

1 Are indigenous territories and community-managed areas effective natural climate solutions? A
2 neotropical analysis using matching methods and geographic discontinuity designs.

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25 Abstract:

26 Indigenous Territories (ITs) and Community Managed Protected Areas (PAs) with less
27 restriction on forest use than integral PAs may represent cost-effective natural climate solutions
28 to meet the Paris agreement. However, the literature has been limited to examining the effect of
29 ITs and Community Managed PAs on deforestation, despite the influence of anthropogenic
30 degradation. Thus, little is known about the temporal and spatial effect of allocating ITs and
31 Community Managed PAs on carbon stocks dynamics that account for losses from
32 deforestation and degradation. Using Amazon Basin countries and Panama at the national
33 level, and Petén (Guatemala) and Acre (Brazil) at the subnational level, this study aims to
34 estimate the temporal and spatial effects of ITs and PAs on carbon stocks. To estimate the
35 temporal effects, we use annual carbon density maps, matching analysis, and linear mixed
36 models. Furthermore, we explore the spatial biases derived from matching analysis and use
37 geographic discontinuity designs to assess the spatial effect of PAs and ITs boundaries on
38 carbon stocks. The temporal effects highlight that allocating ITs preserves carbon stocks and
39 buffer losses as PAs in Panama and Amazon Basin countries. Community Managed PAs
40 temporal effect on carbon stocks surpasses that of integral PAs in Petén (Guatemala) and Acre
41 (Brazil). The geographic discontinuity designs reveal that ITs and Community Managed PAs
42 boundaries secure more extensive carbon stocks than their surroundings, and this difference
43 tends to increase towards the least accessible areas. These results also suggest that
44 indigenous and community land-use in neotropical forests may have a limited and stable spatial
45 impact on carbon stocks. Our findings imply that ITs and Community Managed PAs in
46 neotropical forests support Nationally Determined Contributions (NDCs) under the Paris
47 Agreement. Thus, Indigenous peoples and local communities must become recipients of
48 countries' results-based payments.

49

50

51 Introduction

52 Avoided forest conversion and natural forest management are among the most cost-effective
53 natural climate solutions to meet the Paris Agreement (1). Protected Areas (PAs), cornerstones
54 of biodiversity conservation, may contribute to these cost-effective solutions by preventing
55 carbon stocks losses (2). However, since 1990, South America and Central America have
56 tripled the area of PAs (3) while simultaneously losing 10% and 25% of forest cover,
57 respectively (4). These forest conversion trends stress the need for additional natural climate
58 solutions that could reinforce the role of PAs. In Neotropical countries and across the globe,
59 Indigenous Territories (ITs) cover significant portions of natural lands with minimal human
60 disturbance and tend to overlap with PAs (5). More than 30% of the Amazon Basin forest's
61 aboveground carbon stocks are in ITs, and nearly 7% of these stocks are in areas overlapping
62 with PAs (Overlapped Areas, hereafter OAs) (6). Other locations with shared governance, such
63 as Community Managed PAs, also seem promising in climate change mitigation (7). Thus, ITs
64 and Community Managed PAs with fewer forest use restrictions and non-centralized
65 governance may conserve forests and potentially represent effective natural climate solutions.

66

67 However, the effect of ITs and PAs in forest conservation might be overestimated based on the
68 fact that they tend to be located in higher elevations, steeper slopes, and greater distances to
69 roads and cities than unprotected lands (8,9). To control for this non-random spatial location, an
70 increasing number of studies have relied on a statistical technique called matching analysis
71 (10,11). In these studies, matching analysis samples observations with similar geographical
72 characteristics, removing heterogeneous observations, and allowing to compare protected lands
73 with unprotected areas. For example, using matching analysis, ITs in the Brazilian Amazon

74 have been found to restrain high deforestation pressure more effectively than PAs (12).
75 Panama's PAs and claimed ITs more effectively avoided deforestation than unprotected lands
76 with similar topography and accessibility (13). Matching analysis also allowed identifying
77 decreased deforestation where ITs and other land tenures overlap (e.g., PAs) in Peru
78 (Anderson et al., 2018). Furthermore, Blackman & Veit (2018) concluded that ITs in the Amazon
79 Basin of Colombia, Bolivia and Brazil avoid carbon emissions from deforestation. Similarly,
80 Community Managed PAs, such as community concessions in Petén (Guatemala), and
81 sustainable use areas in Acre (Brazil), have been estimated to avoid deforestation, in contrast to
82 integral PAs that restrict sustainable use, which did not show a discernible effect relative to
83 unprotected lands (16,17). Therefore, controlling for spatial location using matching confirm that
84 ITs and Community Managed PAs are as effective as other PAs to avoid deforestation.

85
86 Despite the influence of anthropogenic degradation and recovery on forest conservation and
87 carbon stocks dynamics, research on matching analysis has been limited to examining the
88 effect of protected lands on deforestation. Shifting cultivation, considered a driver of degradation
89 (18), is common among tropical forest landholders (19). After long fallow periods (>20 years),
90 shifting cultivation can only recover around 50% of mature forests' carbon stocks (20). Logging
91 and fires, other causes of degradation in tropical forests, remove 45% and 22% of forest's
92 carbon stocks and take decades to recover (21). Thus, accounting for forest degradation and
93 recovery in temporal carbon stocks dynamics may influence the effectiveness of protected lands
94 in forest conservation, particularly in those with fewer use restrictions (e.g., ITs and Community
95 Managed PAs). However, little is known about the temporal effect of protected lands on carbon
96 stocks dynamics after controlling for spatial location.

97
98 Matching analysis controls for the spatial location, but it does not guarantee unambiguous
99 estimates of protected lands' effects in forest conservation. Karsenty et al. (22) highlight that

100 matching analysis implies weighting influence to particular deforestation (or degradation)
101 covariates, such as roads or rivers. The choice and omission of covariates influence the
102 observations sampled by matching, potentially excluding relevant areas, and altering the effect
103 attributed to a particular protected land (22). In this regard, some have recognized that sampling
104 through matching analysis might not be independent and exclude observations around the
105 boundaries of protected lands (23–25), rather than exploring the implications of sampling across
106 these boundaries. Conversely, the effect of protected lands' boundaries on deforestation has
107 been estimated through regression discontinuity designs. Bonilla-Mejía & Higuera-Mendieta (26)
108 found that ITs' boundaries are more effective than integral PAs at curbing deforestation in
109 Colombia. Similarly, Baragwanath & Bayi (27) established that ITs' boundaries with granted
110 property rights in Brazil decrease deforestation. However, few studies have used matching
111 analysis in geographic discontinuity designs, control for geographic distance among
112 observations (28), and estimate the effect of protected lands' boundaries on carbon stocks. Nor
113 have they addressed whether different forest use levels, such as ITs and PAs, imply different
114 spatial effects on carbon stocks.

115
116 This study builds upon previous research assessing the effect of protected lands on
117 deforestation through matching analysis and addresses some limitations of this methodology.
118 Using Panama and Amazon Basin Countries at the national scale, and Petén (Guatemala) and
119 Acre (Brazil) at the subnational scale, this study aims to estimate protected lands' temporal and
120 spatial effects on aboveground carbon stocks. The hypothesis is that protected lands with more
121 restrictions on forest use (e.g., PAs) will secure higher carbon stocks than less restrictive lands
122 (e.g., ITs and Community Managed PAs) over time and throughout their boundaries by reducing
123 the influence of anthropogenic degradation. Regardless of forest use restrictions and
124 governance, we find that protected lands preserve carbon stocks and buffer losses temporarily
125 and spatially across neotropical forests.

126 Our study makes three contributions to the literature. First, we provide a consistent use of
127 matching analysis in multiple protected lands and countries, allowing us to compare the effects
128 of ITs, OAs, and PAs across Central America and the Amazon Basin. Conversely, previous
129 studies have analyzed either multiple protected lands on a country scale (e.g., 14,17,18) or
130 single protected lands categories across regions (e.g., 11,16). Second, we use the temporal
131 dynamics of aboveground carbon stocks (2003 to 2016) instead of forest cover, thus making it
132 possible to estimate a more accurate temporal effect of protected lands in climate change
133 mitigation. Furthermore, we explore the spatial biases derived from matching analysis sampling
134 and account for them through geographic discontinuity designs, allowing us to assess the
135 spatial effect of protected lands' boundaries on carbon stocks. To our knowledge, this study is
136 among the first to estimate the effect of multiple protected lands on carbon stocks temporarily
137 (14 years) and spatially (throughout boundaries), providing a quantified estimate of forest
138 conservation and climate change mitigation across Neotropical Forests.

