Title: Intra and inter-annual climatic conditions have stronger effect than grazing intensity on

root growth of permanent grasslands

Running head: Root production in grazed grasslands

List of authors: Catherine Picon-Cochard[1*], Nathalie Vassal[1,2], Raphaël Martin[1], Damien

Herfurth[1,2], Priscilla Note[1,3], Frédérique Louault[1]

Institute of origin:

1. Université Clermont Auvergne, INRAE, VetAgro Sup, UREP, 63000 Clermont-
Ferrand, France
2. Université Clermont Auvergne, VetAgro Sup, 63370 Lempdes, France
3. INRAE, UE1414, Herbipôle, 63122 Saint-Genès-Champanelle, France
14

15   ***: Corresponding author**: Catherine Picon-Cochard, Tel: (+)33 443 761 615; email:

16   catherine.picon-cochard@inrae.fr

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Abstract

Background and Aims. Understanding how direct and indirect changes in climatic conditions, management, and species composition affect root production and root traits is of prime importance for the delivery of carbon sequestration services of grasslands. This study considers the effects of climatic variability and gradients of herbage utilisation by grazing on root production over the course of two years. The root and leaf traits of the plant communities were determined to detect their capacity to predict above- and below-ground net primary production, ANPP and BNPP, respectively. **Methods**. A long-term field experiment was used to compare the effects of abandonment and low (Ca-) and high (Ca+) grazing intensities (resulting in mean residual plant heights of 15.2 cm and 7.7 cm, respectively) induced by grazing rotations on upland fertile grasslands after 10 years of treatment application. Ingrowth cores and exclusion cages were used to measure, respectively, the root and shoot mass production several times each year and at an annual scale. The root and leaf traits of the communities were measured near the vegetation's peak growing season. **Results**. We observed strong seasonal root production across treatments in both a wet and a dry year, but the response to grazing intensity was hardly observable within growing seasons. In the abandonment treatment, the spring and autumn root growth peaks were delayed by approximately one month compared to the two cattle treatments, possibly due to a late plant canopy green-up induced by lower soil temperatures and an accumulation of litter. The BNPP was slightly lower in the abandonment treatment compared to the cattle treatments only during the dry year, whereas a decline of the ANPP in the abandonment treatment compared to the Ca+ treatment was observed during the wet year. In response to drought, which occurred during the second year, the root-to-shoot biomass ratio was stable in the cattle treatments but declined in the abandonment treatment. The higher allocation to root mass could benefit plant communities under drier conditions. **Conclusions**. Rotational grazing pressures and climatic condition variabilities had limited

effects on root growth seasonality, although drought had stronger effects on the BNPP than on

the ANPP. The stability of the root-to-shoot biomass ratio during the dry year evidenced a

higher resistance to drought by grazed versus abandoned grassland communities.

Introduction

Permanent grasslands provide many services that connect to human activities through livestock

products but also contribute to regulate greenhouse gas emissions because their soils

accumulate large amounts of carbon (C) in organic matter fractions. Root activity (growth,

exudation, and turnover) contributes to C and nitrogen (N) inputs and to absorb both nutrients

and water, which are essential to fix atmospheric carbon dioxide (CO_2) and produce plant

biomass. The intensification of management practices may affect these services as well as

climate variability (Conant et al., 2001; Jones and Donnelly, 2004; Soussana and Duru, 2007).

Thus, improving our understanding of grassland roots dynamics under different management

and climatic conditions may help to identify management options to maintain the forage

production and C sequestration abilities of this ecosystem and thus its sustainability.

Different management practices modify forage production and the amount of soil C and N

through both the direct effects of defoliation, fertilisation, or returns of excreta to soil on root

growth and soil abiotic factors and the indirect effects of species composition changes (Bardgett

and Wardle, 2003; Dawson et al., 2000; Soussana et al., 2004). In mown grasslands, the root

mass production is generally lower when grasses are frequently mown and fertilised (Leuschner

et al., 2013; Picon-Cochard et al., 2009). This result may be explained by changes in the root-

to-shoot allocation with an increase of above-ground growth to maximise light capture. The

complexity of these phenomena in grazed grasslands is greater than in mown systems owing to

animals' selective defoliation of plant species and because returns to soil are spatially

heterogeneous (Rossignol et al., 2011). In addition, the level of soil fertility may buffer the

degree of root response to defoliation in grazed grasslands as plants exhibit specific responses

to defoliation in fertile and unfertile grasslands (Duru et al., 1998). Overall, the complexity of

phenomena being direct and indirect effects on plants could explain why no clear trend has been

found for the effects of grazing on above- and below-ground production (e.g., refer to the

syntheses of Milchunas and Lauenroth [1993] and McSherry and Ritchie [2013]), although two

meta-analyses have emphasised the negative effect of grazing intensity on above- and below-

ground C stocks compared to ungrazed systems (Zhou et al., 2017; Li et al., 2018). Furthermore,

repeated defoliations induced by grazing or mowing grasslands can simultaneously increase

soil temperatures and moisture (Moretto et al., 2001; Pineiro et al., 2010; Smith et al., 2014).

Soil moisture can also be modified by high stocking rates through changes in the soil bulk

density due to soil compaction and changes in the leaf area index after defoliation (Pineiro et

al., 2010). These direct effects of grazing on soil abiotic factors should affect the root growth

of grazed grasslands, although these phenomena have not been well documented in field

conditions.

Species composition change induced by management is also an important determinant of

above- and below-ground responses in grazed grasslands. Intensive practices (e.g., high grazing

intensity or fertilisation) generally favour the development of fast-growing species (plant

exploitative strategy), whereas the opposite extensive practices (e.g., low grazing intensity or

the absence of fertilisation) favour slow-growing species (plant conservative strategy; Klumpp

et al., 2009; Louault et al., 2005; Soussana and Lemaire, 2014; Wardle et al., 2004). The root-

to-shoot biomass allocation, as well as the functional traits (used as proxies of ecosystem

properties such as the ANPP or BNPP; e.g., Laliberté and Tylianakis, 2012), are thus likely to

change in response to an intensification of practices, such as from ungrazed to intensively

grazed temperate grasslands, alpine meadows, steppes or desert steppes (Klumpp and Soussana,

2009; Zeng et al., 2015). According to Ziter and MacDougall (2013), the uncertainty

surrounding nutrient-defoliation responses makes it difficult to predict whether C storage would

be higher in managed versus unmanaged grasslands. Thus, soil fertility should be considered

when comparing different grazing intensities in grasslands (Louault et al., 2005).

Increased climate variability is another source of response uncertainty in managed

ecosystems. As more frequent and longer periods of drought associated with heat waves may

threaten and shape the long-term dynamics of perennial ecosystems such as grasslands

(Brookshire and Weaver, 2015), it is important to understand how above- and below-ground

compartments respond to climate variability. However, the data on above- and below-ground

biomass responses to drought for grasslands are limited (Byrne et al., 2013; Wilcox et al., 2015;

Li et al., 2018), although some evidence shows that resource conservation strategy is associated

with drought tolerance (Pérez-Ramos et al., 2012; Reich, 2014). Changes in root morphology

and functioning may thus be important determinants in plants' adaptive strategies to drought

and have been less studied than above-ground plant responses (Biswell and Weaver, 1933;

Dawson et al., 2000; McInenly et al., 2010). However, there are not enough data to make

generalisations about the combined impacts of management and climatic condition variabilities,

such as precipitation reduction, on root and shoot biomass production and plant traits defining

plant strategies related to resource use and grazing intensity.

