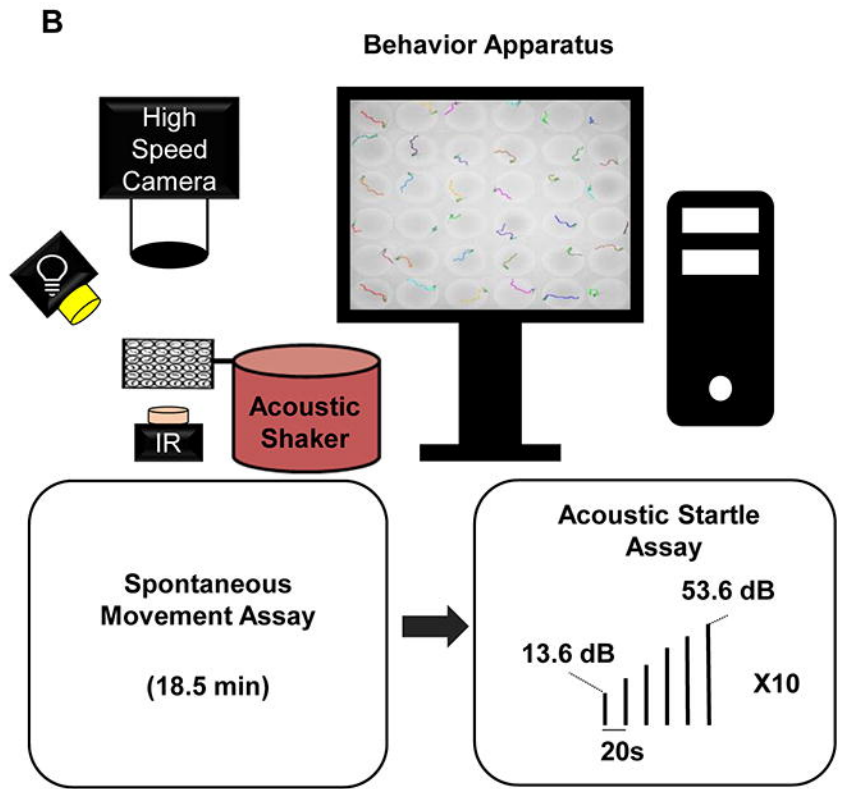
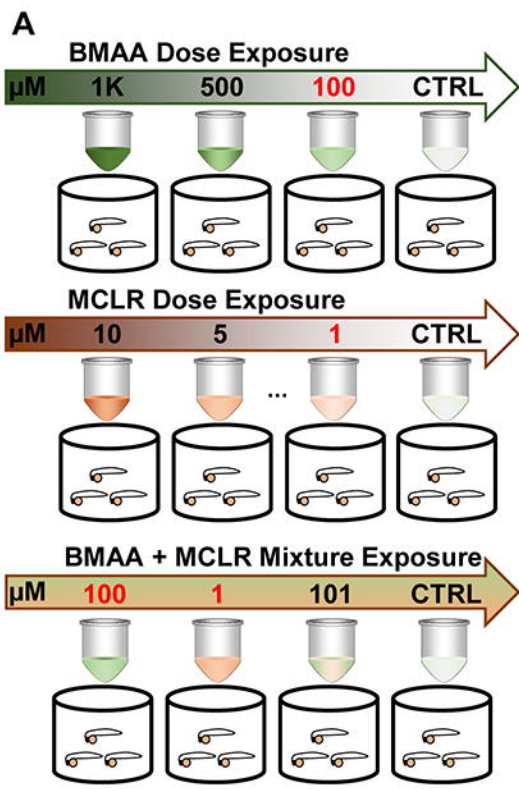
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