

Climate-risk to European fisheries and coastal communities

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30 **Abstract**

31 **With the majority of the global human population living in coastal regions, identifying the climate**
32 **risk that ocean-dependent communities and businesses are exposed to is key to prioritising the finite**
33 **resources available to support adaptation. Here we apply a climate-risk analysis across the European**
34 **fisheries sector for the first time to identify the most at-risk fleets and sub-national regions. We**
35 **combine a trait-based approach with ecological niche models to differentiate climate hazards between**
36 **populations of fish and use them to assess the relative climate risk for 380 fishing fleets and 105**
37 **coastal regions in Europe. Countries in SE Europe and the UK have the highest risks to both their**
38 **fishing fleets and their communities while, in other countries, the risk-profile is greatest at either the**
39 **fleet or community level. These results reveal the diversity of challenges posed by climate-change to**
40 **European fisheries: climate adaptation, therefore, needs to be tailored to each country's and even**
41 **each region's specific situation. Our analysis supports this process by highlighting where adaptation**
42 **measures are needed and could have the greatest impact.**

43 Manuscript Body

44 The ocean provides human societies with a wide variety of goods and services, ranging from food and
45 employment to climate regulation and cultural nourishment¹. Climate change is already shifting the
46 abundance, distribution, productivity and phenology of living marine resources²⁻⁴ and, thereby, many of
47 the ecosystem services that we depend upon⁵. These impacts, however, are not being experienced uniformly
48 by human society but depend on the characteristics and context of the community or business affected.
49 Raising awareness and understanding the risk to human systems is, therefore, a key first step⁶ to developing
50 and prioritising appropriate adaptation options in response to the challenges of the climate crisis⁷.

51 Over the past decades, climate risk assessments (CRAs) and climate vulnerability assessments (CVAs) have
52 been developed to support such a prioritisation. The approach, developed by the Intergovernmental Panel
53 on Climate Change (IPCC), has shifted over time from a focus on “vulnerability” to a focus on “risk”⁸, in
54 part due to criticisms of the negative framing that “vulnerability” implies⁹. The modern CRA framework
55 considers risk as the intersection of hazard, exposure and vulnerability¹⁰. CVAs, and more recently CRAs,
56 have been applied widely in the marine realm, for example in coastal communities in northern Vietnam¹¹,
57 Kenya¹² and the USA¹³, at the national level across coastal areas of the USA^{14,15} and Australia^{16,17}, across
58 regions such as Pacific island nations^{18,19} and globally^{6,20,21}. Several ‘best practice’ guides have also been
59 developed^{7,22}.

60 CRAs and CVAs covering European waters, however, are notable by their absence. The lack of attention
61 to climate risk in European fisheries may arise in part from the results of early global CVAs⁶ that ranked
62 European countries as having low vulnerabilities due to their affluence and, therefore, high ‘adaptive
63 capacity’. Yet the European region poses unique challenges when assessing climate-risks due to its wide
64 range of species, biogeographical zones and habitats. Fishing techniques and the scale of fisheries vary
65 widely, from large fleets of small vessels in the Mediterranean Sea²³ to some of the largest fishing vessels
66 in the world (e.g. the 144-m long *Annelies Ilena*). Furthermore, although fisheries contribute very little to
67 national GDP, food or income-security for most countries²⁴, in specific communities and regions fishing is
68 the mainstay of employment²⁵. Adapting European fisheries to a changing climate, therefore, requires the
69 development of robust analyses capable of assessing the climate-risk across this extremely diverse
70 continent.

71 We conducted a detailed CRA across the European marine fisheries sector, estimating the climate risk of
72 both fishing fleets and coastal regions in one integrated analysis. Our analyses span more than 50 degrees
73 of latitude from the Black Sea to the Arctic and encompass the United Kingdom, Norway, Iceland and
74 Turkey in addition to the 22 coastal nations of the European Union. We apply an approach that incorporates
75 fine-scale geographical differences in the climate hazard of fish and shellfish populations and then assess
76 the climate-risk of both European fishing fleets and coastal regions in two separate CRAs. Since both CRAs
77 are based on the same underlying climate hazard, these analyses can be combined to compare the relative
78 importance of climate-hazard to fleets and coastal communities within a country.

