Modeling of symbiotic bacterial biofilm growth with an example of the *Streptococcus-Veillonella* sp. system

Dianlei Feng[1,*], Insa Neuweiler[1], Regian Nogueira[2], Udo Nackenhorst[3]

1 Institute of Fluid Mechanics and Environmental Physics in Civil Engineering, Leibniz Universität Hannover, Germany
2 Institute for Sanitary Engineering and Waste Management, Gottfried Wilhelm Leibniz Universität Hannover, Germany
3 Institute of Mechanics and Computational Mechanics, Leibniz Universität Hannover, Germany

* feng@hydromech.uni-hannover.de

Abstract

We present a multi-dimensional continuum mathematical model for modeling the growth of a symbiotic biofilm system. We take a dual-species namely, the *Streptococcus - Veillonella* sp. biofilm system as an example for numerical investigations. The presented model describes both the cooperation and competition between these species of bacteria. The coupled partial differential equations are solved by using an integrative finite element numerical strategy. Numerical examples are carried out for studying the evolution and distribution of the bio-components. The results demonstrate that the presented model is capable of describing the symbiotic behavior of the biofilm system. However, homogenized numerical solutions are observed locally. To study the homogenization behavior of the model, numerical investigations regarding on how random initial biomass distribution influences the homogenization process are carried out. We found that a smaller correlation length of the initial biomass distribution leads to faster homogenization of the solution globally, however, shows more fluctuated biomass profiles along the biofilm thickness direction. More realistic scenarios with bacteria in patches are also investigated numerically in this study.

1   Introduction

More than 90% of microbes live in biofilms which can be defined as "assemblages of bacterial cells attached to a surface and enclosed in an adhesive matrix secreted by the cells"(Madigan, 2012). A biofilm *in natura* is usually found as a multi-component, multi-species, heterotopic matter with multi-phase properties. Competition and cooperation among species of bacteria are normally involved in multi-species biofilm systems (Yang et al., 2011). A biological system is symbiotic when cooperation happens between two different organisms.

Mathematical and numerical modeling is a powerful tool for understanding both the physical and bio-chemical processes during the formation and development of the biofilms. Many modeling strategies and models have been developed for describing biofilm processes. However, mathematical modeling of symbiotic biofilm systems has not been well studied. Modeling a symbiotic biofilm system naturally requires the

consideration of a multi-species biofilm problem which is usually challenging, especially for multi-dimensional scenarios.

Several studies on modeling multi-dimensional multi-species biofilm system have been carried out in the literature. (Noguera et al., 1999) presented a Cellular Automaton (CA) model for modeling a dual-species (*D. vulgaris-M. formicicum*) biofilm system. Recently, (Tang and Liu, 2017) developed a multi-species multi-dimensional CA model to study syntrophic and dissimilatory metal reducing bacterial bioiflm system. In their study, a syntrophic ecological system is considered. (Martin et al., 2017) studied a dual-species (*S. gordonii-P. gingivalis*) oral biofilm system with a further developed CA model. Three models on the relationship between these two species of bacteria, namely independence for substrates, competition for substrate, and inhibition of one to another, are compared in their study. Individual-based Modeling (IbM) have also been used for studying various multi-species biofilm problems, such as modeling ammonium oxidizer bacteria (AOB) and nitrite oxidizer bacteria (NOB) system (Picioreanu et al., 2004), modeling dormant cell formation (Chihara et al., 2015) process, investigation of denitrifying bacteria - sulfate reducing bacteria - methanogens biofilm systems (Martin et al., 2015) and so forth. These models are not continuum and are usually categorized as the so-called discrete element based models (Fujikawa, 1994; Kreft et al., 2001; Noguera et al., 1999; Picioreanu et al., 1998; Wimpenny and Colasanti, 1997).

Although the reaction model presented in this paper can also be applied with discrete element based models, we present our model within the framework of continuum biofilm models which are fully described by partial differential equations (Alpkvist and Klapper, 2007; Cogan, 2004; Duddu et al., 2009; Eberl et al., 2001; Klapper and Dockery, 2002; Lindley et al., 2012; Wanner and Gujer, 1986; Zhang et al., 2008b). (Alpkvist and Klapper, 2007) presented a multi-dimensional multi-species biofilm model and applied it for studying an autotrophs-heterotrophs-inert bifilm system. The growth of biofilm is modeled as an advective movement of a potential flow due to the production or reduction of biomass. Instead of modeling the biofilms grow advectively, (Rahman et al., 2015) presented a multi-species multi-dimensional continuum diffusion-reaction model. A cross-diffusion process is additionally introduced into the model to overcome the internal over-mixing problem. The over-mixing of different species of biomass is a common flaw in many diffusion-reaction continuum models as well as some discrete element based models for multi-species multi-dimensional biofilm modeling (Tang and Valocchi, 2013). Since the advection-reaction biofilm model presented by Alpkvist and Klapper (Alpkvist and Klapper, 2007) does not involve physical diffusion, it is expected to be less diffusive, although the numerical dissipation is not avoidable. However, when the reactions become complex, whether locally homogenized biomass distribution will be obtained or not is still an open question. In this paper, we look into such a problem by studying the presented symbiotic biofilm system with an advection-reaction type model.

We take an oral dual-species, namely the *Streptococcus-Veillonella* sp. biofilm system (Chalmers et al., 2008), as an example of modeling symbiotic biofilm systems. Both *Streptococcus* sp. and *Veillonella* sp. play vital roles in the formation of dental biofilms. Dental bacterial biofilms are often of special interested due to their roles in the treatments of dental implants and diseases (Kommerein et al., 2017; Paquette et al., 2006). Different from many other types of biofilms, the bacteria that form the dental biofilms need salivary glycoproteins to attach to the teeth surface. Meanwhile, saliva is also the main nutrient source for the microbes in oral cavities (Chalmers, 2008; Chalmers et al., 2008). Among the *Streptococcus* group bacteria, *Streptococcus gordonii* is a well-known Gram-Positive commensal bacterium in oral cavities which can cause caries and demineralization of teeth (Nascimento et al., 2009). *Streptococcus gordonii* is also the first batch of bacteria that attach to the teeth surface and is known as a species of early colonizers (Kolenbrander et al., 2010). The early colonizers can build up a

foundation for the later colonizers to attach to the biofilm structure. This makes them important for studying oral biofilms.

Similar to the *S. gordonii*, *Veillonella* sp. is also a group of early colonizers which congregate with *S. gordonii* (Periasamy and Kolenbrander, 2010). It has been discovered that the *Veillonella* sp. cannot grow alone in saliva. However, it colonizes when there is *Strptococcus* sp. (e.g. *S. gordonii*) coexisting in the system (Chalmers et al., 2008). Similar behavior has also been observed in (Kara et al., 2007; Mashima and Nakazawa, 2015). A biological interpretation of such phenomenon is that the *Veillonella* sp. cannot utilize the carbon source in saliva directly but can ferment the metabolic production of *S. gordonii*, namely lactic acid, instead (Periasamy and Kolenbrander, 2010). It is worth noting that the lactic acid not only provides a carbon source for the growth of the *Veillonella* sp. but also causes a decrease of the pH in the system. It has been experimentally observed that the *Veillonella* sp. can tolerate more acid than the *S. gordonii* (Bradshaw and Marsh, 1998). The presence of the *Veillonella* which utilizes the lactic acid may thus be beneficial to the growth of the *S. gordonii* indirectly.

In this study, we develop a mathematical model for the symbiotic *S. gordonii* - *Veillonella* biofilm system under the hypothesis that the lactic acid has a negative influence on the growth of *S. gordonii* and thus the symbiotic biofilm system studied in this paper is essentially a mutualistic ecological system. The flow velocity in oral cavities is known to be very small and does not have much impact on the growth of the *S. gordonii* biofilm within a typical characteristic period (12h-24h) (Rath et al., 2017). The mathematical model developed by Alpkvist and Klapper (Alpkvist and Klapper, 2007) has been validated for the *S. gordonii* biofilm (Feng et al., 2018; Rath et al., 2017). Many studies have reported that the *Streptococcus* sp. dominates in a multi-species biofilm, which consists of the *Streptococcus* sp. and *Veillonella* sp. (Chalmers et al., 2008; Kommerein et al., 2017; Periasamy and Kolenbrander, 2010). For this reason, we develop the symbiotic biofilm model based on the advection-reaction biofilm growth model.

In the model presented in this paper, the lactic acid, which is produced inside of the biofilm system, plays a vital role in the chain of the symbiotic bio-chemical process is considered explicitly. Despite the saliva cannot function as a carbon source for the *Veillonella* sp., saliva still may provide other nutrients for the growth of the *Veillonella* sp. due to its complex mixture substance property. Therefore, a competition of the saliva between *S. gordonii* and *Veillonella* sp. is considered in the model. The model presented in this paper can be used for quantitatively understanding of the *Streptococcus*-*Veillonella* symbiotic biofillm system and in general can be modified to model a variety of biofilm systems with similar bio-chemical properties. To the best of our knowledge, a continuum model for such kind of symbiotic biofilm systems has not been well developed. This study may shed light on understanding the numerical behaviors of multi-dimensional multi-species continuum (advection-reaction) biofilm models which involve complex reactions.

The presented biofilm model is highly non-linear. To better understand the solution behaviors, we investigated how the initial condition influences the solution. As the acid plays a vital role in the model, we also carry out numerical investigations on the acid transport equation's boundary conditions and the inhibition parameter.

This paper is structured as follows. Details of the model will be presented in section 2 (material and methods). 2D numerical simulations with different initial biomass distributions are carried out and compared in section 3 (results and discussion). Numerical investigations on how initial biomass distributions with different spatial correlation lengths are also carried out. Furthermore, more realistic scenarios with initial conditions of one species randomly embedded in another are studied. Simulation results of the distribution and evolution of the biomass and substrates are studied and

discussed. Then, the study is summarized in section 4. Information regarding the ¹¹⁸
dimensionless form of the governing equations, as well as the corresponding numerical ¹¹⁹
method, are presented in Appendix A and B. ¹²⁰

2   Mathematical model   ¹²¹

The *Streptococcus-Veillonella* biofilm system is illustrated in Figure 2. Two biomass ¹²²
components, namely the *S. gordonii* and *V. PK1910* interact with each other in a ¹²³
mutualistic relationship. Especially, *S. gordonii* produces lactic acid, which is the ¹²⁴
carbon source for the growth of *Veillonella*. On the other hand, *Veillonella* sp. helps to ¹²⁵
prevent the environment from being too acid, and thus benefits the growth of *S.* ¹²⁶
gordonii. We develop the mathematical model based on the following assumptions: ¹²⁷

- A layer of biomass has been presented as an initial condition. The model does not ¹²⁸
  capture the process in which individual bacteria grow into a biofilm. On the other ¹²⁹
  hand, the growth process after forming an initial layer of biofilm is modeled. ¹³⁰
  However, the initial layer can be sufficient thin. ¹³¹

- Two species of bacteria, namely the *Streptococcus* sp., and *Veillonella* sp. are ¹³²
  considered in the model. The volume fractions of the bacteria sum to one inside of ¹³³
  the biofilm, which implies that the pore space and EPS are not explicitly modeled ¹³⁴
  (Alpkvist and Klapper, 2007). ¹³⁵

- Saliva is considered as a common nutrient source consumed by both species of ¹³⁶
  bacteria. However, the growth of *Veillonella* is limited by both the nutrient and ¹³⁷
  the lactic acid, which is produced by *Streptoccus* sp (Chalmers et al., 2008). ¹³⁸

- The concentration of hydrogen is not modeled explicitly, but the concentration of ¹³⁹
  the lactic acid is considered instead as an indicator for it (Gordeeva et al., 2017). ¹⁴⁰

