

1 **Pronounced phenotypic differentiation in a wide-dispersing marine**
2 **snail in response to contrasting selection pressures at a local scale**

3

4 Nicolás Bonel^{1,2,3*}, Jean-Pierre Pointier⁴ & Pilar Alda^{1,2}

5

6

7

8 ¹ Laboratorio de Zoología de Invertebrados I, Departamento de Biología, Bioquímica y Farmacia,
9 Universidad Nacional del Sur (UNS), San Juan 670, (B8000ICN) Bahía Blanca, Argentina.

10

11

12 ² Consejo Nacional de Investigaciones Científicas y Técnicas (CONICET), Argentina.

13

14

15 ³ Centre d'Écologie Fonctionnelle et Évolutive, UMR 5175, CNRS—Université de Montpellier,
16 Université Paul-Valéry Montpellier—École Pratique des Hautes Études—IRD, 34293 Montpellier Cedex
17 05, France.

18

19 ⁴ PSL Research University, USR 3278 CNRS–EPHE, CRIOBE Université de Perpignan, Perpignan,
20 France.

21

22 * To whom correspondence should be addressed: nicobonel@gmail.com (N. Bonel)

23

24

25

26

27 **Running head:** Pronounced phenotypic differentiation in a wide dispersing snail

28

29

30

31

32 **Key words:** phenotypic plasticity, intertidal zonation, contrasting selection pressures,
33 shell characters, genital morphology, high dispersal potential, planktotrophic snail.

34

35 **Abstract**

- 36 1. Heterogeneous environments pose a particular challenge for organisms because the
37 same phenotype is unlikely to perform best regardless of the types of stress it
38 encounters. The grain size theory predicts that species with high dispersal potential
39 experience a more heterogeneous, fine-grained environment where phenotypic
40 plasticity may evolve to cope with habitat heterogeneity.
- 41 2. To understand how species meet this challenge, we investigated the extent to which
42 contrasting selection pressures induced ecological and phenotypic responses in a
43 natural population of a wide-dispersing marine snail.
- 44 3. We collected, measured external and internal characters, weighted, and dissected
45 individuals of *Heleobia australis* (Rissooidea: Cochliopidae) from heterogeneous
46 habitats from the intertidal area of the Bahía Blanca estuary, Argentina. We also
47 conducted molecular analyses by amplifying the COI gene in individuals sampled
48 from each habitat.
- 49 4. We found that subpopulations of *H. australis*, inhabiting close to each other and
50 without physical barriers, exhibited a strong phenotypic differentiation in shell
51 characters and body weight in response to environmental conditions (thermal, saline,
52 and dehydration stress), crab predation, and parasites. We proved that this
53 differentiation occurred even early in life as most of the characters observed in
54 juveniles mirrored those found in adults. We also found a clear variation in penis size
55 in snails collected from each habitat and raised in common garden laboratory
56 conditions. The COI gene analysis confirmed that the individuals studied constituted
57 a single species despite the strong phenotypic difference among subpopulations.
- 58 5. The pronounced phenotypic differentiation in *H. australis* is all the more remarkable
59 because it occurred at a very small geographical scale, which is rarely documented
60 for a wide-dispersing species. Our findings provide a reasonable ground for
61 advocating that *H. australis* experienced a fine-grained environment and, thus,
62 benefited from the combined effect of directional selection and plasticity to evolve
63 locally adapted phenotypes to contrasting habitat conditions at a local scale.

64 **Introduction**

65 If organisms have difficulties in adapting to human-induced global change and fail to
66 track projected environmental changes, populations become vulnerable to decline and
67 extinction (Hill, Griffiths, & Thomas, 2011; Hoffmann & Sgrò, 2011). Phenotypic
68 plasticity, a common feature in nature across many taxa, is the ability of a single
69 genotype to express different phenotypes in dissimilar biotic or abiotic environments
70 (Travis, 1994; West-Eberhard, 2003). Many studies indicate a role of plasticity in
71 shaping phenotypic responses as an effective mechanism that can greatly improve the
72 conditions for persistence of populations facing abrupt environmental shift by
73 facilitating rapid adaptation to the new selection pressures (e.g. Chevin & Lande, 2009;
74 Ghalambor, McKay, Carroll, & Reznick, 2007; Price, Qvarnström, & Irwin, 2003). An
75 adaptive evolutionary response to environmentally induced non-heritable variation
76 occurs when phenotypic plasticity operates jointly with directional selection causing the
77 expression of potentially beneficial traits (De Jong, 2005; Fusco & Minelli, 2010;
78 Merilä & Hendry, 2014) and allowing the population to reach the fitness optimum (i.e.
79 ‘genetic assimilation’; Pigliucci, 2006; Waddington, 1953; West-Eberhard, 1989). In
80 this sense, phenotypic plasticity is crucial for the long-term persistence of populations
81 facing new environmental stress brought about by human-induced global change (Hill et
82 al., 2011).

83

84 Strategies to cope with environmental variation will depend upon the organism’s
85 dispersal ability since it sets the environmental grain and plays a crucial role in
86 determining levels of gene flow among populations, and thus, in the evolution of
87 phenotypic traits within those populations (De Jong, 1999, 2005; Levins, 1968; Mather,
88 1955; Scheiner, 1998; Via & Lande, 1985). According to the grain-size model, the
89 evolution of traits will follow different pathways in species with contrasting abilities to
90 disperse. For instance, aquatic snails display a wide range of reproductive modes and
91 larval development types, ranging from species with immobile larvae to those with
92 planktotrophic (free-floating) larvae (Janson, 1987; Johannesson, 2003; Parsons, 1997,
93 1998; Yamada, 1987). Specifically, snails with restricted dispersal (absence of
94 planktotrophic larvae) may experience a coarse-grained environment, as they tend to
95 stay in the same habitat all their life, and their offspring are likely to remain in that
96 habitat. In such conditions, local adaptation may be ensured by genetic polymorphism

97 because immigration is rare and the intrapopulation gene pool is highly reliant upon
98 local recruitment whereat individuals are exposed to local selection pressures giving rise
99 to different ecotypes. By contrast, wide-dispersing species will experience greater gene
100 flow and encounter different types of habitats more or less at random, thus experiencing
101 a fine-grained environment and exhibiting greater genetic homogeneity than in
102 relatively immobile species. This corresponds to a generalist strategy where phenotypic
103 plasticity may evolve to cope with heterogeneous environments (Bourdeau et al., 2015;
104 Chapman, 1995; Conde-Padín, Grahame, & Rolán-Alvarez, 2007; Hollander, Collyer,
105 Adams, & Johannesson, 2006; Johannesson, 2003; Levins, 1968; Reid, 1996; Vermeij,
106 1982; Yamada, 1987).

107

108 Temperate marine habitats, in particular the intertidal zone, exhibit great variability in
109 environmental factors mainly driven by the fall and rise of the tides, creating areas on
110 the shore that are alternately immersed and exposed (Bourdeau et al., 2015; Denny &
111 Paine, 1998). The transitional nature of the intertidal habitat from marine to terrestrial
112 conditions strongly influences the physiology and the ecology of intertidal organisms
113 due to the increase in environmental harshness (e.g. desiccation, extremes in
114 temperature and salinity) along the vertical zonation of the intertidal area (Denny &
115 Paine, 1998; Mouritsen, Sørensen, Poulin, & Fredensborg, 2018; Vermeij, 1972). From
116 this perspective, intertidal organisms, such as aquatic gastropods, are likely to exhibit
117 increased phenotypic plasticity in response to such environmental selective pressures
118 (Berrigan & Scheiner, 2004).

119

120 Phenotypic variation in shell traits is particularly widespread in gastropods (Bourdeau et
121 al., 2015) and their shells offer an easily measured and permanent record of how the
122 organism responded to local abiotic (e.g. water chemistry, temperature) and biotic (e.g.
123 predation, parasitism) agents (Dillon, Jacquemin, & Pyron, 2013). Major environmental
124 factors as prolonged submersion times, desiccation, high temperature, and extreme
125 salinity across the vertical zonation create contrasting selective pressures that give rise
126 to opposite shell trait responses (Johannesson & Johannesson, 1996; Struhsaker, 1968).
127 For instance, prolonged submersion times enhance foraging times and the absorption of
128 calcium carbonate from water leading to higher shell growth rates, which result in larger
129 and narrower shells and increased body mass (Vermeij, 1973; review by Chapman,

130 1995). By contrast, snails from upper intertidal habitats that are periodically exposed to
131 hot and dry conditions and have only a limited time each day to feed exhibit the
132 opposite shell and body mass responses and are characterized by having a smaller
133 aperture size, which has been correlated to an increase in resistance to high desiccation
134 in upper intertidal habitats (Chapman, 1995; Machin, 1967; Melatunan, Calosi, Rundle,
135 Widdicombe, & Moody, 2013; Schweizer, Triebkorn, & Köhler, 2019; Vermeij, 1973).

136
137 Predators are one component of the heterogeneity across the intertidal zone because
138 they are often distributed patchily in time and space creating a steep selection gradient
139 in predation risk, which, in turns, can promote an adaptive evolutionary shell response
140 (Boulding & Hay, 1993; Bourdeau, 2011). Predator-induced plasticity primarily linked
141 to shell defenses are not limited to shell thickening (more resistant to crushing) as snails
142 are capable of plastically altering different aspects of their shell shape (i.e. increased
143 globosity; lower shell length to width ratio) or aperture (i.e. more elongated; higher
144 aperture length to width ratio), which prevents the crab from pulling the soft parts out of
145 the shell (Bourdeau et al., 2015; Dillon et al., 2013; Johannesson, 2003). However, not
146 all phenotypic plasticity triggered by environmental conditions is adaptive. Some plastic
147 trait responses, usually those imposed by the biochemistry and physiology of the
148 organism, can be reversed over short time scales, which is not the case for
149 developmental plasticity as it tends to be irreversible or takes longer to be reversed
150 (Pigliucci et al., 2006). Parasites, for instance, can induce morphological, behavioral,
151 and physiological change of individual snail hosts thereby influencing several aspects of
152 host life history that can significantly alter size-structure, demography, resource-use,
153 and intra and interspecific interactions of host population (Fredensborg, Mouritsen, &
154 Poulin, 2005; Miura, Kuris, Torchin, Hechinger, & Chiba, 2006; Mouritsen et al.,
155 2018). They have been also reported to induce changes in microhabitat choice (Curtis,
156 1987) and in body size in snails (Alda, Bonel, Cazzaniga, & Martorelli, 2010; Levri,
157 Dillard, & Martin, 2005; Miura et al., 2006; Mouritsen & Jensen, 1994; Probst & Kube,
158 1999). Such pressure exerted by parasites also differs along the intertidal zone (Alda et
159 al., 2010; Alda, Bonel, Cazzaniga, Martorelli, & Lafferty, 2019; Smith, 2001) and can
160 cause important shifts in the expression of host phenotypic traits, creating pronounced
161 phenotypic differences between infected and uninfected hosts (Fredensborg, Mouritsen,
162 & Poulin, 2006; Poulin & Thomas, 1999).

163

164 These contrasting conditions along the vertical distribution pose a particular challenge
165 for intertidal organisms: the same phenotype is unlikely to perform best regardless of
166 the types of stress it encounters. We can then have as a result (i) the same phenotype
167 everywhere (probably with maladaptation), (ii) a phenotypic plasticity promoting
168 different phenotypes under different conditions, or (iii) a local genetic differentiation,
169 contributing to the phenotypic variation between habitats, favored by restricted gene
170 flow through habitat selection (and probably partial genetic isolation leading to
171 sympatric speciation or cryptic species). Limited gene flow can occur when habitat
172 conditions disrupt dispersal of planktotrophic larvae thereby increasing localized
173 recruitment (Johnson, Watts, & Black, 1994) whereat individuals would be subjected to
174 local (a)biotic selective pressures (Johannesson & Johannesson, 1996; Struhsaker,
175 1968). Hence, to understand how species meet this challenge, the first step is to observe
176 phenotypic variation and its association in the field with microhabitats characterized by
177 different sources of stress. If we observe different morphs associated with different
178 conditions, the next step is to explore whether such phenotypic variation results from a
179 direct impact of stress (e.g. trait shift induced by parasites) or is it because we are
180 dealing with a cryptic or incipient sympatric species that find alternative niches (e.g.
181 different microhabitats) preventing competition.

182

183 Here we investigated how different selective pressures induced ecological and
184 phenotypic responses in an intertidal mud snail with planktotrophic larval dispersal. For
185 our study, we collected and measured external and internal characters, weighted, and
186 dissected the snail *Heleobia australis* (Rissooidea: Cochliopidae) over the four seasons
187 throughout a year from three distinct habitats from the intertidal area of the Bahía
188 Blanca estuary, Southwestern Atlantic, Argentina. The (a)biotically contrasting habitat
189 conditions of the study area create the potential for niche differences that could affect
190 snail density and induce different phenotypic responses among individuals from each
191 subpopulation. Finding strong evidence of phenotypic differentiation in external
192 characters (shell traits and body mass), we then investigated whether internal characters
193 as the genital morphology varied among snails from each subpopulation. We found that
194 juveniles collected from each habitat, and raised in common garden laboratory
195 conditions until they were adults, showed a remarkable difference in penis size. We also

196 explored whether individuals living in these contrasting habitats belong to the same
197 species, and not to a species complex, by amplifying the cytochrome oxidase subunit 1
198 (COI) gene. We confirmed that the individuals here studied constitute a single *Heleobia*
199 species. Collectively, our findings clearly advocate that the strong phenotypic
200 differentiation in *H. australis* might have resulted from the combined effect of
201 directional selection and plasticity that led to the evolution of locally adapted
202 phenotypes in response to contrasting selection pressures at a local scale.