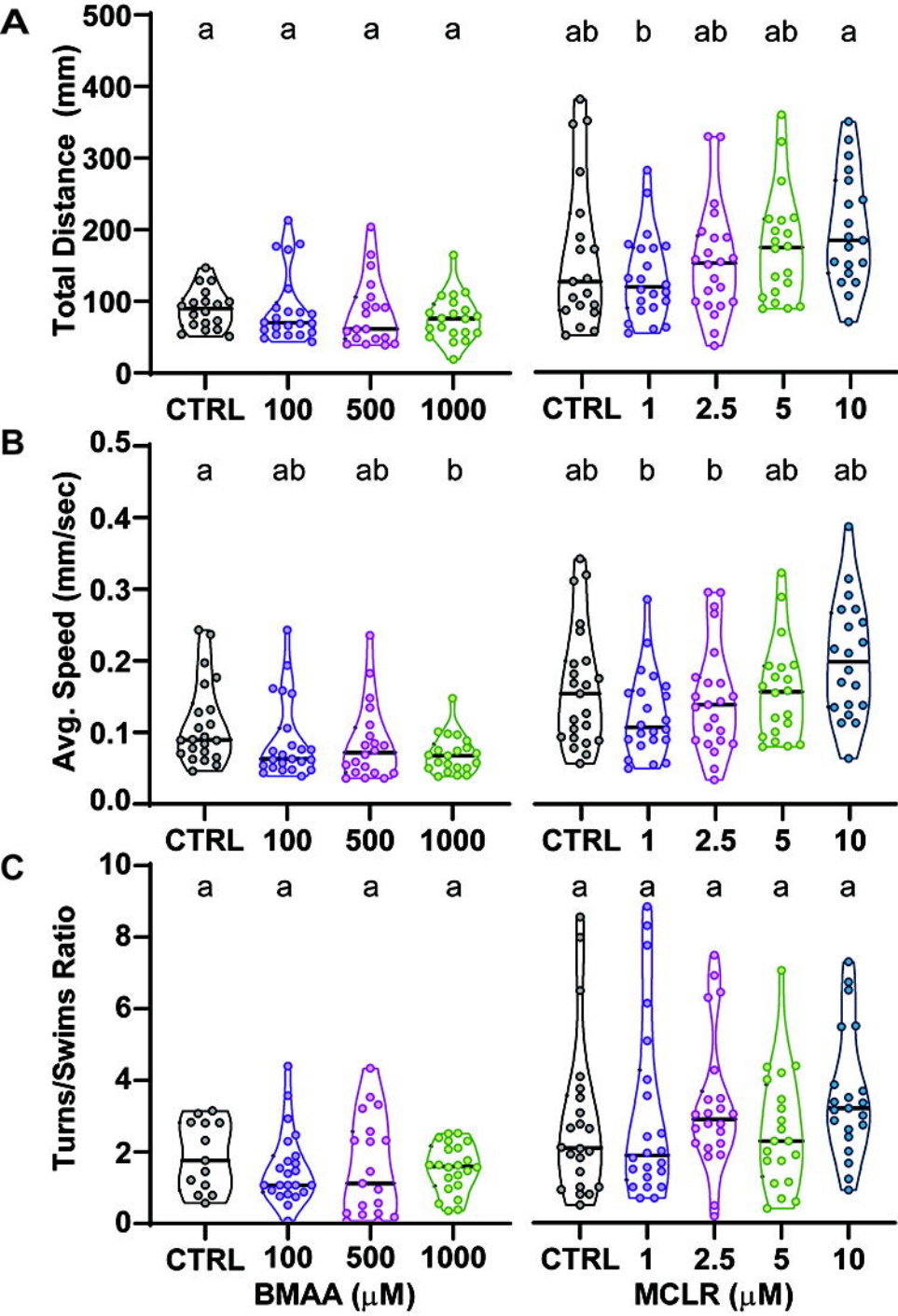
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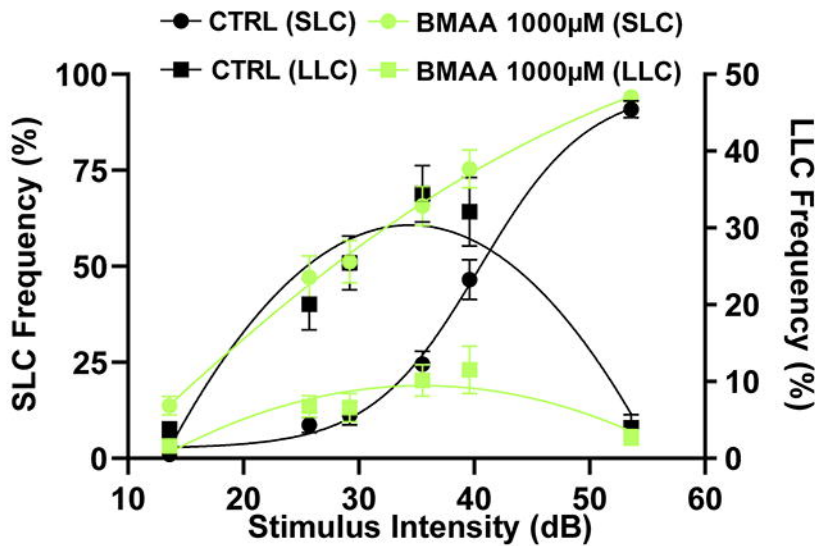
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