- The fluid-structure interaction process is not considered. This assumption is based ¹⁴¹
  on our previous experimental observations of the formation process of *S. gordonii* ¹⁴²
  (Rath et al., 2017). ¹⁴³

- No new void spaces of bacteria can develop during the growth of the biofilm. In ¹⁴⁴
  other words, the biofilm domain is simply connected (see Figure 1). ¹⁴⁵

- The nutrients in saliva are fully mixed with a distance above the biofilm. A ¹⁴⁶
  Dirichlet boundary condition of the nutrients thus exists (Picioreanu et al., 2006). ¹⁴⁷

- The growth velocity is described by an irrotational vector field (Klapper and ¹⁴⁸
  Dockery, 2002). ¹⁴⁹

- Diffusion coefficients are constant both inside and outside of the biofilm (Alpkvist ¹⁵⁰
  and Klapper, 2007). ¹⁵¹

To make the model simple, we introduced assumptions more than necessary. One ¹⁵²
can include the EPS as an additional biomass component in the model, as in (Xavier ¹⁵³
et al., 2005). It is also convenient to assign heterogeneous modeling parameters, such as ¹⁵⁴
diffusion coefficient, density, or even growth rate in the model. The pH can also be ¹⁵⁵
modeled by slightly extending the model, as presented in (Khassehkhan and Eberl, ¹⁵⁶
2008). ¹⁵⁷
   However, main limitation of the presented model is the requirement of being simply ¹⁵⁸
connected (biofilm domain) during biofilm growth. This is mostly due to the difficulty ¹⁵⁹
of enforcing proper potential (or pressure) boundary conditions for the newly generated ¹⁶⁰

(a) simply connected

(b) not simply connected

Figure 1. Typical biofilm patterns generated during the merge of two colonies.

biofilm-fluid surface. Such a problem may come up when trying to model the merge of different colonies. As shown in Figure 1, the model fails to model the merge of multiple colonies if the biofilm can not maintain simply connected. As illustrated in Figure 1b, when there is a new biofilm-fluid interface generated during the merge of colonies, it is not straightforward to give boundary conditions of the potential on Γ_1 and Γ_2. On the other hand, the model can be used to model the merge of different colonies if only the scenario, as shown in Figure 1a, occurs. Modeling the merge of colonies is beyond the aim of this paper. For the readers interested in this topic, we want to refer to the three-dimension numerical study in (Alpkvist and Klapper, 2007). 161 162 163 164 165 166 167 168 169

To build up the model based on the above assumptions, the mass balance of each substrate and component of biomass in the biofilm is considered. Eventually, the mathematical model, which is essentially a free boundary problem, consists of five partial differential equations (PDEs) including four coupled nonlinear time dependent advection-diffusion-reaction (or advection-reaction) equations and a Poisson's equation. 170 171 172 173 174

Figure 2. Illustration of the symbiotic biofilm system. The *Veillonella* sp. PK1910 and *Streptococcus gordonii* are taken as example species of the bacteria in the illustration. Both species of bacteria compete for the nutrients in saliva. The *Streptococcus gordonii* produces lactic acid, which is a carbon source for *Veillonella* sp. The lactic acid also prevents the growth of *gordonii*.

Taking a 2D model as an example, the biofilm is modeled within a computational 175

domain of $\Omega = \{\mathbf{x} = (\mathbf{x}, \mathbf{z}) : \mathbf{0} \le \mathbf{x} \le \mathbf{W}, \mathbf{0} \le \mathbf{z} \le \mathbf{H}\}$ with the corresponding boundary $\partial\Omega$. The top border is denoted as Γ_{ht} and the other sides are denoted by $\Gamma_{s1} = \partial\Omega \not\supseteq \Gamma_{ht}$. The outward-pointing normal vectors of Γ_{s1} and $\partial\Omega$ are denoted as $\mathbf{n}_{s1}$ and $\mathbf{n}_{s2}$ respectively. A sub-domain $B_t \in \Omega$ refers to the biofilm domain above which a bulk fluid domain F_t is assumed. The biofilm-fluid interface $\Gamma_{int} = B_t \cap F_t$, which is a free boundary, moves over time induced by the growth of the biofilm.

Figure 3. Two dimensional illustration of the computational domains of the symbiotic biofilm model.

2.1 Mass balance of biomass

The growth of both *S. gordonii* and *Veillonella* are modeled as advective movements driven by the production and reduction of biomass inside of the biofilm. Growth velocities of different components of biomass are assumed to be the same as the biofilm growth velocity **u**. We also assume the density of each bio-component is constant. Writing the mass balance equations of *S. gordonii* and *Veillonella* sp. in terms of their volume fraction ϑ reads

$$\frac{\partial \vartheta_1}{\partial t} + \nabla \cdot (\mathbf{u}\vartheta_1) = \frac{g_1}{\rho_s}, \quad \mathbf{x} \in \Omega, \tag{1a}$$

$$\frac{\partial \vartheta_2}{\partial t} + \nabla \cdot (\mathbf{u}\vartheta_2) = \frac{g_2}{\rho_v}, \quad \mathbf{x} \in \Omega, \tag{1b}$$

where $\mathbf{u}\ [LT^{-1}]$ denotes the growth velocity of the biofilm, indexes "1" and "2" are used to refer physical quantities of the *S. gordonii* and *Veillonella* bacteria respectively and, ρ_s and ρ_v denote the density of the corresponding biomass. In equation (1), g refers to the mass production/reduction rate of the biomass. A no flux boundary is applied for both equation (1a) and equation (1b)

$$\mathbf{n}_{s2} \cdot (\mathbf{u}\vartheta_i) = 0, \quad (i = 1, 2), \quad \mathbf{x} \in \partial\Omega. \tag{2}$$

It has been experimentally found that the *S. gordonii* can tolerate less acid than the *Veillonella* sp. and the *Veillonella* sp. may become dominant even in the environment with pH < 5.0 (Bradshaw and Marsh, 1998). The growth of *S. gordonii* is sensitive to the acid and, on the other hand, *Veillonella* can grow in small a pH environment. Therefore, we only model the negative influence of lactic acid on the growth of *S. gordonii*. The mass production process of *S. gordonii* is limited by the concentration of saliva (s_1) and suppressed by the lactic acid (s_2)

$$g_1 = \vartheta_1 \rho_s \mu_1 \frac{s_1}{k_{11} + s_1} \frac{k_l}{k_l + s_2}. \tag{3}$$

In equation (3), μ_1 $[T^{-1}]$ and k_{11} $[ML^{-3}]$ refer to the maximum growth rate and the half-rate constant of the *S. gordonii* in the Monod kinetic respectively. We introduce a parameter k_l $[ML^{-3}]$ for modeling the suppression process induced by the lactic acid. A benefit of choosing g_1 in the form as equation (3) is, when $s_2 \to +\infty$, $g_1 \to 0$. This means the growth of the *S. gordonii* stops when the concentration of the lactic acid is infinitely large.

The growth of *Veillonella* is limited by both the concentration of the nutrients in saliva s_1 and the concentration of the lactic acid s_2

$$g_2 = \vartheta_2 \rho_v \mu_2 \frac{s_1}{k_{21} + s_1} \frac{s_2}{k_{22} + s_2}. \tag{4}$$

In equation (4), μ_2 $[T^{-1}]$ denotes the maximum growth rate of *Veillonella* and k_{21} $[ML^{-3}]$ and k_{22} $[ML^{-3}]$ are the half-rate constant in the Monod kinetic.

2.2   Mass balance of saliva and lactic acid

The chemical reactions of components in the saliva environment are very complex, and the corresponding reaction parameters are mostly unknown. Instead of describing detailed chemical reactions of all involved chemical components in saliva, we consider the saliva itself as one substance demanded by both species of bacteria. Since the time scale of the transport process of the substrate (solutions) is much smaller than the time scale of the biofilm growth (Picioreanu et al., 2000), the mass balance of the saliva is modeled as a stationary diffusion-reaction process as

$$-D_1 \nabla^2 s_1 = r_{11} + r_{12}, \quad \mathbf{x} \in \Omega, \tag{5}$$

where r_{11} and r_{12} $[\text{ML}^{-3}\text{T}^{-1}]$ denote the saliva (nutrients in saliva) consumption rates by *S. gordonii* and *Veillonella* sp. respectively. D_1 $[L^2T^{-1}]$ denotes the diffusion coefficient of the saliva.

The nutrient in saliva comes from the outside of the modeling domain Ω. A constant concentration $\bar{s}_1$ is set at the top of the computational domain Γ_{ht} (as shown in Figure 3)

$$s_1 = \bar{s}_1, \quad \mathbf{x} \in \Gamma_{ht}. \tag{6}$$

A no flux boundary is applied on the other borders of the computational domain as

$$\mathbf{n}_{s1} \cdot \nabla s_1 = 0, \quad \mathbf{x} \in \Gamma_{s1}, \tag{7}$$

where $\mathbf{n}_{s1}$ denotes the normal vector of Γ_{s1}.

The reaction term corresponding to the consumption of saliva by *S. gordonii* is described by using the Monod kinetics and reads

$$r_{11} = -\frac{1}{Y_{s1}} g_1, \tag{8}$$

where $Y_{s1}[-]$ denotes the corresponding yield of the *S. gordonii* by consuming rate of the saliva. Similarly, the consumption rate of the saliva due to the *Veillonella* reads

$$r_{12}(s_1, s_2, \vartheta_2) = -\frac{1}{Y_{v1}} g_2, \tag{9}$$

where Y_{v1} $[-]$ is the yield of the *Veillonella*.

The lactic acid which is secreted by the *S. gordonii* accumulates over time in the system. If a batch system is considered, the lactic acid is produced inside of the biofilm domain B_t and a fully mixed domain in the fluid does not exist, which is different from

the nutrients (saliva). In this case, the mass balance of the lactic acid cannot be modeled as a quasi-steady state since the absence of the Dirichlet boundary condition. On the other hand, in a flow chamber s For these reason, the mass balance of the acid has to be modeled as a time dependent process. The mass transport of acid is diffusion-reaction dominated. Therefore, one can simplify the model by neglecting the advection process and achieve comparable, accurate results. Thanks to the A-stability property of the presented space-time numerical method, there is no difficulty in solving the mass balance equation of the acid together with other time-dependent PDEs in the model by considering the advection process. In this paper, we model the advection process of the acid since it does not increase the computational expense and can improve the accuracy of the model as well. Moreover, the production and consumption of the lactic acid by *S. gordonii* and *Veillonella* sp. are also considered respectively. The mass balance of the lactic acid reads

$$\frac{\partial s_2}{\partial t} + \nabla \cdot (\mathbf{u} s_2) - D_2 \nabla^2 s_2 = r_{21} + r_{22}, \quad \mathbf{x} \in \Omega, \tag{10}$$

where D_2 $[L^2 T^{-1}]$ is the diffusion coefficient of the lactic acid. r_{21} $[ML^{-3}T^{-1}]$ and r_{22}
$[ML^{-3}T^{-1}]$ describe the production of the lactic acid by *S. gordonii* and the
consumption rate of the acid by the *Veillonella* respectively.

    S. gordonii grows by consuming saliva and producing lactic acid at the same time. Thus the reaction term r_{21} can be written as

$$r_{21} = \lambda g_1, \tag{11}$$

where λ $[-]$ is a parameter which describes the mass production rate of acid by unit mass of *S. gordonii*. Only the *Veillonella* sp. in the system consumes lactic acid, therefore, the reaction kinetic can be written as

$$r_{22}(s_1, s_2, \vartheta_2) = -\frac{1}{Y_{v2}} g_2, \tag{12}$$

where Y_{v2} refers to the yield of *Veillonella* by consuming the lactic acid. A no flux boundary is applied on $\partial\Omega$ for equation (10) as

$$\mathbf{n}_{s2} \cdot (\mathbf{u}\nabla s_2 - D_2 \nabla s_2) = 0, \quad \mathbf{x} \in \partial\Omega, \tag{13}$$

where $\mathbf{n}_{s2}$ denotes the normal vector on $\partial\Omega$.