203

204 **Material and Methods**

205 **Study species**

206 The intertidal mud snail *Heleobia australis* has a wide geographic range inhabiting
207 marine and estuarine ecosystems from tropical to temperate regions (from Brazil: 22°
208 54' S, to Argentina: 40° 84' S; De Francesco & Isla, 2003) and is the most common
209 benthic macrofaunal species in the estuaries and coastal lagoons. It is a gonochoristic
210 species with internal fertilization (Neves, Valentin, & Figueiredo, 2010). Mature
211 females lay egg capsules, containing one fertilized egg (N. Bonel, *personal*
212 *observation*, but see Neves et al., 2010), preferably on shells of conspecific
213 individuals, but they may also be laid on shells of other species, on sand grains, or
214 algae, which later develop into a veliger larva that exhibits a short pelagic larval life
215 (Neves et al., 2010), which is, on average, 10±3 days in standard laboratory conditions
216 (12:12 photoperiod, 25 °C, water salinity 30 PSU, and *ad libitum* food in the form of
217 boiled ground lettuce; N. Bonel *personal observation*). Adult snails attain a maximum
218 shell size ranging from 7 to 8 mm and individuals have an averaged shell length of 4.7
219 mm (1.1 SD) with a lifespan of 2.9 years in a temperate climate and two well-defined
220 recruitment episodes (Carcedo & Fiori, 2012). However, in tropical and subtropical
221 areas breeding and recruitment occur year-round (Neves et al., 2010) suggesting that it
222 is likely that *H. australis* have shorter generation times. This species is capable of
223 creating an air bubble inside its pallial cavity, allowing juveniles and adults to
224 temporarily float in the water column, whereat they are carried along by the tide and
225 wind driven current (Echeverría, Neves, Pessoa, & Paiva, 2010), increasing its
226 dispersion potential. Moreover, *H. australis* is an obligate intermediate host of a large
227 diversity of parasites that infect several other species hosts. For instance, *H. australis*
228 has been reported to host 18 larval trematode species and 16 of these were from the

229 Bahía Blanca estuary (Argentina), being all but one parasitic castrators, with an
230 individual worm eventually filling up the digestive gland and gonad of the snail host
231 (Alda & Martorelli, 2014).

232

233 **Study area and (a)biotically heterogeneous habitats**

234 This study was conducted in the Villa del Mar saltmarsh-mudflat located in the middle
235 reaches of the Bahía Blanca estuary, Argentina (38° 51' S – 62° 07' W). The intertidal
236 mudflat extends for more than 1 km across the tidal gradient. Its topography is
237 characterized by a very gently sloping ramp on the seaward side (Pratolongo, Kirby,
238 Plater, & Brinson, 2009) and affected by strong semidiurnal tides and high seasonal
239 variation (Perillo, Piccolo, Parodi, & Freije, 2001). Close to the mean high tide level,
240 marshes covered by cordgrass (*Spartina alterniflora*) form a narrow, 150 m wide strip
241 of vegetation followed by a mudflat area with no vegetation (Pratolongo, Perillo, &
242 Piccolo, 2010). During low tide, these upper areas of the intertidal zone are subjected to
243 high desiccation and temperature and salinity fluctuations as the tide takes *ca.* 10 hours
244 to cover them (P. Pratolongo, *personal communication*; 26 February 2019) whereas low
245 areas located close the seaward edge remain covered by water during low tide.

246

247 The existence of a tidal cycle exposes *H. australis* to extreme, but predictable, changes
248 in abiotic conditions (at least twice every day) likely determining their vertical
249 distribution (zonation patterns). Snails in the upper intertidal zone must remain
250 quiescent for long periods of time during daylight hours or during low tide, being
251 exposed to aerial conditions for longer periods than individuals in the lower intertidal
252 zone, which remain cover by water during low tide. The intertidal zone can therefore be
253 characterized in three distinct habitats: flats, marshes, and pans. Flats and marshes are
254 located in the upper zone and they drain at low tide, though flats are free of vegetation
255 and marshes are covered by cordgrass. Pans are free of vegetation but remain covered
256 by water during low tide and are located close the seaward edge. Thermal, saline, and
257 dehydration stress are strong selective forces occurring mainly in the upper intertidal
258 area (flats and marshes; Bourdeau et al., 2015; Denny & Paine, 1998) whereas these
259 inducing agents are weaker in pans, which exhibit low environmental stress condition
260 (Fig. S1).

261

262 Biotic agents also vary along the vertical distribution of the intertidal zone (Fig. S1). On
263 the one hand, the grapsid crab *Neohelice granulata* is one of the most abundant
264 macroinvertebrates of intertidal areas of the SW Atlantic estuaries where it commonly
265 inhabits the upper vegetated area (marshes) but uses the entire intertidal zone (Alvarez
266 et al., 2013; Angeletti & Cervellini, 2015; Angeletti, Lescano, & Cervellini, 2014;
267 Spivak, Anger, Luppi, Bas, & Ismael, 1994). This burrowing crab could exert a strong
268 pressure on *Heleobia australis* since it can drastically reduce snail density in vegetated
269 areas due to an intense bioturbation activity (Alvarez et al., 2013; Angeletti &
270 Cervellini, 2015; Angeletti et al., 2014; Spivak et al., 1994) and/or through snail
271 predation (Barutot, D’Incao, & Fonseca, 2011; D’Incao, Silva, Ruffino, & Braga, 1990).
272 On the other hand, parasite pressure also differs along the intertidal zone. Trematode
273 infection not only inevitably leads to snail castration reducing its fitness to zero (Alda et
274 al., 2019) but also gives rise to smaller shell-sized morphs for infected snails (Alda et
275 al., 2010). The prevalence (percentage of individuals infected) of trematodes is higher in
276 the lower area of the intertidal zone, where parasite infection is predominately caused
277 by one extremely prevalent trematode, *Microphallus similimus*. Such pressure is
278 stronger in pans than in flats and marshes because prolonged submersion times allow
279 snails to increase time spent foraging and thus ingesting parasite eggs (Alda et al.,
280 2019).

281

282 **Field sampling and laboratory procedure**

283 We sampled individuals of the intertidal mud snail *H. australis* in summer (March 6),
284 autumn (July 15), winter (September 20), and spring (December 7) of 2012 in a one-
285 hectare plot. We randomly took nine samples from each habitat (flats, marshes, and
286 pans) with 10 cm diameter and 2 cm deep circular samplers (Area = 78.5 cm²). Snails
287 were sieved from the sediment through a 1 mm-mesh, then transported alive to the lab,
288 kept in aquaria, and fed *ad libitum* with flake fish food.

289

290 All 10,367 snails were used to estimate snail density and were also photographed using
291 a camera attached to a dissecting microscope to analyze shell and aperture
292 morphometric (see below). A random subset of snails (n = 6,250) was used to account
293 for infected and uninfected snails (hereafter infection status) in each habitat. The
294 remaining uncrushed snails were used to estimate snail shell and body weight (n =

295 3,057; see below) and parasite biomass (n = 1,060; N. Bonel and P. Alda, *unpublished*
296 *data*).

297

298 **Snail density and infection status**

299 To test how contrasting habitat conditions affected snail density, we counted all
300 sampled individuals per habitat and sampling date to analyze spatial and temporal
301 variation of snail density. Then, to determine infection status, snails from each habitat
302 were crushed using a mortar and a pestle, tissue was examined under a dissecting
303 microscope, and trematodes were identified under a compound microscope following
304 Alda & Martorelli (2014). Previous studies show that the most prevalent parasite in the
305 study area is *Microphallus simillimus* (Microphallidae), which makes up 87% of the
306 overall prevalence (Alda et al., 2019; Alda & Martorelli, 2014). This parasite has life
307 history traits such as abbreviated life cycle (the same snail host can serve as the first-
308 and second-intermediate host), meaning that once snails ingest parasite eggs, the
309 metacercariae encyst within the sporocyst in the infected snail, thereby maximizing
310 transmission success (Alda & Martorelli, 2014). One possible approach to the
311 contribution of parasites to the phenotypic variance in host populations is to compare
312 phenotypic responses of uninfected and infected hosts maintained under identical
313 conditions (Poulin & Thomas, 1999). Thus, to test for this, we only considered snails
314 infected by the most prevalent trematode *M. simillimus* and also from pans (the habitat
315 with the highest prevalence; Alda et al., 2019), which implies that both infected and
316 uninfected snails grew in the same environmental conditions. By doing so, we can
317 ensure that differences in phenotypic responses can only be attributed to *M. simillimus*
318 effect and not caused by other trematodes or habitat stress conditions.

319

320 As older hosts have greater cumulative risk of infection than do young hosts, each
321 sample should be standardized by age, collected at the same time from an area in which
322 hosts are likely to mingle and where they have experienced relatively uniform risks of
323 infection (Lafferty, Sammond, & Kuris, 1994). We therefore identified age cohorts by
324 means of the length-frequency distributions in each sampling date and habitat and
325 removed those individuals outside the lower 95% confidence limit of the cohort with
326 larger individuals (see Appendix S1 in Supporting Information, ‘*Snail length-frequency*
327 *distribution and cohort identification*’ for details; Fig. S2).

328

329 **Shell and aperture morphometric**

330 To analyze variation in shell morphometric across habitat conditions and infection
331 status, we measured four linear variables from each photograph using ImageJ software:
332 shell length (SL), shell width (SW), aperture length (AL), and aperture width (AW). We
333 followed this approach rather than the geometric morphometric analysis because it
334 allowed us for measuring a larger number of individuals (i.e.; 10,367 snails). Similar to
335 some other caenogastropod species (e.g. *Littoridina* sp.), the shell of *H. australis* is
336 generally small and conical and the axis of coiling lying at an angle of about 45° above
337 the plane of the elliptical aperture (Hershler & Thompson, 1992; Vermeij, 1973). We
338 thus used SL and SW to calculate the shell size by calculating the volume of a cone:

339

340
$$\text{Shell size: Volume of a cone} = 1/3 \pi (\text{SW}/2)^2 \text{SL}$$

341

342 To analyze the shell shape, we estimated the SL to SW ratio. Likewise, we analyzed
343 aperture shape by calculating the AL to AW ratio. We used AL and AW to calculate
344 aperture size by calculating the area of an ellipse:

345

346
$$\text{Aperture size: Area of ellipse} = (\text{AL}/2) (\text{AW}/2) \pi$$

347

348 **Shell thickness and body mass**

349 To test whether snails subjected to a stronger predation pressure exhibit thicker shells,
350 (considered to be a shell defense trait against predators as it is more resistant to
351 crushing; Bourdeau et al., 2015; Dillon et al., 2013; Johannesson, 2003) and whether
352 contrasting environmental stress conditions affected body mass, we estimated and
353 compared shell weight and ash-free dry-weight among habitats. We used a subsample of
354 individuals (n = 70), representative of the population shell length-frequency
355 distributions obtained previously for all individuals sampled on each date and habitat.
356 We removed sediment and epibiota with a scouring pad before weighing and measuring
357 the organic content of the snail as ash-free dry weight (AFDW, calculated as the
358 difference between the dry weight and the weight of the incombustible component of
359 the shell), which was considered as body mass. To achieve this, we dried snails
360 individually in porcelain crucibles for 48 h at 60 °C, weighed them with a digital scale

361 (precision 0.1 mg), ashed the snails for 5 h in a muffle furnace at 500 °C and then
362 reweighed them (Bonel & Lorda, 2015). The ash shell weight (the incombustible
363 component of the shell) was considered to be an equivalent of shell thickness and used
364 to estimate and compare its variation among individuals along the vertical gradient of
365 the intertidal area.

366

367 **Variation in penis size**

368 To test for differences in genital morphology, we analyzed variation in the penis size
369 and morphology of individuals from contrasting habitats conditions living in sympatry
370 but raised in common garden laboratory conditions. We sampled snails of *H. australis*
371 from the three distinct habitats of the intertidal area of the Villa del Mar saltmarsh-
372 mudflat in November (Spring) of 2017. Snails were transported to the lab and they were
373 split into adults and juveniles based on their shell-length distribution (as mentioned
374 above). Adults were preserved in alcohol 96% and used to conduct molecular analysis.
375 Juveniles were kept them in three different aquaria (one for each habitat) in common
376 garden laboratory conditions (12:12 photoperiod, 25 °C, water salinity 30 PSU, and *ad*
377 *libitum* food in the form of boiled ground lettuce) for 21 months, which ensured that all
378 individuals considered in the analysis were adults. Prior to dissection, we collected, on
379 average, 30 snails from each aquaria and preserved them in the Railliet–Henry’s
380 solution (930 ml distilled water, 6 g NaCl [0.85%], 50 ml formaldehyde [37%], 20 ml
381 glacial acetic acid). Then, they were all dissected under the dissecting microscope and
382 the reproductive system of *ca.* 15 males per aquarium was drawn using a camera lucida
383 attachment (Pointier, Noya, Amarista, & Théron, 2004). These were scanned and the
384 penis surface area was calculated using the ImageJ software. Shells were photographed
385 and analyzed as mentioned above to estimate shell size and shape.

386

387 **COI analysis**

388 To confirm that individuals living in these contrasting habitats belong to the same
389 species and not to a species complex, we conducted molecular analyses by amplifying
390 the cytochrome oxidase subunit 1 (COI) gene in 10 individuals sampled from each of
391 the three sites. Then, we built a gene tree and a haplotype network to verify if
392 individuals gather in clusters depending on the habitat they come from. We also applied
393 a species-delimitation method: Automatic Barcode Gap Detection (ABGD; Puillandre,

394 Lambert, Brouillet, & Achaz, 2012). ABGD detects barcode gaps, which can be
395 observed whenever the divergence among organisms belonging to the same species is
396 smaller than divergence among organisms from different species (see Supporting
397 Information for specific details on these procedures).

398

399 **Statistical analyses**

400 To test for spatial and temporal variation of snail density, we performed a two-way
401 ANOVA including habitats (flats, marshes, and pans) and seasons (summer, autumn,
402 winter, spring) as fixed effect, and their interaction. We transformed density (natural-
403 log) data to meet assumptions of normality and homoscedasticity.

404

405 To test for phenotypic variation in shell characters, we defined three categories (cat.) of
406 snails that comprised: (i) uninfected adults, (ii) infected adults, and (iii) uninfected
407 juveniles. These three categories (cat. *i-iii*) were considered to test whether contrasting
408 (a)biotic stress conditions (environmental stress and crab predation) induced variation in
409 shell traits (morphometric and thickness) and body mass. To evaluate the effect exerted
410 by the trematode *Microphallus similimus* on shell morphometry of *Heleobia australis*
411 from pans (the habitat with the highest prevalence), we created a fourth category that
412 included uninfected and infected adult snails (cat. *iv*). We performed four Principal
413 Components Analyses (PCAs) to visualize the main components of the morphological
414 shell variation for each snail category. Then, we tested for differences in shell
415 morphology (size and shape of the shell and aperture) by fitting independent linear
416 mixed models (LMMs), with a Gaussian error distribution, for the above-mentioned
417 four categories (cat. *i-iv*).

