Activity, energy use and species identity affect eDNA shedding in freshwater fish

Bettina Thalinger[1,2*], Andreas Rieder[1], Anna Teuffenbach[1], Yannick Pütz[1], Thorsten Schwerte[1], Josef Wanzenböck[3] & Michael Traugott[1,4]

[1] *Department of Zoology, University of Innsbruck, Technikerstr. 25, 6020 Innsbruck, Austria*

[2] *Centre for Biodiversity Genomics, University of Guelph, 50 Stone Road E, N1G2W1, Guelph, Ontario, Canada*

[3] *Research Department for Limnology, University of Innsbruck, Mondseestr. 9, 5310 Mondsee, Austria*

[4] *Sinsoma GmbH, Lannes 6, 6176 Voels, Austria*

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Corresponding Author:

Bettina Thalinger

Bettina.Thalinger@gmail.com

Abstract

The quantitative measurement of eDNA form field-collected water samples is gaining

importance for the monitoring of fish communities and populations. The interpretation of

these signal strengths depends, among other factors, on the amount of target eDNA shed

into the water. However, shedding rates are presumably associated with species-specific

traits such as physiology and behavior. Although such differences between juvenile and adult

fish have been previously detected, the general impact of movement and energy use in a

resting state on eDNA release into the surrounding water remains hardly addressed.

In an aquarium experiment, we compared eDNA shedding between seven fish

species occurring in European freshwaters. The investigated salmonids, cyprinids and

sculpin exhibit distinct adaptions to microhabitats, diets, and either solitary or schooling

behavior. The fish were housed in aquaria with constant water flow and their activity was

measured by snapshots taken every 30 s. Water samples for eDNA analysis were taken

every 3 h and energy use was determined in an intermittent flow respirometer. After

controlling for the effect of fish mass, our results demonstrate a positive correlation between

target eDNA quantities as measured with digital PCR, fish activity and energy use, as well as

species-specific differences. For cyprinids, the model based on data from individual fish was

only partly transferable to groups, which exhibited lower activity and higher energy use.

Our findings highlight the importance of fish physiology and behavior for the

comparative interpretation of taxon-specific eDNA quantities. Species traits should therefore

be incorporated into eDNA-based monitoring and conservation efforts.

Introduction

The sensitivity, non-invasiveness and cost-efficiency of environmental DNA (eDNA) based methods has been proven for diverse habitats and species making them powerful new tools for conservation biology and biodiversity assessments (Barnes and Turner, 2016; Deiner et al., 2017; Huerlimann et al., 2020). Regarding the detection of fish species, eDNA-based monitoring outperforms traditional methods such as electrofishing: for example, for the detection of the endangered European weather loach, *Misgurnus fossilis* (Sigsgaard et al., 2015), the assessment of fish communities in Australian streams (McColl-Gausden et al., 2020), and the distribution of brook trout, *Salvelinus fontinalis* in a US watershed (Evans et al., 2017). The manifold successes of eDNA-based species detection lead to a call for more standardization and better reporting practices (Goldberg et al., 2016; Minamoto et al., 2020; Thalinger et al., 2020a) and to an international effort for implementing the technology into routine species monitoring (Leese et al., 2016; Pilliod et al., 2019). Although reporting the presence/absence of particular species is the starting point of these endeavors, a more quantitative interpretation of field-derived eDNA data is key for the broader applicability of this technology.

Different processes influence the distribution of eDNA in space and time and the detection probabilities of species from environmental samples, namely the origin, degradation, suspension, resuspension, and transport of eDNA (Barnes and Turner, 2016; Harrison et al., 2019). The latter processes are directly linked to local hydrology (e.g. flow and substrate type (Shogren et al., 2017; Pont et al., 2018; Thalinger et al., 2020b)) and environmental conditions (e.g. water temperature, pH, UV-radiation (Strickler et al., 2015; Lacoursière-Roussel et al., 2016; Tsuji et al., 2017)). The amount of eDNA in the water column is directly linked to fish biomass and originally, this was confirmed for common carp (*Cyprinus carpio*) in an aquarium trial and experimental ponds (Takahara et al., 2012). In subsequent experiments, this positive relationship was confirmed for a range of freshwater and marine fish species (Evans et al., 2016; Lacoursière-Roussel et al., 2016; Sassoubre et

al., 2016; Doi et al., 2017; Horiuchi et al., 2019; Jo et al., 2020). However, these results were

primarily obtained for individuals at the same life stage.

eDNA is released into the environment in the form of mucus, feces, scales, and

gametes (Merkes et al., 2014; Barnes and Turner, 2016; Sassoubre et al., 2016; Bylemans

et al., 2017). Under natural conditions, differences in fish physiology, diet and behavior are

likely to affect this process and confound the interpretation of eDNA-based results from a

water body (Klymus et al., 2015). For perch and eel, Maruyama et al. (2014) and Takeuchi et

al. (2019), respectively, found lower eDNA shedding rates for adults in comparison to

juveniles, which is likely caused by the ontogenetic decrease in metabolic rates (Winberg,

1960). However, these findings could not be confirmed in an experiment with a salmonid

species (Mizumoto et al., 2018). In general, the metabolic rate and activity differ between fish

species due to distinct physiology and behavior with pelagic species being more active and

displaying higher resting metabolic rates than benthic species (Johnston et al., 1988; Killen

et al., 2010). These factors are frequently hypothesized as underlying causes for spiking

eDNA levels at the beginning of aquarium experiments and mismatching quantitative results

in studies comparing eDNA levels between species in the same water body (Takahara et al.,

2012; Maruyama et al., 2014; Evans et al., 2016).

