

# 1 Hybridisation capture allows DNA damage analysis of ancient marine eukaryotes

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13

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15

## 16 Abstract

17

18 Marine sedimentary ancient DNA (*sedaDNA*) is increasingly used to study past ocean  
19 ecosystems, however, studies have been severely limited by the very low amounts of DNA  
20 preserved in the seafloor, and the lack of bioinformatic tools to authenticate *sedaDNA* in  
21 metagenomic data. We applied a hybridisation capture ‘baits’ technique to target marine  
22 eukaryote *sedaDNA* (specifically, phytoplankton, ‘Phytobaits1’; and harmful algal bloom  
23 taxa, ‘HABbaits1’), which resulted in up to 4- and 9-fold increases, respectively, in the  
24 relative abundance of eukaryotes compared to shotgun sequencing. We further used the  
25 new bioinformatic tool ‘HOPS’ to authenticate the *sedaDNA* component, establishing a new

26 proxy to assess *sedaDNA* authenticity, the Ancient:Default (A:D) sequences ratio, here  
27 positively correlated with seafloor depth, and generated the first-ever DNA damage  
28 profiles of a key phytoplankton, the ubiquitous coccolithophore *Emiliana huxleyi*. Our study  
29 opens new options for the detailed investigation of marine eukaryotes and their evolution  
30 over geological timescales.

31

## 32 **1 Introduction:**

33

34 Over the past decade marine sedimentary ancient DNA (*sedaDNA*) has become  
35 increasingly used to study past ocean ecosystems and oceanographic conditions. The  
36 novelty of using *sedaDNA* lies in its enormous potential to detect genetic signals of taxa  
37 that do and don't fossilise – meaning that in theory it is possible to go beyond standard  
38 environmental proxies and facilitate reconstruction of past marine ecosystems across the  
39 entire food web. For example, *sedaDNA* has revealed relationships between past marine  
40 community composition and paleo-tsunami episodes in Japan over the past 2,000 years  
41 (Szczuciński et al., 2016), oxygen minimum zone expansions in the temperate Arabian Sea  
42 region over 43 thousand years (kyr) (More et al., 2018), and Arctic sea-ice conditions  
43 spanning 100kyr (DeSchepper et al., 2019). While the logistical challenge of acquiring  
44 undisturbed sediment cores from the deep seafloor remains, the field of *sedaDNA* research  
45 is rapidly advancing due to new ship-board core sampling procedures that allow far greater  
46 contamination control, and improvements in sample processing, sequencing technologies  
47 and bioinformatic tools (Armbrecht et al., 2019).

48

49 Among the huge diversity of marine eukaryotes, phytoplankton are particularly useful  
50 targets to study past ocean conditions. Phytoplankton are free-floating, unicellular

51 microalgae fulfilling two important functions: (1) they form the base of the marine food web  
52 supporting virtually all higher trophic organisms (e.g., Verity and Smetacek, 1990), and (2)  
53 are highly useful environmental indicators due to their sensitivity to changing physical and  
54 chemical oceanographic conditions (Hays et al., 2005). After phytoplankton die, they sink to  
55 the seafloor where small proportions of their DNA are able to become entombed and  
56 preserved in sediments under favorable conditions, over time forming long-term records of  
57 past ocean and climate conditions. Using the small subunit ribosomal RNA gene (18S  
58 rRNA, a common taxonomic marker gene), we recently determined the fraction of marine  
59 eukaryote *sedaDNA* preserved in Tasmanian coastal sediments to be a mere 1.37% of the  
60 total *sedaDNA* pool (Armbrecht et al., 2020). A slightly higher proportion of eukaryote  
61 *sedaDNA* (and also higher diversity) may be captured by combining multiple taxonomic  
62 markers, e.g., the small and large subunit ribosomal RNA gene (Armbrecht, 2020).  
63 However, rather than analysing only part of the total *sedaDNA* pool (such as eukaryote  
64 marker genes within a large metagenomic dataset), it would be much more cost-effective to  
65 increase marine eukaryote *sedaDNA* yield by optimising extraction and laboratory  
66 protocols, to maximise sequencing of *sedaDNA* from the intended target organisms.

67

68 Metagenomic approaches extract and analyse the ‘total’ DNA in a sample (‘shotgun’ style),  
69 irrespective of the source organism, facilitating recovery of DNA sequences from any  
70 organism in proportion to their original presence in that sample. As a result, metagenomic  
71 approaches are well suited to the study of microbial and environmental ancient DNA (e.g.,  
72 Taberlet et al., 2012; Pedersen et al., 2015; Weyrich et al., 2017), including *sedaDNA*. The  
73 use of metagenomics does not prescribe the target DNA fragment size and preserve DNA  
74 damage patterns characteristic of ancient DNA. Importantly, the combination of DNA

75 fragment size variability and damage patterns are vital to assess the authenticity of  
76 potential ancient genetic signals.

77

78 Hybridisation capture techniques are an increasingly popular method to focus the  
79 metagenomic analysis towards loci of interest, such as specific sequences to investigate  
80 particular groups of organisms (Horn et al., 2012; Foster et al., 2020). Hybridisation capture  
81 uses short RNA probes (also called ‘baits’) designed to be complementary to DNA  
82 sequences of interest (e.g., taxonomic marker genes; Fig. 1). By binding to the target  
83 sequence, these genetic baits ‘capture’ DNA fragments from DNA extracts in a manner that  
84 preserves size variability, along with DNA-damage patterns that can be used to examine  
85 whether sequences appear ancient. Additionally, careful bait design (i.e., selection of target  
86 sequences) and optimisations of the application protocol (e.g., hybridisation-temperature  
87 settings) allow differing levels of specificity in the capture process. While such ‘baits’  
88 approaches have previously been used to investigate human, animal and even  
89 environmental DNA (Paijmans et al., 2013; Li et al., 2015; Murchie et al., 2020), its  
90 application to marine sediments to capture *sedaDNA* from key primary producers and  
91 environmental indicator organisms (e.g., eukaryotic phytoplankton) remains untested.

92

93 The assessment of *sedaDNA* authenticity has been hindered by a lack of established  
94 approaches to identify and analyse DNA damage patterns of rare ancient microorganisms  
95 in metagenomic samples (such as eukaryotes in marine *sedaDNA*). For example, software  
96 commonly used to detect DNA damage patterns, such as ‘mapDamage’, computes  
97 nucleotide misincorporation and fragmentation patterns by mapping next-generation  
98 sequencing reads against a reference genome (Ginolhac et al., 2012; Jónsson et al., 2013).  
99 This requires high-quality modern reference genomes, or species where ancient DNA is

100 available in sufficient quantity (e.g., animals or humans; Llamas et al., 2015; Tobler et al.,  
101 2017), but neither is generally possible with marine eukaryote *sedaDNA*. There is a lack of  
102 high-quality reference sequences for the thousands of marine organisms occurring in the  
103 global ocean, and the threshold of ~250 reads per species required to analyse and plot  
104 DNA damage patterns in mapDamage (Collin et al., 2020) is often not reached in *sedaDNA*.  
105 Recently, Hübler et al. (2020) developed a new bioinformatic tool HOPS - 'Heuristic  
106 Operations for Pathogen Screening' - based on the mapDamage algorithm, to identify and  
107 authenticate bacterial pathogens in ancient metagenomic samples and extract this  
108 information for further downstream analysis. In combination with hybridisation capture to  
109 generate a larger number of ancient eukaryote sequences, HOPS has the potential to allow  
110 the assessment of *sedaDNA* authenticity based on DNA damage profiles from key marine  
111 eukaryotes, even if only very few sequences are available ( $\geq 50$  reads per species, Hübler  
112 et al., 2020).

113

114 Here, we develop and apply two hybridisation capture bait sets for the first such analysis of  
115 marine sediments, targeting (i) marine phytoplankton very broadly for general paleo-  
116 monitoring (Phytobaits1), and (ii) selected microalgae (including key phytoplankton groups  
117 such as diatoms, dinoflagellates and coccolithophores) that are highly abundant and/or the  
118 cause of harmful algal blooms (HABs) in our study region of the East Australian coast  
119 (HABbaits1). Based on samples from two coastal sediment cores collected near Maria  
120 Island, Tasmania, we demonstrate: 1) the suitability of Phytobaits1 and HABbaits1 as  
121 effective tools to maximise *sedaDNA* originating from eukaryote targets relative to shotgun  
122 data; 2) the authenticity of both shotgun- and baits-derived sequencing data via HOPS; 3)  
123 examine relationships between the 'ancient' DNA fraction and subseafloor depth through

124 the development of a new *sedaDNA* proxy; and 4) generate the first-ever DNA damage  
125 profile for a keystone marine phytoplankton, the coccolithophore *Emiliania huxleyi*.