This study comprised of a long-term field experiment for which controlled grazing intensity

was applied for 10 years. We compared grazing abandonment and two levels of herbage

utilisation by grazing based on five rotations per year. In two consecutive years, the ingrowth

core method was used to measure the monthly root biomass production and calculate the BNPP.

The ANPP was measured using grazing exclusion cages, and the community-weighted mean

(CWM) leaf and root traits were assessed the first year. We tested the following hypotheses: (i)

high grazing intensity increases above-ground mass at the expense of root production as a result

of the direct negative effect of defoliation on root growth, whatever the climatic conditions; (ii)

inter-annual climatic conditions modulate the above- and below-ground biomass production

response to grazing intensity as a consequence of the higher presence of defoliation-tolerant

and drought-sensitive species (*Lolium perenne* or *Trifolium repens*) in the high grazing intensity

treatment; (iii) root traits respond to treatments and are a determinant of the BNPP, as observed

for leaf traits and the ANPP.

Materials and methods

<u>Site characteristics</u>

The experiment transpired at the long-term observatory network (ACBB, https://www.anaee-france.fr/en/infrastructure-services/in-natura-experimentation/agrosystem/acbb) located in St-Genès-Champanelle, France (45° 43′N, 03° 01′E, 880 m a.s.l.). The local climate is semi-continental with oceanic influences and a mean annual temperature and precipitation of 8.5°C

and 784 mm, respectively (Table 1). The site supports mesotrophic multi-specific permanent

grasslands dominated by species with high Ellenberg indicator values for N (Schaffers and

Sykora, 2000), indicating a high level of fertility for the site (Table S1; Louault et al., 2017).

The soil is a cambisol with a sandy loam texture developed on a granitic bedrock. Differences

in the local soil composition and profile led us to consider two distinct blocks characterised by

an eutric cambisol (54% sand, 26% silt, 20% clay, 7.0% organic matter, and pH: 5.9) and a

colluvic cambisol (50% sand, 26% silt, 24% clay, 7.4% organic matter, and pH: 6.0), including

some volcanic materials.

<u>Management</u>

Prior to the installation of this experiment in 2005, the study area had been used for intensive

hay and silage production (combining grazing, mowing, and fertilisation) with mineral

fertilisation, and two years preceding the start of the experiment (2003 and 2004), the grassland

site was mown three times per year without fertilisation. From 2005, the grassland was managed

for 10 years with a gradient of grazing intensity resulting from three treatments: abandonment,

low herbage utilisation (Ca-), and high herbage utilisation (Ca+). The Ca- and Ca+ treatments

were obtained by modifying the stocking density (6.9 and 13.8 livestock units (LSU) ha^{-1},

respectively) with five grazing rotations each year occurring at mid-April, late May, early July,

September, and November and lasting on average 9.6, 9.0, 10.7, 8.6, and 2.1 days, respectively.

The two cattle treatments corresponded to two levels of herbage utilisation by grazing and had

on average a residual plant height of 15.2 cm ± 0.5 (mean ± SE) for Ca- and 7.7 cm ± 0.2 for

Ca+ at the end of each grazing rotation. For each treatment, two replicate plots were created per

block, resulting in four replicates per treatment and a total of 12 plots (2 blocks x 2 plots x 3

treatments). The average distance between the two blocks was approximately 230 m, and all

the treatments were randomised within each block. The size of the plots differed according to

the treatment: 2,200 m^2 for the two cattle treatments and 400 m^2 for the abandonment treatment.

<u>Climatic and edaphic conditions</u>

The daily precipitation (mm) and air temperature (°C) were measured for the two years and

recorded with an onsite meteorological station. A climatic water balance was calculated as

precipitation minus the potential evapotranspiration (P - PET, mm) with the Penman-Monteith

equation. The daily soil temperature (°C) was measured with thermocouple sensors (homemade

copper-constantan sensors) inserted at a 20-cm depth in each plot and recorded with an HOBO

data logger (U12-014, Onset Instruments, MA, USA). The daily volumetric soil water content

(SWC; m^3 m^{-3}) of each plot was measured with two probes (ECHO-10, Decagon, USA) inserted

horizontally at a 20-cm depth and connected to data loggers (EM5 and EM50, Decagon, USA).

From January 2014 to November 2015 (DOY 132–326), the SWC was measured every 30 min

and averaged at a daily scale. For each plot, the average values of the two probes were used.

The daily relative soil water content (RSWC) data are shown and calculated as the ratio:

$RSWC = \frac{SWC - SWCmin}{SWCmax - SWCmi}$, where SWC is the soil moisture at a given day, SWCmin is the

minimum value of soil moisture, and SWCmax is the maximum value of soil moisture, all

observed during the two years. For the soil temperature and RSWC, the values were averaged

according to the root growth timescale.

<u>Root growth and root mass</u>

Six months before the start of the experiment, shallow (0-20 cm) soil was collected on each of

the two blocks, sieved (using a 5 mm mesh size) to remove stones and coarse organic matter,

and then left unused outside covered under a shelter and protected from direct sunlight. This

air-dried soil was subsequently used to fill the ingrowth cores each month.

In December 2013, soil cores were collected with an auger (which had an 8-cm diameter and

0-20 cm depth) for each of the 12 plots at four locations representative of the plant community

in the treatment. On average, the mean distance between locations was 19.8 m ± 0.2 for Ca+,

21.7 m ± 0.1 for Ca-, and 17.2 m ± 0.2 for Ab (mean ± standard deviation; refer to Figure S1).

After the core harvest, each hole was filled with an 8-mm mesh sized plastic net containing a

fixed volume of air-dried sieved soil (called thereafter ingrowth core) collected six months

beforehand. For the next two years and approximately each month (2 x 10 times), the ingrowth

cores, containing soil and the root and rhizome material that had grown therein, were extracted

and then replenished with another fixed volume of air-dried sieved soil. Thus, the monthly and

annual root production (BNPP, g m^{-2} y^{-1}) were measured from February 2014 to December

2015. The root production period ranged on average 36.5 days, but with longer periods in the

winter and shorter periods in the spring and summer (Table 1). In periods without precipitation,

a fixed volume of water was added to adjust soil humidity to field conditions. After collection,

the ingrowth cores were transported to the laboratory and immediately stored at 4°C before

being processed in the next five days. The roots were washed under tap water with a 200-µm

sieve and then oven dried for 48 h at 60°C.

To measure the root mass stock, the soil cores were collected three times (December 2013 and

March and June 2014) with the same auger and near the ingrowth core locations. These samples

were stored in the freezer at -18°C; after defrosting, the roots were washed with the same

procedure as that used for the ingrowth cores and then oven dried for 48 h at 60°C.