Here, we investigate the effect of fish activity (i.e. movement), energy use (i.e. oxygen

use × oxycaloric factor), and species identity in an aquarium experiment with seven fish

species commonly occurring in European rivers and streams. The examined species

comprised four salmonids (*Salmo trutta, Salvelinus fontinalis, Oncorhynchus mykiss,*

Thymallus thymallus), two cyprinids (*Phoxinus phoxinus*, *Squalius cephalus*) and one sculpin

(*Cottus gobio*). *Salmo trutta* is a rhithral species, territorial in later life stages, and primarily

feeds on benthic organisms and insects drift on the surface. *Salvelinus fontinalis* and *O.*

mykiss were anthropogenically introduced into European freshwaters; they are omnivorous

and less territorial than *S. trutta*. If possible, these three species choose areas with reduced

current close to the main riverbed as preferential microhabitat. *Thymallus thymallus* is also a

rhithral species, but its scales are larger and adults primarily use the main riverbed (Spindler,

1997; Freyhof and Kottelat, 2007). *Phoxinus phoxinus* is a schooling, small fish species in

the rhithral. It feeds on a mixture of plant debris, algae and small invertebrates. The juveniles

prefer vegetation-rich microhabitats without current, while adults switch to gravel substrate

with low to intermediate flow. *Squalius cephalus* is eurytopic and can occur in habitats with

strong to low current. Its juveniles are schooling and omnivorous with adults predominantly

preying on fish. *Cottus gobio* is a rheophilic and benthic species primarily feeding on small

bottom invertebrates. It has no swim bladder and mostly resides in interstices between large

boulders or on coarse gravel characterized by low current (Spindler, 1997; Freyhof and

Kottelat, 2007).

We hypothesized that higher activity (i.e. movement) leads to higher eDNA

concentrations as there is more shearing between the fish surface and the surrounding

water, and higher volumes are pumped through the gills due to the elevated oxygen demand.

Independent of activity, fish species with higher energy use in a resting state potentially also

emit more eDNA. Additionally, the species-specific composition of the constantly renewed

cutaneous mucus layer (Ángeles Esteban, 2012) might lead to differences between

individual taxa.

Materials and Methods

Experimental setup

The aquarium experiment was carried out between 2nd March and 17th July 2017 at the

Research Department for Limnology Mondsee of the University of Innsbruck, Austria (further

on "ILIM"). The salmonid species were purchased from commercial hatcheries, *P. phoxinus*

and *S. cephalus* were caught with permission in Lake Mondsee and *C. gobio* were caught

with permission in rivers in Tyrol (Austria). Fish individual sizes were chosen as similar as

possible within and between species. Until the start of the experiment, the fish species were

kept separately in aquaria fed with lake water at the ILIM.

In accordance with the regulations of the Austrian Animal Experiment Act (December

28, 2012) (Tierversuchsrechtsänderungsgesetz, part 1, section 1, §1, point 2), and with the

Directive 2010/63/EU of the European Parliament and of the Council of the European Union

(September 22, 2010) on the protection of animals used for scientific purposes (chapter 1,

article 1, point 5a), all fish were reared according to regular agriculture (aquaculture)

practice, including provision of appropriate tank size, sufficient rate of waterflow, natural

photoperiod, ad libitum food supply, and temperatures within the species' thermal tolerance

range. This ensured that no pain, suffering, distress or lasting harm was inflicted on the

animals, confirmed by the fact that mortality rates were low and equal between rearing

groups. Based on the legislative provisions above, no ethics approval and no IACUC protocol

was required for the experiments performed. In particular the respirometry experiments were

discussed with the legislative authorities (Austrian Federal Ministry of Education, Science

and Research and University of Veterinary Medicine, Vienna) and the conclusion was that

the assessment of basic metabolism under these conditions (small fish sizes in relatively

large chambers) does not incur pain, suffering or distress to the fish and no formal animal

experimentation protocol was required.

Five aquaria (60 l) and corresponding plastic lids were used in the experiment, each

of which was thoroughly cleaned with sodium hypochlorite (5 %) and then rinsed with tap

water (fish-DNA free) prior to each experimental run (i.e. changing the fish under

investigation). The flow-through rate (aquaria were tap-water fed) was set to 5.45 l/min to

mimic natural conditions and keep eDNA concentrations in the fish tanks constant based on

previous test runs (data not shown). Water temperature was kept stable at 15 °C for inflowing

water automatically stabilizing temperature in the aquaria at this level. Each tank was further

equipped with an air-stone to ensure water mixing. At the start of each experimental run, a

water sample (negative control) was taken from one of the aquaria and processed as

described below. Then, five fish individuals of each species were selected aiming at similar

sizes across all fish species. Each fish was placed individually in an aquarium using DNA-

free fishnets (Fig. 1). For *P. phoxinus* and *S. cephalus*, the experiment was carried out twice:

once with individual fish, and once with groups of three fish per aquarium. The day before the

experiment and for its duration, the respective fish were not fed to avoid contamination by

fish feed and minimize effects of defecation. Each run started with one day of familiarization

in the aquaria.