126

## 127 **2 Methods:**

128

### 129 *2.1 Samples*

130 Cores were collected during the *RV Investigator* voyage IN2018\_T02 (19 and 20 May 2018,  
131 respectively, Fig. 2) to Tasmania, from sites in the Mercury Passage and Maria Island (Fig.  
132 2). We collected one KC Denmark Multi-Core (MCS3, 36 cm long, estimated to cover the  
133 last ~135 years based on  $^{210}\text{Pb}$  dating at the Australian Nuclear Science and Technology  
134 Organisation (ANSTO, Lucas Heights, Sydney) in the Mercury Passage (MP, 42.550 S,  
135 148.014 E; 68 m water depth), and one gravity core (GC2; 3 m long) offshore from Maria  
136 Island (42.845 S, 148.240 E; 104 m) composed of 2 sections; GC2A (bottom) and GC2B  
137 (top) estimated to cover the last ~8,000 years based on  $^{14}\text{C}$  dating, ANSTO). The untreated  
138 cores were initially stored on-board at 10°C, followed by transport to and storage at 4 °C at  
139 ANSTO. To minimise contamination during core slicing and subsampling (October, 2018,  
140 ANSTO), we wiped working benches, sampling and cutting tools with bleach and 80%  
141 EtOH, changed gloves immediately when contaminated with sediment, and wore  
142 appropriate PPE at all times (gloves, facemask, hairnet, disposable lab gown). We removed  
143 the outer ~1 cm of the working core-half (working from bottom to the top of the core), then  
144 collected plunge samples by pressing sterile 15 mL centrifuge tubes (Falcon) ~2 cm deep  
145 into the sediment core centre at 5 cm depth intervals. All *sedaDNA* samples were  
146 immediately frozen at -20°C and transported to the Australian Centre for Ancient DNA  
147 (ACAD), Adelaide. For this study, a total of 30 samples were selected from both cores,

148 representing ~2 cm depth intervals within the upper 36 cm of MCS3 and GC2, and ~20 cm  
149 depth intervals in GC2 downcore from 36 cm below seafloor (cmbsf).

150

## 151 *2.2 SedaDNA extractions*

152 We prepared *sedaDNA* extracts and sequencing libraries at ACAD's ultra-clean ancient  
153 (GC2) and forensic (MCS3) facilities following ancient DNA decontamination standards  
154 (Willerslev and Cooper, 2005). All sample tubes were wiped with bleach on the outside prior  
155 to entering the laboratory for subsampling. Our extraction method followed the optimised  
156 ("combined") approach outlined in detail in Armbrecht et al. (2020), with a minor  
157 modification in that we stored the final purified DNA in TLE buffer (50  $\mu$ L Tris HCL (1M), 10  
158  $\mu$ L EDTA (0.5M), 5 mL nuclease-free water) instead of customary Elution Buffer (Qiagen)  
159 (see Supplementary Material Methods). To monitor laboratory contamination, we used  
160 extraction blank controls (EBCs) by processing 1-2 (depending on the extraction-batch size)  
161 empty bead-tubes through the extraction protocol. A total of 30 extracts were generated  
162 from sediment samples and 7 extracts from EBCs.

163

## 164 *2.3 RNA-baits design*

165 We designed two RNA hybridisation bait-sets, one targeting phytoplankton for a more  
166 detailed overview of phytoplankton diversity (hereafter 'Phytobaits1'), and one targeting  
167 specific plankton organisms and their predators to enable detailed investigation of HABs,  
168 especially those caused by dinoflagellates, in coastal marine ecosystems (hereafter,  
169 'HABbaits1'). Phytobaits1 was based on 18S-V9 and 16S-V4 sequences of major phyto-  
170 and zooplankton groups, whereas we designed HABbaits1 from a collection of LSU, SSU,  
171 D1-D2-LSU, COI, rbcL and ITS sequences for specific marine target organisms often  
172 associated with HABs in our study region (Table 1; Supplementary Material Methods). In

173 collaboration with Arbor Biosciences, USA, we designed RNA baits based on these target  
174 sequences, with Phytobaits1 containing a total of 15,952 RNA baits targeting the 18S-V9  
175 region of a broad diversity of phytoplankton and their predators and the 16S-V4 region of  
176 three cyanobacteria, and HABbaits1 contained 15,310 RNA baits targeting commercially  
177 important toxic microalgae and their predators (see Supplementary Material Methods).

178

#### 179 *2.4 Library preparations*

180 We prepared metagenomic libraries from all DNA extracts following Weyrich et al. (2017),  
181 with the following modifications. A 20 µL aliquot of DNA was repaired (15 min, 25 °C) in a  
182 40 µL reaction using T4 DNA polymerase (New England Biolabs). After purifying the DNA  
183 (MinElute™ Reaction Cleanup Kit, Qiagen), a ligation step followed (T4 DNA ligase,  
184 Fermentas) in which truncated Illumina-adapter sequences containing two unique 5 base-  
185 pair (bp) barcodes were attached to the double-stranded DNA (60 min, 22 °C) (Meyer and  
186 Kircher, 2010). DNA purification (MinElute™ Reaction Cleanup Kit, Qiagen) was performed,  
187 followed by a fill-in reaction with adapter sequences (Bst DNA polymerase, New England  
188 Biolabs; 30 min, 37 °C, with polymerase deactivation for 10 min, 80 °C). For metagenomic  
189 shotgun library preparations we followed the protocol outlined in detail in Armbrecht et al.  
190 (2020), with slight modifications described in Supplementary Material Methods. For  
191 sequencing library preparations for the hybridisation capture we followed the MyBaits®  
192 Manual v4.1 April, 2018; Arbor Biosciences, USA, with modifications detailed in  
193 Supplementary Material Methods. Sequencing was performed at the Australian Cancer  
194 Research Foundation Cancer Genomics Facility & Centre for Cancer Biology, Adelaide,  
195 Australia, and at the Garvan Institute of Medical Research, KCCG Sequencing Laboratory  
196 (Kinghorn Centre for Clinical Genomics), Darlinghurst, Australia.

197



## 198 2.5 Data analysis

### 199 2.5.1 Bioinformatics

200 Bioinformatic processing and filtering of the sequencing data, hereafter referred to as  
201 datasets ‘Shotgun’, ‘Phytobaits1’ and ‘HABbaits1’, followed established protocols previously  
202 described in Armbrecht et al. (2020), with the exception that we used the NCBI Nucleotide  
203 database (<ftp://ftp.ncbi.nlm.nih.gov/blast/db/FASTA/nt.gz>, downloaded November 2019) as  
204 the reference database to align our *sedaDNA* sequences to (allowing us to run all three  
205 datasets against the same database; see Supplementary Material Methods). All species  
206 detected in EBCs (Supplementary Material Table 1) were subtracted from the sample data,  
207 and hereafter the term ‘samples’ refers to sediment-derived data post-EBC subtraction. For  
208 each dataset (Shotgun, Phytobaits1 and HABbaits1), we used MEGAN6 Community Edition  
209 V6.18.10 to rank our assigned reads by domain and exported these read counts. We  
210 determined relative abundances per domain per sample, and the average and standard  
211 deviation per domain across all samples from MCS3 and GC2 (separately for each site due  
212 to relatively high variability in relative abundance between them, see results). To quantify  
213 the increase in the proportion of our target domain Eukaryota using Phytobaits1 and  
214 HABbaits1 relative to Shotgun, we determined the ratio between the average relative  
215 abundance per domain between Phytobaits1:Shotgun, and HABbaits1:Shotgun.

216

### 217 2.5.2 Ancient DNA authenticity assessment and damage analysis

218 To assess the authenticity of our Shotgun, Phytobaits1 and HABbaits1 *sedaDNA* we ran  
219 the ‘MALTEExtract’ and ‘Postprocessing’ tools of the HOPS v0.33-2 pipeline (Hübler et al.,  
220 2020). For specific configuration settings see Supplementary Material Methods. We  
221 processed each dataset using the ‘def\_anc’ mode, which provided results for all filtered  
222 reads (‘default’; D) as well as all reads that had at least one damage lesion in their first 5

223 bases from either the 5' or 3' end ('ancient'; A) (Hübler et al., 2020). Generally, HOPS  
224 determines DNA damage patterns separately for individual taxa, i.e., requires an input list of  
225 target taxa for which to compare the *sed*aDNA sequences identified in our samples to their  
226 modern references. We used two taxa screening lists with the aim to generate *sed*aDNA  
227 damage profiles for a representative regional phytoplankton species: (a) 'Eukaryota', to  
228 allow a general assessment of the amount of eukaryote sequences categorised as 'default'  
229 or 'ancient' in each of our samples; and (b) a set of selected marine organisms known to be  
230 common in our Tasmanian study region (Table 1). Subsequently to (a) we used the HOPS-  
231 generated 'RunSummary' output to determine the ratio of ancient to default in each sample  
232 for each dataset ('A:D' ratio hereafter). Eukaryote taxa recovered in the EBCs  
233 (Supplementary Material Table 1) were excluded from the calculation. Subsequent to (b),  
234 we used the MaltExtract Interactive Plotting Application (MEx-IPA, by J. Fellows Yates;  
235 <https://github.com/jfy133/MEx-IPA>) to visualise *sed*aDNA damage profiles of the target  
236 phytoplankton *Emiliana huxleyi* Shotgun, Phytobaits1 and HABbaits1 (ancient reads only).

237

### 238 2.5.3 Statistics

239 To determine relationships between the A:D ratio and subseafloor depth and test the A:D  
240 ratio's validity as *sed*aDNA authenticity proxy, we performed two-tailed Pearson correlation  
241 analyses between the A:D ratios of Shotgun, Phytobaits1 and HABbaits1 (n = 27 each, as  
242 no data was retrieved for 3 samples, see section 3.1) and subseafloor depth using the  
243 software PAST (Hammer et al., 2001).

