

1 **Article Title:** Using publicly available data to conduct rapid assessments of extinction risk

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3 **Authors:** Michael O. Levin¹, Jared B. Meek¹, Brian Boom², Sara M. Kross¹, Evan A. Eskew³

4 ¹ Department of Ecology, Evolution, and Environmental Biology, Columbia University, New York,
5 NY

6 ² Center for Conservation Strategy, New York Botanical Garden, New York, NY

7 ³ Department of Biology, Pacific Lutheran University, Tacoma, WA

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12 **Corresponding Author:**

13 Michael Levin

14 1190 Amsterdam Ave,

15 New York, NY 10027

16 mol2112@columbia.edu

17 **Abstract**

18 The IUCN Red List plays a key role in setting global conservation priorities. Species are added
19 to the Red List through a rigorous assessment process that, while robust, can be quite time-
20 intensive. Here, we test the rapid preliminary assessment of plant species extinction risk using a
21 single Red List metric: Extent of Occurrence (EOO). To do so, we developed REBA (Rapid
22 EOO-Based Assessment), a workflow that harvests and cleans data from the Global Biodiversity
23 Information Facility (GBIF), calculates each species' EOO, and assigns Red List categories
24 based on that metric. We validated REBA results against 1,546 North American plant species
25 already on the Red List and found ~90% overlap between REBA's rapid classifications and those
26 of full IUCN assessments. Our preliminary workflow can be used to quickly evaluate data
27 deficient Red List species or those in need of reassessment, and can prioritize unevaluated
28 species for a full assessment.

29

30

31 Introduction

32 The International Union for the Conservation of Nature's (IUCN; www.iucn.org) Red List is one of the
33 most widely used frameworks to assess extinction risk. The representation of plants on the Red List,
34 however, suffers from pervasive biases that plague conservation science generally (Di Marco et al., 2017;
35 Nic Lughadha et al., 2020). For example, the proportion of described plant species added to the Red List
36 is well below that of vertebrates (10% and 72%, respectively, as of 2020;
37 <https://www.iucnredlist.org/resources/summary-statistics>, Table 1a). Assessed plants are primarily trees,
38 taxa of particular interest to IUCN Specialist Groups, and species linked to commercial and horticultural
39 interests, among other biases (Bachman et al., 2019; Brummitt, Bachman, & Moat, 2008; Sharrock,
40 2020). Furthermore, research shows that the Red List may be vastly underestimating plant extinctions, a
41 concerning finding considering that the modern rate of plant extinction is at least 500 times greater than
42 the background extinction rate (Humphreys, Govaerts, Ficinski, Nic Lughadha, & Vorontsova, 2019).

43
44 Red listing can also be hampered by features of the assessment process itself. Extant species can be
45 placed into one of six categories: Critically Endangered (CR), Endangered (EN), Vulnerable (VU), Near
46 Threatened (NT), Least Concern (LC), or Data Deficient (DD). The DD category lacks an explicit risk
47 status and is meant to temporarily hold species that lacked sufficient data to be fully assessed. However,
48 already limited conservation resources are rarely diverted to revisit DD listings (Morais et al., 2013), and,
49 as the category has swelled, the estimated cost to fully reassess all DD species is over USD 300 million
50 (Bland et al., 2015). Outside the DD category, many previously assessed species fail to receive mandated
51 regular reassessments; seventeen percent of all assessments were already out of date (> 10 years old) on
52 the 2012 Red List, with the median age of Red List assessments estimated to reach 36 years by 2050
53 (Rondinini, Di Marco, Visconti, Butchart, & Boitani, 2014).

54
55 To address these biases and limitations, several tools have been developed to facilitate rapid, preliminary
56 assessments (Nic Lughadha et al., 2019), particularly for plants (i.e. S. Bachman, Walker, Barrios,
57 Copeland, & Moat, 2020; Callmander, Schatz, & Lowry, 2005; Davis, Govaerts, Bridson, & Stoffelen,
58 2006; Le Breton et al., 2019; Miller et al., 2013; Utteridge, Nagamasu, Teo, White, & Gasson, 2005).
59 Many of these tools rely upon a single criterion from the full assessment, Criterion B, which focuses on
60 geographic range and is cited in more than 60% of all IUCN assessments (Le Breton et al., 2019).
61 Criterion B relies predominantly upon two measures: Extent of Occurrence (EOO) and Area of
62 Occupancy (AOO). EOO is related to geographic range and measures "the degree to which risks from
63 threatening factors are spread spatially across the taxon's geographical distribution," while AOO
64 correlates with population size and approximates a species' resistance to stochastic events (IUCN
65 Standards and Petitions Committee, 2019; Le Breton et al., 2019). Both have thresholds linked to
66 additional measures of population dynamics and trends that dictate their classification (i.e., if EOO < 100
67 km², a species could be classified CR).

68
69 Previous efforts in this vein have measured the accuracy of an EOO-based assessment method (Miller et
70 al., 2013), used EOO to assess the status of DD plant species (Roberts, Taylor, & Joppa, 2016), and, in
71 one recent instance, produced a streamlined tool that draws from publicly available data sets to identify
72 LC species and submit them to the Red List, allowing the attention of the full assessment to be redirected
73 towards species with higher extinction risk (Bachman et al., 2020). However, few of these rapid
74 assessment frameworks have examined the factors that might influence classification success and none
75 have been tested on large suites of species across broad geographic scales.