    As a remark, we applied a no flux boundary condition in the mass balance equation of the saliva. Such a boundary condition is simple, but it implies that a batch system is considered here. The lactic acid will accumulate in the domain over time, and the acid concentration can be high. Meanwhile, it would be more accurate to write (5) as a time-dependent equation (Wanner and Reichert, 1996) and lose the corresponding Dirichlet boundary condition. In this study, the saliva is sufficient in the system, and the gradient of it plays a more important role than its concentration far away from the biofilm. For this reason. as well as for saving computation resource, we applied a quasi-steady state saliva model in this paper.

    For flow chamber systems, a Robin boundary condition which allows the acid leaves the domain, as presented in (Ghasemi et al., 2018), can be very useful. Also, a diffusive boundary layer shall be considered for modeling the saliva transport (Feng et al., 2017). However, both of these boundary conditions require additional parameters. In this paper, we only apply simple boundary conditions, as presented in equation (6) and (13). However, one can easily choose the boundary conditions that needed in specific studies.

2.3 Potential equation for the movement of the biofilm

We model the movement of the biofilm as a potential flow driven by the production of the biomass inside of the biofilm following (Alpkvist and Klapper, 2007; Klapper and Dockery, 2002). It is assumed that the two species of bacteria compose the whole biofilm. Therefore, the volume fractions of the *S. gordonii* and *Veillonella* sum up to one in the biofilm domain B_t as

$$\sum_{i=1}^{2} \vartheta_i = 1, \quad \mathbf{x} \in \partial B_t. \tag{14}$$

Following the previous potential flow assumption, the biofilm growth velocity is described as the gradient of the potential Φ. Substituting equations (1a) and (1b) into equation (14) yields the potential equation

$$\nabla \cdot \mathbf{u} = \nabla^2 \Phi = \frac{g_1}{\rho_s} + \frac{g_2}{\rho_v}, \quad \mathbf{x} \in B_t, \tag{15a}$$

$$\Phi = 0, \quad \mathbf{x} \in \Gamma_{\text{int}}, \tag{15b}$$

$$\mathbf{n}_\Phi \cdot \nabla \Phi = 0, \quad \mathbf{x} \in \Gamma_\Phi. \tag{15c}$$

In equation (15), $\mathbf{n}_\Phi$ denotes the norm vector of Γ_Φ. Additionally, $\mathbf{u} = 0$ in the domain outside of the biofilm B_t.

2.4 Dimensionless form of the governing equations

Equations (1), (5), (10) and (15a) together with their corresponding boundary conditions compose the full mathematical model which describes the symbiotic biofilm system illustrated in Figure 2.

To obtain the dimensionless form of the governing equations, we introduce the following dimensionless variables

$$\mathbf{X} = \frac{\mathbf{x}}{H}, \qquad T = \frac{t}{t_d}, \qquad \chi = \frac{t_d \Phi}{H^2}, \qquad S_1 = \frac{s_1}{\bar{s}_1},$$

$$S_2 = \frac{s_2}{\bar{s}_1}, \qquad D_2^* = \frac{D_2 t_d}{H^2}, \qquad \Theta_{11}^2 = \frac{H^2 \rho_s \mu_1}{Y_{s1} D_1 \bar{s}_1}, \qquad \Theta_{12}^2 = \frac{H^2 \rho_v \mu_2}{Y_{v1} D_1 \bar{s}_1},$$

$$\Theta_{21}^2 = \frac{\lambda \rho_s \mu_1 t_d}{\bar{s}_1}, \qquad \Theta_{22}^2 = \frac{\rho_v \mu_2 t_d}{\bar{s}_1 Y_{v2}}, \qquad \Psi_1 = \mu_1 t_d, \qquad \Psi_2 = \mu_2 t_d, \tag{16}$$

$$K_{11} = \frac{k_{11}}{\bar{s}_1}, \qquad K_{21} = \frac{k_{21}}{\bar{s}_1}, \qquad K_{22} = \frac{k_{22}}{\bar{s}_1}, \qquad K_L = \frac{k_l}{\bar{s}_1}.$$

In equation (16), H and t_d are characteristic length and time which are taken as the height of the computational domain and the time of a day respectively in this study.

Substituting equation (16) into equations (1), (5), (10) and (15a) yields the dimensionless form of the governing equations. The detailed dimensionless governing equations are listed in A.

3 Results

3.1 Biofilm growth

Two-dimensional simulations are carried out in this section with the modeling parameters listed in Table 1. A wave shape initial biofilm-fluid interface in the

dimensionless domain is applied

$$\Gamma^0_{\text{int}} :\to Z = 0.1 + 0.05 \sin\left(4\pi X + \frac{\pi}{2}\right), \quad \mathbf{X} \in \Omega^*. \tag{17}$$

Table 1. Parameters used for two-dimensional symbiotic biofilm simulation

Quantity name	Symbol	Value	Unit	Source
Length of the computational domain	W	300	μm	(Alpkvist and Klapper, 2007)
Height of the computational domain	H	300	μm	(Alpkvist and Klapper, 2007)
Saliva concentration at the top boundary	$\bar{s}_1$	10	kg/m^3	estimated
Diffusion coefficient of saliva	D_1	5×10^{-10}	m^2/s	(Corbin et al., 2011)
Diffusion coefficient of lactic acid	D_2	1×10^{-9}	m^2/s	(Ribeiro et al., 2005)
Density of *S. gordonii*	ρ_s	1100	kg/m^3	estimated from (Duddu et al., 2009)
Density of *Veillonella*	ρ_v	1100	kg/m^3	estimated from (Duddu et al., 2009)
Biofilm yield of *S. gordonii* by consuming saliva	Y_{s_1}	10^{-1}	$[-]$	estimated
Biofilm yield of *Veillonella* by consuming saliva	Y_{v_1}	10^{-1}	$[-]$	estimated
Biofilm yield of *Veillonella* by consuming lactic acid	Y_{v_2}	10^{-1}	$[-]$	**estimated**
Production coefficient of lactic acid	λ	5×10^{-1}	$[-]$	estimated
Maximum growth rate of *S. gordonii*	μ_1	3×10^{-5}	s^{-1}	(Rath et al., 2017)
Maximum growth rate of *Veillonella*	μ_2	8×10^{-5}	s^{-1}	estimated
Monod half-rate constant	k_{11}	0.878	kg/m^3	(Rath et al., 2017)
Monod half-rate constant	k_{21}	0.878	kg/m^3	estimated
Monod half-rate constant	k_{22}	0.0878	kg/m^3	estimated
Monod half-rate constant	k_l	1	kg/m^3	estimated

Since there is *S. gordonii* in the system initially, there should also be lactic acid presented in the system. Ideally, the initial condition shall be determined based on experimental measurements, and the initial condition influences the numerical solutions. In this study, we aim to investigate the reactive dynamics of the symbiotic biofilm system instead of a specific experimental scenario. In the following simulations, we apply an *ad hoc* initial condition for S_2 by assuming no lactic acid in the system. Also, in this numerical example, we consider that these two species of bacteria are initially evenly mixed as

$$S_2\left(\mathbf{X}, t = 0\right) = 0, \quad \mathbf{X} \in \Omega^*, \tag{18a}$$
$$\vartheta_1\left(\mathbf{X}, t = 0\right) = \vartheta_2\left(\mathbf{X}, t = 0\right) = 0.5, \quad \mathbf{X} \in \Omega^*. \tag{18b}$$

A 200×200 spatial mesh with four-node bilinear elements is applied and the time step size is taken as $\Delta t = \frac{1}{200}$ [day]. The TDG-FIC scheme (see B) with a 3rd order time accuracy is used in the study.

(a) **(b)**

(c) **(d)**

Figure 4. Simulation results of the symbiotic biofilm system at different times (from top to bottom are at $t = 9.6$h$(T = 0.4)$ and $t = 48$h$(T = 2.0)$). Left column: Volume fraction distributions of *S. gordonii* and *Veillonella*; Right column: Volume fraction profiles of *S. gordonii* and *Veillonella* and total biomass in depth at $X = 0.5$.

3.1.1 Biomass evolution and distribution

Simulation results of biomass distribution of both *S. gordonii* and *Veillonella* at early and late times are shown in Figure 4. Since the numerical solutions are symmetric regarding the middle line $X = 0.5$, only the half plane of a solution of each species of biomass is shown in the figure. The white curve in Figure 4 represents the biofilm-fluid interface. The simulation results demonstrate that a flat biofilm pattern is generated after the growth of the biofilm for a certain time. This is due to the nutrients are not scarce in the system. However, we still can observe that the waved biofilm-fluid interface has quite an impact on the local distribution of the lactic acid (see later discussions).

Biomass volume fraction profiles of *S. gordonii*, *Veillonella* and total biomass at different times along $X = 0.5$ are shown in the right column in Figure 4. It is observed that *S. gordonii* moves up as a result of winning the competition for nutrients in saliva in the vicinity of the biofilm-fluid interface. Such a phenomenon has also been observed in experimental studies (Chalmers, 2008) as shown in Figure 5. In the figure, *S. gordonii* and *Veillonella* PK1910 are colored in green and blue respectively. We observe that the *S. gordonii* locates above the *Veillonella* in the areas within the red squares in Figure 5. As a remark, we are not trying to reproduce the biofilm pattern shown in Figure 5 in this paper. The pattern of biofilms highly depends on nutrients conditions. A diluted saliva (Chalmers, 2008) is used as a nutrient in the experiments corresponding to Figure 5. We applied the calibrated model parameters corresponding to the growth of *S. gordonii* from our previous experimental study(Rath et al., 2017). The *S. gordonii* is cultured in the Tryptic Soy Broth medium instead of the diluted saliva. Under such a condition, we find the *S. gordonii* build up a biofilm of layering pattern (Kommerein et al., 2017; Rath et al., 2017). It has been observed that the *Streptococcus* sp.

dominates when they are coaggregated with the *Veillonella* sp. Therefore, all numerical 258
examples in this paper are set up with layered initial conditions.

Figure 5. Spatial distribution of *S. gordonii* (in green) and *Veillonella* PK1910 (in blue) observed with confocal laser scanning microscopy in 2h flowcell biofilms (Chalmers, 2008).

It should be noted that the simulation results correspond to an evenly mixed initial 260
biomass volume condition. In nature, it is observed that *S.gordonii* usually attaches to 261
a surface first and *Veillonella* attaches to *Streptococcus* afterwards. They form the 262
so-called "corn cob" pattern (Kolenbrander et al., 2010) *in vitro*. This demonstrates 263
that the initial condition of either the geometry of the biofilm-fluid interface or the 264
biomass distribution in such a symbiotic biofilm system is heterogeneous. To study such 265
complex heterogeneity properties that are influenced by the initial conditions, an 266
attachment process with considerations of bacteria receptors has to be considered. The 267
attachment process normally happens at an individual bacterium scale and can hardly 268
be considered only with a continuum model. To capture the attachment process, 269
modeling the whole biofilm system with a multi-scale model will be promising. However, 270
this is beyond the scope of this paper. Furthermore, Egland *et al.* presented 271
experimental evidence of signaling between *S. gordonii* and *V. atypica* (Egland et al., 272
2004). The diffusive short distance signaling induced by the quorum sensing (QS) might 273
also another key reasons that lead to the "corn cob" pattern of the biofilm system. If 274
the goal is to reproduce the "corn cob" spatial pattern, a QS biofilm model such as 275
(Emerenini et al., 2015) could be considered. However, the detailed QS mechanism is 276
still unclear (Egland et al., 2004) and may result in many uncertainties in the model. 277

The inhibition parameter k_l has a significant influence on the growth of both species 278
of bacteria. According to (3), the lactic acid does not prevent the growth of *S. gordonii* 279
when k_l is an infinitely large number. In this case, one would expect the *S. gordonii* 280
grows at a large rate. As shown in Figure 6, we observe a faster growth of both species 281
of bacteria by using a larger value of k_l. As a remark, it is not surprised to see that the 282
Veillonella sp. also grows faster associatively with the *S. gordonii*. This is due to the 283
mutualistic relationship between these two species. 284

3.1.2 Lactic acid
285

Distribution of the lactic acid in terms of its dimensionless concentration S_2 at different 286
times is presented in Figure 7. The lactic acid is produced by the *S. gordonii* and 287
consumed by the *Veillonella*. The acid is transported diffusively and accumulates in the 288
fluid domain. It is observed that a higher concentration of acid accumulates above the 289
biofilm-fluid interface, especially, after the *S. gordonii* becomes dominant at the top 290
layer of the biofilm. Since there is more *S. gordonii* at the valley of the biofilm surface 291

(a) Volume fraction evolution profiles of *Strep-tococcus* sp. with different inhibition parameter k_l.

