The effect of activity, energy use, and species identity on environmental DNA shedding of freshwater fish

Bettina Thalinger[1,2,3*], Andreas Rieder[1], Anna Teuffenbach[1], Yannick Pütz[1], Thorsten Schwerte[1], Josef Wanzenböck[4] & Michael Traugott[1,5]

[1] *Department of Zoology, University of Innsbruck, Technikerstr. 25, 6020 Innsbruck, Austria*

[2] *Centre for Biodiversity Genomics, University of Guelph, 50 Stone Road E, N1G2W1, Guelph, Ontario, Canada*

[3] *Department of Integrative Biology, College of Biological Sciences, University of Guelph, 50 Stone Road E, N1G2W1, Guelph, Ontario, Canada*

[4] *Research Department for Limnology, University of Innsbruck, Mondseestr. 9, 5310 Mondsee, Austria*

[5] *Sinsoma GmbH, Lannes 6, 6176 Voels, Austria*

* Correspondence:

Bettina Thalinger

Bettina.Thalinger@gmail.com

Keywords: eDNA, aquarium experiment, respirometry, video-analysis, digital PCR, fish tank

Abstract

The quantitative measurement of environmental DNA (eDNA) from field-collected water

samples is gaining importance for the monitoring of fish communities and populations. The

interpretation of these signal strengths depends, among other factors, on the amount of

target eDNA shed into the water. However, shedding rates are presumably associated with

species-specific traits such as physiology and behavior. Although such differences between

juvenile and adult fish have been previously detected, the general impact of movement and

energy use in a resting state on eDNA release into the surrounding water remains hardly

addressed.

In an aquarium experiment, we compared eDNA shedding between seven fish

species occurring in European freshwaters. The investigated salmonids, cyprinids, and

sculpin exhibit distinct adaptions to microhabitats, diets, and either solitary or schooling

behavior. The fish were housed in aquaria with constant water flow and their activity was

measured by snapshots taken every 30 s. Water samples for eDNA analysis were taken

every 3 h and energy use was determined in an intermittent flow respirometer. After

controlling for the effect of fish mass, our results demonstrate a positive correlation between

target eDNA quantities as measured with digital PCR, fish activity, and energy use, as well

as species-specific differences. For cyprinids, the model based on data from individual fish

was only partly transferable to groups, which exhibited lower activity and higher energy use.

Our findings highlight the importance of fish physiology and behavior for the

comparative interpretation of taxon-specific eDNA quantities. Species traits should therefore

be incorporated into eDNA-based monitoring and conservation efforts.

1. Introduction

The sensitivity, non-invasiveness, and cost-efficiency of environmental DNA (eDNA) based

methods have been proven for diverse habitats and species making them powerful new tools

for conservation biology and biodiversity assessments (Barnes and Turner, 2016; Deiner et

al., 2017; Huerlimann et al., 2020). Regarding the detection of fish species, eDNA-based

monitoring outperforms traditional methods such as electrofishing: for example, for the

detection of the endangered European weather loach, *Misgurnus fossilis* (Sigsgaard et al.,

2015), the assessment of fish communities in Australian streams (McColl-Gausden et al.,

2020), and the distribution of brook trout, *Salvelinus fontinalis* in a US watershed (Evans et

al., 2017). The manifold successes of eDNA-based species detection lead to a call for more

standardization and better reporting practices (Goldberg et al., 2016; Minamoto et al., 2020;

Thalinger et al., 2020a) and to an international effort for implementing the technology into

routine species monitoring (Leese et al., 2016; Pilliod et al., 2019). Although reporting the

presence/absence of particular species is the starting point of these endeavors, a more

quantitative interpretation of field-derived eDNA data is key for the general application of this

technology.

Different processes influence the distribution of eDNA in space and time and the

detection probabilities of species from environmental samples, namely the origin,

degradation, suspension, resuspension, and transport of eDNA (Barnes and Turner, 2016;

Harrison et al., 2019). The latter processes are directly linked to local hydrology (e.g. flow

and substrate type (Shogren et al., 2017; Pont et al., 2018; Thalinger et al., 2020b)) and

environmental conditions (e.g. water temperature, pH, UV-radiation (Strickler et al., 2015;

Lacoursière-Roussel et al., 2016; Tsuji et al., 2017)). The amount of eDNA in the water

column is directly linked to fish biomass and originally, this was confirmed for common carp

(*Cyprinus carpio*) in an aquarium trial and experimental ponds (Takahara et al., 2012). In

subsequent experiments, the positive relationship was confirmed for a range of freshwater

and marine fish species (Evans et al., 2016; Lacoursière-Roussel et al., 2016; Sassoubre et

al., 2016; Horiuchi et al., 2019; Jo et al., 2020). However, these results were primarily

obtained for individuals at the same life stage.

eDNA is released into the environment in the form of mucus, feces, scales, and

gametes (Merkes et al., 2014; Barnes and Turner, 2016; Sassoubre et al., 2016; Bylemans

et al., 2017). Under natural conditions, differences in fish physiology, diet and behavior are

likely to affect this process and confound the interpretation of eDNA-based results from a

water body (Klymus et al., 2015). For perch and eel, Maruyama et al. (2014) and Takeuchi et

al. (2019), respectively, found lower mass-specific eDNA shedding rates for adults in

comparison to juveniles, which is likely caused by the scaling in metabolic rates, excretion

rates, and surface area with body mass (discussed in Yates et al., 2020). However, these

findings could not be confirmed in another experiment with a salmonid species (Mizumoto et

al., 2018). In general, the metabolic rate and activity differ between fish species due to

distinct physiology and behavior with pelagic species being more active and displaying

higher resting metabolic rates than benthic species (Johnston et al., 1988; Killen et al.,

2010). A stress response characterized by elevated metabolism and activity is frequently

hypothesized as underlying cause for spiking eDNA levels at the beginning of aquarium

experiments. Furthermore, metabolism and activity could generally explain mismatching

quantitative results in studies comparing eDNA levels between species in the same water

body (Takahara et al., 2012; Maruyama et al., 2014; Evans et al., 2016).