418

419 To test the effect of habitat conditions on shell traits (cat. *i-iii*), we considered habitat as
420 a fixed effect. To test for morphometric differences between infected and uninfected
421 (cat. *iv*), we included infection status (infected and uninfected) as a fixed effect as we
422 only considered individuals from one habitat (i.e. same habitat conditions for infected
423 and uninfected snails). For all the four categories (cat. *i-iv*), we included sex (male and
424 female) as a fixed factor (and its interaction with infection status or habitat). We added
425 sampling season as a random factor in all models as it can explain a significant portion
426 of the variance in measured variables. Likewise, we also considered shell size as a

427 covariate when testing for shell and aperture shape, but the models were all ran with and
428 without this effect to check the extent to which the effects were mediated by shell size.
429 We also incorporated shell size as a covariate when testing for shell thickness and body
430 weight (by means of LMMs) and estimates were statistically corrected for variation that
431 could be explained by the covariate. Statistical significance of the fixed effects was
432 obtained from model comparisons using likelihood-ratio tests. Random effects were
433 separately tested using chi-square likelihood-ratio tests with the corrections indicated by
434 Zuur (2009). To test for differences in penis size (area of the penis), we performed a
435 linear model with habitat as a factor and we added shell size (volume) as a covariate,
436 which was used to statistically correct estimates for variation explained by the covariate.
437 All individuals analyzed were free of parasites, meaning that the differences found are
438 not due to parasite effect.

439

440 Post-hoc tests were performed when the effects were significant ($P < 0.05$), using Holm-
441 Bonferroni correction for multiple testing to compare their effects. Values are given as
442 means \pm 1 SE unless otherwise stated. All analyses and figures were performed with R
443 v.3.3.3 packages *lme4* (Bates, Mächler, Bolker, & Walker, 2014), *nlme* (Pinheiro, Bates,
444 DebRoy, Sarkar, & R Core Team, 2017), *car* (Fox et al., 2011), *effects* (Fox, 2003),
445 *MASS* (Ripley et al., 2013), *plyr* (Wickham, 2011), *dplyr* (Wickham, François, Henry,
446 & Müller, 2019), *devtools* (Wickham, Hester, & Chang, 2019), *ggbiplot* (Vu, 2011),
447 *psych* (Revelle, 2018), *ggplot2* (Wickham & Chang, 2008), and *outliers* (Komsta,
448 2011).

449

450 **Results**

451 **Overview**

452 Overall, the mud snail *Heleobia australis* showed a strong variation in snail density and
453 a remarkable variation in several traits linked to contrasting (a)biotic selective pressures
454 (environmental stress, predation, or parasite infection) occurring along the vertical
455 zonation of the intertidal area. Tables 1 to 3 summarize descriptive statistics on density
456 and shell traits measured for the first and fourth snail categories (cat. *i* and *iv*). In the
457 appendix, we reported results of the snail density split by age classes (juvenile and
458 adult; Table S1), principal components analyses (Table S2), descriptive statistics and

459 the linear mixed models for the four snail categories on shell traits (Tables S3–S8) and
460 for shell weight and body mass (Table S9).

461

462 **Strong habitat and seasonal effect on snail density**

463 We found spatial differences in total density (including juveniles and adult snails; Table
464 1; Fig. 1). We observed no significant interaction between habitats and seasons ($F_{(6, 106)}$
465 $= 1.56$, $P = 0.167$). Marshes showed the lowest density whereas flats and pans showed a
466 marginal non-significant difference (Table 1). We found significant differences between
467 seasons (Table 1). Density was higher in spring than in autumn, winter, and summer
468 (Table 1). Snail density in autumn did not differ from that of summer or winter, these
469 latter two being not different either (Table 1), but it largely increased from autumn to
470 spring (particularly in pans as compared to flats and marshes). This increase across
471 seasons was mainly driven by juvenile recruitment (Table S1; Fig. S3). By contrast, the
472 effect of habitat conditions and seasons on adult density showed a significant interaction
473 because density was higher in pans from autumn to winter whereas it decreased in
474 marshes for the same period, and in flats it decreased from summer to autumn and the
475 than remained fairly constant (Table S1; Fig. S3).

476

477 **Main components of the morphological shell variation within each snail group**

478 For each of the snail categories (cat. *i–iv*), principal component analyses decomposed
479 the shell morphometric measures (shell and apertures size and shape) into four principal
480 components of which the first two explained, overall, more than 73% of the variance in
481 the original data (Table S2; Fig. S4). The component loadings showed that the shell and
482 aperture size are highly correlated one another and with PC1, whereas PC2 highly
483 contrast shell shape with aperture shape. In other words, the first component was
484 strongly correlated with size indices, and the second component with shape (further
485 details in Table S2). The distribution of individual data points from flats and marshes
486 highly overlapped, and this was consistent in all the three snail categories (cat. *i–iii*).
487 This constituted one group that clearly differed from the other group of data points that
488 was composed by individual observations from pans (Fig. S2). Likewise, the scatter of
489 the data showed two groups when decomposing the variation of morphometric measures
490 between infected and uninfected adult snails (cat. *iv*).

491

492 **Strong habitat effect on shell morphometric**

493 The morphometric responses to habitat conditions were similar in all the three snail
494 categories analyzed (cat. *i–iii*). For simplicity, we present results and figures only for
495 uninfected adult snails (cat. *i*) in the main text and those for infected adults (cat. *ii*) and
496 juvenile snails (cat. *iii*) are shown in Supporting Information (Tables S4-S7; Figs. S5-
497 S6).

498
499 ***Shell size***

500 We found a significant interaction between fixed effects (habitat and sex; $X^2_2 = 8.42$, P
501 $= 0.015$). This was mainly driven by differences in this trait between sexes in pans,
502 whereas such difference between males and females was not detected in snails from
503 flats and marshes. In other words, males exhibited a larger shell size (mm^3) compared to
504 females in pans, meaning that the sex effect was much stronger in creating size
505 differences between sexes compared to flats and marshes where the sex effect was
506 weaker resulting in no difference in size between sexes. By contrast, the effect of habitat
507 conditions was stronger in both sexes. Both male and female snails from flats and
508 marshes showed a smaller shell size (a decrease of 12 and 22 % in shell size;
509 respectively) relative to individuals from pans (Tables 2 and S3; Fig. 2).

510

511 ***Shell shape***

512 The interaction between fixed effects was significant ($X^2_2 = 8.20$, $P = 0.017$), but when
513 we statistically corrected for shell size, it became marginally non-significant ($X^2_2 =$
514 5.69 , $P = 0.058$). Snails from pans (sex pooled) exhibited the most elongated shells
515 relative to individuals from marshes and flats. At the sex level (habitats pooled), males
516 had a more elongated shell shape than females (Table 2 and S3; Fig. 2).

517

518 ***Aperture size***

519 We found a strong habitat and sex effect on aperture size; the interaction between these
520 variables was not significant ($X^2_2 = 3.05$, $P = 0.217$). Across habitats (sex pooled),
521 snails from pans showed the largest aperture size (mm^2) whereas in flats and marshes it
522 was 9 to 12% smaller, respectively. Moreover, we observed a significant difference in
523 aperture size between sexes (habitats pooled), males had bigger apertures than females
524 and this difference was consistent across habitats (Table 2 and S3; Fig. 2).

525

526 ***Aperture shape***

527 We found no significant difference in aperture shape between sexes ($X^2_1 = 0.65$, $P =$
528 0.724), and this pattern was similar even after correcting for shell size ($X^2_2 = 2.20$, $P =$
529 0.333). However, individuals from pans (sex pooled) showed the most rounded shape
530 relative to those from flats and marshes, which showed a more elongated aperture shape
531 (Table 2 and S3; Fig. 2).

532

533 **Pronounced trematode effect on shell morphometric**

534 Infection caused by the trematode *M. simillimus* strongly shifted the mean value of most
535 phenotypic traits of *H. australis* to lower values when comparing infected and
536 uninfected snails from pans.

537

538 ***Shell size***

539 Shell size of infected snails (sex pooled) decreased 10% relative to uninfected ones,
540 whereas females (infection status pooled) were 5% smaller than males (Table 3; Fig. 3).
541 We found no significant interaction between the fixed effects (status and sex; $X^2_1 =$
542 0.70 , $P = 0.402$).

543

544 ***Shell shape***

545 We found a significant interaction between status and sex ($X^2_1 = 15.86$, $P < 0.001$). This
546 was because, at the status level, uninfected males were more elongated (or exhibited
547 higher SL to SW ratio) than uninfected females, but infected males and females showed
548 no difference in shell shape. We found the same pattern at the sex level; that is, infected
549 male and female snails have a more elongated shell relative to uninfected male and
550 female individuals (Table 3; Fig. 3).

551

552 ***Aperture size***

553 We observed a strong effect of infection status and sex in aperture size. Infected snails
554 (sex pooled) and females (status pooled) had a smaller aperture than uninfected and
555 male snails (15 and 3% respectively; Table 3; Fig. 3). We found no significant
556 interaction between fixed effects ($X^2_1 = 2e-04$, $P = 0.990$).

557

558 ***Aperture shape***

559 Uninfected snails exhibited a more elongate aperture shape (higher AL to AW ratio)
560 than infected individuals whereas we found no sex effect on this shell trait (Table 3;
561 Fig. 3). We found no significant interaction between fixed effects and no sex effect on
562 aperture shape, even after correcting by shell size ($X^2_1 = 0.33$, $P = 0.564$).

563

564 **Thicker shells and lower body mass in habitats with high (a)biotic stress**

565 We found that snails differed in shell weight (~ thickness) across habitats ($X^2_2 = 48.93$,
566 $P < 0.001$). Snails from pans showed the thinnest shells (11.53 ± 0.23 mg) compared to
567 snails from marshes (12.54 ± 0.15 mg; $P_{\text{Pans} < \text{Marshes}} < 0.001$) and flats (12.00 ± 0.15 mg;
568 $P_{\text{Pans} < \text{Flats}} = 0.001$). By contrast, in marshes individuals showed the heaviest/thickest
569 shells ($P_{\text{Marshes} > \text{Flats}} < 0.001$; Fig. 4). Body mass also differed across habitats ($X^2_2 = 19.67$,
570 $P < 0.001$). Snails from pans exhibited the heaviest body mass (0.50 ± 0.05 mg) relative
571 to individuals from marshes (0.32 ± 0.03 mg) and flats (0.33 ± 0.04 mg) ($P_{\text{Pans} > \text{Marshes}}$
572 < 0.001 ; $P_{\text{Pans} > \text{Flats}} < 0.001$), which showed no differences in body weight ($P_{\text{Marshes} < \text{Flats}} =$
573 0.390 ; Fig. 4). Further details on observations, corrected and uncorrected mean values,
574 and statistics are indicated in the appendix (Table S9; Fig. S7).

575

576 **Variation in penis size**

577 The penis shape confirmed that individuals from each habitat of the Bahía Blanca
578 estuary belong to the same species (Fig. S8; Gaillard & Castellanos, 1976). We found,
579 however, a clear difference in penis size ($F_{(2, 39)} = 8.42$, $P < 0.001$). Individuals from
580 marshes showed the largest size relative to flats ($P_{\text{flats} < \text{marshes}} < 0.001$) and pans
581 ($P_{\text{pans} < \text{marshes}} = 0.034$); snails collected from these two habitats showed a marginal non-
582 significant difference ($P_{\text{flats} < \text{pans}} = 0.090$; Fig. 5). We found no significant effect of shell
583 size, meaning that differences in penis size were not due differences in individual's shell
584 size ($F_{(1, 39)} = 0.27$, $P = 0.608$). Further, we found no differences in shell shape, even after
585 controlling by shell size ($F_{(2, 43)} = 0.851$, $P = 0.435$).

586

587 **Molecular analysis of the cytochrome oxidase subunit 1 (COI) gene**

588 The species-delimitation analysis implemented in ABGD found only one partition (prior
589 maximal distance, $P = 0.001$) that confirmed that the individuals here studied constitute a
590 single *Heleobia* species, which is in agreement with the similar penis shape found

591 among habitats. In fact, we did not observe any type of structure among individuals.
592 The gene tree and haplotype network showed that individuals did not gather in clusters
593 depending on the habitat they come from (Figs. S9-S10).

594

595 **Discussion**

596 This study provides clear evidence of ecological and phenotypic trait variation in
597 response to contrasting biotic and abiotic conditions at a local scale. Snail density and
598 most of mean values of phenotypic traits measured shifted to lower values in habitats
599 with high physical stress conditions (flats and marshes), crab-predation (marshes), and
600 parasite (pans) pressure (summarized in Fig. 6). The pronounced phenotypic
601 differentiation in *Heleobia australis* is all the more remarkable because it occurred at a
602 very small geographical scale, which is rarely documented for a wide-dispersing
603 species. These findings support the standpoint that this species might have experienced
604 a fine-grained environment where the combined effect of directional selection and
605 plasticity led to the evolution of locally adapted phenotypes in response to contrasting
606 selection pressures.

607

608 **Reduced snail density under high environmental and predatory stress**

609 Density of *Heleobia australis* was the lowest in marshes (vegetated area that drains at
610 low tide) relative to flats (unvegetated area that drains at low tide) and pans (covered by
611 water at low tide) whereas flats and pans showed no difference. However, when
612 analyzed by age classes, adult density was higher in pans during winter but remarkably
613 low in marshes, even compared to flats (both habitats from the upper area). This is
614 consistent with the idea that prolonged submersion times could increase the survival
615 rate of snails during harsh climatic conditions (Fig. S3). As for juvenile snails, their
616 density increased during winter season in all the three habitats (Fig. S3). This was
617 somehow unexpected as we hypothesized that intense environmental stress in the upper
618 area of the intertidal zone, combined with harsh climatic conditions, could strongly
619 decrease their survival rate affecting their abundance. It is puzzling, however, that
620 density in marshes was low relative to flats because *H. australis* is positively associated
621 to marsh plants (the smooth cordgrass *Spartina alterniflora*, which dominates the lower
622 marsh), as they buffer physical stress factors (thermal and dehydration stress) relative to

623 uncovered areas, promoting snail aggregation (Canepuccia et al., 2007 and references
624 therein).

