- 1 Title: Multidimensionality in the thermal niches of dung beetles could limit species'
- 2 responses to temperature changes
- 3 **Running Title:** Thermal niche multidimensionality
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19 Abstract

20 Understanding the consequences of climate change requires understanding how temperature controls species' responses across key biological aspects, as well as the 21 22 coordination of thermal responses across these aspects. We study the role of temperature in determining the species' diel, seasonal, and geographical occurrence, using dung 23 beetles as a model system. We found that temperature has relatively low -but not 24 25 negligible- effects in the three spatiotemporal scales, once accounting for alternative 26 factors. More importantly, the estimated thermal responses were largely incongruent 27 across scales. This shows that species have multidimensional thermal niches, entailing 28 that adjustments to fulfil temperature requirements for one biological aspect, such as 29 seasonal ontogenetic cycles, may result in detrimental effects on other aspects, like diel 30 activity. These trade-offs can expose individuals to inadequate temperatures, reducing populations' performance. Paradoxically, the relatively weak effects of temperature we 31 32 found may have serious consequences for species' responses to warming if temperature 33 regulates essential aspects of species' biology in divergent ways.

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35 **Keywords:** biological scale, daily activity, geographic distribution, niche dimensionality,

36 phenology, physiological trade-offs

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38 Introduction

Temperature is fundamental for the efficient capture and management of the energy that 39 maintains living organisms (Brown et al. 2004). Temperature variations affect the 40 41 abundance and distribution of species (Angilletta 2009), the variability of ecological 42 systems (Wang et al. 2009), and even the history of life and biodiversity on Earth itself 43 (Schwartzman 1999, Mayhew et al. 2008). Indeed, temperature plays a critical role in controlling key aspects such as species' spatiotemporal distribution, physiological activity 44 45 or individual growth rates (Somero 2005, Thackeray et al. 2016, Scranton & Amarasekare 2017, Madrigal-González et al. 2018), among many other things. Here, the 46 47 effects of temperature on species' geographic distributions and seasonal and diel activities are of particular interest since variation in these aspects can have dramatic consequences 48 for their ecological performance and persistence (Edwards & Richardson 2004, 49 50 Schweiger et al. 2008, Rader et al. 2013). The ongoing climate changes are drastically 51 modifying the spatial and temporal organization of biodiversity (Chapin III & Diaz 52 2020), which is leading to spatial and seasonal decouples of interacting species (Sheldon 53 et al. 2011) and, thus, to the disruption of food webs and ecosystem services (Román-Palacios & Wiens 2020). Ecologists and climatologists have accumulated a large amount 54 of evidence on these effects during recent decades, which are especially relevant for 55 ectotherms (Paaijmans et al. 2013). Despite this evidence, how temperature responses 56 57 integrate across different species' aspects is largely unknown. To obtain this knowledge 58 is crucial because incongruous responses can lead to incompatible adjustments to temperature changes along biological aspects, compromising the species' performances 59 60 under climate change.

Delimiting the actual effect of environmental temperature on the distribution and 61 62 abundance of species may become difficult when other variables that are either spatially 63 or temporally correlated with temperature are considered simultaneously. For instance, the latitudinal distribution of species in the Northern Hemisphere is associated with 64 65 historical events and dispersal limitations, whose effects generate geographical patterns 66 that can be confounded with those of temperature variations (Araújo et al. 2008, Hortal et al. 2011, Calatayud et al. 2016, 2019). Similarly, the apparent relationships between 67 68 temperature and either seasonal or diel activities may be indeed conditioned by life-

history constraints related with the time required to complete individual development, species' voltinism, the phase in which overwintering occurs, photoperiod limitations, light requirements, and the reliance on solar radiation independently on the environmental temperature (Bradshaw & Holzapfel 2007, 2010, Teder 2020). Hence, assessing the predictive value of temperature in accounting for the spatial and temporal variations in species occurrence and abundance would require considering any alternative variables that could play a significant role in these variations.

76 Experimental setups can help unravel the "true" role of temperature in driving 77 geographical, seasonal and diel patterns for some model organisms while controlling for 78 other variables (Angilletta 2009). However, experiments based on artificial thermal 79 gradients can subject individuals to new and unrealistic stress conditions, thereby providing overestimated projections of species responses (Guo et al. 2020). Alternatively, 80 81 one could explore the contribution of temperature using observational data where the 82 variations in temperature and other complementary predictors are decoupled. For 83 example, the effects of temperature and solar radiation can be teased apart using diel 84 activity from consecutive days that showed substantial variations in temperature (i.e., 85 while presenting almost equal sunlight incomings). Similarly, the effects of temperature 86 and day length can be teased apart using seasonal data along steep temperature gradients, 87 with nearly equal day lengths (such as e.g., elevational gradients). Finally, the role of temperature in determining the species' distribution can be assessed by comparing 88 89 geographical areas with different temperature regimens. That is, if temperature is an important variable, we should find similar responses under different background 90 91 temperatures.

92 The relevance of temperature in accounting for the spatiotemporal variation in 93 species occurrence and abundance may thus be estimated from observational data, 94 comparing the results from including or not alternative predictors to account for 95 complementary causal factors. Temperature will stand out as a relevant factor across different biological scales if its association with several species' responses is high 96 97 throughout different spatiotemporal dimensions, but also if such responses are congruent across dimensions. The congruence in thermal responses to diel, seasonal and 98 99 geographical gradients would support the universal and homogeneous role of temperature 100 in delimiting the occurrence and abundance of species. Note here that expectations are 101 that different mechanisms are behind the response to temperature variations associated 102 with geography, seasonality and diel rhythms. For instance, daily temperature variations

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should also be related to changes in light or other environmental factors that can generate 103 behavioural, endocrine, and physiological diel rhythms (Levy et al. 2019). In contrast, 104 responses to seasonal temperatures should be associated with the annual rhythms and the 105 106 need to synchronize life history phases with seasonal variations in climate (Saunders 107 2020). On the other hand, responses to geographical variations in temperature should 108 relate to local adaptation processes acting at the population level, and likely involving the 109 above-mentioned individual tolerances and ontogenetic timing, as well as other essential 110 species attributes (Sunday et al. 2019).

111 Despite these differences, a certain level of congruence in the responses would indicate the consistent role of temperature as a holistic and predictable driver of key 112 biological aspects. Such congruence would be evident, for example, if species occurring 113 in colder regions are also active during colder periods of the year and at colder hours of 114 115 the day in areas of milder climate. Such hypothesized thermal congruence is fundamental to respond adequately to global warming, as decoupling responses across different 116 spatiotemporal gradients may expose local populations to critical temperatures, thus 117 118 compromising their long-term persistence. For instance, if seasonal and diel responses to 119 temperature are decoupled, species might not be able to adjust seasonal cycles as much as 120 it would be necessary to prevent individuals from facing critical temperatures during diel 121 activities. Following this line of evidence, studying the congruence of thermal responses across evolutionary lineages is also important because a marked phylogenetic signal in 122 thermal niches would also point to the relevance of temperature changes. That is, if 123 thermal adaptations are evolutionarily conserved, species might present limited ability to 124 modify their thermal responses, being unable to cope with climate warming. Should this 125 126 be true, phylogenetic biases in the potential effects of climate warming would be also 127 expectable. Despite the relevance of studying the consistency of thermal responses across 128 spatiotemporal gradients and evolutionary lineages, integrative studies on this topic are lacking. 129

Here we study the thermal responses associated with geographical, seasonal, and diel temperature variations using several temperate dung beetle species as a model system. Dung beetles are capable to self-regulate their body temperature and produce heat depending on their body size (Verdú & Lobo 2008, Verdú et al. 2012) a physiological adaptation directly linked to the need of a quick dispersal response to exploit an ephemeral resource. In addition, they feed on cattle from domestic and wild animals, participating in nutrient cycling and seed dispersion (Nervo et al. 2017, Milotić et al 2019), and thus, providing important ecosystem functions. These characteristics makedung beetles an ideal and important group to study thermal responses.

Specifically, we evaluated the responses of dung beetles to changes in 139 140 temperature associated to: (i) diel rhythms across three consecutive days with contrasted 141 temperatures; (ii) seasonal rhythms across six sites located at different elevations; and 142 (iii) geographical ranges along five river basins in the Iberian Peninsula (Fig. 1). We 143 hypothesized that if temperature is the main factor determining the activity and 144 distribution of dung beetles, its effect should be observed along the three considered 145 spatiotemporal gradients, and its relevance would be higher if the effects of other alternative and/or complementary factors are low. Furthermore, congruence in the 146 different species' thermal responses to diel, seasonal and geographical changes would be 147 expected if the importance of temperature is independent of the spatiotemporal context. 148 149 On the contrary, a low explanatory capacity of temperature and a lack of congruence in its effects across the three spatiotemporal gradients would support a limited and 150 dissimilar role of temperature depending on the spatiotemporal context. Finally, if species 151 152 are evolutionarily limited to adapt to new thermal regimens, we expect thermal niches to 153 be phylogenetically conserved.