76
77 Here, we use a publicly available database to gather plant occurrence records for Red Listed species on a
78 continental scale for the first time and analyze the resulting data using a rapid, EOO-based assessment
79 (hereafter, REBA) to assign species a Red List category. We assessed the concordance between our
80 automated classifications and the existing full IUCN classifications, classified DD species into extinction

81 risk categories, and fit statistical models to highlight plant traits and threats that affect the probability of
82 “correct” classification using REBA. Ultimately, our results provide a proof-of-concept for a rapid
83 conservation classification workflow which can be applied to a wide range of species at various scales.
84 This method can serve as a prioritization tool for optimizing resources and effort toward producing full
85 IUCN assessments.

86 Methods

87 Automated Red List Classification

88 The REBA workflow begins by using the R package rGBIF (Chamberlain & Boettiger, 2017) to query
89 GBIF for georeferenced occurrence records, which we cap at 50,000 per species to reduce computation
90 time. To further clean the data, we remove records not belonging to kingdom Plantae and filter records to
91 only include “HUMAN_OBSERVATION” or “OBSERVATION” record types to eliminate records that
92 might be georeferenced to a museum location rather than the location of sample collection. We then use
93 the “cc_sea()” function within the R package CoordinateCleaner (Zizka et al., 2019) to remove
94 occurrence records that do not lie over land. Next, REBA uses the R package rCAT to conduct an EOO-
95 based Red List classification (Moat & Bachman, 2020). rCAT calculates EOO as the area of a minimum-
96 convex polygon drawn around known occurrence records (a minimum of 3 is required) and uses IUCN-
97 defined thresholds to classify species as CR, EN, VU, NT, or LC, with EOO values of $< 100\text{km}^2$, $<$
98 $5,000\text{km}^2$, $< 20,000\text{km}^2$, $< 30,000\text{km}^2$, and $\geq 30,000\text{km}^2$, respectively. REBA relies exclusively on EOO
99 because there is precedent for such an approach in the literature (see Davis et al., 2006; Miller et al.,
100 2013), and we believe that a metric designed to measure the spatial spread of risk itself (EOO) is more
101 robust for this analysis than one designed to approximate a species’ insurance against that risk (AOO).
102

103 Testing REBA on North American Plant Species

104 We tested the efficacy of the REBA workflow on each of the 2,662 North American plant species on the
105 Red List. We gathered data on extinction risk, ‘Plant Type’, and ‘Threats’ from the IUCN using the Red
106 List’s advanced search feature (<https://www.iucnredlist.org/search>; accessed March 23,
107 2020). After removing 109 species with no GBIF occurrence points and 23 with taxonomic
108 discrepancies, we initially harvested 13,232,845 occurrence records representing 2,530 species. While all
109 of these species are found in North America, not all are native to the continent. Non-native species
110 identified as part of the North American flora by the IUCN were retained for this analysis (hereafter:
111 “North American species”).
112

113 After passing through the data cleaning portion of the workflow we were left with 6,566,297 records from
114 1,829 unique plant species. We joined this occurrence data with Red List assessment data by species and
115 eliminated records from the year of or years following the IUCN’s assessment to ensure REBA was not
116 influenced by data unavailable during the original Red List assessment process. After eliminating species
117 with fewer than 3 cleaned occurrence records, REBA produced EOO-based Red List classifications for
118 1,546 plant species.
119

120 To visualize REBA’s accuracy we generated a tile plot illustrating the overlap between Red List Category
121 classifications generated by the IUCN and by REBA. We then calculated the number of “correctly”
122 classified species (i.e., REBA matched the existing Red List classification), over-classified species (i.e.,
123 REBA produced a higher extinction risk category than the existing Red List classification), or under-
124 classified species (i.e., REBA produced a lower extinction risk category than the existing Red List
125 classification). We also utilized the REBA workflow to classify plant species from North America
126 currently classified as DD.
127

128 Statistical Modeling

129 We derived two smaller data subsets to model the effects of ‘Plant Type’ and ‘Threats’ on the probability
130 of correct classification. The first contained all species that had ‘Plant Type’ data available from the
131 IUCN. In this dataset, we condensed the 18 original ‘Plant Type’ categories assigned by the IUCN into
132 ten new biologically relevant categories (Appendix 1), effectively eliminating those that would otherwise
133 be represented by a limited sample of species. This modeling dataset contained 1,533 plant species. The
134 second dataset represented all species that had ‘Threats’ data available from the IUCN. ‘Threats’
135 categories remained unaltered, and the dataset contained 396 species.

136
137 We then constructed two separate multilevel Bayesian models with a binomial outcome distribution
138 where correct classification was the variable of interest. Both models included the number of occurrence
139 points as a main effect. One included ‘Plant Type’ categories as varying effects, and the other included
140 ‘Threats’ categories as varying effects. These models estimate the effect of number of occurrence points,
141 ‘Plant Type’, and ‘Threats’ on REBA’s probability of correct classification. We specified and fit our
142 models using the “map2stan()” function within the ‘rethinking’ R package (Carpenter et al., 2017;
143 McElreath, 2016). We fit each model using four independent MCMC chains, specifying 7,500 total model
144 iterations, 2,500 of which were considered warmup. As a result, our model inferences are based on 20,000
145 posterior samples from each model (5,000 post-warmup samples per chain with four chains total). After
146 model fitting, we inspected parameter trace plots and R-hat values to confirm convergence and good
147 model fits (Gelman & Rubin, 1992). We report parameter estimates using posterior means and 99%
148 highest posterior density intervals (HPDIs) where appropriate.