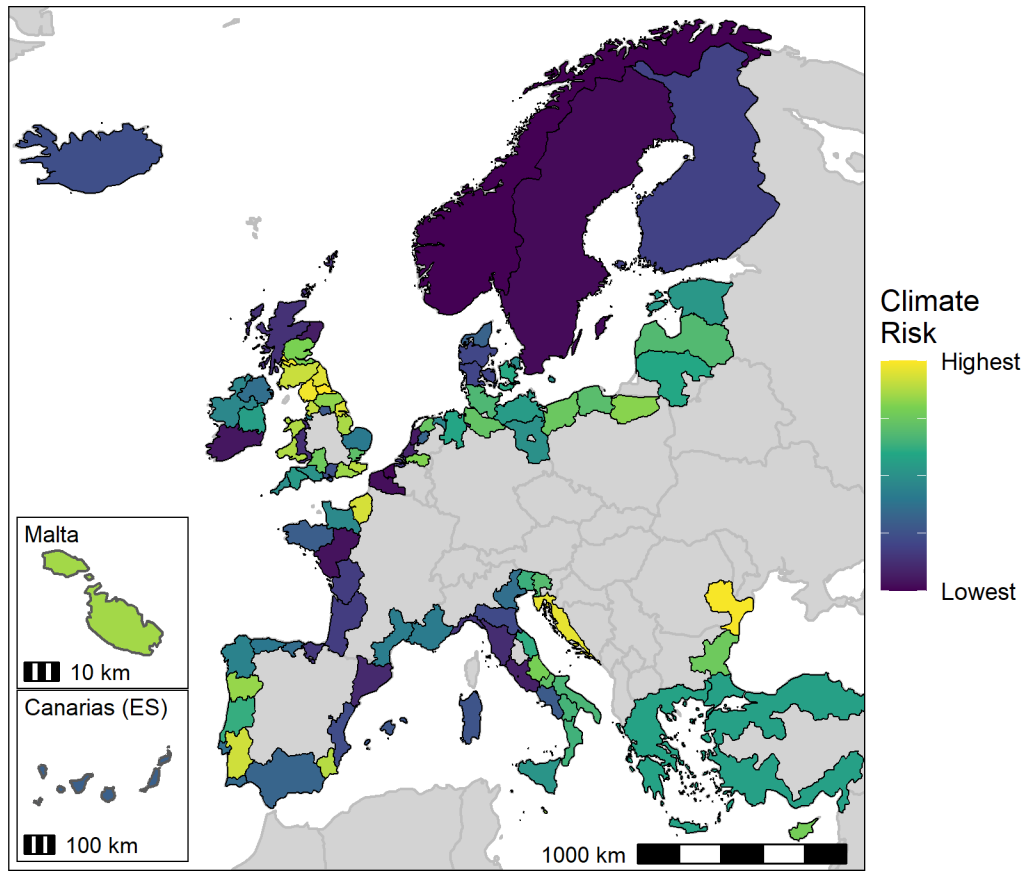
79 Our index of climate hazard is derived from the biological traits of the species being harvested, together
80 with modelled distribution data. Species trait data were gathered for 157 fish and shellfish species harvested
81 in European waters, representing 90.3% of the total value of landings in Europe and at least 78% (and
82 typically more than 90%) of national value. We accounted for the expected large differences in climate
83 hazard throughout a species range (i.e. from the cold to warm edges of the distribution) by focusing on
84 “populations” (i.e. a single species in a single FAO subarea). Population-level climate hazards were then
85 defined based on the thermal-safety margin (TSM) between the temperature in that subregion and the upper
86 thermal tolerance of the species^{26,27}. Climate hazards could then be calculated for 556 significant

87 “populations” in 23 FAO subareas, based on the TSM of the population and the inherent traits of the
88 species^{15,28,29}.

89 We then calculated the climate risk for 105 coastal regions across 26 countries in the European region
90 (Figure 1). Population-level climate hazards of fish were integrated to regions, weighted by the relative
91 value of landings in that region. We defined exposure metrics based on the diversity^{30,31} of these landings,
92 and vulnerability based on regional socio-economic metrics⁶. We focused our analysis on coastal regions,
93 as these are the communities most directly dependent on the ocean: regions far from the sea but within a
94 coastal nation were explicitly excluded (e.g. Bavaria in Germany).

95 The analysis reveals appreciable variation in the climate risk within the European region and even within a
96 single country. In the United Kingdom, for example, climate risk is greatest in the north of England, while
97 Scotland and the south of England show the least risk. Indeed, six of the 10 regions with the highest climate
98 risk, including the overall top region (Tees Valley & Durham), are part of the UK (Table S8). These results
99 reflect high hazard scores for the species landed in these regions, together with high vulnerability due to
100 low GDP per capita in some of these regions.

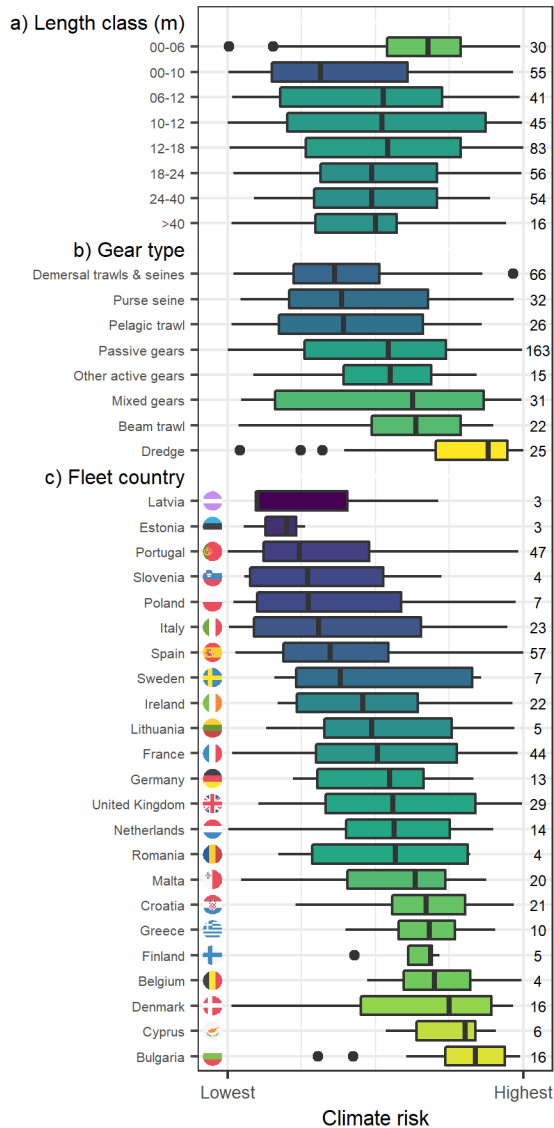
101 Larger-scale patterns are also apparent. South-east Europe stands out with consistently high climate risk,
102 with coastal Romania and Croatia in the top five. Both countries have high vulnerability scores due to low
103 GDP per capita of their coastal communities, and high exposure scores due to fisheries that target only a
104 few species (e.g. the value of Romania’s fisheries is more than 70% veined rapa whelk, *Rapana venosa*).
105 Many northern European countries, including Belgium, the Netherlands and Scandinavian nations have
106 relatively low climate risks due to their wealth (high GDP per capita), diverse fisheries and the relatively
107 low climate hazard of the fish populations landed.



108
109 **Figure 1 Map of the regional climate risk.** Colour scale is linear in the value of the corresponding score, but is presented without
110 values, as they have little direct meaning. National-level borders are shown for reference. Insets at bottom-left show small regions.

