

# 1 **Competition for electrons favors N<sub>2</sub>O reduction in denitrifying *Bradyrhizobium*** 2 **isolates**

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## 17 18 **Abstract**

19 The legume-rhizobium symbiosis accounts for the major part of the biological N<sub>2</sub> fixation in  
20 agricultural systems. Despite their lower need for synthetic nitrogen fertilizers, legume-cropped  
21 fields are responsible for substantial N<sub>2</sub>O emissions. Several economically important legumes fix  
22 N<sub>2</sub> through symbiosis with bacteria belonging to the genus *Bradyrhizobium*. Many bradyrhizobia  
23 are also denitrifiers, and inoculation of legumes with N<sub>2</sub>O-reducing strains has been suggested to  
24 mitigate N<sub>2</sub>O emissions. Here, we analyzed the phylogeny and denitrification capacity of  
25 *Bradyrhizobium* strains, most of them isolated from peanut nodules. All performed at least partial  
26 denitrification, but only 37% were complete denitrifiers. This group shared a common phenotype  
27 with a strong preference for N<sub>2</sub>O- over NO<sub>3</sub><sup>-</sup>-reduction. We tested if this could be due to low  
28 quantities of NO<sub>3</sub><sup>-</sup> reductase (periplasmic Nap) but found that Nap was more abundant than N<sub>2</sub>O  
29 reductase (Nos). This corroborates a recently proposed mechanism that the electron pathway  
30 from quinols to Nos via the cytochrome *bc*<sub>1</sub> complex outcompetes that to Nap via the membrane-  
31 bound NapC. We propose that this applies to all organisms which, like many bradyrhizobia, carry  
32 Nos and Nap but lack membrane-bound NO<sub>3</sub><sup>-</sup> reductase (Nar). Supporting this, *Paracoccus*  
33 *denitrificans*, which has Nap and Nar, reduced N<sub>2</sub>O and NO<sub>3</sub><sup>-</sup> simultaneously.

34

## 35 Introduction

36 Emissions of nitrous oxide (N<sub>2</sub>O), the third most important anthropogenic greenhouse gas, have  
37 been rising steadily during the past 150 years. Predictions for the rest of the century show either  
38 continued N<sub>2</sub>O emission increases or rates of decrease slower than the decreases occurring in  
39 CO<sub>2</sub> emissions [1,2]. Agriculture is the main source of anthropogenic N<sub>2</sub>O emissions, with  
40 denitrification playing the major role in most soils [3]. The primary cause of the emissions is  
41 excessive use of synthetic fertilizers [2,4], but cultivation of legumes also contributes  
42 substantially with a global average of 1.29 (range 0.03-7.09) kg N<sub>2</sub>O-N ha<sup>-1</sup> year<sup>-1</sup> while the  
43 corresponding estimate for fertilized agricultural fields was 3.22 (range 0.09-12.67) [5]. In both  
44 cases, the ultimate driver of the emission is input of reactive nitrogen which undergoes a series  
45 of redox reactions and eventually returns to the atmosphere as N<sub>2</sub>O or N<sub>2</sub>. Our aim should be to  
46 minimize the N<sub>2</sub>O/N<sub>2</sub> emission ratio of the system [6]. To achieve this, novel mitigation options  
47 beyond “good agronomic practice” will be needed [7]. Several technical abatement options have  
48 been suggested, including precision fertilization and use of nitrification inhibitors [8].

49 An alternative strategy is to enhance the populations of N<sub>2</sub>O reducing microorganisms in the  
50 soil. Selective *in situ* stimulation of N<sub>2</sub>O reducing, indigenous soil organisms appears unrealistic,  
51 while some options for the addition of N<sub>2</sub>O reducing microbes may be feasible. One is the  
52 suggestion to inoculate soybeans with rhizobia that are not only efficient N<sub>2</sub>-fixers but also have  
53 the capacity to reduce N<sub>2</sub>O [9-11]. At present, N<sub>2</sub>O reduction is not taken into account when  
54 screening strains for their suitability as inoculants and, in fact, several of the commercial  
55 inoculants currently used are only partial denitrifiers [12]. For example, only one of the five  
56 *Bradyrhizobium* strains most commonly used as soybean inoculants in South America could  
57 reduce N<sub>2</sub>O [13]. The growing global need for plant-based proteins is expected to require new  
58 areas to be used for legume cropping and will likely increase the demand for new legume  
59 varieties. Not all combinations of legume cultivar and rhizobial strain result in an efficient  
60 symbiosis [14,15]. Importantly, indigenous soil populations of rhizobia may not be compatible  
61 with a certain legume crop, especially if it has not been grown in that area before. Therefore, the  
62 expansion of legume cropping is creating an increasing demand for a wider variety of inoculants  
63 to ensure nodulation and optimal N<sub>2</sub> fixation efficacy [16]. This offers a golden opportunity to

64 mitigate N<sub>2</sub>O emissions, providing that the inoculants are selected among rhizobial strains that  
65 can reduce N<sub>2</sub>O.

66 The purpose of the present study was to determine how denitrification phenotypes vary  
67 across taxonomically diverse groups of bradyrhizobia, with particular focus on their capacity for  
68 N<sub>2</sub>O reduction, and to understand cellular mechanisms that make these organisms potential sinks  
69 for N<sub>2</sub>O. We determined the phylogenetic position of a set of *Bradyrhizobium* strains, obtained  
70 from different culture collections, and analyzed their capacity for denitrification. Most of the  
71 strains were isolated from nodules of peanut plants (*Arachis hypogaea*) growing in different  
72 regions of China. Peanut is the fourth most important legume crop globally, with China being the  
73 main producer [17]. We also included some bradyrhizobia from other parts of the world, isolated  
74 from peanut and other legumes.

75 Several taxonomic groups within the genus *Bradyrhizobium* have been reported to denitrify  
76 with many strains having truncated denitrification pathways lacking one or more of the steps [12,  
77 18, 19]. A complete denitrification pathway in bradyrhizobia comprises the periplasmic reductase  
78 Nap for dissimilatory NO<sub>3</sub><sup>-</sup> reduction, encoded as a part of the *napEDABC* operon; the Cu-  
79 containing nitrite reductase NirK encoded by *nirK* only or as a part of the *nirKV* operon, although  
80 *cd<sub>1</sub>*-type NirS nitrite reductase has also been reported [20]; the cytochrome *c* dependent nitric  
81 oxide reductase cNor (*norCBQD* operon); and a clade I N<sub>2</sub>O reductase Nos (*nosRZDYFLX* operon)  
82 [21-25]. While the bradyrhizobia along with a number of other denitrifiers, only have the  
83 periplasmic Nap for dissimilatory NO<sub>3</sub><sup>-</sup> reduction, other denitrifiers have the membrane-bound  
84 Nar, or both. Nap is a periplasmic heterodimer (NapAB) consisting of the small electron transfer  
85 subunit NapB and the catalytic NapA, which draws electrons from quinol via the membrane-  
86 bound *c*-type tetrahaeme cytochrome NapC. Nar is a three-subunit enzyme (NarGHI) anchored  
87 in the membrane by NarI, which is a transmembrane protein that draws electrons from quinol  
88 to the catalytic NarG via NarH, both facing the cytoplasm [26, 27]. Nir, Nor and Nos receive  
89 electrons from the *bc<sub>1</sub>* complex via periplasmic small *c*-type cytochromes [28, 29]. Clade I type  
90 Nos reductase may in addition draw electrons from quinol via the membrane bound NosR [30].

91 In a recent study, Mania et al. [19] screened strains within the genus *Bradyrhizobium* for their  
92 denitrification phenotype. Most of them could perform two or more of the denitrification steps,

93 but only half of them could reduce N<sub>2</sub>O. All the N<sub>2</sub>O reducing strains displayed the same strong  
94 preference for N<sub>2</sub>O over NO<sub>3</sub><sup>-</sup> reduction when incubated under anoxic conditions, which was  
95 suggested to be due to competition between the electron pathways to Nap and Nos. The  
96 quantification of denitrification gene transcripts showed 5-8 times lower transcript numbers of  
97 *nap* compared to *nos*, however. Therefore, it could not be excluded that the phenomenon is due  
98 to low abundance of Nap compared to Nos. Here we have investigated the denitrification geno-  
99 and phenotypes for another set of bradyrhizobia and deepened our understanding of the  
100 competition for electrons between Nap and Nos by refined analyses of the electron flow kinetics  
101 in combination with the determination of the relative amounts of Nap and Nos by proteomics.  
102 In addition, we compared the denitrification kinetics of the bradyrhizobia with the denitrification  
103 model bacterium *Paracoccus denitrificans*, which holds a complete denitrification pathway  
104 including both Nap and Nar [26, 29]. The results showed that in all bradyrhizobia with complete  
105 denitrification, NO<sub>3</sub><sup>-</sup> reduction was almost completely hampered in the presence of N<sub>2</sub>O. The  
106 proteomics analysis provided evidence that this was not due to a low abundance of Nap. As  
107 expected, the NarG carrying *P. denitrificans* showed a different denitrification phenotype with  
108 simultaneous reduction of NO<sub>3</sub><sup>-</sup> and N<sub>2</sub>O.

109

## 110 **Materials and methods**

### 111 *Bacterial strains and growth conditions*

112 The rhizobial isolates were obtained from various culture collections. Table 1 shows a list of host  
113 plant and geographic origin of the strains used in the phylogenetic analysis (all listed isolates;  
114 details see [31-37]) and the strains analyzed for their denitrification capacity (all except the four  
115 strains from Ethiopia earlier determined to have a complete denitrification pathway [19]).  
116 Preparation of *Bradyrhizobium* cultures and growth conditions followed Mania et al. [19]. All  
117 cultures were grown at 28 °C in Yeast Mannitol Broth (YMB; [38]) (10 g l<sup>-1</sup> D-mannitol, 0.5 g l<sup>-1</sup>  
118 K<sub>2</sub>HPO<sub>4</sub>, 0.2 g l<sup>-1</sup> MgSO<sub>4</sub>·7H<sub>2</sub>O, 0.1 g l<sup>-1</sup> NaCl and 0.5 g l<sup>-1</sup> yeast extract; pH 7.0). The cultures were  
119 streaked for purity and single colonies were picked and checked for possible contaminating  
120 bacteria by sequencing of the 16S rRNA genes (see below). Single colonies were grown in YMB  
121 and were preserved in 15% (v/v) glycerol at -80 °C. For experiments involving gas kinetics and

122 proteomics, cultures were raised from the frozen stocks in stirred batches of 50 ml YMB  
123 supplemented with filter sterilized  $\text{KNO}_3$  and/or  $\text{KNO}_2$  solutions. The flasks were sealed with  
124 sterilized, gas tight butyl rubber septa (Matrix AS, Norway). Air was removed by applying vacuum  
125 for 6 x 360 s, followed by 30 s He-filling. The desired amount of  $\text{O}_2$  and/or  $\text{N}_2\text{O}$  was injected into  
126 the headspace, and the over-pressure released [39]. Denitrification kinetics of *Bradyrhizobium*  
127 strains with complete denitrification pathways were compared to the model bacterium  
128 *Paracoccus denitrificans* PD1222, which was grown under the same conditions as the  
129 *Bradyrhizobium* strains, except that Sistrom's medium supplemented with 2 mM  $\text{KNO}_3$  was used  
130 [40,41].

### 131 132 *Denitrification end products and gas kinetics analyses*

133 Denitrification end-products were determined for all strains as described previously [42].  
134 Detailed gas kinetics were measured following Mania et al. [19]. Pre-cultures were added to  
135 incubation vials (n=3) supplied with 1 mM initial  $\text{KNO}_3$  and, when indicated, also with 0.5 mM  
136  $\text{KNO}_2$ . Unless stated otherwise, 0.7 mL  $\text{O}_2$  (1 vol% in headspace) and 0.7 ml  $\text{N}_2\text{O}$  (approximately  
137 50  $\mu\text{mol N vial}^{-1}$ ) were added to the headspace at the start of the experiment (0 hpi, hours post  
138 inoculation). For the determination of end-products, the amounts of  $\text{O}_2$ , NO,  $\text{N}_2\text{O}$  and  $\text{N}_2$  in the  
139 headspace were measured at the start and after 10 days of incubation.