149

150 Counterfactual Predictions of Classification Accuracy

151 We visualized our model results with counterfactual plots using the full posterior probability distributions
152 to show the implied relationships between the probability of correct classification and ‘Plant Type’ or
153 ‘Threats’, respectively. We plotted the implied probability of correct classification using four occurrence
154 point sample sizes: 100, 1,000, 10,000, and 20,000. In effect, we imagine applying REBA for a species
155 with a given ‘Plant Type’ or ‘Threats’ classification across a range of arbitrary sample sizes, where the
156 implied probability of correct classification is informed by the observed data that was used to fit the
157 model.

158

159 The inference derived from our fit models can also be applied to DD species, allowing us to generate
160 posterior distributions for the implied probability of correct Red List classification analogous to the
161 counterfactual analyses just described. All 13 DD species with REBA-generated classifications had ‘Plant
162 Type’ data, but only 5 had ‘Threats’ data. Therefore, we used the full posterior distributions from our
163 ‘Plant Type’ model to generate the implied probability of correct classification for each of these species,
164 given their ‘Plant Type’ and the actual number of cleaned occurrence points available. Thus, our analyses
165 for DD species represent the implied probability of correct classification for the exact species-level data
166 as analyzed through REBA.

167 Results

168 Classification Overlap and Modeling Probability of Correct Classification

169 REBA correctly classified 1,379 of 1,533 species (89.95%) in our dataset of North American plant
170 species. An overwhelming majority of correct classifications (99.49%) were for LC species (Fig. 1). We
171 under-classified 58 species (3.78%) and over-classified 96 (6.26%).

172

173 ‘Geophytes’ contained the highest proportion of under-classified species among ‘Plant Type’ (17.02%;
174 Fig. 2) and exhibited the strongest negative effect on the probability of correct classification (mean effect
175 on logit scale [99% HPDI]: -1.14 [-2.25, 0.00]; Fig. 3). ‘Annuals’, ‘Ferns’, and ‘Graminoids’ contained

176 no under-classified species (0%; Table S1, Fig. 2) and `Annuals` exhibited the strongest positive effect
177 (1.13 [-0.22, 3.48]; Fig. 3). In the `Plant Type` model, posterior estimates for effect of the number of
178 points (NOP) maintained support for positive values across the entire 99% HPDI (1.60 [0.63, 3.06]).

179
180 Among `Threats` categories, `Human Intrusions and Disturbance` had the highest proportion of under-
181 classified species (27.03%; Fig. 4), while `Residential and Commercial Development` exhibited the most
182 negative effect on the probability of correct classification (-0.55 [-1.20, 0.03]; Fig. 5). `Climate Change
183 and Severe Weather` had the lowest proportion of under-classified species (12.62%; Table S2, Fig. 4) and
184 exhibited the largest positive effect on correct classification (0.24 [-0.33, 0.91]; Fig. 5). In the `Threat`
185 model, posterior estimates for NOP again had support constrained to only positive values (0.99 [0.16,
186 2.26]).

187

188 Classifying Data Deficient Species

189 Thirteen North American species classified by the IUCN as DD remained after data filtering (Table S3).
190 REBA classified twelve as LC and one as NT. Their mean EOO was 56,612,597.68 km², with a minimum
191 of 22,002.41 km² and a maximum of 162,676,407.50 km². Their mean NOP was 4,322.80, with a
192 minimum of 9 and a maximum of 40,298. All DD species had mean implied probabilities of correct
193 classification above 0.70 (Table S3, Fig. 6). *Ulmus glabra* had the highest mean implied probability of
194 correct classification (1.00 [.99, 1.00]) and *Zingiber zerumbet* had the lowest (.73 [.49, .90]; Table S3,
195 Fig. 6). These thirteen species represent a third of the total number of North American species classified
196 as DD (n = 39). The 26 species that were removed in our data cleaning process contained fewer than three
197 cleaned occurrence records, precluding an EOO calculation. They remain priorities for further data
198 collection.

199 Discussion

200 REBA can swiftly and accurately produce preliminary Red List assessments on a continental scale that
201 match existing Red List assessments approximately 90% of the time. Our modeling efforts indicated that
202 the number of points available for a species is the most important contributor to the probability of correct
203 classification. REBA classified 13 DD species into non-threatened categories with more than 70%
204 probability of presumed correct classification. Results from the `Threats` model are less easily interpreted
205 than those of the `Plant Type` model. The mean number of `Plant Type` per species was 1.29, while for
206 `Threats` it was 2.35. Because many species are affected by multiple threats, it is more difficult to parse
207 individual effects of each one.

208

209 One of our most significant concerns regarding REBA came from under-classified species, which
210 represent species of current conservation concern that would be masked under this assessment framework.
211 REBA under-classified all seven North American species listed as CR, labelling six as LC: five *Fraxinus*
212 species threatened by the Emerald Ash Borer (*Agilus planipennis*), and the American Chestnut
213 (*Castanea dentata*), threatened by the chestnut blight caused by *Cryphonectria parasitica*. These are
214 widespread species, and their large EOOs mask the tremendous risk that invasive species and disease
215 represent across their range. These particular threats are thus potential confounders of the REBA
216 framework, particularly for species with large EOOs. This matches observations made in other
217 applications of similar methods (S. Bachman et al., 2020).