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155 Material and methods

156 Data origin

We use data on 16 Iberian dung beetle species of the family Scarabaeidae (ten from 157 Aphodiinae and six from Scarabaeinae subfamilies). These species were selected because 158 they occurred in at least 10% of the samples of the three datasets considered, covering 159 different spatial and temporal extents (see below). All considered species (Table 1) are of 160 161 small body size, with body weights far smaller than 1.9 g (0.2 g at most), the threshold from which endothermy is thought to appear in this group of beetles (Verdú et al. 2006). 162 Temperature-occurrence associations for all these species were examined along: (i) five 163 164 geographical areas of similar extent but different temperature regimes within the Iberian 165 Peninsula (geographical dataset or GD); (ii) six sites placed across a steep elevational 166 range in Central Iberia, and sampled during the same dates but differing in their environmental temperatures (seasonal dataset or SD); and (iii) three consecutive days 167 with similar daily variations but different weather conditions in a single locality near the 168 169 centre of the same elevational range (diel dataset or DD).

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Geographical Dataset. The GD is divided in five study areas, corresponding to the major 171 172 river basins of the Iberian Peninsula (Ebro, Duero, Tajo, Guadiana and Guadalquivir; limits extracted from HydroBASINS data available at www.hydrosheds.org, Lehner & 173 Grill 2013, Fig. 1a). These natural areas were used since their borders correspond with 174 marked geographical accidents, which are expected to act as dispersal barriers. 175 Furthermore, they are relatively similar in extent (areas ranging from 5.6 x 10^4 to 9.7 x 176 10^4 km²) and almost follow a latitudinal gradient, hence showing contrasting 177 environmental temperatures (Fig. 1b). In each of these basins, we collected all 178 179 georeferenced occurrences of the selected species from GBIF (www.gbif.org, accessed May 2020) and additional published sources (Hortal & Lobo 2011). As this kind of data 180 is biased due to historically uneven sampling effort (Lobo et al. 2018), the occurrences 181 182 were pooled within UTM grid cells of 10 x 10 km spatial resolution. This grain was 183 selected because it corresponds to the effective resolution of most of the occurrence information in the dataset, and it is appropriate to avoid the effects of oversampled 184 localities while retaining a reasonable climatic detail. The frequency of each species' 185 occurrence data in temperature bins of 1°C (ranging from -3 to 20°C, n=24) was 186 187 calculated for each river basin ($24 \times 5 = 120$), and these figures were used as dependent 188 variables in the subsequent regression analyses.

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Seasonal dataset. Six sites along an elevational gradient located in the Sierra de 190 Guadarrama (Central Spain) (Fig. 1a, Espinoza 2016) were used to explore the effect of 191 temperature variations in SD. Elevations range from 755 to 1900 m a.s.l., separating sites 192 193 approximately 200 m a.s.l.. Each survey site was sampled approximately every three weeks, totalling fourteen times from May 2012 to June 2013. We choose this elevation 194 195 gradient because these sites show considerable variations in temperature during the whole 196 period of the surveys (Fig. 1b). The sampling protocol in each periodical sample 197 consisted of five pitfall-traps baited with fresh cattle dung and separated around 30 m 198 from each other. Traps were placed in open habitats to avoid potential habitat and shadow 199 effects and were active during 48 h. The individuals of these traps were pooled together, obtaining an estimation of each species' abundance per elevation site and date (6 x 14 =200 201 84), which were used as response variables in subsequent statistical analyses.

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203 *Diel Dataset.* Temperature effects on diel activity were assessed using dung beetle data from a grassland located next to El Ventorrillo MNCN field station, placed in the Sierra 204 205 de Guadarrama at an approximate elevation of 1500 m a.s.l. (Fig. 1a). This locality was 206 chosen as it shows a high diversity of dung beetles (between 30 and 40 species belonging 207 to the considered subfamilies; Cuesta & Lobo 2019). We sampled three consecutive days (April 28th-30th 2015) that showed contrasting temperatures, with around 8 °C of 208 difference between the mean temperatures of the coldest and the hottest days (Fig. 1b). 209 210 Each day, ten pitfall traps baited with fresh cattle dung were distributed around a 211 circumference of approximately 50 m. of radius (*i.e.*, traps were at least 30 m apart from 212 each other). Since we intended to measure the flight activity during short periods, the bait 213 was introduced into a nylon stocking piece to avoid the stagnancy of beetle individuals within the dung bait along different sampling events. We checked all traps every 30 min. 214 215 from dawn to dusk (approximately from 7:30 am to 7:00 pm, n=23), collecting all 216 individuals to subsequently identify them in the laboratory. Traps were also checked 217 during the night to discard nocturnal activity. Individuals from the ten traps were pooled 218 together, obtaining an estimation of the abundance of active individuals from each species each 30 min $(23 \times 3 = 69)$, which were further used as dependent variables. 219

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221 Temperature measures and alternative correlates

222 Temperature measures were obtained from different standardized methods for each one of the different spatio-temporal scales considered, but trying to maintain a considerable 223 224 degree of congruence among them. For the Geographical Dataset, we obtained mean 225 annual temperatures at a 30 sec resolution from the WorldClim database (see 226 www.worldclim.org, Hijmans et al. 2005). We preferred mean annual temperatures over 227 monthly average figures since the precise seasonal activity over the complete study area 228 was unknown for most of the species. Nevertheless, spring and autumn temperatures (the 229 seasons when phenological peaks occur for most species) were positively correlated with 230 mean annual temperatures (Pearson's r = 0.99 and 0.97, respectively), so we assume that 231 mean annual temperature is a reasonable proxy for both of them.

For the *Seasonal Dataset*, we set up a temperature data logger in each of the elevational points during the whole period of the study. This device was placed in the shadow at one meter from the ground to escape from extreme temperatures due to insolation, mimicking the meteorological stations on which WorldClim data are based on. Temperature was recorded each 10 min. and we used the mean daily temperature whenpitfall-traps were active.

238 In the case of the *Diel Dataset*, temperature measurements were taken using five 239 data loggers placed in the study site just in the centre of the circumference formed by the 240 traps. Data loggers were placed to recover temperature measurements from the different 241 microclimatic conditions available for dung beetles: two at one meter over the ground, in 242 the sun and shadow; another two directly on the ground, also both in the sun and shadow; 243 and one buried at 10 cm depth. Preliminary results showed that the mean temperatures 244 from the data logger placed on the ground in the sun were those that best correlated with 245 the species' diel activity, so we used these measurements in subsequent analyses. 246 Temperature was recorded each minute, and average temperatures during the 30 min 247 before traps were checked were used as predictors.

248 As previously stated, the effects of temperature measurements might be 249 overestimated due to its collinearity with other factors with which it shares spatial (in the 250 case of GD) or temporal structure (in the case of SD and DD). We quantified this 251 potential overestimation effect by using different "contrast variables", alternative 252 predictors which are often partly correlated with temperature but are either measures or 253 proxies of other potential causal factors for dung beetle spatial and temporal responses. 254 These alternative predictors were temperature availability and survey effort in the case of GD, day of the year for SD, and hour of the day in the case of DD. The effect of 255 temperature on the frequency of occurrence (GD) or abundance (SD and DD) that is 256 257 independent of these contrast variables was assessed as the "pure" effect of temperature 258 variations that is independent of the range of temperatures available (GD), the period of 259 the year (SD), and the hour of the day (DD) (see analytical methods below).