Here, we investigate the effect of fish activity (i.e. movement), energy use (i.e. oxygen

use × oxycaloric factor), and species identity in an aquarium experiment with seven fish

species commonly occurring in European rivers and streams (Fig. 1). We hypothesized that

higher activity (i.e. movement) leads to higher eDNA concentrations as there is more

shearing between the fish surface and the surrounding water, and higher volumes are

pumped through the gills due to the elevated oxygen demand. Independent of activity, fish

species with higher energy use in a resting state potentially also emit more eDNA.

Additionally, the species-specific composition of the constantly renewed cutaneous mucus

layer (Ángeles Esteban, 2012) might lead to differences between individual taxa.

2. Materials and methods

2.1. Study species

The examined species comprised four salmonids (*Salmo trutta, Salvelinus fontinalis,*

Oncorhynchus mykiss, Thymallus thymallus), two cyprinids (*Phoxinus phoxinus*, *Squalius*

cephalus), and one sculpin (*Cottus gobio;* Fig. 1). *Salmo trutta* is a rhithral species, territorial

especially in later life stages, and primarily feeds on benthic organisms and insect drift on the

surface. *Salvelinus fontinalis* and *O. mykiss* were anthropogenically introduced into

European freshwaters and are less territorial than *S. trutta* (Freyhof and Kottelat, 2007). If

possible, these three species choose areas with reduced current close to the main riverbed

as preferential microhabitat. *Thymallus thymallus* is also a rhithral species, but its scales are

larger and adults primarily use the main riverbed (Spindler, 1997; Freyhof and Kottelat,

2007)*. Phoxinus phoxinus* is a schooling, small fish species in the rhithral. It feeds on a

mixture of plant debris, algae, and small invertebrates. The juveniles prefer vegetation-rich

microhabitats without current, while adults switch to gravel substrate with low to intermediate

flow. *Squalius cephalus* is eurytopic (rhithral to potamal) and can occur in habitats with

strong to low current. Its juveniles are schooling and omnivorous with adults predominantly

preying on fish. *Cottus gobio* is a rheophilic and benthic species primarily feeding on small

bottom invertebrates. It has no swim bladder and mostly resides in interstices between large

boulders or on coarse gravel characterized by low current (Muus and Dahlström, 1968;

Spindler, 1997; Freyhof and Kottelat, 2007).

2.2.   Experimental setup

The aquarium experiment was carried out between 2nd March and 17th July 2017 at the

Research Department for Limnology Mondsee of the University of Innsbruck, Austria. The

juvenile salmonid individuals were purchased from commercial hatcheries, *P. phoxinus* and

S. cephalus were caught with permission in Lake Mondsee and *C. gobio* were caught with

permission in rivers in Tyrol (Austria). Fish individual sizes were chosen as similar as

possible within and between species. As *P. phoxinus* and *C. gobio* are smaller in comparison

to the other species (Fig. 1), these individuals were supposedly closer to reproductive

maturity. Until the start of the experiment, the fish species were kept separately in aquaria

fed with lake water.

In accordance with the regulations of the Austrian Animal Experiment Act (December

28, 2012) (Tierversuchsrechtsänderungsgesetz, part 1, section 1, §1, point 2), and with the

Directive 2010/63/EU of the European Parliament and of the Council of the European Union

(September 22, 2010) on the protection of animals used for scientific purposes (chapter 1,

article 1, point 5a), all fish were reared according to regular agriculture (aquaculture)

practice, including the provision of appropriate tank size, sufficient rate of waterflow, natural

photoperiod, *ad libitum* food supply, and temperatures within the species' thermal tolerance

range. This ensured that no pain, suffering, distress or lasting harm was inflicted on the

animals, confirmed by the fact that mortality rates were low and equal between rearing

groups. Based on the legislative provisions above, no ethics approval and no IACUC protocol

was required for the experiments performed. In particular the respirometry experiments were

discussed with the legislative authorities (Austrian Federal Ministry of Education, Science

and Research and the University of Veterinary Medicine, Vienna) and the conclusion was

that the assessment of basic metabolism under these conditions (small fish sizes in relatively

large chambers) does not incur pain, suffering or distress to the fish and no formal animal

experimentation protocol was required.

Five aquaria (60 l) and corresponding plastic lids were used in the experiment, each

of which was thoroughly cleaned with sodium hypochlorite (5 %) and then rinsed with tap

water (fish-DNA-free) prior to each experimental run (i.e. changing the fish under

investigation). The flow-through rate for the tap-water fed aquaria was set to 5.45 l/min to

mimic natural conditions and keep eDNA concentrations in the fish tanks constant based on

the results of previous test runs (Supplementary Material (SM) 1). The water temperature in

the aquaria was stabilized at 15 °C by centrally heating the inflowing water to this

temperature. Each tank was further equipped with an air-stone to ensure water mixing. At the

start of each experimental run, a water sample (negative control) was taken from one of the

aquaria and processed as described below. Then, five fish individuals per species were

selected aiming at similar size. Each fish was placed individually in an aquarium using DNA-

free fishnets (Fig. 2). For *P. phoxinus* and *S. cephalus*, the experiment was carried out twice:

once with individual fish, and once with groups of three fish per aquarium. The day before the

experiment and for its duration, the respective fish were not fed to avoid contamination by

fish feed and minimize the effects of defecation. Each run started with one day of

familiarization.

2.3.   Water sampling, filtration, and pH

All equipment used for this process was cleaned with sodium hypochlorite (5 %) and rinsed

with tap water prior to each use; DNA-free gloves were always worn. On the second day, 2 l

water samples were taken every three hours from 9:00 AM to 12:00 AM (six samples) at the

back end of each aquarium (opposite to the inflow) using flexible tubes and 2 l wide neck

bottles (Fig. 2). Due to the high flow rates (entire water volume replaced every 11 min), the

water level in each aquarium self-adjusted automatically after every sampling. The water

samples were immediately filtered in an adjacent laboratory using glass microfiber filters

(1.2 µm pore width, 47 mm diameter, Whatman GF/C) and one negative control (2 l MilliQ-

water) was included per sampling event. Thereafter, the filters were individually placed in

2 ml reaction tubes and stored at −28 °C until further processing in a special diagnostic

molecular laboratory at the Department of Zoology, University of Innsbruck (Austria). After

each sampling, pH was measured in three arbitrarily selected aquaria using a Hach HQ40

device.