625

626 One possible explanation for the low density in marshes, relative to flats, is if this
627 resulted from a negative interaction between *H. australis* and the grapsid crab *Neohelice*
628 *granulata*, whose aggregation is also facilitated by marsh plants (Alvarez et al., 2013;
629 Angeletti & Cervellini, 2015; Angeletti et al., 2014; Spivak et al., 1994) where it attains
630 density peaks in late spring and early summer but it is absent throughout the year in
631 unvegetated areas (flats or pans; Angeletti & Cervellini, 2015). It is therefore likely that
632 the low density of *H. australis* in vegetated crab-rich habitats resulted from an increased
633 bioturbation activity caused by *N. granulata*, particularly during that period of time
634 (Alvarez et al., 2013). Another mutually non-exclusive hypothesis is that this difference
635 between marshes and flats could result from a higher predation pressure caused by this
636 crab (Barutot et al., 2011; D’Incao et al., 1990), in which case it might have induced
637 predatory shell defenses as shell thickening or narrower apertures (Boulding & Hay,
638 1993; Johannesson & Johannesson, 1996; Rolán-Alvarez, 2007; review by Bourdeau et
639 al., 2015). Accordingly, snail from marshes exhibited a strong variation in shell
640 characters that support this hypothesis, which we discuss later in this section.

641

642 **Snails exhibited larger, narrower, and thinner shells and heavier body mass in**
643 **habitats with low stress conditions**

644 Snails subjected to different levels of stress showed a strong phenotypic variation. In all
645 categories analyzed (cat. *i-iii*), individuals showed a clear shift in their phenotype likely
646 in response to different selective pressures. Overall, the distribution of trait values
647 tended to represent two distinct groups, one for flats and marshes and another for pans.
648 Such responses seem to be primarily linked to the level of exposure to physical stress
649 (temperature and salinity fluctuation and desiccation) of individuals during low tide.
650 Prolonged submersion times enhance foraging times and the absorption of calcium
651 carbonate from water increasing shell growth rate, which, in turn, gives rise to larger but
652 more elongated shells (review by Chapman, 1995). As there is a maximal rate at which
653 calcium carbonate can be absorbed from the water, rapidly growing individuals produce
654 thinner shells for the same amount of body weight than slower growing individuals.
655 This results in a larger internal volume allowing for accommodating a higher body

656 mass, which is related to higher growth rate favored by longer foraging times
657 (Chapman, 1995; Kemp & Bertness, 1984; Palmer, 1981; Trussell, 2000a).
658 Accordingly, snails that remain covered by water during low tide (pans) showed larger,
659 narrower, and thinner shells and heavier body mass relative to individuals from habitats
660 with high environmental stress conditions (flats and marshes). As higher body weight is
661 directly linked to higher fecundity (i.e. egg production; Fredensborg et al., 2006), it is
662 therefore likely that the heaviest body weight of individuals from pans is related to a
663 higher recruitment, which would explain the increased juvenile abundance compared to
664 flats and marshes (Fig. S3).

665
666 Juvenile and adult snails from flats and marshes showed a smaller aperture size (smaller
667 aperture surface area) with respect to individuals from pans. It could be that higher
668 environmental stress conditions favored the expression of a smaller aperture size as a
669 potentially beneficial trait increasing resistance to high desiccation in these habitats that
670 drain at low tide. This is consistent with other studies showing that the size of the shell
671 aperture in gastropods is smaller in response to hot and dry conditions (Chapman, 1995;
672 Machin, 1967; Melatunan et al., 2013; Schweizer et al., 2019; Vermeij, 1973).
673 Moreover, individuals from flats and marshes showed thicker shells and lower body
674 mass. Shell thickening could be due to a higher deposition of calcium carbonate at a
675 slow grow rate (likely in response to unfavorable conditions in more exposed habitats
676 from the upper intertidal) that traded off against investment in body mass.

677

678 **Crab predation pressure would favor the expression of shell defenses in *Heleobia***
679 ***australis***

680 Snails from marshes showed a more rounded and thicker shell, reduced body mass, and
681 a narrow aperture shape for three snail categories analyzed (cat. *i-iii*). Such
682 characteristics are consistent with shell defenses probably induced by the high presence
683 of the predatory crab *Neohelice granulata* in that habitat in particular (e.g. Appleton &
684 Palmer, 1988; Bourdeau et al., 2015; Palmer, 1990). Considering that crushing
685 predators exert a strong selective pressure driving the evolution of behavioral, chemical,
686 and morphological defense traits (Appleton & Palmer, 1988; Bourdeau et al., 2015;
687 Johannesson, 2003; Palmer, 1990; Trussell, 2000b), our results, combined with its
688 remarkably low density, support the hypothesis that predation pressure is stronger in

689 marshes than in flats and pans. Yet, specific information on whether crabs actively
690 predate on *H. australis* is limited or inexistent. Earlier studies reported that *N.*
691 *granulata* consumes mollusks but in a low frequency (Barutot et al., 2011; D’Incao et
692 al., 1990). These studies, however, have used visual examination of gut/stomach
693 contents. This procedure may have under-estimated true predation levels as crabs only
694 eat the soft tissue after crushing/peeling the shell whereby the remains found in guts or
695 stomach become unidentifiable due to maceration and digestion. Thus, further studies
696 using molecular-based tools are needed for detecting the presence of *Heleobia*’s tissue
697 in crab’s stomachs as an alternative or complementary approach to visual identification
698 (e.g. Albaina et al., 2010; Collier, Fitzgerald, Hice, Frisk, & McElroy, 2014). By doing
699 so, we would be able to establish the trophic link between these two species that is
700 highly likely to exist. Evidence found in this study clearly reveals that *H. australis*
701 phenotypically responded to the presence of predatory crabs, likely even early in life, as
702 juveniles from marshes also showed a narrow aperture shape and a more rounded shell.
703 Our results therefore warrant further investigation to understand the adaptive value of
704 the plastic shell responses of gastropods induced by predators.

705

706 **Strong parasite effect on shell characters: a trade-off between growth and early**
707 **reproduction**

708 Infected and uninfected snails from pans showed a clear phenotypic differentiation in
709 shell traits. Trematode infection inevitably leads to snail host castration drastically
710 reducing host fitness, which, in evolutionary terms, is equivalent to death of the host
711 (Fredensborg et al., 2006; Lafferty, 1993). In this sense, being castrated by a parasite is
712 similar as being eaten by a predator (Kuris, 1974). This means that parasitic castrators
713 can exert a strong selective force that could favor the expression of behavioral,
714 physiological, and morphological adaptations to minimize the negative impact of
715 parasitism on host fitness (Lafferty, 1993). Early maturation, which results from fast
716 growing individuals, is an effective strategy that increases current reproductive effort
717 over future reproduction (Cole, 1954; Lewontin, 1965; Roff, 1992) which equals to 0 in
718 castrated hosts. In *Heleobia*, castration caused by *Microphallus* can have a profound
719 impact on snail fitness as this parasite shows an extensive host exploitation occupying
720 the entire gonad and most of digestive gland (see Fig. S2 in Alda et al., 2019). Our
721 results showed that infected snails had a more elongated shell shape but a smaller size

722 compared to uninfected snails. Despite infected snails were smaller, they exhibited the
723 same number of whorls (six) as uninfected snails, which indicates that individuals
724 analyzed were adults (Gaillard & Castellanos 1976). It could therefore be that the
725 elongated shape resulted from an initial higher growth rate (review by Chapman, 1995)
726 likely followed by an energy reallocation from growth to early reproduction (Agnew,
727 Bedhomme, Haussy, & Michalakis, 1999; Lafferty, 1993), which could explain the
728 smaller shell size compared to uninfected snails (Alda et al., 2010, 2019; this study).
729 Infected snails also showed a smaller aperture size but a more rounded shape relative to
730 uninfected individuals. This change in aperture size and shape could be a side effect of
731 selection for high rate of shell growth (Boulding & Hay, 1993) induced by parasites.
732 Together, these findings support the idea that infection by trematodes exerted a strong
733 selective pressure on the snail host causing important shifts in the expression of shell
734 traits towards lower mean values, creating a pronounced phenotypic difference between
735 infected and uninfected snails (Poulin & Thomas, 1999).

736

737 **Difference in penis size in a single *Heleobia* species**

738 Interestingly, individuals from each habitat showed a clear variation in penis size, which
739 is all the more remarkable as they were raised in common garden laboratory conditions
740 (no habitat effect) and, when dissected, we observed that none of them was infected by
741 parasites (no parasite effect). It could be that this variation resulted from different
742 growth rates as well as the differences in shell shape observed for wild snails. However,
743 we found no differences in shell shape, even after controlling by shell size, among male
744 individuals reared in laboratory conditions. This result suggests that the differences in
745 penis size might not be related to differences in growth rate. Future studies are therefore
746 needed to exclude other processes (e.g. genetic differences) that may have contributed
747 to this difference in penis size and to understand whether this genital divergence could
748 be an indicative of pre-zygotic reproductive isolation among subpopulations of
749 *Heleobia australis* living in sympatry (Hollander, Smadja, Butlin, & Reid, 2013;
750 Kameda, Kawakita, & Kato, 2009). Despite the difference in size, its shape indicates
751 that individuals from each subpopulation belong to the same species, more specifically
752 to *Heleobia australis australis* (see Gaillard & Castellanos, 1976 for further detail on
753 taxonomy). Accordingly, the protein-coding COI gene confirmed that specimens
754 considered in this study constitute a single species.

755

756 **Does the pronounced phenotypic differentiation in *Heleobia australis* reflect an**
757 **adaptive evolutionary response to local selection pressures?**

758 The intertidal mud snail *H. australis* exhibited a strong phenotypic variation in shell
759 morphology and other characteristics associated to contrasting biotic (parasitism and
760 predation) and abiotic (high temperature, salinity, and desiccation stress) factors among
761 habitats that are separated just a few meters away from each other within the intertidal
762 area. Such phenotypic differentiation at a very small geographical scale has been rarely
763 documented for a natural population of a wide-dispersing species. It does seem
764 reasonable to interpret the observed variation as an adaptive phenotypic response to
765 contrasting selection pressures occurring along the vertical gradient of the intertidal area
766 (Levins, 1968; Pigliucci, 2001). Environmentally-induced phenotypic change has been
767 reported for a large number of snail species (review by Bourdeau et al., 2015). The
768 appearance of an environmentally induced novel phenotype relies on plasticity unique
769 ability to generate an immediate phenotypic response to the surrounding habitat
770 conditions (West-Eberhard, 2003). If genetic variation in plasticity exists (or arises),
771 then an environmentally induced novel phenotype can be refined by selection on the
772 expression of such phenotype through processes such as genetic accommodation or
773 assimilation (see Pigliucci et al., 2006; Waddington, 1953; West-Eberhard, 1989, 2003,
774 2005 for further details). Both processes are potentially relevant in fostering divergent
775 phenotypes within populations and subsequently driving diversification (e.g. Pfennig et
776 al., 2010 and references therein; Suzuki & Nijhout, 2006; West-Eberhard, 2003, 2005).
777 Multilocus or genomic analyses are therefore needed to unveil whether the phenotypic
778 differentiation of external and internal characters might have become genetically
779 assimilated, likely due to by local disruption to dispersal caused by environmental stress
780 conditions thereby increasing localized recruitment (e.g. Parsons, 1997; Struhsaker,
781 1968). If so, the individuals studied herein would be part of a species complex that have
782 recently diverged, which would explain why individuals showed, for instance,
783 differences in the genital morphology but did not cluster when analyzing the conserved
784 COI gene (Janzen et al., 2017; Matos & Maraví, Wahlberg, Antonelli, & Penz, 2019).

785

786 **Conclusion**

787 Our results show a remarkable shift in the expression of phenotypic traits, even early in
788 life, in response to strong selective pressures exerted by biotic (parasites, predation) and
789 abiotic (temperature, salinity, desiccation) stressors, which also affected ecological
790 processes such as snail density (summarized in Fig. 6) in a natural population at a very
791 small geographical scale. While this study was not designed to test for an adaptive basis
792 to phenotypic variation, these findings provide a reasonable ground for advocating that
793 *H. australis* benefited from the combined effect of directional selection (likely through
794 genetic assimilation) and plasticity to evolve locally adapted phenotypes to contrasting
795 habitat conditions. The demonstration of the adaptive nature of these differences in life
796 history features will be the object of a future study.

797

798 **Acknowledgements**

799 We would like to express our gratitude to Kevin Lafferty, Patrice David, Antonio A.
800 Vazquez Perera, Philippe Jarne, and Romain Villoutreix for their constructive
801 suggestions on earlier versions of the manuscript. We also thank Leandro A. Hünicken
802 and Micaela Folino for field and laboratory assistance and to Néstor J. Cazzaniga for
803 providing insightful comments on the systematics of *Heleobia* group. This work was
804 partially funded by a research grant from the Universidad Nacional del Sur (UNS
805 24/B153 and UNS 24/B199). NB was partially supported by the “Programa de
806 Financiamiento Parcial de Estadías en el Exterior para Investigadores Asistentes”,
807 National Scientific and Technical Research Council CONICET (Res. N° 558 1236/08;
808 4118/16”).

809

810 **Authors’ contributions**

811 NB and PA conceived the ideas and designed methodology; NB, JPP, and PA collected
812 the data; NB analyzed the data and led the writing of the manuscript. All authors
813 contributed critically to the drafts and gave final approval for publication.