140 All strains with a complete denitrification pathway according to the end-point analysis, as  
141 well as all HAMBI strains that had NO as end-product, were investigated further by analyzing gas  
142 kinetics in response to oxygen depletion. To secure an inoculum in which any previously  
143 synthesized denitrification enzymes had been diluted out by aerobic growth, all cultures were  
144 precultured aerobically for at least 6-8 generations. This was done in uncapped vials covered with  
145 aluminum foil (to allow oxygen transport), with vigorous stirring (600 rpm) at 28 °C. To further  
146 ensure the absence of hypoxia, the cultures were transferred to new vials before the  $\text{OD}_{600}$   
147 reached 0.3. Portions of the precultures were inoculated into 120 ml vials containing 50 ml YMB  
148 supplemented with 1, 2 or 3 mM  $\text{KNO}_3$  and with He,  $\text{O}_2$  and  $\text{N}_2\text{O}$  in headspace as described above  
149 for the end-product measurements. The vials were placed in a robotized incubation system that  
150 frequently measured  $\text{O}_2$ , NO,  $\text{N}_2\text{O}$  and  $\text{N}_2$  in headspace [39, 43]. Leakage and sampling losses  
151 were taken into account when calculating the rates of production/consumption of each gas, see

152 Molstad et al [39] for details. Earlier experiments determined that no gas production (NO, N<sub>2</sub>O  
153 or N<sub>2</sub>) took place from chemical reactions between of NO<sub>2</sub><sup>-</sup> and the medium under the conditions  
154 used [19].

155

#### 156 *Nitrite and growth measurements*

157 Samples for NO<sub>2</sub><sup>-</sup> and growth measurements (0.1-1.0 ml) were taken manually from the liquid  
158 phase through the rubber septum of the vials using a sterile syringe. NO<sub>2</sub><sup>-</sup> concentrations were  
159 determined immediately after each sampling by drawing a 10 µL liquid sample which was injected  
160 into a purging device containing 50% acetic acid with 1% (w/v) NaI (at room temperature), by  
161 which NO<sub>2</sub><sup>-</sup> is reduced instantaneously to NO [44, 45]. The quantities of resultant NO were  
162 measured by chemiluminescence using a NO analyser (Sievers NOA™ 280i, GE Analytical  
163 Instruments). The system was calibrated by injecting standards. Sampling for NO<sub>2</sub><sup>-</sup> measurements  
164 was done as frequently as the gas sampling from the

165 *Bradyrhizobium* and *P. denitrificans* cultures.

166 Aerobic and anaerobic growth was investigated in detail for *Bradyrhizobium* sp. HAMBI 2125  
167 to determine the growth yield per mol electrons transferred to O<sub>2</sub>, N<sub>2</sub>O, and to the entire  
168 denitrification pathway. Four treatments were compared: 1) 7% O<sub>2</sub> in headspace; 2) 0% O<sub>2</sub>, 4 mM  
169 NO<sub>3</sub><sup>-</sup>; 3) 0% O<sub>2</sub> and 10 ml N<sub>2</sub>O; 4) 7% O<sub>2</sub>, 4 mM NO<sub>3</sub><sup>-</sup> and 10 ml N<sub>2</sub>O. Pre-cultures were started  
170 from single colonies and grown in YMB under fully oxic conditions (21% O<sub>2</sub>) for inoculation of the  
171 aerobic batch cultures (treatments 1 and 4), and under anoxic conditions for inoculation of the  
172 anaerobic batches (treatments 2 and 3). The OD<sub>600</sub> of the aerobic pre-cultures was kept < 0.3 by  
173 regularly inoculating portions of the culture into vials with fresh medium. There were 6 replicate  
174 vials for each treatment. Liquid samples were taken from three of them every 6 h (0.85 ml) for  
175 growth measurements (OD<sub>600</sub>). The three other replicates were only monitored for gas kinetics  
176 and were then used to measure final cell dry weight, to assess the growth yield per mol electrons  
177 to O<sub>2</sub> and N-oxides. To determine cell dry weight, entire culture volumes were centrifuged,  
178 washed twice in distilled water (resuspension and centrifugation), dried at 100 °C in Eppendorf  
179 tubes, and weighed. Cell numbers were determined by microscopic counts at selected sampling  
180 times, which gave a conversion factor of 5.8\*10<sup>8</sup> cells ml<sup>-1</sup> \* OD<sub>600</sub><sup>-1</sup>. The estimated yield from

181 growth on O<sub>2</sub>, N<sub>2</sub>O and NO<sub>3</sub><sup>-</sup> was 12.0, 8.5 and 6.7 E12 cells mol<sup>-1</sup> electrons, respectively (Table  
182 S1). The growth rates were 0.063, 0.031 and 0.024 h<sup>-1</sup> for cells grown with O<sub>2</sub>, N<sub>2</sub>O and NO<sub>3</sub><sup>-</sup>  
183 respectively (Fig. S1). The yields were used to calculate cell density throughout the incubations,  
184 enabling the calculation of cell specific rates of electron flows to the various electron acceptors,  
185 for details see Fig. S2 and Table S1.

186

### 187 *DNA extraction and gene sequence analyses*

188 Cultures were started from single colonies and incubated in YMB medium under oxic conditions  
189 at 28 °C for five days. DNA was extracted using QIAamp DNA Mini Kit (Qiagen). Purity of the  
190 strains was checked by Sanger sequencing of 16S rRNA amplicons (for primers and PCR conditions  
191 see Table S2). The genome sequences were obtained from another ongoing project: (JGI Gold  
192 study ID: Gs0134353), and their Genome accession numbers are listed in Table S3. Genome  
193 sequences were analyzed using RAST (Rapid Annotations using Subsystems Technology). Whole-  
194 genome average nucleotide identity (ANI) comparisons were done using FastANI v.0.1.2.

195 A phylogenetic tree that included all HAMBI and CCBAU strains from this study (Table 1), as  
196 well as the Ethiopian strains AC70c, AC86d2, AC87j1 and AC101b and the type strain of all species  
197 of the genus *Bradyrhizobium* (February 2020, <http://www.bacterio.net/bradyrhizobium.html>),  
198 was constructed based on multilocus sequence analysis (MLSA) of the four housekeeping genes  
199 *atpD* (ATP synthase F1, beta subunit), *glnII* (glutamine synthetase type II), *recA* (recombinase  
200 A), *rpoB* (RNA polymerase, beta subunit). The sequences from all HAMBI strains used in the MLSA  
201 analysis were obtained from the present study. Amplicons of *rpoB* and *atpD* of the Ethiopian  
202 strains AC70c, AC86d2 and AC101b and *recA* from AC70c were Sanger sequenced. For primers  
203 and PCR conditions, see Table S2. All other sequences were obtained from GenBank  
204 (<http://www.ncbi.nlm.nih.gov/genbank>). The GeneBank accession numbers for all housekeeping  
205 genes used are listed in Table S5. The gene sequences were aligned by ClustalW [46] in Bioedit  
206 v7.0.5.3 [47], and the alignments were edited manually. The best-fit models of nucleotide  
207 substitution were selected based on Akaike information criterion (AIC) applying MEGA7 [48].  
208 The Bayesian phylogenetic analyses of the concatenated loci (*atpD-recA-glnII-rpoB*) of 78  
209 *Bradyrhizobium* strains were performed using the algorithm Metropolis coupled Markov chain

210 Monte Carlo (MCMCMC) twice for ten million generations using MrBayes v3.2.7a program [49].  
211 *Neorhizobium galegae* was used to root the tree. The general time reversible plus gamma  
212 distribution plus invariable sites (GTR + G + I) was chosen as the best-fit model of phylogenetic  
213 analyses of the four partition of the concatenated dataset (*atpD-glnII-recA-rpoB*). The algorithms  
214 were analyzed using Tracer v1.7 [50], and the phylogenetic trees were visualized by Figtree v1.4.3  
215 (<https://beast.community/figtree>). For comparison with the MLSA tree based on four  
216 housekeeping genes, a phylogenetic tree of all the HAMBI test strains listed in Table 2 was  
217 constructed using the Genome Taxonomy Database Toolkit (GTDB-Tk) v0.3.2, which is based on  
218 a phylogeny inferred from the concatenation of 120 ubiquitous, single-copy proteins [51].

219

## 220 *Proteomics*

221 Relative quantification of the denitrification reductases was done for the strain *Bradyrhizobium*  
222 sp. HAMBI 2125, chosen as a representative for the complete denitrifiers (Fig. 1), which all  
223 showed the same preference for N<sub>2</sub>O over NO<sub>3</sub><sup>-</sup> reduction (Fig. 2A, S3, S4, S5). Cells were first  
224 raised from frozen stock under strictly aerobic conditions as explained above. As the aerobic  
225 preculture reached OD<sub>600</sub> = 0.05 (2.9\*10<sup>7</sup> cells ml<sup>-1</sup>), it was used to inoculate 18 vials (1 ml  
226 inoculum vial<sup>-1</sup>) containing YMB with 1 mM NO<sub>3</sub><sup>-</sup> and with 1 ml N<sub>2</sub>O (approximately 80 μmol N<sub>2</sub>O-  
227 N flask<sup>-1</sup>) and 1 % O<sub>2</sub> in headspace. Samples for protein analysis were taken i) from the aerobic  
228 inoculum; ii) just before transition to anaerobic respiration (22 hpi, when O<sub>2</sub> reached ~4.6 μM);  
229 iii) during the period of rapid reduction of exogenous N<sub>2</sub>O (27 and 32 hpi); and iv) during the  
230 period of NO<sub>3</sub><sup>-</sup> reduction to N<sub>2</sub> (36 hpi). Three replicate vials were harvested at each time point.  
231 Gas kinetics, NO<sub>2</sub><sup>-</sup> concentrations and sampling times for proteomics are shown in Fig. S6. After  
232 measuring the OD<sub>600</sub>, the samples were centrifuged (10 000 g, 4 °C, 10 min) and the cell pellets  
233 were stored at -20 °C for protein extraction. In addition, we harvested and froze cells from the  
234 aerobic preculture.

235 To extract the proteins, the cell pellets were thawed, resuspended in lysis buffer (20 mM Tris-  
236 HCl pH8, 0.1 % v/v Triton X-100, 200 mM NaCl, 1 mM DTT) and treated with 3x45 s bead beating  
237 with glass beads (particle size ≤ 106 μm Sigma) at maximum power and cooling on ice between  
238 the cycles (MP Biomedicals™ FastPrep- 24™, Thermo Fischer Scientific Inc). Cell debris was



239 removed by centrifugation (10 000 g; 5 min) and the supernatant, containing water soluble  
240 proteins, was used for proteomic analysis based on Orbitrap-MS as described before [52]. To  
241 quantify the denitrification reductases we calculated the fraction of the individual enzymes as  
242 percentages of the sum of all LFQ (label-free quantitation of proteins) abundances for each time  
243 point.