Water sampling, filtration and pH

All equipment used for this process was cleaned with sodium hypochlorite (5 %) and rinsed

with tap water prior to each use; DNA-free gloves were always worn. On the second day, 2 l

water samples were taken every three hours from 9 o'clock to midnight at the back end of

each aquarium (opposite to the inflow) using flexible tubes and 2 l wide neck bottles (Fig. 1).

The water samples were immediately filtered in an adjacent laboratory using glass microfiber

filters (1.2 µm pore width, 47 mm diameter, Whatman GF/C) and one negative control

consisting of 2 l MilliQ-water was included per sampling event. Thereafter, the filters were

individually placed in 2 ml reaction tubes and stored at -28 °C until further processing in a

special diagnostic molecular laboratory at the Department of Zoology, University of Innsbruck

(Austria). After each sampling, pH was measured in three arbitrarily selected fish tanks using

a Hach HQ40 device.

Activity measurement

During the familiarization time (day 1) and between water samplings, fish swimming activity

was quantified using a custom-made activity monitoring system consisting of one high-

definition USB camera (Ziggi HD Plus, IPEVO.COM, Great Britain) per aquarium. The

cameras were placed at the front of each tank and the focus was set towards the back end

(Fig. 1). To enable recordings during the night, aquaria were lighted throughout the two

recording days. Additionally, white polystyrene plates were used to cover the bottom and the

sides of each fish tank to exclude influences from neighboring aquaria and standardize

reflections. The signals from the cameras were acquired with a frame rate of 2 fpm ("frames

per minute") with a macro using the image analysis software FIJI (http://fiji.sc/, a distribution

of ImageJ) for MacOS (Schindelin et al., 2012; Rueden et al., 2017). For each aquarium, a

region of interest ("ROI") excluding the inflow, air-stone and sidewalls was set manually

(Supporting Information (SI) 1). Subsequent frames were arithmetically subtracted and the

average gray scale within the region of interest, as a quantification of fish activity, was

extracted from the difference-images. The dataset was manually checked to exclude artifacts

produced by changes in illumination (light/dark illumination of the fish), water sampling,

measurement of abiotic factors, fogged-up aquarium front and few other camera movements

sometimes leading to a changing region of interest in the recordings (SI 1).

Respirometry

A custom made intermittent-flow respirometer was used (Forstner, 1983; Svendsen et al.,

2016) including three measurement chambers placed in a larger tank (Fig. 1). The device

was cleaned prior to each fish change using a mixture of 3 % hydrogen peroxide and 3 l of

tap water. The volume of each chamber was determined prior to the experiment and oxygen

saturation (100 %) and temperature (8 – 9 °C) were kept constant in the tank via an airstone

and a laboratory heating/cooling device (Lauda DLK 10 and Alpha 1, Lauda Germany). The

three chambers of the respirometer were connected to the respirometers' water circuit,

constantly pumping O_2-saturated water from the large tank through the three chambers. For

measurements of oxygen consumption, a chamber was cut off from this circuit and a short,

closed loop for this chamber was established. Dissolved oxygen was measured in this

chamber every 30 s for a period of 15 min. using a YSI ProODO probe (YSI Inc. Yellow

Springs, USA) and logged to a computer before the system switched to the next chamber for

a 15 min measuring period. On the third day of an experimental run, three of the five fish

were placed individually in the chambers avoiding air bubbles and kept there for 24 h for

familiarization. On the fourth day, respirometer measurements were carried out for 24 h.

Thereafter, the remaining two fish individuals were placed in two measurement chambers for

one day of familiarization followed by one day of measurements (day five and six; Fig. 1).

The third chamber was left empty, but measured as well, to evaluate potential

microorganism-induced oxygen decrease. After the respirometer measurement day, the

mass [g] and total length [mm] of each fish was determined before placing them together in a

fish tank. For respirometer measurements of fish groups, the three individuals previously

sharing an aquarium, were put together in a respirometer chamber.

Filter processing and molecular analysis

After defrosting, each filter was soaked with 200 µl of lysis buffer consisting of TES-buffer

(0.1 M TRIS, 10 mM EDTA, 2 % sodium dodecyl sulphate; pH 8) and proteinase K (20 mg/m)

in a ratio of 19:1 and incubated at 56 °C over night in a rocking platform. On the next day,

filters were transferred with DNA-free forceps to a perforated inset which was repositioned in

the top half of the original 2 ml reaction tube and centrifuged for 10 min at 14 000 rpm.

Afterwards, filters were discarded and the lysate at the bottom of the reaction tube (300-

800 µl) used for DNA extraction. Insets were cleaned in sodium hypochlorite (2.5 %) for at

least 30 min, thoroughly washed with MilliQ-water (10 wash steps) and reused.