139

140 Methods

141 Geographic scope

142 This study emerged from the annual meeting of the "Red Amazónica de Información
143 Socioambiental Georeferenciada" RAISG (Amazon Georeferenced Socio-Environmental
144 Information Network) carried out in Quito (Ecuador) in August 2018. The authors belong to
145 diverse organizations (Academic, Research, International and National NGOs) that participate
146 or collaborate with RAISG. Additionally, some of the authors have also previously collaborated
147 with the "Coordinadora de las Organizaciones Indígenas de la Cuenca Amazónica" - COICA
148 (Coordinator of Indigenous Organizations of the Amazon River Basin), and the "Alianza
149 Mesoamericana de Pueblos y Bosques" - AMPB (Mesoamerican Alliance of Peoples and

150 Forests). These authors' collaborations and the aim to provide indigenous peoples,
151 communities, and countries with a quantified estimate of their forest conservation activities
152 define our study's geographical scope: the Department of Petén (Guatemala), Panama, and the
153 Amazon Basin portions from Colombia, Ecuador, Peru, and Brazil. The study sites cover an
154 area surpassing 10,000 million ha and contain a mosaic of protected lands that include PAs and
155 ITs.

156

157 Our study has two scales of analysis: the country-level and the subnational-level. The country-
158 level analysis includes Panama and Amazon Basin Countries (Fig 1). PAs at the country-level
159 encompass national and sub-national jurisdictions with governance by governments, private
160 governance, and shared governance that allow sustainable use from privates and communities
161 (Table 1). ITs without official titles or in the process of official recognition (i.e., claimed lands)
162 were also included at the country-level, except in Colombia, where the data was not available.
163 All ITs overlapping with PAs were defined as OAs. To capture the influence of different
164 governance categories of PAs, we analyzed at the subnational-level two case studies (Fig 2):
165 the Maya Biosphere Reserve in the Department of Petén (Guatemala) and the State of Acre
166 (Brazil). PAs were classified into three governance categories (Table 2): Community Managed
167 PAs, Sustainable Use PAs, and Integral PAs. Community Managed PAs corresponded to
168 Community Concessions from Petén (Guatemala) and Extractive Reserves from Acre (Brazil),
169 where local communities share forests' governance with governments. Sustainable Use PAs are
170 equivalent to IUCN Categories IV, V and VI areas that share governance with governments but
171 do not exclusively involve community management. Finally, Integral PAs corresponded to
172 National Parks with a limited human presence and emphasize on biodiversity conservation
173 equivalent to IUCN I-III. All private and public lands outside ITs, OAs and PAs were defined as
174 other lands in the national and subnational level analyses.

175

176 **Fig 1. National-level analysis.** Panama and the Amazon Basin portions of Colombia, Ecuador,
 177 Peru, and Brazil. Land tenure is classified as PAs (green), ITs (orange), OAs (yellow), and
 178 Other Land (grey).
 179

National Jurisdiction	Protected Lands	
	Protected Areas (PAs)	Indigenous Territories (ITs)
Panama	National Park Protective Forest Wildlife Refugee Multiple Use Area Forest Reserve Hydrological Reserve Zone of hydrological protection	Titled: "Comarcas" Titled: Collective Territories Claimed/Untitlled
Colombia	National Park National Protective Forest Reserve National Forest Reserve Civil Society Nature Reserve Fauna and Flora Sanctuary	Titled: Indigenous Reserve
Ecuador	National Park National Reserve Protective Forests Ecological Conservation Area Biological Reserve Ecological Reserve Fauna Production Reserve Wildlife Refugees	Titled Declared Claimed/Untitlled
Peru	National Park National Sanctuary Historical Sanctuary Protective Forest Landscape Reserve Communal Reserve Hunting Reserve Reserved Zone	Titled / Declared: Native community Titled/ Declared: Peasant community Claimed/Untitlled

Brazil	National Park	Titled/ Declared: Indigenous Area
	Environmental Protection Area	Titled/ Declared: Native Community
	Area of Relevant Ecological Interest	Titled / Declared: Indigenous Reserve
	Ecological Station	Titled / Declared: Indigenous Territory
	State Forest	Claimed/Untitlled
	State Park	
	Wildlife Refugee	
	Biological Reserve	
	Sustainable Use Reserve	
	Extractive Reserve	

180 **Table 1. Protected lands included at the national-level analysis.** PAs are equivalent to IUCN
181 categories I-VI. All areas in which ITs overlap with PAs are considered Overlapped Areas
182 (OAs).

183
184 **Fig 2. Subnational-level analysis.** (A) The department of Petén in Guatemala and the state of
185 Acre in the Brazilian Amazon. Protected Areas (PAs) in Petén (Guatemala) (B) and Acre (Brazil)
186 (C). PAs are classified into different governance categories: Community Managed PAs (light
187 green), Sustainable Use PAs (blue-green) and Integral PAs (dark green). Other PAs (white) in
188 Petén correspond to buffer zones and multiple-use areas. Protected Lands (Excluded) (black
189 lines) were not analyzed and correspond to PAs outside the Maya Biosphere Reserve in Petén
190 (Guatemala), and ITs in Acre (Brazil).

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Protected Areas (PAs)

Subnational Jurisdiction	Community Managed	Sustainable Use	Integral
Peten (Guatemala)	Community Concessions	Private Concessions and Cooperatives	National Parks Protected Biotopes
Acre (Brazil)	Extractive Reserves	National Forests	National Parks Ecological Stations

196 **Table 2. PAs included at the Subnational-level analysis.** PAs classified into different
197 governance categories in Petén (Guatemala) and Acre (Brazil).

198

199 Spatial data and processing

200 The boundaries of protected lands were curated by ACOFOP (Asociación de Comunidades
201 Forestales de Petén) for Petén (Guatemala); the Neotropical Ecology Laboratory (McGill
202 University, Smithsonian Tropical Research Institute) for Panama; and RAISG (Amazon Geo-
203 referenced Socio-Environmental Information Network) in the case of Amazon Basin Countries.

204

205 We used Annual carbon density maps based on raster data (~500 m resolution) that was
206 generated by the Woodwell Climate Research Centre between 2003 and 2016 and explained in
207 detail by Baccini et al. (29,30) and Walker et al. (6). These estimations derive from combining
208 LiDAR data and field measurements that calibrate a machine learning algorithm that generates
209 annual carbon density estimates from MODIS satellite imagery. These carbon density maps can
210 detect annual losses and gains in carbon density, aggregating changes from deforestation,
211 forest degradation, and recovery.

212

213 Elevation, slope and the distance to roads, settlements and rivers were included as covariates
214 to establish the spatial location conditions associated with annual carbon density across

215 national and subnational jurisdictions (S1 Appendix). Elevation and slope were obtained from
216 the satellite imagery of the SRTM (Shuttle Radar Topographic Mission - Arc Second Global).
217 The distance to roads was calculated from geospatial data produced by national institutions in
218 Petén and Panama (S1 Appendix). Road distance corresponding to Amazon Basin countries
219 was based on the geospatial data curated by RAISG. The distances to rivers and settlements (>
220 5000 people) were calculated from geospatial data produced by national institutions. Land
221 tenure and covariate data were resampled to the spatial resolution of carbon density, creating
222 observation units of ~500-m resolution across different land tenures with estimates for
223 covariates and carbon density. All geoprocessing was performed in ArcGIS (ESRI, 2018).
224 Finally, we established protected lands' non-random spatial location by estimating the mean
225 covariate differences between other lands and protected lands in each jurisdiction using Mann
226 Whitney tests (S2 Appendix).

227

228 Coarsened exact matching

229 We performed matching analysis and linear mixed models to control for spatial location and
230 infer the effect of PAs, ITs and OAs (i.e., protected lands) on carbon stocks relative to other
231 lands (Fig 3). Matching analysis preprocesses datasets to reduce the association of a treatment
232 variable with covariates by removing heterogeneous observations and creating a subset of
233 treatment and control observation units with similar covariate values (31). Here, the treatment
234 variable corresponded to land tenure, and matching created subsets of observation units of
235 ~500 m resolution in protected lands and other lands with similar slope, elevation and distance
236 to roads, towns and rivers. To account for the size and heterogeneity of the Brazilian Amazon,
237 we included the states as covariates in this country.

238

239 **Fig 3. Workflow to infer the temporal and spatial effect of ITs, OAs, and PAs on carbon**
240 **stocks.**

241
242 Specifically, we used coarsened exact matching (CEM) (32) with the R package *MatchIt* (33) for
243 each protected land across the national and subnational levels analyses. Following steps from
244 Iacus et al. (34), we first defined coarsening choices for each covariate (S3 Appendix). For
245 example, the elevation was coarsened in multiple categories based on 100 meters intervals.
246 This coarsening choice meant that protected lands and other lands with elevation values
247 between 900 and 1000 m were considered "equivalent". Then, CEM located control and
248 treatment observation units in matching sub-groups with equivalent coarsened values for all
249 covariates. The third step pruned matching sub-groups that did not have at least one treatment
250 and one control observation with equivalent coarsened covariate values. These steps were
251 reiterative until the coarsening choices produced standardized mean differences between
252 protected lands and other lands below 5% for all covariates (S4 Appendix).