(b) Volume fraction evolution profiles of *Veil-lonella* sp. with different inhibition parameter k_l.

Figure 6. Sensitivity analysis of the inhibition parameter k_l on the growth of bacteria.

at $T = 1.0$ and $T = 1.6$, two punctiform sources of the acid at the biofilm valley are observed. This is why we observe the maximum acid concentration appears in the middle of the computational domain, and the acid diffuses to the top boundary. However, the gradient of S_2 gets smaller in the fluid domain over time due to acid accumulation. Such a process leads to a fully mixed condition later (e.g., at $T = 2.0$). If a Dirichlet boundary of S_2 is applied, such a fully mixed domain will not appear (See later discussion). The *Veillonella* wins the competition at the lower part of the biofilm (see biomass profiles in Figure 4) where the lactic acid is largely consumed by the *Veillonella*. What should be noted is that the distribution of the lactic acid is also influenced by the biofilm patterns and the generation of biofilm patterns might be influenced by the distribution of the lactic acid as well.

Figure 8 shows profiles of the lactic acid at $X = 0.5$ at different times. For each plot, the concentration of the lactic acid increases from the bottom of the biofilm to the top of it and decreases slightly afterwards in the bulk fluid. The profiles demonstrate that the location of the maximum concentration moves together with the biofilm interface. The profiles also show that the lactic acid accumulates in the bulk fluid domain over time. Comparing the two profiles of $t = 1.6$ day$(T = 1.6)$ and $t = 2.0$ day$(T = 2.0)$, the difference of the acid concentrations at the bottom of the biofilm is rather small. It seems the solution reaches a steady state locally there.

Since we apply a no-flow boundary to model the acid transport in a batch system. The acid accumulates in the fluid domain over time. If the aim is to model the biofilm growth in a flow system, one can apply a Dirichlet boundary for s_2. By assuming the acid is removed immediately at the top of the computation domain, equation (13) can be replaced by

$$s_2 = 0, \quad \mathbf{x} \in \partial\Omega. \tag{19}$$

As shown in Figure 9, the lactic acid profiles are different than the profiles shown in Figure 8. In this case, the acid does not accumulate as much as in the case with no-flow boundary condition. In Figure 10, the profiles of the lactic acid at $T = 0.4$, $T = 1.6$ and $T = 2.0$ by using no-flow and Dirichlet boundaries are compared. At an early time, the difference in acid concentration is small in the biofilm, which results in comparable biofilm thickness. However, the influence of the boundary condition increases in time. The biofilm grows slower by using the Dirichlet boundary condition as (19) comparing to using the no-flow boundary. The dimensionless thickness of the biofilm reduces 3% at $T = 1.6$ by using the Dirichlet boundary. At $T = 2.0$, the reduction is increased to 6%.

(a) $T = 0.4$

(b) $T = 1.0$

(c) $T = 1.6$

(d) $T = 2.0$

Figure 7. Distribution of lactic acid of the symbiotic biofilm system at different times ($T = 0.4$, 1.0, 1.6, 2.0). The white curve in the figures denotes the biofilm-fluid interface.

The lactic acid plays a twofold role in biofilm growth. On the one hand, it is the carbon source for *Veillonella*. On the other hand, the acid inhibits the growth of *Streptococcus*. In this study, we applied a relatively large value of k_l (see Table 1), which reduces the second effect. Since the acid does not accumulate in the bulk fluid when the Dirichlet boundary is applied, the acid concentration is smaller, especially at a late time (e.g., the peak value of S_2 at $T = 2.0$). Therefore, biofilm grows slower. As a remark, by using a smaller inhibition parameter k_l, one may obtain a different solution behavior.

In most of continuum multi-species biofilm models, the mass transport of substrates is considered as a quasi-steady-state process (Alpkvist and Klapper, 2007; Merkey et al., 2009). In this study, we model the intermediate product (lactic acid) additionally as a transient advection-diffusion-reaction process and solve it together with the mass balance of the two species of bacteria. The lactic acid transport equation, which is diffusion and (or) reaction dominated, has very different mathematical properties than the other two time-dependent equations (hyperbolic PDEs). We manage to solve these three coupled equations monolithically benefiting from the robust numerical framwork which is based on the space-time finite element method (see Appendix B).

The simulation results of this 2D example shed a light on the properties of the multi-dimensional numerical solutions of the presented biofilm model. First, heterogeneous biomass distribution is obtained with this model even though a fully mixed initial condition is applied. On the other hand, the biomass is homogenized locally in the biofilm. Second, the biofilm-fluid interface is flatted in time. However, the non-flat interface has an influence on the lactic acid distribution. Based on those observations, we ask ourselves: 1) Will the local homogenization process always happen in this model regardless of the initial biomass distribution? If so, will the initial distribution influence the homogenization process? 2) How will a heterogeneous initial biomass distribution influence the morphology of the biofilm interface as well as the components (including the biomass and the lactic acid) distributions during the growth?

Figure 8. Dimensionless concentration profiles of lactic acid at $X = 0.5$ at different times ($T = 0.4$, 1.0, 1.6, 2.0) with a no-flow boundary condition.

Figure 9. Dimensionless concentration profiles of lactic acid at $X = 0.5$ at different times ($T = 0.4$, 1.0, 1.6, 2.0) with a Dirichlet boundary condition.

3.2 Sensitivity to initial biomass distribution

In order to answer these questions raised in the previous section, we study the sensitivity of the presented mathematical model to different types of initial biomass distributions. As a matter of fact, the initial distribution of different species of bacteria in a biofilm can be never evenly mixed in nature. The numerical setups presented in this section may better reflect the properties of real biofilm systems. Two categories of initial biomass distributions are considered. The first type is a random biomass distribution with different spatial correlation lengths (Cirpka, 2003), and the second one is by considering embedded biomass initial distributions (as discussed in the following sections).

As shown in Figure 11, we consider three cases of the random initial distribution of biomass. Random fields are generated by using given correlation lengths. The correlation length measures the distance of two correlated points (e.g., the points with correlated material properties or physical quantities). Thus, the correlation length is often seen as a measure of the roughness of surfaces. In this study, to generate different random fields, different correlation lengths of $2\Delta x$, $4\Delta x$, and $8\Delta x$ are applied in each case, in which $\Delta x = \frac{1}{200}$ denotes the length of a spatial element in the horizontal direction. As a remark, the total initial mass (or volume) of each species of bacteria is the same.

To simplify the problem, the initial biofilm-fluid interface is modeled as a flat line, and a constant initial biofilm thickness of 60 μm ($Z = 0.2$) is assumed. Following our

Figure 10. Comparison of lactic acid profiles at $X = 0.5$ with different boundary conditions at $T = 0.4$, 1.6 and 2.0.

previous numerical setup, we assume that there is no lactic acid in the system initially.   368

As shown in Figure 12, the *Veillonella* patches are stretched and thus generating plume-like structures during the biofilm formation in all three cases. This indicates that processes happen mainly in the vertical direction, and the horizontal movement is not obvious. By increasing the correlation length of the initial biomass distribution, the size of the plumes increases. This results in that the biomass volume fraction profiles along a vertical line, for instance at $x = 150$ μm ($X = 0.5$) as shown in Figure 13, get less fluctuated with a larger correlation length. However, this does not mean that with a larger correlation length, the biomass homogenizes faster. In order to study how different biomass initial distributions influence the homogenization time, we compare the numerical solutions corresponding to the random initial distributions to the solution with an evenly mixed initial biomass distribution. Since the simulation involves moving domains, measuring the homogenization process is nontrivial. We calculate the difference of the volume fractions of the *Veillonella* between the evenly mixed and randomly distributed cases to measure the homogenization process

$$M_{\mathrm{diff},i} = \frac{1}{N} \left\| \boldsymbol{\vartheta}_2 \big|_{\mathrm{mixed}} - \boldsymbol{\vartheta}_2 \big|_i \right\|, \qquad (20)$$

where $\boldsymbol{\vartheta}_2$ is the solution vector of the volume fraction of *Veillonella*, N is the number of   369
the unknowns in the vector and indexes "mixed" refers to the evenly mixed initial   370
condition. The index $i = 1, 2, 3$ refers to scenarios with different correlation lengths of   371
$2\Delta x$, $4\Delta x$ and $8\Delta x$. Since the evenly mixed case always presents a locally homogenized   372
solution which produces a comparable biofilm thickness evolution as the other cases at   373
an early time, a smaller M_{diff} corresponds to a solution which is more homogenized.   374

The evolution of M_{diff} corresponding to these three scenarios is illustrated in Figure   375
14. The results demonstrate that the numerical solutions are homogenized in all these   376
three cases as soon as the simulation starts. The case corresponding to a smaller   377
correlation length requires less time to reach the state of a given value of M_{diff} at the   378
early time. In other words, the system homogenizes faster when the correlation length is   379
smaller. However, as mentioned earlier, more fluctuated vertical biomass profiles are   380
observed corresponding to the scenarios with larger correlation length. Moreover, M_{diff} 381
increases at a later time (as shown in Figure 14). This is due to even though the initial   382
biomass in the homogeneous case is the same as these three heterogeneous cases, the   383
distribution of the lactic acid is different from case to case which results in different   384
biomass production in the heterogeneous cases than in the homogeneous case. Such a   385
difference may accumulate over time and leads to an increase in the measured mass   386
difference M_{diff} at late time. In a word, the presented heterogeneous system will not   387

(a) correlation length: $2\Delta x$

(b) correlation length: $4\Delta x$

(c) correlation length: $8\Delta x$

Figure 11. Initial random volume fraction distribution of *Veillonella* with different correlation lengths.

Figure 12. Volume fractions of *Veillonella* at different times corresponding to different initial distribution with various corrlation lengths. Figure from left to right refer to simulation results corresponding to different initial correlation lengths of $2\Delta x$, $4\Delta x$ and $8\Delta x$. Figure from top to bottom refer to simulation results at different times of $t = 0.1$ day ($T = 0.1$), $t = 0.5$ day ($T = 0.5$) and $t = 1.5$ day ($T = 1.5$).

(a) $t = 0.1$ day

(b) $t = 0.5$ day

(c) $t = 1.5$ day

Figure 13. Volume fraction profiles of *Veillonella* at $X = 0.5$ at $t = 0.1$ day ($T = 0.1$; early time), $t = 0.5$ day ($T = 0.5$; middle time) and $t = 1.5$ day ($T = 1.5$; late time).

Figure 14. Homogenization times of three scenarios corresponding to initial biomass distribution with different correlation lengths $2\Delta x$, $4\Delta x$ and $8\Delta x$.