DNA extraction was carried out with the Biosprint 96 instrument (Qiagen) using the

Biosprint 96 DNA blood Kit (Qiagen) and the Biosprint 96 tissue extraction protocol in

accordance with the manufacturer's instructions except for using 100 µl of TE-buffer instead

of AE-buffer for DNA elution. Extractions were carried out in 96-well plates and four negative

controls (containing TES-buffer instead of lysate) were included per plate. To process the

whole lysate volume, a custom DNA-uptake program was set up: three uptake plates were

used and 300 µl of lysate, 300 µl AL-buffer and 300 µl isopropanol were mixed per well in

each plate. Missing lysate volumes (i.e. if only a total of 400 µl were available after

centrifugation) were replaced by TES-buffer. Additionally, 30 µl MagAttract was added per

well in the first plate. Using custom "binding" steps of the robotic platform, the DNA contained

in the first plate was transferred to the second one, next a binding step was carried out in the

second plate before transferring and releasing the entire collected DNA into the third plate,

which was then used for the Biosprint 96 tissue extraction protocol. After extraction, each

eluate was transferred to a 1.5 µl reaction tube for subsequent PCR.

All used primers (Table 1) have been previously published after extensive specificity

and sensitivity testing (Thalinger et al., 2016, 2020b) and additional specificity tests were

carried out on the digital PCR (dPCR) system (see below) confirming the specificity of the

molecular assays under the following conditions: each 22 µl dPCR master mix for droplet

generation on the QX200 AutoDG (Biorad) consisted of one-time EvaGreen Supermix

(Biorad), 0.25 µM forward and reverse primer (Table 1) and up to 10.5 µl DNA extract.

Depending on the results of initial tests with capillary electrophoresis PCR (i.e. the Relative

Fluorescence Units (RFU) of the resulting band; see SI 2), extracts were diluted with

molecular grade water for dPCR as follows: RFU < 0.2: undiluted; $0.2 \leq$ RFU < 1.3: 1:1

dilution; $1.3 \leq$ RFU < 2: 1:3 dilution; $2 \leq$ RFU: 1:7 dilution. Optimized thermo-cycling

conditions were 5 min at 95°C, 40 cycles of 30 s at 95°C, 1 min at 58°C (*O. mykiss*,

P. phoxinus, and *S. cephalus*) or 60°C (*C. gobio*, *S. fontinalis*, *S. trutta*, and *T. thymallus*),

1 min at 72°C, followed by one step of 5 min at 4°C and 5 min at 90°C. dPCR results were

analyzed on the QX200™ Droplet Reader and the corresponding QuantaSoftTM Analysis

Pro Software (Version 1.7; Biorad). As target signal amplitude varied with the length of the

amplified fragment, amplitude thresholds were set individually per primer pair (Table 1) prior

to determining target copy numbers per µl for each DNA extract. Per primer pair, a positive

(DNA extract from target species) and a negative control (molecular grade water) were

included in dPCR, all of which resulted positive and negative, respectively.

Statistical analysis

All calculations and visualizations were carried out in R Version 4.0.2 (R Core Team, 2020)

using the packages "ggplot2" (Wickham, 2016), "ggpubr" (Kassambara, 2019), "nlme"

(Pinheiro et al., 2020), "AICcmodavg" (Mazerolle, 2020), "rsq" (Zhang, 2020) and "sjPlot"

(Lüdecke, 2020). As pH was not measured in all aquaria after each water sampling, missing

values were estimated by averaging measurements taken at the respective fish tank before

and after the skipped time step. If measurements at the first or last water sampling were

missing, the values of the following or previous time step, respectively, were carried over.

The cleared activity dataset was visually inspected and summarized for each time

step: for example, data obtained during the preceding day were associated with the first

eDNA sampling event at 9 AM and measurements between 9 AM and 12 noon were

considered relevant for the second water sampling at 12 noon. Mean activity was calculated

per time interval. No cleared activity data was available for one *S. trutta* and *S. fontinalis*

individual, respectively, and for one *P. phoxinus* and *T. thymallus* individual at a single time

step each.

The total respirometry dataset was cleared of all 15 min measurement series showing

an increase in dissolved O_2. As this value is expected to decrease linearly over the course of

a measurement, a linear regression for the oxygen decrease in a measurement chamber

over time was calculated for each measurement series. All intervals for which the obtained

values showed a bad fit to a linear decrease ($R^2 < 0.8$) were also excluded from further

analyses. For each of the remaining measurement intervals, oxygen consumption (OC) in

mg / h was calculated as $OC = -s \times 60 \times vol$ where "s" denotes the slope of the linear

regression and "vol" the volume of the respective measurement chamber minus the mass of

the fish. Per fish species, the obtained value was corrected for the mean oxygen

consumption in the empty chamber before calculating total energy use (oxygen consumption

× 13.6 J/mg (oxycaloric factor (Brett and Groves, 1979)) per fish. Finally, energy use [J/h]

was averaged across the values obtained from individual measurement intervals for each

fish and fish group. Due to data clearing, this was not possible for one individual and one

group of *C. gobio* and *S. cephalus*, two individuals of *S. fontinalis* and *S. trutta* and three

individuals of *T. thymallus*. For these fish, energy use was estimated as the mean of the

available values.