253
254 After isolating the effect of spatial location through matching, we made temporal estimates
255 regarding the effect of allocating protected lands on carbon stocks at national and subnational
256 levels. This effect was calculated using linear mixed models in the R package *lme4* (35). Annual
257 carbon density was the outcome variable, and land tenure (protected land or other land) a fixed
258 effect. Additionally, elevation, slope and the distance to roads, towns, and rivers were also
259 defined as fixed effects, spanning any remaining imbalances from the matched subsets. The
260 matched sub-group (matched observation units between protected lands and other lands with
261 similar covariate values) was a random effect to account for the structure of matched subsets.
262 These linear mixed models were estimated annually between 2003 and 2016 in all study sites,
263 for each protected land category separately. Two parameters derived from the linear mixed
264 models were used to determine the effect of protected lands on carbon stocks after controlling

265 for spatial location: the intercept and land tenure. The intercept or the annual carbon density
266 found in other lands represents the carbon stocks baseline for protected lands. The land
267 tenure's effect refers to the annual average differences of carbon stocks between protected
268 lands and other lands after matching, defined as the temporal effect.

269

270 Geographic discontinuity designs

271 After calculating the distance of matched observation units around the boundaries of protected
272 lands, we explored geographic discontinuity designs to estimate the spatial effect of
273 administrative boundaries (36). Specifically, we assessed how protected lands' boundaries
274 influence carbon stocks compared with other neighbouring lands. Our geographic discontinuity
275 designs followed Keele et al. (28), which uses matching methods to find treatment and control
276 observation units with similar covariates, including the geographic distance between observation
277 units. To implement these designs, we created subsets of observation units with buffer zones
278 inside and outside of protected lands boundaries, of 0–1 km, 0–5 km, 0–10 km, and 0–15 km.
279 The boundaries between protected lands (e.g., ITs and PAs) were not tested except in Petén
280 (Guatemala). Some protected lands in the Maya Biosphere Reserve (Petén, Guatemala), such
281 as community concessions (Community Managed PAs) and National Parks (Integral PAs), do
282 not share boundaries with other lands. Therefore, the geographic discontinuity designs in Petén
283 (Guatemala) compared the boundaries of Community Managed PAs with Integral PAs,
284 Sustainable Use PAs, and Other protected lands (i.e., buffer zones and multiples use zones).

285

286 After defining the geographic discontinuity designs, we also performed CEM with the buffer
287 zones subsets, including slope, elevation and distance to roads, towns and rivers as covariates.
288 Additionally, we controlled for the geographic distance among observation units according to
289 buffer zones. For example, for buffer zones 0–1 km, we included matches across a 2-km radius,

290 and for 0–15-km buffer zones, a 30-km radius. The differences between average carbon stocks
291 stored inside and outside the boundaries of protected lands, or the spatial effect, was also
292 estimated through the linear mixed models aforementioned. The geographic discontinuity
293 designs estimated the effect of protected lands' boundaries on carbon stocks in 2003 and 2016.
294

295 Results

296 The temporal effect of protected lands on carbon stocks across 297 neotropical countries

298 Matching analysis controlled the influence of spatial location, allowing to estimate the temporal
299 effect of allocating protected lands on carbon stocks. This temporal effect represents the annual
300 mean difference of carbon stocks between protected lands and other lands (Fig 4). Across
301 Panama and Amazon Basin countries, the carbon stocks from 2003 to 2016 in protected lands
302 were usually higher than other lands (i.e. the baseline), resulting in positive temporal effects.
303 Country-level comparisons of temporal effects reveal three regional patterns. Protected lands in
304 Panama had low carbon stocks baselines in other lands (< 65 t C/ha) and substantial temporal
305 effects that represented an increase in carbon stocks above 30%. Brazil's protected lands
306 displayed moderate baselines (< 115 t C/ha) and temporal effects (< 18%). The carbon stocks
307 baselines in western Amazon Basin countries exceeded those of Brazil (> 115 t C/ha), while the
308 temporal effects were moderate (< 10%). Hence, the temporal effects seem substantial in
309 countries with reduced carbon stocks in other lands.

310

311 **Fig 4. The temporal effects of protected lands on aboveground carbon stocks across**
312 **neotropical countries in 2003 and 2016.** Significant temporal effects ($p < 0.05$) are
313 represented as colored bars and percentages, indicating the additional/fewer carbon stocks

314 secured by allocating ITs (orange), OAs (yellow), and PAs (green) relative to the baseline
315 (Other Lands, grey) after controlling for spatial location. Error bars reflect 95% confidence
316 intervals for the baselines and temporal effects.

317
318 The positive temporal effects also reveal the additional amount of carbon stocks secured by
319 allocating protected lands in a particular year across Panama and Amazon Basin countries
320 (Figure 4). During 2003, PAs in Panama secured 95% (37 t C/ha) larger carbon stocks than
321 their baseline (39 t C/ha). Relative to more substantial baselines (> 55 t C/ha), Panama's IT's
322 and OA's accounted for 35% (19 t C/ha) and 71% (44 t C/ha) additional carbon stocks. Similar
323 to Panama, protected lands in Amazon Basin countries represented positive temporal effects in
324 2003. Brazil's ITs and PAs represented 6% (~ 6 t C/ha, respectively) additional carbon stocks
325 compared to their baselines (~ 105 t C/ha), and this effect nearly doubled in OAs (12%, 14 t
326 C/ha). Western Amazon Basin countries displayed similar temporal effects in 2003, ranging
327 between 5–5.7% (i.e., 5 - 7 t C/ha) in PAs from Peru and Colombia, 3.5–5.7 % (i.e., 5 - 7 t C/ha)
328 in PAs from the same countries, and 0.7– 4 % (i.e., 0.5 - 5 t C/ha) in OAs from Colombia and
329 Ecuador. Despite regional differences, these results suggest that in 2003 OAs and ITs had a
330 similar effect on carbon stocks compared to PAs in neotropical countries.

331
332 Overall, protected lands' temporal effects on carbon stocks remained stable or increased
333 relative to other lands until 2016 (Fig 4, S5 Appendix). These effects remained stable in PAs
334 and ITs from Ecuador and did not vary more than 0.5%. ITs in other Amazon Basin countries
335 exhibited increases in temporal effects, reaching between $\sim 3\%$ (4 t C/ha) in Peru and $\sim 10\%$ (10
336 t C/ha) in Brazil. Similarly, Amazon basin PAs had increases that resulted in temporal effects
337 between $\sim 4\%$ (~ 11 t C/ha) and $\sim 9.1\%$ (9.5 t C/ha) for Peru and Colombia, respectively. The
338 temporal effects considerably varied in Amazon Basin OAs during 2016, showing no differences
339 with the baseline in Colombia and the largest increase in Brazil (17.2%, 19 t C/ha). Conversely,

340 protected lands in Panama experienced decreases in temporal effects ($> -5\%$) that seem to be
341 driven by the recovery of carbon stocks in other lands (S6 Appendix). Thus, stable and
342 increasing temporal effects reflect that allocating protected lands buffered losses and secured
343 the stability of carbon stocks relative to the other lands. Furthermore, these results reveal that
344 indigenous lands (ITs and OAs) and PAs secured similar amounts of carbon stocks until 2016.
345
346 The subnational-level analysis in Petén (Guatemala) and Acre (Brazil) estimated the temporal
347 effects on carbon stocks of PAs with varied forest governance and restrictions on forest use (Fig
348 5, S7 Appendix) relative to other lands (S8 Appendix). Community Managed PAs in Petén
349 (Guatemala) (i.e., Community Concessions) had carbon stocks $\sim 60\%$ (28 t C/ha) larger than
350 other lands (95.5 t C/ha) in 2003, and this effect increased to 78% (34 t C/ha) in 2016. Acre's
351 Community Managed PAs (i.e., Extractive reserves) also registered in 2003 larger carbon
352 stocks (6.2%, 8 t C/ha) than other lands (109.4 t C/ha) and an increase in 2016 (9.1%, 11 t
353 C/ha). These temporal effects of Community Managed areas were always greater than
354 Sustainable Use PAs in Petén (Guatemala) and Acre (Brazil). Community Managed PAs at
355 least doubled the temporal effect on carbon stocks of Private Concessions and Cooperatives
356 (Petén, Guatemala), and National Forests (Acre, Brazil). The differences in temporal effects
357 were narrower between Integral PAs and Community Managed PAs. During 2016, Integral PAs
358 exhibited an effect of $\sim 58\%$ (25t C/ha) in Petén (Guatemala) and $\sim 7\%$ (10 t C/ha) in Acre
359 (Brazil). These findings indicate that between 2003 and 2016, Community Managed PAs had a
360 larger effect on carbon stocks than Sustainable Use PAs and integral PAs and highlight the
361 variation among PAs at a subnational scale.

362

363

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365

366 **Fig 5. The temporal effects of PAs in Petén (Guatemala) and Acre (Brazil) on carbon**
367 **stocks in 2003 and 2016.** Significant temporal effects ($p < 0.05$) are represented as colored
368 bars and percentages, indicating the additional/fewer carbon stocks secured by allocating
369 Community Managed PAs (light green), Sustainable Use PAs (cyan), and Integral PAs (dark
370 green) relative to the Baseline (Other Lands, grey) after controlling for spatial location. Error
371 bars reflect 95% confidence intervals for the baselines and temporal effects.