814

815 **References**

- 816 Agnew, P., Bedhomme, S., Haussy, C., & Michalakis, Y. (1999). Age and size at maturity of
817 the mosquito *Culex pipiens* infected by the microsporidian parasite *Vavraia culicis*. *Proceedings*
818 *of the Royal Society of London. Series B: Biological Sciences*, 266(1422), 947–952. doi:
819 10.1098/rspb.1999.0728
- 820 Albaina, A., Fox, C. J., Taylor, N., Hunter, E., Maillard, M., & Taylor, M. I. (2010). A TaqMan

- 821 real-time PCR based assay targeting plaice (*Pleuronectes platessa* L.) DNA to detect predation
822 by the brown shrimp (*Crangon crangon* L.) and the shore crab (*Carcinus maenas* L.)—Assay
823 development and validation. *Journal of Experimental Marine Biology and Ecology*, 391(1–2),
824 178–189. doi: 10.1016/j.jembe.2010.06.029
- 825 Alda, P., Bonel, N., Cazzaniga, N. J., & Martorelli, S. R. (2010). Effects of parasitism and
826 environment on shell size of the South American intertidal mud snail *Heleobia australis*
827 (Gastropoda). *Estuarine, Coastal and Shelf Science*, 87(2), 305–310. doi:
828 10.1016/j.ecss.2010.01.012
- 829 Alda, P., Bonel, N., Cazzaniga, N. J., Martorelli, S. R., & Lafferty, K. D. (2019). A strong
830 colonizer rules the trematode guild in an intertidal snail host. *Ecology*, e02696. doi:
831 10.1002/ecy.2696
- 832 Alda, P., & Martorelli, S. (2014). Larval trematodes infecting the South-American intertidal
833 mud snail *Heleobia australis* (Rissooidea: Cochliopidae). *Acta Parasitologica*, 59(1). doi:
834 10.2478/s11686-014-0209-3
- 835 Alvarez, M. F., Esquius, K. S., Addino, M., Alberti, J., Iribarne, O., & Botto, F. (2013).
836 Cascading top-down effects on estuarine intertidal meiofaunal and algal assemblages. *Journal*
837 *of Experimental Marine Biology and Ecology*, 440, 216–224. doi: 10.1016/j.jembe.2012.12.015
- 838 Angeletti, S., & Cervellini, P. M. (2015). Population structure of the burrowing crab *Neohelice*
839 *granulata* (Brachyura, Varunidae) in a southwestern Atlantic salt marsh. *Latin American Journal*
840 *of Aquatic Research*, 43(3), 539–547.
- 841 Angeletti, S., Lescano, L., & Cervellini, P. (2014). Caracterización biosedimentológica y
842 mineralógica de dos sectores intermareales del estuario de Bahía Blanca. *GeoActa*, 39(2), 54–
843 67.
- 844 Appleton, R. D., & Palmer, A. R. (1988). Water-borne stimuli released by predatory crabs and
845 damaged prey induce more predator-resistant shells in a marine gastropod. *Proceedings of the*
846 *National Academy of Sciences*, 85(12), 4387–4391. doi: 10.1073/pnas.85.12.4387
- 847 Barutot, R. A., D’Incao, F., & Fonseca, D. B. (2011). Natural diet of *Neohelice granulata* (Dana,
848 1851) (Crustacea, Varunidae) in two salt marshes of the estuarine region of the Lagoa dos Patos
849 lagoon. *Brazilian Archives of Biology and Technology*, 54(1), 91–98. doi: 10.1590/S1516-
850 89132011000100012
- 851 Bates, D., Mächler, M., Bolker, B., & Walker, S. (2014). Fitting Linear Mixed-Effects Models
852 using lme4. *ArXiv:1406.5823 [Stat]*. Retrieved from <http://arxiv.org/abs/1406.5823>
- 853 Berrigan, D., & Scheiner, S. (2004). Modeling the evolution of phenotypic plasticity. In
854 *Phenotypic plasticity: Functional and conceptual approaches* (DeWitt TJ, Scheiner, SM (eds),
855 pp. 82–97). New York: Oxford, Univ. Press.
- 856 Bonel, N., & Lorda, J. (2015). Growth and Bbody weight variability of the invasive mussel
857 *Limnoperna fortunei* (Mytilidae) across habitat and season. *Malacologia*, 58(1–2), 129–145.
858 doi: 10.4002/040.058.0202
- 859 Boulding, E. G., & Hay, T. K. (1993). Quantitative genetics of shell form of an intertidal snail:
860 constraints on short-term response to selection. *Evolution*, 47(2), 576. doi: 10.2307/2410072
- 861 Bourdeau, P E, Butlin, R. K., Brönmark, C., Edgell, T. C., Hoverman, J. T., & Hollander, J.
862 (2015). What can aquatic gastropods tell us about phenotypic plasticity? A review and meta-

- 863 analysis. *Heredity*, 115(4), 312–321. doi: 10.1038/hdy.2015.58
- 864 Bourdeau, Paul E. (2011). Constitutive and inducible defensive traits in co-occurring marine
865 snails distributed across a vertical rocky intertidal gradient. *Functional Ecology*, 25(1), 177–
866 185. doi: 10.1111/j.1365-2435.2010.01762.x
- 867 Canepuccia, A. D., Escapa, M., Daleo, P., Alberti, J., Botto, F., & Iribarne, O. O. (2007).
868 Positive interactions of the smooth cordgrass *Spartina alterniflora* on the mud snail *Heleobia*
869 *australis*, in South Western Atlantic salt marshes. *Journal of Experimental Marine Biology and*
870 *Ecology*, 353(2), 180–190. doi: 10.1016/j.jembe.2007.09.009
- 871 Carcedo, M. C., & Fiori, S. M. (2012). Long-term study of the life cycle and growth of *Heleobia*
872 *australis* (Caenogastropoda, Cochliopidae) in the Bahía Blanca estuary, Argentina. *Ciencias*
873 *Marinas*, 38(4), 589–597. doi: 10.7773/cm.v38i4.2079
- 874 Chapman, M. G. (1995). Spatial patterns of shell shape of three species of co-existing littorinid
875 snails in New South Wales, Australia. *Journal of Molluscan Studies*, 61(2), 141–162. doi:
876 10.1093/mollus/61.2.141
- 877 Chevin, L.-M., & Lande, R. (2009). When do adaptive plasticity and genetic evolution prevent
878 extinction of a density-regulated population? *Evolution*, 64(4), 1143–1150. doi: 10.1111/j.1558-
879 5646.2009.00875.x
- 880 Cole, L. C. (1954). The population consequences of life history phenomena. *The Quarterly*
881 *Review of Biology*, 29(2), 103–137.
- 882 Collier, J. L., Fitzgerald, S. P., Hice, L. A., Frisk, M. G., & McElroy, A. E. (2014). A new PCR-
883 based method shows that blue crabs (*Callinectes sapidus* (Rathbun)) consume winter flounder
884 (*Pseudopleuronectes americanus* (Walbaum)). *PLoS ONE*, 9(1), e85101. doi:
885 10.1371/journal.pone.0085101
- 886 Conde-Padín, P., Grahame, J. W., & Rolán-Alvarez, E. (2007). Detecting shape differences in
887 species of the *Littorina saxatilis* complex by morphometric analysis. *Journal of Molluscan*
888 *Studies*, 73(2), 147–154. doi: 10.1093/mollus/eym009
- 889 Curtis, L. (1987). Vertical distribution of an estuarine snail altered by a parasite. *Science*,
890 235(4795), 1509–1511. doi: 10.1126/science.3823901
- 891 De Francesco, C. G., & Isla, F. I. (2003). Distribution and abundance of hydrobiid snails in a
892 mixed estuary and a coastal lagoon, Argentina. *Estuaries*, 26(3), 790–797. doi:
893 10.1007/BF02711989
- 894 De Jong, G. (1999). Unpredictable selection in a structured population leads to local genetic
895 differentiation in evolved reaction norms. *Journal of Evolutionary Biology*, 12(5), 839–851. doi:
896 10.1046/j.1420-9101.1999.00118.x
- 897 De Jong, Gerdien. (2005). Evolution of phenotypic plasticity: patterns of plasticity and the
898 emergence of ecotypes: Research review. *New Phytologist*, 166(1), 101–118. doi:
899 10.1111/j.1469-8137.2005.01322.x
- 900 Denny, M. W., & Paine, R. T. (1998). Celestial mechanics, sea-level changes, and intertidal
901 ecology. *The Biological Bulletin*, 194(2), 108–115. doi: 10.2307/1543040
- 902 Dillon, R. T., Jacquemin, S. J., & Pyron, M. (2013). Cryptic phenotypic plasticity in populations
903 of the freshwater prosobranch snail, *Pleurocera canaliculata*. *Hydrobiologia*, 709(1), 117–127.
904 doi: 10.1007/s10750-012-1441-1

- 905 D’Incao, F., Silva, K. G., Ruffino, M. L., & Braga, A. C. (1990). Hábito alimentar do
906 caranguejo *Chasmagnathus granulata* Dana, 1851 na barra do Rio Grande, RS (Decapoda,
907 Grapsidae). *Atlântica*, 12(2), 85–93.
- 908 Fox, J. (2003). Effect displays in R for generalised linear models. *Journal of Statistical*
909 *Software*, 8(15), 1–27.
- 910 Fox, J., Weisberg, S., Adler, D., Bates, D., Baud-Bovy, G., Ellison, S., & Heilberger, R. (2011).
911 *Package “car”: Companion to applied regression*.
- 912 Fredensborg, B. L., Mouritsen, K. N., & Poulin, R. (2005). Impact of trematodes on host
913 survival and population density in the intertidal gastropod *Zeacumantus subcarinatus*. *Marine*
914 *Ecology Progress Series*, 290, 109–117. doi: 10.3354/meps290109
- 915 Fredensborg, B. L., Mouritsen, K. N., & Poulin, R. (2006). Relating bird host distribution and
916 spatial heterogeneity in trematode infections in an intertidal snail—from small to large scale.
917 *Marine Biology*, 149(2), 275–283. doi: 10.1007/s00227-005-0184-1
- 918 Fusco, G., & Minelli, A. (2010). Phenotypic plasticity in development and evolution: facts and
919 concepts. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 365(1540),
920 547–556. doi: 10.1098/rstb.2009.0267
- 921 Gaillard, M. C., & Castellanos, Z. A. (1976). Mollusca Gasteropoda Hydrobiidae. In *Fauna de*
922 *agua dulce de la República Argentina* (Ringuélet R. A., Vol. 15, pp. 1–40). Buenos Aires:
923 FECIC.
- 924 Ghalambor, C. K., McKay, J. K., Carroll, S. P., & Reznick, D. N. (2007). Adaptive versus non-
925 adaptive phenotypic plasticity and the potential for contemporary adaptation in new
926 environments. *Functional Ecology*, 21(3), 394–407. doi: 10.1111/j.1365-2435.2007.01283.x
- 927 Hershler, R., & Thompson, F. (1992). A review of the aquatic gastropod subfamily
928 Cochliopinae (Prosobranchia, Hydrobiidae). *Malacological Review*, 5.
- 929 Hill, J. K., Griffiths, H. M., & Thomas, C. D. (2011). Climate change and evolutionary
930 adaptations at species’ range margins. *Annual Review of Entomology*, 56(1), 143–159. doi:
931 10.1146/annurev-ento-120709-144746
- 932 Hoffmann, A. A., & Sgrò, C. M. (2011). Climate change and evolutionary adaptation. *Nature*,
933 470(7335), 479–485. doi: 10.1038/nature09670
- 934 Hollander, J., Collyer, M. L., Adams, D. C., & Johannesson, K. (2006). Phenotypic plasticity in
935 two marine snails: constraints superseding life history. *Journal of Evolutionary Biology*, 19(6),
936 1861–1872. doi: 10.1111/j.1420-9101.2006.01171.x
- 937 Hollander, J., Smadja, C. M., Butlin, R. K., & Reid, D. G. (2013). Genital divergence in
938 sympatric sister snails. *Journal of Evolutionary Biology*, 26(1), 210–215. doi:
939 10.1111/jeb.12029
- 940 Janson, K. (1987). Allozyme and shell variation in two marine snails (*Littorina*, Prosobranchia)
941 with different dispersal abilities. *Biological Journal of the Linnean Society*, 30(3), 245–256. doi:
942 10.1111/j.1095-8312.1987.tb00299.x
- 943 Johannesson, B., & Johannesson, K. (1996). Population differences in behaviour and
944 morphology in the snail *Littorina saxatilis*: phenotypic plasticity or genetic differentiation?
945 *Journal of Zoology*, 240(3), 475–493. doi: 10.1111/j.1469-7998.1996.tb05299.x

- 946 Johannesson, K. (2003). Evolution in Littorina: ecology matters. *Journal of Sea Research*,
947 49(2), 107–117. doi: 10.1016/S1385-1101(02)00218-6
- 948 Kameda, Y., Kawakita, A., & Kato, M. (2009). Reproductive Character Displacement in Genital
949 Morphology in *Satsuma* Land Snails. *The American Naturalist*, 173(5), 689–697. doi:
950 10.1086/597607
- 951 Kemp, P., & Bertness, M. D. (1984). Snail shape and growth rates: Evidence for plastic shell
952 allometry in *Littorina littorea*. *Proceedings of the National Academy of Sciences*, 81(3), 811–
953 813. doi: 10.1073/pnas.81.3.811
- 954 Komsta, L. (2011). outliers: Tests for outliers (Version 0.14). Retrieved from [https://CRAN.R-](https://CRAN.R-project.org/package=outliers)
955 [project.org/package=outliers](https://CRAN.R-project.org/package=outliers)
- 956 Kuris, A. M. (1974). Trophic Interactions: Similarity of Parasitic Castrators to Parasitoids. *The*
957 *Quarterly Review of Biology*, 49(2), 129–148. doi: 10.1086/408018
- 958 Lafferty, K. D. (1993). The marine snail, *Cerithidea californica*, matures at smaller sizes where
959 parasitism is high. *Oikos*, 68(1), 3. doi: 10.2307/3545303
- 960 Lafferty, K. D., Sammond, D. T., & Kuris, A. M. (1994). Analysis of larval trematode
961 communities. *Ecology*, 75(8), 2275. doi: 10.2307/1940883
- 962 Levins, R. (1968). *Evolution in changing environments: some theoretical explorations*. No. 2.
963 Princeton University Press.
- 964 Levri, E. P., Dillard, J., & Martin, T. (2005). Trematode infection correlates with shell shape
965 and defence morphology in a freshwater snail. *Parasitology*, 130(6), 699–708. doi:
966 10.1017/S0031182005007286
- 967 Lewontin, R. C. (1965). Selection for colonizing ability. *Proceedings of the First International*
968 *Union of Biological Sciences Symposia on General Biol*, 77–94. Academic Press.
- 969 Machin, J. (1967). Structural adaptation for reducing water loss in three species of terrestrial
970 snail. *Journal of Zoology*, 152(1), 55–65.
- 971 Mather, K. (1955). Polymorphism as an outcome of disruptive selection. *Evolution*, 9, 52–61.
- 972 Melatunan, S., Calosi, P., Rundle, S., Widdicombe, S., & Moody, A. (2013). Effects of ocean
973 acidification and elevated temperature on shell plasticity and its energetic basis in an intertidal
974 gastropod. *Marine Ecology Progress Series*, 472, 155–168. doi: 10.3354/meps10046
- 975 Merilä, J., & Hendry, A. P. (2014). Climate change, adaptation, and phenotypic plasticity: the
976 problem and the evidence. *Evolutionary Applications*, 7(1), 1–14. doi: 10.1111/eva.12137
- 977 Miura, O., Kuris, A. M., Torchin, M. E., Hechinger, R. F., & Chiba, S. (2006). Parasites alter
978 host phenotype and may create a new ecological niche for snail hosts. *Proceedings of the Royal*
979 *Society B: Biological Sciences*, 273(1592), 1323–1328. doi: 10.1098/rspb.2005.3451
- 980 Mouritsen, K. N., & Jensen, K. T. (1994). The enigma of gigantism: effect of larval trematodes
981 on growth, fecundity, egestion and locomotion in *Hydrobia ulvae* (Pennant)(Gastropoda:
982 Prosobranchia). *Journal of Experimental Marine Biology and Ecology*, 181(1), 53–66.
- 983 Mouritsen, K. N., Sørensen, M. M., Poulin, R., & Fredensborg, B. L. (2018). Coastal
984 ecosystems on a tipping point: Global warming and parasitism combine to alter community
985 structure and function. *Global Change Biology*, 24(9), 4340–4356. doi: 10.1111/gcb.14312