244

## 245 3 Results

246

### 247 3.1 Proportions of Eukaryota in Shotgun, Phytobaits1 and HABbaits1

248 After filtering, we retained between 4.6 (GC2A 15 - 16.5 cm) and 16.2 M (GC2B 5 - 6.5 cm)  
249 reads per sample for Shotgun, between 0.1 (MCS3 4 - 5.5 cm) and 4.6 M (GC2A 115 -  
250 116.5 cm) reads per sample for Phytobaits1 and between 0.2 (GC2A 45 - 46.5 cm) and 2.8  
251 M (GC2A 115 - 116.5 cm) reads for HABbaits1. We retrieved no data for 3 out of 30  
252 samples and these samples were excluded from downstream processing. The 3 samples  
253 with no data were MCS3 0 - 1.5 cm with Shotgun, GC2B 115 - 116.5 cm with Phytobaits1  
254 and HABbaits1, and GC2A 85 - 86.5 cm with HABbaits1 - likely due to low template DNA  
255 concentrations. Our EBCs for Shotgun, Phytobaits1 and HABbaits1 detected a total of 121,  
256 69 and 28 eukaryote taxa (Supplementary Material Table 1), emphasising the importance of  
257 sequencing controls and filtering *sedaDNA* data accordingly to remove contaminants.

258

259 Based on alignments using the NCBI Nucleotide database, the majority of Shotgun reads  
260 were assigned to Bacteria ( $86 \pm 5\%$  and  $63 \pm 16\%$  for MCS3 and GC2, respectively; Fig.  
261 3a,b), and a relatively small portion to Eukaryota ( $5 \pm 2\%$  and  $28 \pm 15\%$  to for MCS3 and  
262 GC2, respectively, Fig. 3a,b). This small proportion of Eukaryota increased to 21 and 53%  
263 in MCS3 and GC2 using Phytobaits1 (4.4x and 1.9x over Shotgun, respectively), and 47  
264 and 76% in MCS3 and GC2 using HABbaits1 (9.6x and 2.7x over Shotgun respectively)  
265 (Fig. 3). Phytobaits1 and HABbaits1 were efficient in the targeted enrichment of Eukaryota  
266 *sedaDNA* from marine sediments, with comparatively little 'bycatch' of Bacteria and  
267 Archaea (i.e., a decrease in the proportion of Bacteria and a  $<2.1x$  increase in Archaea  
268 relative to Shotgun; Fig. 3c,d). Phytobaits1 included three cyanobacterial targets, therefore,  
269 some capture of bacterial sequences was expected; less than Shotgun but more than  
270 HABbaits1 (Fig. 3a,b).

271

### 272 3.2 Assessment of *sedaDNA* authenticity

273 For both inshore MCS3 and offshore GC2, the A:D ratio determined per sample increased  
274 with sub-seafloor depth for each of the three datasets Shotgun, Phytobaits1 and HABbaits1  
275 (Fig. 4). At the seafloor surface, we determined an A:D ratio of approximately 0.05 for  
276 MCS3 and GC2, which slightly increased with depth until ~25 cmbsf, before a steeper  
277 increase between ~25 - 35 cmbsf, and, in offshore GC2, remained relatively stable at ~0.3  
278 below 35 cmbsf (>1000 years of age). Correlation analyses showed that this increase of the  
279 A:D ratio with increasing subseafloor depth was highly significant for each dataset (Table  
280 2). Additionally, the A:D ratios of the three different datasets (Shotgun, Phytobaits1 and  
281 HABbaits1) were significantly positively correlated with each other, indicating that the  
282 original proportions of *sedaDNA* damage patterns preserved in Shotgun were maintained in  
283 our hybridisation capture approach using both Phytobaits1 and HABbaits1.

284

### 285 3.3 DNA damage profiles of the marine coccolithophore *Emiliana huxleyi*

286 The *sedaDNA* damage analysis provided DNA damage profiles for most of the target taxa  
287 on our selected taxa list (taxa list 'b'). However, the number of ancient sequences assigned  
288 to the ubiquitous coccolithophore *Emiliana huxleyi* was much higher, allowing the  
289 generation of more detailed DNA damage profiles. Ancient *E. huxleyi* sequences ranged  
290 from a total of 0 - 34 reads in inshore MCS3 and 5 - 2,651 in offshore GC2 for Shotgun,  
291 from 0 - 7 in MCS3 and 1 - 947 in GC2 for Phytobaits1, and from 0 - 7 in MCS3 and 1 -  
292 1183 in GC2 for HABbaits1. A lower representation of 'ancient' sequences in inshore MCS3  
293 is consistent with our observation of a lower A:D ratio in sediments above ~35 cmbsf (i.e.,  
294 the complete length of MCS3) (see section 3.2). Damage profiles for *E. huxleyi* *sedaDNA*  
295 are much more variable in inshore MCS3 (and in the upper ~25 cmbsf of GC2; Fig. 5,6)  
296 than the profiles of deeper, more stable offshore GC2 samples, likely resulting from a

297 scarcity of reads in the upper sediment layers and DNA damage patterns not being as  
298 pronounced as in deeper GC2 samples.

299

300 The *E. huxleyi* *sed*aDNA damage profiles of Shotgun, Phytobaits1 and HABbaits1 were  
301 consistent across samples (Fig. 5,6), suggesting that the hybridisation capture technique  
302 reliably preserves the DNA damage patterns of the original sample (represented by  
303 Shotgun) and is well-suited for the capture of past marine eukaryote *sed*aDNA. Further,  
304 HOPS provided a valid approach to authenticate *sed*aDNA from marine eukaryotes. We  
305 were unable to generate clear DNA damage profiles from the upper ~25 – 35 cmbsf in both  
306 MCS3 (spanning the last ~125 years; Fig. 5) and GC2 (~900 years, Fig. 6), indicating that  
307 DNA damage is not as pronounced in the upper (younger) sediment layers at our study  
308 location and detectable only below that depth. Below ~35 cmbsf in GC2 the *E. huxleyi* DNA  
309 damage profiles assumed a typical U-shape as the number of mismatches at the end of  
310 DNA fragments increases (Fig. 6). Our *E. huxleyi* *sed*aDNA damage profiles are the first  
311 generated for a marine eukaryote - and extend over an 8,000-year timescale.

312

#### 313 **4 Discussion**

314

315 In this study we designed two new RNA bait sets and applied the hybridisation capture  
316 technique to inshore and offshore marine sediments to investigate marine eukaryotes more  
317 broadly (Phytobaits1) and in a more tailored approach focusing on selected taxa common  
318 and often harmful in our study region (HABbaits1). Our results showed that hybridisation  
319 capture improved the genetic yield of eukaryote *sed*aDNA, and preserved DNA damage  
320 patterns that allowed us to make an assessment of *sed*aDNA authenticity, as well as

321 enabling us to generate the first ancient DNA damage profiles of a keystone marine  
322 phytoplankton organism, the ubiquitous coccolithophore *Emiliana huxleyi*.

323

#### 324 4.1 Targeted enrichment of marine Eukaryota sedaDNA

325 Both Phytobaits1 and HABbaits1 successfully captured sedaDNA of eukaryote organisms  
326 in two sediment cores collected off the Tasmanian east coast. Eukaryote sedaDNA has  
327 been shown repeatedly to be present in low amounts in seafloor sediments, which has  
328 limited the metagenomic analysis and detailed reconstruction of past marine ecosystems.  
329 While both Phytobaits1 and HABbaits1 achieved a considerable enrichment in Eukaryota  
330 for our inshore site MCS3 (4- to 9-fold, respectively), this increase was about half at the  
331 offshore site GC2. The difference in Eukaryota increase may be due to the initial difference  
332 in Eukaryota proportions at the two sites. Shotgun showed Eukaryota contributed ~5% to  
333 the total pool of sedaDNA at MCS3, while contributing ~28% at GC2. The latter high  
334 proportion is primarily a result of a sharp increase in the relative abundance of Eukaryota in  
335 GC2 below 35 cmbsf. It is possible that this initially relatively high presence of Eukaryota  
336 sequences in the GC2 sedaDNA extracts saturated the baits in our hybridisation reaction.  
337 This would explain a less pronounced increase in GC2 Eukaryota proportions using either  
338 bait set. To further increase the Eukaryota signal in future studies, it may be beneficial to  
339 add a larger volume of baits (>3  $\mu$ L) to sedaDNA extracts that either are expected or have  
340 been shown (e.g., by shallow shotgun sequencing prior to enriching) to have a relatively  
341 high Eukaryota sedaDNA content.

342

#### 343 4.2 Assessment of sedaDNA authenticity Interesting

344 The HOPS bioinformatic tool (Hübler et al., 2020) proved highly valuable in identifying and  
345 analysing ancient eukaryote sequences in our sedaDNA. The HOPS generated output of

346 our 'Eukaryota' (taxa list a) run enabled the determination of the A:D ratio, a parameter that  
347 can be used as a proxy of *sedaDNA* authenticity in the future. Here, the A:D ratio was quite  
348 low (0.05 – 0.3), which might point towards our *sedaDNA* being relatively well preserved,  
349 and thus a high proportion of reads passing the default filtering criteria. The latter criteria  
350 used a minimum percent identity (mpi) level of 95%, a relatively stringent cut-off while still  
351 retaining the majority of reads. Increasing the mpi cut-off may have resulted in a higher A:D  
352 ratio due to less reads passing the filtering criteria, however, this would also eliminate the  
353 majority of reads available for analyses.