<u>Root traits</u>

Subsamples of the washed roots collected with the ingrowth cores in June 2014 were fresh

weighed and then frozen at -18°C before performing a morphology analysis. After defrosting,

the roots were stained with methylene blue (5 g L^{-1}) for approximately 5–10 minutes, rinsed in

water, spread in a transparent glass box containing a thin layer of water, and covered with a

transparent plastic sheet. High resolution images were recorded with a double light scanner

(800 dpi, perfection V700, Epson, JA) and analysed using WinRhizo software (PRO 2012b,

Regent Instruments, CA) with the automatic procedure. Two scans per location were recorded

and separately analysed to measure the root length (m), root volume (cm^3), root surface area

(m^2), average root diameter (mm), and root length by class diameter (there are 13 classes: 11

with a 0.1-mm interval and 2 with a 0.5-mm interval). The specific root length (SRL; m g^{-1}),

root tissue density (RTD; g cm^{-3}), and specific root area (SRA; m^2 g^{-1}) were calculated for fine

roots as in Picon-Cochard et al. (2012).

<u>Botanical composition</u>

The species contribution (%) was visually observed on a 20-cm diameter circle around each

ingrowth core location in April (cattle treatments) and May (abandonment treatment) 2014. For

each zone, a score on a 10-point scale was allocated to the species present according to their

volume occupancy, and the percentage of each species was calculated at the plot scale by

averaging the values of the four zones. Table S2 lists the species and their relative contributions.

Above-ground biomass production

On each plot and on each sampling date, four fenced sampling areas (0.6×0.6 m) were used to measure the accumulation of above-ground biomass after the above-ground standing biomass was clipped to 5.5 cm, oven dried, and weighed. Measurements were made five times over the year, once before each grazing event in the Ca+ and Ca- treatment plots and three times (spring, summer, and autumn) in the abandonment plots. The sampling areas were moved within the plot at each measurement date during the year. The annual above-ground net primary production (ANPP, g m^{-2} y^{-1}) was calculated as the sum of the successive biomass accumulations throughout the year.

Leaf traits

The CWM trait values of the leaf dry matter content (LDMC), specific leaf area (SLA) and reproductive plant height were calculated for each ingrowth core zone using both the relative contribution of the dominant species to the community (i.e., species that account for at least 85% of the cumulated species contribution to the community) measured in 2014 and the leaf trait measurements made at plot scale in 2006 and 2007. The traits were measured on 10 vegetative plants using standard protocols (refer to the methods in Louault et al., 2005). The reproductive plant height was measured on mature plants located in fenced zones to allow full plant development. The CWM is expressed with the following equation: $\text{CWM} = \sum p_i \times \text{trait}_i$, where p_i is the relative contribution of species i to the community and trait$_i$ is the trait of species i.

Statistical analyses

For a given date, data averages of the root mass and root traits collected at each of the four locations (four ingrowth cores in each plot) were used to obtain a single value for each of the 12 plots and test for the effect of the treatment and dates. Before variance analysis (ANOVA), the normality of residuals was inspected with quantile-quantile plots of model residuals, and

variance homogeneity was confirmed by checking the plots of model residuals versus the fitted

values. Data were transformed if they deviated from the ANOVA assumptions (square root, ln,

reciprocal). Linear mixed effects models as available in the R 'nlme' package (Pinheiro et al.,

2015) were used to perform repeated measure ANOVAs to test the effects of treatments and

dates, as well as their interactions on values of root growth, soil temperature, RSWC, and root

mass stock, with plots nested within block as a random factor accounting for temporal pseudo-

replication (S4). For the root growth dynamics, soil temperature, and RSWC (Figure 1; Table

S3), the dates correspond to 20 dates, whereas for the root mass stock, the dates correspond to

three harvest dates (Table 2). For the BNPP, ANPP, and root-to-shoot ratio (BNPP/ANPP), the

data were analysed using a nested mixed model procedure with the treatments and year used as

fixed factors and with plot nested within block used as random factors. For the leaf and root

traits data, the treatments were used as fixed factors with plots nested within block used as a

random factor. Post-hoc tests were performed to compare the significance levels across fixed

factors with a Tukey test ('lsmeans' package). Principal component analyses (PCAs) were

performed for each year to analyse the relationships between the leaf and root traits, soil

temperature, RSWC, root mass stock, ANPP, and BNPP measured at the plot level; the

treatments were considered supplementary categories ('FactoMineR' package). This statistics

approach allows sets of trait and property relationships to be compared to detect response and

effect traits and to analyse multiple dimensions of trait relationships, which is not possible with

pairs of correlations. All statistical analyses were performed in an R environment (Version

3.5.2, R Core Team, 2012) using RStudio (Version 1.1.463). The scripts are shown in S4.

Results

Climatic conditions during the experiment

Compared with the average long-term climatic data for the site, the first and second years of

the experiment received higher (+92 mm) and lower (-199 mm) precipitation, respectively

(Table 1). The PET in the second year was also higher than the long-term average (a 73-mm

difference), leading to a negative annual climatic water balance (P – PET = -181 mm and a

deficit of 271 mm compared to the long-term average). The annual temperature in the two

experimental years was similar and approximately 0.8°C higher than the long-term average for

the site (Table 1). At a monthly timescale and during part of the growing season (March to

September), in comparison with the first year, the second year had a cumulated water deficit

difference of -266 mm and warmer temperatures by an average of 1.9°C. Larger differences

between the two years occurred in June and July with an average higher temperature (+6°C),

higher water deficit (P – PET = -152.6 mm), and less precipitation (-81%) in the second year.

Dynamics of soil temperature and relative soil water content

The soil temperature was significantly affected by the treatment, date, and treatment × date

(Figure 1; Table S3). For most of the dates (February to October), the abandonment treatment

corresponded with a lower soil temperature (1.76°C on average) than the cattle treatments,

whereas the Ca- treatment soil temperature was significantly lower (0.64°C) than the Ca+

treatment. However, this difference was only observed to be significant for a limited number of

dates in early summer of both years. The RSWC fluctuated from 0.6–0.7 at the beginning of

spring to 0.38 in June during the wet year and 0.2 during the dry year, which accords with the

variation of the climatic water balance (P – PET). In the case of the dry year, from summer until

autumn, the RSWC remained lower than 0.4 and P - PET was negative.

Root growth dynamics

Root growth was affected by the date and treatment × date interaction (Figure 1). Each year,

the root growth peak occurred twice, in spring and autumn, and growth was markedly reduced

in summer and winter. Only in the second year did growth cease in summer. Second year root

growth was significantly lower than the first year. Regarding the treatment effect, the

abandonment treatment showed significant lower root growth than the two cattle treatments for the spring period in both years and for the autumn period of the second year. In autumn 2014 (end of September), a delay of growth peaks was also observed, which led to a twofold higher root growth for the abandonment treatment versus the two cattle treatments. The two cattle treatments experienced similar root growth across the years and seasons.