2.4. Activity measurement

During the familiarization time (day 1) and between water samplings, fish swimming activity

was quantified using a custom-made activity monitoring system consisting of one high-

definition USB camera (Ziggi HD Plus, IPEVO.COM) per aquarium. The cameras were

placed at the front of each tank and the focus was set towards the back end (Fig. 2). To

enable recordings during the night, aquaria were lighted throughout the two recording days.

Additionally, white polystyrene plates were used to cover the bottom and the sides to exclude

influences from neighboring aquaria and standardize reflections. The signals from the

cameras were acquired with a frame rate of 2 frames per minute (fpm) with a macro using

the image analysis software FIJI (http://fiji.sc/, a distribution of ImageJ) for MacOS

(Schindelin et al., 2012; Rueden et al., 2017). For each aquarium, a region of interest (ROI)

excluding the inflow, air-stone and sidewalls was set manually (SM 2). Subsequent frames

were arithmetically subtracted and the average gray-scale within the region of interest, as a

quantification of fish activity, was extracted from the difference-images. The dataset was

manually checked to exclude artifacts produced by changes in illumination (light/dark

illumination of the fish), water sampling, measurement of abiotic factors, fogged-up aquarium

front and few other camera movements sometimes leading to a changing region of interest in

the recordings (SM 2).

2.5.   Respirometry

A custom-made intermittent-flow respirometer was used (Forstner, 1983; Svendsen et al.,

2016) including three measurement chambers placed in a larger tank (Fig. 2). The device

was cleaned prior to each fish change using a mixture of 3 % hydrogen peroxide and 3 l of

tap water. The volume of each chamber was determined prior to the experiment and oxygen

saturation (100 %) and temperature (8 – 9 °C) were kept constant in the tank via an airstone

and heating/cooling device (Lauda DLK 10 and Alpha 1, Lauda Germany). The three

chambers of the respirometer were connected to the respirometers' water circuit, constantly

pumping O_2-saturated water from the large tank through the three chambers. For

measurements of oxygen consumption, a chamber was cut off from this circuit and a short,

closed-loop for this chamber was established. Dissolved oxygen was measured in this

chamber every 30 s for a period of 15 min. using a YSI ProODO probe (YSI Inc.) and logged

to a computer before the system switched to the next chamber for a 15 min measuring

period. On the third day of an experimental run, three of the five fish were placed individually

into the chambers avoiding air bubbles and kept there for 24 h for familiarization. On the

fourth day, respirometer measurements were carried out for 24 h. Thereafter, the remaining

two fish individuals were placed in two measurement chambers for one day of familiarization

followed by one day of measurements (days five and six; Fig. 2). The third chamber was left

empty, but measured as well, to evaluate potential microorganism-induced oxygen decrease.

After the respirometer measurement day, the mass [g] and total length [mm] of each fish

were determined before placing them together in a fish tank. For respirometer measurements

of fish groups, the three individuals previously sharing an aquarium were put together in a

respirometer chamber.

2.6.   Filter processing and molecular analysis

After defrosting, each filter was soaked with 200 μl of lysis buffer consisting of TES-buffer

(0.1 M TRIS, 10 mM EDTA, 2 % sodium dodecyl sulphate; pH 8) and proteinase K (20

mg/ml) in a ratio of 19:1 and incubated at 56 °C over night in a rocking platform. On the next

day, filters were transferred with DNA-free forceps to a perforated inset which was

repositioned in the top half of the original 2 ml reaction tube and centrifuged for 10 min at

20 000 *g*. Afterwards, filters were discarded and the lysate at the bottom of the reaction tube

(300-800 μl) was used for DNA extraction. Insets were cleaned in sodium hypochlorite

(2.5 %) for at least 30 min, thoroughly washed with MilliQ-water (10 wash steps) and reused.

DNA extraction was carried out with the Biosprint 96 instrument (Qiagen) using the

Biosprint 96 DNA blood Kit (Qiagen) and the Biosprint 96 tissue extraction protocol following

the manufacturer's instructions except for using 100 μl of TE-buffer instead of AE-buffer for

DNA elution. Extractions were carried out in 96-well plates and four negative controls

(containing TES-buffer instead of lysate) were included per plate. To process the whole

lysate volume, a custom DNA-uptake program was set up: three uptake plates were used

and 300 μl of lysate, 300 μl AL-buffer and 300 μl isopropanol were mixed per well in each

plate. Missing lysate volumes (i.e. if only a total of 400 μl were available after centrifugation)

were replaced by TES-buffer. Additionally, 30 μl MagAttract was added per well in the first

plate. Using custom "binding" steps of the robotic platform, the DNA contained in the first

plate was transferred to the second one. Next, a binding step was carried out in the second

plate before transferring and releasing the entire collected DNA into the third plate, which

was then used for the Biosprint 96 tissue extraction protocol. After extraction, each eluate

was transferred to a 1.5 μl reaction tube for subsequent PCR.

All used primers (Table 1) have been previously published after extensive specificity

and sensitivity testing (Thalinger et al., 2016, 2020b) and additional specificity tests were

carried out on the digital PCR (dPCR) system (see below) confirming the specificity of the

molecular assays under the following conditions: each 22 µl dPCR master mix for droplet

generation on the QX200 AutoDG (Biorad) consisted of one-time EvaGreen Supermix

(Biorad), 0.25 µM forward and reverse primer (Table 1) and up to 10.5 µl DNA extract.

Depending on the results of initial tests with capillary electrophoresis PCR (i.e. the Relative

Fluorescence Units (RFU) of the resulting band; see SM 3), extracts were diluted with

molecular grade water for dPCR as follows: $RFU < 0.2$: undiluted; $0.2 \leq RFU < 1.3$: 1:1

dilution; $1.3 \leq RFU < 2$: 1:3 dilution; $2 \leq RFU$: 1:7 dilution. Optimized thermo-cycling

conditions were 5 min at 95°C, 40 cycles of 30 s at 95°C, 1 min at 58°C (*O. mykiss*,

P. phoxinus, and *S. cephalus*) or 60°C (*C. gobio*, *S. fontinalis*, *S. trutta*, and *T. thymallus*),

1 min at 72°C, followed by one step of 5 min at 4°C and 5 min at 90°C. dPCR results were

analyzed on the QX200™ Droplet Reader with the corresponding QuantaSoftTM Analysis

Pro Software (Version 1.7; Biorad). As target signal amplitude varied with the length of the

amplified fragment, amplitude thresholds were set individually per primer pair (Table 1) prior

to determining target copy numbers per µl for each DNA extract. Per primer pair, a positive

(DNA extract from target species) and a negative control (molecular grade water) were

included in dPCR, all of which resulted positive and negative, respectively. All filtration and

extraction controls resulted negative as well.