372

373 **Insight at a finer scale: The role of protected lands' boundaries on** 374 **carbon stocks**

375 To identify the spatial implications of matching analysis in quantifying forest conservation, we
376 estimated the distance of observation units to the boundaries of protected lands (Fig 6, Table 3).
377 Matched observation units in protected lands had a range of average distances to their
378 boundaries, between 1.3 km (± 2.26) in PAs from Ecuador and 10.15 km (± 11.70) in PAs from
379 Peru. The distance of matched observation units in other lands to the boundaries of protected
380 lands ranged between 3.10 km (± 3.13) (Ecuador) and 9.52 km (± 7.72) (Panama). Not
381 surprisingly, the spatial distributions imply that observations along the boundaries of protected
382 lands are more likely to share spatial features (i.e., elevation, slope, and distance to roads,
383 towns, and rivers). In the case of observations in protected lands, these sampling outcome
384 suggests that matching analysis selects the most accessible areas, omitting the core and
385 possibly more intact forests. Thus, the spatial distribution from matching indicates that the
386 aforementioned temporal effects of protected lands are conservative.

387

388 **Figure 3. Observation units sampled through matching analysis in protected lands from**
389 **Panama and Acre.** (A). Panama is National-level analysis, including PAs, ITs, and OAs. (B).
390 Acre (Brazil) is a subnational-level analysis and includes different PAs categories.

391

Jurisdiction	Land Tenure	Mean distance of Matched Observation Units to Protected Lands' boundaries (km)	SD
Panama	Other Lands	9.51	7.72
	PAs	1.04	1.41
	ITs	2.37	2.99
	OAs	2.25	2.75
Colombia	Other Lands	10.57	10.70
	PAs	6.32	5.25
	ITs	9.35	1.34
	OAs	8.69	7.55
Ecuador	Other Lands	3.10	3.13
	PAs	1.30	1.55
	ITs	1.39	2.25
	OAs	1.48	1.80
Peru	Other Lands	6.57	6.72
	PAs	10.15	11.70
	ITs	1.94	3.24
	OAs	6.37	5.62
Brazil	Other Lands	6.19	4.26
	PAs	6.12	4.25
	ITs	6.11	4.25
	OAs	5.86	4.20
Peten (Guatemala)	Other Lands	42.19	29.34
	Strict PAs	3.70	2.89
	Community Managed PAs	2.89	2.35
	Sustainable Use PAs	4.01	2.52
Acre (Brazil)	Other Lands	5.81	10.31
	Strict PAs	4.72	4.92
	Community Managed PAs	2.47	2.99
	Sustainable Use PAs	1.37	1.53

392

393 **Table 3. Mean distance to protected lands' boundaries of observation units sampled**
 394 **through matching analysis by jurisdiction and land tenure.**

395

396 Considering the spatial distribution of matched observations, we performed geographic
397 discontinuity designs to understand how carbon stocks varied spatially throughout the
398 boundaries of protected lands in 2003 and 2016. The geographic discontinuity designs estimate
399 spatial effects. That is, the mean differences of carbon stocks inside and outside protected
400 lands for various distances around their boundaries, after controlling for spatial location. Overall,
401 the geographic discontinuity designs show that carbon stocks increase inside the boundaries of
402 protected lands in 2003 and 2016 (Figure 7). As discussed below, the geographic discontinuity
403 designs reveal spatial and spatial-temporal patterns across protected lands.

404

405 **Figure 7. The spatial effect of protected lands' boundaries on carbon stocks during 2003**
406 **and 2016 in neotropical countries.** Significant temporal effects ($p < 0.05$) are represented as
407 points and percentages, indicating the additional/fewer carbon stocks secured inside the
408 boundaries of ITs (orange), OAs (yellow), and PAs (green) relative to surrounding lands at a
409 certain buffer distance. Empty points and dashed lines represent the spatial effects in 2003,
410 while full points and continuous lines the spatial effects in 2016. The values in parentheses
411 represent the percentual increase/decrease in spatial effects between 2003 and 2016. Error
412 bars reflect 95% confidence intervals for the temporal effects.

413

414 The spatial patterns of geographic discontinuity designs exhibit how protected lands influence
415 carbon stocks within their boundaries. We found that the spatial effects of protected lands tend
416 to increase with the buffer distance to boundaries, and they are more pronounced in the first 5
417 km (Figure 7, S9 Appendix). For instance, ITs from Brazil in 2016 had carbon stocks 10.3% (21
418 t C/ha) larger than surrounding areas (102 t C/ha) when comparing a 1 km buffer. This spatial
419 effect increased to 15% (27 t C/ha) at 5 km, 17% (~30 t C/ha) at 10 km, and 19% (~34 t C/ha) at
420 15 km. ITs in Panama and western Amazon Basin countries displayed a similar spatial effect.
421 Except for Peru, OAs also had increasing spatial effects, and their influence on carbon stocks

422 exceeded that of ITs and PAs. For example, OAs' carbon stocks in Colombia did not differ from
423 surrounding areas at 1km (120 t C/ha) in 2016 but had a spatial effect on carbon stocks of 2.5%
424 (~7 t C/ha) at 5 km, which is over five times higher than ITs' and PAs' effect in the same country.
425 The spatial influence of PAs varied across countries. Relative to 10 km buffer comparisons, PAs
426 spatial effects on carbon stocks reduce at 15 km in Brazil and Peru. At 1 and 5 km buffers,
427 Colombia's PAs had 0.80% and 0.46% fewer carbon stocks than surrounding lands,
428 respectively. These resulting spatial patterns imply that allocating ITs and OAs generate
429 boundaries that effectively conserve carbon stocks as PAs. Furthermore, the increasing effects
430 on carbon stocks along with the distance to boundaries, more frequent in ITs and OAs, indicate
431 that protected lands shape forest landscapes by preserving the core and least accessible areas.
432

433 A spatial-temporal comparison of geographic discontinuities between 2003 and 2016 may
434 indicate whether the boundaries of protected lands bring stability to carbon stocks. We found
435 that from 2003 to 2016, the differences of carbon stocks inside and outside protected lands
436 increased, except for ITs in Colombia (Figure 7). Colombia's ITs secured larger carbon stocks
437 within their boundaries in 2016, but their spatial effect reduced 0.2% at 5km and 10km,
438 potentially driven by a recovery in surrounding areas (S9 Appendix). The most substantial
439 increases in spatial effects occurred among OAs. In Brazil, OAs spatial effect on carbon stocks
440 increased by 11% (~34 to 53 t C/ha) at 15 km in 2016, while ITs and PAs by 5.4% and 3.7%
441 respectively. Similarly, Ecuador's OAs increased their spatial effects on carbon stocks 2.2% at
442 15km, contrasting national PAs (0.6%) and ITs (0.2%). These increases between 2003 and
443 2016 in spatial effects suggest carbon stocks losses in surrounding areas that were buffered
444 inside the boundaries of protected lands, especially in OAs.

445
446 The geographic discontinuity designs in multiple categories of PAs from Acre (Brazil) and Petén
447 (Guatemala) represent different geographic settings (Figure 8, S10 Appendix). As with country-

448 level results, geographic discontinuities in Acre (Brazil) compared carbon stocks inside and
449 outside PAs' boundaries. At 5km buffers from their boundaries, Community Managed PAs had
450 8.4% larger carbon stocks than surrounding lands (126 t C/ha) at 5km from their boundaries in
451 2016. Acre's Sustainable Use PAs and Integral PAs had a more moderate effect, exhibiting a
452 7.2% and 5.2% increase of carbon stocks inside their boundaries at the same distance and
453 during the same year. Moreover, the spatial effects of Community Managed and Sustainable
454 Use PAs consistently increased with the buffer distance to boundaries, contrasting Integral PAs.
455 The spatial effects in Acre (Brazil) during 2016 exceeded those of 2003, implying the stability of
456 carbon stocks within all PA's boundaries and a decrease outside in other lands (S10 Appendix).
457 Petén (Guatemala) represented another geographic setting where PAs share boundaries inside
458 the Maya Biosphere Reserve. After comparing the boundaries of Community Managed with
459 other PAs categories, at 5 km buffers, we found that these areas stored 3% more carbon than
460 integral PAs, but 0.82% fewer than sustainable use PAs. Compared with 2003, these spatial
461 effects reduced, partially driven by the recovery of PAs surrounding Community Managed PAs
462 (S10 Appendix). Despite the different geographic settings of Acre (Brazil) and Peten
463 (Guatemala), these case studies suggest that the boundaries of Community Managed PAs
464 secure higher carbon stocks than Integral PAs.

465

466 **Figure 8. The spatial effect of PAs' boundaries on carbon stocks during 2016 in Petén**
467 **(Guatemala) and Acre (Brazil).** Significant temporal effects ($p < 0.05$) are represented as
468 points and percentages, indicating the additional/fewer carbon stocks secured inside the
469 boundaries of Community Managed PAs (light green), Sustainable Use PAs (cyan), and Strict
470 PAs (dark green) relative to surrounding lands at a certain buffer distance. Empty points and
471 dashed lines represent the spatial effects in 2003, while full points and continuous lines the
472 spatial effects in 2016. The values in parentheses represent the percentual increase/decrease
473 in spatial effects between 2003 and 2016. Error bars reflect 95% confidence intervals for the

474 temporal effects. The spatial effect in the Maya Biosphere Reserve (Petén) compares
475 Community Managed PAs with other Protected Lands, Sustainable Use PAs, and Integral PAs.
476 PAs in Acre are compared with Other Lands.