Seasonal root mass stock, below- and above net primary production, and root-to-shoot biomass ratio

The stock of root mass did not change throughout the seasons or across treatments (Table 2). The BNPP, ANPP, and root-to-shoot biomass ratio (R/S) were significantly lower during the second year, with a stronger effect on the BNPP (-44% on average) than the ANPP (-24% on average; Figure 2; Table 3). Only the abandonment treatment maintained its ANPP value in the second year, which led to a 48% decline in the R/S (significant treatment × year, P < 0.01; Table 3). Accordingly, a treatment effect was only observed for the BNPP during the second year, with a decline of 24% for the abandonment treatment compared to the cattle treatments. For the ANPP during the first year, the Ca+ treatment had 22% and 68% higher values, respectively, than the Ca- and abandonment treatments, while the Ca- treatment had a 38% higher ANPP than the abandonment treatment.

Species composition, leaf and root traits

The abandonment treatment was characterised by the dominance of tall grass species (76% in total) with 27.2% of *Alopecurus pratensis*, 18.8% of *Elytrigia repens*, 11.3% of *Poa pratensis*, and 10.3% of *Arrhenatherum elatius*. Some forbs (19%) were also present, whereas legumes were absent (Table S2). The two cattle treatments differed from the abandonment treatment by the equal presence of *Taraxacum officinale* (18% on average) and *Trifolium repens* (17% on average). Differences also existed concerning grass species (56% in total), with the dominance

of *Dactylis glomerata* (22.2%), *A. pratensis* (7.6%), and *Festuca arundinacea* (5.6%) for the

Ca- treatment and *Lolium perenne* (13.6%), *D. glomerata* (9.1%), and *Poa trivialis* (7.2%) for

the Ca+ treatment. Thus, the Ca+ treatment had a higher percentage of *L. perenne* than the Ca-

treatment (Table S2).

The CWM leaf traits were significantly modified by the treatments. The plant height and

LDMC were significantly higher ($P < 0.05$ and $P < 0.0001$, respectively) and the SLA was

lower ($P < 0.05$) in the abandonment treatment than in the two cattle treatments (Table 4).

Unlike the leaf traits, the root traits were only slightly affected by the treatments. The SRL (P

< 0.1) and SRA ($P < 0.05$) were lower in the abandonment treatment than in the Ca- treatment,

but not than in the Ca+ treatment. For other root traits (diameter, RTD, and root length

percentage by class diameter), no between-treatment differences were observed (Table 4).

<u>Co-variation of traits and production</u>

The two main axes of the standardised PCA explained 60.1% and 56.8% of the community trait

and production variation in 2014 and 2015, respectively (Figure 3). For the first year, the first

PCA axis (PC1), which accounted for 43.4% of the total variation, was significantly related to

the leaf and root traits, ANPP, and soil temperature. The soil temperature, SRA, and ANPP had

positive loadings, while the diameter, plant height, and LDMC had negative loadings (Table 5).

The second PCA axis (PC2), which accounted for 16.7% of the total variation, was significantly

and positively related to the root diameter and negatively related to the SRA. For the second

year, the PC1 accounted for 37.4% of the total variation and was significantly related to the leaf

and root traits, ANPP, and BNPP. The BNPP and SRA had negative loadings, whereas the root

diameter, plant height, and ANPP had positive loadings (Table 5). The PC2, which accounted

for 19.4% of the total variation, was significantly and positively related to the RSWC and the

stock of root mass averaged across three dates. The abandonment treatment was significantly

related to the PC1s, with negative and positive loadings for the first and second years,
respectively.

Discussion

Ten years of contrasted management strongly modified the functional diversity and above-

ground production of this fertile upland grassland as previously shown for the same site

(Herfurth et al., 2015; Louault et al., 2017). Accordingly, we expected that the above-ground

biomass patterns would be mirrored below ground, especially during the periods of grazing. In

this section, we first discuss the within-year differences of root growth, followed by the inter-

annual variation responses to grazing intensity and climatic condition variabilities between the

two contrasting years. Finally, we analyse the relationships between traits and the above- and

below-ground production.

<u>Seasonality of root growth in relation to grazing intensity and climatic conditions</u>

As expected, the permanent grassland root growth was affected by seasons and the spring and

autumn peaks (Garcia-Pausas et al., 2011; Pilon et al., 2013; Steinaker and Wilson, 2008), but

unexpectedly, grazing pressure applied by rotations and climatic condition variabilities had

very limited effects on this seasonality. Accordingly, at the below-ground level, the plant

community behaviour was not affected by either the rotational grazing management or the

climatic condition variabilities, although a severe drought occurred in summer of the second

year. Only the abandonment treatment showed a delayed root growth peak in spring. This delay

was probably the result of a slower shoot budburst and a reduced capacity to produce new green

leaves in the dense litter canopy, especially at the beginning of the growing season in spring

(data not shown). Moreover, the tall and dense canopy of the abandonment treatment strongly

modified the soil temperature; as expected, the soil conditions were cooler in the abandoned

vegetation (Picon-Cochard et al., 2006; Zhou et al., 2017; Zhu et al., 2016). As shown in some studies, light, soil, water, and nutrient availability (Edwards et al., 2004; Garcia-Pausas et al., 2011; Steinaker and Wilson, 2008) are other abiotic factors determining root growth dynamics in grasslands, for example root peaks were observed before the summer soil temperature peak when a negative climatic water balance occurred, especially in the second year. Nevertheless, plants growing in the abandonment treatment offset their slower root growth by producing a similar annual root biomass, especially during the wet year. The presence of tall grass species, such as *A. pratensis*, *A. elatius*, and *E. repens*, having plant trait syndromes related to both disturbance and resource conservation strategies (i.e. lower SLA and SRL and higher plant height and root depth; Pagès and Picon-Cochard, 2014) might explain the tall grass capacity to produce higher root biomasses before the canopy senescence onset. Additionally, pre-existing soil fertility can be maintained in conditions of very low levels of herbage utilisation (near-abandonment) because of the absence of biomass exportation and increased internal recycling of N within senescent plants, both of which contribute to an increase in the total N available for plant growth (Loiseau et al., 2005).

The similar root growth dynamics of the two cattle treatments was unexpected since infrequent defoliation and moderate excreta returns to the soil might increase the root biomass production at the expense of the shoot biomass production (Klumpp et al., 2009). The absence of effect on the root growth and BNPP indicates that the grazing applied to the plant communities by rotations was too short but sufficient to observe the effect on the ANPP in wet conditions.

Different methods exist globally to manage grasslands by grazing (Huyghe et al., 2014), including rotational or permanent grazing options with different stocking densities, durations, and types of herbivores. In general, grazing management creates high spatial heterogeneity within the plots due to animals' selective defoliation of plant species and because returns to soil are spatially heterogeneous. Thus, in grazed grasslands, disturbances create patches of

vegetation, which should affect the local root growth and below-ground biomass of the plant

communities if the intensity of grazing is sufficient. The complexity of these phenomena in

grazed grasslands is greater than in mown systems (Rossignol et al., 2011).