Concerning the fish-eDNA copy numbers obtained from dPCR, 21 filtered water

samples did not lead to an amplification. They were removed from the dataset, as other fish

individuals of comparable size and other samplings reliably produced positive results and

hence, errors in sample processing might have occurred. One group of *P. phoxinus* had to

be excluded from further analyses, as two of three individuals were accidentally chosen

S. cephalus. To determine whether the pH measurements, mean activity and eDNA copy

numbers were significantly influenced by sampling (i.e. time of the day), a one-way repeated

measurements ANOVA with rank transformation was calculated for each variable using a

combination of fish species and aquarium as random factor. A significant trend could not be

detected (Table 2). Despite efforts to standardize the mass of the chosen fish individuals

within and between species, fish mass was identified as confounding variable (SI 3). Hence,

eDNA copies, mean activity, and energy use were all normalized by the mass of the

respective fish individual prior to all further analyses.

Generalized Linear Models (GLM) for a Gamma-distributed dependent variable (i.e.

eDNA copies; positive continuous variable) and a log-link function were set up to investigate

the effects of fish species, energy use, mean activity, and pH (Faraway, 2016). Data

obtained from fish groups were excluded from the comparison of model performance. Due to

the small number of tested fish individuals and species examined, no models with random

factors were considered, but sampling and fish individuals were included in several models to

show their ineffectiveness in explaining the obtained data (Table 3). The variable "fish

species" was entered via dummy coding into the models using "*C. gobio*" as base category.

Corresponding with the focus of this study to investigate the effect of species identity, fish

physiology and activity on eDNA shedding, a set of six candidate models was chosen (Table

3). AICc, ΔAICc, and AICc weights (ω) were used to evaluate the strength of the six models

for describing the data including Nagelkerke, Cragg & Uhler's pseudo-R^2 values (Burnham

and Anderson, 2002). Simulated, scaled residuals were calculated based on the best-

performing candidate model, (package: DHARMa (Florian Hartig, 2020); function:

"simulateResiduals"; n = 1000); the best performing model passed the consecutive check for

outliers and overdispersion.

To test the differences between single and grouped fish in the different stages of the

experiment, a data subset containing only values obtained from single and grouped

P. phoxinus and *S. cephalus* was analyzed. Target eDNA copies, energy use, and mean

activity (all normalized by fish mass) of the four distinct fish categories were tested for

normality and homogeneity of variance with Shapiro-Wilk and Bartlett tests. Then,

differences between groups were examined via Kruskal-Wallis tests followed by Wilcoxon

rank sum tests with Benjamini-Hochberg-corrected p-values. In a final step, target eDNA

copies for groups of *P. phoxinus* and *S. cephalus* were predicted using the model previously

established for single fish. Pairwise Wilcoxon tests were used to verify whether there was a

significant difference between predicted and measured target eDNA copy numbers for both

species separately and combined.

Results

The mean mass of individually housed fish was 3.06 g ± 1.56 g (SD) and *C. gobio* individuals

had the highest mass (5 g ± 2.1 g (SD)). Water samples from *P. phoxinus* and *T. thymallus*

aquaria had the highest eDNA copy numbers per µl extract and gram fish mass (31.13 ±

53.23 (SD) and 47.68 ± 41.13 (SD), respectively; Fig. 2). The normalized mean activity was

highest for *S. fontinalis* (1.08 ± 0.33 (SD)) and lowest for *C. gobio* (0.34 ± 0.10 (SD); Fig. 2).

The energy use per gram fish mass was highest for *O. mykiss* (1.81 J/h ± 0.91 J/h (SD)),

while *S. fontinalis* and *S. trutta* aquaria had the lowest pH.

The ΔAICc-based comparison of model weight (single fish only) resulted in model #3

outperforming five other candidate models (Table 3 and Table 4). Therein, mean activity,

energy use and fish species were contained as explanatory variables (pseudo-R^2_{NCU} = 0.60;

dispersion parameter = 1.02). Increased activity had a significantly positive effect on eDNA

copy numbers ($p < 0.01$) and *P. phoxinus*, *S. cephalus* and *T. thymallus* displayed

significantly higher copy numbers compared to *C. gobio* (base group), after controlling for the

effect of fish mass. The relationship between energy use and copy numbers was also found

to be positive, but not significant ($p = 0.05$; Table 5 and Fig. 3).

For single and grouped individuals of *P. phoxinus* and *S. cephalus*, significant

differences were detected between the four groups regarding target eDNA copies

(Chi² = 10.59; p < 0.05), mean activity (Chi² = 80.95; p < 0.001) and energy use

(Chi² = 36.77; p < 0.001): mean activity was significantly higher when fish were kept solitary

compared to having them in groups for both species (p < 0.001). Contrastingly, energy use

was significantly higher for grouped individuals in both species (p < 0.01). Target eDNA

copies were significantly higher for grouped *P. phoxinus* (42.61 ± 48.04 (SD)) compared to

single and grouped *S. cephalus* and characterized by few outliers with particularly high eDNA

concentration (Fig. 2 and Fig. 4)

To test the suitability of model #3 for describing eDNA shedding also for grouped fish,

model #3-predicted eDNA copies were compared to the measured copy numbers in the

group treatments. For the two species combined, there was no significant difference between

predicted and measured copy numbers (W = 0.98, p = 0.34). For *P. phoxinus* alone, no such

difference was detected either (W = 270; p = 0.72; Fig. 5), while predicted and measured

copy numbers of *S. cephalus* showed a significant difference (W = 614; p < 0.05; Fig. 5).