244

## 245 **Results**

### 246 *Phylogenetic distribution of denitrifying bradyrhizobia*

247 The phylogenetic position of the 23 test strains was determined by MLSA analysis of four protein-  
248 coding housekeeping genes (*atpD-glnII-recA-rpoB*) (Fig. 1). The analysis comprised 78  
249 bradyrhizobial strains and included all hitherto described *Bradyrhizobium* reference species. All  
250 but one of the 23 test strains listed in Table 1 were assigned to the *Bradyrhizobium japonicum*  
251 supergroup (Fig. 1). The exception was *Bradyrhizobium lablabi* HAMBI 3052<sup>T</sup>, which is a member  
252 of the *Bradyrhizobium jicamae* supergroup according to the new genome-based taxonomy of  
253 bradyrhizobia proposed by Avontuur et al. [53]. The test strains were placed in 11 lineages (clades  
254 A-K) with high posterior probability (PP=1.00). Most test strains were clustered within the clades  
255 of seven diverse bradyrhizobial species. Strains *Bradyrhizobium* sp. HAMBI 2134 and 2142 formed  
256 a monophyletic cluster (clade G) with high posterior probability (1.00). *Bradyrhizobium* sp. strain  
257 HAMBI 2116 was accommodated in a lineage (clade A) distant from other bradyrhizobia. The  
258 phylogenetic relationships revealed by the four-gene MLSA tree (Fig. 1) were supported by the  
259 results from an analysis of the HAMBI strains (listed in Table 2), based on the GTDB Tool kit (Fig.  
260 S7). In both trees strains HAMBI 2299, 2153 and 2116 were most closely related to the reference  
261 species *B. yuanmingense*; strains HAMBI 2125 (and 2126) and 2127 were closest to *B. ottawaense*;  
262 strain HAMBI 2130 was closest to *B. shewense*; strains HAMBI 2115 was closest to *B. japonicum*  
263 USDA6; strains HAMBI 2128 (as well as HAMBI 2129 and 2150, Table S4) were closest to *B.*  
264 *arachidis*; strain HAMBI 3052 was closest to *B. lablabi*. Strains HAMBI 2149, 2133, 2136, 2137,  
265 2151 and 2152 formed one cluster in the GTDB tree together with *B. japonicum* USDA 135. This  
266 cluster was close to *B. liaoningense*, as also found in the MLSA analysis (Fig. 1), which did not

267 include *B. japonicum* USDA 135. Strains HAMBI 2134 (and 2135, Table S4) and 2142 formed a  
268 separate cluster in both trees.

269 We originally included 21 isolates from the HAMBI culture collection in our study, all  
270 deposited with different strain names. Pairwise whole genome comparisons revealed that some  
271 strains were very similar with ANI values >99.2% and in most cases >99.9 % (Table S4), which was  
272 also supported by the clustering of similar strains in the GTDB tree (Fig. S7). Since the strains  
273 within such clusters with ANI values >99% also showed practically identical denitrification  
274 phenotypes, we present the result for only one of the strains in each such cluster in the main  
275 paper, and the results for the others as supplementary information.

276 The denitrification end-product analysis showed that all test strains were denitrifiers, able to  
277 reduce  $\text{NO}_3^-$  and/or  $\text{NO}_2^-$  to  $\text{NO}$ ,  $\text{N}_2\text{O}$  or  $\text{N}_2$ . Seven of the 19 strains tested for denitrification in the  
278 present study were complete denitrifiers that reduced  $\text{NO}_3^-$  to  $\text{N}_2$ . The four Ethiopian strains  
279 AC87j1, AC86d2, AC101b and AC70c were determined earlier to be complete denitrifiers [19]. All  
280 complete denitrifiers except *B. lablabi* (clade K) belonged to the *B. japonicum* supergroup and  
281 were positioned in clades E, H, I and J. The clades E, H and J represented the species *B.*  
282 *guangxiense*, *B. ottawaense* and *B. shewense*, respectively. Clade I, accommodating strains  
283 *Bradyrhizobium* sp. HAMBI 2125 and AC70c, formed a separate, monophyletic group. The other  
284 test strains had truncated denitrification pathways and were able to reduce  $\text{NO}_3^-$  and  $\text{NO}_2^-$  to  $\text{NO}$   
285 (5 strains) or  $\text{N}_2\text{O}$  (7 strains). The seven strains having  $\text{N}_2\text{O}$  as end-product were distributed in  
286 clades A, C, D and G. Three strains grouped with *B. arachidis* in clade C. *Bradyrhizobium* sp. strain  
287 HAMBI 2115 was placed in clade D along with *B. japonicum* USDA6<sup>T</sup>. Three strains were  
288 accommodated in clades A and G, neither of which included any of the hitherto described  
289 *Bradyrhizobium* species. The strains with  $\text{NO}$  as denitrification end-product were found in clades  
290 B, C and F, which also harbored the species *B. yuanmingense*, *B. arachidis* and *B. liaoningense*,  
291 respectively. Strains HAMBI 2153 and 2149, as well as the strains highly similar to HAMBI 2149  
292 (Table S4), were also examined using detailed gas kinetics to further examine whether they could  
293 reduce  $\text{NO}$  (Fig. S8). Like in the end-product investigation they produced large amounts of  $\text{NO}$   
294 (14-15  $\mu\text{mol vial}^{-1}$ ) while no reduction of  $\text{NO}$  or  $\text{N}_2\text{O}$  was detected (Fig. S8).

295

296 *Consistency between denitrification gene operons and end products*

297 The genetic potential for denitrification, based on whole genome sequencing, and the  
298 corresponding phenotype, determined by end-point analysis and gas kinetics, were compared  
299 for 13 strains (Table 2), as well as for isolates with genomes with ANI values >99.2% to these  
300 strains (Table S4). The five strains able to reduce  $\text{NO}_3^-$  to  $\text{N}_2$  had complete operons for all four  
301 reduction steps. For most of the other strains, with  $\text{NO}$  or  $\text{N}_2\text{O}$  as denitrification end -products,  
302 the lack of a function was explained by complete lack of the corresponding operon. However,  
303 two strains (HAMBI 2153 and HAMBI 2299) showed a discrepancy between denitrification  
304 genotype and phenotype. They were apparently unable to reduce  $\text{NO}$  and accumulated  $\mu\text{M}$   
305 concentrations of  $\text{NO}$ , despite having the *nor* operon. Comparison of these strains with strains  
306 having functional  $\text{NO}$  reduction revealed that strain HAMBI 2153 had a frame shift in *norB*, while  
307 *norQ* was not in the correct reading frame. Strain HAMBI 2299 carried *nap*, *nir* and *nor* operons  
308 with all expected genes present. The end-point analysis of this strain showed that after 10 days  
309 of incubation with 1 mM  $\text{NO}_3^-$ , 0.5 mM  $\text{NO}_2^-$  and 1 ml  $\text{N}_2\text{O}$  in headspace, it had produced toxic  
310 levels (>10  $\mu\text{M}$ ) of  $\text{NO}$ . Inspection of the amino acid sequences did not reveal any obvious  
311 explanations for lack of Nor activity.

312

313 *All  $\text{N}_2\text{O}$ -reducing Bradyrhizobium strains preferred  $\text{N}_2\text{O}$  over  $\text{NO}_3^-$  as electron acceptor*

314 To determine if the complete denitrifiers identified in this study showed a similar, strong  
315 preference for  $\text{N}_2\text{O}$  over  $\text{NO}_3^-$  as reported by Mania et al. [19] for other *Bradyrhizobium* strains,  
316 detailed gas kinetics was analyzed. The results showed remarkably similar kinetics for all strains  
317 during and after the transition from aerobic to anaerobic respiration in vials to which  $\text{N}_2\text{O}$  had  
318 been added to the headspace (Figs. 2A, S3, S4, S5). Strain HAMBI 2125 was chosen as a  
319 representative for the complete denitrifiers. The  $\text{NO}_2^-$  concentration in cultures of this strain was  
320 measured along with the gas kinetics (Figs. 2A, S3), allowing the estimation of  $\text{NO}_3^-$   
321 concentrations by N-mass balance.

322 The first phase, marked in blue, was dominated by  $\text{O}_2$  respiration. When the  $\text{O}_2$  concentration  
323 in the medium reached 4.6  $\mu\text{M}$ , some  $\text{NO}_3^-$  reduction took place, measured as a slight  
324 accumulation of  $\text{NO}_2^-$  and  $\text{NO}$  (0.7 and 1.7  $\mu\text{mol}$ ; not clearly visible in the figure due to scaling).

325 The small but significant decrease in N<sub>2</sub>O during the aerobic phase was caused by sampling losses,  
326 not by N<sub>2</sub>O reduction or other reactions, as explained in Fig. S9. The same is seen for experiments  
327 presented in Figs. 2C and S3-S5. Reduction of exogenous N<sub>2</sub>O started when the O<sub>2</sub> concentration  
328 in the liquid was approximately 3.0 μM and rapidly reached a rate of 10.4 ± 1.2 μmol N flask<sup>-1</sup> h<sup>-1</sup>  
329 <sup>1</sup>, matching the corresponding N<sub>2</sub> production rate of 9.7 ± 1.0 μmol N flask<sup>-1</sup> h<sup>-1</sup>. This phase of  
330 rapid and stoichiometric reduction of N<sub>2</sub>O to N<sub>2</sub>, marked in pink, lasted until most of the  
331 exogenous N<sub>2</sub>O was reduced. No NO<sub>3</sub><sup>-</sup>-reduction took place during reduction of the exogenous  
332 N<sub>2</sub>O. When the exogenous N<sub>2</sub>O was depleted, the rate of NO<sub>3</sub><sup>-</sup> reduction increased and the rate  
333 of N<sub>2</sub> production decreased instantaneously to 2.22 ± 0.18 μmol flask<sup>-1</sup> h<sup>-1</sup> (phase marked in  
334 green). When NO<sub>3</sub><sup>-</sup> reduction started, the NO<sub>2</sub><sup>-</sup> concentration increased sharply and reached  
335 approximately 200-350 μM (9.8-17.4 μmol vial<sup>-1</sup>), accounting for 20-35% of the initial 50 μmol  
336 NO<sub>3</sub><sup>-</sup>-N. The N<sub>2</sub> curve for strain HAMBI 2125 reached a stable plateau at 123 μmol N<sub>2</sub>-N, which  
337 accounts for the sum of N<sub>2</sub>O- and NO<sub>3</sub>-N present initially. The same experiment was conducted  
338 with other strains, but without measuring NO<sub>2</sub><sup>-</sup> (Figs. S4, S5), and they all showed the same fast  
339 stoichiometric conversion of N<sub>2</sub>O to N<sub>2</sub> and a subsequent slower production of N<sub>2</sub> from NO<sub>3</sub><sup>-</sup>,  
340 lasting until the N<sub>2</sub>O+NO<sub>3</sub><sup>-</sup>-N was completely recovered as N<sub>2</sub>-N. Strain HAMBI 2125 showed good  
341 control of NO during the entire incubation period, reaching a small peak of 11 nmol vial<sup>-1</sup> (6.9 nM  
342 in the liquid) around 41 hpi. For the other strains NO concentrations never exceeded 50 nM, and  
343 most strains kept NO concentrations <20 nM during the entire incubation (Figs. 2A, S3, S4, S5).

344

345 *Nar* carrying *Paracoccus denitrificans* showed no preference for N<sub>2</sub>O over NO<sub>3</sub><sup>-</sup>

346 In accordance with our hypothesis, the *Nar* carrying *P. denitrificans* reduced NO<sub>3</sub><sup>-</sup> simultaneously  
347 with the exogenously supplied N<sub>2</sub>O, indicating no hampering of NO<sub>3</sub><sup>-</sup> reduction by N<sub>2</sub>O (Fig. 2C).  
348 In contrast to all the examined *Bradyrhizobium* strains (Figs. 2A, S3, S4, S5), *P. denitrificans*  
349 started to reduce NO<sub>3</sub><sup>-</sup> and headspace N<sub>2</sub>O concomitantly when the O<sub>2</sub> concentration was close  
350 to depletion and stoichiometrically accumulated NO<sub>2</sub><sup>-</sup> from NO<sub>3</sub><sup>-</sup> reduction. Nitrite reduction and  
351 emergence of NO started when most of the NO<sub>3</sub><sup>-</sup> had been reduced to NO<sub>2</sub><sup>-</sup>.