477

478 Discussion

479 In this study, we aim to estimate the temporal and spatial effects of allocating protected lands,
480 namely ITs, PAs and OAs, on carbon stocks across Neotropical Forests from Central America
481 and the Amazon Basin. Considering that these protected lands presumably experience low
482 anthropogenic pressures of forest loss, we control the effect of spatial location. Contrary to our
483 hypothesis, ITs, OAs, and Community Managed PAs generally preserve carbon stocks and
484 buffer losses as much as other PAs with more forest use restrictions. Over time, these protected
485 lands secure more stable and higher carbon stocks than other lands between 2003 and 2016.
486 Spatially, the geographic discontinuity designs show that carbon stocks increase inside the
487 boundaries of protected lands. These temporal and spatial effects were conservative and had
488 varied patterns across protected lands and jurisdictions.

489

490 The effectiveness of protected lands in conserving forests and 491 carbon stocks

492
493 Our findings highlight the need for a "spatially explicit" understanding of matching analysis
494 regarding protected lands and forest conservation. Other studies recognize the spatial biases of
495 matching analysis and incorporate more "spatially explicit" methodologies. Gaveau et al. (2013),
496 for example, provides the spatial distribution of matched observation units among timber
497 concessions, PAs and oil palm concessions in Kalimantan (Indonesia). Bowker et al. (2017) in
498 Africa and Zhao et al. (24) in China exclude from matching analysis other lands in a 10-km
499 buffer around PAs. These studies attempt to avoid spatial autocorrelation by controlling

500 sampling distance, while Negret et al. (37), test different post-matching models to control this
501 bias and assess avoided deforestation in PAs from Colombia. Other studies use regression
502 discontinuity designs to isolate some effects of spatial location and test the role protected lands'
503 boundaries (26,27). Our study presents an integrated approach. On the one hand, the temporal
504 effect resembles matching methods that are not spatially explicit on sampled observation units
505 (10,15,17,38). After exploring the spatial distribution of matched observation units, our findings
506 point that they are biased towards geographic boundaries, causing conservative estimates
507 about protected lands. On the other hand, we use geographic discontinuity designs with
508 matching analysis to directly control for spatial location and the geographic distance among
509 observations, generating valid counterfactuals inside and outside protected lands' boundaries
510 and maintaining conservative estimates (28). Hence, our study makes a novel methodological
511 contribution to research by integrating matching analysis and geographic discontinuity designs
512 to test the effectiveness of PAs' and ITs' boundaries in conserving carbon stocks across
513 neotropical countries.

514

515 By using conservative estimates, our findings support growing evidence indicating that ITs and
516 Community Managed PAs can be as effective as other PAs in forest conservation (6,16,39,40).
517 After controlling for spatial location, we found that allocating indigenous lands (i.e., ITs and OAs)
518 secured similar or even larger carbon stocks than PAs between 2003 and 2016 in Panama and
519 Amazon Basin countries. These findings are in line with Nolte et al. (12), who showed that
520 indigenous lands (ITs and OAs) are more effective than PAs at curbing deforestation pressure
521 in Brazil. By comparing indigenous lands (ITs and OAs) and PAs, our findings complement
522 Blackman & Veit's (15) estimates of avoided emissions from deforestation in ITs from Colombia
523 and Brazil (15). However, they did not detect a discernible effect from Ecuador's ITs, while our
524 results estimated a positive effect on carbon stocks. Similarly, our results from Panama, where
525 OAs had the most considerable effect on carbon stocks, partially contrast another study where

526 PAs were the most effective in avoiding deforestation (13). These differences with previous
527 studies might be attributable to our outcome variable (annual carbon stocks) that integrates
528 deforestation, degradation, and recovery. Estimating carbon stocks changes offer more
529 accurate estimates regarding the effectiveness of protected lands, especially in countries where
530 degradation emissions equal or exceed those from deforestation (e.g., Colombia, Ecuador, and
531 Peru) (6). Thus, our results demonstrate that indigenous governance or shared governance of
532 forests is as effective as state-based governance in conserving carbon stocks, suggesting that
533 titling ITs and formalizing shared governance in PAs represent a significant impact for avoided
534 land-use emissions in neotropical countries.

535

536 Similarly, Community Managed PAs, here, Community Concessions (Petén, Guatemala) and
537 Extractive Reserves (Acre, Brazil), were more effective in conserving carbon stocks than other
538 PAs with sustainable use and strict protection between 2003 and 2016. Regarding the
539 effectiveness of Community Concessions over Integral PAs, our results are consistent with
540 Blackman's (16) estimates of avoided deforestation. Additionally, we established that
541 Community Concessions are more effective than Private Concessions and Cooperatives in
542 preserving carbon stocks between 2003 and 2016. Other studies also established that PAs with
543 sustainable use in Acre (Brazil) significantly reduce deforestation (17). According to our results,
544 the greatest impact in avoiding the loss of carbon stocks is found in Extractive Reserves, while
545 the impact of National Forests was marginal. Hence, our results indicate that Community
546 Managed PAs, as ITs and OAs, actively avoid the loss of carbon stocks and bring stability to
547 forests. Moreover, our results imply that allocating Community Managed PAs, ITs, and OAs,
548 while providing material and cultural benefits to their inhabitants, can have a pivotal role in
549 climate change mitigation as other PAs.

550

551 Our geographic discontinuity designs provide conservative estimates regarding protected land's
552 effect on carbon stocks within their boundaries. Although the assessments of PA's boundaries is
553 common in the literature (41), they do not control for spatial location or compare different
554 categories protected lands. Overall, our findings indicate that carbon stocks increase inside the
555 boundaries of protected lands. However, the spatial effects are variable among PAs. For
556 example, PAs from Colombia seem only to avoid carbon stock losses more than 5 km inside
557 their boundaries in 2003 and 2016. These spatial patterns are not due to recent anthropogenic
558 pressures and confirm the inability of PAs' to reduce forest loss inside their boundaries
559 (26,42,43). Conversely, our results show that ITs, OAs, and Community Managed PAs tend to
560 secure larger carbon stocks than their surroundings, and this difference tends to increase
561 towards the least accessible areas. Similar results were found in ITs with granted property rights
562 in Brazil (27) and titled IT's in Colombia (26) which gradually decrease deforestation inside their
563 boundaries. These gradual reductions in deforestation and degradation imply that indigenous
564 and community land use, presumably for local livelihoods, reduce carbon stocks in the most
565 accessible forests while conserving core areas. Other studies have shown on a local scale the
566 limited impacts of indigenous land use, such as shifting agriculture and agroforestry, on carbon
567 stocks (44,45). Additionally, our results reveal that these spatial effect of protected lands
568 remains temporarily stable. Case studies from Mexico and Ecuador suggest that land-use is
569 temporarily stable in ITs (46,47). Our results, after controlling for spatial location, are among the
570 first to establish that indigenous and community land-use in neotropical forests may have a
571 limited and stable spatial impact on carbon stocks.

572

573 National contexts matter

574

575 Nonetheless, the current and future effects of allocating protected lands on carbon stocks is
576 influenced by national contexts. Overall geographical trends indicate that protected lands in

577 Panama and Brazil have wider temporal and spatial effects on carbon stocks than Colombia,
578 Ecuador and Peru. These geographical differences reflect past trends of extensive forest loss in
579 other lands from Panama (48) and the Brazil (49). Moreover, the increasing differences in
580 carbon stocks among protected lands and other lands that we found even after controlling for
581 spatial location, highlight a growing pressure on neotropical forests. Consequently, the
582 capacity of protected lands to preserve or reduce carbon stock losses is likely to change.
583 Between 2000 and 2013, tropical South America and Guatemala lost 7.3% and 13% of intact
584 forest lands, respectively, mostly caused by the expansion of agriculture (50). Community
585 Concessions in Petén (Guatemala) are challenged by changing concession statuses and land
586 invasions (7). PAs in Colombia are witnessing an increase in deforestation around their
587 boundaries after the Peace Agreement with the Revolutionary Armed Forces (FARC) (43). ITs
588 and PAs in southern Peru are threatened by growing road infrastructure, land invasions, illegal
589 gold mining, and coca production (51). Oil blocks in the Ecuadorian Amazon will expand in
590 cover from 32% to 68%, overlapping with biodiversity hotspots in PAs and ITs (52). In Brazil,
591 limited law enforcement to prevent forest loss from soy, meat, and timber production in the
592 Amazon Basin converge with recent setbacks in the land tenure security of ITs (53). Land
593 invasions and deforestation in Panama also pose a threat to ITs (54). In this sense, as
594 deforestation and degradation persist, countries' climate benefits from forests are increasingly
595 dependent on the stability of ITs and PAs carbon stocks. The increasing dependence on stable
596 forests points to the need to protect them through land use planning and resource allocation in
597 institutions at the international, national, and sub-national level (55,56).