111 The risks associated with climate change will also be felt by the individual fishing vessels and fleets whose
112 fishers' livelihoods depend on the ocean. We therefore performed a second CRA to examine the climate
113 risk of European fishing fleets. As the basis for this analysis, we followed the EU definition of a "fleet
114 segment" based on the size classes of the vessels, the country of registration, the gear being used and the
115 geographical region being fished (Atlantic or Mediterranean)²³. We scaled fish population-level climate
116 hazards up to the fleet segment level based on the composition of landings by value of that fleet, while we
117 based exposure on the diversity and dominance of landings and vulnerability on the profitability of the fleet.
118 Coverage of our analysis at this fleet segment level was poorer than at the national level: nevertheless, we
119 still cover 75% or more of total fishery catch value for more than 70% of the 380 fleet segments within the
120 EU and UK.

121 The smallest class of vessels (0-6m) had an appreciably higher climate risk than all other size classes (Figure
122 2a). For the most part, these fleets operated in the Mediterranean region, particularly in Croatia, Bulgaria,
123 France, Malta and Greece (Table S9). This result reflects, in part, the higher climate risk of stocks in this
124 area, but is also driven by the poor profitability (and therefore higher vulnerability) of these fleets. On the
125 other hand, the high catch diversity of these fleets reduces exposure and helps to reduce the net climate-
126 risk.



127

128 **Figure 2 Climate risk of European fleet segments.** The climate risk across 380 fleet segments is plotted as a function of a) the
 129 size range of the vessels (m), b) the gear type employed (sorted by median risk) and c) the country of origin of the fleet (sorted by
 130 median risk). Risk is represented on a linear scale from highest to lowest: the absolute values are not shown, as they have little
 131 direct meaning. The distribution of risk is shown as a boxplot, where the vertical line is the median, the box corresponds to the
 132 interquartile range (IQR), and the whiskers cover all points less than 1.5 times the IQR from the box. Outliers are plotted as points.
 133 Boxes are coloured based on the median climate risk for that category. The number of fleet segments in each class is shown at right.
 134 Note that the STECF definitions of small length classes (less than 12m) vary between countries and therefore have a degree of
 135 overlap. STECF gear codes are aggregated to our “Gear Types” to ensure comparability between Atlantic and Mediterranean
 136 fisheries (Table S4).

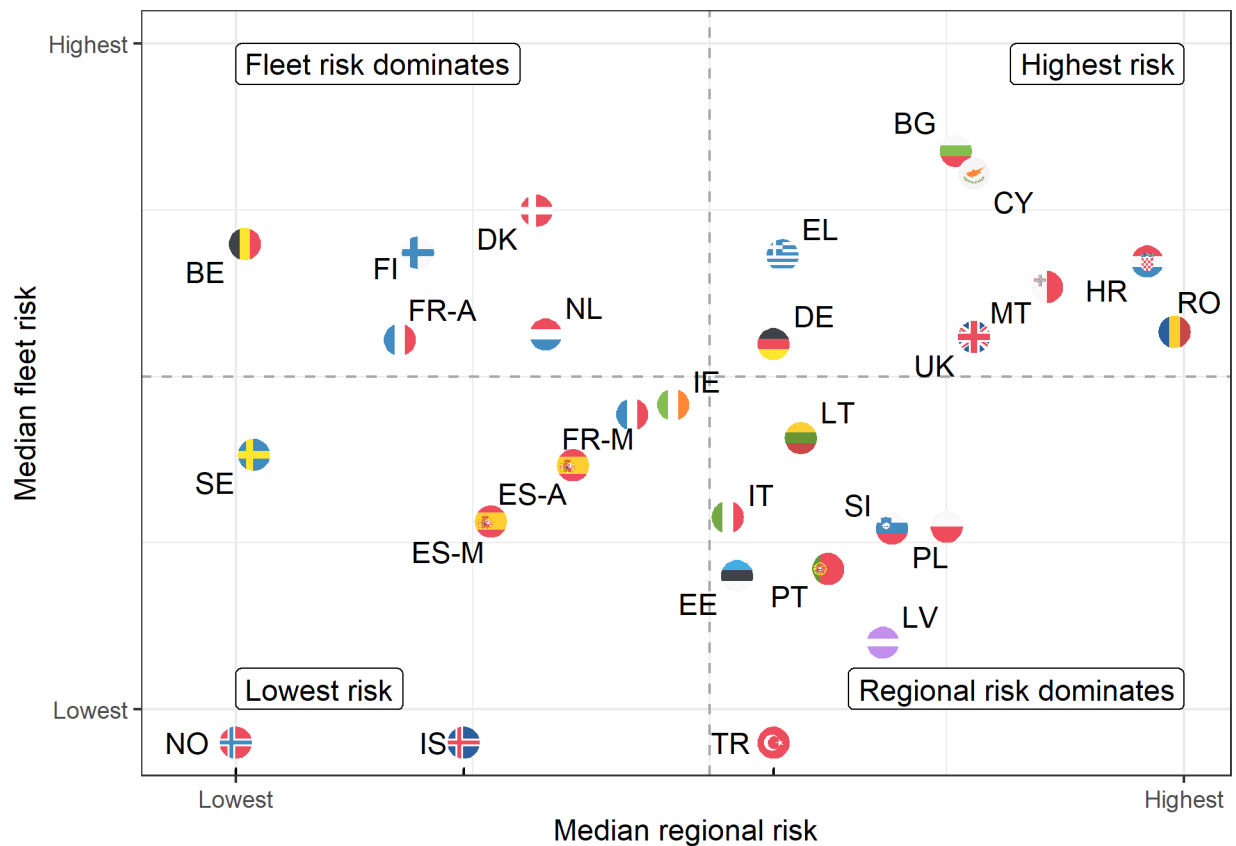
137 Systematic differences in climate-risk were revealed among gear types (Figure 2b), with dredgers having
 138 the highest climate risk. These fleets generally target populations with high climate hazards and have low
 139 species diversity in their catches (giving high exposure): good profitability, on the other hand, lowers their
 140 vulnerability and somewhat reduces overall risk (Table S9). Fleets using pelagic and demersal trawls
 141 together with purse seines have the lowest climate risks, primarily due to the low hazard associated with
 142 the species on which they fish.