352

353 *Electron flow rates*

354 The rates of electron flow to the various reductases ( $\text{fmol e}^- \text{ cell}^{-1} \text{ h}^{-1}$ ) were calculated based on  
355 the measured gas- and  $\text{NO}_2^-$  kinetics. Since the growth yield per mol electron (to  $\text{O}_2$  and nitrogen-  
356 oxides) was determined (Fig. S2 and Table S1), the cell density could be calculated for each time  
357 point throughout the incubation, based on the measured cumulated electron flow, thus enabling  
358 an estimation of cell specific electron flows throughout. In *Bradyrhizobium* strain HAMBI 2125  
359 (Fig. 2B), the cell specific electron flow was sustained at a more or less constant level ( $5\text{-}6 \text{ fmol e}^-$   
360  $\text{ cell}^{-1} \text{ h}^{-1}$ ) through the transition from respiring  $\text{O}_2$  to respiring  $\text{N}_2\text{O}$ , until the external  $\text{N}_2\text{O}$  was  
361 depleted (phases marked blue and pink in the corresponding gas kinetics Fig. 2A), but declined  
362 to  $\sim 3\text{-}4 \text{ fmol e}^- \text{ cell}^{-1} \text{ h}^{-1}$  when depending on the activity of Nap. In *P. denitrificans* (Fig. 2D) there  
363 was instead a simultaneous electron delivery to Nar and Nos during the phase when exogenous  
364  $\text{N}_2\text{O}$  was being reduced (marked pink in the corresponding gas kinetics, Fig. 2C). This was later  
365 followed by electron flow to Nir, Nor and Nos, after the depletion of nitrate. The cell specific  
366 respiratory metabolism was  $10\text{-}13 \text{ fmol e}^- \text{ cell}^{-1} \text{ h}^{-1}$  during the phase dominated by aerobic  
367 respiration, then increased to approximately  $15\text{-}16 \text{ fmol e}^- \text{ cell}^{-1} \text{ h}^{-1}$  when respiring  $\text{NO}_3^-$  and  $\text{N}_2\text{O}$ ,  
368 followed by a decrease to approximately  $8 \text{ fmol e}^- \text{ cell}^{-1} \text{ h}^{-1}$  during respiration of the accumulated  
369  $\text{NO}_2^-$ .

370  
371 *Injection of  $\text{N}_2\text{O}$  to actively denitrifying cultures resulted in an immediate arrest of  $\text{NO}_3^-$ -reduction*  
372 To investigate whether the preferred electron flow to  $\text{N}_2\text{O}$  in the foregoing experiments could be  
373 explained by regulation at the transcriptional level (i.e. early expression of *nos*),  $\text{N}_2\text{O}$  was injected  
374 into a culture that was actively reducing  $\text{NO}_3^-$  to  $\text{N}_2$ . A transcriptional regulation would not result  
375 in an immediate switch to the exogenous  $\text{N}_2\text{O}$  as electron acceptor but would instead be seen as  
376 a simultaneous reduction of all N-oxides present. This was tested on strain HAMBI 2125 (Fig. 3).  
377 The cultures were incubated in vials containing YMB with 2 mM initial  $\text{NO}_3^-$ , and with 1%  $\text{O}_2$  but  
378 without exogenous  $\text{N}_2\text{O}$  in the headspace. Denitrification (from  $\text{NO}_3^-$  reduction) commenced  
379 when  $\text{O}_2$  reached  $4.3 \mu\text{M}$  and proceeded at a rate of  $2.6 \pm 0.4 \mu\text{mol N flask}^{-1} \text{ h}^{-1}$ . At 32 hpi, when  
380 the cells were actively reducing  $\text{NO}_3^-$  to  $\text{N}_2$ , approximately  $300 \mu\text{mol N}_2\text{O-N}$  was injected. This  
381 injection led to a complete suppression of  $\text{NO}_3^-$  reduction. The measured  $\text{N}_2$  production rate was  
382 similar to the measured  $\text{N}_2\text{O}$  reduction rate ( $23.9 \pm 3.6$  and  $26.4 \pm 3.6 \mu\text{mol N flask}^{-1} \text{ h}^{-1}$ ,

383 respectively). NO accumulation during this phase declined rapidly to zero consistent with an  
384 absence of  $\text{NO}_3^-$  reduction. The  $\text{NO}_3^-$  reduction recovered once the external  $\text{N}_2\text{O}$  was depleted,  
385 as reflected by recovery of NO, and an  $\text{N}_2$  production rate similar as that before the injection of  
386  $\text{N}_2\text{O}$ .

387 In an additional experiment, including all seven  $\text{N}_2\text{O}$ -reducing strains, exogenous  $\text{N}_2\text{O}$  was  
388 present in the headspace from the start of the incubation (Fig. S5). As expected from the first  
389 experiment (Fig. 2A), the exogenous  $\text{N}_2\text{O}$  was rapidly reduced, followed by slow production of  $\text{N}_2$   
390 from  $\text{NO}_3^-$  reduction. A second injection of  $\text{N}_2\text{O}$  into the actively denitrifying cultures resulted in  
391 an immediate increase in the  $\text{N}_2$  production rate, similar to that during reduction of the initial  
392  $\text{N}_2\text{O}$ . These results supported the notion that  $\text{N}_2\text{O}$  was preferred as electron acceptor over  $\text{NO}_3^-$ .

393

394 *The preference for  $\text{N}_2\text{O}$  over  $\text{NO}_3^-$  was not explained by the relative numbers of the different*  
395 *reductases*

396 A proteomics analysis was done for strain HAMB1 2125. The abundance of NapA, NirK and NosZ  
397 molecules, measured as intensities, were determined at different time points in cells sampled  
398 during aerobic respiration (21%  $\text{O}_2$ ), and during and after transition from 1% oxygen respiration  
399 to denitrification (Fig. 4). The level of NorC molecules detected was two orders of magnitude  
400 lower than those for the other reductases, suggesting that membrane-bound enzymes require a  
401 different extraction protocol. Therefore, only values for the periplasmic reductases NapA, NirK  
402 and NosZ were included here. The aerobic pre-incubation over several generations effectively  
403 removed most of the old denitrification reductases, seen from the signals from NapA and NirK  
404 being 293-585 times lower than those observed during denitrification. Nos values were only 16-  
405 47 times lower, indicating low levels of constitutive transcription of the *nos* gene and translation  
406 of the corresponding mRNA under aerobic conditions. All three reductases increased in  
407 abundance during the  $\text{O}_2$  depletion phase (blue) and when  $\text{O}_2$  approached depletion, the  
408 NapA/Nos and NapA/NirK ratios were  $1.4 \pm 0.2$  and  $4.1 \pm 1.8$  (sd, n=3). Interestingly, the NapA/Nos  
409 ratio was stable during the period of strong respiration of exogenous  $\text{N}_2\text{O}$ , marked pink in Fig. 4,  
410 with values of  $1.5 \pm 0.3$  and  $1.2 \pm 0.2$ , although practically no  $\text{NO}_3^-$  reduction took place. The  
411 NapA/Nos ratio decreased to 0.7 during the phase of  $\text{NO}_3^-$  reduction to  $\text{N}_2$  (after depletion of

412 exogenous N<sub>2</sub>O). A similar pattern was found for the NapA/NirK ratios, while the NosZ/NirK ratios  
413 stayed around 1.0 during the anaerobic phase.

414

#### 415 **Discussion**

416 The study comprised a set of taxonomically diverse *Bradyrhizobium* strains, as shown by the  
417 phylogenetic tree (Figs. 1 and S7). This genus is commonly found among the most dominant  
418 organisms in soil microbial communities [54,55] and of large environmental and economic value  
419 for their N<sub>2</sub>-fixation ability in symbiosis with legumes. Based on the MLSA analysis of four  
420 concatenated housekeeping genes (Fig. 1) the strains clustered in eleven clades, covering various  
421 parts of the phylogenetic tree. Five strains in three clades did not group with described  
422 *Bradyrhizobium* species, suggesting that three new species may be delineated for these three  
423 clades. Our results from the denitrification analyses support other studies showing that complete  
424 denitrification pathways are more common in the *B. japonicum* supergroup than in other  
425 phylogenetic branches of *Bradyrhizobium* [18,19]. The complete denitrifiers in the present study  
426 belonged to clusters related to *B. shewense*, *B. ottawaense* and *B. guangxiense*, with the  
427 exception of one strain which clustered with the distantly related *B. lablabi* (Fig. 1). Although this  
428 suggests that some taxonomic groups of bradyrhizobia have higher frequency of complete  
429 denitrifiers than others, taxonomic position cannot be used as an indicator of the denitrification  
430 end-product since even very closely related strains may differ in this regard [19].

431 Recent research has revealed several transcriptional and post-transcriptional mechanisms  
432 that influence the amounts of intermediate products released from denitrifying organisms [56-  
433 60]. Although of value by itself, such information may be anecdotal. To understand the  
434 environmental relevance of a regulatory mechanism, it is important to know how widespread it  
435 is. We found the same preference for N<sub>2</sub>O over NO<sub>3</sub><sup>-</sup> as electron acceptor in all strains with  
436 complete denitrification, as did Mania et al. [19] for a different set of isolates. This indicates a  
437 phenomenon common to bradyrhizobia. In theory, it could be explained by lower numbers of  
438 NapA molecules compared to NosZ molecules. The results by Mania et al. [19] for strain AC 87j1,  
439 which is closely related to strain HAMBI 2125, could not disprove this, since transcription analysis  
440 showed 5-8 times lower copy numbers of transcripts coding for NapA compared to NosZ. This

441 prompted us to perform a proteomics analysis in the present study. The results (Fig. 4) provided  
442 clear evidence that the strong N<sub>2</sub>O reduction was not due to higher abundance of NosZ vs NapA,  
443 thus supporting the hypothesis by Mania et al. [19] that the electron pathway to NosZ is  
444 inherently more competitive than that to NapA.

445 Electrons from the quinol pool are channeled to Nap via the membrane-bound NapC, and to  
446 Nos, Nor and Nir via the membrane-bound *bc*<sub>1</sub> complex and periplasmic cytochromes [61]. Most  
447 likely, the electron pathway to NosZ via *bc*<sub>1</sub> is inherently more competitive than that to NapAB  
448 via NapC, plausibly due to a combination of stronger affinity for quinol and a higher *k*<sub>cat</sub> of *bc*<sub>1</sub>  
449 versus NapC. This implies a general mechanism which may apply not only to bradyrhizobia, but  
450 possibly to all denitrifiers that carry Nap and Nos but lack the membrane-bound Nar. Verifying  
451 this will require testing of a wide range of organisms. As a first step, we included in this study a  
452 comparison of the denitrification kinetics of the bradyrhizobia with *P. denitrificans*, which carries  
453 Nar in addition to Nap and Nos. The results supported our hypothesis by clearly demonstrating  
454 that, while the *Bradyrhizobium* strains left NO<sub>3</sub><sup>-</sup> untouched as long as exogenously supplied N<sub>2</sub>O  
455 was present (Figs. 2A, S3, S4, S5), *P. denitrificans* reduced these electron acceptors  
456 simultaneously (Fig. 2C). This suggests that the electron pathways from the quinol pool to NarG  
457 via NarI and NarH and to NosZ via the *bc*<sub>1</sub> complex and cytochrome c compete equally well for  
458 electrons, at least in *P. denitrificans*.