218

219 Another concern arises from the data itself. We found that the probability of correct classification
220 increases substantially across all `Plant Types` and `Threats` with increasing NOP: REBA assessments
221 based on > 20,000 records approach a 100% probability of correct classification (Fig. 3, Fig. 5). However,
222 including more data may expose REBA to well-founded data quality concerns for publicly available data
223 (see Meyer, Weigelt, & Kreft, 2016). Our data cleaning relied primarily on GBIF metadata and only
224 implemented spatial filters to remove non-terrestrial points. Bachman et al. used a similar methodology

225 but employed a stricter spatial filter that limited occurrence data to those records that overlapped with a
226 species' native range (Bachman et al., 2020). However, such strict filtering potentially masks valuable
227 occurrence records of non-native but naturalized species. Critical spatial filtering and further assessment
228 of more rigorous cleaning methods to improve data quality is central to further refinement of rapid
229 assessment tools.

230
231 While acknowledging these concerns, REBA did produce correct classifications in ~90% of cases. The
232 majority were for LC species, which could be a result of geographic bias in our study; North American
233 species are well-studied and likely to have significant data available (Meyer et al., 2016), which may
234 underlie the high number of LC classifications that REBA produced. By a similar logic, the more
235 occurrence points a species has, the more likely it is to be of minimal conservation concern—a species
236 with > 10,000 records is likely to be widespread. However, the identification of LC species is valuable for
237 assessment prioritization. Once recognized, LC species can be put aside in the assessment pipeline in
238 favor of those more likely to have a higher extinction risk that would benefit from more immediate
239 attention (Bachman et al., 2020).

240
241 REBA's optimal function is to work alongside the full Red List assessment process, helping to focus
242 limited human and financial capital on species in need of immediate attention by discovering species of
243 least concern. We have demonstrated that it can accurately and rapidly leverage publicly available data to
244 operate on a continental scale to prioritize plant conservation efforts (Antonelli et al., 2020). It can be
245 applied to swiftly narrow the growing pool of DD species, address the growing backlog of species in need
246 of reassessment, and provide a preliminary pass for unassessed species. Further refinement of REBA, as
247 well as broader spatial and taxonomic applications, are necessary and welcomed. The need for action is
248 immediate—there is little time to waste.

249 Data Accessibility Statement

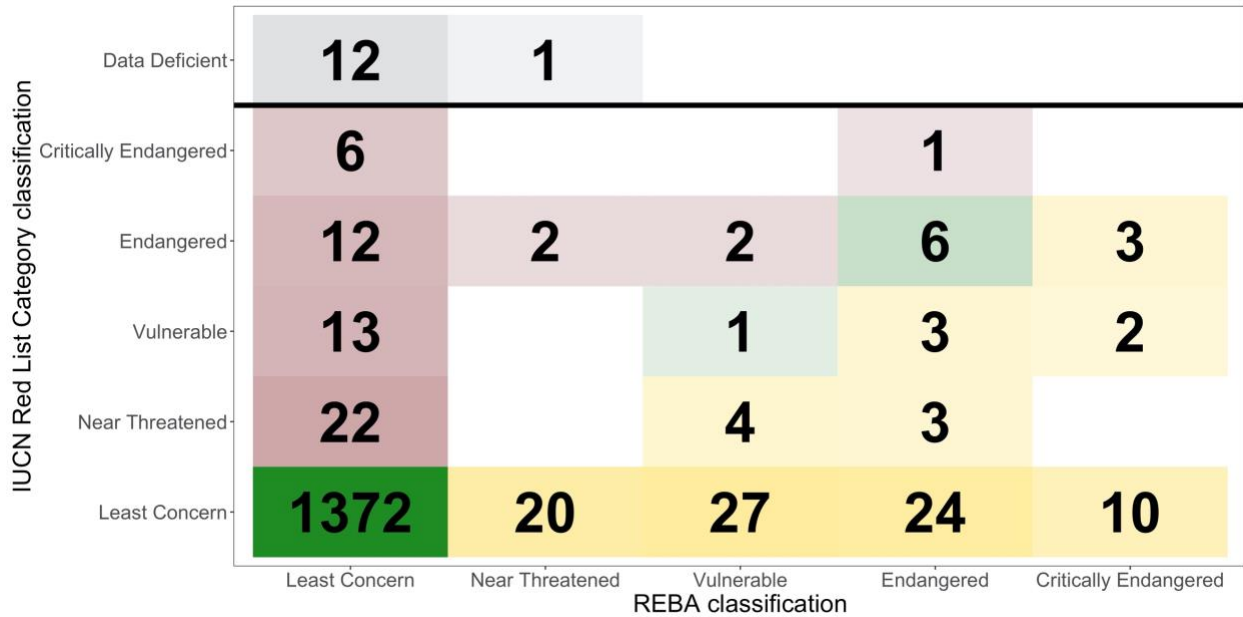
250 Data downloaded from the IUCN, along with R scripts used for GBIF data collection, filtering, analyses,
251 and visualization are freely available on GitHub: https://github.com/eveskew/plant_rapid_assessment.