2.7.   Statistical analysis

All calculations and visualizations were carried out in R Version 4.0.2 (R Core Team, 2020)

using the packages "ggplot2" (Wickham, 2016), "gridExtra" (Auguie, 2017), "ggpubr"

(Kassambara, 2019), "lme4" (Bates et al., 2015), "AICcmodavg" (Mazerolle, 2020), "MuMIn"

(Barton, 2019), "rsq" (Zhang, 2020) and "sjPlot" (Lüdecke, 2020). As pH was not measured in

all aquaria after each water sampling, missing values were estimated by averaging

measurements taken at the respective fish tank before and after the skipped time step. If

measurements at the first or last water sampling were missing, the values of the following or

previous time step, respectively, were carried over.

The cleared activity dataset was visually inspected and summarized for each time

step: for example, data obtained during the preceding day were associated with the first

eDNA sampling event at 9:00 AM and measurements between 9:00 AM and 12:00 PM were

considered relevant for the second water sampling at 12:00 PM. Mean activity was

calculated per time interval. No cleared activity data was available for one *S. trutta* and

S. fontinalis individual, respectively, and for one *P. phoxinus* and *T. thymallus* individual at a

single time step each.

The total respirometry dataset was cleared of all 15 min measurement series showing

an increase in dissolved O_2. As this value is expected to decrease linearly over the course of

a measurement (Svendsen et al., 2016), a linear regression for the oxygen decrease in a

measurement chamber over time was calculated for each measurement series. All intervals

for which the obtained values showed an insufficient fit to a linear decrease ($R^2 < 0.8$) were

also excluded from further analyses. For each of the remaining measurement intervals,

oxygen consumption (OC) in mg / h was calculated as $OC = -s \times 60 \times vol$ where "s"

denotes the slope of the linear regression and "vol" the volume of the respective

measurement chamber minus the mass of the fish. Per fish species, the obtained value was

corrected for the mean oxygen consumption in the empty chamber before calculating total

energy use (oxygen consumption × 13.6 J/mg (oxycaloric factor (Brett and Groves, 1979))

per fish. Finally, energy use [J/h] was averaged across the values obtained from individual

measurement intervals for each fish and fish group. Due to data clearing, this was not

possible for one individual and one group of *C. gobio* and *S. cephalus*, two individuals of

S. fontinalis and *S. trutta* and three individuals of *T. thymallus*. For these fish, energy use

was estimated as the mean of the available values.

Concerning the fish eDNA copy numbers obtained from dPCR, 21 filtered water

samples did not lead to an amplification. They were removed from the dataset, as other fish

individuals of comparable size and other samplings reliably produced positive results and/or

eDNA was detected in celPCR. Hence, processing errors during sampling and in the

laboratory were deemed the most likely cause for failing amplification. One group of

P. phoxinus had to be excluded from further analyses, as two of three individuals were

identified as *S. cephalus* when removed from the aquarium after the experiment. To

determine whether the pH measurements, mean activity and eDNA copy numbers were

significantly influenced by sampling (i.e. time of the day), a one-way repeated measurements

ANOVA with rank transformation was calculated for each variable using a combination of fish

species and aquarium as random factor. A significant trend could not be detected (Table 2).

Despite efforts to standardize the mass of the chosen fish individuals within and between

species, fish mass was identified as confounding variable (SM 4). Hence, eDNA copies,

mean activity, and energy use were normalized by the mass of the respective fish individual

prior to all further analyses.

Generalized Linear Mixed-Effects Models (GLMM) for a Gamma-distributed

dependent variable (i.e. eDNA copies) were set up with a log-link function to investigate the

effects of mean activity, energy use, fish species, and pH (Faraway, 2016). Data obtained

from fish groups were excluded from the comparison of model performance. Fish individuals

were used as random intercept to account for repeated measurements and models were fit

using Gauss-Hermite quadrature (nAGQ = 20) and the BOBYQA algorithm (Bolker et al.,

2009; Powell, 2009). The variable "fish species" was entered via dummy coding into the

models using *C. gobio* as base category. Corresponding with the focus of this study to

investigate the effect of species identity, energy use, and activity on eDNA shedding, a set of

six candidate models was chosen (Table 3). AICc, ΔAICc, and AICc weights (ω) were used

to evaluate the strength of the six models for describing the data including marginal and

conditional pseudo-R^2 values (Burnham and Anderson, 2002; Nakagawa and Schielzeth,

2013). Simulated, scaled residuals were calculated based on the best-performing candidate

model, (package: DHARMa (Hartig, 2020); function: "simulateResiduals"; n = 1000). The best

performing model passed the consecutive check for outliers and overdispersion; the 95%CI

for its fixed effects were derived via bootstrapping (200 simulations).

To test the differences between single and grouped fish in the different stages of the

experiment, a data subset containing only values obtained from single and grouped

P. phoxinus and *S. cephalus* was analyzed. Target eDNA copies, energy use, and mean

activity (all normalized by fish mass) of the four distinct fish categories were tested for

normality and homogeneity of variance with Shapiro-Wilk and Bartlett tests. Then,

differences between groups were examined via Kruskal-Wallis tests followed by Wilcoxon

Rank Sum tests with Benjamini-Hochberg-corrected p-values. In a final step, target eDNA

copies for groups of *P. phoxinus* and *S. cephalus* were predicted using the model previously

established for single fish (only possible when not incorporating the random effect of fish

individual). Pairwise Wilcoxon tests were used to verify whether there was a significant

difference between predicted and measured target eDNA copy numbers for both species

separately and combined.