(a) case a)   **(b)** case b)

Figure 15. Two scenarios of initial biomass distribution (blue:bulk fluid; purple: *S. gordonii*; green: *Veillonella*).

converge to the homogeneous system where the biomass is evenly mixed initially.   388

3.3   Initial biomass distribution with patch-shape *Veillonella*   389

One could also argue that a random initial distribution of biomass is not realistic. It is   390
more often found in mature biofilm systems that the *Veillonella* group bacteria are   391
embedded in *Streptococcus* group bacteria (Chalmers et al., 2008; Kommerein et al.,   392
2017). Therefore, we carry out studies with two different initial biomass distribution   393
scenarios as shown in Figure 15. The irregular geometry of the *Veillonella* clusters,   394
numbered from 1 to 4 (from left to right), are generated by using an implicit   395
superquadric function which is defined by an implicit equation with random parameters   396
(Wellmann and Wriggers, 2012) and is rotated by a random angle. In case a), a flat   397
initial biofilm-fluid interface is generated, while a non-flat initial interface is generated   398
in case b).   399
    Simulation results of these two cases with different initial biomass distributions are   400
shown in Figure 16 and Figure 17. Similar to the numerical results presented in the   401
previous sections (with random initial biomass distributions), the *S. gordonii* moves to   402
the top of the biofilm. High (and low) lactic acid concentration area is observed locally   403
in space (as shown in the third column in Figure 16 and 17). Since the acid is produced   404

by the *S. gordonii* and consumed by the *Veillonella*, high concentration areas are observed mostly at the location where there is more *S. gordonii*.

The simulation results demonstrate that locally homogenization of the biomass occurs at the late time (at 1.0 day). Clearly, some *Veillonella* patches are still observable even at late time. One may expect that the patches will be homogenized if we run the simulation long enough. However, the homogenization time scale for each patch can be different (see later discussion). Specially, a non-flat biofilm-fluid interface can be generated during the development of the biofilm even if the interface is flat initially (as shown in Figure 16). This is due to the heterogeneous initial biomass distribution which results in a heterogeneous biofilm growth velocity field at an early time and thus leads to a change of the interface morphology. With the homogenization of the solution, the biofilm interface gets flatted again.

Different biomass distributions (in different cases, or at different areas in the same case) naturally result in different growth velocity fields locally. Therefore, the movements of *Veillonella* clusters behave differently as shown in Figure 18b. In the figure, contours of the volume fraction of *Veillonella* $\vartheta_2 = 0.6$ of each case at $t = 0$ and at $t = 1.0$ day are plotted.

As shown in Figure 18a, the second *Veillonella* cluster (numbered in Figure 15) seems to be rotated anticlockwise. However, the third cluster moves up together with the growth of the biofilm. The forth one is simply stretched and shrunken. In case b), as shown in Figure 18b, we observe an interesting feature of the numerical solution that the third and the forth *Veillonella* clusters cannot be captured by the contour line $\vartheta_2 = 0.6$ at 1.0 day. This demonstrates that those two clusters are homogenized faster than the other two clusters. The simulation results suggest that the homogenization time varies spatially which might be induced by the heterogeneous reactions.

The presented model shows the homogenized distribution of biomass at a later time, which can be different from the experimental observations. In this case, the model may fail to predict some biological behaviors of the system. Future validations are needed for determining to which extend the model is useful. Moreover, we focus on the biological/reactive dynamics of the system in this study. Therefore, we did not consider the interaction between the biofilm structure and the flow (Towler et al., 2007), the flow channels in the biofilms (Wilking et al., 2013), heterogeneous distributions of the biomass properties like the EPS, density, and so forth (Bryers and Drummond, 1998; Wimpenny et al., 2000). However, these are also essential processes in the life cycle of biofilms, and they may also influence the homogenization process. There is still a gap between understanding and modeling of the associated problems, which requires further studies.

3.4   The role of the symbiotic reactions

The presented reaction model may be one of the main reasons that cause local biomass homogenization. To understand the role of the reaction terms in the mathematical model, we compare the simulation results presented in Figure 18b to the results by using a model where the symbiotic reactions are switched off. The same model setup, as the case shown in Figure 18b, has been applied in the comparison. We simplify the symbiotic model (SM) presented in section 2 by applying:

$$g_1 = \vartheta_1 \rho_s \mu_1 \frac{s_1}{k_{11} + s_1}, \tag{21}$$

$$g_2 = \vartheta_2 \rho_v \mu_2 \frac{s_1}{k_{21} + s_1}, \tag{22}$$

and

$$r_{22} = 0. \tag{23}$$

Figure 16. Simulation results of the symbiotic biofilm system at different times (corresponding to the case in Figure 15a). The white curve denotes the biofilm-fluid interface. From top to bottom are at $t = 0.05$ day ($T = 0.05$), $t = 0.5$ day ($T = 0.5$) and $t = 1.0$ day ($T = 1.0$). Left column: Volume fraction distributions of *S. gordonii*; Middle colume: Volume fraction distributions of *Veillonella*; Right column: Dimensionless concentration of lactic acid (black curves denote isolines of S_2).

Figure 17. Simulation results of the symbiotic biofilm system at different times (corresponding to the case in Figure 15b). The white curve denotes the biofilm-fluid interface. From top to bottom are at $t = 0.05$ day ($T = 0.05$), $t = 0.5$ day ($T = 0.5$) and $t = 1.0$ day ($T = 1.0$). Left column: Volume fraction distributions of *S. gordonii*; Middle colume: Volume fraction distributions of *Veillonella*; Right column: Dimensionless concentration of lactic acid (black curves denote isolines of S_2).

(a) case a)   **(b)** case b)

Figure 18. Movement of *Veillonella* clusters over time. The purper curve and the black curve denote contour lines of volume fraction of *Veillonella* $\vartheta_2 = 0.6$ at $t = 0$ day ($T = 0$) and $t = 1.0$ day ($T = 1.0$) respectively.

(a) $t = 0.25$ day   **(b)** $t = 0.5$ day   **(c)** $t = 1.0$ day

Figure 19. Simulation results of volume fractions of *Veillonella* sp. at different times by using the competitive model. The white curve denotes the biofilm-fluid interface. The black and grey curves denote contour lines of volume fraction of *Veillonella* $\vartheta_2 = 0.6$ and $\vartheta_2 = 0.8$ respectively.

Thus, the model reduces to a competitive model (CM) in which the growth of both species of bacteria is limited only by the nutrient (saliva). In this case, the lactic acid does not play a role in biofilm growth, and it accumulates in space. One would expect that by applying CM, the biofilm grows faster than using SM due to the absent limitation of the lactic acid. Meanwhile, the species which has a larger maximum growth rate becomes dominant during the growth. As shown in Figure 19, the simulation results of the *Veillonella* sp. by using the competitive model prove the above expectations. The *Veillonella* sp. becomes dominating during the growth and wins the competition. The simulation results of the competitive model (as shown in Figure 19) do not show homogenized solutions like the results corresponding to the symbiotic model (see Figure 18b). This suggests that the symbiotic interaction can be one of the main reasons that lead to the homogenization process.

As a remark, the advection-diffusion biofilm model can be reformulated to a sharp interface model which prevents mixing of different species entirely if they are not mixed initially. However, by taking volume fractions as primary variables, it is difficult to reproduce sharp interface solutions numerically due to the present of numerical dissipation.

4 Summary and discussion

4.1 Summary

We have presented a new mathematical model for modeling the growth of symbiotic biofilms in this paper. An intermediate product, namely the lactic acid produced by the *Streptococcus* sp., has been explicitly modeled. The lactic acid functions as a growth-limiting substrate of the *Veillonella*, and suppresses the growth of the *Streptococcus*. We found that the volume fractions of different species of bacteria homogenize at a later simulation time.

To look more into the model, we studied how the initial biomass distribution influences the homogenization process. Random distributions of the biomass volume fractions with different correlation lengths were taken as initial conditions. We compared the simulation results of the case where the bacteria were initially evenly mixed. As would be expected, the biomass homogenized faster associating with an initial state with a smaller correlation length. However, the random distribution cases do not converge to the same concentration distribution as the evenly mixed one, although the same initial mass and morphology of the biofilm surface (flat surface) have been assigned.

For a better understanding of the role of the initial distributions, we investigated scenarios with patch-shape biomass. The simulation results demonstrate that the local growth field can lead to changes in the morphology of the biofilm surface and different movements of bacteria patches. Moreover, we also observe different local homogenization speeds that are induced by the heterogeneity of the growth velocity field. The presented biofilm model is further simplified to a competitive model by switching off the symbiotic reactions. The simulation results of the competitive model do not get homogenized, like in the symbiotic model. This demonstrates that the symbiotic reaction model presented in this paper is the main reason that leads to spatial homogenization.

4.2 Discussion

It is well-known in microbiology that bacterial biofilms' biological behaviors are a lot more complicated than a mere competition of "food" in nature. Many efforts have been put in the past decades to understand how different bacteria species interact with each other in a biofilm. The symbiotic relationship between two species of bacteria is one of the most interesting phenomena in microbiology systems.

Mathematical modeling has been proved to be a powerful tool for understanding biofilm's biological behaviors in the past years (Mattei et al., 2017). Tens or even hundreds of mathematical models have been developed for modeling the bacterial biofilm systems since the 1980s (Rittmann and McCarty, 1980). Nowadays, biofilm models are getting more comprehensive than earlier ones due to a deeper understanding of the biological mechanisms. Based on the previous pioneer continuum biofilm models (Alpkvist and Klapper, 2007; Klapper and Dockery, 2002), we studied the symbiotic behavior of a dual-species biofilm system by using mathematical modeling.

To the best of the authors' knowledge, this study provides the first comprehensive investigation of the dynamics of a biofilm's symbiotic reactions using a continuum mathematical model. We modeled the biofilm growth as an advective movement of a reactive potential flow in this study. The model has the advantage of capturing a sharp biofilm interface with introducing additional parameters; however, it suffers limitations on modeling merge of different colonies (see discussion in Section 2). To overcome this drawback, one can improve the model by considering the fluid phase explicitly and thus lead to the mixture models (Zhang et al., 2008a). Alternatively, one can model the biofilm growth as a diffusion process (Ghasemi et al., 2018; Rahman et al., 2015), and

apply the reaction models presented in this paper directly.

Instead of modeling the colonies' merge, we are more interested in the reaction model developed for the symbiotic biofilm systems. Even though the simulation results demonstrate that the model can represent the species of bacteria's symbiotic behaviors, we found that the reactive model can lead to homogenized solutions. The presented model may fail to predict the biological behavior of the symbiotic biofilms after a certain time when the numerical solution becomes too homogenized. Therefore, it is necessary to investigate what factors influence the homogenization speed in the model. It turned out that the speed of the homogenization depends on the initial biomass distribution.

Individual bacteria are never resolved in a continuum biofilm model. This might be another potential reason which leads to the over-mixing of the biomass. However, the discretized element based biofilm models can also result in considerable internal mixing, especially for multi-species problems (Tang and Valocchi, 2013). It would be interesting to compare the mixing behavior occurred in different continuum and discretized element based biofilm models in future studies. As a matter of fact, there are still many open questions on modeling multi-species biofilm problems. A third possible reason that results in the unwanted homogenized solution is missing important bio-processes in the model. There could be specific quorum sensing (QS) mechanisms that prevent the mixing process in nature. Mathematical modeling the QS in biofilm systems has already attracted much attention in recent years (Emerenini et al., 2015; Ghasemi et al., 2018; Zhao and Wang, 2017). Many more investigations are required to uncover the QS mechanisms in multi-species biofilms.