354

355 At both our inshore and offshore site, we observed a significant increase in A:D ratio with  
356 subseafloor depth, demonstrating that eukaryote *sedaDNA* shows increased DNA damage  
357 with increasing age of sediments. However, at both sites the A:D ratio was consistently  
358 lower in the upper 25 - 35 cmbsf (~0.05), then increased sharply down-core from this depth  
359 in both (~0.3), and remained at this level towards the bottom of in GC2. Whether reaching  
360 such a 'limit' in the A:D ratio is a pattern characteristic of our study location, or indicative of  
361 a 'critical depth' below which *sedaDNA* degradation accelerates, remains to be  
362 investigated. Future *sedaDNA* studies should investigate how the A:D ratio varies in much  
363 older sediment records (older than Holocene) and depending on sediment properties (e.g.,  
364 clay-rich sediments that appear to benefit DNA preservation; Vuillemin et al., 2019).

365

366 The strong positive correlation between the A:D ratios amongst Shotgun, Phytobaits1 and  
367 HABbaits1, demonstrates that DNA damage signals present in *sedaDNA* are preserved  
368 throughout the hybridisation capture approach. This is important as it allows the  
369 authentication of *sedaDNA* using bioinformatic tools (see section 4.3), which any ancient  
370 DNA study should incorporate (Hübler et al., 2020). Through hybridisation capture more

371 target sequences are available as input for DNA damage analysis software such as HOPS,  
372 which increases the robustness of such analyses (Hübler et al., 2020), thus is strongly  
373 recommended for *sedaDNA* analyses. While future refinement of A:D ratios may be  
374 necessary, our analyses show that it can be used as a proxy for *sedaDNA* authenticity in  
375 sediment records. Generally, for marine *sedaDNA* investigations of eukaryote taxa, the  
376 capacity to assess DNA damage provides a crucial advantage over metabarcoding where  
377 DNA damage-based authenticity assessment is impossible.

378

### 379 4.3 *Emiliania huxleyi* *sedaDNA* damage profiles

380 Running our data through the HOPS pipeline (taxa list *b*) and MEx-IPA allowed us to  
381 generate DNA damage plots for a key marine phytoplankton species, *Emiliania huxleyi*.  
382 This ubiquitous calcareous nanoplankton has thrived in the oceans since the Cretaceous, is  
383 one of the most abundant phytoplankton species in the global ocean and is ubiquitous from  
384 tropical to temperate to Antarctic Australian waters (Hallegraeff 1984; Cubillos et al. 2007).  
385 Consistent with its biogeographic distribution in the modern ocean, we expected to detect  
386 traces of this species in our *sedaDNA*, and in higher relative abundances offshore.

387

388 We retained the maximum number of reads throughout our analyses (by examining  
389 proportions rather than rarefying our data), which enabled us to generate *E. huxleyi* DNA  
390 damage profiles from all three datasets, Shotgun, Phytobaits1 and HABbaits1. The damage  
391 profiles generated by Shotgun, Phytobaits1 and HABbaits1 per sample were very similar,  
392 indicating preservation of DNA damage patterns in our original sample (Shotgun) and in our  
393 enriched samples after hybridisation capture. Consistent with our finding of low A:D ratio in  
394 the upper 25 - 30 cmbsf, no clear *E. huxleyi* damage patterns could be determined from  
395 these depths. *sedaDNA* damage patterns with a typical U-shape were found only below



396 ~25 cmbsf in GC2, again suggesting the existence of a critical depth below which DNA  
397 degradation becomes more pronounced, reinforcing the importance of investigating  
398 whether this phenomenon is of wider importance, and possibly correlated with the age or  
399 physical or chemical properties of marine sediments.

400

#### 401 *4.4 Significance of hybridisation capture and marine sedaDNA damage analysis*

402 The study of marine *sedaDNA* offers huge potential for the comprehensive reconstruction of  
403 past marine ecosystems (including viruses, archaea, prokaryotes and eukaryotes).  
404 Eukaryotes (phytoplankton and higher organisms) are particularly popular study organisms  
405 due to their importance as primary producers and use as environmental indicators.  
406 However, *sedaDNA* studies focussing on eukaryotes have been severely limited by the very  
407 low amounts of DNA preserved in the subseafloor, and the lack of bioinformatic tools to  
408 authenticate these miniscule amounts of eukaryote *sedaDNA* in metagenomic data. To  
409 date, no marine *sedaDNA* study exists that had proven authenticity (i.e., the DNA recovered  
410 is ancient and free from modern contamination) through bioinformatic approaches such as  
411 *sedaDNA* damage analysis, a routine procedure in ancient DNA studies focussing on  
412 humans and megafauna. Our study provides a key advance in that we (1) used a  
413 hybridisation capture technique to enrich target marine eukaryote *sedaDNA* independent of  
414 DNA fragment size, and (2) applied the recently developed bioinformatic tool HOPS for  
415 *sedaDNA* damage analysis and to authenticate our marine *sedaDNA*. These advances are  
416 of importance as we are now able to bioinformatically discriminate the authentic *sedaDNA*  
417 component to more accurately estimate paleo-community composition.

418

#### 419 **Conclusions**

420

421 In this study we show the reliability of the hybridisation capture as a novel tool for  
422 investigating changing patterns of abundance of marine eukaryotes from their *sedaDNA* in  
423 seafloor sediments. We furthermore applied a new bioinformatic approach for metagenomic  
424 *sedaDNA* damage analysis, which allowed us to develop a new proxy for *sedaDNA*  
425 authenticity (the A:D ratio) that changes with subseafloor depth. Through our *sedaDNA*  
426 damage analysis we were also able to generate *sedaDNA* damage profiles of the ubiquitous  
427 coccolithophore *E. huxleyi*, the first ever such profiles generated for a marine eukaryote –  
428 extending over an 8,000-year timescale. Our study provides a major step forward for the  
429 future investigation of eukaryotes from marine *sedaDNA*, enabling detailed insights into  
430 past marine ecosystem composition over geological timescales.

431

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433

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444

#### 445 **Author contributions**

446 L.A. designed this research, carried out laboratory work, bioinformatic and statistical  
447 analyses and wrote the first draft of the manuscript. G.H., C.B., and C.W. collected and  
448 provided the core samples. C.W. provided sediment core dating. A.C. provided guidance on  
449 the hybridisation capture technique and ancient DNA analyses. G.H., C.B. and A.C.  
450 assisted with data interpretation and secured funding for this project. All co-authors  
451 provided comments and feedback on manuscript drafts, and edited the final manuscript  
452 submitted for publication.

453

#### 454 **Competing interests**

455 All co-authors declare that there are no competing interests.

456

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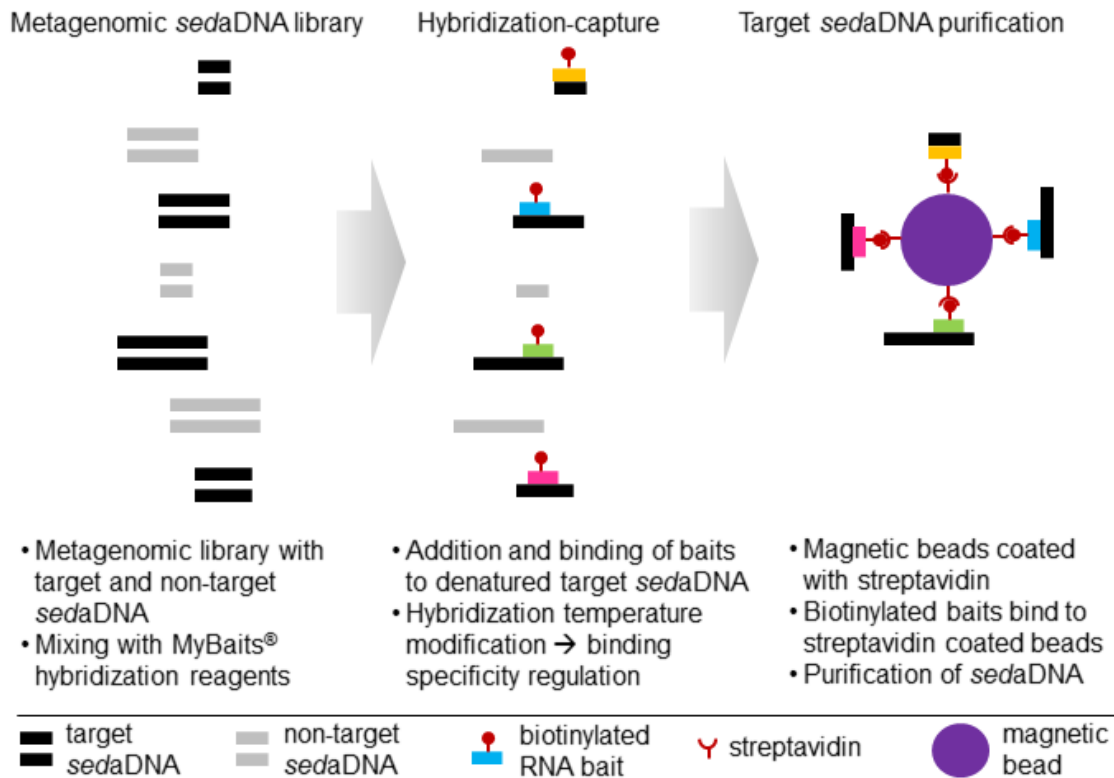
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592 **Main Text Figures and Tables**