143 The strongest differentiation in climate risk amongst fleet segments, however, is at the national level (Figure
144 2c). A clear cluster of high climate-risk fleet segments can be seen in south-east Europe, particularly in
145 Croatia, Greece, Bulgaria, Cyprus and Romania (Figure S1). The risk profiles underlying each of these
146 cases, however, are quite different, emphasising the need to understand the components in detail. Greek
147 and Cypriot fleets have high climate risks due to poor profitability and, therefore, high vulnerability, while
148 Bulgarian and Romanian fleets active in the Black Sea have extremely low catch diversities, giving them
149 unusually high exposures (Table S9). It is also important to note that there is substantial variation among
150 fleets within a country. For example, two of the five most at-risk fleets (including the most at risk) are
151 Spanish (Table S9), even though the national level median for Spain is amongst the lowest in Europe. A
152 detailed examination of the individual elements of the risk-profile is therefore critical to understanding the
153 underlying factors responsible for these results.

154 A strength of the analysis performed here is that the results of the fleet and region CRAs can be directly
155 compared. While the fleets and regions are all exposed to the same base set of hazards, the relative
156 importance of each fish or shellfish population (and therefore hazard) differs. Each region and fleet also
157 has its own intrinsic exposure and vulnerability profiles, further modulating the climate risk. However, as
158 the base set of hazards that our analysis starts from is the same in both CRAs, a direct comparison of the
159 two cases is possible, allowing the relative importance of climate risk to fleets and regions to be gauged.

160 Systematic differences in risk between fleets and regions can be seen among European countries (Figure 3)
161 with several characteristic types of responses apparent. Countries in south-eastern Europe, together with
162 the United Kingdom, have consistently the highest risk across both fleets and regions. The regional climate
163 risk scores of states on the south coast of the Baltic (Latvia, Lithuania, Estonia and Poland) are higher than
164 their fleet level scores, while the high fleet-risk of NW European states is moderated by their relative
165 affluence and therefore low regional risk. Spain and Sweden generally have low climate risks in both
166 sectors.

167



168

169 **Figure 3 Comparison of the median fleet- and region-based risks for European countries.** Labels indicate the country code.
 170 In addition, France (FR) and Spain (ES) are split into their Atlantic (-A suffix) and Mediterranean (-M suffix) seabords. As the
 171 fleet-segment analysis only covers fleets from the EU and UK, no data are available for Turkey, Norway and Iceland: their regional
 172 risk results are plotted in the horizontal margin. Dashed lines divide the coordinate system into quarters. Country codes: BE:
 173 Belgium. BG: Bulgaria. CY: Cyprus. DE: Germany. DK: Denmark. EE: Estonia. EL: Greece. ES: Spain. FI: Finland. FR: France.
 174 HR: Croatia. IE: Ireland. IS: Iceland. IT: Italy. LT: Lithuania. LV: Latvia. MT: Malta. NL: Netherlands. NO: Norway. PL: Poland.
 175 PT: Portugal. RO: Romania. SE: Sweden. SI: Slovenia. TR: Turkey. UK: United Kingdom.

176 Our analysis highlights the wide variety of challenges facing European countries and regions with adapting
 177 their fisheries sectors to a changing climate. In some cases, such as in the southern-Baltic states, a focus on
 178 strengthening the resilience of local regions and communities would be of most benefit e.g. by creating
 179 alternative employment opportunities. In other regions, fleet risks dominate, and therefore increasing the
 180 efficiency and diversity of the fleet would appear to be a priority. However, some areas, such as the UK
 181 and south-east Europe appear to require both types of intervention, and therefore present the greatest
 182 adaptation challenges. There is, however, no “one-size-fits-all” solution that can be applied across all
 183 European waters or even, in some cases, across a country (e.g. the UK): climate adaptation plans therefore
 184 need to be tailored to these realities.

185 Climate risk and vulnerability analyses do, however, have a key role to play in shaping the development of
 186 such plans. By increasing awareness of the elements that contribute to a fleet or region’s risk⁶, CVAs and
 187 CRAs can help prioritise adaptation actions to mitigate this risk³² and thereby maximise the effectiveness
 188 of limited resources. Previous socio-economic linked analyses have focused on adaptive capacity (in the
 189 CVA framework) as a focal point for action^{6,12}. However, the diversity of European risk profiles found here
 190 highlights the need and potential for adaptation actions across all components of the risk portfolio.

191 First and foremost of these actions is ensuring sustainable management of the living marine resources upon
192 which the sector rests. The future impact of over-exploitation of these resources can be more important than
193 that stemming from climate change, particularly in the heavily fished North Atlantic region³³. Maintaining
194 these stocks at a higher abundance leads to increases in genetic diversity, meta-population complexity, and
195 age structure, all of which make stocks more resilient to the challenges of a changing environment^{34,35}. The
196 ensuing increase in productivity and incomes will simultaneously benefit both fishing fleets and regions,
197 generating a “win-win” effect³⁶. Fisheries scientists already have many of the tools necessary to ensure that
198 management systems are robust to climate change and climate variability³⁷, while emerging tools, such as
199 seasonal-to-decadal marine ecological forecasts³⁸, can potentially provide the basis for further coping
200 strategies³⁹. A focus on sustainable management will therefore reduce the climate risk that both fleets and
201 regions are exposed to.

202 Diversification of the sector is a second key tool to reduce climate risk. Fishing fleets and regions relying
203 on only a few species have an elevated risk of climate impacts: increasing this spread reduces exposure and
204 therefore buffers fleets and communities against climate risk^{31,40,41}. Diversification of catches and landings
205 can take place autonomously as fishers respond to changes in the abundance and distribution of the fish
206 they depend on^{32,37}. For example, changes in the distribution of fish species in surrounding waters^{42–44} have
207 led to the development of new fisheries in the UK for squid, seabass and red mullet, amongst others⁴⁵. There
208 are, however, barriers to diversification^{31,41}, most notably in the form of the variety of resources available:
209 the limited catch diversity and therefore high exposure of fleets and regions adjoining the Black Sea, for
210 example, arises at least in part from the relatively low biodiversity of this region. The ability to diversify
211 may also be limited by existing quota agreements⁴⁶, a particularly challenging issue under the “relative
212 stability” agreements of the EU Common Fisheries Policy.