252 Conflict of Interest

253 The authors declare no conflict of interest.

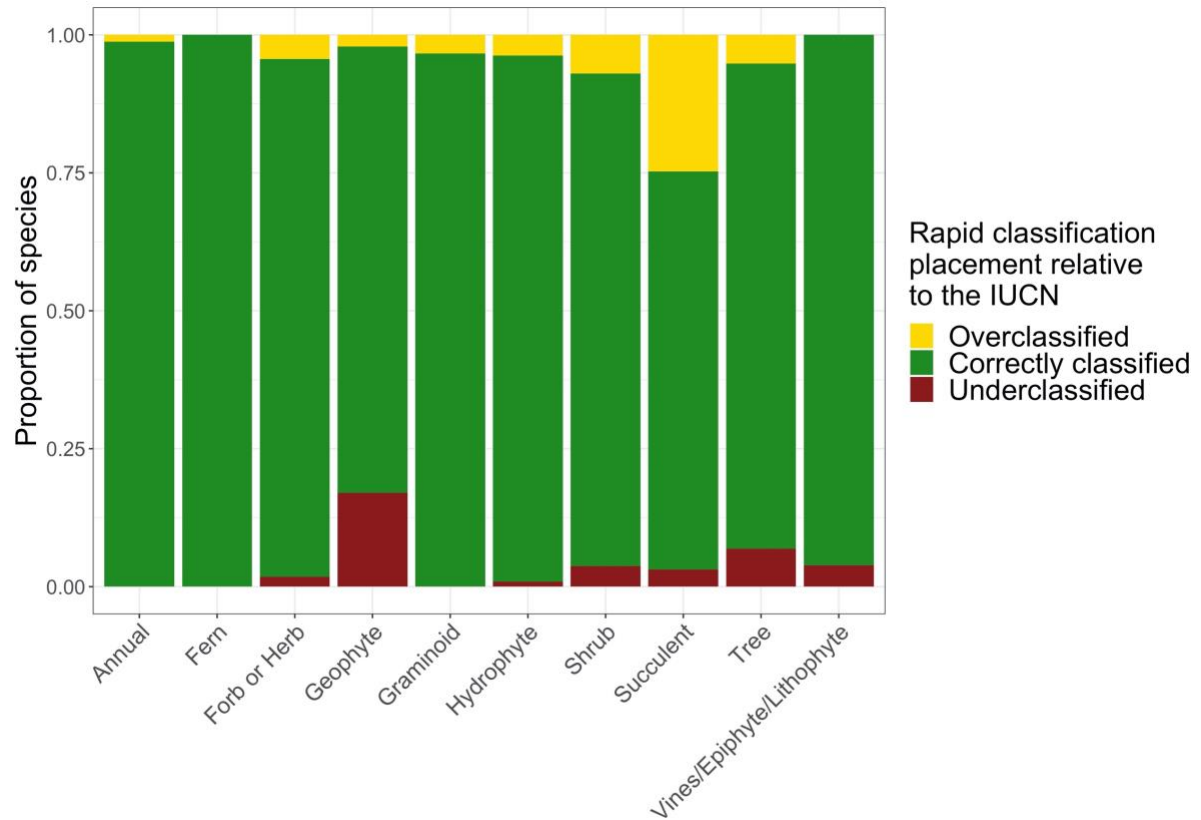
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255 **Fig. 1** Below the bold line, each tile represents an intersection of IUCN threat category
 256 classifications: those assigned by the IUCN and those assigned using REBA. Green tiles along
 257 the diagonal represent matching classifications, where both the IUCN and REBA classified
 258 species into the same categories. Yellow tiles below the diagonal are those we over-classified,
 259 where REBA placed species into a higher extinction risk category than that produced by the
 260 IUCN's classification. Red tiles above the diagonal represent those species we under-classified,
 261 where REBA placed species into a lower extinction risk category than that produced by the
 262 IUCN's classification. Tile transparency is a function of the number of species associated with
 263 that classification combination. Grey tiles above the bold line represent the threat categories into
 264 which we classified 13 DD species using REBA.