Discussion

This experiment confirms the hypothesized positive relationship between eDNA shedding,

fish activity and energy use. Additionally, species identity and thereby associated

physiological differences were found to influence the amount of released eDNA.

Furthermore, our data show that models of eDNA shedding cannot always be generalized

from experiments with individual fish to fish groups. For a conclusive habitat-scale estimation

of fish communities with eDNA-based methods it is therefore necessary to incorporate

species physiology and behavior into the analysis.

In early aquarium experiments, the strongest eDNA signals were found right after the

introduction of fish into tanks without water circulation and often explained by elevated stress

levels through handling and adaption to the new environment (Takahara et al., 2012; Klymus

et al., 2015; Maruyama et al., 2019). Hence, many recent studies allow for one or several

days of accommodation prior to eDNA sampling (Lacoursière-Roussel et al., 2016; Jo et al.,

2019; Takeuchi et al., 2019). We can confirm the positive relationship between fish activity

(i.e. movement) and eDNA shedding independent of the introductory phase of an

experiment. However, it was not possible to determine the actual reason for the elevated

eDNA levels associated with higher activity, as both higher metabolic rates during movement

and higher water volumes shearing against the fish body could be responsible for this effect.

For eDNA-based field studies this result indicates that signals emitted by highly active fish

(e.g. during spawning or predatory behavior) potentially mimic higher levels of fish biomass.

Energy use in a resting state as measured with an intermittent-flow respirometer, was

also positively correlated with eDNA shedding, albeit with a smaller effect size. Potentially,

this finding can be attributed to the higher metabolic rate and larger gill size of active species

in combination with higher water volumes pumped through them (Wegner et al., 2009).

However, the elevated eDNA signals could also stem from other physiological processes

(e.g. defecation), which are known to positively influence eDNA shedding rates (Klymus et

al., 2015). As fish were not fed during the entire experiment, the latter factor is potentially

negligible. Except for *T. thymallus* the energy use of the species preferring microhabitats with

strong currents and preying on fish as adults (primarily *O. mykiss* and *S. cephalus*) was

higher than for *C. gobio* and *P. phoxinus*. This is in concordance with general differences in

resting metabolic rates between these ecological guilds (Roberts, 1975; Johnston et al.,

1988; Killen et al., 2010).

There were distinct differences in eDNA shedding between the species, with

T. thymallus, *P. phoxinus* and *S. cephalus* emitting the most eDNA. The adaptation to

habitats with stronger currents (Freyhof and Kottelat, 2007), namely an increased mucus

production in combination with comparably large scales, might explain this result for

T. thymallus and *S. cephalus*. The underlying taxonomy could also contribute to this pattern if

cyprinids (*P. phoxinus* and *S. cephalus*) generally pass more DNA into the surrounding water

via their gills, feces or mucus. Another explanation for the high eDNA shedding of cyprinids

in this experiment could be the stress induced by solitary housing. The model estimating

eDNA concentrations for individual fish could not fully explain the findings obtained for

grouped fish: the activity of both *P. phoxinus* and *S. cephalus* was significantly lower when

fish were held in groups, while their energy use was significantly higher. However, the

opposing direction of these two effects potentially explains why differences between

predicted and measured eDNA quantities were only significant for *S. cephalus.*

Generally, the measured eDNA concentrations per μl DNA extract were right-skewed

and a few exceptionally high values showed considerable influence on the size of standard

deviations. These results were independent of fish handling and stress during the

introduction phase as eDNA sampling started only after 24 h and the aquaria had constant

flow with the entire volume being renewed every 11 min. Such "outliers" were also detected

in other aquarium experiments (Klymus et al., 2015; Wilcox et al., 2016) and cell-

conglomerates released into the surrounding water were previously deemed responsible for

this pattern (Wilcox et al., 2016). Additionally, the size distribution of eDNA particles (from

< 0.2 μm to >180 μm) and commonly detected fragment sizes suggest intact cells or

organelles as the primary source of eDNA in the water column (reviewed by Harrison et al.,

2019). Our data support the hypothesis of constant eDNA shedding rates at constant

environmental conditions. We could not observe any effects of sampling time, possibly due to

the constant illumination of the aquaria. Hence, this aspect is not necessarily transferable to

natural environments where fish are known to exhibit distinct diurnal movement patterns

(Helfman, 1986).

The influence of fish mass on eDNA concentrations was not in the focus of this

experiment and fish individuals were as similar in size/mass as possible. However, adult fish

of *P. phoxinus* and *C. gobio* are considerably smaller in comparison to the other species

(Freyhof and Kottelat, 2007) and the respective juveniles were thus in a later life stage.

Based on the ontogenetic decrease in metabolic rate (Winberg, 1960), this could be an

alternative explanation for the low energy use of these two species. For studies investigating

eDNA shedding directly from live animals, biomass will always be an influential and

potentially confounding variable and should thus be considered carefully already during

experimental design. Depending on the actual main source of eDNA (feces, mucus...) it is

furthermore questionable, if fish mass is the best index variable to describe study animals.