A Dimensioless form of governing equations

Mass balance for saliva

$$-\nabla^2 S_1 = -\Theta_{11}^2 \vartheta_1 \frac{S_1 K_L}{(K_{11} + S_1)(K_L + S_2)} - \Theta_{12}^2 \vartheta_2 \frac{S_1 S_2}{(K_{21} + S_1)(K_{22} + S_2)}, \quad \mathbf{X} \in \Omega^*. \tag{24}$$

Mass balance for lactic acid

$$\frac{\partial S_2}{\partial T} + \nabla \chi \cdot \nabla S_2 - D_2^* \nabla^2 S_2 =$$
$$\Theta_{21}^2 \vartheta_1 \frac{S_1 K_L}{(K_{11} + S_1)(K_L + S_2)} - \Theta_{22}^2 \vartheta_2 \frac{S_1 S_2}{(K_{21} + S_1)(K_{22} + S_2)} - \tag{25}$$
$$S_1 S_2 \left[\frac{\vartheta_1 \Psi_1 K_L}{(K_{11} + S_1)(K_L + S_2)} + \frac{\vartheta_2 \Psi_2 S_2}{(K_{21} + S_1)(K_{22} + S_2)} \right], \quad \mathbf{X} \in \Omega^*.$$

Mass balance for *S. gordonii*

$$\frac{\partial \vartheta_1}{\partial T} + \nabla \chi \cdot \nabla \vartheta_1 =$$
$$\vartheta_1 S_1 \left[\frac{(1 - \vartheta_1) \Psi_1 K_L}{(K_{11} + S_1)(K_L + S_2)} - \frac{\vartheta_2 \Psi_2 S_2}{(K_{21} + S_1)(K_{22} + S_2)} \right], \quad \mathbf{X} \in \Omega^*. \tag{26}$$

Mass balance for *Veillonella*

$$\frac{\partial \vartheta_2}{\partial T} + \nabla \chi \cdot \nabla \vartheta_2 =$$
$$\vartheta_2 S_1 \left[\frac{(1 - \vartheta_2) \Psi_2 S_2}{(K_{21} + S_1)(K_{22} + S_2)} - \frac{\vartheta_1 \Psi_1 K_L}{(K_{11} + S_1)(K_L + S_2)} \right], \quad \mathbf{X} \in \Omega^*. \tag{27}$$

Potential equation

$$\nabla^2 \chi = S_1 \left[\frac{\vartheta_1 \Psi_1 K_L}{(K_{11} + S_1)(K_L + S_2)} + \frac{\vartheta_2 \Psi_2 S_2}{(K_{21} + S_1)(K_{22} + S_2)} \right], \quad \mathbf{X} \in \mathrm{B}_{\mathrm{t}}^*. \tag{28}$$

B   Numerical methods

Two types of PDEs, namely the elliptic equations (24) and (28) and the time dependent advection-diffusion-reaction (ADR) equations (26) and (27) (or advection-reaction equations) are involved in the current biofilm model. The mass balance equations of the lactic acid and the biomass (*S. gordonii* and *Veillonella*) are solved simultaneously in this study. The Time-discontinuous Galerkin (TDG) method (Hughes and Hulbert, 1988; Hulbert, 1992) is applied for solving the time dependent PDEs. The finite increment calculus (FIC) method (Oñate, 1998; Oñate et al., 2007) is used to stabilize the numerical solutions. We would like to refer to (Feng et al., 2017; Sapotnick and Nackenhorst, 2012) for more detailed interpretation of the TDG-FIC method on solving the time dependent ADR equations. In this study, 4-node bi-linear iso-parametric elements are applied for spatial discretization of the mass balance equations of the saliva, lactic acid and biomass. 8-node second order iso-parametric elements are used for solving the potential equation. The free boundary (moving biofilm-fluid interface) at each time step is captured by using an iso-line of 80% of total biomass volume fraction. Detailed information on the numerical schemes involved in the numerical strategy will be presented in the following part. As a remark, only one and two- dimensional problems are studied in this paper even through the mathematical model also can be applied for three-dimensional problems.

B.1   Finite element approximation of nonlinear Poisson equation with C^0 elements

Equations (24) and (28) are Poisson equations. In this study, they are solved by using conventional Galerkin finite element method with C^0 type of elements. Both of these two equations can be written generally as

$$-\nabla^2 q = f, \quad q \in \mathcal{B}, \tag{29a}$$

$$q = c, \quad q \in \Gamma_D, \tag{29b}$$

$$\mathbf{n} \cdot \nabla q = 0, \quad q \in \Gamma_N, \tag{29c}$$

where q represents the primary variables as S_1 or χ in the governing equations. f denotes a function as $f = -\Theta_{11}^2 \vartheta_1 \frac{S_1 K_L}{(K_{11}+S_1)(K_L+S_2)} - \Theta_{12}^2 \vartheta_2 \frac{S_1 S_2}{(K_{21}+S_1)(K_{22}+S_2)}$ for equation (24) and $f = -S_1 \left[\frac{\vartheta_1 \Psi_1 K_L}{(K_{11}+S_1)(K_L+S_2)} + \frac{\vartheta_2 \Psi_2 S_2}{(K_{21}+S_1)(K_{22}+S_2)} \right]$ for equation (28). $\mathcal{B}$ refers to the computational domain with the corresponding Dirichlet and Neuman boundaries Γ_D and Γ_N. $\mathbf{n}$ is the norm vector on the Neuman boundary and c is a constant value corresponding to the Dirichlet boundary. The weak form of equation (29) can be written as

$$a(q, w) = F(w), \quad \forall w \in \mathcal{V}, \tag{30}$$

where,

$$a(q, w) = \int_{\mathcal{B}} \nabla q \nabla w \mathrm{d}V, \tag{31a}$$

$$F(w) = \int_{\mathcal{B}} f w \mathrm{d}V. \tag{31b}$$

w is an arbitrary test function in space $\mathcal{V}$ which is a subspace of the Hilbert space $\mathcal{H}^1$. As a remark, the Newton-Raphson method (Wriggers, 2008) is used when the equation is nonlinear.

B.2 TDG-FIC approximation of time dependent advection-diffusion-reaction equation

Equations (25)-(27) are time-dependent PDEs. Therefore, they are solved simultaneously by using the TDG-FIC method. Those three equations can be written as a set of time dependent advection-diffusion-reaction partial differential equations as

$$\frac{\partial \boldsymbol{q}}{\partial t} + \tilde{\mathbf{u}}^T \tilde{\boldsymbol{\nabla}} \boldsymbol{q} - \tilde{\boldsymbol{\nabla}}^T \tilde{\boldsymbol{D}} \tilde{\boldsymbol{\nabla}} \boldsymbol{q} - \boldsymbol{f}(\boldsymbol{q}) = \boldsymbol{0}, \tag{32}$$

where $\boldsymbol{q} = (S_2, \vartheta_1, \vartheta_2)^T$ denotes the unknown vector. Detailed information on the matrix form of each term in equation (32) can be found in C.

Applying discontinuous Galerkin method for time discretization and standard Galerkin discretization in space yields a time-space weak form of equation (32) over a time-space domain $\Omega \times \mathcal{T}_n$ (see Figure 2 in (Feng et al., 2019))

$$\int_\Omega w \int_{\mathcal{T}_n} \nu \left(\frac{\partial \boldsymbol{q}^h}{\partial t} + \tilde{\mathbf{u}}^T \tilde{\boldsymbol{\nabla}} \boldsymbol{q}^h - \tilde{\boldsymbol{\nabla}}^T \tilde{\boldsymbol{D}} \tilde{\boldsymbol{\nabla}} \boldsymbol{q}^h - \boldsymbol{f}(\boldsymbol{q}^h) \right) \mathrm{d}t \mathrm{d}\Omega + \\ \int_\Omega w \nu_{n-1}^+ [\![\boldsymbol{q}_{n-1}^h]\!] \mathrm{d}\Omega = 0, \tag{33}$$

where Ω and $\mathcal{T}_n$ denote the spatial and temporal computational domains, respectively. $\boldsymbol{q}^h$ denotes the time-space approximation of the variable vector. Since the nodal value at a time point is discontinuous, a jump value of $\boldsymbol{q}_{n-1}^h$ at t_{n-1}, which is notated as $[\![\boldsymbol{q}_{n-1}^h]\!]$, is introduced as

$$[\![\boldsymbol{q}_{n-1}^h]\!] = \boldsymbol{q}_{n-1}^{h+} - \boldsymbol{q}_{n-1}^{h-}. \tag{34}$$

Due to the non-linearity of the reaction terms $\boldsymbol{f}(\boldsymbol{q}^h)$, the Newton-Raphson method (Wriggers, 2008) is applied and equation (33) is linearized by introducing an incremental process with $\boldsymbol{q}^{h,0} = \boldsymbol{0}$ as

$$\boldsymbol{q}^{h,\beta+1} = \boldsymbol{q}^{h,\beta} + \Delta \boldsymbol{q}^h, \tag{35a}$$

$$\boldsymbol{f}(\boldsymbol{q}^{h,\beta+1}) \approx \boldsymbol{f}(\boldsymbol{q}^{h,\beta}) + \left. \frac{\partial \boldsymbol{f}}{\partial \boldsymbol{q}} \right|_{\boldsymbol{q}^{h,\beta}} \Delta \boldsymbol{q}^h. \tag{35b}$$

Substituting (35) into (33) yields the linearized time-space weak form

$$\int_\Omega w \int_{\mathcal{T}_n} \nu \left(\frac{\partial \Delta \boldsymbol{q}^h}{\partial t} + \tilde{\mathbf{u}}^T \tilde{\boldsymbol{\nabla}}(\Delta \boldsymbol{q}^h) - \tilde{\boldsymbol{\nabla}}^T \tilde{\boldsymbol{D}} \tilde{\boldsymbol{\nabla}}(\Delta \boldsymbol{q}^h) - \left. \frac{\partial \boldsymbol{f}}{\partial \boldsymbol{q}} \right|_{\boldsymbol{q}^{h,\beta}} \Delta \boldsymbol{q}^h \right) \mathrm{d}t \mathrm{d}\Omega \\ + \int_\Omega w \nu_{n-1}^+ \Delta \boldsymbol{q}_{n-1}^{h+} \mathrm{d}\Omega = \int_\Omega w \nu_{n+1}^+ \boldsymbol{q}_{n-1}^{h-,\beta} \mathrm{d}\Omega - \int_\Omega w \nu_{n-1}^+ \boldsymbol{q}_{n-1}^{h+,\beta} \mathrm{d}\Omega - \\ \int_\Omega w \int_{\mathcal{T}_n} \nu \left(\frac{\partial \boldsymbol{q}^{h,\beta}}{\partial t} + \tilde{\mathbf{u}}^T \tilde{\boldsymbol{\nabla}} \boldsymbol{q}^{h,\beta} - \tilde{\boldsymbol{\nabla}}^T \tilde{\boldsymbol{D}} \tilde{\boldsymbol{\nabla}} \boldsymbol{q}^{h,\beta} - \boldsymbol{f}(\boldsymbol{q}^{h,\beta}) \right) \mathrm{d}t \mathrm{d}\Omega. \tag{36}$$

The unknown field of $\boldsymbol{q}^h$ is interpolated over the time-space space by

$$\boldsymbol{q}^h = \sum_{k=1}^{n_t} \hat{N}_k \sum_{j=1}^{n_n} \tilde{N}_j \hat{\tilde{\boldsymbol{q}}}_{jk}, \tag{37}$$

where n_t and n_n refer to temporal and spatial total degrees of freedoms respectively. $\hat{N}$ and $\tilde{N}$ are the temporal and spatial shape functions and $\hat{\tilde{\boldsymbol{q}}}$ denotes the time-space nodal variable vector.