213 Thirdly, governance can coordinate and drive actions to reduce the vulnerability of fleets and regions.
214 Investments and support for developing new, and switching between, fishing, storage, transport and
215 processing technologies can increase the efficiency of fleet operations and, therefore, reduce
216 vulnerability^{18,37,47}. Increasing regional development, including employment opportunities outside the
217 fisheries sector, reduces regional vulnerability and risk^{6,48}. Furthermore, both fishing fleets and regions can
218 also potentially benefit from governance-led actions that increase the flexibility, ability to learn, social
219 organisation and the power and freedom to respond to challenges⁴⁹. Regional, national and European
220 governments therefore have a critical role to play in adapting fisheries and ocean-dependent regions to the
221 risks presented by climate change.

222 Several key caveats of these results need to be highlighted. Our analysis focused solely on the sensitivity
223 to ocean warming, ignoring other climate-driven processes, such as ocean acidification, deoxygenation, and
224 changes in storms or circulation patterns^{5,30} that, while important, are viewed as second order effects. Spatial
225 differences in warming across European regional seas were also not accounted for here but these differences
226 (1 to 2°C by 2050) are much less than the variability in thermal safety margins between populations (range
227 15°C). The treatment of uncertainty in CVAs and CRAs varies greatly between studies^{15,50} but in such a
228 semi-quantitative analysis the choice of metrics is usually the most important aspect⁵¹. We believe that such
229 “structural uncertainty”⁵² is best addressed by focusing on a limited, but transparent and readily
230 interpretable set of indicators, rather than by quantifying uncertainties or increasing complexity. Finally,
231 while we have considered European fisheries on fish stocks in the Mediterranean Sea, we have not
232 incorporated coastal communities in African countries that also fish on these same stocks. The relatively
233 low GDP per capita of these communities suggests that they would have correspondingly high regional
234 vulnerabilities and therefore climate risks but it is not possible to draw robust conclusions in the absence of
235 appropriate data sets: the population-level hazards generated here (Table S7) could be readily applied to
236 aid such analyses in the future.

237 This study has shown that even though climate-risk to European countries is, on average, moderate
238 compared to many other countries across the globe^{6,21}, major differences exist across the European region.
239 This is not only true for coastal regions, where especially south-east European and various UK coastal
240 regions were found to be subject to the greatest climate-risk, but also for the different European fishing
241 fleets, with (small-sized) fleet segments in south-east Europe at greatest risk. This corroborates with fine-
242 scale spatial differences among fishing communities documented in eastern North America^{13,53} and the
243 Caribbean^{30,54}, each requiring very different adaptation actions. Our detailed analyses allow a distinction
244 between climate hazard, exposure and vulnerability as key sources of climate-risk, and highlight where
245 (and which) adaptation measures can have greatest impact in increasing resilience.