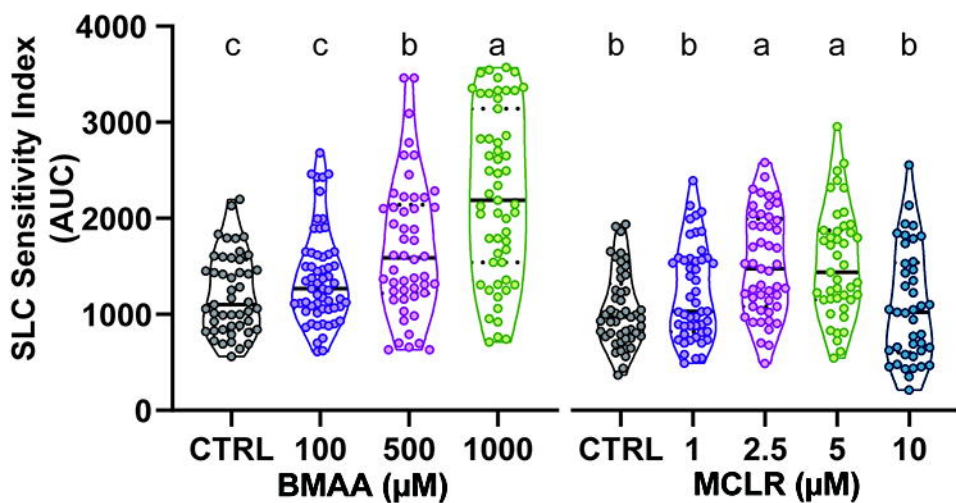




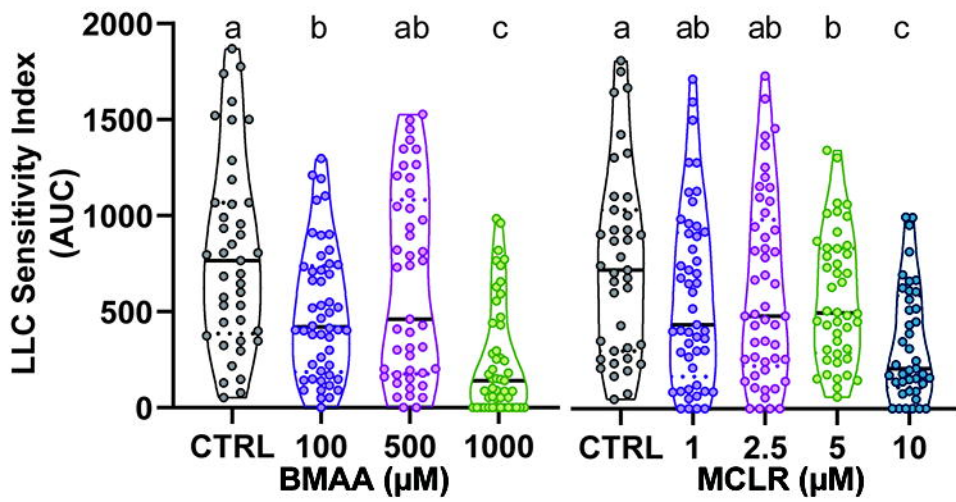
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