598

599 Finally, our study has some limitations. Regarding covariates of spatial location, the data
600 included in our analysis do not capture the influence of rapidly changing roads. Nevertheless,
601 the covariates included in this study still create a general classification of accessibility and forest
602 loss pressures to control for the non-random location of protected lands. Despite using a

603 stratified sampling matching, known to effectively reduce covariate imbalances and the
604 variability of treatment effects (e.g., temporal and spatial effects) (57), further research would
605 benefit from comparing stratified and random sampling matching. We also aimed to identify the
606 overall influence of protected land categories, but they may represent different realities in each
607 country. For instance, OAs in Colombia are subject to a policy that requires National Park
608 Authorities to establish co-management agreements with Indigenous communities (58), which is
609 not necessarily the case in other countries. Even in our sub-national case study in Petén
610 (Guatemala), Community Concessions represent a diverse mosaic of community-based
611 organizations with particular land-use dynamics (16,59). The outcome variable also brings
612 limitations because it does not differentiate carbon stock losses due to deforestation and
613 degradation, rather it provides a comprehensive measure (i.e., aboveground carbon stocks) that
614 captures the effectiveness of protected lands beyond deforestation.

615

616 Conclusions

617 After controlling the influence of spatial location, we found that protected lands with fewer
618 restrictions on forest use represent effective natural climate solutions. Particularly, indigenous
619 lands (ITs, OAs) and PAs have similar temporal and spatial effects on carbon stocks in Panama
620 and Amazon Basin countries. A similar effect also emerges when comparing Community
621 Concessions in Petén (Guatemala) and Extractive Reserves in Acre (Brazil) with Integral PAs.
622 Considering that the observation units sampled by matching are located along the boundaries of
623 protected lands, these temporal and spatial effects are conservative. Consequently, our findings
624 show that indigenous peoples and local communities are supporting Nationally Determined
625 Contributions (NDCs) under the Paris Agreement. Brazil and Ecuador expect to receive their
626 first results-based payments from the Green Climate Fund corresponding to 96.5 and 18.6
627 million USD, respectively (60). For the critical role they play in reducing net carbon emissions,

628 indigenous peoples and local communities must become recipients of such benefits,
629 independent of the opportunity costs of avoided deforestation and degradation (61).
630

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830 **Supplementary Material**

831

Jurisdiction	Geo-spatial Information	Source
All regions	Annual carbon density (2003 – 2016)	Woods Hole Research Center (Baccini et al., 2012; Baccini et al., 2017).
All regions	Elevation and sope	Shuttle Radar Topographic Mission – USGS (United States Geological Survey)
Petén (Guatemala)	Roads, Towns (> 5000 inhabitants), Rivers.	IDEG ("Infraestructura de datos espaciales de Guatemala").
Petén (Guatemala)	PAs, Forest Concessions, and Multiple Use Zones.	ACOFOP ("Asociación de Comunidades Forestales de Petén"), Alianza Mesoamericana de los Pueblos, Prisma.
Panama	Roads, Towns (> 5000 inhabitants), Rivers, PAs.	STRI ("Smithsonian Tropical Research Institute").
Panama	Titled and Claimed ITs	Neotropical Ecology Laboratory (Prof. Catherin Potvin Universidad de McGill/STRI), COONAPIP (Coordinadora Nacional de los Pueblos Indígenas de Panamá).
Amazon Basin (Colombia,	PAs and ITs, Roads	RAISG (Red Amazónica de Información Socio-Ambiental Geo-Referenciada).

Ecuador, Perú,

Brasil).

Colombia	Rivers	IGAC ("Instituto Geográfico Agustín Codazzi").
Colombia	Settlements (> 5000 inhabitants)	DANE (Departamento Administrativo Nacional de Estadística).
Ecuador	Settlements (> 10 000 inhabitants), Rivers.	IGM ("Instituto Geográfico Militar").
Peru	Rivers	IGN ("Instituto Geográfico Nacional").
Peru	Settlements (> 5000 inhabitants)	INEI ("Instituto Nacional de Estadística e Informática") and Ministry of Education.
Brazil	Settlements (> 5000 inhabitants), Rivers.	IBGE ("Instituto Brasileiro de Geografia e estadística").

832

833 **S1 Appendix. Geospatial Information and its sources by jurisdiction.**

834

835

Jurisdiction	Protected Lands	Roads (km)	Settlements (km)	Rivers (km)	Elevation (m)	Slope (%)
Panama	PAs	-25.91***	-6.04***	-0.23***	213.00***	-2.00***
	ITs	-52.36***	-8.20***	-0.33***	-154.00***	-2.00***
	OAs	-109.12***	-13.86***	-0.01***	-3.00***	-3.00***
Colombia	PAs	-29.10***	-37.56***	-0.25***	12.00***	0.00***
	ITs	-13.92***	-62.62***	0.16*	177.00*	1.00*
	OAs	-28.87***	-70.74***	1.07***	186.00***	1.00***
Ecuador	PAs	-12.50***	-8.66***	-0.53***	-309.00***	-2.00***
	ITs	-28.49***	-16.21***	0.04***	415.00***	3.00***
	OAs	-23.36***	-25.11***	-0.13***	447.00***	3.00***

Peru	PAs	-28.05***	-46.95***	-0.04***	-22.00***	0.00***
	ITs	2.06***	5.07***	2.08***	29.00***	0.00***
	OAs	-29.54***	3.26***	-2.85***	235.00***	2.00***
Brazil	PAs	-27.22***	-33.14***	-3.54***	9.00***	0.00***
	ITs	-32.17***	-91.69***	-7.91***	-53.00***	0.00***
	OAs	-15.65***	-126.83***	-15.65***	-49.00***	-1.00***
Petén (Guatemala)	Integral PAs	-2.65***	-9.14***	-3.12***	70.00***	0.00***
	Community Managed PAs	2.01***	-6.18***	-2.72***	23.00***	0.00***
	Sustainable Use PAs	-1.38***	-5.33***	-1.26***	65.00***	0.00***
Acre (Brazil)	Integral PAs	19.24***	-284***	-29.30***	-47.80***	-0.01***
	Community Managed PAs	24.56***	5.11***	18.08***	29.99***	0.02***
	Sustainable Use PAs	43.79***	-25.50***	17.15***	24.45***	0.06***

836 **S2 Appendix. Covariates mean differences between other lands and protected lands by**

837 **jurisdiction and their statistical significance from Mann Whitney U tests. *** p < 0.001, ** p**

838 **< 0.01, * p < 0.05.**

839

Jurisdiction	Protected Land	Covariates				
		Roads (km)	Settlements (km)	Rivers (km)	Elevation (m)	Slope (%)
Panama	PAs	0.5	0.5	0.5	50	1
	ITs	2.5	2	1	100	1.5
	OAs	1	2	1	150	1.5
Colombia	PAs	1	1	0.5	50	1.5
	ITs	2	1	1	50	1.5
	Overlapped Areas	2	2	1	100	1.5
Ecuador	PAs	1	1	1	50	2
	ITs	1	2	0.8	150	2
	OAs	0.5	2	0.7	150	2
Peru	PAs	2	1	2.5	200	2

	ITs	1	1	1	150	2
	OAs	1	1	1	150	2
Brazil	PAs	1	1	1	100	1
	ITs	1	1	1	200	1.5
	OAs	1	1	1	200	1.5
Petén (Guatemala)	Integral PAs	1	1	1	100	1
	Community Managed PAs	1	1	1	100	1
	Sustainable Use PAs	1	1	1	100	1
Acre (Brazil)	Integral PAs	1	1	1	100	1
	Community Managed PAs	1	1	1	200	1.5
	Sustainable Use PAs	1	1	1	200	1.5

840

841 **S3 Appendix. Coarsening Choices applied through Coarsened Exact Matching (CEM) by**
 842 **jurisdiction across protected lands (PAs, ITs, and OAs).**

843

844 **S4 Appendix. Standard mean differences of covariates among other lands and protected**
 845 **lands before (Pre-Match) (full circles) and after matching analysis (Matched) (empty**
 846 **circles) across national-level and subnational-level jurisdictions.** At the National-level
 847 (Panama, Colombia, Ecuador, Peru, and Brazil) protected lands are classified as ITs (orange),
 848 OAs (yellow), and PAs (green). At the subnational-level (Petén and Acre), protected lands
 849 correspond to Community-Managed PAs (light green), Sustainable Use PAs (cyan), and Integral
 850 PAs (dark green). The standard mean difference expresses the size of the covariate imbalance
 851 between other lands and a particular protected land relative to their pooled standard deviation.
 852 Negative values imply higher covariate values in protected lands than other lands.

853

854 **S5 Appendix. The temporal effect of ITs (orange), OAs (yellow), and PAs (green) on**
 855 **aboveground carbon stocks across neotropical countries.** Each point represents the
 856 significant annual effects ($p < 0.05$) of protected lands (ITs, OAs, and PAs). The temporal
 857 effects are the annual differences of carbon stocks in protected lands and other lands after

858 controlling for the spatial location through matching analysis and linear mixed models. Error
859 bars reflect 95% confidence intervals for the temporal effect derived from the linear mixed
860 models.