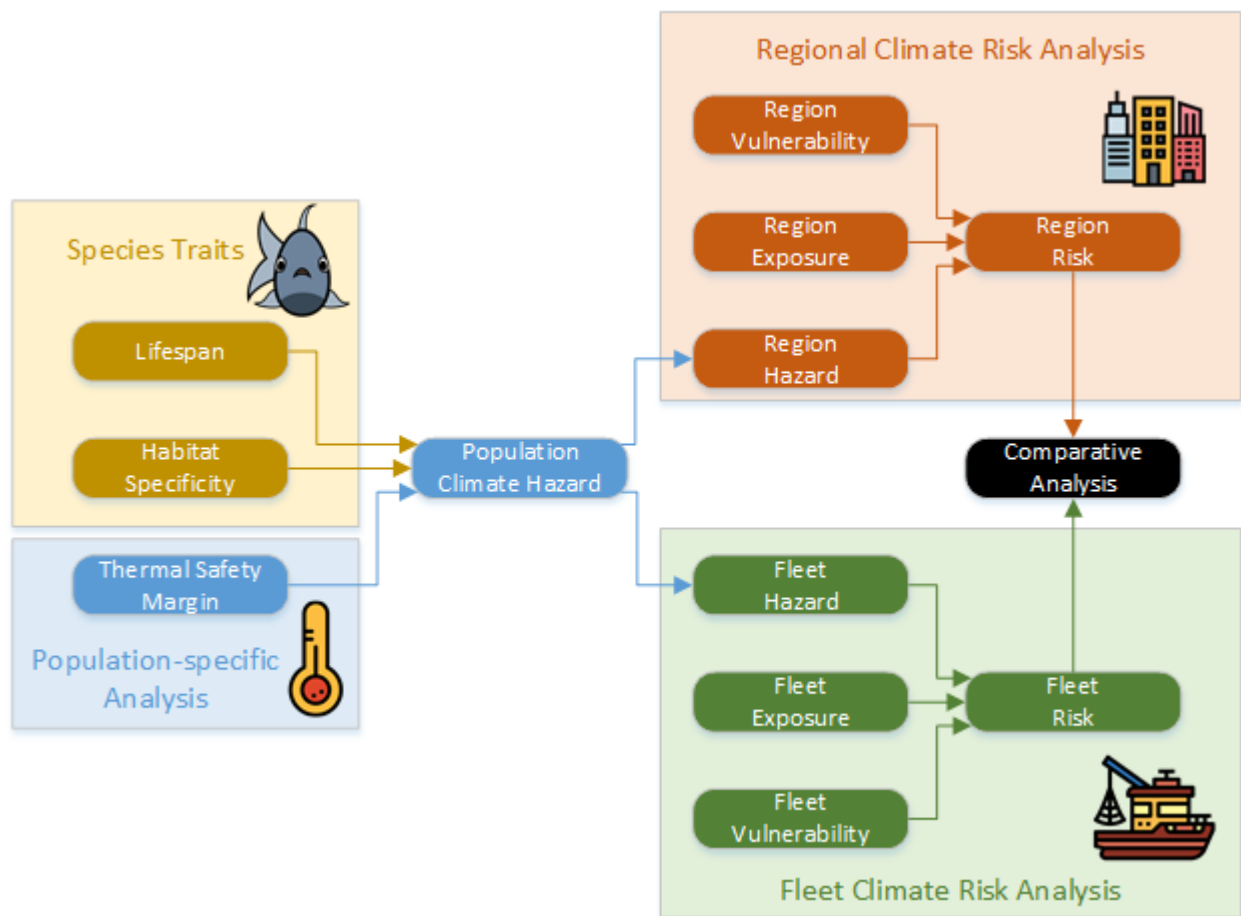
246 **Acknowledgements**

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248 programme under grant agreement No 678193 (CERES – Climate change and European Aquatic
249 Resources). The results generated by this analysis can be explored using an online tool available at
250 https://markpayne.shinyapps.io/CERES_climate_risk/ Source code is available at
251 https://github.com/markpayneatwork/CERES_vulnerability. "Fishing Boat", "Urban" and "Thermometer"
252 icons in Figure 4 by smalllikeart from www.flaticon.com.

253 Methods

254 General approach

255 We have applied an integrated approach to a climate risk assessment (CRA) across the European fisheries
256 sector. The CRA has three major components (Figure 4; Figure S2). The first and most fundamental of
257 these is the *population hazard* component, where the hazard associated with climate-change impacts on
258 both species and individual fish populations is quantified. We then use these hazard metrics as inputs into
259 two parallel climate-risk assessments focussing on coastal regions and fishing fleets in turn. In each of these
260 cases, the population hazard is integrated up to the region or fleet level based on information about the
261 relative importance of each fish population to that unit, to form the region- or fleet-specific hazards. These
262 hazard data are then complemented with region- and fleet-focused *exposure* and *vulnerability* metrics to
263 produce a *climate-risk* for each.



264

265 **Figure 4 Schematic diagram** illustrating the approach used here to estimate climate risk in European fishery-dependent
266 communities and fishing fleets. Species traits and population specific analyses of the thermal safety margin are combined to give
267 a population-specific climate hazard. This hazard then forms the basis for the regional and fleet level CRAs, based on the
268 combination of hazard, exposure and vulnerability. Finally, the regional and fleet risks are combined again into a comparative
269 analysis. A detailed flow diagram is presented in the supplementary material (Figure S2).

270 **Scope and Data Sources**

271 We aimed to assess the climate risk for the European marine fisheries sector, including all 22 EU countries
272 with marine borders, the United Kingdom, Norway, Iceland and Turkey. We based our analysis primarily
273 on catch data from FAO Areas 21, 27, 34 and 37 held in the EUROSTAT database (Table S1), excluding
274 distant water fleets. While this database covers more than 1200 species, many of these are economically
275 minor. We therefore aimed to cover the largest 90% of the value of the marine fish and shellfish sector in
276 each country and across Europe as a whole. Two species predominately inhabiting freshwater, European
277 perch (*Perca fluviatilis*) and pike-perch (*Sander lucioperca*), were removed from the database. Alternative
278 (or misspelled) scientific names were corrected where we could identify these (following World Register
279 of Marine Species, WoRMS) (Table S3).

280 Regional analyses were performed for European coastal regions based on NUTS2 statistical units. Sub-
281 national indicators of landings composition were derived from monthly harbour-level “first-sales” data
282 from the EU Market Observatory for Fisheries and Aquaculture (EUMOFA) (Table S1). In cases where
283 this data covered more than one NUTS2 unit within a country (10 countries), the harbour data was
284 aggregated up to NUTS2 units based on the geographical coordinates of the harbours. Where EUMOFA
285 data coverage was insufficient, the coastal NUTS2 units of that country were merged into one “region”
286 (Table S5) and EUROSTAT national landings data were assigned to it (Table S1). Socio-economic data for
287 the NUTS2 units was also obtained from EUROSTAT and integrated up to our “regions”, if relevant.

288 The Annual Economic Report (AER) provided by the EU Scientific, Technical and Economic Committee
289 for Fisheries (STECF)²³ formed the basis of the fishing fleet analysis (Table S1). This dataset has the
290 advantage of providing a single coherent dataset for fleet segments (the combination of fishing technique
291 and a vessel length category) across all of the European Union and United Kingdom: however, it does not
292 include data on fleets from Norway, Iceland or Turkey, and in the absence of comparable datasets, these
293 countries were not included in this part of the analysis.

294 All data was averaged over the period 2010-2018, where available.

295 **Hazard Metrics**

296 The *hazard* dimension of our CRA measures the strength and severity of climate change on the unit of
297 interest: in this case, fish populations in European waters. Many previous CVAs and CRAs do not
298 distinguish between the positive and negative effects of climate change, and simply highlight elements of
299 their study system that will change, making interpretation difficult. In contrast, we made a conscious
300 decision to focus on “negative” impacts in order to have an unambiguous interpretation. We consider the
301 hazard due to climate change impacts on living marine resources as being the combination of both species-
302 specific and population-specific processes, as follows.

303 **Species-specific processes**

304 A trait-based approach was employed to characterise the hazard of a species to climate change. Such an
305 approach is well established in climate-risk and vulnerability analyses^{15,17,28}, due to its ability to draw on
306 general understanding of the response of species to climate change. Trait data was collated from previously
307 published databases⁵⁵⁻⁵⁸ and complemented with data from Fishbase⁵⁹ and Sealifebase⁶⁰ (accessed April-
308 July 2019) (Table S1). Of the original set of species from EUROSTAT, 24 taxa were only at the genus
309 level, and appropriate trait sets were therefore identified based on ‘exemplar species’: in some cases
310 different exemplar species were used for the North Atlantic (FAO Area 27) and Mediterranean regions

311 (FAO Area 37) (Table S2). Barnacles (*Pollicipes pollicipes*) and solen razor clams (*Solen* spp.) were also
 312 removed owing to a lack of biological traits data and difficulties identifying suitable exemplar species.