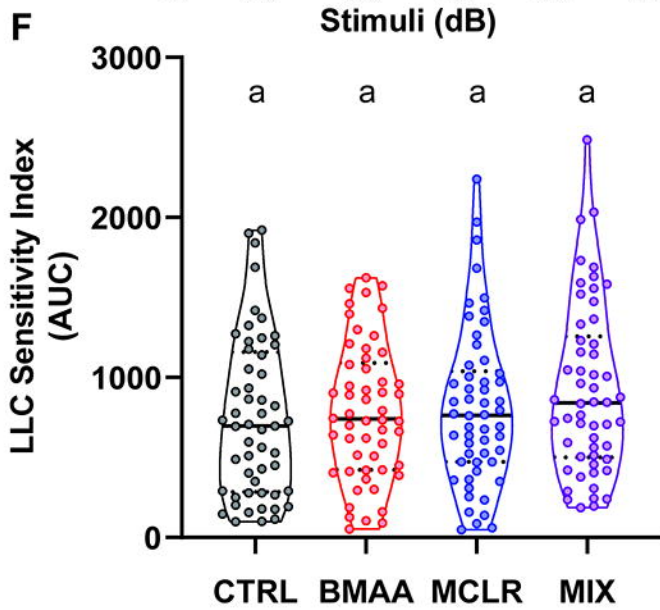
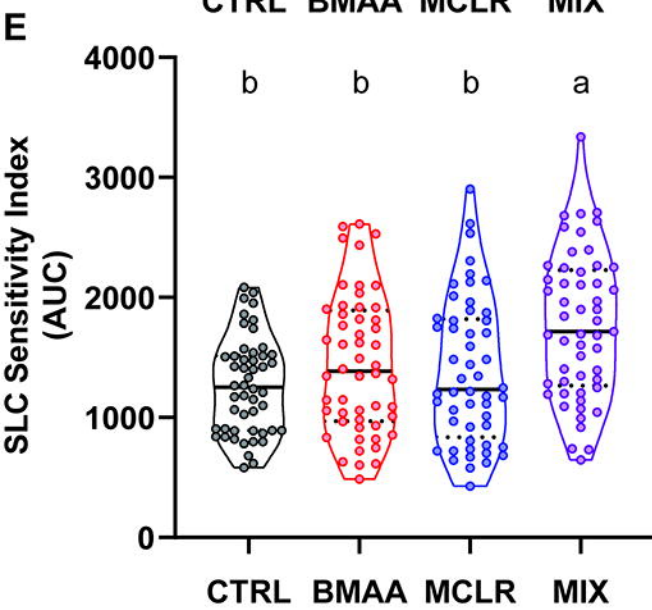
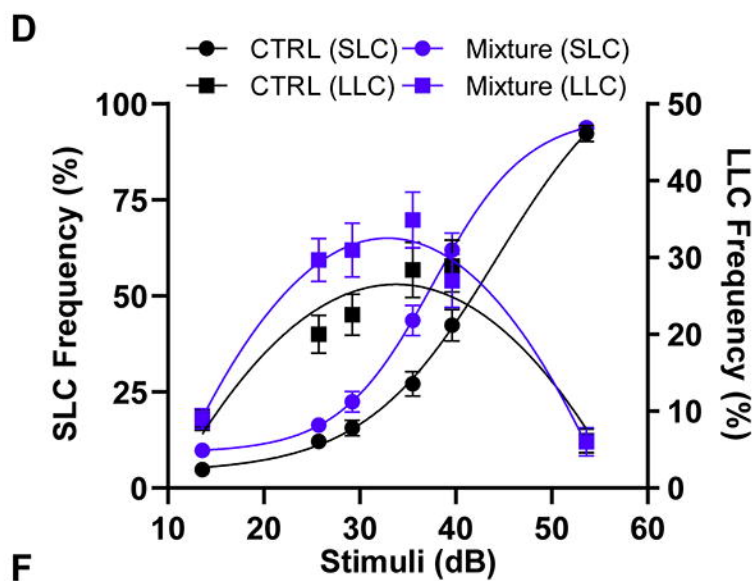