260 Temperature availability for each basin is the relative frequency of 10 x 10 km 261 UTM cells in each 1°C temperature bin. This variable aims to represent the thermal spectrum available in each basin. Hence, a high explanatory capacity of this variable on 262 263 the frequency of occurrence of a species would imply that the apparent thermal 264 preference of this species can be simply because its spatial pattern of occurrence mimics 265 the distribution of temperatures in the analyzed basin. Further, the typical correlation 266 between the observed pattern of occurrence of a species and the spatial distribution of 267 survey effort can also generate spurious correlations between species' frequency and temperature in each basin. This potential source of error was considered here by 268 269 calculating the relative frequency for each 1°C temperature bin of all dung beetle records

270 included in the formerly mentioned databases and pooled within the 10 x 10 km UTM cells. Nevertheless, we found that this estimation of survey effort and temperature 271 272 availability were highly and positively correlated in all basins (Pearson's r ranging from 273 0.97 to 0.99), since the most frequent temperatures have been also surveyed more often, 274 which implies that the surveys are randomly allocated within the available temperatures. 275 Consequently, we discarded using survey bias as contrast variable, considering that the 276 effect of temperature availability also includes differences in survey effort. In the case of SD data, the day of the year was obtained by first ordering the available dates from the 277 day corresponding to the summer solstice (June $21^{th} = 0$ or 360), to subsequently convert 278 these figures into radians and obtaining two circular variables by calculating their cosine 279 and sine values. Thus, the summer-winter oscillation is represented by the cosine of the 280 date and oscillates from 1 to -1, whereas the spring-autumn transition is represented by 281 282 the sine of the date scale 1 to -1. Finally, the hour of the day (DD data) is simply codified 283 as the number of minutes from dawn.

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285 Statistical analyses

286 *Explanatory capacity of temperature*

287 We first explored the independent capacity of temperature to explain variations in dung 288 beetle data in GD, SD, and DD. For each dataset, we conducted Generalized Linear Regression Models of the relative frequency or the abundance of each species as a 289 function of temperature values. All data coming from the five basins (in GD), the six 290 291 elevational sites (SD), and the three days (in DD) were considered at the same time in 292 each one of the three models. A curvilinear quadratic function of temperature was 293 included in all the cases to account for the typical unimodal performance curves of 294 ectotherms (Huey & Kingsolver 1989). A negative binomial error distribution for the 295 dependent variable was assumed to avoid overdispersion issues associated with the 296 Poisson error distribution (Blasco-Moreno et al. 2019), and it was related to the set of 297 predictors via a logarithmic link function. It is important to note that we did not include a 298 term in the models to account for the different spatial (i.e., basins and elevations) and 299 temporal (*i.e.*, days) units. By doing so, we were ignoring other factors that may affect 300 the distribution and activity of dung beetles, besides temperature and the contrast 301 variables analyzed. However, this allows us to tease apart the effects of temperature and 302 the formerly mentioned contrast variables with a spatial and temporal structure while 303 avoiding model overparameterization. Hence, we fitted three supplementary GLMs

304 representing different hypotheses regarding the importance of temperature: (i) a full model where both temperature and contrast variables are included altogether, (ii) a model 305 306 including only these contrast variables, and (iii) a null model where only the intercept was included. We assumed a linear relationship between the density of occurrence and 307 308 temperature availability (GD); whereas in SD and DD, we assumed curvilinear relationships between abundance and contrast variables by including a quadratic term of 309 both the number of minutes from dawn, and date sine and cosine. We used a deviance 310 311 partition approach (Legendre 1993, see also Calatayud et al. 2019 for the same approach) to calculate the deviance explained by each set of variables alone (*i.e.*, temperature vs. 312 contrast variables; herein, total pseudo R^2) and once accounting for the collinearity with 313 other variables (herein, partial pseudo R^2). Model performance was assessed using the 314 Akaike Information Criterion corrected for small sample size (AICc). 315

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317 *Thermal niche attributes*

Deriving thermal niches from occurrence data typically provides a partial description of 318 319 the whole potential response of the species (Sánchez-Fernández et al. 2012, Saupe et al. 320 2018). However, occurrence-based thermal niches may nevertheless be characterized by 321 different attributes such as the optimum temperature and niche breadth (Gouveia et al. 322 2014, Löffler & Pape 2020, Fig. 2). The temperature optimum of each species was assessed by fitting quadratic curves in a GLM and calculating the maxima as their 323 inflection point (see Villén-Pérez & Carrascal 2015 for a similar procedure). Thermal 324 325 niche breadth was also obtained as the area under the curve of these fitted curves. Fitted values were normalized to reach a maximum value of one to make calculations 326 327 comparable among datasets and species.

We evaluated the intraspecific dissimilarity in the thermal niches across different spatial and temporal scales, herein called "thermal lability", using data from the different study units used in each dataset; that is, between river basins, elevation sites, and days (Fig. 2). Thermal lability between pairs of units was measured using the Simpson index as follows:

 $333 \qquad S = \frac{\min(b,c)}{a + \min(b,c)},$

where *a* represents the area under the curves where both curves overlap, and *b* and *c* represent the independent areas under the curves in study units (see Fig. 2). The larger the overlap between the curves obtained at different scales, the smaller the thermal lability

will be. We computed this index for all pairs of units in each dataset (*i.e.*, for each pair of
basins, each pair of elevations, and each pair of days) and then considered the maximum
dissimilarity among all pairs from the same dataset, as this measure will provide a more
realistic estimate of the potential thermal lability of each species.

341

342 *Congruencies in thermal niches*

The congruence in the thermal niches of the different species derived from the three 343 datasets (i.e., GD, SD, and DD) was assessed using Spearman rank correlations between 344 the deviance explained by temperature (*i.e.*, both for the total and partial pseudo $R^{2}s$), as 345 well as the obtained temperature optima, thermal niche breadths and thermal labilities for 346 347 each pair of datasets. In addition to these descriptors, we explored the congruence in the overlap of thermal niches estimated from different datasets. To do this, we examined 348 349 whether interspecific thermal niche dissimilarities were correlated between the different 350 datasets. We computed dissimilarities between the models' normalized fitted values where the temperature was the only explanatory variable using the Simpson index as 351 previously explained, but in this case between pairs of species (see also Fig. 2). By doing 352 353 so, we created a thermal niche pairwise dissimilarity matrix for each dataset. Then, we 354 conducted Mantel tests based on Spearman's p correlation coefficient to assess the 355 relationship between dissimilarity matrices obtained from the different datasets. Significance was evaluated by comparing observed ρ coefficients with 999 null values 356 obtained by permuting the dissimilarity matrices. 357

358

359 Phylogenetic signal

360 The potential lability of thermal niches shall be also assessed from an evolutionary point of view. In this sense, a marked phylogenetic signal would indicate both potential 361 362 evolutionary constrains for temperature variation responses, and phylogenetically-363 structured effects of global warming. We reconstructed a Bayesian phylogenetic 364 hypothesis for 18 species present in our datasets based on two mitochondrial (COI and 365 COII) and one nuclear markers (28S RNA, see Appendix S1 for details on phylogenetic 366 reconstruction). DNA markers were sequenced for this study and retrieved from Genbank (Table S1, accessed in June 2016). Pagel's λ test (Pagel, 1999) and Blomberg's K 367 statistics (Blomberg et al. 2003) were used to explore the phylogenetic signal in the five 368 369 variables considered (total and partial deviance explained by temperature, temperature 370 optimum, thermal niche breadth, and thermal lability). , Significance for Pagel's λ was 371 assessed with a likelihood ratio test comparing the negative log likelihood obtained from 372 the original tree topology with the negative log likelihood from a topology transformed to 373 remove the signal (*i.e.*, $\lambda = 0$). In the case of Blomberg's K, we tested for significance by 374 randomizing the labels of the phylogenetic tips and comparing observed and random K 375 values. Finally, we also investigated for phylogenetic signal in the thermal niche 376 dissimilarities for each dataset. To do so, Spearman correlations between thermal 377 dissimilarities and phylogenetic distances were used, assessing significance by comparing 378 observed correlations with null values where the labels of the tips of the phylogeny were 379 randomized. In all cases where tip labels were randomized, p-values were calculated as 380 the proportion of null values being equal or higher than observed values.

All analyses were conducted in R environment (R Core Team 2020), using the AICcmodavg package (Mazerolle 2019) to calculate AICc values, the function "sintegral" as implemented in the *Bolstad2* packed (Curran 2013) to assess areas under the curves, the *vegan* package (Oksanen et al. 2019) for the Mantel tests, and the *phytools* package (Revell 2012) to calculate Pagel's λ and Blomberg's K.