3. Results

The mean mass of individually housed fish was 3.06 g ± 1.56 g (SD) and *C. gobio* individuals

had the highest mass (5 g ± 2.1 g (SD); Table 4). Water samples from *P. phoxinus* and

T. thymallus aquaria had the highest eDNA copy numbers per µl extract and gram fish mass

(31.13 ± 53.23 (SD) and 47.68 ± 41.13 (SD), respectively; Fig. 3, Table 4). The normalized

mean activity was highest for *S. fontinalis* (1.08 ± 0.33 (SD)) and lowest for *C. gobio* (0.34 ±

0.10 (SD); Fig. 3, Table 4). The energy use per gram fish mass was highest for *O. mykiss*

(1.81 J/h ± 0.91 J/h (SD)), while *S. fontinalis* and *S. trutta* aquaria had the lowest pH.

The ΔAICc-based comparison of model weight (single fish only) resulted in model #3

outperforming five other candidate models (Table 3 and Table 5). Therein, mean activity,

energy use, and fish species were contained as explanatory variables (conditional pseudo-

$R^2 = 0.59$). Increased activity had a significantly positive effect on eDNA copy numbers

($p < 0.05$) and *P. phoxinus*, *S. cephalus*, and *T. thymallus* displayed significantly higher copy

numbers compared to *C. gobio* (base group) after controlling for the effect of fish mass. The

relationship between energy use and copy numbers was also positive, but not significant

($p = 0.08$; Table 6 and Fig. 4).

For single and grouped individuals of *P. phoxinus* and *S. cephalus*, target eDNA

copies per gram fish were significantly higher for grouped fish in general (28.96 ± 35.44 (SD)

compared to 22.44 ± 38.64 (SD); Chi² = 5.96; $p < 0.05$). Specifically, they were significantly

higher for grouped *P. phoxinus* (42.61 ± 48.04 (SD)) compared to single *P. phoxinus* and

single and grouped *S. cephalus* and characterized by few outliers with particularly high eDNA

concentration (Fig. 3 and Fig. 5). Significant differences were also detected between the four

groups regarding mean activity (Chi² = 80.95; $p < 0.001$) and energy use (Chi² = 36.77;

$p < 0.001$): mean activity was significantly higher when fish were kept solitary compared to

having them in groups for both species ($p < 0.001$). Contrastingly, energy use was

significantly higher for grouped individuals of *P. phoxinus* and *S. cephalus* ($p < 0.01$).

To test the suitability of model #3 for describing eDNA shedding also for grouped fish,

model #3-predicted eDNA copies were compared to the measured copy numbers in the

group treatments. For the two species combined, there was no significant difference between

predicted and measured copy numbers (W = 1612, $p = 0.35$). For *P. phoxinus* alone, no such

difference was detected either (W = 274; $p = 0.78$; Fig. 6), while predicted and measured

copy numbers of *S. cephalus* showed a significant difference (W = 609; $p < 0.05$; Fig. 6).

4. Discussion

This experiment confirms the hypothesized positive relationship between eDNA shedding

and fish activity. The species identity and thereby associated physiological differences were

found to influence the amount of released eDNA, and the positive relationship between

energy use and eDNA signals was not significant. Furthermore, our data show that models of

eDNA shedding cannot always be generalized from experiments with individual fish to fish

groups. For a conclusive habitat-scale estimation of fish communities with eDNA-based

methods it is therefore necessary to incorporate species physiology and behavior into the

analysis.

390   In early aquarium experiments, the strongest eDNA signals were found right after the

introduction of fish into tanks without water circulation and often explained by elevated stress

levels through handling and adaption to the new environment (Takahara et al., 2012;

Maruyama et al., 2014, 2019; Klymus et al., 2015). Hence, many recent studies allow for one

or several days of accommodation prior to eDNA sampling (Lacoursière-Roussel et al., 2016;

Jo et al., 2019; Takeuchi et al., 2019). We can confirm the positive relationship between fish

activity (i.e. movement) and eDNA shedding independent of the introductory phase of an

experiment. However, it was not possible to determine the actual reason for the elevated

eDNA levels associated with higher activity, as both higher metabolic rates during movement

and higher water volumes shearing against the fish body could be responsible for this effect.

For eDNA-based field studies, this result indicates that signals emitted by highly active fish

(e.g. during spawning or predatory behavior) potentially mimic higher levels of fish biomass.

Energy use in a resting state as measured with an intermittent-flow respirometer, was

also positively correlated with eDNA production, albeit not significant. In case this trend is

confirmed in the future, it could be attributed to the higher metabolic rate and larger gill size

of active species in combination with higher water volumes pumped through them (Wegner

et al., 2009). However, the elevated eDNA signals could also stem from other physiological

processes (e.g. defecation), which are known to positively influence eDNA production rates

(Klymus et al., 2015). As fish were not fed during the entire experiment, the latter factor is

potentially negligible, albeit it might substantially influence eDNA levels under natural

conditions. Except for *T. thymallus*, the energy use of the species preferring microhabitats

with strong currents and preying on fish as adults (primarily *O. mykiss* and *S. cephalus*) was

higher than for *C. gobio* and *P. phoxinus*. This is in concordance with general differences in

resting metabolic rates between these ecological guilds (Roberts, 1975; Johnston et al.,

1988; Killen et al., 2010). In this experiment, the smaller sized *S. fontinalis* and *S. trutta*

individuals, were more difficult to measure with the chosen respirometer setup (i.e. fewer

measurements passed our quality filtering), which could be the cause for the weak

relationship between energy use and eDNA copy numbers. In the future, more emphasis

should be placed on a ratio of 20-50 between the volume of the measurement chamber in

the respirometer and the fish individual to facilitate respirometer measurements (Svendsen et

al., 2016).