459 The proteomics analysis of HAMBI strain 2125 showed that NosZ was synthesized under fully  
460 aeriated conditions, while NapA and NirK were not. Before sampling for proteomics analysis, the  
461 cultures had been incubated in vials with vigorous stirring for at least 10 generations to secure  
462 that denitrification reductases produced during earlier exposure to denitrifying conditions were  
463 diluted to near extinction in the cells. This was the case for NapA and NirK in the sample from  
464 21% O<sub>2</sub> (Fig. 4). NosZ molecules were, however, approximately 10 times more abundant than  
465 NapA and NirK during this phase, which suggests constitutive, albeit low, synthesis of NosZ. The  
466 transcription analysis of strain AC87j1 reported by Mania et al. [19] provides no clear explanation  
467 to this, though. In the presence of 5 μM O<sub>2</sub> in the liquid, there were < 0.3 *nosZ* and *nirK* transcripts  
468 cell<sup>-1</sup> with only slightly higher levels of *nosZ* than *nirK*. The *napA* transcription was somewhat  
469 higher with 0.8 copies cell<sup>-1</sup>. It is possible, still, that a low but steady production of Nos takes



470 place in these organisms irrespective of O<sub>2</sub> concentrations, in line with the higher β-galactosidase  
471 activity observed for a *nos-lacZ* fusion of *B. japonicum* compared to a *nap-lacZ* fusion [62, 63].  
472 This could be a strategy for cells to be prepared for anaerobic conditions and avoid being  
473 entrapped in anoxia without energy to produce the denitrification machinery [56]. If so, it  
474 suggests that these bacteria may be ready to reduce any N<sub>2</sub>O produced by other organisms in  
475 their environment, thus acting as net sinks for this greenhouse gas.

476 Seven strains lacked the last step of denitrification and thus had N<sub>2</sub>O as end-product,  
477 explained by the absence of the entire *nos* operon in all cases (Table 2). When exposed to  
478 denitrifying conditions in the environment, such organisms will reduce all available N-oxides to  
479 N<sub>2</sub>O and thus act as net producers of this greenhouse gas. Three strains had NO as denitrification  
480 end-product, but only HAMBI 2149 lacked the entire *nor* operon (Table 2). The other two strains,  
481 HAMBI 2153 and 2299, contained *nor* operons but showed some sequence differences compared  
482 to strains with functional NO reduction. Further investigations trying to elucidate the reason for  
483 the lacking NO reduction would, however, be beyond the scope of the present study. Organisms  
484 that produce NO but are incapable of reducing it are apparently not uncommon [19, 64, 65]. In  
485 the environment such organisms are most likely found in consortia with organisms that express  
486 Nor or other NO-reducing enzymes.

487 The estimated growth yields were calculated based on the electron flow to the different  
488 reductases (Table S1) and the proportions were expected to match that of the charge separations  
489 per electron transport for the various respiratory pathways, as listed by [66]. The number of  
490 charge separations per 10 electrons ( $q/10e^-$ ) are 50 for low affinity oxidase, 40 for high affinity  
491 oxidases, 30 for Nos alone, and 26 for the entire Nap, Nir, Nor, Nos denitrification pathway (NO<sub>3</sub><sup>-</sup>  
492 to N<sub>2</sub>), which would a yield ratio of 100:60:52 for the three pathways (O<sub>2</sub>, N<sub>2</sub>O, NO<sub>3</sub><sup>-</sup>→N<sub>2</sub>) if we  
493 assume that low affinity oxidases are used, and 100:75:65 for high affinity oxidases. Our  
494 measured yields were 12, 8.5 and 6.7 \*10<sup>12</sup> cells per mol electrons for the three pathways, i.e. a  
495 yield ratio of 100:71:56, which agrees reasonably with that predicted for low affinity oxidases.

496 The emerging evidence that bradyrhizobia with complete denitrification deplete N<sub>2</sub>O in their  
497 surroundings before reducing available NO<sub>3</sub><sup>-</sup> is promising from an environmental point of view  
498 since they have the potential to act as net sinks for N<sub>2</sub>O produced by other organisms in their

499 vicinity. In legume fields, nodules and other N-rich organic material will be degraded, releasing  
500  $\text{NH}_4^+$ , which will rapidly be metabolized by nitrifying bacteria and archaea, producing  $\text{NO}_3^-$  while  
501 consuming  $\text{O}_2$ . This will create optimal conditions for denitrification. Legume cropped fields  
502 account for a significant share of the global  $\text{N}_2\text{O}$  emissions [5], and it is likely that coupled  
503 nitrification-denitrification is a main cause [67]. One way of mitigating these emissions is to  
504 enhance the number of efficient  $\text{N}_2\text{O}$  reducing bacteria in the soil microbial community. Rhizobial  
505 inoculants are used since long to increase the yield of leguminous crops and at the same time  
506 avoid the need for synthetic fertilizers. Different combinations of rhizobial strain and legume  
507 cultivar may differ substantially in nitrogen fixation effectiveness, which is also affected by  
508 environmental conditions [15, 68], and large efforts are put into the development of new  
509 inoculants that are compatible with various types of legumes as well as soil types and climates  
510 [69]. The present study shows that *Bradyrhizobium* strains also differ with respect to  $\text{N}_2\text{O}$   
511 reduction, and suggests that new inoculants should be selected among rhizobia that express Nos.

512

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517

## 518 **References**

- 519 1. Rogelj J, Shindell D, Jiang K, Fifita S, Forster P, Ginzburg V et al. Mitigation pathways compatible  
520 with 1.5 °C in the context of sustainable development. In, Masson-Delmotte V, Zhai P, Pörtner HO,  
521 Roberts D, Skea J, Shukla PR, et al. (eds). *Global Warming of 1.5 °C. An IPCC Special Report on the*  
522 *impacts of global warming of 1.5°C above pre-industrial levels and related global greenhouse gas*  
523 *emission pathways, in the context of strengthening the global response to the threat of climate*  
524 *change, sustainable development, and efforts to eradicate poverty*. 2018;  
525 [https://www.ipcc.ch/site/assets/uploads/sites/2/2019/02/SR15\\_Chapter2\\_Low\\_Res.pdf](https://www.ipcc.ch/site/assets/uploads/sites/2/2019/02/SR15_Chapter2_Low_Res.pdf).  
526 2. Thompson RL, Lassaletta L, Patra PK, Wilson C, Wells KC, Gressent A, et al. Acceleration of global

- 527 N<sub>2</sub>O emissions seen from two decades of atmospheric inversion. *Nat Clim Chang* 2019; **9**: 993–998.
- 528 3. Thomson AJ, Giannopoulos G, Pretty J, Baggs EM, Richardson DJ. Biological sources and sinks of  
529 nitrous oxide and strategies to mitigate emissions. *Philos Trans R Soc B* 2012; **367**: 1157-1168.
- 530 4. Reay DS, Davidson EA, Smith KA, Smith P, Melillo JM, Dentener F, et al. Global agriculture and  
531 nitrous oxide emissions. *Nat Clim Chang* 2012; **2**: 410–416.
- 532 5. Jensen ES, Peoples MB, Boddey RM, Gresshoff PM, Hauggaard-Nielsen H, Alves JR, et al. Legumes  
533 for mitigation of climate change and the provision of feedstock for biofuels and biorefineries. A  
534 review. *Agron Sustain Dev* 2012; **32**: 329–364.
- 535 6. Schlesinger WH. On the fate of anthropogenic nitrogen. *Proc Natl Acad Sci* 2009; **106**: 203–208.
- 536 7. Davidson EA, Kanter D. Inventories and scenarios of nitrous oxide emissions. *Environ Res Lett* 2014;  
537 **9**: 105012.
- 538 8. Winiwarter W, Höglund-Isaksson L, Klimont Z, Schöpp W, Amann M. Technical opportunities to  
539 reduce global anthropogenic emissions of nitrous oxide. *Environ Res Lett* 2018; **13**: 014011.
- 540 9. Hénault C, Revellin C. Inoculants of leguminous crops for mitigating soil emissions of the  
541 greenhouse gas nitrous oxide. *Plant Soil* 2011; **346**: 289-296.
- 542 10. Itakura M, Uchida Y, Akiyama H, Hoshino YT, Shimomura Y, Morimoto S, et al. Mitigation of nitrous  
543 oxide emissions from soils by *Bradyrhizobium japonicum* inoculation. *Nat Clim Chang* 2013; **3**: 208–  
544 212.
- 545 11. Akiyama H, Hoshino YT, Itakura M, Shimomura Y, Wang Y, Yamamoto A, et al. Mitigation of soil N<sub>2</sub>O  
546 emission by inoculation with a mixed culture of indigenous *Bradyrhizobium diazoefficiens*. *Sci Rep*  
547 2016; **6**: 32869.
- 548 12. Zilli JÉ, Alves BJR, Rouws LFM, Simões-Araujo JL, de Barros Soares LH, Cassán F, et al. The  
549 importance of denitrification performed by nitrogen-fixing bacteria used as inoculants in South  
550 America. *Plant Soil* 2019; **451**: 5-24.
- 551 13. Obando M, Correa-Galeote D, Castellano-Hinojosa A, Gualpa J, Hidalgo A, Alché JD, et al. Analysis  
552 of the denitrification pathway and greenhouse gases emissions in *Bradyrhizobium* sp. strains used  
553 as biofertilizers in South America. *J Appl Microbiol* 2019; **127**: 739–749.
- 554 14. Thrall PH, Laine A-L, Broadhurst LM, Bagnall DJ, Brockwell J. Symbiotic effectiveness of rhizobial  
555 mutualists varies in interactions with native Australian legume genera. *PLoS ONE* 2011; **6**: e23545.
- 556 15. Woliy K, Degefu T, Frostegård Å. Host Range and Symbiotic Effectiveness of N<sub>2</sub>O Reducing  
557 *Bradyrhizobium* Strains. *Front Microbiol* 2019; **10**: 2746.
- 558 16. Santos MS, Nogueira MA, Hungria M. Microbial inoculants: reviewing the past, discussing the

- 559 present and previewing an outstanding future for the use of beneficial bacteria in agriculture. *AMB*  
560 *Expr* 2019; **9**: 205.
- 561 17. Arya SS, Salve AR, Chauhan S. Peanuts as functional food: a review. *J Food Sci Technol* 2016; **53**:  
562 31–41.
- 563 18. Sameshima-Saito R, Chiba K, Minamisawa K. New method of denitrification analysis of  
564 *Bradyrhizobium* field isolates by gas chromatographic determination of <sup>15</sup>N-labeled N<sub>2</sub>. *Appl*  
565 *Environ Microbiol* 2004; **70**: 2886–2891.
- 566 19. Mania D, Woliy K, Degefu T, Frostegård Å. A common mechanism for efficient N<sub>2</sub>O reduction in  
567 diverse isolates of nodule-forming bradyrhizobia. *Environ Microbiol* 2020; **22**: 17–31.
- 568 20. Sánchez C, Minamisawa K. Redundant roles of *Bradyrhizobium oligotrophicum* Cu-type (NirK) and  
569 cd1-type (NirS) nitrite reductase genes under denitrifying conditions. *FEMS Microbiol Lett* 2018;  
570 **365**: fny015.
- 571 21. Velasco L, Mesa S, Delgado MJ, Bedmar EJ. Characterization of the *nirK* gene encoding the  
572 respiratory, Cu-containing nitrite reductase of *Bradyrhizobium japonicum*. *Biochim Biophys Acta*  
573 2001; **1521**: 130–134.
- 574 22. Velasco L, Mesa S, Xu C, Delgado MJ, Bedmar EJ. Molecular characterization of *nosRZDFYLX* genes  
575 coding for denitrifying nitrous oxide reductase of *Bradyrhizobium japonicum*. *Antonie Van*  
576 *Leeuwenhoek* 2004; **85**: 229–235.
- 577 23. Mesa S, Velasco L, Manzanera ME, Delgado MJ, Bedmar EJ. Characterization of the *norCBQD* genes,  
578 encoding nitric oxide reductase, in the nitrogen fixing bacterium *Bradyrhizobium japonicum*.  
579 *Microbiology* 2002; **148**: 3553–3560.
- 580 24. Delgado MJ, Bonnard N, Tresierra-Ayala A, Bedmar EJ, Müller P. The *Bradyrhizobium japonicum*  
581 *napEDABC* genes encoding the periplasmic nitrate reductase are essential for nitrate respiration.  
582 *Microbiology* 2003; **149**: 3395–3403.
- 583 25. Bedmar EJ, Robles EF, Delgado MJ. The complete denitrification pathway of the symbiotic,  
584 nitrogen-fixing bacterium *Bradyrhizobium japonicum*. *Biochem Soc Trans* 2005; **33**: 141–144.
- 585 26. Richardson DJ. Bacterial respiration: a flexible process for a changing environment. *Microbiology*  
586 2000; **146**: 551–571.
- 587 27. Richardson DJ, Berks BC, Russell DA, Spiro S, Taylor CJ. Functional, biochemical and genetic  
588 diversity of prokaryotic nitrate reductases. *Cell Mol Life Sci* 2001; **58**: 165–178.
- 589 28. Berks BC, Ferguson SJ, Moir JWB, Richardson DJ. Enzymes and associated electron transport  
590 systems that catalyse the respiratory reduction of nitrogen oxides and oxyanions. *Biochim Biophys*