- 986 Neves, R. A. F., Valentin, J. L., & Figueiredo, G. M. (2010). Morphological description of the
987 gastropod *Heleobia australis* (Hydrobiidae) from egg to hatching. *Brazilian Journal of*
988 *Oceanography*, 58(3), 247–250. doi: 10.1590/S1679-87592010000300007
- 989 Palmer, A. R. (1981). Do carbonate skeletons limit the rate of body growth? *Nature*, 292(5819),
990 150–152.
- 991 Palmer, A. R. (1990). *Effect of crab effluent and scent of damaged conspecifics on feeding,*
992 *growth, and shell morphology of the Atlantic dogwhelk *Nucella lapillus* (L.).* 28.
- 993 Parsons, K. E. (1997). Role of dispersal ability in the phenotypic differentiation and plasticity of
994 two marine gastropods. *Oecologia*, 110(4), 461–471. doi: 10.1007/s004420050181
- 995 Parsons, K. E. (1998). The Role of dispersal ability in the phenotypic differentiation and
996 plasticity of two marine gastropods II. Growth. *J. Exp. Mar. Biol. Ecol.*, 25.
- 997 Perillo, G. M. E., Piccolo, M. C., Parodi, E., & Freije, R. H. (2001). The Bahia Blanca Estuary,
998 Argentina. In U. Seeliger & B. Kjerfve (Eds.), *Coastal Marine Ecosystems of Latin America*
999 (Vol. 144, pp. 205–217). Berlin, Heidelberg: Springer Berlin Heidelberg. doi: 10.1007/978-3-
1000 662-04482-7_15
- 1001 Pfennig, D. W., Wund, M. A., Snell-Rood, E. C., Cruickshank, T., Schlichting, C. D., &
1002 Moczek, A. P. (2010). Phenotypic plasticity's impacts on diversification and speciation. *Trends*
1003 *in Ecology & Evolution*, 25(8), 459–467. doi: 10.1016/j.tree.2010.05.006
- 1004 Pigliucci, M., Murren, C. J., & Schlichting, C. D. (2006). Phenotypic plasticity and evolution by
1005 genetic assimilation. *Journal of Experimental Biology*, 209(12), 2362–2367. doi:
1006 10.1242/jeb.02070
- 1007 Pigliucci, Massimo. (2001). *Phenotypic plasticity: beyond nature and nurture*. Baltimore: Johns
1008 Hopkins University Press.
- 1009 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team. (2017). *nlme: linear and*
1010 *nonlinear mixed effects models. R package version 3.1-131*. Retrieved from [https://CRAN.R-](https://CRAN.R-project.org/package=nlme)
1011 [project.org/package=nlme](https://CRAN.R-project.org/package=nlme)
- 1012 Pointier, J., Noya, O., Amarista, M., & Théron, A. (2004). *Lymnaea cousini* Jousseaume, 1887
1013 (Gastropoda: Lymnaeidae): first record for Venezuela. *Memórias Do Instituto Oswaldo Cruz*,
1014 99(6), 567–569. doi: 10.1590/S0074-02762004000600005
- 1015 Poulin, R., & Thomas, F. (1999). Phenotypic Variability Induced by Parasites: *Parasitology*
1016 *Today*, 15(1), 28–32. doi: 10.1016/S0169-4758(98)01357-X
- 1017 Pratolongo, P., Kirby, J., Plater, A., & Brinson, M. (2009). Temperate coastal wetlands:
1018 morphology, sediment processes, and plant communities. In *An Integrated Ecosystem*
1019 *Approach, Coastal Wetlands* (Perillo, G.M.E., Wolanski, E., Cahoon, D.R., Brinson, M.M.
1020 (Eds.), pp. 185–210). Amsterdam: Elsevier.
- 1021 Pratolongo, P., Perillo, G. M. E., & Piccolo, M. C. (2010). Combined effects of waves and
1022 plants on a mud deposition event at a mudflat-saltmarsh edge in the Bahía Blanca estuary.
1023 *Estuarine, Coastal and Shelf Science*, 87(2), 207–212. doi: 10.1016/j.ecss.2009.09.024
- 1024 Price, T. D., Qvarnström, A., & Irwin, D. E. (2003). The role of phenotypic plasticity in driving
1025 genetic evolution. *Proceedings of the Royal Society of London. Series B: Biological Sciences*,
1026 270(1523), 1433–1440. doi: 10.1098/rspb.2003.2372

- 1027 Probst, S., & Kube, J. (1999). Histopathological effects of larval trematode infections in
1028 mudsnails and their impact on host growth: what causes gigantism in *Hydrobia ventrosa*
1029 (Gastropoda: Prosobranchia)? *Journal of Experimental Marine Biology and Ecology*, 238(1),
1030 49–68. doi: 10.1016/S0022-0981(99)00002-7
- 1031 Puillandre, N., Lambert, A., Brouillet, S., & Achaz, G. (2012). ABGD, Automatic Barcode Gap
1032 Discovery for primary species delimitation: ABGD, automatic barcode gap discovery.
1033 *Molecular Ecology*, 21(8), 1864–1877. doi: 10.1111/j.1365-294X.2011.05239.x
- 1034 Reid, D. G. (1996). *Systematics and evolution of Littorina*. London: Ray Soc. Publs.
- 1035 Revelle, W. (2018). psych: Procedures for Personality and Psychological Research (Version
1036 1.8.12.). Northwestern University, Evanston, Illinois, USA. Retrieved from [https://CRAN.R-](https://CRAN.R-project.org/package=psych)
1037 [project.org/package=psych](https://CRAN.R-project.org/package=psych)
- 1038 Ripley, B., Venables, B., Bates, D. M., Hornik, K., Gebhardt, A., Firth, D., & Ripley, M. B.
1039 (2013). Package ‘mass’ (Version 538).
- 1040 Roff, D. A. (1992). *The evolution of life histories: theory and analysis*. New York: Chapman &
1041 Hall.
- 1042 Rolán-Alvarez, E. (2007). Sympatric speciation as a by-product of ecological adaptation in the
1043 Galician *Littorina saxatilis* hybrid zone. *Journal of Molluscan Studies*, 73(1), 1–10. doi:
1044 10.1093/mollus/eyl023
- 1045 Scheiner, S. M. (1998). The genetics of phenotypic plasticity. VII. Evolution in a spatially-
1046 structured environment. *Journal of Evolutionary Biology*, 11(3), 303–320. doi: 10.1046/j.1420-
1047 9101.1998.11030303.x
- 1048 Schweizer, M., Triebkorn, R., & Köhler, H. (2019). Snails in the sun: Strategies of terrestrial
1049 gastropods to cope with hot and dry conditions. *Ecology and Evolution*, 9(22), 12940–12960.
1050 doi: 10.1002/ece3.5607
- 1051 Smith, N. F. (2001). Spatial heterogeneity in recruitment of larval trematodes to snail
1052 intermediate hosts. *Oecologia*, 127(1), 115–122. doi: 10.1007/s004420000560
- 1053 Spivak, E., Anger, K., Luppi, T., Bas, C., & Ismael, D. (1994). Distribution and habitat
1054 preferences of two grapsid crab species in Mar Chiquita Lagoon (Province of Buenos Aires,
1055 Argentina). *Helgoländer Meeresuntersuchungen*, 48(1), 59–78. doi: 10.1007/BF02366202
- 1056 Struhsaker, J. W. (1968). Selection mechanisms associated with intraspecific shell variation in
1057 *Littorina picta* (Prosobranchia: Mesogastropoda). *Evolution*, 22(3), 459–480. doi:
1058 10.1111/j.1558-5646.1968.tb03986.x
- 1059 Suzuki, Y., & Nijhout, H. F. (2006). Evolution of a polyphenism by genetic accommodation.
1060 *Science*, 311(5761), 650–652. doi: 10.1126/science.1118888
- 1061 Travis, J. (1994). Evaluating the adaptive role of morphological plasticity. In *Ecological*
1062 *Morphology: Integrative Organismal Biology* (eds P.C. Wainwright & S.M. Reilly, pp. 99–122).
1063 Chicago: University of Chicago Press.
- 1064 Trussell, G. C. (2000a). Phenotypic clines, plasticity, and morphological trade-offs in an
1065 intertidal snail. *Evolution*, 54(1), 151–166.
- 1066 Trussell, G. C. (2000b). Predator-induced plasticity and morphological trade-offs in latitudinally
1067 separated populations of *Littorina obtusata*. *Evolutionary Ecology Research*, 2(6), 803–822.

- 1068 Vermeij, G. J. (1972). Intraspecific shore-level size gradients in intertidal molluscs. *Ecology*,
1069 53(4), 693–700. doi: 10.2307/1934785
- 1070 Vermeij, G. J. (1973). Morphological patterns in high-intertidal gastropods: Adaptive strategies
1071 and their limitations. *Marine Biology*, 20(4), 319–346. doi: 10.1007/BF00354275
- 1072 Vermeij, G. J. (1982). Phenotypic evolution in a poorly dispersing snail after arrival of a
1073 predator. *Nature*, 299(5881), 349–350.
- 1074 Via, S., & Lande, R. (1985). Genotype-Environment Interaction and the Evolution of
1075 Phenotypic Plasticity. *Evolution*, 39(3), 505. doi: 10.2307/2408649
- 1076 Vu, V. Q. (2011). ggbiplot: A ggplot2 based biplot (Version 0.55). Retrieved from
1077 <http://github.com/vqv/ggbiplot>
- 1078 Waddington, C. H. (1953). Genetic assimilation of an acquired character. *Evolution*, 7, 118–
1079 126.
- 1080 West-Eberhard, M. J. (1989). Phenotypic plasticity and the origins of diversity. *Annual Review*
1081 *of Ecology and Systematics*, 2, 249–278.
- 1082 West-Eberhard, M. J. (2003). *Developmental plasticity and evolution*. Oxford; New York:
1083 Oxford University Press.
- 1084 West-Eberhard, M. J. (2005). Developmental plasticity and the origin of species differences.
1085 *Proceedings of the National Academy of Sciences*, 102(Supplement 1), 6543–6549. doi:
1086 10.1073/pnas.0501844102
- 1087 Wickham, H. (2011). The Split-Apply-Combine Strategy for Data Analysis. *Journal of*
1088 *Statistical Software*, 40(1), 1–29.
- 1089 Wickham, H., & Chang, W. (2008). ggplot2: an Implementation of the Grammar of Graphics
1090 (Version 0.7. 0.7). Retrieved from <http://CRAN.R-Project.Org/package=ggplot2>
- 1091 Wickham, H., François, R., Henry, L., & Müller, K. (2019). dplyr: A Grammar of data
1092 manipulation (Version 0.8.3). Retrieved from <https://CRAN.R-project.org/package=dplyr>
- 1093 Wickham, H., Hester, J., & Chang, W. (2019). devtools: Tools to Make Developing R Packages
1094 Easier (Version 2.1.0). Retrieved from <https://CRAN.R-project.org/package=devtools>
- 1095 Yamada, S. B. (1987). Geographic variation in the growth rates of *Littorina littorea* and *L.*
1096 *saxatilis*. *Marine Biology*, 96(4), 529–534. doi: 10.1007/BF00397970
- 1097 Zuur, A. F. (Ed.). (2009). *Mixed effects models and extensions in ecology with R*. New York,
1098 NY: Springer.
- 1099

1100 **Tables**

1101 **Table 1.** Summary of means (\pm SE) and statistical significance of the two-way ANOVA testing for habitat and seasonal effect on total snail
 1102 density (ind./per sample; one sample being 78.5 cm⁻²) of the intertidal mud snail *Heleobia australis* from the Bahía Blanca estuary, Argentina.
 1103 Number of samples are indicated between parentheses. Value in bold indicate overall mean. We reported raw estimates whereas the model fits
 1104 for Ln-transformed density. See Methods section for details.

1105

	Seasons	Habitats			Mean by Season	Habitat effect	Habitat comparison	Effect size	Season effect	Season comparison	Effect size	
		Flats (F)	Marshes (M)	Pans (P)								
Total density	Summer (Su)	132±17 (9)	52±14 (9)	128±20 (9)	104±12 (27)	$F_{(2, 106)} = 15.48^{***}$	P vs. M	5.49 ^{***}	$F_{(3, 106)} = 6.22^{***}$	Au vs. Sp	4.26 ^{***}	
	Autumn (Au)	65±13 (9)	87±42 (9)	70±14 (9)	74±15 (27)		M vs. F			-3.54 ^{***}	Wi vs. Sp	2.69*
	Winter (Wi)	92±20 (8)	70±24 (9)	206±30 (9)	124±18 (26)		P vs. F			1.91	Su vs. Sp	2.57*
	Spring (Sp)	187±36 (9)	114±36 (9)	317±34 (9)	206±26 (27)						Su vs. Au	1.68
	Mean by Habitat	120±14 (35)	81±15 (36)	180±20 (36)	127±10 (107)					Au vs. Wi	1.53	
										Su vs. Wi	0.14	

1106

1107 * $P < 0.05$; ** $P < 0.01$; *** $P < 0.001$. We applied the Holm-Bonferroni correction for multiple testing when comparing effect sizes.