861

862 **S6 Appendix. The carbon stocks baseline of ITs (orange), OAs (yellow), and PAs (green)**

863 **across neotropical countries.** Each point represents the mean annual carbon stocks found in
864 other lands (i.e., carbon stocks baseline) that share a spatial location similar to protected lands
865 (ITs, OAs, PAs) after matching analysis and linear mixed models. Error bars reflect 95%
866 confidence intervals for the carbon stocks baselines derived from the linear mixed models.

867

868 **S7 Appendix. The temporal effect of Protected Areas (PAs) in Petén (Guatemala) and**

869 **Acre (Brazil) on carbon stocks.** The Community Managed PAs (light green) correspond to
870 Community Concessions in the Maya Biosphere Reserve in Petén (Guatemala) and Extractive
871 Reserves in Acre, Brazil (IUCN VI). Sustainable Use PAs (cyan) correspond to Private
872 Concessions and Cooperatives in the Maya Biosphere Reserve (Petén, Guatemala) and
873 National and State Forests in Acre (Brazil). Integral PAs (dark green) comprise IUCN categories
874 I-IV in each jurisdiction. Each point represents the significant temporal effects of PAs. These
875 temporal effects are the annual differences of carbon stocks in PAs and other lands after
876 controlling for the spatial location through matching analysis and linear mixed models. Error
877 bars reflect 95% confidence intervals for the temporal effects derived from the linear mixed
878 models.

879

880 **S8 Appendix. The carbon stocks baseline of PAs categories in Petén (Guatemala) and**

881 **Acre (Brazil) on carbon stocks.** The Community Managed PAs (light green) correspond to
882 Community Concessions in the Maya Biosphere Reserve in Petén (Guatemala) and Extractive
883 Reserves in Acre, Brazil (IUCN VI). Sustainable Use PAs (cyan) correspond to Private

884 Concessions and Cooperatives in the Maya Biosphere Reserve (Petén, Guatemala) and
885 National and State Forests in Acre (Brazil). Integral PAs (dark green) comprise IUCN categories
886 I-IV in each jurisdiction. Each point represents the mean annual carbon stocks found in other
887 lands (i.e., carbon stocks baseline) that share a spatial location similar to PAs categories after
888 matching analysis and linear mixed models. Error bars reflect 95% confidence intervals for the
889 carbon stocks baselines derived from the linear mixed models.

890

891 **S9 Appendix. The carbon stocks baseline outside the boundaries of across neotropical**
892 **countries.** Full (2016) or empty (2003) points represent the mean annual carbon stocks found
893 in other lands (i.e., carbon stocks baseline) outside the boundaries of ITs (orange), OAs
894 (yellow), and PAs (green) at a certain distance,. Error bars reflect 95% confidence intervals for
895 the carbon stocks baselines derived from the linear mixed models.

896

897 **S10 Appendix. The carbon stocks baseline outside PAs' boundaries in Petén (Guatemala)**
898 **and Acre (Brazil).** The Community Managed PAs (light green) correspond to Community
899 Concessions in the Maya Biosphere Reserve in Petén (Guatemala), and Extractive Reserves in
900 Acre, Brazil (IUCN VI). Sustainable Use PAs (cyan) correspond to Private Concessions and
901 Cooperatives in the Maya Biosphere Reserve (Petén, Guatemala) and National and State
902 Forests in Acre (Brazil). Integral PAs (dark green) comprise IUCN categories I-IV in each
903 jurisdiction. Each dot represents the mean annual carbon stocks found in other lands (i.e.,
904 carbon stocks baseline) outside PAs' boundaries at a certain distance, according to geographic
905 discontinuity designs. Error bars reflect 95% confidence intervals for the carbon stocks
906 baselines derived from the linear mixed models.