- 591            *Acta* 1995; **1232**: 97–173.
- 592    29. Zumft WG. Cell biology and molecular basis of denitrification. *Microbiol Mol Biol Rev* 1997; **61**:  
593            533–616.
- 594    30. Zhang L, Wüst A, Prasser B, Müller C, Einsle O. Functional assembly of nitrous oxide reductase  
595            provides insights into copper site maturation. *Proc Natl Acad Sci* 2019; **116**: 12822–12827.
- 596    31. Zhang XP, Nick G, Kaijalainen S, Terefework Z, Paulin L, Tighe SW, et al. Phylogeny and Diversity of  
597            *Bradyrhizobium* Strains isoalted from the Root Nodules of Peanut (*Arachis hypogaea*) in Sichuan,  
598            China. *Syst Appl Microbiol* 1999; **22**: 378-386.
- 599    32. Van Rossum D, Muyotcha A, Van Verseveld HW, Stouthamer AH, Boogerd FC. Effects of  
600            *Bradyrhizobium* strain and host genotype, nodule dry weight and leaf aera on groundnut (*Arachis*  
601            *hypogaea* L. ssp. *fastigiata*) yield. *Plant and Soil* 1993; **154**: 279-288.
- 602    33. Van Rossum D, Muyotcha A, Van Verseveld HW, Stouthamer AH, Boogerd FC. Siderophore  
603            production by *Bradyrhizobium* spp. Strains nodulating groundnut. *Plant and soil* 1994; **163**: 177-  
604            187.
- 605    34. Chang YL, Wang ET, Sui XH, Zhang XX, Chen WX. Molecular diversity and phylogeny of rhizobia  
606            associated with *Lablab purpureus* (Linn.) grown in Southen China. *Syst Appl Microbiol* 2011; **34**:  
607            276-284.
- 608    35. Wang R, Chang YL, Zheng WT, Zhang D, Zhang XX, Sui XH, et al. *Bradyrhizobium arachidis* sp. Nov.,  
609            isolated from effective nodules of *Arachis hypogaea* grown in China *Syst Appl Microbiol* 2013; **36**:  
610            101-105.
- 611    36. Li YH, Wang R, Zhang XX, Young PW, Wang ET, Sui XH, et al. *Bradyrhizobium guangdongense* sp.  
612            nov. and *Bradyrhizobium guangxiense* sp.nov., isolated from effectvie nodules of peanut. *Int J Syst*  
613            *Evol Microbiol* 2015; **65**: 4655-4661.
- 614    37. Degefu T, Wolde-meskel E, Woliy K, Frostegård Å. Phylogenetically diverse groups of  
615            *Bradyrhizobium* isolated from nodules of tree and annual legume species growing in Ethiopia. *Syst*  
616            *Appl Microbiol* 2017; **40**: 205-214.
- 617    38. Schwartz W, Vincent JM, a manual for the practical study of the root-nodule bacteria (IBP  
618            Handbuch no. 15 des international biology program, London). XI u. 164 S., 10 Abb., 17 tab., 7 Taf.  
619            Oxford-Edinburgh 1970: Blackwell scientific Publ.,45 s. *Z Allg Mikrobiol* 1972; **12**: 1521–4028.
- 620    39. Molstad L, Dörsch P, Bakken LR. Robotized incubation system for monitoring gases (O<sub>2</sub>, NO, N<sub>2</sub>O  
621            N<sub>2</sub>) in denitrifying cultures. *J Microbiol Methods* 2007; **71**: 202–211.
- 622    40. Sistrom WR. The kinetics of the synthesis of photopigments in *Rhodopseudomonas spheroides*. *J*

- 623 *Gen Microbiol* 1962; **28**: 607–616.
- 624 41. Bergaust L, Mao Y, Bakken LR, Frostegård Å. Denitrification response patterns during the transition  
625 to anoxic respiration and posttranscriptional effects of suboptimal pH on nitrogen oxide reductase  
626 in *Paracoccus denitrificans*. *Appl Environ Microbiol* 2010; **76**: 6387–6396.
- 627 42. Lycus P, Bøthun KL, Bergaust L, Shapleigh JP, Bakken LR, Frostegård Å. Phenotypic and genotypic  
628 richness of denitrifiers revealed by a novel isolation strategy. *ISME J* 2017; **11**: 2219–2232.
- 629 43. Molstad L, Dörsch P, Bakken LR. Improved robotized incubation system for gas kinetics in batch  
630 cultures. *ResearchGate* 2016; <https://doi.org/10.13140/RG.2.2.30688.07680>.
- 631 44. Cox RD. Determination of nitrate and nitrite at the parts per billion level by chemiluminescence.  
632 *Anal Chem* 1980; **52**: 332–335.
- 633 45. MacArthur PH, Shiva S, Gladwin MT. Measurement of circulating nitrite and S-nitrosothiols by  
634 reductive chemiluminescence. *J Chromatogr B* 2007; **851**: 93–105.
- 635 46. Larkin MA, Blackshields G, Brown NP, Chenna R, Mcgettigan PA, McWilliam H, et al. Clustal W and  
636 Clustal X version 2.0. *Bioinformatics* 2007; **23**: 2947–2948.
- 637 47. Hall TA. BioEdit: a user-friendly biological sequence alignment editor and analysis program for  
638 Windows 95/98/NT. *Nucleic Acids Symposium Series* 1999; **41**: 95–98.
- 639 48. Kumar S, Stecher G, Tamura K. MEGA7: Molecular Evolutionary Genetics Analysis Version 7.0 for  
640 Bigger Datasets. *Mol Biol Evol* 2016; **33**: 1870–1874.
- 641 49. Ronquist F, Teslenko M, Van Der Mark P, Ayres DL, Darling A, Höhna S, et al. Mrbayes 3.2: Efficient  
642 bayesian phylogenetic inference and model choice across a large model space. *Syst Biol* 2012; **61**:  
643 539–542.
- 644 50. Rambaut A, Drummond AJ, Xie D, Baele G, Suchard MA. Posterior summarization in Bayesian  
645 phylogenetics using Tracer 1.7. *Syst Biol* 2018; **67**: 901–904.
- 646 51. Chaumeil PA, Mussig AJ, Hugenholtz P, Parks DH. GTDB-Tk: a toolkit to classify genomes with the  
647 Genome Taxonomy Database. *Bioinformatics* 2020; **36**: 1925–1927.
- 648 52. Conthe M, Lycus P, Arntzen MØ, Ramos da Silva A, Frostegård Å, Bakken LR, et al. Denitrification as  
649 an N<sub>2</sub>O sink. *Water Res* 2019; **151**: 381–387.
- 650 53. Avontuur JR, Palmer M, Beukes CW, Chan WY, Coetzee MPA, Blom J, et al. Genome-informed  
651 *Bradyrhizobium* taxonomy: where to from here? *Syst Appl Microbiol* 2019; **42**: 427–439.
- 652 54. Delmont TO, Prestat E, Keegan KP, Faubladièr M, Robe P, Clark IM, et al. Structure, fluctuation and  
653 magnitude of a natural grassland soil metagenome. *ISME J* 2012; **6**: 1677–1687.
- 654 55. Xu ZF, Hansen MA, Hansen LH, Jacquiod S, Sørensen SJ. Bioinformatic Approaches Reveal

- 655 Metagenomic Characterization of Soil Microbial Community. *PLoS ONE* 2014; **9**: e93445.
- 656 56. Lycus P, Soriano-Laguna MJ, Kjos M, Richardson DJ, Gates AJ, Milligan DA, et al. A bet-hedging  
657 strategy for denitrifying bacteria curtails their release of N<sub>2</sub>O. *Proc Natl Acad Sci* 2018; **115**: 11820–  
658 11825.
- 659 57. Torres MJ, Simon J, Rowley G, Bedmar EJ, Richardson DJ, Gates AJ, et al. Nitrous oxide metabolism  
660 in nitrate-reducing bacteria: physiology and regulatory mechanisms. In: Poole RK (eds). *Advances in*  
661 *microbial physiology*, first edition. Elsevier, 2016; pp 353-432.
- 662 58. Liu B, Mao Y, Bergaust L, Bakken LR, Frostegård Å. Strains in the genus *Thauera* exhibit remarkably  
663 different denitrification regulatory phenotypes. *Environ Microbiol* 2013; **15**: 2816–2828.
- 664 59. Gaimster H, Hews CL, Griffiths R, Soriano-Laguna MJ, Alston M, Richardson DJ, et al. A central small  
665 RNA regulatory circuit controlling bacterial denitrification and N<sub>2</sub>O emissions. *mBio* 2019; **10**:  
666 e01165-19.
- 667 60. Jiménez-Leiva A, Cabrera JJ, Bueno E, Torres MJ, Salazar S, Bedmar EJ, et al. Expanding the Regulon  
668 of the *Bradyrhizobium diazoefficiens* NnrR Transcription Factor: New Insights Into the  
669 Denitrification Pathway. *Front Microbiol* 2019; **10**: 1926.
- 670 61. Shapleigh JP. The prokaryotes: Prokaryotic physiology and biochemistry. In Rosenberg E, DeLong  
671 EF, Lory S, Stackebrandt E, Thompson F (eds). *The prokaryotes*. Springer: Berlin Heidelberg, 2013;  
672 pp 405–425.
- 673 62. Bueno E, Robles EF, Torres MJ, Krell T, Bedmar EJ, Delgado MJ, et al. Disparate response to  
674 microoxia and nitrogen oxides of the *Bradyrhizobium japonicum* *napEDABC*, *nirK* and *norCBQD*  
675 denitrification genes. *Nitric Oxide* 2017; **68**: 137–149.
- 676 63. Torres MJ, Bueno E, Jiménez-Leiva A, Cabrera JJ, Bedmar EJ, Mesa S, et al. FixK<sub>2</sub> Is the Main  
677 Transcriptional Activator of *Bradyrhizobium diazoefficiens* *nosRZDYFLX* Genes in Response to Low  
678 Oxygen. *Front Microbiol* 2017; **8**: 1621.
- 679 64. Falk S, Liu B, Braker G. Isolation, genetic and functional characterization of novel soil *nirK*-type  
680 denitrifiers. *Syst Appl Microbiol* 2010; **33**: 337–347.
- 681 65. Casella S, Shapleigh JP, Toffanin A, Basaglia M. Investigation into the role of the truncated  
682 denitrification chain in *Rhizobium sulae* strain HCNT1. *Biochem Soc Trans* 2006; **34**: 130-132.
- 683 66. Van Spanning R, Richardson D, Ferguson S. Introduction to the biochemistry and molecular biology  
684 of denitrification. In: Bothe H, Ferguson SJ, Newton WE (eds). *Biology of the nitrogen cycle*. Elsevier:  
685 Amsterdam, The Netherlands, 2007; pp 3–21.
- 686 67. Bakken LR, Frostegård Å. Sources and sinks for N<sub>2</sub>O, can microbiologist help to mitigate N<sub>2</sub>O

- 687 emissions? *Environ Microbiol* 2017; **19**: 4801-4805.
- 688 68. Lindström K, Mousavi SA. Effectiveness of nitrogen fixation in rhizobia. *Microb. Biotechnol.* 2019;  
689 <https://doi.org/10.1111/1751-7915.13517>.
- 690 69. Martínez J, Negrete-Yankelevich S, Godinez LG, Reyes J, Esposti MD, Martínez Romero E. Short-  
691 Term Evolution of Rhizobial Strains Toward Sustainability in Agriculture. In *Microbial Models: From*  
692 *Environmental to Industrial Sustainability*. Springer: Singapore, 2016; pp. 277–292.
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697 **Table 1 Strains used in the current study.** Strains AC 70c, AC 86d2, AC 86j1 and AC 101b (below dashed  
698 line) were included in the phylogenetic analysis in the present study but were not examined for  
699 denitrification since this was reported elsewhere [19].