1108 **Table 2.** Uninfected adult snails. Summary of means (\pm SE) and statistical significance of habitat and sex effect on shell and aperture
 1109 morphometry of uninfected adult individuals of the mud snail *Heleobia australis* from the intertidal area of the Bahía Blanca estuary, Argentina.
 1110 Variables are the same as indicated in table 2. Number of observations are indicated between parentheses, which are only shown for shell size but
 1111 are the same for other traits measured.
 1112

Traits measured	Habitat type	Sex		Sex pooled	Habitat effect	Habitat comparison	Effect size	Sex effect	Sex comparison	Effect size
		Males (♂)	Females (♀)							
Shell size Φ	Flats (F)	6.19 \pm 0.18 (388)	6.17 \pm 0.18 (396)	6.17 \pm 0.16 (784)	<i>Males</i> $X_2^2 = 346.22^{***}$	P vs. F	17.91 ^{***}	<i>Flats</i> $X_1^2 = 0.20$; n.s.	♂ vs. ♀	0.50
	Marshes (M)	6.78 \pm 0.13 (266)	6.64 \pm 0.11 (245)	6.71 \pm 0.11 (511)		P vs. M	11.01 ^{***}			
	Pans (P)	7.97 \pm 0.36 (600)	7.55 \pm 0.40 (738)	7.74 \pm 0.40 (1338)	<i>Females</i> $X_2^2 = 249.33^{***}$	M vs. F	4.08 ^{***}	<i>Marshes</i> $X_1^2 = 1.57$; n.s.	♂ vs. ♀	1.25
	<i>Habitats pooled</i>	7.14 \pm 0.22 (1254)	6.95 \pm 0.25 (1379)			P vs. F	15.20 ^{***}			
						P vs. M	9.01 ^{***}	<i>Pans</i> $X_1^2 = 24.30^{***}$	♂ vs. ♀	4.93 ^{***}
						M vs. F	3.62 ^{***}			
Shell shape †	Flats	2.27 \pm 0.02	2.26 \pm 0.02	2.27 \pm 0.02	$X_2^2 = 101.02^{***}$	P vs. F	10.13 ^{***}	$X_1^2 = 29.02^{***}$	♂ vs. ♀	5.39 ^{***}
	Marshes	2.32 \pm 0.01	2.29 \pm 0.01	2.31 \pm 0.01		P vs. M	5.31 ^{***}			
	Pans	2.35 \pm 0.01	2.31 \pm 0.01	2.33 \pm 0.01	M vs. F	3.61 ^{***}				
	<i>Habitats pooled</i>	2.32 \pm 0.01	2.29 \pm 0.01							
Aperture size	Flats	1.82 \pm 0.05	1.83 \pm 0.06	1.82 \pm 0.05	$X_2^2 = 362.62^{***}$	P vs. F	17.10 ^{***}	$X_1^2 = 8.54^{**}$	♂ vs. ♀	2.93 ^{**}
	Marshes	1.90 \pm 0.04	1.86 \pm 0.03	1.88 \pm 0.03		P vs. M	12.79 ^{***}			
	Pans	2.10 \pm 0.09	2.05 \pm 0.10	2.07 \pm 0.09	M vs. F	1.63				
	<i>Habitats pooled</i>	1.97 \pm 0.06	1.94 \pm 0.07							
Aperture shape	Flats	1.73 \pm 0.03	1.73 \pm 0.03	1.73 \pm 0.03	$X_2^2 = 535.73^{***}$	P vs. F	-22.91 ^{***}	$X_1^2 = 0.04$; n.s.	♂ vs. ♀	-0.20
	Marshes	1.70 \pm 0.02	1.71 \pm 0.01	1.70 \pm 0.02		P vs. M	-15.35 ^{***}			
	Pans	1.60 \pm 0.02	1.60 \pm 0.02	1.60 \pm 0.02	M vs. F	-3.86 ^{***}				
	<i>Habitats pooled</i>	1.66 \pm 0.02	1.66 \pm 0.03							

1113 Φ As there was a significant interaction between Habitat and Sex when testing shell size, we report effect sizes separately for this trait.
 1114 † We report estimates of shell shape corrected for shell size, as the pattern observed was not the same when removing the covariate from the model.
 1115 * $P < 0.05$; ** $P < 0.01$; *** $P < 0.001$. We applied the Holm-Bonferroni correction for multiple testing when comparing effect sizes
 1116
 1117
 1118

1119 **Table 3.** Infected vs. Uninfected adult snails. Summary of means (\pm SE) and statistical significance of status (infected of uninfected) and sex
 1120 effect on shell and aperture morphometry of the intertidal mud snail *Heleobia australis* from the Bahía Blanca estuary, Argentina. In these
 1121 analyses we only considered individuals from pans, which allowed for preventing habitat effect on morphometry. Shell size estimated as the
 1122 volume of a cone (mm^3), shell shape as the length to width ratio (SL/SW), aperture size as the area of an ellipse (mm^2), and aperture shape as the
 1123 ratio between aperture length and width (AL/AW). Number of observations are indicated between parentheses, which are only shown for shell
 1124 size but are the same for other traits measured.
 1125

Traits measured	Habitat type	Sex		Sex pooled	Status effect	Status comparison	Effect size	Sex Effect	Sex comparison	Effect size	
		Male (♂)	Female (♀)								
Shell size	Uninfected (U)	7.97 \pm 0.36 (600)	7.55 \pm 0.40 (738)	7.74 \pm 0.38 (1338)	$X_1^2 = 149.69^{***}$	U vs. I	12.42 ***	$X_1^2 = 31.68^{***}$	♂ vs. ♀	5.65 ***	
	Infected (I)	7.01 \pm 0.14 (432)	6.76 \pm 0.20 (573)	6.86 \pm 0.16 (1005)							
	Status pooled	7.59 \pm 0.28 (1032)	7.22 \pm 0.32 (1311)								
Shell shape Φ	Uninfected	2.36 \pm 0.01	2.32 \pm 0.01	2.34 \pm 0.01	$X_1^2 = 11.82^{***}$	U vs. I	3.43 ***	$X_1^2 = 37.51^{***}$	♂ vs. ♀	6.12 ***	
	Infected	2.40 \pm 0.01	2.39 \pm 0.02	2.40 \pm 0.01							
	Status pooled	2.38 \pm 0.01	2.35 \pm 0.01								
					<i>Males</i>		<i>Uninfected</i>				
					$X_1^2 = 99.50^{***}$		U vs. I		9.98 ***		$X_1^2 = 0.09$; n.s.
					<i>Females</i>		<i>Infected</i>				
Aperture size	Uninfected	2.10 \pm 0.09	2.05 \pm 0.10	2.07 \pm 0.09	$X_1^2 = 408.65^{***}$	U vs. I	21.11 ***	$X_1^2 = 14.95^{***}$	♂ vs. ♀	3.88 ***	
	Infected	1.80 \pm 0.04	1.76 \pm 0.06	1.77 \pm 0.05							
	Status pooled	1.98 \pm 0.07	1.92 \pm 0.08								
Aperture shape	Uninfected	1.60 \pm 0.02	1.60 \pm 0.02	1.60 \pm 0.02	$X_1^2 = 60.57^{***}$	U vs. I	7.83 **	$X_1^2 = 1.3e-3$; n.s.	♂ vs. ♀	0.03	
	Infected	1.56 \pm 0.02	1.56 \pm 0.02	1.56 \pm 0.02							
	Status pooled	1.58 \pm 0.02	1.58 \pm 0.02								

1126
 1127 Φ As there was a significant interaction between Status and Sex when testing shell shape, we report effect sizes separately for this trait.
 1128 * $P < 0.05$; ** $P < 0.01$; *** $P < 0.001$. We applied the Holm-Bonferroni correction for multiple testing when comparing effect sizes.
 1129

1130 **Figure Captions**

1131 **Figure 1.** Variation in total density (juveniles and adults pooled) of the planktotrophic
1132 snail *Heleobia australis* across habitats and seasons in the intertidal area of the Bahía
1133 Blanca estuary, Argentina. Bars represent ± 1 SE.

1134

1135 **Figure 2.** Phenotypic variation in shell and aperture morphometrics of uninfected adult
1136 snails *Heleobia australis* (cat. *i*) in the Bahía Blanca estuary, Argentina. Blue and red
1137 dots indicate mean values of each variable and sex in each habitat (flats, marshes, and
1138 pans). Mean values of shell shape were statistically corrected for shell size. Bars
1139 represent ± 1 SE.

1140

1141 **Figure 3.** Phenotypic variation in shell and aperture morphometrics of infected and
1142 uninfected adult snails *Heleobia australis* (cat. *iv*) in response to a strong parasite
1143 pressure from a habitat with low environmental stress conditions but high parasite
1144 prevalence (pans) in the Bahía Blanca estuary, Argentina. Blue and red dots indicate
1145 mean values of each variable and sex (males and females, respectively). Mean values of
1146 aperture size were statistically corrected for shell size. Bars represent ± 1 SE.

1147

1148 **Figure 4.** Shell weight (as a proxy of thickness) and body mass variation of the
1149 intertidal mud snail *Heleobia australis* across habitats with different biotic and abiotic
1150 selective pressures in the Bahía Blanca estuary, Argentina. Bars represent ± 1 SE.

1151

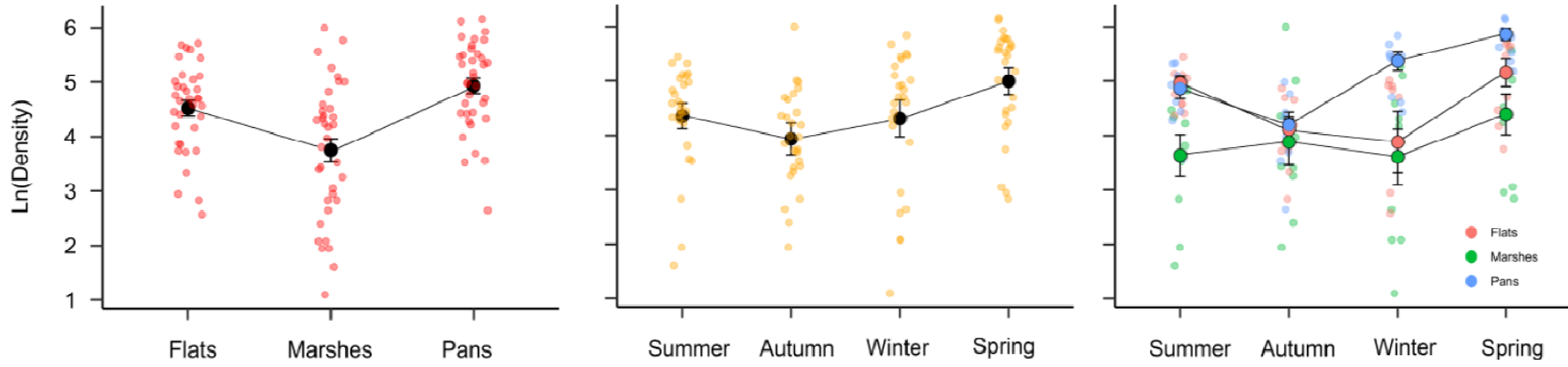
1152 **Figure 5.** Differences in penis size (mm^2) of uninfected adult snails *Heleobia australis*
1153 collected from three environmentally distinct habitats from the Bahía Blanca estuary,
1154 Argentina and kept in standard laboratory conditions for 21 months, which ensured that
1155 individuals analyzed were all adults. Bars represent ± 1 SE.

1156

1157 **Figure 6.** Summary of the responses observed in snail population density, shell and
1158 aperture morphometry, and shell and body weight of the intertidal mud snail *Heleobia*
1159 *australis* from the Bahía Blanca estuary, Argentina. Such responses were induced by: (i)
1160 environmental pressure, whereby snails experiencing prolonged submersion times in
1161 habitats that remain covered by water during low tide (pans) exhibited higher density
1162 and higher mean values of morphological traits relative to those from more