Discretization of equation (36) in space and time by adopting equation (37) within single time-space element $\Omega^e \times \mathcal{T}_n$ yields

$$
\begin{aligned}
&\left[(\tilde{\boldsymbol{N}}^T, \tilde{\boldsymbol{N}})^e \otimes \mathbf{T}_a^n + \right.\\
&\left. \left((-\tilde{\boldsymbol{B}}^T \tilde{\mathbf{u}}, \tilde{\boldsymbol{N}})^e + (\tilde{\boldsymbol{B}}^T \tilde{\boldsymbol{D}}, \tilde{\boldsymbol{B}})^e - (\tilde{\boldsymbol{N}}^T \left.\frac{\partial \boldsymbol{f}}{\partial \boldsymbol{q}}\right|_{\boldsymbol{q}^{h,\beta}}, \tilde{\boldsymbol{N}})^e \right) \otimes \mathbf{T}_b^n \right] \Delta \hat{\tilde{\boldsymbol{q}}} \\
&= (\hat{\tilde{\boldsymbol{q}}}_{n-1}^-, \tilde{\boldsymbol{N}}^T)^e \otimes \mathbf{T}_c^n - (\tilde{\boldsymbol{N}}^T, \tilde{\boldsymbol{N}})^e \otimes \mathbf{T}_a^n \hat{\tilde{\boldsymbol{q}}}^\beta - \\
&\left[(-\tilde{\boldsymbol{B}}^T \tilde{\mathbf{u}}, \tilde{\boldsymbol{N}})^e \hat{\tilde{\boldsymbol{q}}}^\beta + (\tilde{\boldsymbol{B}}^T \tilde{\boldsymbol{D}}, \tilde{\boldsymbol{B}})^e \hat{\tilde{\boldsymbol{q}}}^\beta - (\boldsymbol{f}\left(\hat{\tilde{\boldsymbol{q}}}^\beta\right), \tilde{\boldsymbol{N}}^T)^e \right] \otimes \mathbf{T}_b^n,
\end{aligned}
\tag{38}
$$

where

$$
(m, n)^e = \int_{\Omega^e} mm \mathrm{d}\Omega, \tag{39}
$$

$\hat{\tilde{\boldsymbol{q}}}$ refers to the time-space approximation of the nodal variables. $\mathbf{T}_a$, $\mathbf{T}_b$, $\mathbf{T}_c$ are time matrixes with only temporal shape function $\hat{\boldsymbol{N}}$ involved

$$
\mathbf{T}_a = \int_{\mathcal{T}_n} \hat{\boldsymbol{N}}^T \frac{\mathrm{d}\hat{\boldsymbol{N}}}{\mathrm{d}t} \mathrm{d}t + \left(\hat{\boldsymbol{N}}_{n-1}^+\right)^T \hat{\boldsymbol{N}}_{n-1}^+, \tag{40a}
$$

$$
\mathbf{T}_b = \int_{\mathcal{T}_n} \hat{\boldsymbol{N}}^T \hat{\boldsymbol{N}} \mathrm{d}t, \tag{40b}
$$

$$
\mathbf{T}_c = \left(\hat{\boldsymbol{N}}_{n-1}^+\right)^T \hat{\boldsymbol{N}}_{n-1}^-. \tag{40c}
$$

$\tilde{\boldsymbol{N}}$ is extended spatial shape function depending on the spatial element type (see C for detailed matrix form of $\tilde{\boldsymbol{N}}$).

The diffusion term in (32) only affects the mass balance equation of the lactic acid and does not appear in the mass balance equations of biomass. Therefore, the system of the mass balance equations in this model is still hyperbolic dominated. It is well known that solving hyperbolic PDEs with the conventional Galerkin method suffers from convective instability (Donea and Huerta, 2003). One way to relieve such a problem is by adding an additional artificial diffusion (balancing diffusion) term into the system. This yields the so-called "stabilization methods" (Codina, 1998; Franca et al., 1992; Lian et al., 2016) which have been developed for Galerkin methods over the past decades. In this paper, one of the widely used stabilization methods, namely the finite increment calculus (FIC) method, which introduces a nonlinear balancing artificial diffusion term in (38) is adopted. Thus, the diffusion coefficient matrix $\tilde{\boldsymbol{D}}$ is replaced by

$$
\tilde{\boldsymbol{D}}^{STB} = \tilde{\boldsymbol{D}} + \tilde{\boldsymbol{D}}_{FIC}, \tag{41}
$$

where

$$
\tilde{\boldsymbol{D}}_{FIC} = \begin{pmatrix} \boldsymbol{D}_I^*(S_2) & \boldsymbol{0} & \boldsymbol{0} \\ \boldsymbol{0} & \boldsymbol{D}_{II}^*(\vartheta_1) & \boldsymbol{0} \\ \boldsymbol{0} & \boldsymbol{0} & \boldsymbol{D}_{III}^*(\vartheta_2) \end{pmatrix} \tag{42}
$$

$\boldsymbol{D}_I^*$, $\boldsymbol{D}_{II}^*$ and $\boldsymbol{D}_{III}^*$ are 2×2 diagonal matrices depending on the variables S_2, ϑ_1 and ϑ_2 respectively as in (42). Detailed calculation procedures of these $\boldsymbol{D}^*$ matrices for multi-dimensional problems can be found in (Feng et al., 2017). As a remark, $\tilde{\boldsymbol{D}}_{FIC}$ depends on the solution of each spatial element at the current time step. Moreover, the artificial balancing diffusion is introduced along the directions of the principle curvatures of the solution and thus it must be calculated in a local coordinate system and then transferred to the global Cartesian coordinate system afterwards. Therefore, the FIC stabilized algorithm becomes nonlinear and iterations at each time step are required even if a linear problem is considered.

C   Matrices in the advection-reaction equation

For two-dimensional problems, the advection and diffusion coefficients matrices read

$$\tilde{\mathbf{u}} = \mathbf{I}_{3\times 3} \otimes \mathbf{u} = \begin{pmatrix} 1 & 0 & 0 \\ 0 & 1 & 0 \\ 0 & 0 & 1 \end{pmatrix} \otimes \begin{pmatrix} u_x \\ u_y \end{pmatrix} = \begin{pmatrix} u_x & 0 & 0 \\ u_y & 0 & 0 \\ 0 & u_x & 0 \\ 0 & u_y & 0 \\ 0 & 0 & u_x \\ 0 & 0 & u_y \end{pmatrix}, \tag{43a}$$

$$\tilde{\boldsymbol{\nabla}} = \mathbf{I}_{3\times 3} \otimes \nabla = \begin{pmatrix} 1 & 0 & 0 \\ 0 & 1 & 0 \\ 0 & 0 & 1 \end{pmatrix} \otimes \begin{pmatrix} \frac{\partial}{\partial x} \\ \frac{\partial}{\partial y} \end{pmatrix} = \begin{pmatrix} \frac{\partial}{\partial x} & 0 & 0 \\ \frac{\partial}{\partial y} & 0 & 0 \\ 0 & \frac{\partial}{\partial x} & 0 \\ 0 & \frac{\partial}{\partial y} & 0 \\ 0 & 0 & \frac{\partial}{\partial x} \\ 0 & 0 & \frac{\partial}{\partial y} \end{pmatrix}, \tag{43b}$$

$$\tilde{\mathbf{D}} = \begin{pmatrix} \boldsymbol{D_2^*} & \mathbf{0} & \mathbf{0} \\ \mathbf{0} & \mathbf{0} & \mathbf{0} \\ \mathbf{0} & \mathbf{0} & \mathbf{0} \end{pmatrix}, \tag{43c}$$

where u_x and u_y are the advection velocity components along x and y coordinates respectively. $\mathbf{I}$ denotes an identity matrix and $\otimes$ is the Kronecker product. $\mathbf{I}$ denotes an identical matrix and $\otimes$ is the Kronecker product. $\boldsymbol{D_2^*}$ denotes a 2×2 diagonal matrix with D_2^* placed at the main diagonal and $\mathbf{0}$ denotes a 2×2 null matrix. Furthermore, $\boldsymbol{f}(\boldsymbol{q})$ reads

$$\boldsymbol{f}(\boldsymbol{q}) = \left(f_1\left(S_2, \vartheta_1, \vartheta_2\right), f_2\left(S_2, \vartheta_1, \vartheta_2\right), f_3\left(S_2, \vartheta_1, \vartheta_2\right) \right)^T, \tag{44}$$

where f_1, f_2 and f_3 denote the reaction terms of equations (25)-(27) respectively.

For four-node 2D bi-linear element, the shape function is extended to

$$\tilde{\boldsymbol{N}} = \boldsymbol{N} \otimes \mathbf{I}_{3\times 3}, \tag{45}$$

where $\boldsymbol{N} = [N_1, N_2, N_3, N_4]$ is the shape function of the four-node 2D bi-linear element. Similarly,

$$\tilde{\boldsymbol{B}} = \boldsymbol{B} \otimes \mathbf{I}_{3\times 3}. \tag{46}$$

Acknowledgements

The authors would like to thank Dr. Henryke Rath, Dr. Nadine Kommerein and Mrs. Natascha Brandhorst for many helpful discussions. Dianlei Feng would also like to thank the German Research Council (DFG) under grant No. 426819984.

References

Alpkvist, E. and Klapper, I. (2007). A multidimensional multispecies continuum model for heterogeneous biofilm development. *Bulletin of Mathematical Biology*, 69(2):765–789.

Bradshaw, D. J. and Marsh, P. D. (1998). Analysis of pH-driven disruption of oral microbial communities in vitro. *Caries Research*, 32(6):456–462.

Bryers, J. D. and Drummond, F. (1998). Local macromolecule diffusion coefficients in structurally non-uniform bacterial biofilms using fluorescence recovery after photobleaching (FRAP). *Biotechnology and Bioengineering*, 60(4):462–473.

Chalmers, N. I. (2008). *Multispecies oral biofilms studied at the single community level as a model system for spatiotemporal development of biofilms and interspecies interactions*. Ph.d. thesis, University of Maryland, Baltimore, Maryland, United States.

Chalmers, N. I., Palmer, R. J., Cisar, J. O., and Kolenbrander, P. E. (2008). Characterization of a *Streptococcus* sp.-*Veillonella* sp. community micromanipulated from dental plaque. *Journal of Bacteriology*, 190(24):8145–8154.

Chihara, K., Matsumoto, S., Kagawa, Y., and Tsuneda, S. (2015). Mathematical modeling of dormant cell formation in growing biofilm. *Frontiers in Microbiology*, 6:534.

Cirpka, O. A. (2003). Generation of random, autocorrelated, periodic fields.

Codina, R. (1998). Comparison of some finite element methods for solving the diffusion-convection-reaction equation. *Computer Methods in Applied Mechanics and Engineering*, 156(1):185–210.

Cogan, N. G. (2004). The role of the biofilm matrix in structural development. *Mathematical Medicine and Biology*, 21(2):147–166.

Corbin, A., Pitts, B., Parker, A., and Stewart, P. S. (2011). Antimicrobial penetration and efficacy in an in vitro oral biofilm model. *Antimicrobial Agents and Chemotherapy*, 55(7):3338–3344.

Donea, J. and Huerta, A. (2003). *Finite element methods for flow problems*. John Wiley & Sons.

Duddu, R., Chopp, D. L., and Moran, B. (2009). A two-dimensional continuum model of biofilm growth incorporating fluid flow and shear stress based detachment. *Biotechnology and Bioengineering*, 103(1):92–104.

Eberl, H. J., Parker, D. F., and van Loosdrecht, Mark C. M. (2001). A new deterministic spatio-temporal continuum model for biofilm development. *Journal of Theoretical Medicine*, 3(3):161–175.

Egland, P. G., Palmer, R. J., and Kolenbrander, P. E. (2004). Interspecies communication in streptococcus gordonii–veillonella atypica biofilms: Signaling in flow conditions requires juxtaposition. *Proceedings of the National Academy of Sciences*, 101(48):16917–16922.

Emerenini, B. O., Hense, B. A., Kuttler, C., and Eberl, H. J. (2015). A mathematical model of quorum sensing induced biofilm detachment. *PLOS ONE*, 10(7):1–25.