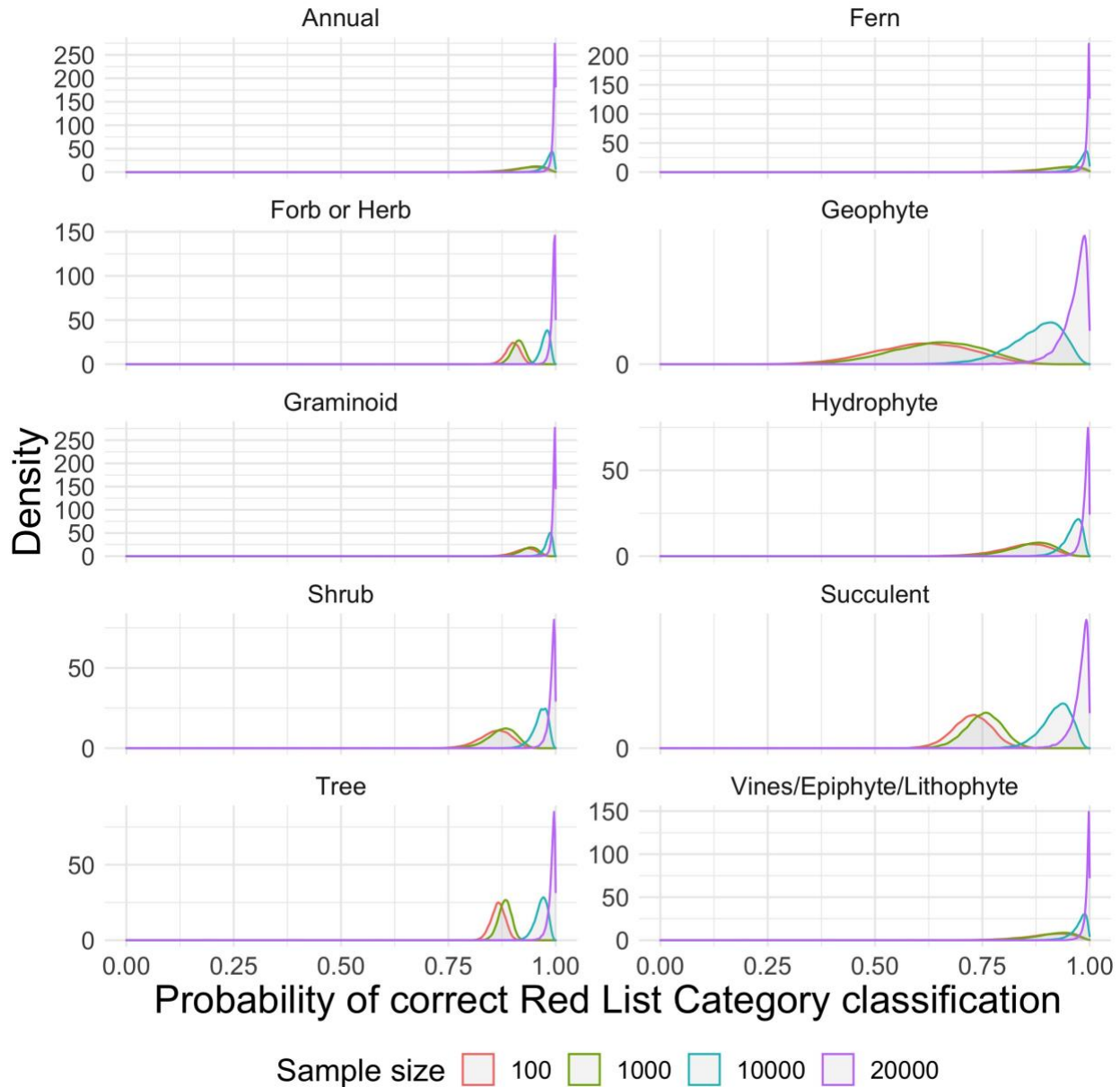
265
266

267 **Fig. 2** Each bar represents the total number of species assigned to a given 'Plant Type' by the
268 IUCN. The proportion of those species that were classified correctly, over-classified, and under-
269 classified are represented respectively by green, yellow, and red sections. It is important to
270 know that some species are counted across multiple plant types, as they were assigned more
271 than one by the IUCN (i.e., *Acer grandidentatum* is classified as both a 'Shrub' and 'Small
272 Tree'). See Table S1 for raw data.



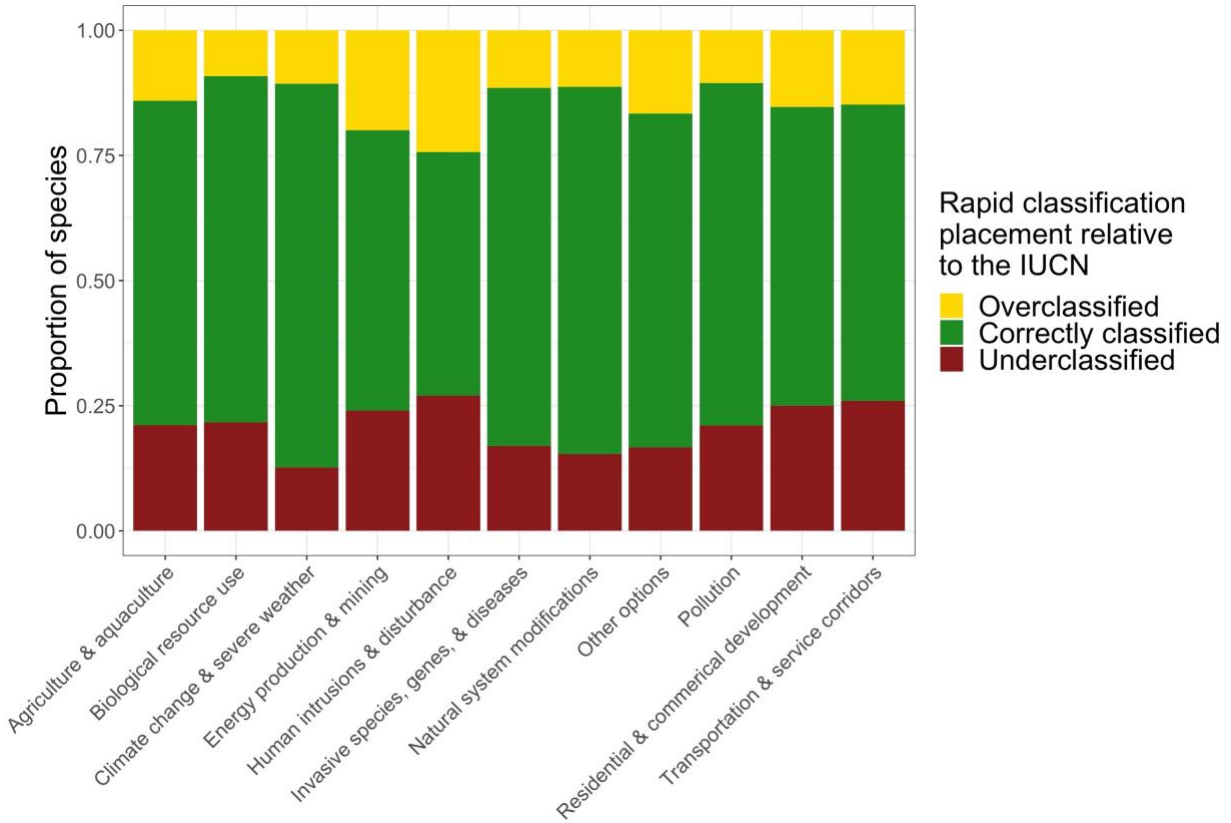
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275 **Fig. 3.** Posterior probability distributions representing the influence of IUCN assigned 'Plant
276 Type' on the probability of correct REBA classification at different sample sizes.