There were distinct differences in eDNA shedding between the species, with

T. thymallus, *P. phoxinus*, and *S. cephalus* emitting the most eDNA. The adaptation to

habitats with stronger currents (Freyhof and Kottelat, 2007), namely an increased mucus

production in combination with comparably large scales, might explain this result for

T. thymallus and *S. cephalus*. The underlying taxonomy could also contribute to this pattern if

cyprinids (*P. phoxinus* and *S. cephalus*) generally release more DNA into the surrounding

water via their gills, feces or mucus. Another explanation for the high eDNA shedding of

cyprinids in this experiment could be the stress induced by solitary housing. The model

estimating eDNA concentrations for individual fish could not fully explain the findings

obtained for grouped fish: the activity of both *P. phoxinus* and *S. cephalus* was significantly

lower when fish were held in groups, while their energy use was significantly higher. A

change in measurement precision regarding activity and energy use (respirometer more

precise, activity measurement less precise for fish groups) could explain these contradictory

results. Nevertheless, eDNA copies of grouped *P. phoxinus* individuals did not differ from

values predicted with a model based on single fish.

Generally, the measured eDNA concentrations per µl DNA extract were right-skewed

and a few exceptionally high values showed a considerable influence on the size of standard

deviations. These results were independent of fish handling and stress during the

introduction phase as eDNA sampling started only after 24 h and the aquaria had constant

flow with the entire volume being renewed every 11 min. Such "outliers" were also detected

in other aquarium experiments (Klymus et al., 2015; Wilcox et al., 2016) and cell-

conglomerates released into the surrounding water were previously deemed responsible for

this pattern (Wilcox et al., 2016). Additionally, the size distribution of eDNA particles (starting

from < 0.2 µm and exceeding >180 µm) and commonly detected fragment sizes suggest

intact cells or organelles as the primary source of eDNA in the water column (reviewed by

Harrison et al., 2019). Our data support the hypothesis of constant eDNA shedding rates at

constant environmental conditions indicating the potential to determine this variable for a

broad range of species and add to the interpretation of field sampling results. We could not

observe any effects of sampling time, possibly due to the constant illumination of the aquaria.

Hence, this aspect is not necessarily transferable to natural environments where fish are

known to exhibit distinct diurnal movement patterns (Helfman, 1986).

The influence of fish mass on eDNA concentrations was not in the focus of this

experiment and fish individuals were as similar in size/mass as possible. However, adult fish

of *P. phoxinus* and *C. gobio* are considerably smaller in comparison to the other species

(Freyhof and Kottelat, 2007) and the respective juveniles were thus closer to sexual maturity.

Hence, the allometric change of metabolic processes (Brown et al., 2004) could be an

alternative explanation for the comparably low energy use of these two species. For studies

investigating eDNA shedding directly from live animals, biomass will always be an influential

and potentially confounding variable and should thus be considered carefully already during

experimental design. Recently, the allometrically scaled mass was found to be the best index

variable for describing eDNA concentrations in lakes (Yates et al., 2020); since excretion

rate, metabolic rate and surface area all scale allometrically too (Brown et al., 2004; O'Shea

et al., 2006; Vanni and McIntyre, 2016), future experiments could greatly benefit from the

incorporation of this concept. For activity measurements via videotaping, fish length had to

be used as an index variable. In this context considerations of body shapes and fins, which

differ a lot between taxa (Freyhof and Kottelat, 2007), are also advisable. Finally, individual

differences are well documented for fish behavior and metabolic rates (Metcalfe et al., 2016).

The number or study animals in future experiments should thus be increased to better control

for such effects within a species.

Our results demonstrate that for the successful application of eDNA-based methods

on a habitat scale it is necessary to incorporate fish physiology and behavior not only in the

study design and sampling process (e.g. by sampling at different depths and in different

micro-habitats (Littlefair et al., 2020)), but also during data analysis (Barnes and Turner,

2016; Thalinger et al., 2020a). Seasonal patterns could have a much stronger effect on

eDNA concentrations in the water column as previously assumed: for instance, many

cyprinids in European freshwaters seek calm areas without current during the winter. Their

eDNA is less likely to spread through the water column and additionally, their decreased

activity lowers the detection probability even further. In the future, the eDNA shedding of

diverse fish species and families in relation to their biomass, activity, and energy use should

be investigated to deepen our understanding of taxon-specific effects. Until then, estimations

of fish biomass from eDNA quantities in field-collected samples should at least take distinct

physiology and behavior into account, especially for comparative analyses between species

or seasons.

5. Acknowledgements

This research was conducted within the eDNA-Alpfish project funded by the Austria

Research Promotion Agency (FFG); project number 853219. We thank R. Vogt for his

support during the experiment, M. Böcker for assistance with the background literature, J.

Harvie for input on the statistic analysis, and C. Moritz and D. Kirschner for their help in

obtaining the *C. gobio* individuals. We thank two anonymous reviewers for their extensive,

constructive feedback on the original manuscript. This manuscript has been uploaded as

preprint to bioRxiv https://doi.org/10.1101/2020.10.28.359653 (Thalinger et al., 2020c).

6. Conflict of interest

MT is the co-founder of Sinsoma GmbH, a for profit company dedicated to DNA analyses in

environmental studies.

7. Author contribution statement

MT and JW, conceived the study; the experiment was designed by BT, MT, TS and JW, and

carried out by AR and AT under the supervision of BT and JW. TS was responsible for the

processing of activity data; AR, AT and YP were responsible for laboratory processing of the

eDNA samples under the supervision of BT who also carried out statistical analysis and

wrote the first draft of the manuscript which was revised by all co-authors.

8. Data Availability Statement

All data on eDNA signals, fish activity, energy use, fish mass and pH have been uploaded to