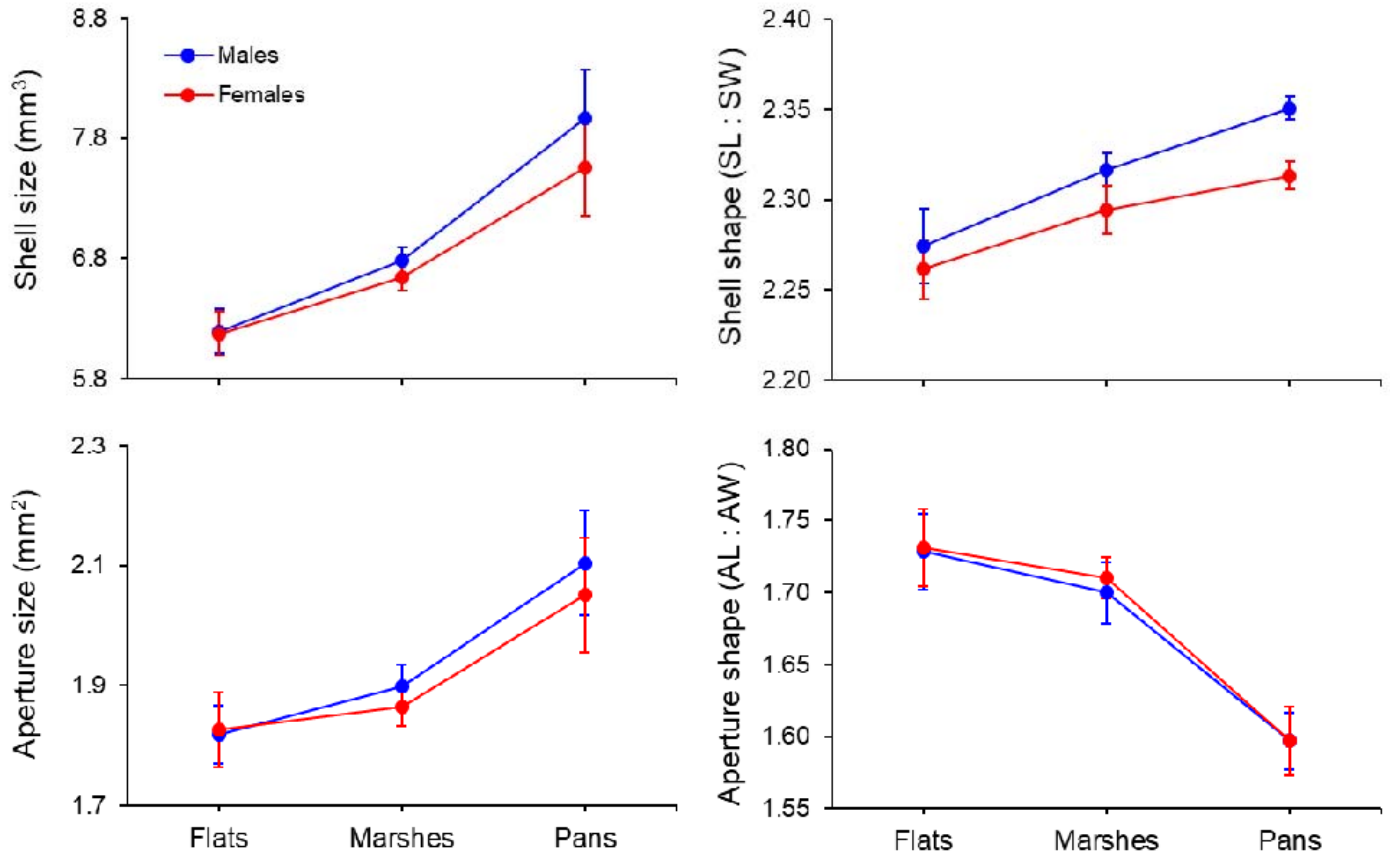
1163 environmental stress conditions (flats and marshes), which showed opposite phenotypic
1164 responses; (ii) predation pressure caused by the presence of the burrowing crab
1165 *Neohelice granulata* could be linked to the remarkable low density in marshes and
1166 could have favored the expression of shell defenses as shell thickening, increased shell
1167 globosity (lower SL to SW ratio), narrower apertures (higher AL to AW ratio); and (iii)
1168 strong parasite pressure caused by *Microphallus simillimus* where morphological traits
1169 of infected (I) snails shifted to lower mean values compared to uninfected (U) ones;
1170 note that polymorphisms are only shown for pans, where parasite prevalence was the
1171 highest. Responses observed in this study are indicated by color-filled boxes, white-
1172 filled boxes indicate possible mechanisms or processes that might explain the observed
1173 responses. Green color refers to responses induced by environmental pressure, orange to
1174 polymorphism induced by crab predation, and pink by parasite pressure. Upper arrows
1175 (↑) indicate increase/higher/longer and down arrows (↓) indicate decrease/lower/shorter.

1176 **Figures**



1177

1178 **Figure 1.**

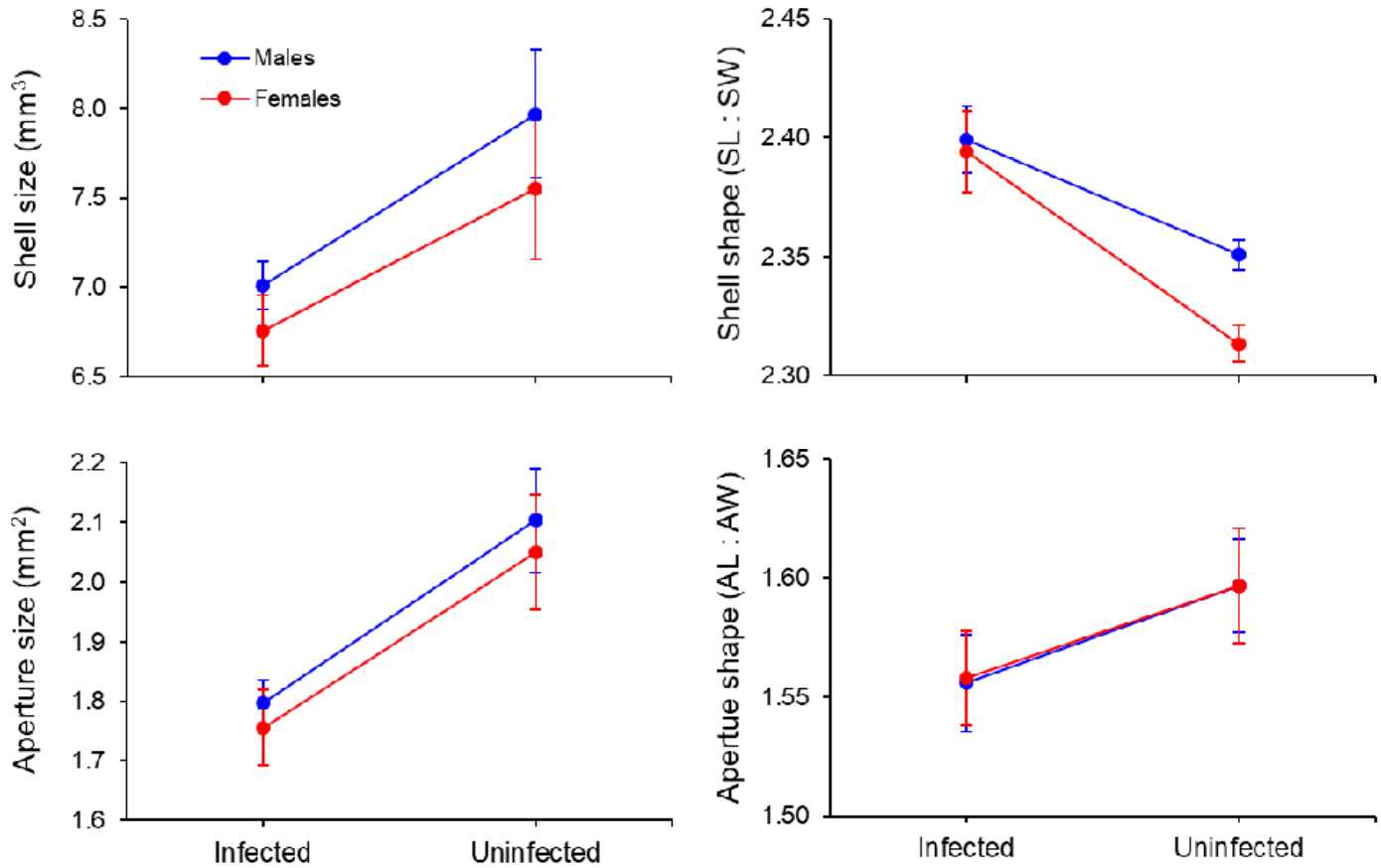


1179

1180

Figure 2.

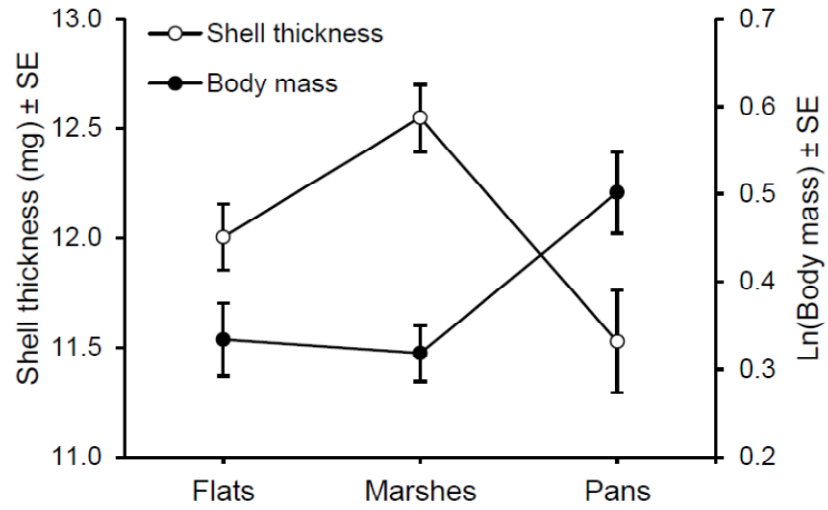
1181



1182

1183

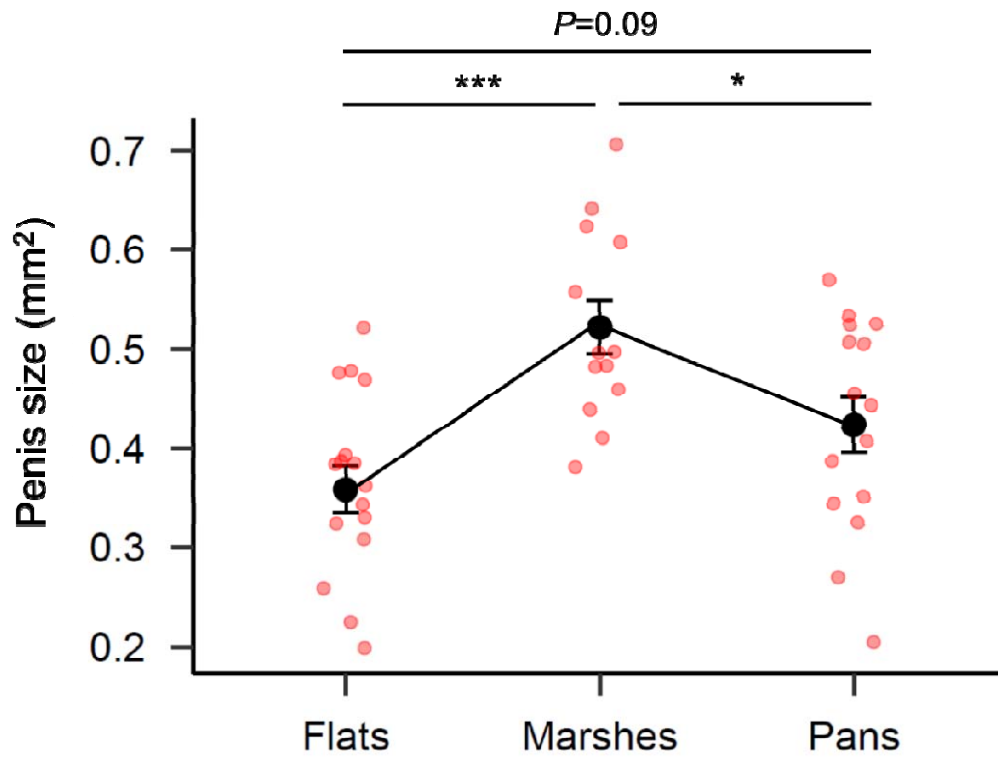
Figure 3.



1184

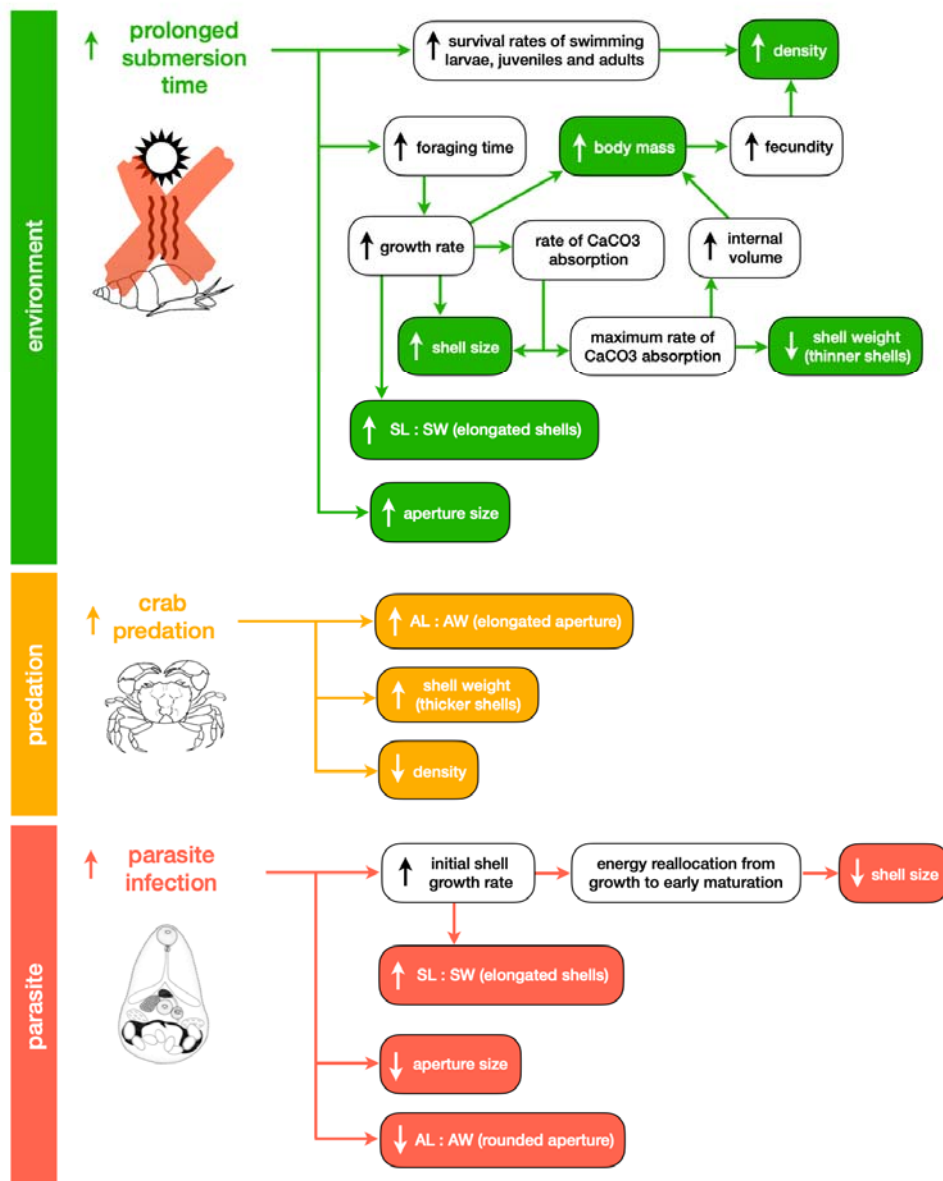
1185 **Figure 4.**

1186



1187

1188 **Figure 5.**



1189

1190 **Figure 6.**

1191