386

387 **Results**

388 There is an evident gradient in the explanatory relevance of temperature towards higher 389 relevance at progressively larger scales (*i.e.*, geographical > seasonal > diel). Model 390 selection revealed that the full model, including temperature and contrast variables, was 391 the most parsimonious for most species in most datasets (Table 1). As exceptions to this general pattern, in the geographical dataset, the model only including temperature was 392 393 equivalent to the full model (according to AICc) for one species, and it was also the best 394 supported for another species. In the seasonal dataset, the model only including 395 temperature was the best supported for four species, whereas the model only including 396 contrast variables was equivalent to the full model for just one species. Finally, the model 397 including minute from dawn in DD data was equivalent to the full model for only two 398 species and even better for one species (Table 1). In general, the total deviance explained by the models including temperature and contrast variables was considerably high (mean 399 pseudo- $R^2s = 0.62$, 0.63, and 0.77; ranges = 0.51-0.75, 0.38-0.86, and 0.64-0.86, 400 401 respectively for GD, SD, and DD; see Fig. 3). Partial regressions revealed that the effects of temperature and contrast variables largely overlap, being the deviance independently 402 explained by temperature considerably low (see Fig. 3). Interestingly, the percentage of 403

404 deviance explained by temperature decreased from the geographical (mean pseudo- $R^2s =$ 405 0.33; range 0.13–0.48), to the seasonal (0.19; 0.05–0.36) and diel datasets (0.08; 0.01– 406 0.20) (see Fig. 3).

Thermal niche attributes derived from the different datasets showed little 407 congruence. Neither the pseudo R^2 explained by temperature alone nor the total pseudo 408 R^2 were positively and significantly correlated between any pair of datasets, and none of 409 the thermal niche attributes were significantly correlated between the three considered 410 411 datasets (Table 2). Moreover, Mantel tests showed that interspecific niche dissimilarities 412 were not correlated among the three studied spatiotemporal scales (Table 2). Finally, we 413 did not find phylogenetic signal for any of these variables in any of the datasets, except in 414 the case of niche breadth for the diel dataset (Table 3).

415

416 Discussion

417 Our results show that the spatial and temporal responses of the studied species show large 418 associations to contrast variables besides temperature, but also that temperature controls 419 to dung beetle occurrence may increase towards larger temporal and spatial scales. This 420 contrasts with our preliminary expectations of a high importance of temperature for dung 421 beetle occurrence and activity based on the known basal ectothermic physiology of the 422 considered species. Further, thermal niches were incongruent across scales for the studied species and lacked phylogenetic signal, indicating that thermal adaptations are highly 423 variable both between and within species. 424

The generally low partial effects of temperature found in our study lead to two 425 important conclusions: (i) the abundance, distribution, and activity of dung beetles are 426 427 controlled by other factors different from temperature, which are at least partially represented by the *ad hoc* contrast variables used here; and (ii) dung beetle species must 428 429 have biological mechanisms that provide them with the plasticity required to cope with the temperature variations associated to each spatiotemporal context. Thermoregulation 430 431 and body heat gain are intimately linked to solar radiation in ectotherms (Angilletta 432 2009). Indeed, empirical evidence suggest that solar radiation is associated with dung 433 beetles' body temperatures (Bartholomew & Heinrich 1978) and temporal variations in 434 their abundance and species richness (Lobo et al. 1998). Hence, it is likely that this factor 435 is a key environmental control of the diel activity of dung beetles. Regarding annual rhythms, photoperiod seems to be a crucial environmental cue regulating insects' 436 437 seasonality (Nijhout 1994, Bradshaw & Holzapfel 2007). This is likely the case for dung

beetles, given the relatively weak effects of temperature on their phenology found in our 438 study. Also, the different life-history phases of an insect need to be synchronized 439 440 seasonally, and these require a minimum amount of time to complete. The development 441 of a dung beetle individual requires from 30 to 80 days depending on the species 442 (Christensen & Dobson 1977, Romero-Samper & Martín-Piera 1995, 2007, Arellano et 443 al. 2017), a time that determines key life-history characteristics such as the number of 444 generations per year or the overwintering phase. These developmental constraints are 445 therefore hard to modify without major evolutionary changes (Teder 2020), thereby 446 limiting the effects of environmental temperature on the seasonal abundance and occurrence of dung beetle species. Finally, many factors contribute to shaping the 447 448 geographical distribution of dung beetle species, including dispersal limitations (Lobo et al. 2006), historical events (Hortal et al. 2011), or the response to other environmental 449 450 variables such as precipitation, soil, habitat, or trophic preferences (Hanski et al. 1991, Hortal et al. 2001, Lobo & Martín-Piera 2002). It is important to note that we have not 451 452 quantified the effects of these variables explicitly, so their inclusion could further weaken 453 the pure effect of temperature.

454 Regardless of the effects of alternative factors, it seems accurate that dung 455 beetles have mechanisms to withstand marked temperature variations, especially those 456 associated with diel and seasonal rhythms. Given the nature of our data and analyses, these mechanisms can be operating either at the population level, at the individual level, 457 or both. At the population level, a high genetic diversity linked to large phenotypic 458 variability can produce the apparently labile thermal responses. That is, as individuals are 459 sorted in time and/or space according to their environmental adaptations, population(s) 460 461 formed by individuals with different thermal preferences would show a certain level of 462 thermal independency. This mechanism seems more plausible to explain results in the 463 geographical datasets, where river basins can act as dispersal barriers, limiting gene flow and enhancing local adaptation to different temperature regimens (Lenormand 2002). 464 However, it seems less likely that this phenotypic variability alone is responsible for the 465 466 responses to diel and seasonal temperature variations, where a high gene flow is expected between the individuals and populations that are active at different elevations or days. 467 468 Physiologically plastic responses allowing individuals to be active at different 469 temperatures seem a more plausible mechanism in this case (Crispo 2008). In any case, these two potential mechanisms (phenotypic variability and individual plasticity) are in 470 471 agreement with the observed lack of phylogenetic signal on species responses to temperature across scales, which indeed suggests a lack of thermal niche conservatism
(Gilbert & Miles 2019). The relative contribution of population phenotypic variability
and individual plasticity remains elusive, calling for further studies directed to unravel the
detailed mechanisms behind the diverse responses to temperature found in our study.

476 Be that as it may, the effects of temperature were significant and not negligible, 477 being larger for species distribution than for seasonal activity, and even smaller for diel 478 activity. The increasing importance towards larger scales may be related to the fact that 479 the effects of temperature on the studied biological aspects are nested. That is, the 480 occurrence in a given location would entail that a species holds the adaptations required 481 to maintain a stable population there, which include physiological and/or behavioural 482 adaptations to cope with the seasonal temperature variations that occur in that locality. In the same way, a population with adults active during a given period of the year should 483 484 present adaptations to handle the daily temperature variations happening during the days when adults are active. Hence, the hierarchically cumulative effects of temperature across 485 486 these biological scales may explain why temperature becomes more important for 487 geographic distributions than for temporal activities. Ascertaining the plausibility of this 488 idea requires further investigation of intraspecific responses to daily temperature 489 variations across seasons and seasonal temperature responses throughout different 490 populations placed across the species' geographic distribution.

Perhaps the most interesting of our results is the lack of congruence in the 491 realized thermal niches across the studied species and spatiotemporal contexts. This 492 means that, for instance, species occurring in colder regions do not appear in colder 493 months nor at colder hours of the day in other regions. This somehow counterintuitive 494 495 result could be related to the uneven relevance of the alternative variables for the 496 different species and spatiotemporal contexts, which facilitates the decoupling of the 497 thermal responses associated with the distribution and activity of dung beetles. It is likely that the processes involved in adult movements, life-history cycles, and population 498 499 maintenance are differently regulated by temperature, despite of their nested nature. In 500 other words, our results suggest that species have multidimensional thermal niches, where 501 each critical biological aspect responds to temperature along a different dimension. 502 Therefore, rather than exerting a universal effect, temperature plays multiple roles in a 503 species' biology and metapopulation dynamics. This lack of congruence, together with the low independent effects of temperature found in our deviance partition analyses, 504 505 indicates that estimates of thermal niches will be, in general, inaccurate and context-

dependent. This calls from being particularly cautious when using responses measured at different scales as proxies for future responses to climate change. Overall, our results show the difficulties in estimating general thermal niches of species, challenging forecasts of species future dynamics under climate warming based on unidimensional thermal niches (Gvoždik, 2018).