Figshare and are available at https://doi.org/10.6084/m9.figshare.13151180.v1

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10. Tables and Figures

Table 1: digital PCR assays used to amplify fish eDNA

target taxon	primer name	primer sequence (5' - 3')	primer conc. in dPCR (μM)	target gene	fragment length (bp)	amplitude threshold (dPCR)	source
Cottus gobio	Cot-gob-S632	GAATAAAGGACTAAACCAAGTGGG	0.25	16S	118	13,500	Thalinger et al. 2016
	Cot-gob-A641	GCTGTAGCTCTCAGTTGTAGGAAAA	0.25				
Salmo trutta	Sal-tru-S1002	TCTCTTGATTCGGGCAGAACTC	0.25	COI	89	8,400	Thalinger et al. 2020
	Sal-tru-A1002	CGAAGGCATGGGCTGTAACA	0.25				
Oncorhynchus mykiss	Onc-myk-S655	TCTCCCTTCATTTAGCTGGAATC	0.25	COI	82	12,500	Thalinger et al. 2016
	Onc-myk-S655	GCTGGAGGTTTTATGTTAATAATGGTC	0.25				
Salvelinus spp.	Sal-vel-S651	ATAGTCGGCACCGCCCTT	0.25	COI	112	14,000	Thalinger et al. 2016
	Sal-vel-A651	TAACGAAGGCATGGGCTGTT	0.25				
Thymallus thymallus	Thy-thy-S653	ATCAAATTTATAATGTGATCGTCACG	0.25	COI	179	14,000	Thalinger et al. 2016
	Thy-thy-A653	AAGAAAGGACGGGGGAAGC	0.25				
Phoxinus phoxinus	Pho-pho-S639	CGTGCAGAAGCGGATATAAATAC	0.25	16S	128	15,750	Thalinger et al. 2016
	Pho-pho-A648	CCAACCGAAGGTAAAGTCTTATTG	0.25				
Squalius cephalus	Squ-cep-S669	CAGTATACCCACCGCTTGCG	0.25	COI	130	14,250	Thalinger et al. 2016
	Squ-cep-A669	TTAATAATTGTGGTAATGAAGTTGACC	0.25				

715 Columns denote the target taxon of each primer combination, primer names, sequences, their respective concentration in dPCR, target gene,

716 amplicon sizes, and threshold value for positive droplets in dPCR. Additionally, the source column shows the original publication. Please note that

717 the *Salvelinus* spp. primer pair was designed to amplify both *S. fontinalis* and *Salvelinus umbla*.

718 **Table 2:** The results of one-way repeated measurements ANOVA with rank transformation

719 examining a potential effect of sampling on pH, mean activity, and target eDNA copy numbers.

		F-value	p-value
pH	(intercept)	34.52	< 0.001
	sampling	2.22	0.053
mean activity	(intercept)	78.78	<0.001
	sampling	0.76	0.57
target copies per µl	(intercept)	39.00	<0.001
	sampling	0.38	0.86

723 **Table 3:** Covariate structures of the candidate Gamma GLMM with a log link function.

model #	covariate structure (fixed effects)
1	mean activity + energy use + fish species + pH + sampling
2	mean activity + energy use + fish species + sampling
3	mean activity + energy use + fish species
4	mean activity + energy use
5	mean activity + fish species
6	fish species

The models were compared for their potential to explain the target eDNA copy numbers per

gram fish and μl extract obtained from single-fish aquaria with the following parameters: fish

species identity (seven species), mean activity (per gram fish), energy use (per gram fish) and

pH. In all models, individual fish were included as random effect (random intercept) to account

for repeated measurements and the sampling event was included primarily to show its

insignificance.

Table 4: The means and standard deviations of mass, activity, and energy use for each fish

species in the experiment.

	mass [g] ± SD [g]	activity ± SD	energy use [J/h] ± SD [J/h]	eDNA copies ± SD
Cottus gobio	5.00 ± 2.10	0.34 ± 0.10	0.68 ± 0.17	2.48 ± 2.94
Oncorhynchus mykiss	3.30 ± 0.31	0.51 ± 0.10	1.81 ± 0.91	9.60 ± 5.16
Phoxinus phoxinus	3.48 ± 0.55	0.55 ± 0.09	0.67 ± 0.16	31.13 ± 53.23
Salvelinus fontinalis	1.40 ± 0.42	1.08 ± 0.33	1.06 ± 0.00	11.46 ± 8.72
Salmo trutta	2.22 ± 0.77	0.68 ± 0.29	0.92 ± 0.00	5.98 ± 8.08
Squalius cephalus	2.72 ± 0.66	0.69 ± 0.09	1.06 ± 0.44	14.14 ± 11.56
Thymallus thymallus	3.53 ± 1.42	0.56 ± 0.30	0.43 ± 0.01	47.68 ± 41.13
Phoxinus phoxinus grouped	12.06 ± 1.86	0.26 ± 0.09	1.00 ± 0.33	42.61 ± 48.04
Squalius cephalus grouped	7.75 ± 1.30	0.32 ± 0.11	1.41 ± 0.50	18.04 ± 13.69

736 The eDNA copies (per µl extract), activity, and energy use are provided per gram fish mass; for

737 grouped fish, the mass is displayed per aquarium (i.e. sum of three fish individuals).

738   **Table 5:** Results of the ordinal ranking based on ΔAICc for the GLMM (Table 3).

model #	K	AICc	ΔAICc	ω	marginal pseudo-R²	conditional pseudo-R²
3	11	159.57	0	0.52	0.56	0.59
5	10	160.28	0.71	0.36	0.56	0.60
6	9	162.74	3.17	0.11	0.53	0.60
2	16	167.10	7.53	0.01	0.57	0.60
1	17	169.55	9.98	0.00	0.57	0.60
4	5	194.18	34.6	0.00	0.04	0.52

Models are sorted from high to low weight and K denotes for the number of estimable

parameters, AICc for the second-order variant of Akaike's Information Criterion, ΔAICc for AICc

difference, ω for Akaike weight, and marginal/conditional pseudo-R² represent the variance

explained by the fixed effects only and by the entire model, respectively.

Table 6: The highest weight (ω = 0.52) GLMM (model #3) describing the measured eDNA copy

numbers via A) the fixed effects: mean activity, energy use and fish species identity, and B) the

random effect fish individual (31 groups, σ = 0.88).

A)	parameter estimate	standard error	lower 95% CI	upper 95% CI	t-value	p-value
intercept	0.19	0.31	-0.55	0.85	0.63	0.53
mean activity	1.00	0.42	0.19	1.95	2.39	0.02*
energy use [J/h]	0.41	0.23	-0.07	0.94	1.77	0.08
Oncorhynchus mykiss	0.79	0.43	-0.21	1.57	1.85	0.06
Phoxinus phoxinus	2.40	0.35	1.71	3.17	6.78	< 0.001***
Salvelinus fontinalis	0.65	0.47	-0.45	1.67	1.38	0.17
Salmo trutta	0.32	0.38	-0.58	1.28	0.83	0.40
Squalius cephalus	1.25	0.38	0.41	2.01	3.28	< 0.01**
Thymallus thymallus	2.93	0.33	2.30	3.58	8.75	< 0.001***

B)	variance	standard deviation
fish individual (intercept)	0.06	0.24

Significant p-values of fish species in the model refer to a significant difference between *Cottus*

gobio (used as base category for dummy coding) and the respective fish species.