Strain	Host plant	Geographic origin	Reference
HAMBI 2115*	<i>Arachis hypogaea</i> L.	Zimbabwe	[31]
HAMBI 2116	<i>A. hypogaea</i> L.	Zimbabwe	[31]
HAMBI 2125	<i>A. hypogaea</i> L.	Sichuan, China	[31]
HAMBI 2127	<i>A. hypogaea</i> L.	Sichuan, China	[31]
HAMBI 2128	<i>Macrotyloma africanum</i>	Zimbabwe	[32]
HAMBI 2130	<i>Vigna unguiculata</i>	Zimbabwe	[33]
HAMBI 2134	<i>A. hypogaea</i> L.	Sichuan, China	[31]
HAMBI 2142	<i>A. hypogaea</i> L.	Sichuan, China	[31]
HAMBI 2149	<i>A. hypogaea</i> L.	Sichuan, China	[31]
HAMBI 2153	<i>A. hypogaea</i> L.	Israel	[31]
HAMBI 2299	<i>A. hypogaea</i> L.	Hawaii, USA	<a href="#">HAMBI**</a>
HAMBI 3052 <sup>T</sup>	<i>Lablab purpureus</i>	Anhui, China	[34]
CCBAU 051107 <sup>T</sup>	<i>A. hypogaea</i> L.	Hebei, China	[35]
CCBAU 23155	<i>A. hypogaea</i> L.	Anhui, China	[35]
CCBAU 33067	<i>A. hypogaea</i> L.	Jiangxi, China	[35]
CCBAU 45332	<i>A. hypogaea</i> L.	Henan, China	[35]
CCBAU 53344	<i>A. hypogaea</i> L.	Guangxi, China	[36]
CCBAU 53363 <sup>T</sup>	<i>A. hypogaea</i> L.	Guangxi, China	[36]
CCBAU 53429	<i>A. hypogaea</i> L.	Guangxi, China	[36]
AC70c	<i>Phaseolus vulgaris</i>	Akaki, Ethiopia	[37]
AC86d2	<i>Cajanus cajan</i>	Awassa, Ethiopia	[37]
AC87j	<i>Millettia ferruginea</i>	Awassa, Ethiopia	[37]
AC101b	<i>Acacia saligna</i>	Leku, Ethiopia	[37]

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701 HAMBI: The culture collection at the Department of Microbiology, University of Helsinki.

702 CCBAU: Culture Collection of China Agricultural University, Beijing, China.

703 AC: The strains are kept in a local collection at Norwegian University of Life Sciences and available upon request (Å. Frostegård).

704 \*) For information on donator and specimen history of HAMBI strains please consult <https://kotka.luomus.fi/culture/bac>

705 \*\*) <https://kotka.luomus.fi/view?uri=http://tun.fi/RAH.1747&type=document&page=1&spot=1>

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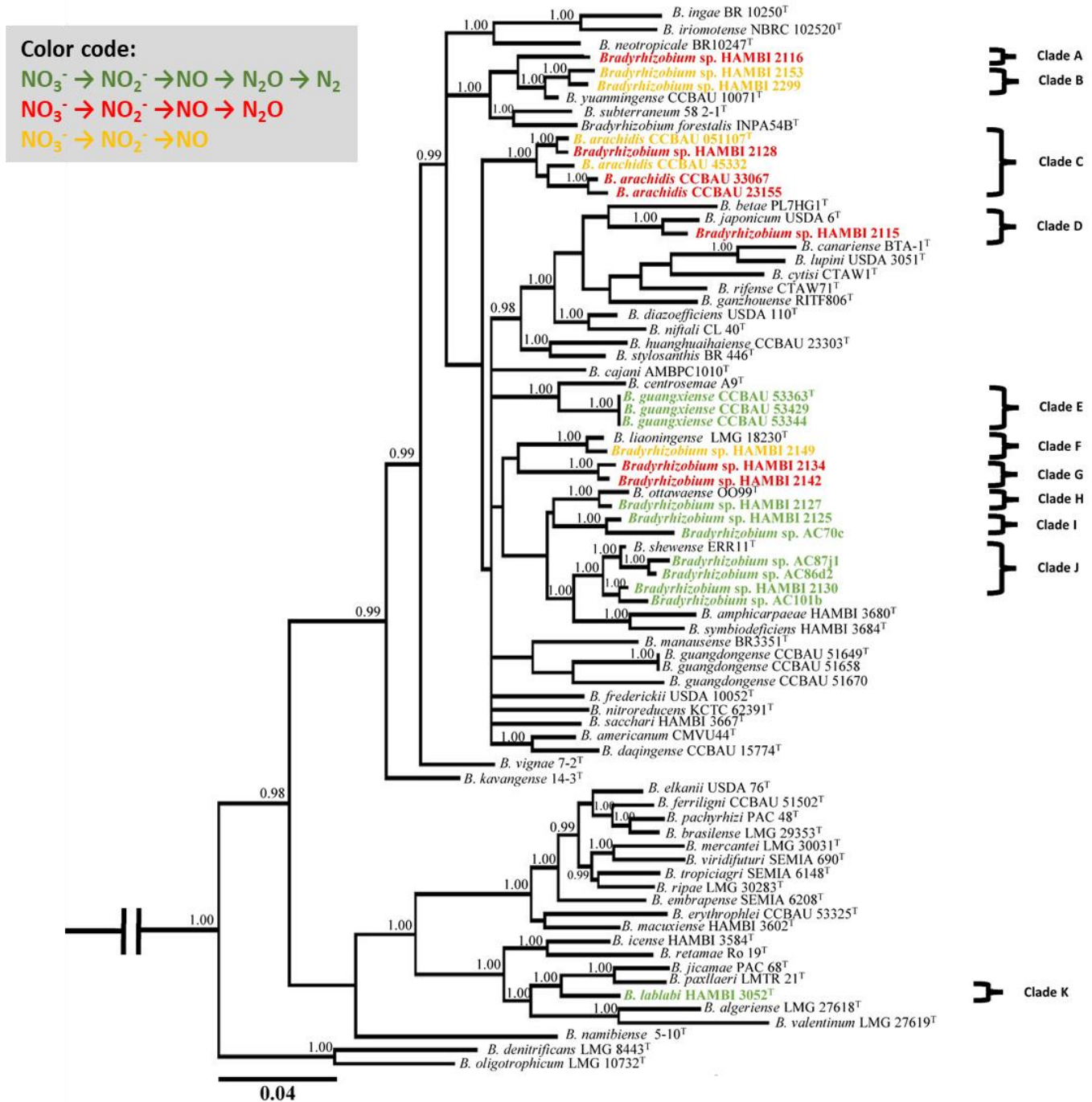
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713 **Table 2** Denitrification gene operons of *Bradyrhizobium* strains, analyzed based on whole genome  
 714 sequencing. Strains placed within the same section of the table were highly similar as seen from pairwise  
 715 genome comparison, with ANI values >99.2, and in most cases >99.9 % (Table S4). They also had the same  
 716 denitrification phenotype. Strains representing each of these clusters are marked in bold and were  
 717 included in the phylogenetic tree (Fig. 1). Presence (+) or absence (-) of functional and  
 718 regulatory/accessory genes are indicated. Colors represent a function detected after analysis of  
 719 denitrification end products (blue= nitrate reduction; yellow = nitrite reduction; red = nitric oxide  
 720 reduction; and green = nitrous oxide reduction). Blank = absence of function. The amino acid sequences  
 721 of the *nor* operons of strains HAMBI 2299 and HAMBI 2153 were compared to those of the strains listed  
 722 here (in bold) that produced functional reductases (except strain CCBAU 53363<sup>T</sup>).

<i>Bradyrhizobium</i> strains	Nitrate reduction					Nitrite reduction		Nitric oxide reduction					Nitrous oxide reduction						
	<i>nap</i>					<i>nir</i>		<i>nor</i>					<i>nos</i>						
	<i>E</i>	<i>D</i>	<i>A</i>	<i>B</i>	<i>C</i>	<i>K</i>	<i>V</i>	<i>C</i>	<i>B</i>	<i>Q</i>	<i>D</i>	<i>E</i>	<i>R</i>	<i>Z</i>	<i>D</i>	<i>Y</i>	<i>F</i>	<i>L</i>	<i>X</i>
<b>HAMBI 2125</b> HAMBI 2126	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
<b>HAMBI 2127</b>	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
<b>HAMBI 2130</b>	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
<b>HAMBI 3052<sup>T</sup></b>	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
<b>*CCBAU 53363<sup>T</sup></b>	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
<b>HAMBI 2115</b>	+	+	+	+	+	+	+	+	+	+	+	+	-	-	-	-	-	-	-
<b>HAMBI 2116</b>	+	+	+	+	+	+	+	+	+	+	+	+	-	-	-	-	-	-	-
<b>HAMBI 2128</b> HAMBI 2129 HAMBI 2150	+	+	+	+	+	+	+	+	+	+	+	+	-	-	-	-	-	-	-
<b>HAMBI 2134</b> HAMBI 2135	+	+	+	+	+	+	+	+	+	+	+	+	-	-	-	-	-	-	-
<b>HAMBI 2142</b>	+	+	+	+	+	+	+	+	+	+	+	+	-	-	-	-	-	-	-
<b>**HAMBI 2153</b>	+	+	+	+	+	+	+	+	-	-	+	+	-	-	-	-	-	-	-
<b>HAMBI 2299</b>	+	+	+	+	+	+	+	+	+	+	+	+	-	-	-	-	-	-	-
<b>HAMBI 2149</b> HAMBI 2133 HAMBI 2136 HAMBI 2137 HAMBI 2151 HAMBI 2152	+	+	+	+	+	+	+	-	-	-	-	-	-	-	-	-	-	-	-

723 \*The accession number for the complete genome sequence of CCBAU 53363 in NCBI is CP022219.

724 \*\* A frame shift in *norB*; *norQ* is not in the correct reading frame.