511 The partial control of temperature on the activity and distribution of dung beetles 512 may be both a blessing and a curse regarding the effects of climate warming. On the one 513 hand, the apparent thermal lability suggests that temperature increases should not 514 strongly modify neither diel and seasonal activities nor the geographic distribution of dung beetles, likely preventing mismatches with interacting species and the subsequent 515 food chain perturbations. This assumption would contradict the results of studies 516 suggesting moderate or even large effects of climate change on dung beetle distributions 517 518 (Dortel et al. 2013, Menéndez et al. 2013, Holley & Andrew 2019). On the other hand, the diel, seasonal, or geographical adjustments are among the fastest responses to climate 519 520 warming (Levy et al. 2019, Duchenne et al. 2020). However, our results suggest that the 521 response towards temperature variations is relatively independent at each spatiotemporal 522 scale. This entails that adjustments to temperature requirements may not be coordinated 523 across key biological aspects. Hence, adjustments to fulfil the temperature requirements 524 for one biological aspect may result in detrimental effects on other aspects, thereby 525 reducing individual and population performance as, e.g., seasonal adjustments may expose individuals to inadequate temperatures during diel activity. In the worst-case 526 scenario, the incapacity of species to adjust their temperature requirements by modifying 527 diel, seasonal, and geographical patterns at convenience will increase the likelihood of 528 529 local extinctions when the individuals are exposed to critical temperatures in their daily or yearly periods of activity. Paradoxically, the partially weak effects of temperature we 530 531 found may have serious consequences for climate warming if temperature regulates important aspects of species' biology in divergent ways (Tsai et al. 2020). 532

533 Overall, our results show that temperature may be less important than other 534 factors in determining dung beetle activity and distribution. Further, the incongruences in 535 thermal niches estimated from the geographic distribution and seasonal and diel activities 536 show the complex effects of temperature on key species aspects, pointing to a truly 537 multidimensional nature of thermal niches. Together with the partially weak control of 538 temperature on species activity and distribution, these incongruences may difficult fast

539	responses to climate warming, potentially exposing individuals to critical, or at least
540	inadequate, temperatures and reducing individual and population's fitness.

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542 **References** 543 Angilletta M I

- Angilletta, M.J., 2009. *Thermal adaptation: a theoretical and empirical synthesis*. Oxford
 University Press, Oxford, UK.
 Araújo, M.B., Nogués-Bravo, D., Diniz-Filho, J.A.F., Haywood, A.M., Valdes, P.J. &
- Rahbek, C. 2008. Quaternary climate changes explain diversity among reptiles
 and amphibians. *Ecography* 31, 8-15.
- Arellano, L., Castillo-Guevara, C., Huerta, C., Gemán-García, A. & Lara, C. 2017.
- 549 Nesting biology and life history of the dung beetle *Onthophagus leontei*550 (Coleoptera: Scarabaeinae). Animal *Biology* 67, 41-52.
- Bartholomew, G.A. & Heinrich, B. 1978. Endothermy in African dung beetles during
- flight, ball making, and ball rolling. *Journal of Experimental Biology* 73, 65-83.
- Blasco-Moreno, A., Pérez-Casany, M., Puig, P., Morante, M. & Castells, E. 2019. What
 does a zero mean? Understanding false, random and structural zeros in ecology. *Methods in Ecology and Evolution* 10, 949-959.
- Blomberg, S.P., Garland Jr., T. & Ives A.R. 2003. Testing for phylogenetic signal in
 comparative data: behavioral traits are more labile. *Evolution* 57: 717-745.
- Bradshaw, W.E. & Holzapfel, C.M. 2007. Evolution of animal photoperiodism. *Annual Review in Ecology, Evolution and Systematic* 38, 1-25.
- Bradshaw, W.E. & Holzapfel, C.M. 2010. Light, time, and the physiology of biotic
 response to rapid climate change in animals. *Annual Review of Physiology* 72, 147-166.
- Brown, J.H., Gillooly, J.F., Allen, A.P., Savage, V.M & West, G.B. 2004. Toward a
 metabolic theory of ecology. *Ecology* 85, 1771-1789.
- Calatayud, J., Hortal, J., Medina, N.G., Turin, H., Bernard, R., Casale, A., Vicente M.,
 O., Lyubomir, P. & Rodríguez, M.Á. 2016. Glaciations, deciduous forests, water
 availability and current geographical patterns in the diversity of European *Carabus* species. *Journal of Biogeography* 43, 2343-2353.
- Calatayud, J., Rodríguez, M.Á., Molina-Venegas, R., Leo, M., Horreo, J.L., & Hortal, J.
 2019. Pleistocene climate change and the formation of regional species pools. *Proceedings of the Royal Society B* 286, 20190291.

572	Chapin III, F.S. & Díaz, S. 2020. Interactions between changing climate and biodiversity:
573	Shaping humanity's future. Proceedings of the National Academy of Science
574	U.S.A. 117, 6295-6296
575	Crispo, E. 2008. Modifying effects of phenotypic plasticity on interactions among natural
576	selection, adaptation and gene flow. Journal of Evolutionary Biology 21, 1460-
577	1469.
578	Christensen, C.M. & Dobson, R.C. 1977. Biological studies on Aphodius fimetarius (L.)
579	(Coleoptera, Scarabaeidae). Journal of Kansas Entomological Society 50, 129-
580	134.
581	Cuesta, E. & Lobo, J.M. 2019. A comparison of dung beetle assemblages (Coleoptera,
582	Scarabaeoidea) collected 34 years apart in an Iberian mountain locality. Journal
583	of Insect Conservation 23, 101-110.
584	Curran, J.M. 2013. Bolstad2: Bolstad functions. R package version 1.0-28.
585	https://CRAN.R-project.org/package=Bolstad2
586	Dortel, E., Thuiller W., Lobo J.M., Bohbot H., Jay-Robert P. & Lumaret J.P. 2013.
587	Potential effects of climate change on the distribution of Scarabaeidae dung
588	beetles in Western Europe. Journal of Insect Conservation 17, 1059-1070.
589	Duchenne, F., Thébault, E., Michez, D., Elias, M., Drake, M., Persson, M., Piot, J.S.,
590	Pollet, M., Vanormelingen, P. & Fontaine, C. 2020. Phenological shifts alter the
591	seasonal structure of pollinator assemblages in Europe. Nature Ecology &
592	Evolution 4: 115-121.
593	Edwards, M. & Richardson, A.J. 2004. Impact of climate change on marine pelagic
594	phenology and trophic mismatch. Nature, 430, 881-884.
595	Espinoza, V.R. 2016. Distribución y estructura de las comunidades de escarabeidos
596	coprófagos (Coleoptera, Scarabaeoidea) a lo largo de gradientes de
597	altitud. Unpublished PhD thesis available at https://eprints.ucm.es/36438/
598	Gilbert, A.L. & Miles, D. B. 2019. Spatiotemporal variation in thermal niches suggests
599	lability rather than conservatism of thermal physiology along an environmental
600	gradient. Biological Journal of the Linnean Society 128, 263-277.
601	Guo, F., Guénerad, B., Economo, E.P., Deutsch, C.A. & Bonebrake, T.C. 2020. Activity
602	niches outperform thermal physiological limits in predicting global ant
603	distributions. Journal of Biogeography 47, 829-842.
604	Gouveia, S.F., Hortal, J., Tejedo, M., Duarte, H., Cassemiro, F.A.S., Navas, C.A. &
605	Diniz-Filho, J.A.F. 2014. Climatic niche at physiological and macroecological

606	scales: the thermal tolerance-geographical range interface and niche
607	dimensionality. Global Ecology and Biogeography 23, 446-456
608	Gvoždik, L. 2018. Just what is the thermal niche? Oikos 127, 1701-1710.
609	Hanski, I. & Cambefort, Y. 1991. Resource partitioning. In: Dung Beetle Ecology.
610	Hanski, I. & Cambefort, Y (eds.) Princeton University Press, Princeton, NJ. Pp.
611	330-349.
612	Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G. & Jarvis, A. 2005. Very high
613	resolution interpolated climate surfaces for global land areas. International
614	Journal of Climatology 25, 1965-1978.
615	Holley, J.M. & Andrew, N.R. 2019. Experimental warming alters the relative survival
616	and emigration of two dung beetle species from an Australian dung pat
617	community. Austral Ecology 44, 800-811.
618	Hortal, J., Lobo, J.M. & Martín-Piera, F. 2001. Forecasting insect species richness scores
619	in poorly surveyed territories: the case of the Portuguese dung beetles (Col.
620	Scarabaeinae). Biodiversity and Conservation 10, 1343-1367.
621	doi:10.1023/A:1016624500023
622	Hortal, J. & Lobo, J.M. 2011. Can species richness patterns be interpolated from a limited
623	number of well-known areas? Mapping diversity using GLM and kriging.
624	Natureza & Conservação 9, 200-207.
625	Huey, R.B. & Kingsolver, J.G. 1989. Evolution of thermal sensitivity of ectotherm
626	performance. Trend in Ecology and Evolution 4, 131-135.
627	Legendre, P. 1993. Spatial autocorrelation: trouble or new paradigm? Ecology 74, 1659-
628	1673 doi: 10. 2307/1939924
629	Lehner, B. & Grill, G. 2013. Global river hydrography and network routing: baseline data
630	and new approaches to study the world's large river systems. Hydrological
631	Processes 27, 2171-2186.
632	Lenormand, T. 2002. Gene flow and the limits to natural selection. Trends in Ecology &
633	Evolution 17, 183-189.
634	Levy, O., Dayan, T., Porter, W.P. & Kronfeld-Schor, N. 2019. Time and ecological
635	resilience: can diurnal animals compensate for climate change by shifting to
636	nocturnal activity? Ecological Monographs 89, e01334.
637	Lobo, J.M., Lumaret, J.P. & Jay-Robert, P. 1998. Sampling dung beetles in the French
638	Mediterranean area: Effects of abiotic factors and farm practices. Pedobiologia
639	42, 252-266.