Notwithstanding, the confounding effect of soil fertility and defoliation may mask a clear

response of the below-ground compartment in grazed grasslands. In consequence, we postulate

that the root growth in the Ca+ treatment was favoured by the higher soil temperature

compensating for the negative effects of frequent defoliation on root growth, while the cooler

soil conditions encountered in the Ca- treatment might have slowed the root growth. Soil

moisture level is a main determinant of plant growth and can be affected by cattle treatments.

Some studies have shown an increase of soil moisture in grazed compared ungrazed treatments

due to the lower leaf area index in the grazed conditions (Moretto et al., 2001; Pineiro et al.,

2010) or an absence of effect in others studies (LeCain et al., 2002; Smith et al., 2014). The

presence of herbivores can increase soil bulk density and consequently modify soil moisture.

In our field conditions after 10 years of treatment applications, the absence of effect on soil

moisture could be due to several of these factors: absence of change of soil density, due to the

rotational grazing as cattle spend less time in the plots than in continuous grazing systems; the

functional composition of the community regarding both their response to defoliation and their

water-use strategies; the monthly-based temporal scale used, which could buffer shorter-term

responses.

We should also consider the level of soil fertility and species composition as drivers of root

growth and trait plasticity (Dawson et al., 2000). The soil fertility of our site, reflected by the

nitrogen nutrition index (NNI; Lemaire and Gastal, 1997), was similar across our grazing

intensity gradient (Table S1), at least in 2014. Thus, our site provided us with the opportunity

to compare grazing intensity effects at equivalent soil fertility levels. Because root trait

plasticity generally shows larger differences with respect to soil fertility than from cutting or

defoliation (Leuschner et al., 2013; Picon-Cochard et al., 2009), we could expect grazing

intensity to produce a less pronounced effect on root growth under similar soil fertility

conditions. Indeed, the higher presence of defoliation-tolerant species, characterised by a

shorter stature and root system (e.g., *L. perenne* and *P. trivialis*) but a higher shoot and root

growth capacity after defoliation and higher rhizosphere activity (Dawson et al., 2000),

probably compensated for the negative effect of defoliation in the Ca+ treatment. The sampling

depth might also have had an effect as we expect that due to species-specific differential root

depth distribution, harvesting root systems deeper than 20 cm should have resulted in a more

contrasting root growth response across the two cattle treatments according to the grass species

composition (Xu et al., 2014). The results indicate that the high soil temperatures, high soil

fertility, and species composition moderated the root growth response across our grazing

intensity gradient. The difficulty of assigning species composition in root mixtures, however,

makes it difficult to draw firm conclusions.

<u>Responses of ANPP, BNPP, and root-to-shoot biomass production ratio across the grazing</u>

<u>intensity gradient and climate variability</u>

According to meta-analyses and recent results (McSherry and Ritchie, 2013; Zeng et al., 2015;

Zhou et al., 2017; Li et al., 2018), grazing intensity generally negatively affects the above- and

below-ground biomass of grasslands regardless of the climatic conditions or vegetation type,

although these effects can be modulated by the level of grazing intensity. The study results do

not confirm these findings because the ANPP and BNPP increased in the wet and dry year,

respectively, in response to grazing intensity compared to the abandonment treatment.

Methodology issues for estimating the ANPP and BNPP should thus be considered for studies

comparison. Some papers report the use of biomass stock or fluxes measured once at peak of

growth or at several periods (Scurlock et al., 2002), other report estimation of the BNPP from

indirect measurements (e.g., Zeng et al., 2015). The root mass, based on stock, provides a

snapshot of plant functioning that generally includes mixtures of living and senescent tissues,

thus depending on abiotic factors and plant growth, whereas measurements based on new shoot

and root biomass reflect the growth potential of grasslands. Although these methods are very

different, the BNPP measured with ingrowth cores produced similar results to the root mass

stock assessed at three seasons. Another point to consider is the number of samples used to

compare treatments and detect significant differences. In grasslands, the coefficient of variation

of the root dry weight in auger samples is generally high, i.e. between 30 and 50% (Bengough

et al., 2000). According to these authors, our sampling protocol (with 16 samples) was adapted

to detect at least 35% differences between treatments, whereas to detect less than 10%

differences, more than 100 replicates should be collected. While it is possible that collecting

more samples would have highlighted significant differences across the treatments, we had to

find a compromise between performing more frequent samplings (20 dates) to study the

seasonal dynamics of root growth and collecting more samples at the plot level but less

frequently.

Negative climatic water balance (P - PET) conditions had stronger effects on the ANPP and

BNPP than grazing intensity because severe drought had a direct negative effect on plant

growth. In comparison with another experiment located along side ours, 80% of canopy

senescence was reached for P – PET of -156 mm (Zwicke et al., 2013). This index reached -

303 mm from March to August, confirming that a severe drought occurred in the second year

of our experiment explaining the root growth cessation during the summer. At an annual scale,

the ANPP of the two cattle treatments showed lower resistance to increased aridity (resistance

is defined as $ANPP_{year2}/ANPP_{year1}$ and equalled 0.63) than the abandonment treatment (ratio =

1). For the BNPP, the results were inversed, leading to a lower resistance for the root-to-shoot

biomass ratio in the abandonment treatment than in the two cattle treatments. The absence of

root growth modification by grazing at an annual scale during the wet year reflects the slight

change in root-to-shoot biomass allocation. Other processes, such as root turnover, are expected

to change in grazed versus ungrazed grasslands. For our study, Herfurth et al. (2015) observed

a similar root mass stock along a grazing disturbance gradient, but by using a simplified C flux

model, these authors showed that the Ca+ treatment tended to accelerate C cycling in plant

communities, resulting in a higher quantity of C allocated to the soil organic matter continuum.

Taken together, these results suggest that the slight BNPP increase under grazing may occur

with an increase in rhizodeposition because the root turnover, calculated as the BNPP to root

mass stock ratio (data not shown; Lauenroth and Gill, 2003), was not different across the

treatments.

Furthermore, our results suggest that cattle treatments slow the negative effect of aridity on

the root-to-shoot biomass ratio, underling that these treatments seem better adapted to buffering

the negative effect of drought on grassland root production than abandonment. This finding is

consistent with previous work showing that moderate grazing could be more beneficial than no

grazing for drought resistance and the recovery of the ANPP and BNPP (Frank, 2007; Xu et al.,

2012). The BNPP has also been shown to have greater resistant than the ANPP to changes in

precipitation (Yan et al., 2013). Other studies observed no prevalent effects of grazing, drought,

or fire on grassland production in North America or South Africa (Koerner and Collins, 2014).

Nevertheless, further research is needed to determine whether grazing pressure has additive or

combined effects on the drought response of grasslands (Ruppert et al., 2015).