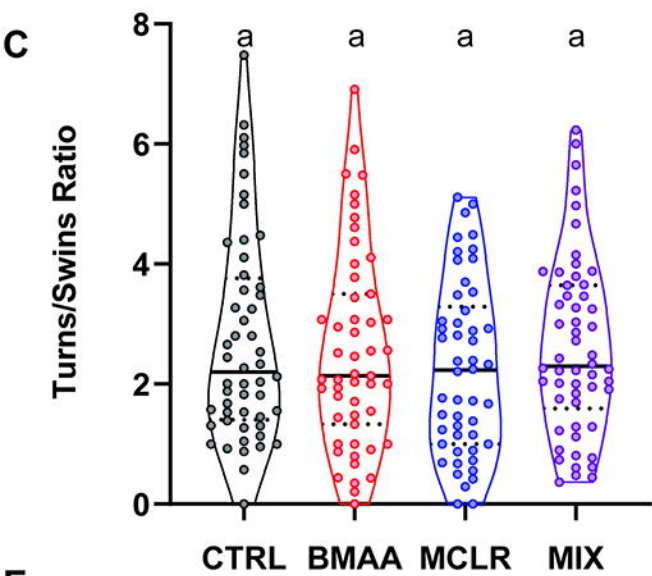
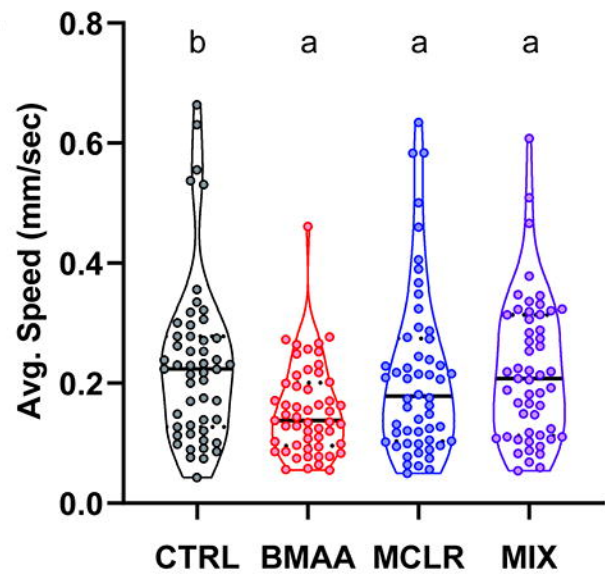
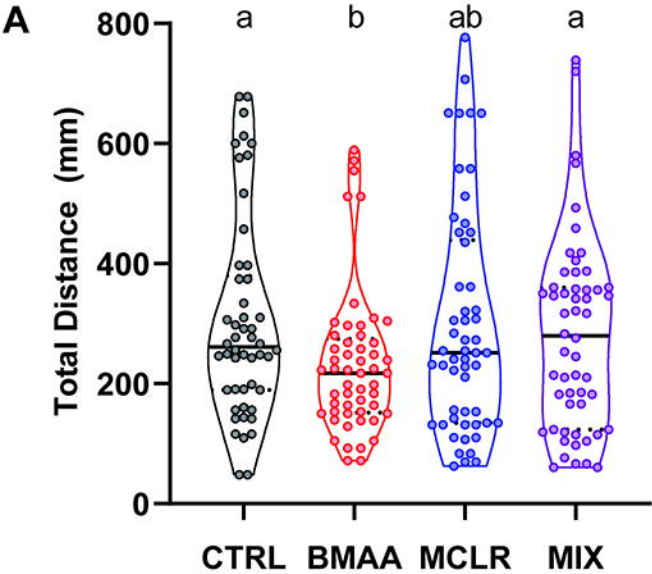


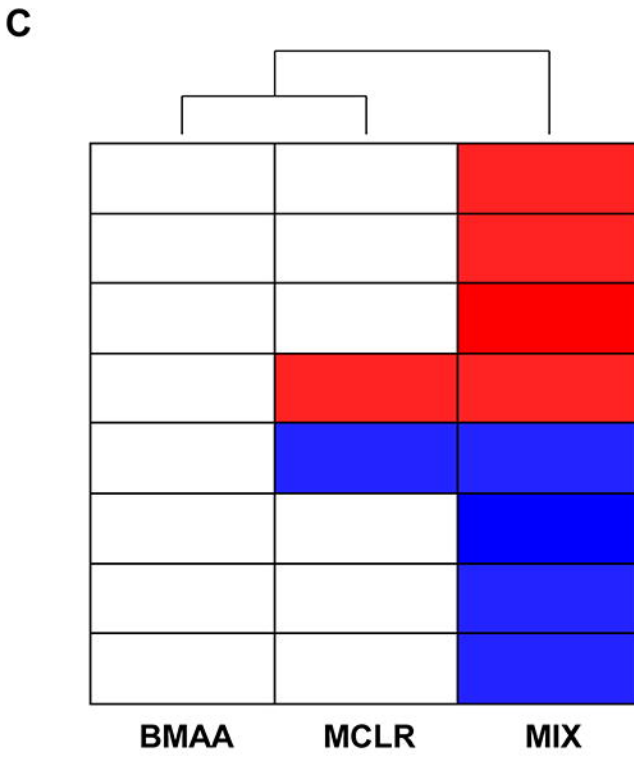