640	Lobo, J.M. & Martín-Piera, F. 2002. Searching for a predictive model for species richness
641	of Iberian dung beetle based on spatial and environmental variables. Conservation
642	<i>Biology</i> 16, 158-173.
643	Lobo, J.M., Verdú, J.R. & Numa, C. 2006. Environmental and geographical factors
644	affecting the Iberian distribution of flightless Jekelius species (Coleoptera:
645	Geotrupidae). Diversity and Distributions 12, 179-188.
646	Lobo, J.M., Hortal, J., Yela, J.L., Millán, A., Sánchez-Fernández, D., García-Roselló, E.,
647	González-Dacosta, J., Heine, J., González-Vilas, L. & Guisande, C. 2018.
648	KnowBR: An application to map the geographical variation of survey effort and
649	identify well-surveyed areas from biodiversity databases. Ecological Indicators
650	91, 241-248.
651	Löffler, J. & Pape, R. 2020. Thermal niche predictors of alpine plant species. Ecology
652	101, e02891.
653	Madrigal-González, J., Andivia, E., Zavala, M.A., Stoffel, M., Calatayud, J., Sánchez-
654	Salguero, R. & Ballesteros-Cánovas, J. 2018. Disentangling the relative role of
655	climate change on tree growth in an extreme Mediterranean environment. Science
656	of the Total Environment 642, 619-628.
657	Mazerolle, M.J. 2019. AICcmodavg: Model selection and multimodel inference based on
658	(Q)AIC(c). R package version 2.2-2. https://cran.r-
659	project.org/package=AICcmodavg.
660	Mayhew, P.J., Jenkins, G.B. & Benton, T.G. 2008. A long-term association between
661	global temperature and biodiversity, origination and extinction in the fossil
662	record. Proceedings of the Royal Society B 275, 47-53.
663	Menéndez, R., González□Megías, A., Jay□Robert, P. & Marquéz□Ferrando, R. 2014.
664	Climate change and elevational range shifts of dung beetles. Global Ecology and
665	Biogeography 23, 646-657. doi:10.1111/geb.12142.
666	Milotić, T., et al. 2019. Functionally richer communities improve ecosystem functioning:
667	Dung removal and secondary seed dispersal by dung beetles in the Western
668	Palaearctic. Journal of Biogeography 46, 70-82.
669	Nervo, B., et al. 2017. Ecological functions provided by dung beetles are interlinked
670	across space and time: evidence from 15N isotope tracing. <i>Ecology</i> 98, 433-446.
671	Nijhout, H.F. 1994. Insect Hormones. Princeton, NJ: Princeton Univversity Press
672	Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D.,
673	Minchin, P.R., O'Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H.,

674	Szoecs, E. & Wagner H. 2019. vegan: Community Ecology Package. R package
675	version 2.5-6. https://CRAN.R-project.org/package=vegan
676	Paaijmans, K.P., Heining, R.L., Seliga, R.A., Blanford, J.I., Blanford, S., Murdock, C.C.
677	& Thomas, M.B. 2013. Temperature variation makes ectotherms more sensitive to
678	climate change. Global Change Biology 19, 2373-2380.
679	Pagel, M. 1999. Inferring the historical patterns of biological evolution. Nature 401, 877-
680	884.
681	R Core Team 2020. R: A language and environment for statistical computing. R
682	Foundation for Statistical Computing, Vienna, Austria. URL https://www.R-
683	project.org/.
684	Rader, R., Reilly, J., Bartomeus, I. & Winfree, R. 2013. Native bees buffer the negative
685	impact of climate warming on honey bee pollination of watermelon crops. Global
686	Change Biology 19, 3103-3110.
687	Revell, L.J. 2012. phytools: An R package for phylogenetic comparative biology (and
688	other things). Methods in Ecology and Evolution 3, 217-223.
689	Román-Palacios, C. & Wiens, J. 2020. Recent responses to climate change reveal the
690	drivers of species extinction and survival. Proceedings of the National Academy
691	of Science U.S.A. 117, 4211-4217.
692	Romero-Samper, J. & Martín-Piera, F. 1995. Nesting behaviour, ontogeny and life -cycle
693	of Onthophagus stylocerus (Coleoptera; Scarabaeidae). European Journal of
694	Entomology 92, 667-679.
695	Romero-Samper, J. & Martín-Piera, F. 2007. Comportamiento reprodutor y ciclo
696	biológico de Aphodius conjugatus (Panzer, 1795) (Coleoptera, Aphodiidae).
697	Boletín Sociedad Entomológica Aragonesa 41, 189-192.
698	Sánchez-Fernández, D., Aragón, P., Bilton, D.T. & Lobo, J.M. 2012. Assesing the
699	congruence of thermal niche estimations derived from distribution and
700	physiological data. A test using diving beetles. PLoS ONE 7, e48163.
701	Saunders, D.S. 2020. Dormancy, diapause, and the role of the circadian system in Insect
702	photoperiodism. Annual Review of Entomology 5, 373-389.
703	Saupe, E.E., Barve, N., Owens, H.L., Cooper, J.C., Hosner, P.A. & Peterson, A.T. 2018.
704	Reconstructing Ecological Niche Evolution When Niches Are Incompletely
705	Characterized. Systematic Biology 67, 428-438
706	Schwartzman, D. 1999. Life, Temperature, and the Earth. Columbia University Press,
707	New York.

708	Schweiger, O., Settele, J., Kudrna, O., Klotz, S., & Kühn, I. 2008. Climate change can
709	cause spatial mismatch of trophically interacting species. Ecology 89, 3472-3479.
710	Scranton, K. & Amarasekare, P. 2017. Predicting phenological shifts in a changing
711	climate. Proceedings of the National Academy of Sciences 114, 13212-13217.
712	Sheldon, K.S., Yang, S. & Tewksbury, J.J. 2011. Climate change and community
713	disassembly: impacts of warming on tropical and temperate montane community
714	structure. Ecology Letters 14, 1191-1200.
715	Somero, G. 2005. Linking biogeography to physiology: Evolutionary and acclimatory
716	adjustments of thermal limits. Frontiers in Zoology, 2: 1-9.
717	Sunday, J., Bennett, J.M., Calosi, P., Clusella-Trullas, S., Gravel, S., Hargreaves, A.L.,
718	Leiva, F.P., Verberk, W.C.E.P., Olalla-Tárraga, M.A. & Morales-Castilla, I. 2019.
719	Thermal tolerance patterns across latitude and elevation. Philosophical
720	Transactions of the Royal Society B 374, 20190036.
721	Tsai, H. Y., et al. 2020. Locally-adapted reproductive photoperiodism determines
722	population vulnerability to climate change in burying beetles. Nature
723	communications 11, 1-12.
724	Teder, T. 2020. Phenological responses to climate warming in temperate moths and
725	butterflies: species traits predict future changes in voltinism. Oikos 129, 1051-
726	1060.
727	Thackeray, S.J. et al. 2016. Phenological sensitivity to climate across taxa and trophic
728	levels. Nature 535, 241-245.
729	Verdú, J.R., Arellano, J. & Numa, C. 2006. Thermoregulation in endothermic dung
730	beetles (Coleoptera: Scarabaeidae): Effect of body size and ecophysiological
731	constraints in flight. Journal of Insect Physiology 52, 854-860.
732	Verdú, J.R. & Lobo, J.M. 2008. Ecophysiology of thermoregulation in endothermic dung
733	beetles: ecological and geographical implications. In Fattorini, S (ed). Insect
734	Ecology and Conservation. Research Singpost, Trivandrum, Kerala, pp. 299-317.
735	Verdú, J.R., Alba-Tercedor, J. & Jimenez-Manrique, M. 2012. Evidence of different
736	thermoregulatory mechanisms between two sympatric Scarabaeus species using
737	infrared thermography and micro-computer tomography. PLoS ONE 7, e33914
738	Villén-Pérez, S. & Carrascal, L.M. 2015. Occurrence data may provide unreliable thermal
	• •

- 740 Wang, Z., Brown, J.H., Tang, Z. & Fang, J. 2009. Temperature dependence, spatial scale,
- and tree species diversity in eastern Asia and North America. *Proceedings of the*
- 742 *National Academy of Sciences* 106, 13388-13392.