Fish length should also be considered as it is a good approximator for activity measurements

via videotaping, an estimator of fish surface, and at least for eels a good estimator of eDNA

shedding (Takeuchi et al., 2019). Since body shapes and fins differ a lot between taxa

(Freyhof and Kottelat, 2007), it could also be beneficial to factor body shape (fusiform,

elongated, flat,...) into future experiments in case the fish surface is the primary source of

eDNA. Finally, individual differences are well documented for fish behavior and metabolic

rates (Metcalfe et al., 2016). The number or study animals in future experiments should thus

be increased to better control for such effects within the same species.

Our results demonstrate that for the successful application of eDNA-based methods

on a habitat scale it is necessary to incorporate fish physiology and behavior not only in the

study design and sampling process (e.g. by sampling at different depths and in different

micro-habitats (Littlefair et al., 2020)), but also during data analysis (Barnes and Turner,

2016; Thalinger et al., 2020a). Seasonal patterns could have a much stronger effect on

eDNA concentrations in the water column as previously assumed: for instance, many

cyprinids in European freshwaters seek calm areas without current during the winter. Their

eDNA is less likely to spread through the water column and additionally, their decreased

activity and energy use lower the detection probability even further. In the future, the eDNA

shedding of manifold fish species and families in relation to their biomass, activity and energy

use should be investigated to deepen our understanding of taxon-specific effects. Until then,

estimations of fish biomass from eDNA quantities in field-collected samples should at least

take distinct physiology and behavior into account, especially for comparative analyses

between species or seasons.

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obtaining the *C. gobio* individuals.

471 Conflict of interest

472 MT is the co-founder of Sinsoma GmbH, a for profit company dedicated to DNA analyses in

473 environmental studies.

476 Author contribution statement

477 MT and JW, conceived the study; the experiment was designed by BT, MT, TS and JW, and

478 carried out by AR and AT under the supervision of BT and JW. TS was responsible for the

479 processing of activity data; AR, AT and YP were responsible for laboratory processing of the

480 eDNA samples under the supervision of BT who also carried out statistical analysis and

481 wrote the first draft of the manuscript which was revised by all co-authors.

484 Data Availability Statement

All data on eDNA signals, fish activity, energy use, fish mass and pH have been uploaded to

Figshare and are available at https://doi.org/10.6084/m9.figshare.13151180.v1

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Tables and Figures

Table 1: PCR assays used to amplify fish-eDNA: columns denote the target taxon of each primer combination, primer names, sequences, their respective concentration in PCR, target gene, amplicon sizes, and threshold value for positive droplets in dPCR. Additionally, the source column states whether a primer pair has been previously published. Please note that the *Salvelinus* sp. primer pair was designed to amplify both *S. fontinalis* and *Salvelinus umbla*.

Table 2: The results of one-way repeated measures ANOVA with rank transformation examining a potential effect of sampling on pH, mean activity and target eDNA copy numbers.

Table 3: Covariate structures of the candidate models compared for their potential to explain the target eDNA copy numbers per gram fish and µl extract obtained from single-fish aquaria with the following parameters: fish species identity (seven species), mean activity (per gram fish), energy use (per gram fish) and pH. Sampling event and fish individual number were included primarily to show their insignificance.

Table 4: Results of the ordinal ranking based on ΔAICc for the Generalized Linear Models (Table 3). Models are ordered from high to low weight and K denotes for the number of estimable parameters, AICc for the second-order variant of Akaike's Information Criterion, ΔAICc for AICc difference and ω for Akaike weight.

Table 5: The highest weight (ω = 0.90) Generalized Linear Model (#3) describing the measured eDNA copy numbers via mean activity, energy use and fish species identity. Significant p-values of fish species in the model refer to a significant difference between *Cottus gobio* (used as base category for dummy coding) and the respective fish species.

Figure 1: The setup of the aquarium experiment carried out with seven fish species: five individual fish were put in fish tanks for water sampling (eDNA) and activity recordings (day 1 and 2) followed by respirometer measurements (three individuals on days 3 and 4; two individuals plus empty control chamber on days 5 and 6). For *Phoxinus phoxinus* and *Squalius cephalus* the experiment was repeated using groups of three individuals.

Figure 2: Key parameters obtained during the experiment for single fish. Boxplots display target eDNA copies per µl extract, energy use [J/h], mean activity, and pH per fish species. Fish species are abbreviated: "Cot gob": *Cottus gobio*; "Onc myk": *Oncorhynchus mykiss*; "Pho pho": *Phoxinus phoxinus*; "Sal fon": *Salvelinus fontinalis*; "Sal tru": *Salmo trutta*; "Squ cep": *Squalius cephalus*; "Thy thy": *Thymallus thymallus*. The variables target eDNA copies, mean activity and energy use were normalized by fish mass to control for the effect of this confounding variable.

Figure 3: Graphic representation of the GLM model coefficients (model #3) best describing the obtained target eDNA copy numbers. Significance codes of denoted fish species indicate differences in comparison to the base category *Cottus gobio*, whiskers denote the 95%-CI.