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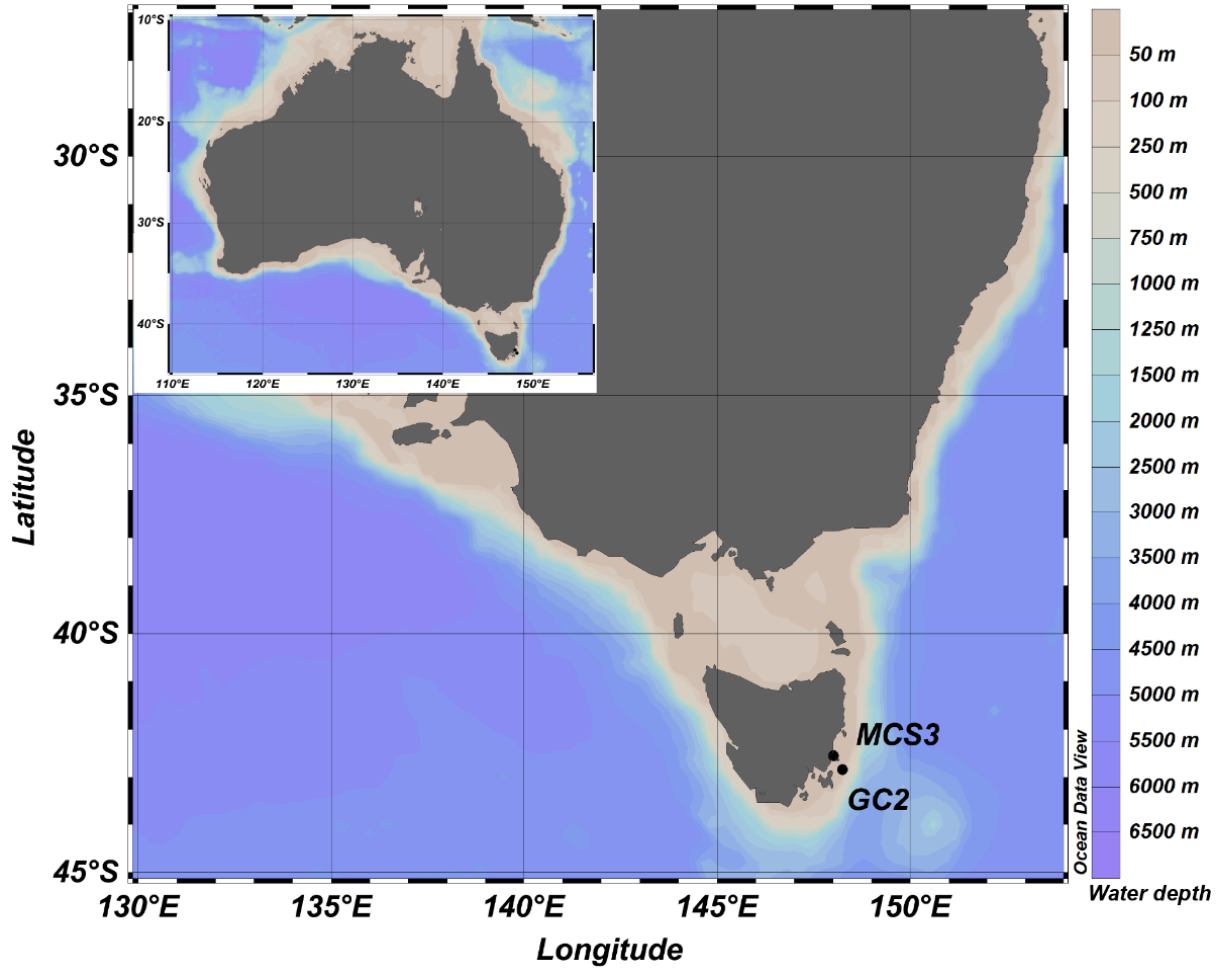


595

596 **Figure 1: Schematic of hybridisation capture applied to marine sedimentary ancient**  
597 **DNA (*sedaDNA*).** The three main steps are the preparation of a metagenomic *sedaDNA*  
598 library, hybridisation capture using RNA baits (in this study: Phytobaits1 and HABbaits1)  
599 that are biotinylated, which enables binding of baits to streptavidin-coated magnetic beads  
600 (multiple baits per bead possible, schematic not to scale). For further technical details see  
601 Methods, MyBaits® Manual V4.01 (2018), and Horn et al. (2012).  
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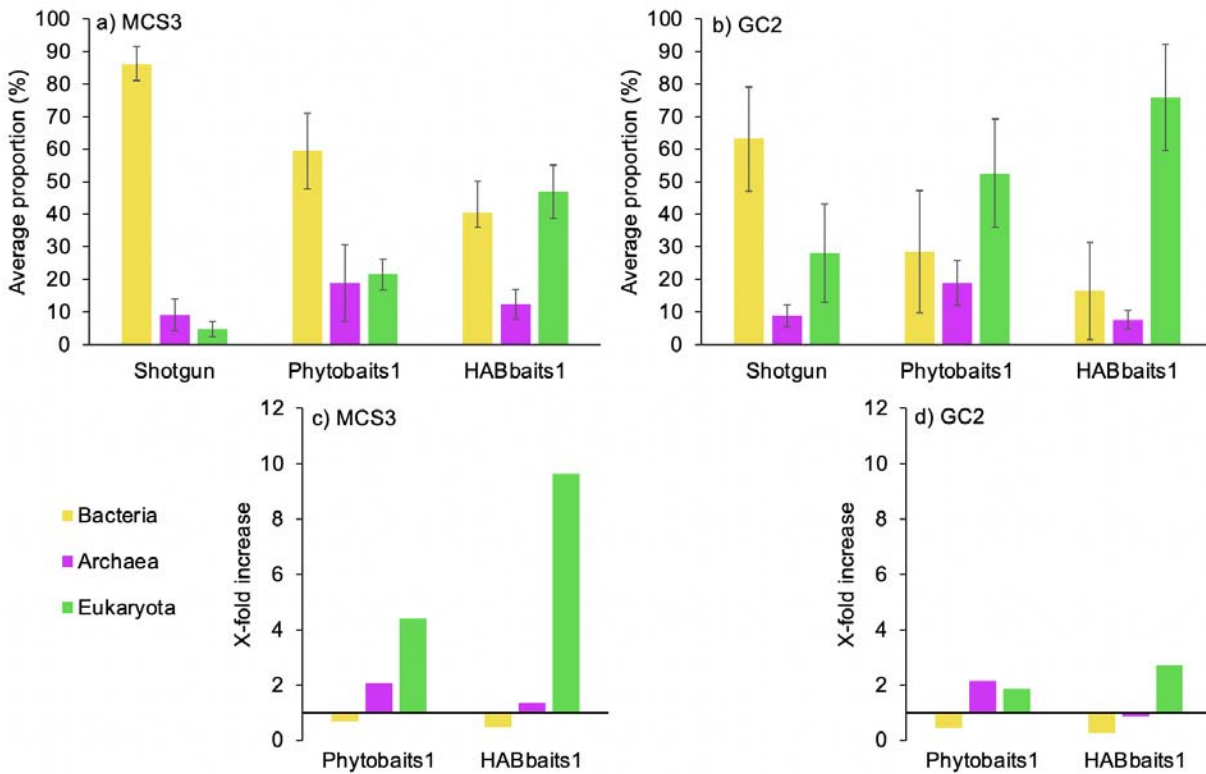
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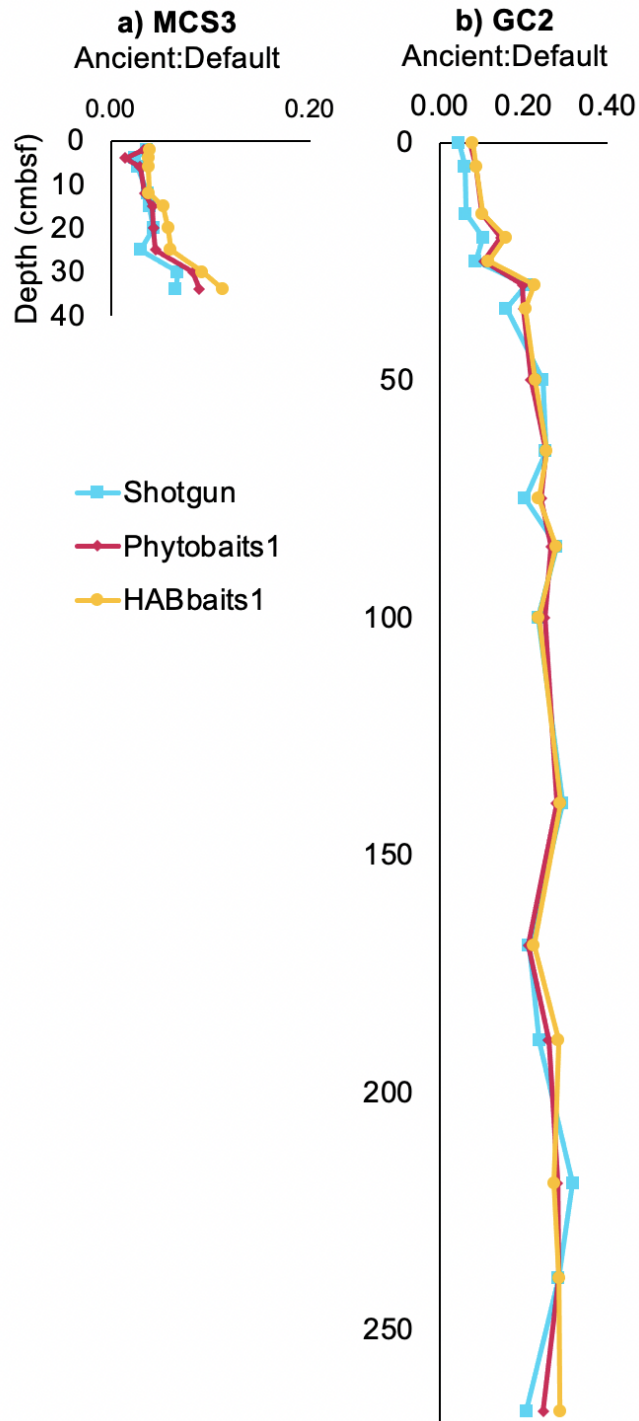
**Figure 2: Map of coring sites, inshore (MCS3) and offshore (GC2B) of Maria Island, Tasmania, South-East Australian Coast.** Map created in ODV (Schlitzer, R., Ocean Data View, <https://odv.awi.de>, 2018).