Feng, D., Neuweiler, I., and Nackenhorst, U. (2017). A spatially stabilized TDG based finite element framework for modeling biofilm growth with a multi-dimensional multi-species continuum biofilm model. *Computational Mechanics*, 59(6):1049–1070.

Feng, D., Neuweiler, I., Nackenhorst, U., and Wick, T. (2019). A time-space flux-corrected transport finite element formulation for solving multi-dimensional advection-diffusion-reaction equations. *Journal of Computational Physics*, 396:31–53.

Feng, D., Rath, H., Neuweiler, I., Stumpp, N., Nackenhorst, U., and Stiesch, M. (2018). A deeper insight of a multi-dimensional continuum biofilm growth model: experimental observation and parameter studies. In Wriggers, P. and Lenarz, T., editors, *Biomedical Technology: Modeling, Experiments and Simulation*, pages 257–272. Springer International Publishing, Cham.

Franca, L. P., Frey, S. L., and Hughes, T. J. (1992). Stabilized finite element methods: I. Application to the advective-diffusive model. *Computer Methods in Applied Mechanics and Engineering*, 95(2):253–276.

Fujikawa, H. (1994). Diversity of the growth patterns of Bacillus subtilis colonies on agar plates. *FEMS Microbiology Ecology*, 13(3):159–168.

Ghasemi, M., Hense, B. A., Eberl, H. J., and Kuttler, C. (2018). Simulation-based exploration of quorum sensing triggered resistance of biofilms to antibiotics. *Bulletin of Mathematical Biology*, 80(7):1736–1775.

Gordeeva, Y. L., Rudakovskaya, E., Gordeeva, E., and Borodkin, A. (2017). Mathematical modeling of biotechnological process of lactic acid production by batch fermentation: A review. *Theoretical Foundations of Chemical Engineering*, 51(3):282–298.

Hughes, T. J. and Hulbert, G. M. (1988). Space-time finite element methods for elastodynamics: formulations and error estimates. *Computer Methods in Applied Mechanics and Engineering*, 66(3):339–363.

Hulbert, G. M. (1992). Time finite element methods for structural dynamics. *International Journal for Numerical Methods in Engineering*, 33(2):307–331.

Kara, D., Luppens, S. B. I., van Marle, J., Ozok, R., and ten Cate, J. M. (2007). Microstructural differences between single-species and dual-species biofilms of *Streptococcus mutans* and *Veillonella parvula*, before and after exposure to chlorhexidine. *FEMS Microbiology Letters*, 271(1):90–97.

Khassehkhan, H. and Eberl, H. J. (2008). Modeling and simulation of a bacterial biofilm that is controlled by ph and protonated lactic acids. *Computational and Mathematical Methods in Medicine*, 9(1):47–67.

Klapper, I. and Dockery, J. (2002). Finger formation in biofilm layers. *SIAM Journal on Applied Mathematics*, 62(3):853–869.

Kolenbrander, P. E., Palmer, JR, R. J., Periasamy, S., and Jakubovics, N. S. (2010). Oral multispecies biofilm development and the key role of cell-cell distance. *Nature Reviews. Microbiology*, 8(7):471–480.

Kommerein, N., Stumpp, S. N., Müsken, M., Ehlert, N., Winkel, A., Häussler, S., Behrens, P., Buettner, F. F. R., and Stiesch, M. (2017). An oral multispecies biofilm model for high content screening applications. *PloS One*, 12(3):e0173973.

Kreft, J. U., Picioreanu, C., Wimpenny, J. W., and van Loosdrecht, M. C. (2001). Individual-based modelling of biofilms. *Microbiology (Reading, England)*, 147(Pt 11):2897–2912.

Lian, Y., Ying, Y., Tang, S., Lin, S., Wagner, G. J., and Liu, W. K. (2016). A Petrov–Galerkin finite element method for the fractional advection–diffusion equation. *Computer Methods in Applied Mechanics and Engineering*, 309:388–410.

Lindley, B., Wang, Q., and Zhang, T. (2012). Multicomponent hydrodynamic model for heterogeneous biofilms: two-dimensional numerical simulations of growth and interaction with flows. *Physical Review. E, Statistical, Nonlinear, and Soft Matter Physics*, 85(3 Pt 1):031908.

Madigan, M. T. (2012). *Brock biology of microorganisms*. Benjamin Cummings, San Francisco, 13th ed. edition.

Martin, B., Tamanai-Shacoori, Z., Bronsard, J., Ginguené, F., Meuric, V., Mahé, F., and Bonnaure-Mallet, M. (2017). A new mathematical model of bacterial interactions in two-species oral biofilms. *PloS One*, 12(3):e0173153.

Martin, K. J., Picioreanu, C., and Nerenberg, R. (2015). Assessing microbial competition in a hydrogen-based membrane biofilm reactor (MBfR) using multidimensional modeling. *Biotechnology and Bioengineering*, 112(9):1843–1853.

Mashima, I. and Nakazawa, F. (2015). The interaction between *Streptococcus* spp. and *Veillonella tobetsuensis* in the early stages of oral biofilm formation. *Journal of Bacteriology*.

Mattei, M. R., Frunzo, L., D'Acunto, B., Pechaud, Y., Pirozzi, F., and Esposito, G. (2017). Continuum and discrete approach in modeling biofilm development and structure: a review. *Journal of Mathematical Biology*.

Merkey, B. V., Rittmann, B. E., and Chopp, D. L. (2009). Modeling how soluble microbial products (smp) support heterotrophic bacteria in autotroph-based biofilms. *Journal of Theoretical Biology*, 259(4):670 – 683.

Nascimento, M. M., Gordan, V. V., Garvan, C. W., Browngardt, C. M., and Burne, R. A. (2009). Correlations of oral bacterial arginine and urea catabolism with caries experience. *Oral Microbiology and Immunology*, 24(2):89–95.

Noguera, D., Pizarro, G., Stahl, D., and Rittmann, B. (1999). Simulation of multispecies biofilm development in three dimensions. *Water Science and Technology*, 39(7):123–130.

Oñate, E. (1998). Derivation of stabilized equations for numerical solution of advective-diffusive transport and fluid flow problems. *Computer Methods in Applied Mechanics and Engineering*, 151(1-2):233–265.

Oñate, E., Miquel, J., and Zárate, F. (2007). Stabilized solution of the multidimensional advection–diffusion–absorption equation using linear finite elements. *Computers & Fluids*, 36(1):92–112.

Paquette, D. W., Brodala, N., and Williams, R. C. (2006). Risk factors for endosseous dental implant failure. *Dental Clinics of North America*, 50(3):361–74, vi.

Periasamy, S. and Kolenbrander, P. E. (2010). Central role of the early colonizer *Veillonella* sp. in establishing multispecies biofilm communities with initial, middle, and late colonizers of enamel. *Journal of Bacteriology*, 192(12):2965–2972.

Picioreanu, C., Kreft, J.-U., and van Loosdrecht, M. C. M. (2004). Particle-based multidimensional multispecies biofilm model. *Applied and Environmental Microbiology*, 70(5):3024–3040.

Picioreanu, C., RITTMANN, B., and Van Loosdrecht, M. (2006). *Mathematical modelling of biofilms*. IWA Publishing, London, UK.

Picioreanu, C., van Loosdrecht, M. C., and Heijnen, J. J. (1998). A new combined differential-discrete cellular automaton approach for biofilm modeling: application for growth in gel beads. *Biotechnology and Bioengineering*, 57(6):718–731. 715 716 717

Picioreanu, C., van Loosdrecht, M. C., and Heijnen, J. J. (2000). Effect of diffusive and convective substrate transport on biofilm structure formation: a two-dimensional modeling study. *Biotechnology and Bioengineering*, 69(5):504–515. 718 719 720

Rahman, K. A., Sudarsan, R., and Eberl, H. J. (2015). A mixed-culture biofilm model with cross-diffusion. *Bulletin of Mathematical Biology*, 77(11):2086–2124. 721 722

Rath, H., Feng, D., Neuweiler, I., Stumpp, N. S., Nackenhorst, U., and Stiesch, M. (2017). Biofilm formation by the oral pioneer colonizer *Streptococcus gordonii*: an experimental and numerical study. *FEMS Microbiology Ecology*, 93(3). 723 724 725

Ribeiro, A. C. F., Lobo, V. M. M., Leaist, D. G., Natividade, J. J. S., Veríssimo, L. P., Barros, M. C. F., and Cabral, A. M. T. D. P. V. (2005). Binary diffusion coefficients for aqueous solutions of lactic acid. *Journal of Solution Chemistry*, 34(9):1009–1016. 726 727 728

Rittmann, B. E. and McCarty, P. L. (1980). Model of steady-state-biofilm kinetics. *Biotechnology and Bioengineering*, 22(11):2343–2357. 729 730

Sapotnick, A. and Nackenhorst, U. (2012). A combined FIC-TDG finite element approach for the numerical solution of coupled advection–diffusion–reaction equations with application to a bioregulatory model for bone fracture healing. *International Journal for Numerical Methods in Engineering*, 92(3):301–317. 731 732 733 734

Tang, Y. and Liu, H. (2017). Modeling multidimensional and multispecies biofilms in porous media. *Biotechnology and Bioengineering*, 114(8):1679–1687. 735 736

Tang, Y. and Valocchi, A. J. (2013). An improved cellular automaton method to model multispecies biofilms. *Water Research*, 47(15):5729–5742. 737 738

Towler, B. W., Cunningham, A., Stoodley, P., and McKittrick, L. (2007). A model of fluid–biofilm interaction using a burger material law. *Biotechnology and Bioengineering*, 96(2):259–271. 739 740 741

Wanner, O. and Gujer, W. (1986). A multispecies biofilm model. *Biotechnology and Bioengineering*, 28(3):314–328. 742 743

Wanner, O. and Reichert, P. (1996). Mathematical modeling of mixed-culture biofilms. *Biotechnology and Bioengineering*, 49(2):172–184. 744 745

Wellmann, C. and Wriggers, P. (2012). A two-scale model of granular materials. *Computer Methods in Applied Mechanics and Engineering*, 205-208:46–58. 746 747

Wilking, J. N., Zaburdaev, V., De Volder, M., Losick, R., Brenner, M. P., and Weitz, D. A. (2013). Liquid transport facilitated by channels in bacillus subtilis biofilms. *Proceedings of the National Academy of Sciences*, 110(3):848–852. 748 749 750

Wimpenny, J., Manz, W., and Szewzyk, U. (2000). Heterogeneity in biofilms. *FEMS Microbiology Reviews*, 24(5):661–671. 751 752

Wimpenny, J. W. and Colasanti, R. (1997). A unifying hypothesis for the structure of microbial biofilms based on cellular automaton models. *FEMS Microbiology Ecology*, 22(1):1–16. 753 754 755

Wriggers, P. (2008). *Nonlinear finite element methods.* Springer, Berlin and London. 756

Xavier, J. B., Picioreanu, C., and Van Loosdrecht, M. C. (2005). A framework for multidimensional modelling of activity and structure of multispecies biofilms. *Environmental microbiology*, 7(8):1085–1103.

Yang, L., Liu, Y., Wu, H., Høiby, N., Molin, S., and Song, Z. (2011). Current understanding of multi-species biofilms. *International Journal of Oral Science*, 3(2):74–81.

Zhang, T., Cogan, N., and Wang, Q. (2008a). Phase field models for biofilms. II. 2-D numerical simulations of biofilm-flow interaction. *Communications in Computational Physics*, 4(1):72–101.

Zhang, T., Cogan, N. G., and Wang, Q. (2008b). Phase field models for biofilms. I. Theory and one-dimensional simulations. *SIAM Journal on Applied Mathematics*, 69(3):641–669.

Zhao, J. and Wang, Q. (2017). Three-dimensional numerical simulations of biofilm dynamics with Quorum Sensing in a flow cell. *Bulletin of Mathematical Biology*, 79(4):884–919.