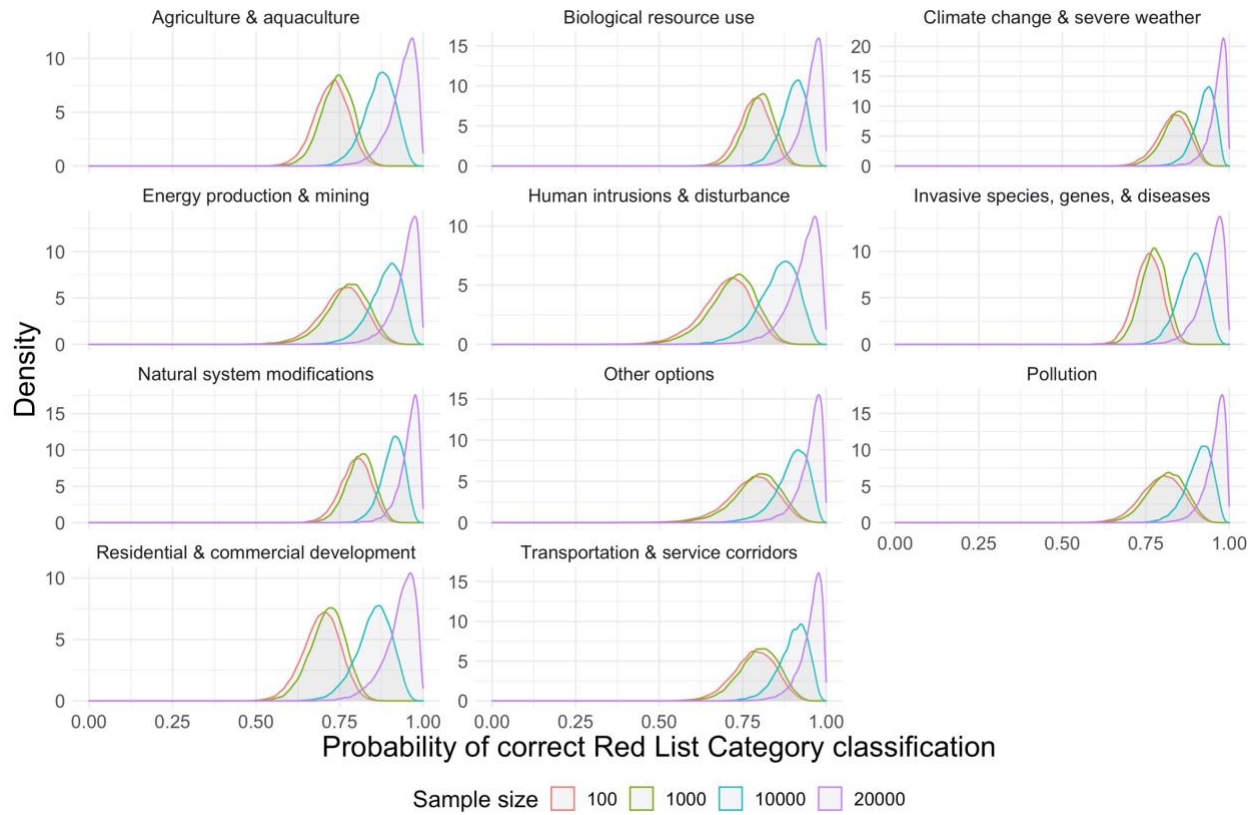
277
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279 **Fig. 4** Each bar represents the total number of species assigned to a given 'Threats' category
280 by the IUCN. The proportion of those species that were classified correctly, over-classified, and
281 under-classified are represented respectively by green, yellow, and red sections. It is important
282 to know that some species are counted across multiple threat types, as they were assigned
283 more than one by the IUCN (i.e., *Acer rubrum* is classified as threatened by 'Natural System
284 Modifications,' 'Invasive and Other Problematic Species, Genes, and Disease,' and 'Climate
285 Change and Severe Weather'). See Table S2 for raw data.