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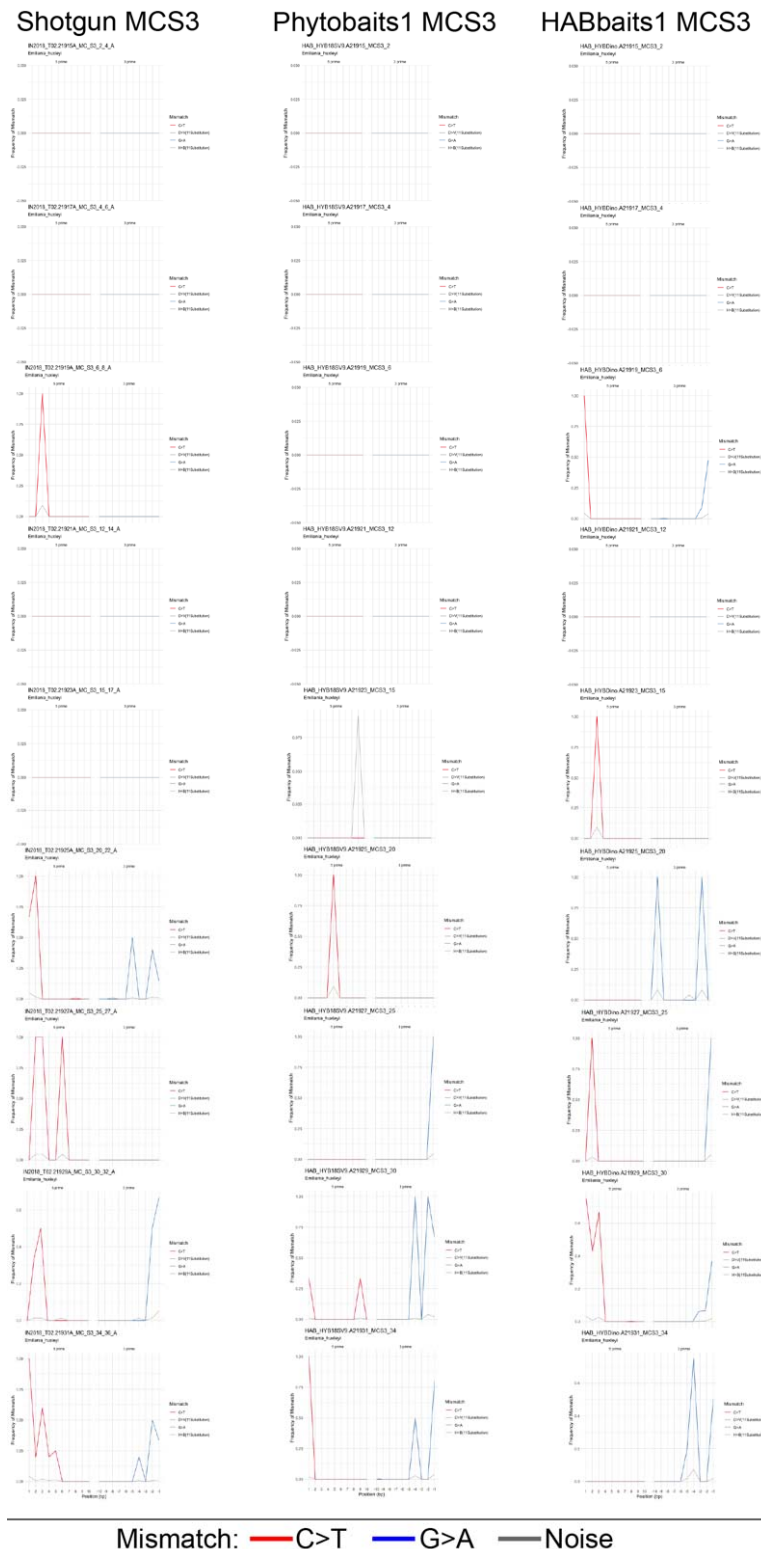
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614 **Figure 3: Proportions of reads assigned to Bacteria, Archaea and Eukaryota using**  
615 **Shotgun, Phytobaits1 and HABbaits1.** a,b) Average proportion of reads and standard  
616 deviation across inshore MCS3 (n = 9) and offshore GC2 (n = 18) samples, respectively.  
617 c,d) Increase in the proportion of Bacteria, Archaea and Eukaryota in Phytobaits1 and  
618 HABbaits1 relative to Shotgun for MCS3 and GC2 samples, respectively, based on average  
619 proportions shown in (a,b).  
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**Figure 4: A:D ratio of reads assigned to Eukaryota with subseafloor depth.** Shown is the increase in the A:D ratio with depth (centimetres below seafloor, cmbsf) in both a) MCS3, b) GC2. See Table 1 for correlation between A:D ratio per dataset (Shotgun, Phytobaits1, HABbaits1) and depth, and amongst the datasets.

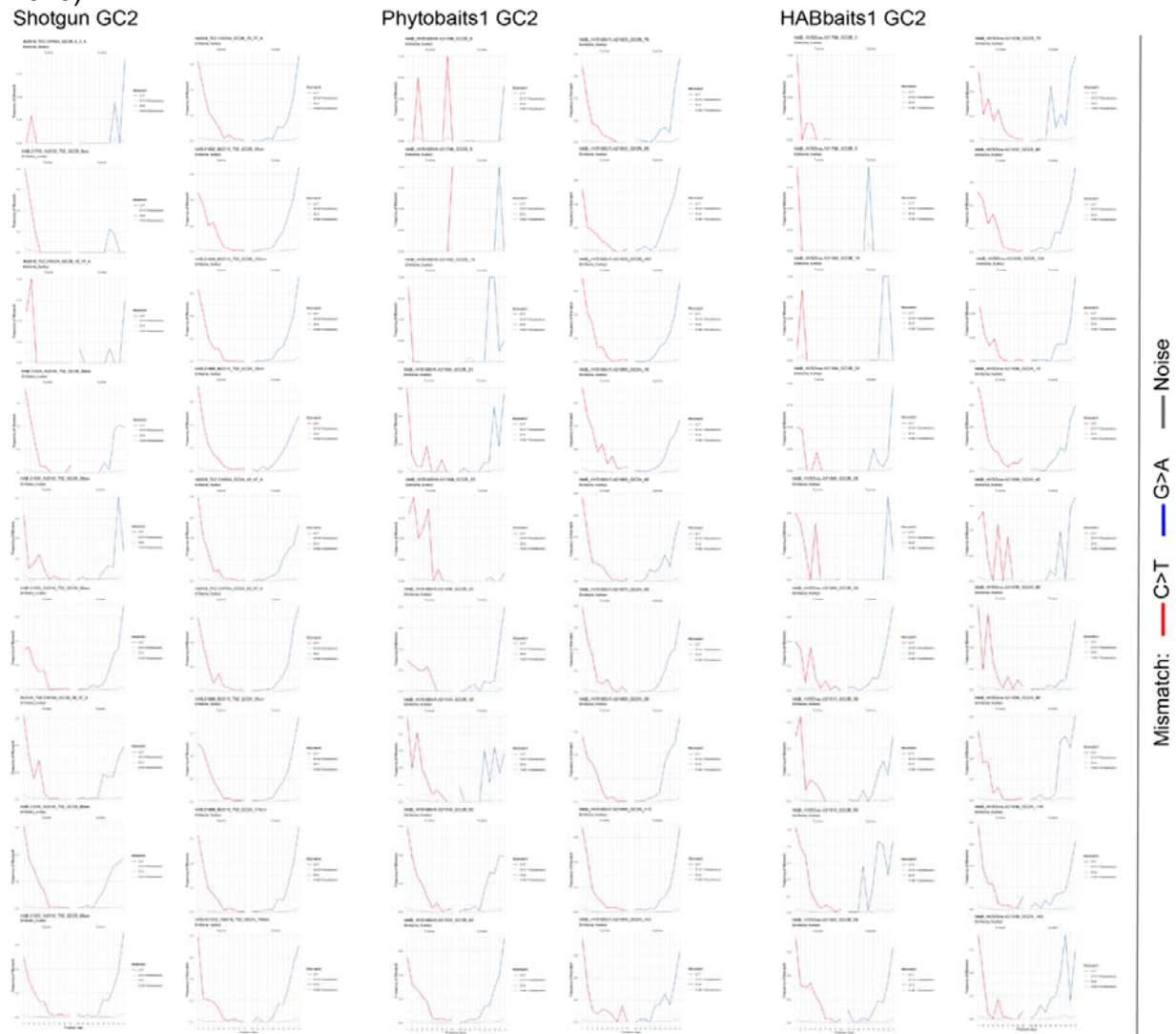


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Mismatch: — C>T — G>A — Noise

628 **Figure 5: SedaDNA damage profiles of *Emiliana huxleyi* in MCS3.** *E. huxleyi* *sedaDNA*  
 629 damage profiles (frequency of mismatch against base pair position) per sample for  
 630 Shotgun, Phytobaits1 and HABbaits1 in MCS3 (listed from top-down). The red and blue  
 631 lines denote C>T substitutions in 5' direction and G>A substitutions in 3' direction,

632 respectively, for all ancient alignments. Grey lines denote estimated noise (Hübler et al.,  
633 2020).