Figure 1: A summary of the morphological and ecological traits of the fish species used in the

aquarium experiment. The provided information describes the situation in Central European

freshwaters and is not necessarily transferable to other geographic regions. Depending on the

source, different maximum fish length measurements were available with "TL" = "total length"

measured from the tip of the snout to the longest tip of the caudal fin and "SL" = "standard

length" measured from the tip of the snout to the base of the caudal fin. (Muus and Dahlström,

1968; Spindler, 1997; Freyhof and Kottelat, 2007)

(https://upload.wikimedia.org/wikipedia/commons/b/bb/CottusGobioSpreadingFins.JPG

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https://upload.wikimedia.org/wikipedia/commons/9/99/Thymallus_thymallus2.jpg separated from

background; Gilles San Martin, CC BY-SA 2.0 https://creativecommons.org/licenses/by-sa/2.0,

via Wikimedia Commons)

Figure 2: The setup of the aquarium experiment carried out with seven fish species: five

individual fish were put in fish tanks for water sampling (eDNA) and activity recordings (day 1

and 2) followed by respirometer measurements (three individuals on days 3 and 4; two

individuals plus empty control chamber on days 5 and 6). For *Phoxinus phoxinus* and *Squalius*

cephalus the experiment was repeated using groups of three individuals per tank and

respirometer chamber.

Figure 3: Key parameters obtained during the experiment for single fish. Boxplots display target

eDNA copies per µl extract, energy use [J/h], mean activity, and pH per fish species. Fish

species are abbreviated: "Cot gob": *Cottus gobio*; "Onc myk": *Oncorhynchus mykiss*; "Pho pho":

Phoxinus phoxinus; "Sal fon": *Salvelinus fontinalis*; "Sal tru": *Salmo trutta*; "Squ cep": *Squalius*

cephalus; "Thy thy": *Thymallus thymallus*. The variables target eDNA copies, mean activity, and

energy use were normalized by fish mass to control for the effect of this confounding variable.

Figure 4: Graphic representation of the GLMM estimates (model #3) best describing the

obtained target eDNA copy numbers: A) fixed effects and B) random effect of individual fish.

Coefficients are exponentiated, significance codes of denoted fish species indicate differences

in comparison to the base category *Cottus gobio*, whiskers display the 95%-CI. Fish species are

abbreviated: "Cot gob": *Cottus gobio*; "Onc myk": *Oncorhynchus mykiss*; "Pho pho": *Phoxinus*

phoxinus; "Sal fon": *Salvelinus fontinalis*; "Sal tru": *Salmo trutta*; "Squ cep": *Squalius cephalus*;

"Thy thy": *Thymallus thymallus* in addition to individual numbers from 1 to 5.

Figure 5: Comparison of target eDNA copies, mean activity, and energy use (normalized by fish

mass) in aquaria obtained from single and grouped individuals of *Phoxinus phoxinus* and

Squalius cephalus. Different lower case letters above boxplots code for significant differences

($p < 0.05$) between categories, which are abbreviated as: "Pho pho": *Phoxinus phoxinus* (single

fish); "Pho pho g": *Phoxinus phoxinus* grouped fish; "Squ cep": *Squalius cephalus* (single fish);

"Squ cep g": *Squalius cephalus* grouped fish.

Figure 6: For groups of *Phoxinus phoxinus* (Pho pho g) and *Squalius cephalus* (Squ cep g)

measured and predicted copy numbers are plotted: left: against each other; middle: predicted

copy numbers are compared between species; right: comparison of measured copy numbers

between the two species. The measured copy numbers were log-transformed to enable a direct

comparison with the values predicted by the Gamma GLMM with log-link function; the random

effect of individual fish could not be taken into account for this prediction. For *S. cephalus* a

significant difference between measured and predicted copy numbers was detected (W = 609;

$p < 0.05$).

	Salmo trutta	*Salvelinus fontinalis*	*Oncorhynchus mykiss*	*Thymallus thymallus*	*Phoxinus phoxinus*	*Squalius cephalus*	*Cottus gobio*
family	Salmonidae	Salmonidae	Salmonidae	Salmonidae	Cyprinidae	Cyprinidae	Cottidae
length	resident stream individuals up to 300 mm SL	in European freshwaters up to 500 mm SL	up to 1000 mm SL	up to 600 mm TL	up to 140 mm TL	up to 600 mm SL	up to 180 mm TL
status and association with river zonation	native; rhithral zone	introduced	introduced; frequently stocked	native; rhithral zone	native; rhithral zone	native; rhithral to potamal zone	native
distinct morphological features	small scales	small scales	small scales	large scales	small scales	large scales	no swim bladder and no true scales
feeding	aquatic invertebrates, insect drift, and small fish	aquatic invertebrates, insect drift, and small fish	aquatic invertebrates, insect drift, and small fish	predominantly drifting insects, small worms and crustaceans	plant debris, algae, and small invertebrates	juveniles: omnivorous; adults: primarily piscivorous	small bottom invertebrates
preferred microhabitat	locations with reduced current close to the main riverbed	locations with reduced current close to the main riverbed	locations with reduced current close to the main riverbed	main river bed with potentially strong current	juveniles: vegetation-rich locations without current; adults: gravel substrate with low to intermediate current	eurytopic; can occur in habitats with strong or low current	benthic: resides in interstices between large boulders or on coarse gravel characterized by low current
social behavior	territorial (already as juveniles)	less territorial and at equal size dominated by *S. trutta*	less territorial than *S. trutta*	schooling	schooling	juveniles: schooling; adults: solitary or in small groups	solitary and territorial

eDNA sampling
day 1: 24 h familiarization
day 2: 2 l water sample every 3 h (6 ×)

2 l

activity measurement
picture of aquaria every 30s (day 1 & 2)

respirometry
24 h familiarization + 24 h measurement
day 3 & 4: fish 1, 2 & 3
day 5 & 6: fish 4 & 5 + empty reference chamber