Figure 4: Comparison of target eDNA copies, mean activity and energy use (normalized by
fish mass) in aquaria obtained from single and grouped individuals of *Phoxinus phoxinus* and
Squalius cephalus. Different lower case letters above boxplots code for significant
differences ($p < 0.05$) between categories, which are abbreviated as: "Pho pho": *Phoxinus*
phoxinus (single fish); "Pho pho g": *Phoxinus phoxinus* grouped fish; "Squ cep": *Squalius*
cephalus (single fish); "Squ cep g": *Squalius cephalus* grouped fish.
720
Figure 5: For groups of *Phoxinus phoxinus* (Pho pho g) and *Squalius cephalus* (Squ cep g)
measured and predicted copy numbers are plotted: left: against each other; middle: predicted
copy numbers are compared between species; right: comparison of measured copy numbers
between the two species. The measured copy numbers were log-transformed to enable a
direct comparison with the values predicted by the Gamma GLM with log-link function. For
S. cephalus a significant difference between measured and predicted copy numbers was
detected ($t = 2.37$; $p = 0.02$).

Table 1

target taxon	primer name	primer sequence (5' - 3')	primer conc. in PCR (μM)	target gene	fragment length (bp)	amplitude threshold (dPCR)	source
Cottus gobio	Cot-gob-S632	GAATAAAGGACTAAACCAAGTGGG	0.25	16S	118	13,500	Thalinger et al. 2016
	Cot-gob-A641	GCTGTAGCTCTCAGTTGTAGGAAAA	0.25				
Salmo trutta	Sal-tru-S1002	TCTCTTGATTCGGGCAGAACTC	0.25	COI	89	8,400	Thalinger et al. 2020
	Sal-tru-A1002	CGAAGGCATGGGCTGTAACA	0.25				
Oncorhynchus mykiss	Onc-myk-S655	TCTCCCTTCATTTAGCTGGAATC	0.25	COI	82	12,500	Thalinger et al. 2016
	Onc-myk-S655	GCTGGAGGTTTTATGTTAATAATGGTC	0.25				
Salvelinus spp.	Sal-vel-S651	ATAGTCGGCACCGCCCTT	0.25	COI	112	14,000	Thalinger et al. 2016
	Sal-vel-A651	TAACGAAGGCATGGGCTGTT	0.25				
Thymallus thymallus	Thy-thy-S653	ATCAAATTTATAATGTGATCGTCACG	0.25	COI	179	14,000	Thalinger et al. 2016
	Thy-thy-A653	AAGAAAGGACGGGGGGAAGC	0.25				
Phoxinus phoxinus	Pho-pho-S639	CGTGCAGAAGCGGATATAAATAC	0.25	16S	128	15,750	Thalinger et al. 2016
	Pho-pho-A648	CCAACCGAAGGTAAAGTCTTATTG	0.25				
Squalius cephalus	Squ-cep-S669	CAGTATACCCACCGCTTGCG	0.25	COI	130	14,250	Thalinger et al. 2016
	Squ-cep-A669	TTAATAATTGTGGTAATGAAGTTGACC	0.25				

Table 2

		F-value	p-value
pH	(intercept)	34.52	< 0.001
	sampling	2.22	0.053
mean activity	(intercept)	78.78	<0.001
	sampling	0.76	0.57
target copies per µl	(intercept)	39.00	<0.001
	sampling	0.38	0.86

Table 3

model #	covariate structures
1	mean activity + energy use + fish species + pH + sampling + fish individual
2	mean activity + energy use + fish species + sampling + fish individual
3	mean activity + energy use + fish species
4	mean activity + energy use
5	mean activity + fish species
6	fish species

Table 4

model #	K	AICc	ΔAICc	ω	pseudo-R^2_{NCU}
3	10	1208.32	0	0.90	0.60
5	9	1212.76	4.44	0.10	0.58
2	19	1221.36	13.04	0	0.62
6	8	1223.20	14.88	0	0.55
1	20	1223.64	15.32	0	0.62
4	4	1346.99	138.67	0	0.04

Table 5

predictor variable	parameter estimate	standard error	lower 95% CI	upper 95% CI	t-value	p-value
intercept	0.26	0.26	-0.23	0.77	0.97	0.33
mean activity	1.05	0.36	0.36	1.75	2.91	< 0.01**
energy use [J/h]	0.40	0.20	0.04	0.77	1.97	0.05
Oncorhynchus mykiss	0.72	0.37	0.06	1.41	1.96	0.05
Phoxinus phoxinus	2.33	0.29	1.76	2.91	7.90	< 0.001***
Salvelinus fontinalis	0.53	0.39	-0.23	1.31	1.34	0.18
Salmo trutta	0.28	0.31	-0.34	0.92	0.91	0.36
Squalius cephalus	1.16	0.32	0.54	1.80	3.65	< 0.001**
Thymallus thymallus	2.96	0.28	2.43	3.49	10.51	< 0.001***

Figure 1

eDNA sampling
day 1: 24 h familiarization
day 2: 2 l water sample every 3 h (6 ×)

2 l

activity measurement
picture of aquariums every 30s (day 1 & 2)

respirometry
24 h familiarization + 24 h measurement
day 3 & 4: fish 1, 2 & 3
day 5 & 6: fish 4 & 5 + empty reference chamber

Figure 2

Figure 3

Figure 4

Figure 5