634

635 **Figure 6: SedaDNA damage profiles of *Emiliana huxleyi* in GC2.** *E. huxleyi* *sedaDNA*  
636 damage profiles (frequency of mismatch against base pair position) per sample for  
637 Shotgun, Phytobaits1 and HABbaits1 in GC2 (listed from top-down with GC2 profiles  
638 continuing in the second column for each dataset). The red and blue lines denote C>T  
639 substitutions in 5' direction and G>A substitutions in 3' direction, respectively, for all ancient  
640 alignments. Grey lines denote estimated noise (Hübler et al., 2020).

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648 **Table 1: Phytobaits1 and HABbaits1.** Target taxa of Phyto- and HABbaits1 genes/gene  
 649 regions and source databases. For HABbaits1, all listed databases were searched for each  
 650 gene (region) per target taxon, and, if available, the longest sequence was selected and  
 651 included.

Bait set	Target taxa	Targeted gene/ gene region	Database from which sequences were acquired
Phytobaits1	Ciliophora, MALV, Dinophyceae, Archaeplastida, Euglenida, Telonemia, Haptophyta, Cryptophyta, Katablepharidophyta, Chlorarachnea, Phaeodarea, Foraminifera, Acantharea, Other_Radiolaria, RAD, Collodaria, MAST, Bicoeca, MOCH, Raphidophyceae, Pinguiphyceae, Phaeophyceae, Chrysophyceae- Synurophyceae, Pelagophyceae, Dictyochophyceae, Bolidophyceae-and- relatives, Bacillariophyta	18SV9	W2_PR2_V9 (De Vargas et al., 2015)
	<i>Trichodesmium erythraeum, Prochlorococcus marinus, Synechococcus sp.</i>	16SV4	SILVA <a href="https://www.arb-silva.de/">https://www. arb-silva.de/</a>
HABbaits1	<b>Dinoflagellates</b>		LSU: SILVA;
	<i>Alexandrium tamarensense</i> Group 1 ( <i>A. catenella</i> )	LSU, D1D2, ITS, COI	SSU: PR <sup>2</sup> (Guillou et al., 2013) or NCBI
	<i>Alexandrium tamarensense</i> Group 2 ( <i>A. mediterraneum</i> )	LSU, D1D2, ITS, COI	( <a href="https://www.ncbi.nlm.nih.gov/">https://www.ncbi.nlm.nih.gov/</a> );
	<i>Alexandrium tamarensense</i> Group 3 ( <i>A. tamarensense</i> )	LSU, D1D2, ITS, COI	D1D2: PHYTOPK28S-D1D2 (Grzebyk et al., 2017);
	<i>Alexandrium tamarensense</i> Group 4 ( <i>A. pacificum</i> )	LSU, D1D2, ITS, COI	ITS: BOLD
	<i>Alexandrium tamarensense</i> Group 5 ( <i>A. australiense</i> )	LSU, D1D2, ITS, COI	( <a href="http://www.boldsystems.org/">http://www.boldsystems.org/</a> , Ratnasingham and Herbert, 2007) or NCBI;
	<i>Gymnodinium catenatum</i>	LSU, D1D2, ITS, COI	

<i>Noctiluca scintillans</i>	LSU, D1D2	
<i>Tripes (Ceratum) furca</i>	LSU, SSU, D1D2	rbcl: BOLD;
<i>Tripes (Ceratum) fusus</i>	LSU, SSU, D1D2, COI	COI: BOLD or NCBI
<i>Tripes</i> sp. (genus)	SSU	
<i>Tripes muelleri</i>	LSU, SSU	
<b>Diatoms</b>		
<i>Pseudo-nitzschia</i> sp. (genus)	LSU, D1D2, SSU, ITS	
<i>Pseudo-nitzschia cuspidata</i>	LSU, D1D2, ITS, rbcl	
<i>Pseudo-nitzschia pungens</i>	LSU, D1D2, ITS, rbcl	
<b>Haptophytes</b>		
<i>Emiliania huxleyi</i>	LSU, D1D2, rbcl, COI	
<b>Cnidarians</b>		
<i>Aurelia</i> spp.	LSU, D1D2, ITS, COI	
<i>Cyanea</i> spp.	LSU, ITS, COI	
<i>Physalia</i>	LSU, ITS, COI	
<b>Molluscs</b>		
<i>Crassostrea gigas</i>	LSU, D1D2, ITS, COI	
<i>Ostrea angasi</i>	LSU, COI	
<i>Mytilus galloprovincialis</i>	LSU, D1D2, ITS, COI	
<i>Modiolus</i> spp.	LSU, D1D2, ITS, COI	
<b>Genes involved in toxin production</b>		
<i>SxtA</i>		

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654 **Table 2: Summary statistics of correlation analysis between the A:D ratio and**  
655 **subseafloor depth.** Pearson correlation coefficients  $r$  (in italics, lower matrix triangle) and  
656 corresponding two-tailed probability that  $r$  is uncorrelated (upper triangle of matrix; *i.e.*, all  
657 values  $<0.005$  denote a significant correlation) between subseafloor depth (cmbsf) and  
658 Shotgun, Phytobaits1 and HABbaits1 A:D ratios ( $n = 27$  each).

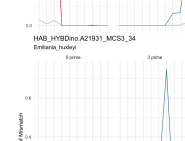
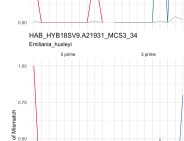
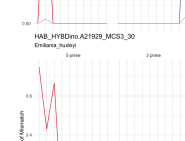
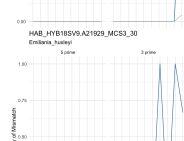
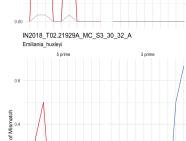
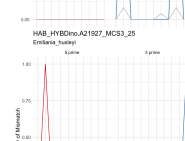
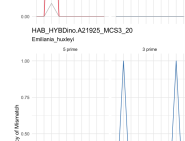
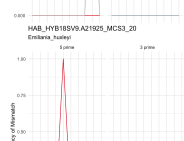
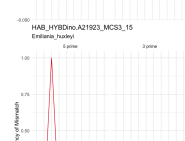
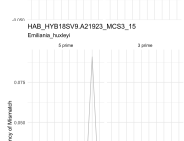
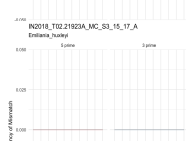
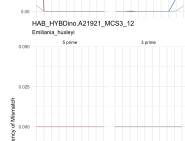
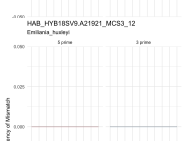
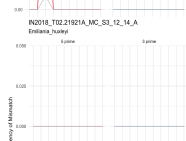
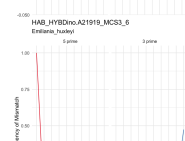
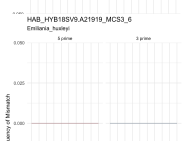
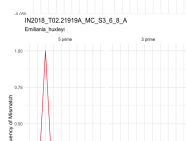
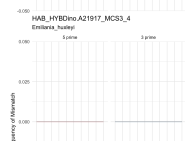
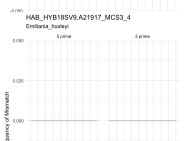
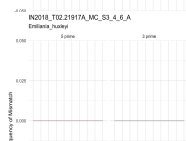
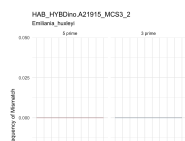
<b>MCS3</b>	<b>Depth (cmbsf)</b>	<b>Shotgun A:D</b>	<b>Phytobaits1 A:D</b>	<b>HABbaits1 A:D</b>
<b>Depth (cmbsf)</b>		0.01180	0.00097	0.00054
<b>Shotgun A:D</b>	<i>0.78722</i>		0.00009	0.00104
<b>Phytobaits1 A:D</b>	<i>0.89909</i>	<i>0.94966</i>		2.82E-05
<b>HABbaits1 A:D</b>	<i>0.91497</i>	<i>0.89704</i>	<i>0.96396</i>	
<b>GC2</b>	<b>Depth (cmbsf)</b>	<b>Shotgun A:D</b>	<b>Phytobaits1 A:D</b>	<b>HABbaits1 A:D</b>
<b>Depth (cmbsf)</b>		0.00171	0.00053	0.00025
<b>Shotgun A:D</b>	<i>0.68497</i>		5.16E-11	2.31E-09
<b>Phytobaits1 A:D</b>	<i>0.73351</i>	<i>0.96787</i>		2.20E-13
<b>HABbaits1 A:D</b>	<i>0.76070</i>	<i>0.94793</i>	<i>0.98386</i>	