313 Trait selection aimed to avoid double-counting information due to inclusion of correlated traits, a commonly
 314 overlooked issue⁵⁶ that impacts many published analyses^{15,28,29,33}. For example, smaller fish are typically
 315 planktivorous, live shorter and grow faster, giving a high correlation between maximum length, lifespan,
 316 growth rates and trophic level. Lifespan is the most commonly available of these metrics and was therefore
 317 chosen as an exemplar for this set of traits. Shorter lifespans are associated with seasonal and variable
 318 environments⁵⁶, implying robustness to change and variability, paralleling the approach used in other
 319 studies^{15,28,29,33}.

320 A “habitat specificity” hazard metric was also developed. Species with spatially restricted habitat
 321 requirements during part or all of their life-history are recognised as being more sensitive to disruption^{61,62}.
 322 In addition, mobile species have the ability to move rapidly to avoid unfavourable conditions in a way that
 323 sedentary species do not, and are therefore at less climate risk³⁰. Traits defining the mobility, and vertical
 324 and horizontal habitats were therefore collated into a single “habitat-specificity score” (Table 1). The final
 325 set of traits is included as supplementary material (Table S6).

326 **Table 1 Combination of mobility, vertical and horizontal habitat traits to generate a habitat specificity score.** Definitions of
 327 traits follow the scheme of Engelhard *et al*⁵⁵.

<i>Habitat Score</i>	<i>Mobility</i>	<i>Vertical habitat</i>	<i>Horizontal habitat</i>
Low (0.00)	Highly migratory species	Any	Any
	Mobile	Any	Oceanic
	Mobile	Bathymedersal Mesopelagic	Slope
Medium (0.33)	Mobile Unknown	Benthopelagic Demersal Pelagic Epipelagic	Slope Shelf Outer shelf
	Unknown	Bathymedersal	Slope
	Mobile	Bathymedersal	Outer shelf
High (0.67)	Mobile (catadromous/anadromous)	Pelagic	Any
	Mobile	Demersal	Inner shelf
	Mobile	Benthopelagic	Coastal
Very high (1.00)	Sedentary	Any	Any
	Mobile	Reef-associated	Any

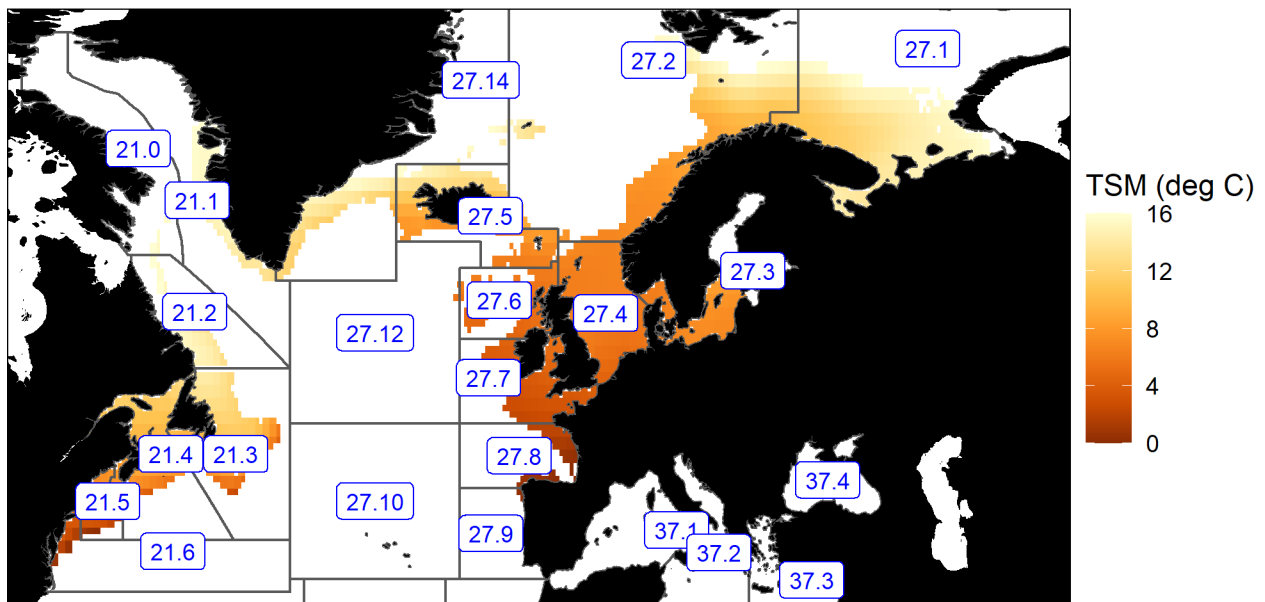
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329 **Population-specific processes**

330 The stress a fish population experiences as the ocean warms depends on the amount of warming, a
 331 commonly employed metric of exposure in CVAs^{6,15}. However, the physiological context of this warming
 332 is also critical but often overlooked. For example, cod (*Gadus morhua*) in the North Sea are close to their
 333 upper thermal limit, and will therefore experience negative impacts of warming, while cod in the Barents
 334 Sea are far from this limit and will experience little or no negative effects of the same amount of warming⁶³.
 335 Such a spatial and physiological context of warming is often overlooked in many CRAs and CVAs, yet is
 336 critical to differentiate the climate hazard between different populations of the same species.