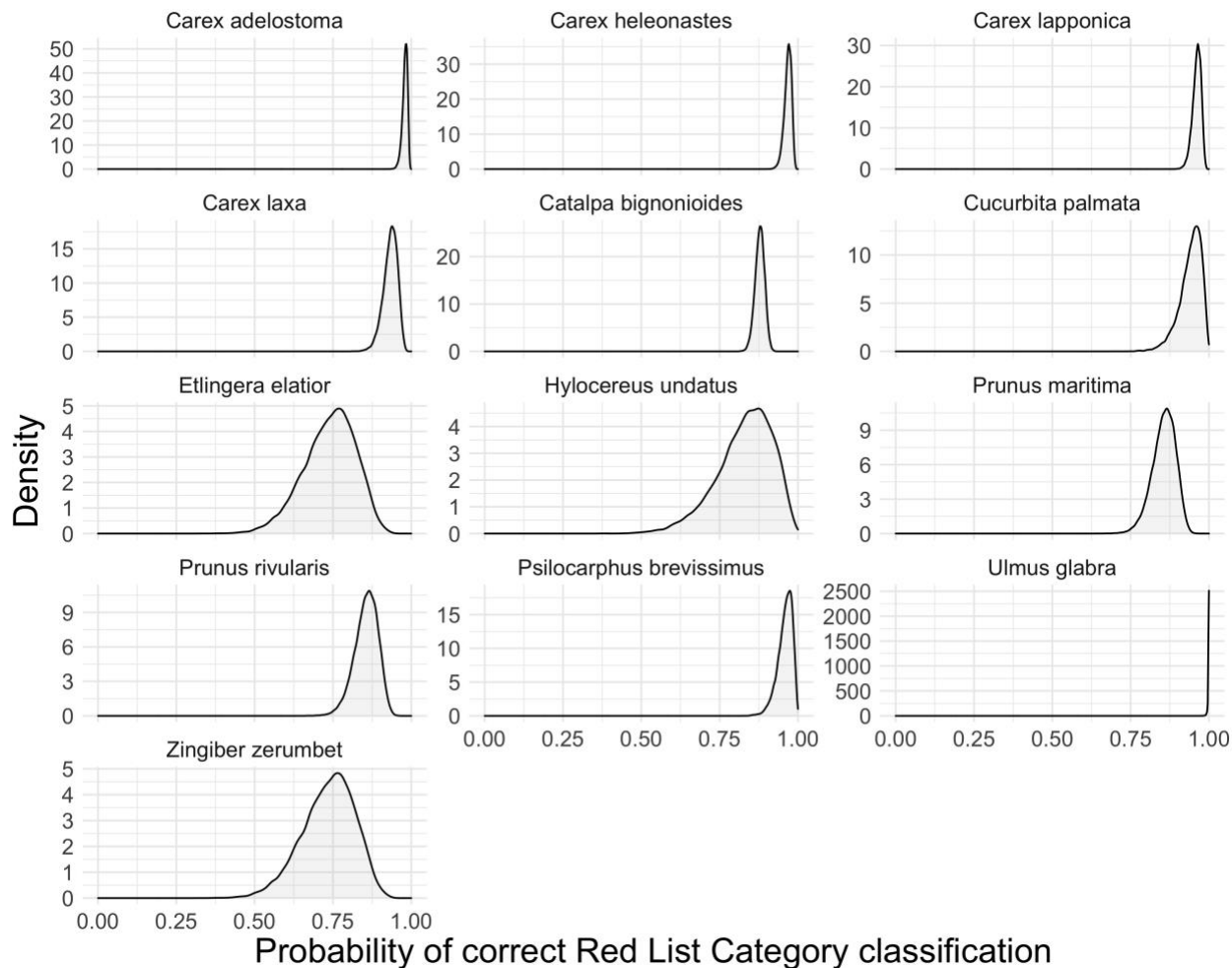
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287

288 **Fig. 5.** Posterior probability distributions representing the influence of IUCN assigned 'Threats'
289 on the probability of correct REBA classification at different sample sizes.



290
291

292 **Fig. 6.** Posterior probability distributions representing the probability of correct REBA
293 classification of 13 DD species. We employ the full posterior distributions of the 'Plant Type'
294 model to produce the implied probability of correct classification for each of these species based
295 on their 'Plant Type' and the actual number of cleaned occurrence points available for that
296 species. See Table S3 for raw data.



297
298

299

300 Appendix 1: Rationale for grouping of ‘Plant Types’

301 We modeled the probability of correct classification among different ‘Plant Types’, which were
302 previously defined by the IUCN (<https://www.iucnredlist.org/resources/classification-schemes>; Plant and
303 Fungal Growth Forms Classification Scheme). The 1,546 species used to validate the REBA classification
304 fell into 18 of the 24 IUCN ‘Plant Type’ categories. We combined some categories to improve statistical
305 analysis and visualization, and these combinations are justified both biologically and by IUCN’s ‘Plant
306 Type’ classification guidelines. For example, IUCN has multiple ‘Tree’ categories differentiated by size
307 (i.e. ‘Tree - size unknown’, ‘Tree - large’, ‘Tree - small’), but admits that the categories based on size are
308 “sub-types which may be dropped at some point in the future”, so we combined these categories into one
309 ‘Tree’ category. Sub-categories for both ‘Succulent’ and ‘Shrub’ categories were combined for similar
310 reasons. We combined the ‘Vine’, ‘Epiphyte’, and ‘Lithophyte’ categories due to low sample sizes, which
311 we thought was biologically reasonable because plants of these types often grow on atypical substrates
312 and tend to display a climbing habit.

313

314

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