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# Shotgun MCS3

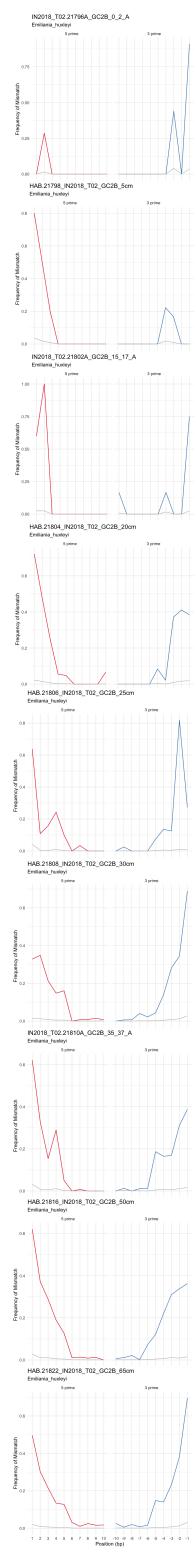
# Phytobaits1 MCS3

# HABbaits1 MCS3

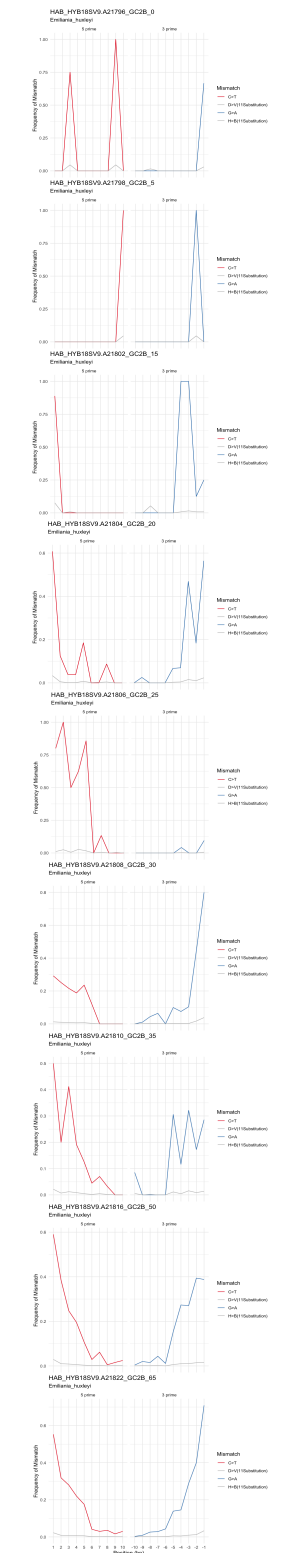
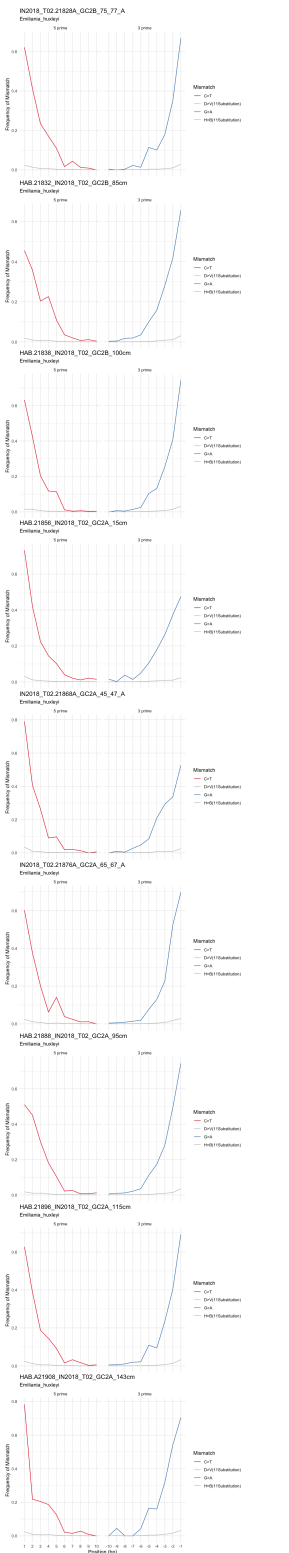


Mismatch: — C>T — G>A — Noise

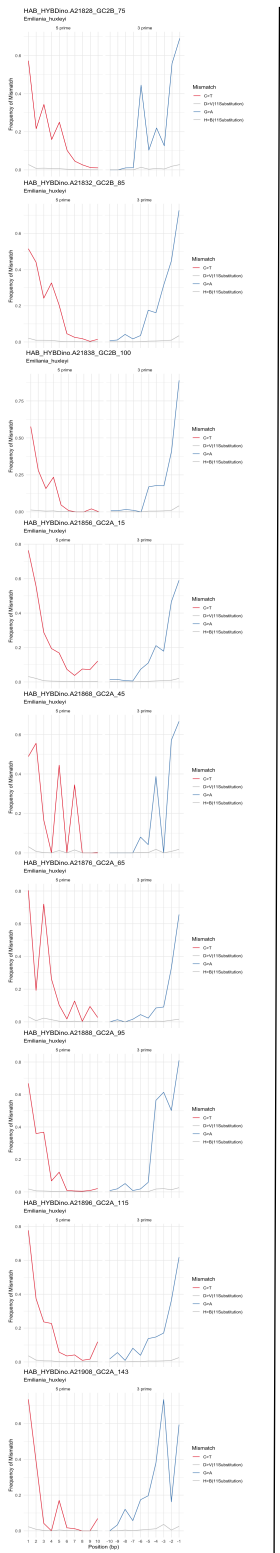
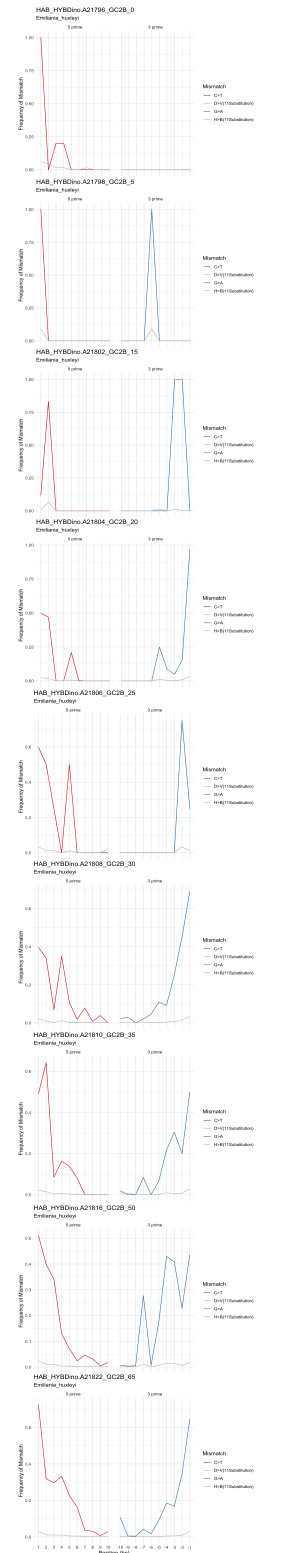
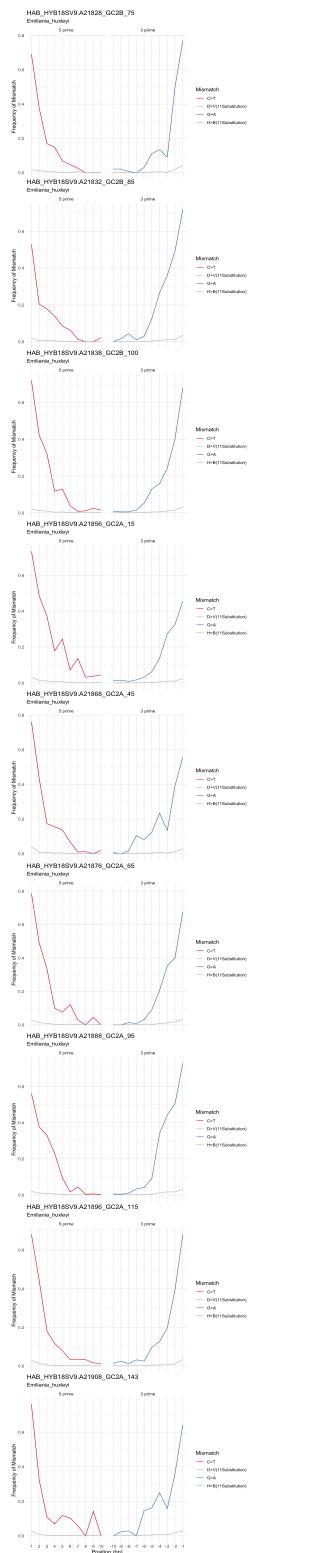
# Shotgun GC2



# Phytobaits1 GC2



# HABbaits1 GC2



Mismatch: — C>T — G>A — G>C — A>T — Noise