337 We resolve this problem in two ways. We first perform our analysis at the “population” level, defined as
338 the combination of species and FAO subarea e.g., cod in subarea 27.4 (North Sea), similar to the approach
339 used to manage many European fish stocks: populations comprising less than 5% of the total catch of the
340 species were excluded from the analysis. Note that we explicitly avoid the use of the term “stock” to refer
341 to this unit of analysis, as this has clear implications in fisheries management but is not entirely the same
342 as our definition “population”. Secondly, we place the degree of warming experienced by these populations
343 in a physiological context using thermal-safety margins (TSM)^{26,27,64,65}. TSM is defined as the difference
344 between the maximum temperature that the species can sustain and the temperature of the environment:
345 high TSMs indicate a high capacity to tolerate warming. Population-specific TSMs therefore permit a fine-
346 grained measure of the hazard from warming.

347 We derived population-specific TSM metrics from the habitat models, parameters and maps provided by
348 Aquamaps www.aquamaps.org⁶⁶ (Table S1). We downloaded “native distribution maps” from the
349 Aquamaps website for the species selected above: where multiple maps were available, choice was guided
350 by the internal map quality ranking system. For the invasive species purple whelk (*Rapana venosa*),
351 originally from waters around Japan, Korea and China but now supporting a large fishery in the Black Sea,
352 the “Suitable Habitat map” was used. From each map we used the “90th percentile” parameter for the
353 temperature response for each species as an estimate of its upper thermal tolerance. Temperatures in a
354 subarea were based on the data underpinning the Aquamaps model (NOAA NCEP Climatology, 1982-
355 1999)⁶⁶, ensuring congruence between the tolerance parameters and the temperature data. Sea-surface or -
356 bottom temperature data, as appropriate for the species and used in the Aquamap, were masked using the
357 habitat model to eliminate unsuitable habitat for each individual species (Figure 5). Population-specific
358 TSM was then calculated as the median difference between the species’ “90th percentile” parameter and
359 temperature across all valid pixels in that subarea.



360

361 **Figure 5 Example of the use of Aquamaps to calculate TSM metrics.** Atlantic cod (*Gadus morhua*) as an example.
362 Environmental data and species thermal tolerance data from Aquamaps are used to calculate the thermal safety margin (TSM) for
363 this species (coloured pixels) and masked using the habitat model to limit data to modelled regions of occurrence. Median TSM
364 values are then calculated within each FAO subarea defining a population (grey polygons, blue labels).

365 **Population-level hazard**

366 Hazard metrics were combined based on their relative ranking for each population. The metrics employed
367 here have little quantitative meaning: rather, it is their relative values that are important. Each metric was
368 therefore converted to a rank percentile, and then combined using a weighted average, with a weight of 0.5
369 for the population TSM (high TSMs give a low hazard), 0.25 for the species' lifespan (shorter-lifespans
370 give a low hazard) and 0.25 for the species' habitat-specificity (low specificity gives a low hazard).

371 Population-level hazard scores were integrated up to fishing fleet and regional levels. In the case of the fleet
372 analysis, this was based on the relative composition (by value) of the populations that each fleet fishes on,
373 while in the case of the regional analysis it was based on the composition (by value) of landings in that
374 region (Figure 4, Figure S2).

375 **Exposure metrics**

376 We define *exposure* as an indicator of how sensitive a community or fishing fleet is to changes in the fish
377 populations it is dependent on. Fleets or fishing communities have lower exposure (higher resilience) if
378 they catch a wide range of different fish species, rather than concentrating on a specific resource^{30,31,41}. If
379 one species is reduced or lost due to the effects of climate change, the impact of that loss is relatively less
380 severe for fleets and communities that are dependent on a broad portfolio of species. We therefore defined
381 our exposure metrics following this logic, using two different metrics to characterise diversity of catch or
382 landings: i) the Shannon diversity index, one of the most commonly used diversity indices in ecology and
383 ii) Simpson's dominance index, a statistic that emphasizes the relative abundance of the most common
384 species in the sample³⁰.

385 For European regions, exposure metrics were based on the value of landings data from EUMOFA and
386 EUROSTAT (Table S1; Figure S2). While EUROSTAT data is species resolved, EUMOFA data is
387 organised in approximately 100 "main commercial species" (MCS) groupings: we therefore harmonised
388 the two datasets by aggregating EUROSTAT data to the MCS groupings based on correlation keys provided
389 by EUMOFA. The Shannon and Simpson metrics were then calculated to estimate the diversity of MCS
390 groups.

391 For fleet segments, the value of landings is available by species code from the STECF Annual Economic
392 Report²³. The two diversity indices could therefore be calculated directly to quantify the diversity of species.

393 In both cases, the exposure index was produced as a composite index of the two indices described above
394 by averaging the percentile ranks.

395 **Vulnerability metrics**

396 *Vulnerability* in this setting refers to the resilience of the analysis unit (either a region or a fleet) and its
397 ability to mitigate the hazard via adaptation.

398 The regional vulnerability metric was based on the gross-domestic product per capita of the region, as
399 calculated from EUROSTAT data at the NUTS2 level (Table S1). Regions with high GDP per capita were
400 viewed as having a high adaptive capacity and therefore low vulnerability. Regional vulnerability was
401 calculated as the percentile rank of this statistic.

402 Fleet segment vulnerability was based on the net profit margin (NPM). This is a standard economic metric,
403 defined as net profit (i.e. revenue minus fixed and variable costs and opportunity cost) divided by the total
404 revenue: it therefore represents how much of the total income generated by the fleet is profit²³. NPM has
405 the feature of taking into account many of the different factors that influence the profitability of the fleet,

406 and is also scale independent (as profitability is divided by the revenue), allowing comparison of both large
407 and small segments. NPM was calculated for each fleet segment based on economic data from the STECF
408 Annual Economic Report²³ (Table S1), and the vulnerability score generated based on percentile rank. Fleet
409 segments with high profitability were viewed as being less vulnerable to the effects of climate change, as
410 they could absorb the anticipated loss associated with any potential negative change in their target species.

411 **Climate risk metrics**

412 For each of the geographic regions, and for each of the fleet segments, the overall *climate risk* was
413 calculated as the unweighted mean of the hazard, exposure and vulnerability percentile ranks.

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