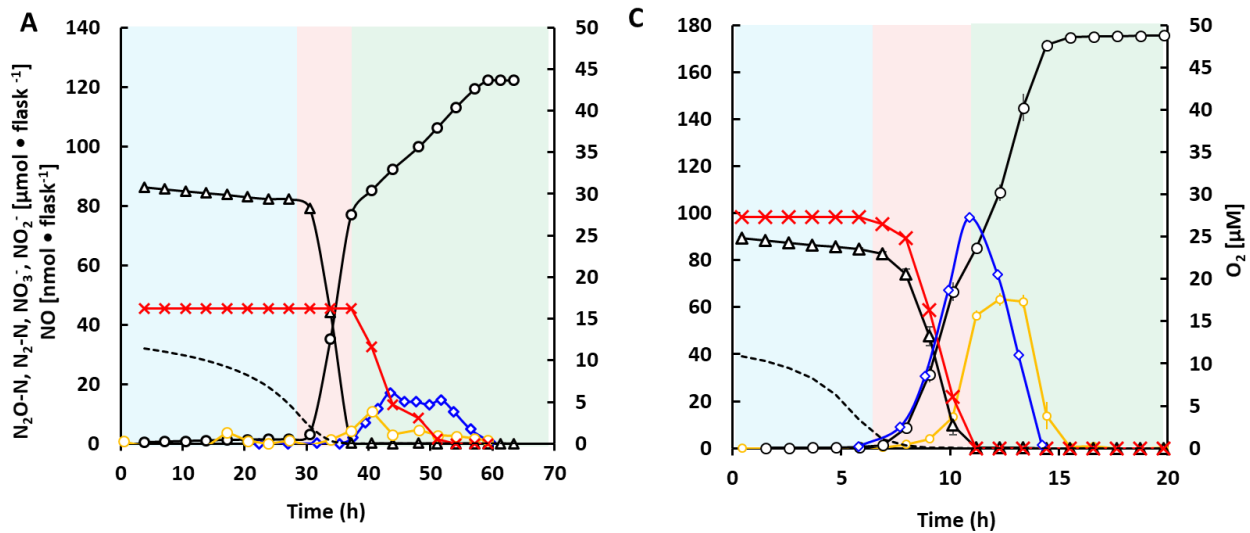


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727 **Fig. 1** Phylogenetic tree of four concatenated housekeeping genes (*atpD*- *glnII*- *recA*- *rpoB*) from  
 728 78 bradyrhizobial strains constructed based on a Bayesian inference analysis. Only the posterior  
 729 probabilities  $\geq 0.95$  are shown in the tree. The genus name *Bradyrhizobium* is abbreviated as  
 730 *B.*, and a "T" at the end of each strain code shows the type strains. Colors indicate denitrification  
 731 end-products, determined in this study, except for strains AC87ji, AC 86d2, AC 70c and AC 101b  
 732 which were reported by [19].

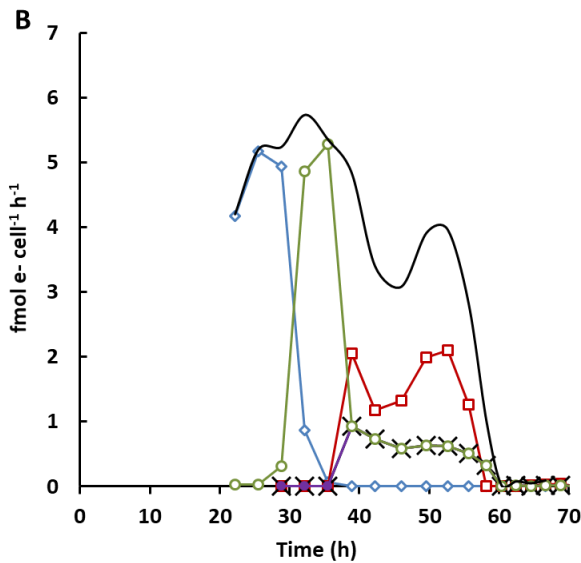
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--- O<sub>2</sub>    × NO<sub>3</sub><sup>-</sup>    ◆ NO<sub>2</sub><sup>-</sup>    ○ NO    ▲ N<sub>2</sub>O    ○ N<sub>2</sub>



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◆ VeO<sub>2</sub>    □ VeNar/Nap    × VeNir    ● VeNor    ○ VeNos    — Vetotal

738 **Fig. 2 Competition for electrons between N<sub>2</sub>O- and NO<sub>3</sub><sup>-</sup> reductase; comparison of Nap- and Nar-carrying**

739 **organisms.** Comparison of gas kinetics and corresponding electron flow to oxidases (aerobic respiration)

740 and the different N-reductases measured in batch cultures of *Bradyrhizobium* strain HAMBI 2125, which

741 carries Nap and Nos but lacks Nar (A and B) and *Paracoccus denitrificans* which carries Nap, Nar and Nos

742 (C and D). The top panels (A & C) show the measured amounts of NO<sub>2</sub><sup>-</sup>, NO, N<sub>2</sub>O and N<sub>2</sub> per vial, and the

743 concentration of O<sub>2</sub> in the liquid. Note that the small but significant decrease in N<sub>2</sub>O during the

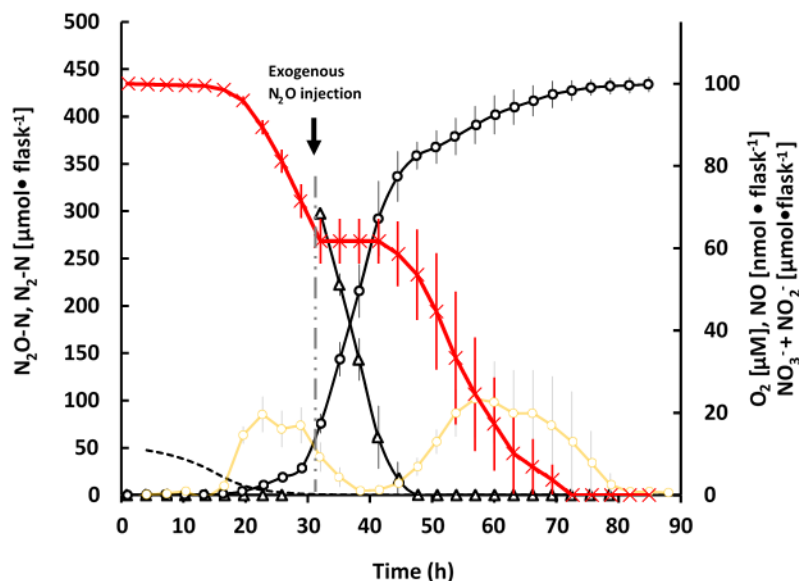
744 aerobic phase was caused by sampling losses, not by N<sub>2</sub>O reduction or other reactions, as

745 explained in Fig. S9. The The bottom panels show the calculated cell specific rates of electron flow (fmol

746 cell<sup>-1</sup> h<sup>-1</sup>) to the terminal oxidases (V<sub>eO<sub>2</sub></sub>) and to the four reductases (V<sub>eNar/Nap</sub>, V<sub>eNir</sub>, V<sub>eNor</sub> and V<sub>eNos</sub>). V<sub>eNir</sub> and

747  $V_{eNor}$  were practically identical throughout (NO remained very low), thus indistinguishable in the panels.  
748 Strain HAMBI 2125 was incubated in 4 replicate flasks with YMB supplemented with 1 mM  $KNO_3$ , of which  
749 only one is shown here (replicate vials are shown in Fig. S3). *P. denitrificans* was incubated in 50 ml  
750 Siström's medium with 2 mM  $KNO_3$  (n=7 replicate vials, standard error indicated by vertical lines). At the  
751 start of the incubation, all vials contained ~1%  $O_2$  in the headspace (~10  $\mu M$  in liquid) and ~ 1.2 ml  $N_2O$   
752 (~90  $\mu mol N_2O-N$ ).  $V_{eNar/Nap}$  and  $V_{eNir}$  in legend (B and D) represents rate of electron transport to  $NO_3^-$  and  
753  $NO_2^-$ , respectively (via NapA and NirK for *Bradyrhizobium* and NarG+Nap and NirS for *P. denitrificans*).

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----  $O_2$      $\times$   $NO_3^- + NO_2^-$      $\circ$  NO     $\triangle$   $N_2O$      $\circ$   $N_2$

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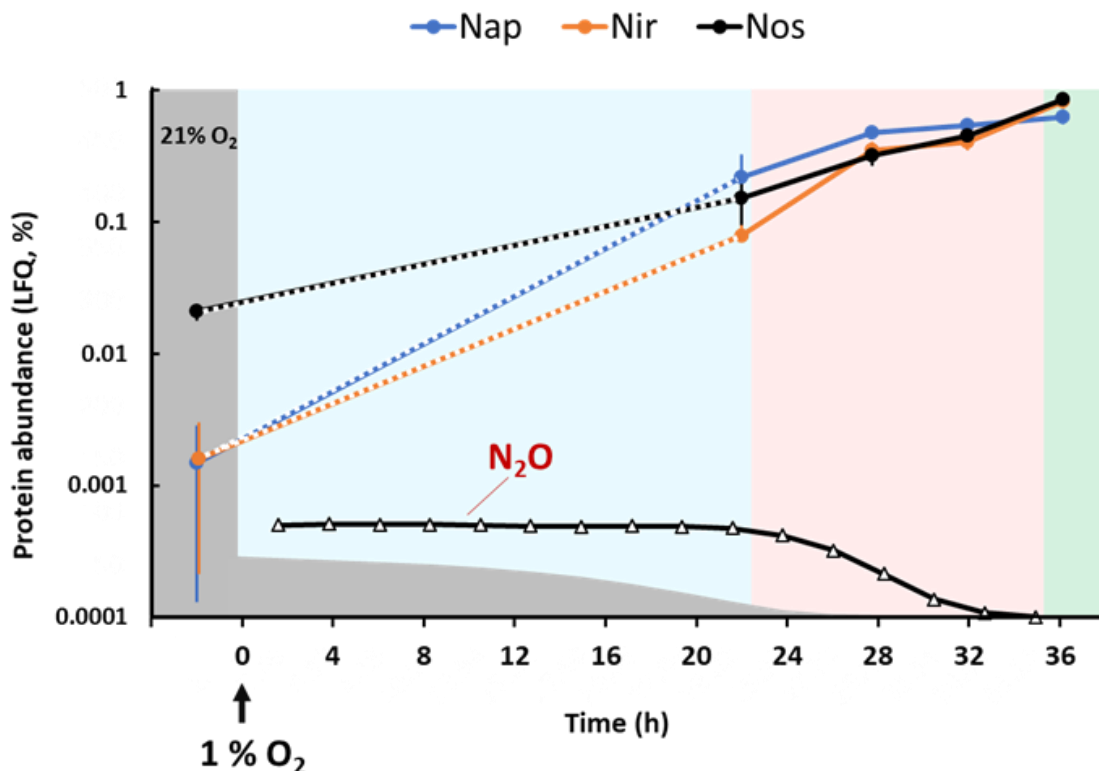
761 **Fig. 3 Effect of injecting  $N_2O$  to an actively denitrifying culture.** Gas kinetics for *Bradyrhizobium* strain  
762 HAMBI 2125 during and after transition from aerobic respiration to N-oxide respiration. Triplicate cultures  
763 were incubated in YMB medium supplemented with 2 mM  $KNO_3$  and with 0.7 ml  $O_2$  in the headspace  
764 (around 10  $\mu M$   $O_2$  in the medium). Approximately 4 ml  $N_2O$  (~300  $\mu mol N_2O-N$ ) were injected at 32 hpi,  
765 indicated by the arrow, into actively denitrifying cultures respiring  $NO_3^-$ . Bars show standard errors (n=3).

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771 **Fig. 4 Quantification of denitrifying reductase molecules throughout transition from aerobic to**  
772 **anaerobic respiration.** The panel shows the concentrations of Nap, Nir and Nos determined by proteomics  
773 analysis together with the concentration of O<sub>2</sub> and N<sub>2</sub>O. The abundance of the individual denitrification  
774 reductases were calculated as percentages of the sum of all LFQ (label-free quantitation of proteins)  
775 abundances for each time point. The first samples were taken from triplicate vials of aerobically grown of  
776 *Bradyrhizobium* strain HAMBI 2125 growing under fully aerobic conditions (21% O<sub>2</sub>). The precultures were  
777 used to inoculate vials with 1% O<sub>2</sub>, which were sampled when reaching hypoxia, and during the  
778 subsequent anoxic respiration. The amounts of O<sub>2</sub> and N<sub>2</sub>O are indicated to illustrate the time of their  
779 depletion, and the colors indicate 4 distinct phases: 1) fully oxic (gray), 2) transition to hypoxic conditions  
780 (blue), 3) early anoxic with respiration based on N<sub>2</sub>O-reduction and 4) anoxic with respiration by NO<sub>3</sub>  
781 reduction (green). For actual values of [O<sub>2</sub>] [N<sub>2</sub>O], [N<sub>2</sub>] and [NO<sub>2</sub><sup>-</sup>], see Fig. S6. Bars show standard deviation  
782 (n=3).

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