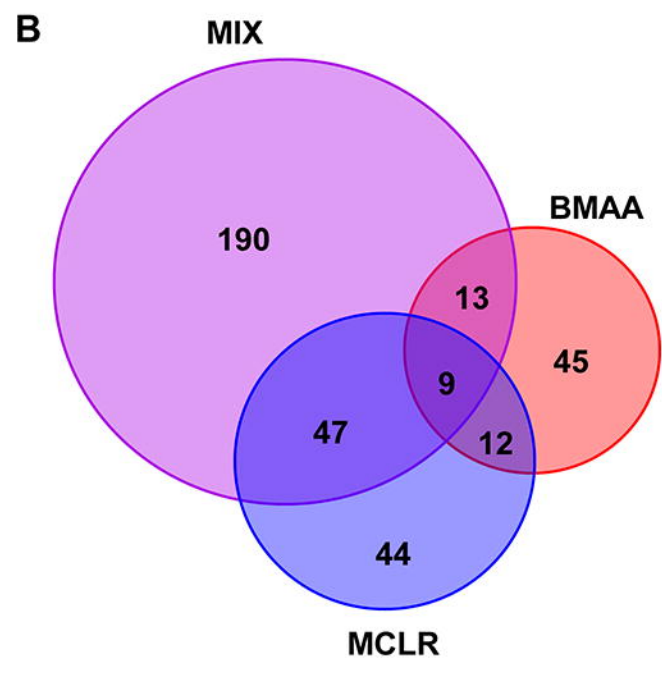
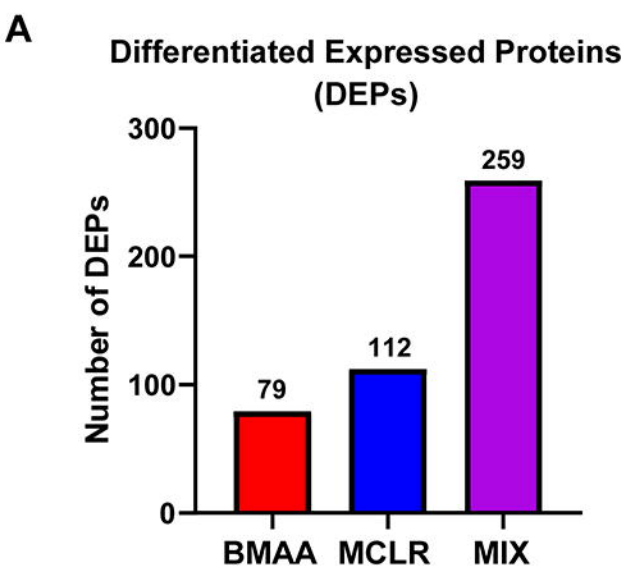
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Canonical Pathways