Table 1. AICc values for the models of each species in each dataset. In all cases, we conducted a complete model (Full) including temperature and the corresponding contrast variables, a model only including temperature (Temp), a model only including contrast variables (Cont), and a null model were no predictor variable was included (Null). Contrast variables were minutes from dawn and its quadratic term for the diel data set; date sine and cosine and their quadratic terms for the seasonal dataset; and temperature availability for the geographic data set. The best models in terms of AICc and the equivalent ones (Δ AICc < 2) are highlighted in bold.

		Diel		Seasonal				Geographic					
Subfamily	Species	Full	Temp	Cont	Null	Full	Temp	Cont	Null	Full	Temp	Cont	Null
Aphodiinae	Acrossus depressus (Kugelann, 1792)	176.53	245.53	187.59	264.43	120.25	126.26	120.67	149.35	172.98	173.17	209.50	209.95
Aphodiinae	Agrilinus constans (Duftschmid, 1805)	140.21	197.94	145.50	214.29	200.92	210.56	206.92	216.62	211.85	218.76	269.52	289.54
Aphodiinae	Aphodius fimetarius (Linnaeus, 1758)	116.91	162.51	121.10	177.61	20 1.68	199.58	2 17.13	214.00	380.68	404.55	419.41	455.12
Aphodiinae	Aphodius foetidus (Herbst, 1783)	42.72	56.56	44.51	64.09	138.39	133.18	146.06	144.19	634.53	679.39	753.81	861.04
Aphodiinae	Colobopterus erraticus (Linnaeus, 1758)	128.81	150.10	136.82	164.96	208.49	251.21	2 17 .59	272.82	343.34	366.05	372.06	410.67
Aphodiinae	<i>Esymus pusillu</i> s (Herbst, 1789)	175.62	227.70	171.67	241.45	151.68	163.52	2 17, 18	191.76	147.93	153.05	177.47	180.48
Aphodiinae	<i>Melinopterus spha</i> cel <i>atus</i> (Panzer, 1798)	471.34	534.10	493.63	577.91	307.12	322.42	321.94	343.35	289.63	304.26	354.89	390.33
Aphodiinae	Teuchestes fossor (Linnaeus, 1758)	194.71	256.95	208.75	280.58	89.06	98.58	96.18	116.19	258.00	268.42	304.76	318.01
Aphodiinae	Trichonotulus scrofa (Fabricius, 1787)	144.87	185.24	160.35	207.64	204.82	226.12	2 16.37	261.94	182.51	186.38	221.98	242.86
Aphodiinae	Volinus sticticus (Panzer, 1798)	305.09	342.35	310.59	371.12	116.05	407.35	121.57	122.05	133.88	132.11	167.53	169.92
Scarabaeinae	Euoniticellus fulvus (Goeze, 1777)	39.24	59.86	49.17	67.31	446.74	473.23	454.04	519.71	285.43	306.31	306.07	352.08
Scarabaeinae	Onthophagus fracticornis (Preyssler, 1790)	255.26	325.6	266.11	350.04	188.32	184.52	200.87	201.75	274.14	279.86	317.51	329.60
Scarabaeinae	Onthophagus lemur (Fabricius, 1781)	117.11	153.81	124.15	170.30	200.57	158.52	205.75	174.36	231.64	238.74	290.70	313.01
Scarabaeinae	Onthophagus opacicollis Reitter, 1892	71.03	80.84	76.13	90.69	343.31	355.11	350.70	356.49	207.58	215.60	218.29	251.59
Scarabaeinae	Onthophagus similis (Scriba, 1790)	256.91	342.23	260.48	359.42	612.48	617.79	6 46 .30	658.81	312.48	328.99	363.43	400.08
Scarabaeinae	<i>Onthophagus vacca</i> (Linnaeus, 1767)	248.03	300.37	248.30	318.63	285.06	296.59	299.48	318.99	315.14	337.44	352.21	409.05

Table 2. Spearman's ρ correlation coefficients and P-values between the considered thermal niche attributes measured by the three studied datasets are detailed.. DD: Diel dataset. SD: Seasonal dataset. GD: Geographical dataset. * Results based on Mantel test.

	DD v	s SD	DD v	s GD	SD vs GD		
	ρ	Р	ρ	Р	ρ	Р	
Breadth	-0.074	0.788	-0.385	0.141	0.100	0.713	
Optimum	-0.262	0.326	-0.179	0.505	0.394	0.132	
Thermal lability	-0.261	0.347	-0.339	0.216	0.132	0.625	
Total pseudo R²	0.029	0.914	-0.016	0.953	-0.200	0.456	
Partial pseudo R²	-0.561	0.024	-0.440	0.088	0.053	0.848	
Niche dissimilarity *	0.260	0.056	0.242	0.051	0.120	0.153	

Table 3. Phylogenetic signal in thermal niches attributes for the three studied datasets (*i.e.*, geographical, seasonal and diel). Significant variables are highlighted in bold. * Results based on Mantel test.

Variable	Dataset	к	Ρ	λ	Ρ	ρ	Р
	Geographical	0.480	0.150	0	1	-	-
Breadth	Seasonal	0.285	0.798	0	1	-	-
	Diel	0.998	0.001	1.096	0.002	-	-
	Geographical	0.292	0.811	0.002	0.990	-	-
Optimum	Seasonal	0.477	0.188	0.361	0.160	-	-
	Diel	0.292	0.710	0	1	-	-
	Geographical	0.475	0. 17 1	0	1	-	-
Thermal lability	Seasonal	0.295	0.900	0	1	-	-
	Diel	0.467	0.228	0.007	0.970	-	-
	Geographical	0.287	0.857	0	1	-	-
Total pseudo R ²	Seasonal	0.324	0.710	0	1	-	-
	Diel	0.3 13	0.726	0	1	-	-
	Geographical	0.367	0.476	0	1	-	-
Partial pseudo R ²	Seasonal	0.388	0.398	0	1	-	-
	Diel	0.425	0.350	0	1	-	-
	Geographical	-	-	-	-	-0.100	0.089
Niche dissimilarity *	Seasonal	-	-	-	-	0.200	0.980
	Diel	-	-	-	-	-0.080	0.139

Figure 1. a) The areas of study for the geographical, seasonal and diel datasets (from left to right). Red squares show the position of the following down-scaled study site. b) Temperature variations in study sites. Lines correspond with predictions of general additive models (GAM) of: (i) temperature availability (measured as the number of 10 km² grid cells whose temperature fell within predefined temperature bins) as function of temperature for the geographic dataset (left); (ii) temperature as a function of days from New Year and minutes form dawn for the seasonal and diel datasets respectively. Analyses were computed independently for each basin, for each elevational site and for each day. GAMs explained an average of 0.90 of deviance across all analyses (median = 0.92, ranging from 0.79 to 0.97).



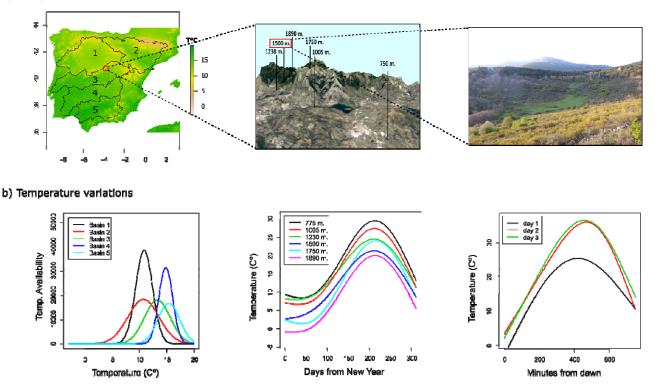
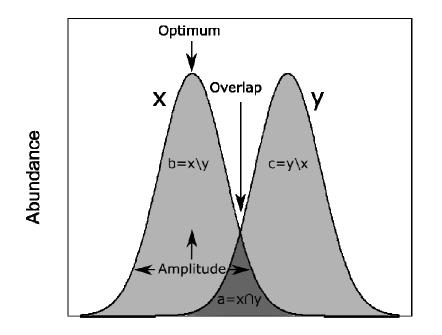
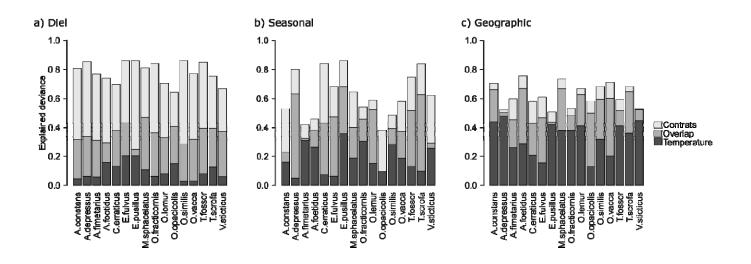


Figure 2. Thermal niche attributes and overlap measure. x and y represent thermal response curves of two species or of a single species in two different study units (*i.e.*, days, elevation or river basins). From this curve we obtained the optimum temperature and the niche amplitude. Further, we used the overlap between them (a) and the two independent areas (b and c) to calculate the Simpson's dissimilarity index, as a measure of the congruence between the responses to temperature of the same species at different scales, and of different species within the same scale.