Community-weighted mean leaf and root traits as predictors of the ANPP and BNPP

As shown by some studies (e.g., Diaz et al., 2007; Laliberté and Tylianakis, 2012; Louault et

al., 2017; Zheng et al., 2015), disturbances induced by grazing pressure profoundly affect plant

community and functional traits by selecting defoliation-tolerant species, such as *L. perenne*,

498 *P. trivialis*, or *T. repens*, with possible cascading effects on multiple ecosystem functions. Given

499 their capacity to regrow quickly after defoliation, these species generally exhibit high values of

500 SLA and low values of LDMC and plant height. They contrast with species that are adapted to

501 fertile soil but exhibit a slower regrowth capacity after defoliation, such as *D. glomerata* or *F.*

502 *arundinacea*, with opposite leaf trait values. In abandonment, competition for light tends to

503 select plants with trait syndromes related to disturbance and conservative strategies (e.g., high

504 plant height, low SLA, and high LDMC values). Thus, the CWM traits of a community depend

505 on the balance between these species groups, which are expected to affect the ANPP and BNPP

506 (Klumpp et al., 2009; Milchunas and Lauenroth, 1993). Although species that are tolerant and

507 intolerant to defoliation were present in both cattle treatments, the leaf trait values were

508 similarly and positively related to the ANPP and only differed from the traits of species present

509 in the abandonment treatment. This means that the cessation of grazing strongly differentiated

510 the plant communities, whereas within the two cattle treatments, differences were slighter.

511  For the below-ground compartment, we expected the above-ground differences to be

512 mirrored by the root growth and traits, assuming that higher root diameter values and lower

513 SRL and SRA values would be associated with a lower BNPP in the abandonment treatment

514 compared with the two cattle treatments. Although the root response to grazing (mainly through

515 defoliation) is generally reported to result in a reduction of root mass or root length (Dawson et

516 al., 2000), our study did not confirm these assumptions. The contrasting results are possibly due

517 to the variable abundance of defoliation-tolerant species or the combined effects of both

518 defoliation and the level of soil fertility on the roots of grazed grasslands (Leuschner et al.,

519 2013; Picon-Cochard et al., 2009; Yan et al., 2013; Ziter and McDougall, 2013). Thus, root

520 growth reductions associated with grazing may have a greater impact in locations where the

521 grazer-mediated N return is spatially decoupled from defoliation (McInenly et al., 2010).

522 Furthermore, the higher SRA observed in the Ca- treatment than in the abandonment or Ca+

treatment should reflect a higher presence of species with fine roots, such as *D. glomerata* or *H. lanatus* (Picon-Cochard et al., 2012), because the soil fertility approximated by the NNI was similar across treatments.

Conclusions

The nearly identical functional traits of the plant communities and similar soil fertility across the two cattle treatments explains the absence of changes in the root mass production for these treatments. Our site disentangled the combined effects of fertility and defoliation on root production, which most previous studies have not addressed. Thus, our results suggest the prevalence of a soil fertility effect on the root production response rather than a defoliation effect. Moreover, the strong effect of climatic condition variabilities on the ANPP and BNPP observed in the short term could increase in the future as more frequent climatic extremes are expected. It is thus necessary to improve our knowledge at a larger timescale of the grazing practices that would increase grasslands resilience to more frequent and intense climatic events, such as drought and heat waves.

Data accessibility

Data are available online: https://zenodo.org/record/4034903#.YA129-fjJPZ

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Conflict of interest disclosure

The authors of this preprint declare that they have no financial conflict of interest with the

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Table 1. Air temperature (°C), precipitation (P, mm), potential evapotranspiration (PET) and climatic
water balance: cumulated (P – PET, mm) and calculated for the 28 y period 1986-2013, mean values ±
SD) and measured for the 10 dates in 2014 and 2015 corresponding to measurements of root growth and
averaged (temperature) or summed (P, PET, P – PET) at annual scale.

Year	Dates	Air temperature	Precipitation	PET	P-PET
	Annual long-term average	**8.5 ± 0.6**	**784 ± 1376**	**693 ± 96**	**91 ± 195**
2014	December 12 – February 23	3.7	98	37.5	60.5
	February 24 – March 23	5.3	27	46.3	-19.3
	March 24 – April 21	7.2	23.5	68.7	-45.2
	April 22 – May 25	9.2	79.5	103.1	-23.6
	May 26 – June 22	14.2	58	110.2	-52.2
	June 23 –July 20	15.1	136.5	93.9	42.6
	July 21 – August 24	14.4	90.5	100.5	-10
	August 25 – September 29	13.7	141.8	79.5	62.3
	September 30 – October 29	11.7	69	36.3	32.7
	October 30 – December 14	5.3	111	10.9	72.1
	Annual	**9.2**	**876**	**691**	**157.7**
2015	December 15 – March 1	1.3	132.5	31	101.5
	March 2 – March 29	4.5	36.5	36.8	-0.3
	March 30 – April 23	8.5	17.5	66.4	-48.9
	April 24 – May 28	11.0	66	113.6	-47.6
	May 29 – June 28	15.5	62.5	129.1	-66.6
	June 29 –July 23	21.1	26	136	-110
	July 24 – August 27	16.4	94.5	124.6	-30.1
	August 28 – September 24	12.8	77	66.3	10.7
	September 25 – October 29	7.8	55	36.1	18.9
	October 30 – December 11	7.0	54.5	25.1	29.4
	Annual	**9.4**	**585**	**766**	**-180.9**

Table 2. a) Repeated measure ANOVA is shown for treatment, date (December 2013, March 2014, June 2014) and interaction effects on root mass (g m^{-2}). Numerator (num), denominator (den) of degree of freedom (DF) and F values are shown. b) Root mass (g m^{-2}) of abandonment, low (Ca-) and high (Ca+) grazing intensity treatments measured in winter (December 12 2013), spring (March 20 2014), summer (June 20 2014) and averaged across the three dates. Means $\pm$ SE are shown, n = 4. Superscripts ^{ns} correspond to P > 0.05.

a)	num/den DF	*F-value*	
Treatment	2/8	1.151^{ns}	
Date	2/18	2.027^{ns}	
Treatment × date	4/18	1.340^{ns}	
b) Date	Abandonment	Ca-	Ca+
December 2013	636.4 ± 133.1	403.3 ± 66.4	496.5 ± 20.6
March 2014	559.1 ± 166.2	609.2 ± 45.3	719.8 ± 47.5
June 2014	574.2 ± 84.8	482.2 ± 38.6	591.2 ± 101.7
3 dates average	589.9 ± 99.9	498.2 ± 43.6	602.5 ± 44.4

Table 3. Repeated measure ANOVA is shown for treatment, year and interaction effects on annual root production (BNPP, g m^{-2} y^{-1}), annual above-ground production (ANPP, g m^{-2} y^{-1}) and root to shoot ratio (R/S). Numerator (num), denominator (den) of degree of freedom (DF), F values are shown. Superscripts $^{ns, *, **, ***}$ correspond to $P > 0.05$, $P < 0.05$, $P < 0.01$, $P < 0.001$, respectively.

	num/den DF	BNPP *F-value*	ANPP *F-value*	R/S *F-value*
Treatment	2/8	2.51^{ns}	8.10^{*}	0.46^{ns}
Year	1/9	70.72^{***}	83.77^{***}	13.09^{**}
Treatment × Year	2/9	3.83^{ns}	22.21^{**}	9.52^{**}