т⁰С

Figure 3. Partial regression results. The deviance explained by temperature alone, the contrast variables alone, and the overlap between them is shown. The contrast variables were minutes from dawn and its quadratic term for the diel data set (a); date sine and cosine and their quadratic terms for the seasonal data set (b), and temperature availability for the geographic data set (c).



Supporting information

Appendix S1

Genomic DNA was extracted from each individual using the BIOSPRINT 15 DNA Kit (Qiagen), following standard manufacturer's protocols for blood, and resuspended in 100 µl of buffer AE. We used COI Sca F, COI Sca R, COII am Sca and COII B 605 Sea (Villalba et al. 2002) and the universal 28S a y 28S 5b primers to amplify fragments of the mitochondrial cytochrome oxidase I (COI), the cytochrome oxidase II (COII) and the 28S genes. Amplifications for all gene fragments were performed in a 50 µl reaction containing 39.7 μ l of H₂O, 5 μ l of 10x PCR buffer, 1 μ l of dNTP mix (10 mM), 0.5 μ l of each primer (10 μ M), 0.3 μ l of AmpliTaq® DNA polymerase (Applied Biosystems) and 3 ul of DNA template. Thermocycling conditions consisted of an initial denaturing step at 94 °C for 4 min, followed by two cycles: (i) a precycle of 5 amplification cycles of 94 °C for 45 sec, 40 °C for 1 min and 72 °C for 1 min, and (ii) a cycle of 35 amplification cycles of 94 °C for 45 sec, 44 °C for 1 min and 72 °C for 1 min, followed by a final elongation step at 72 °C for 10 min and a rapid thermal ramp down to 4 °C. For all reactions, the presence of amplicons of the expected sizes was checked by electrophoresis on a 0.8 % agarose gel. PCR products were purified with the ethanolprecipitation method (Sambrook et al., 1989). Sequencing was performed by Secugen S.L. (Madrid, Spain), using BigDye® and the automated ABI PRISM 3730x1 DNA Analyzer. Sequence chromatograms were read and contigs assembled using Sequencher version 4.7 (Gene Codes Corporation, Ann Arbor, MI). All new sequences were deposited in GenBank (see accession numbers in Table S1).

Sequences were aligned in CLUSTALW and MUSCLE, followed by visual inspection using BioEdit (Hall, 1999). Prior to phylogenetic analysis, jModeltest 2.1.1 (Darriba et al., 2012) was used to choose the best-fit model of nucleotide substitution for each of the four genes, and for combined matrices under the corrected Akaike information criterion (AICc). For the COI and COII, HKY was obtained, while Jukes Cantor for 28S. Phylogenetic analyses were performed in a Bayesian framework using BEAST v 2.4 (Drummond and Rambaut, 2007). We established 3 calibrations points based on Ahrens et. al (2014), setting uniform priors with lower and upper boundaries. The calibrations represent the basal split of the following taxa: Aphodiinae (58.7 – 55.8 Million years ago), *Aphodius* (37.2 – 33.9 Mya) and Scarabaeinae (92 – 83.5 Mya). For the age of the rest of the nodes, we set a LogNormal relaxed molecular clock for each

gene and let the software estimate the rate from the priors. The MCMC chain ran for 100.000.000 steps, sampled every 10.000 steps. Posterior distribution of all the parameters were checked using Tracer, as well as all ESS values being above 200. We built the tree using Tree Annotator, using the Maximum Clade Credibility implemented method after discarding the first 25% samples as a burn-in.

References

- Ahrens, D., Schwarzer, J., & Vogler, A. P. 2014. The evolution of scarab beetles tracks the sequential rise of angiosperms and mammals. *Proceedings of the Royal Society B: Biological Sciences* 281, 20141470
- Bouckaert, R., et al. 2014. BEAST 2: a software platform for Bayesian evolutionary analysis. *PLoS Computational Biology* 10, e1003537.
- Darriba, D., Taboada, G. L., Doallo, R., & Posada, D. 2012. jModelTest 2: more models, new heuristics and parallel computing. *Nature methods* 9, 772-772.
- Hall, T.A. 1999. BioEdit: A User-Friendly Biological Sequence Alignment Editor and Analysis Program for Windows 95/98/NT. *Nucleic Acids Symposium Series* 41, 95-98.
- Sambrook, J., Fritsch, E. F., & Maniatis, T. 1989. Molecular cloning: a laboratory manual, 2nd ed. Cold Spring Harbor, New York: Cold Spring Harbor Laboratory Press
- Villalba, S., Lobo, J. M., Martín-Piera, F., & Zardoya, R. 2002. Phylogenetic relationships of Iberian dung beetles (Coleoptera: Scarabaeinae): insights on the evolution of nesting behavior. *Journal of molecular evolution* 55, 116-126.

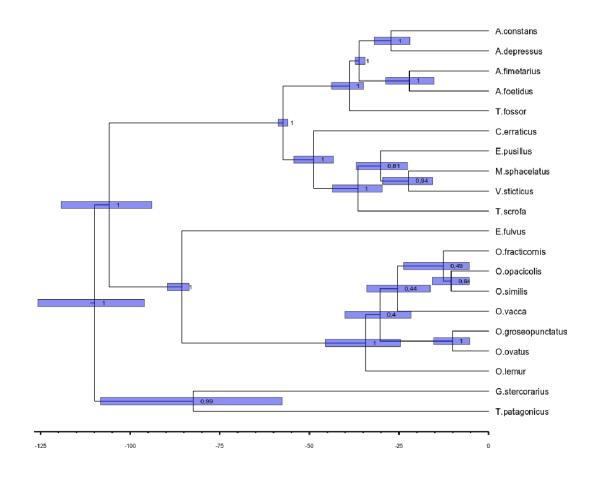


Fig. S1. Bayesian phylogenetic hypothesis for the studied species. Posterior probabilities are provided. Blue bars represent the 95% credible interval around node ages.

Species	28	COI	COII
Aphodius constans	-	AY039372	AY039372
Aphodius depressus	ABXXXX	ABXXXX	ABXXXX
Aphodius erraticus	ABXXXX	ABXXXX	ABXXXX
Aphodius fimetarius	ABXXXX	ABXXXX	ABXXXX
Aphodius foetidus	-	ABXXXX	ABXXXX
Aphodius fossor	ABXXXX	ABXXXX	ABXXXX
Aphodius pusillus	ABXXXX	ABXXXX	ABXXXX
Aphodius scrofa	ABXXXX	ABXXXX	ABXXXX
Aphodius sphacelatus	ABXXXX	ABXXXX	ABXXXX
Aphodius sticticus	-	ABXXXX	-
Euoniticellus fulvus	ABXXXX	ABXXXX	ABXXXX
Geotrupes stercorarius (OUT)	KP419463	AY039377	AY039377
Onthophagus fracticornis	ABXXXX	-	-
Onthophagus grossepunctatus	ABXXXX	AY039347	AY039347
Onthophagus lemur	ABXXXX	AY039353	AY039353
Onthophagus opacicollis	-	ABXXXX	ABXXXX
Onthophagus ovatus	ABXXXX	AY039351	AY039351
Onthophagus similis	ABXXXX	ABXXXX	ABXXXX
Ontophagus vacca	ABXXXX	AY039359	AY039359
Taurocerastes patagonicus (OUT)	KP419662	GU984611	GU984611

Table S1. GenBank accession numbers of the used sequences. Outgroup species are indicated.