Table 4. Root traits measured from ingrowth core collected in June 2014 and leaf traits measured from botanical observation in abandonment (May 2014), Ca- and Ca+ (April 2014) treatments. Diameter: root diameter (mm); SRL: specific root length (m g^{-1}); RTD: root tissue density (g cm^{-3}); SRA: specific root area (m^2 g^{-1}); % 0-0.1 mm: percentage of length in the class diameter 0-0.1 mm; % 0.1-0.2 mm: percentage of length in the class diameter 0.1-0.2 mm; % 0.2-0.3 mm: percentage of length in the class diameter 0.2-0.3 mm; % > 0.3 mm: percentage of length in the class diameter > 0.3 mm; Community-weighted mean (CWM) Height: plant height (cm); SLA: specific leaf area (cm^2 g^{-1}); LDMC: leaf dry matter content (g g^{-1}). Means ± SE are shown (n = 4). num/den DF: numerator and denominator of degree of freedom. Superscripts $^{ns, +, *, **, ***}$ correspond to P > 0.1, P ≤ 0.1, P < 0.05, P < 0.01, P < 0.001, respectively. For SRL, SRA, height, SLA and LDMC, different letters correspond to significant differences between treatments.

	num/den DF	F-value	Abandonment	Ca-	Ca+
Root traits					
Diameter	2/8	1.61^{ns}	0.240 ± 0.015	0.210 ± 0.006	0.222 ± 0.015
SRL	2/8	3.71^{+}	237.2 ± 26.3 b	332.7 ± 30.4 a	277.8 ± 23.8 ab
RTD	2/8	0.55^{ns}	0.099 ± 0.007	0.095 ± 0.003	0.102 ± 0.007
SRA	2/8	4.96*	0.137 ± 0.011 b	0.182 ± 0.008 a	0.155 ± 0.01 ab
% 0-0.1 mm	2/8	1.28^{ns}	28.5 ± 1.1	32.9 ± 5.5	28.8 ± 2.6
% 0.1-0.2 mm	2/8	0.46^{ns}	37.7 ± 4.4	37.7 ± 2.2	39.1 ± 1.8
% 0.2-0.3 mm	2/8	0.30^{ns}	16.6 ± 1.2	16.2 ± 2.4	17.1 ± 1.9
% > 0.3 mm	2/8	1.22^{ns}	17.2 ± 5.0	13.2 ± 1.3	15.1 ± 2.1
Leaf traits					
CWM_Height	2/8	8.45*	93.0 ± 3.5 a	72.8 ± 7.0 b	68.6 ± 3.8 b
CWM_SLA	2/8	5.30*	205.1 ± 5.7 b	231.8 ± 7.3 a	225.5 ± 7.1 ab
CWM_LDMC	2/8	11.22***	0.261 ± 0.008 a	0.227 ± 0.007 b	0.213 ± 0.010 b

Table 5. Contribution of the different variables to the first two axes of the principal component analysis (PCA) calculated for 2014 and 2015. Variables used in the PCA were annual relative soil water content (RSWC), annual soil temperature (Tsoil, °C), root diameter (Diam, mm), specific root area (SRA, m^2 g^{-1}), root mass averaged over three dates (RootMass, g m^{-2}), annual root production (BNPP, g m^{-2} y^{-1}), plant height (Height, cm), leaf dry matter content (LDMC, g g^{-1}), annual above-ground production (ANPP, g m^{-2} y^{-1}). Treatments were added as supplementary categories.

Variable	2014		2015	
	Axis 1 (43.4 %)	Axis 2 (16.7 %)	Axis 1 (37.4 %)	Axis 2 (19.4 %)
RSWC	0.62	0.44	-0.21	**0.64**
Tsoil	**0.91**	0.09	-0.58	0.52
Diam	**-0.64**	**0.75**	**0.78**	0.53
SRA	**0.62**	**-0.58**	**-0.69**	-0.48
RootMass	-0.06	0.22	-0.07	**0.60**
BNPP	0.21	-0.23	**-0.71**	0.35
Height	**-0.82**	-0.07	**0.83**	-0.19
LDMC	**-0.83**	-0.12	**0.61**	0.03
ANPP	**0.71**	0.54	**0.57**	0.20
Suppl. Categories				
Abandonment	**-2.62**	-0.24	**2.04**	-0.27
Ca-	1.07	-0.55	-1.21	-0.62
Ca+	0.70	0.18	-0.83	0.90

Contribution in bold indicates significant correlation of the variables on the PCA axis ($P < 0.05$).

Figure captions

Figure 1. The root growth (g m^{-2} day^{-1}), soil temperature (°C), relative soil water content, and climatic

water balance index (P - PET, mm; hashed bars) dynamics measured over two years for the

abandonment, low (Ca-), and high (Ca+) grazing intensity treatments. The vertical bars correspond to 1

SE (n = 4). The insets indicate P values from the repeated measure two-tailed ANOVA, where treat. is

the treatment; dates, and interaction for the main treatments. *: $P < 0.05$; x: $P \leq 0.1$. For soil temperature,

*# corresponds to significant differences between all treatments (Abandonment < Ca- < Ca+).

Figure 2. The annual root biomass production (BNPP, g m^{-2} y^{-1}), annual above-ground biomass

production (ANPP, g m^{-2} y^{-1}), and root-to-shoot biomass ratio measured in 2014 and 2015 for the

abandonment, low (Ca-), and high (Ca+) grazing intensity treatments. The vertical bars correspond to 1

SE (n = 4). Within a year, different letters correspond to significant differences at $P < 0.05$.

Figure 3. A principal component analysis (PCA) combining the leaf and root traits, above- and below-

ground net primary production, root mass stock, relative soil water content, and soil temperature

measured in 2014 (a) and 2015 (b) for the abandonment, low (Ca-), and high (Ca+) grazing intensity

treatments. Data from each plot were used in each PCA. The first two axes are shown. The arrows

indicate projections of the variables within the PCA where RSWC is the relative soil water content,

Tsoil is the soil temperature (°C), Diam is the root diameter (mm), SRA is the specific root area (m^2 g^{-1}

), RootMass is the root mass averaged over three dates (g m^{-2}), BNPP is the annual root production (g

m^{-2} y^{-1}), Height is the plant height (cm), LDMC is the leaf dry matter content (g g^{-1}), and ANPP is the

annual above-ground production (g m^{-2} y^{-1}).

883   Figure 3

Table S1. Nitrogen nutrition index (NNI %, Lemaire and Gastal 1997, Cruz et al., 2006) measured on forage regrowth of May in 2014 and 2015 on the non-leguminous part to assess the effect of treatments on N availability according to grazing intensity. When legumes were below 4.5% in the herbage mass, NNI was assessed using the procedure defined by Cruz et al. (2006) based on the total forage and the legume contribution. The P-values are associated with a nested mixed model: treatment used as fixed factor with plots nested in blocks as random factors. Mean ± SE is shown (n = 4). For each year, different letters correspond to significant differences at $P < 0.05$.

Year	P-value	Abandonment	Ca-	Ca+
2014	0.146	65.64 ± 3.10 a	59.54 ± 1.78 a	63.72 ± 2.86 a
2015	0.018	69.72 ± 1.19 a	61.71 ± 1.53 b	69.25 ± 2.09 a

For each year, different letters correspond to significant differences at *: $P < 0.05$; **: $P < 0.01$; ***: $P < 0.001$; ns: $P > 0.05$.

References

Lemaire G, Gastal F (1997) N uptake and distribution on plant canopy. In: Lemaire, G (ed.) Diagnosis of the nitrogen status in crops, pp. 3-43. Springer-Verlag, Berlin, DE.
Cruz P, Jouany C, Theau J-P, Petibon P, Lecloux E, Duru M (2006) L'utilisation de l'indice de nutrition azotée en prairies naturelles avec présence de légumineuses. Fourrages 187:369-376.

Table S2. Species contribution (%) in the community present around the ingrowth core measured in April and May 2014 for Ca-, Ca+ and Abandonment, respectively. Mean ± SE is shown (n = 4). For each species, different letters correspond to significant differences at *: P < 0.05; **: P < 0.01; ***: P < 0.001; ns: P > 0.05.

Group	Species	*P-value*	Abandonment	Ca-	Ca+
Grasses	*Agrostis capillaris*	ns	0.0 ± 0.0	0.6 ± 0.6	1.7 ± 1.2
	Arrhenatherum elatius	ns	10.3 ± 6.8	2.2 ± 2.2	2.5 ± 2.5
	Alopecurus pratensis	**	27.2 ± 7.9 a	7.8 ± 3.3 b	3.3 ± 1.7 b
	Dactylis glomerata	*	3.1 ± 2.7 b	22.2 ± 9.8 a	9.1 ± 3.8 ab
	Elytrigia repens	*	18.8 ± 9.9 a	2.8 ± 1.8 b	3.8 ± 2.7 b
	Festuca arundinacea[x]	ns	5.0 ± 2.3	5.6 ± 2.1	6.3 ± 2.2
	Holcus lanatus	*	0.0 ± 0.0 b	4.7 ± 1.6 a	3.4 ± 1.9 a
	Lolium perenne	***	0.0 ± 0.0 b	0.9 ± 0.9 b	13.6 ± 3.8 a
	Poa pratensis	ns	11.3 ± 2.2	3.1 ± 1.5	3.4 ± 2.5
	Poa trivialis	*	0.0 ± 0.0 b	5.0 ± 2.5 a	7.2 ± 2.4 a
	Trisetum flavescens	ns	0.0 ± 0.0	2.2 ± 1.3	0.6 ± 0.4
Forbs	*Achillea millefolium*	ns	1.3 ± 0.9	3.8 ± 2.4	3.1 ± 2.3
	Anthriscus sylvestris	ns	2.5 ± 2.1	0.0 ± 0.0	0.0 ± 0.0
	Cerastium fontanum	ns	0.0 ± 0.0	1.3 ± 0.9	0.0 ± 0.0
	Cerastium glomeratum	ns	0.0 ± 0.0	0.0 ± 0.0	0.3 ± 0.3
	Cirsium arvense	ns	5.0 ± 3.5	0.0 ± 0.0	0.0 ± 0.0
	Hypocheris radicata	ns	0.0 ± 0.0	0.9 ± 0.9	0.0 ± 0.0
	Ranunculus acris	ns	0.0 ± 0.0	0.0 ± 0.0	3.8 ± 3.8
	Stellaria graminea	ns	0.6 ± 0.6	0.6 ± 0.4	0.0 ± 0.0
	Taraxacum officinale agg.	**	0.0 ± 0.0 b	17.5 ± 1.8 a	19.1 ± 6.0 a
	Urtica dioïca	*	9.7 ± 4.9 a	0.0 ± 0.0 b	0.0 ± 0.0 b
	Veronica serpyllifolia	ns	0.0 ± 0.0	0.3 ± 0.3	0.0 ± 0.0
Legumes	*Lathyrus pratensis*	ns	0.0 ± 0.0	0.3 ± 0.3	0.3 ± 0.3
	Trifolium pratense	ns	0.0 ± 0.0	0.0 ± 0.0	0.3 ± 0.3
	Trifolium repens	***	0.0 ± 0.0 b	16.3 ± 4.0 a	17.7 ± 2.5 a

[x]: new species name: *Schedonorus arundinaceus*

Table S3. Repeated measure ANOVA is shown for root growth (g m^{-2} day^{-1}), soil temperature (Tsoil, °C) and relative soil water content (RSWC) responses to treatment, dates (d1 to d20) and interaction effects. Numerator (num), denominator (den) of degree of freedom (DF) and F values are shown. Superscripts $^{ns, **, ***}$ correspond to P > 0.05, P < 0.001, P < 0.0001, respectively.

Variables	Treatment		Dates		Treat. x Dates	
	num/den DF	F-value	num/den DF	F-value	num/den DF	F-value
Root growth	2/8	1.80^{ns}	19/171	50.40***	38/171	2.096**
Tsoil	2/8	33.93***	19/166	944.83***	38/166	9.75***
RSWC	2/8	1.914^{ns}	19/163	25.287***	38/163	1.097^{ns}

S4: R scripts used in the paper

1) Root growth, soil temperature and relative soil water content
```
> lme1 <- lme (variable ~ trait*date, random = ~ 1 | bloc/ID,
+            data=IGC_ID, method="ML")
> anova(lme1)

plot(lme1)
> par(mfrow=c(2,2))
> plot(residuals(lme1))
> qqPlot(residuals(lme1))
> hist(residuals(lme1))

> lme1.lsmeans <-lsmeans(lme1, pairwise ~  trait | date)
> print(lme1.lsmeans[[2]])
```

2) Root mass stock measured at three dates
```
> lme1 <- lme (StockRac ~ trait*date, random = ~ 1 | bloc/ID,
+         data=Stock_ID, method="ML")
> anova(lme1)

plot(lme1)
> par(mfrow=c(2,2))
> plot(residuals(lme1))
> qqPlot(residuals(lme1))
> hist(residuals(lme1))

> lme1.lsmeans <-lsmeans(lme1, pairwise ~  trait | date)
> print(lme1.lsmeans[[2]])
```

3) Root mass stock: averaged of the three dates, root and leaf traits
```
> lme1 <- lme (variable ~ trait, random = ~ 1 | bloc/ID,
+            data=Stock3, method="ML")
> anova(lme1)

plot(lme1)
> par(mfrow=c(2,2))
> plot(residuals(lme1))
> qqPlot(residuals(lme1))
> hist(residuals(lme1))
```

```
> lme1.lsmeans <-lsmeans(lme1, pairwise ~  trait)
> print(lme1.lsmeans[[2]])
```

4) BNPP, ANPP, Root shoot mass ratio, NNI

```
> lme1 <- lme (variable ~ trait*year, random = ~ 1 | bloc/ID,
+             data=BNPP, method="ML")
> anova(lme1)

plot(lme1)
> par(mfrow=c(2,2))
> plot(residuals(lme1))
> qqPlot(residuals(lme1))
> hist(residuals(lme1))

> lme1.lsmeans <-lsmeans(lme1, pairwise ~  trait | year)
> print(lme1.lsmeans[[2]])
```

984   Figure S1: Scheme of the plots and blocks on the experimental site