907

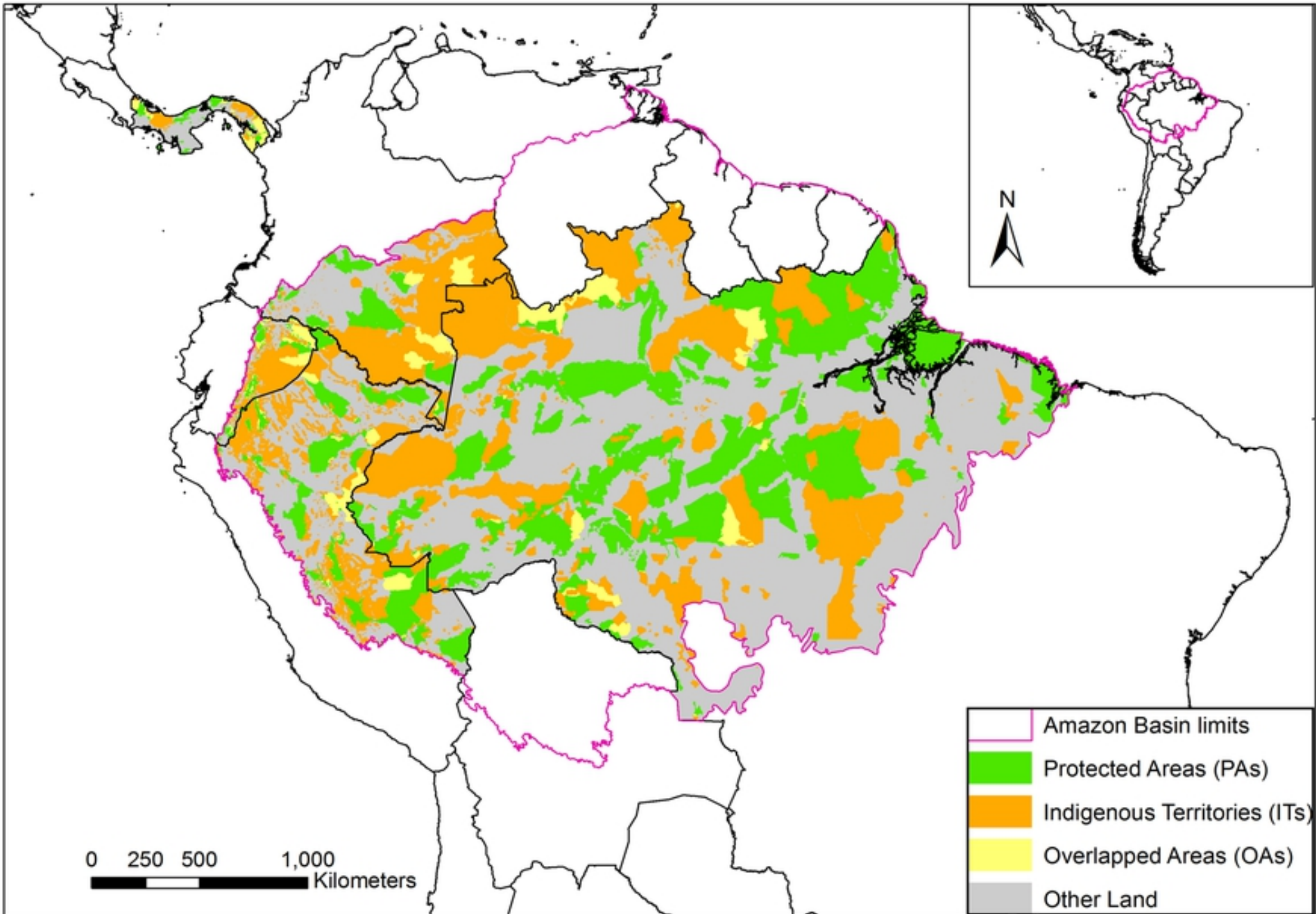


Fig 1

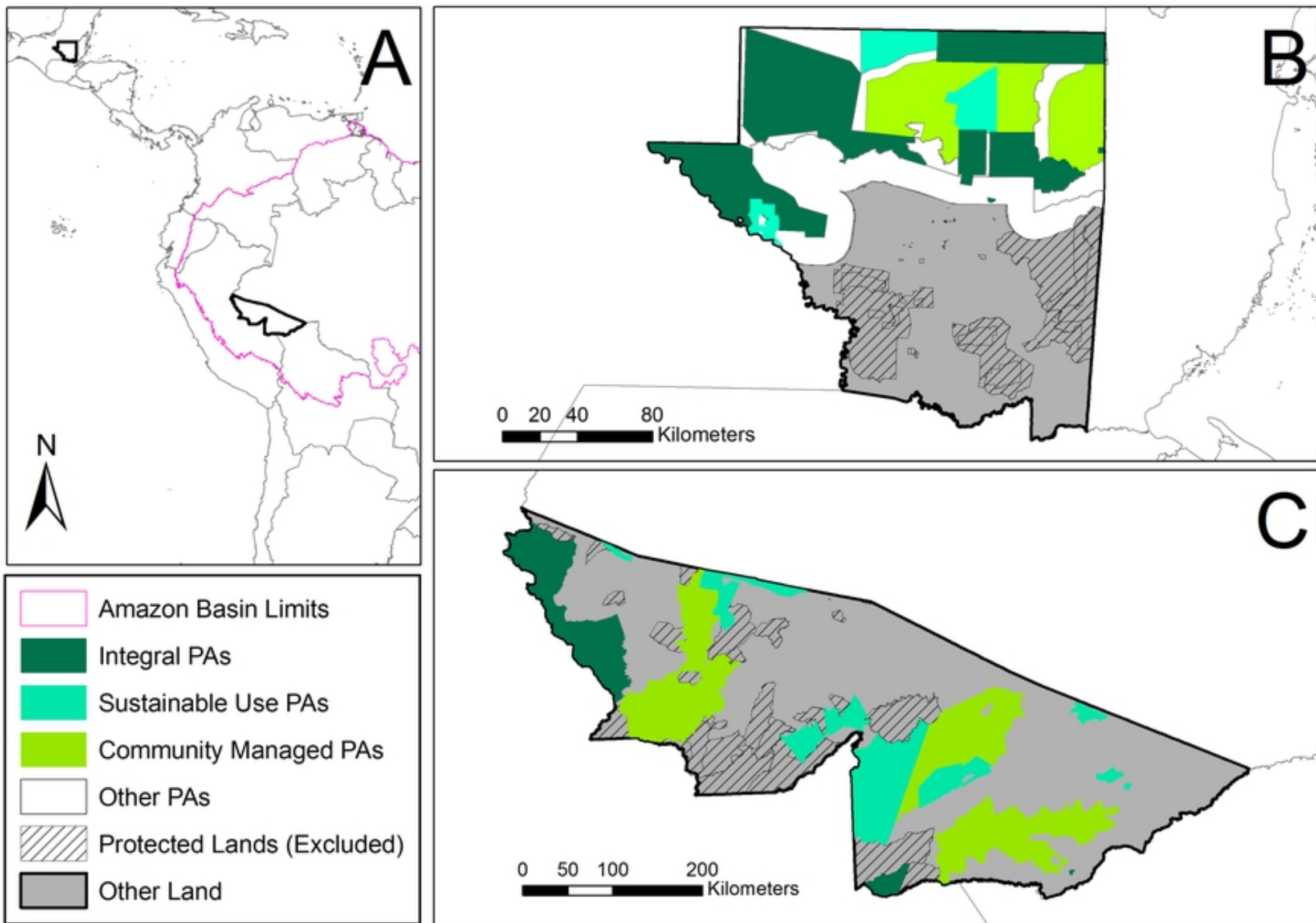


Fig 2

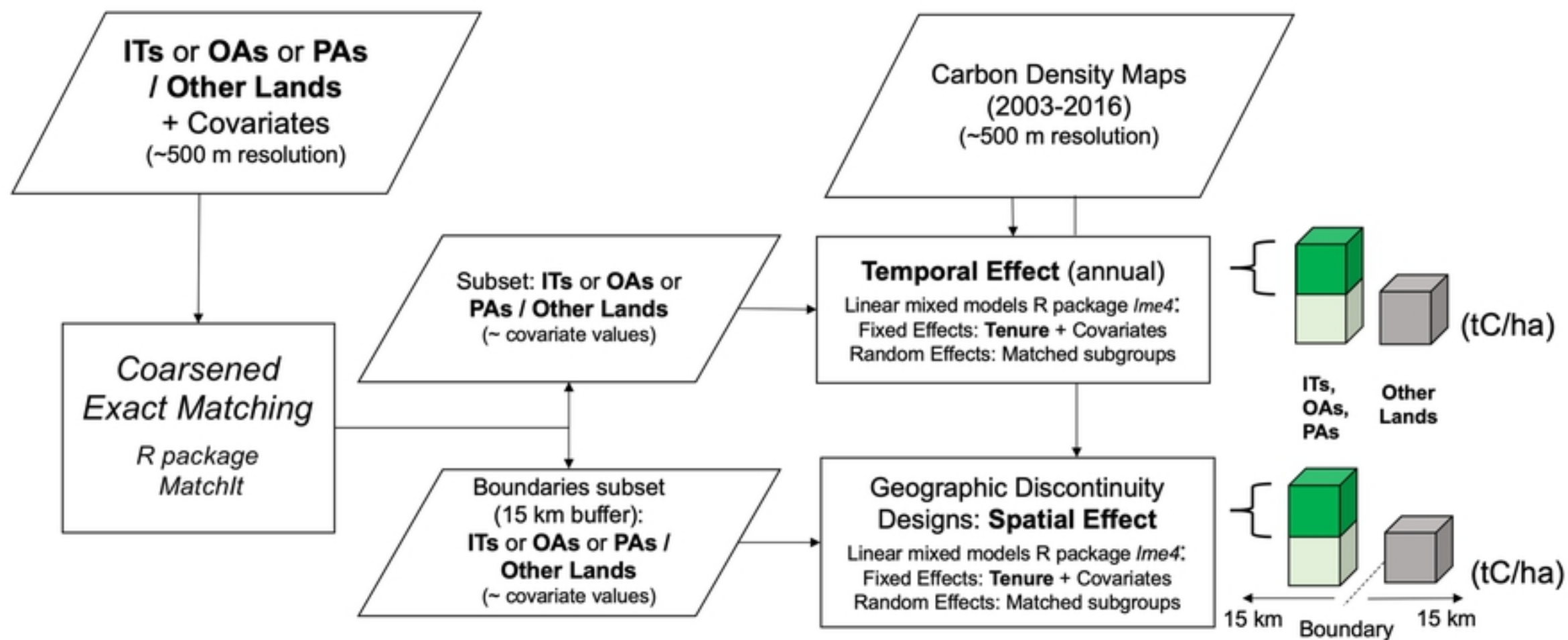


Fig 3

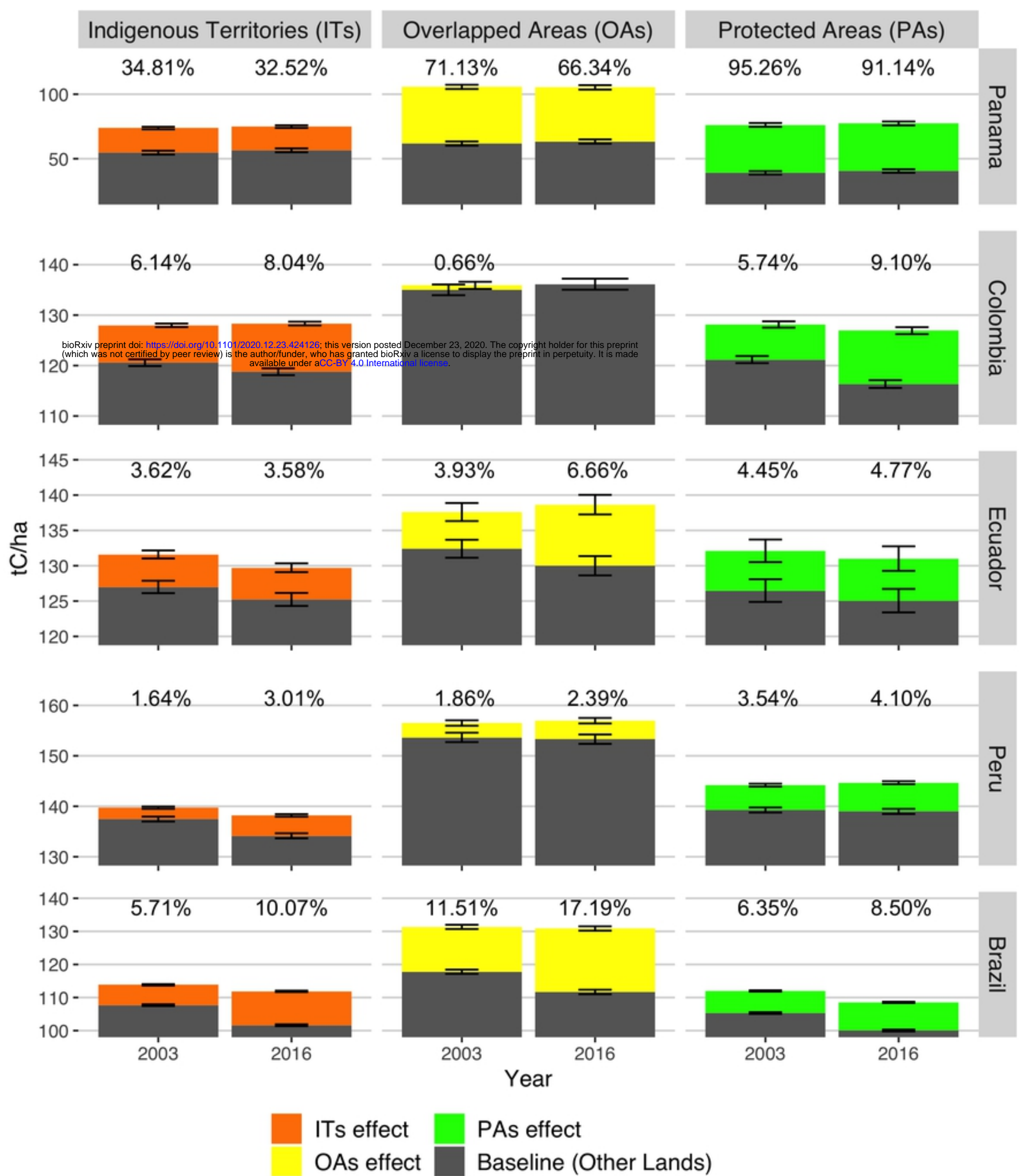


Fig 4

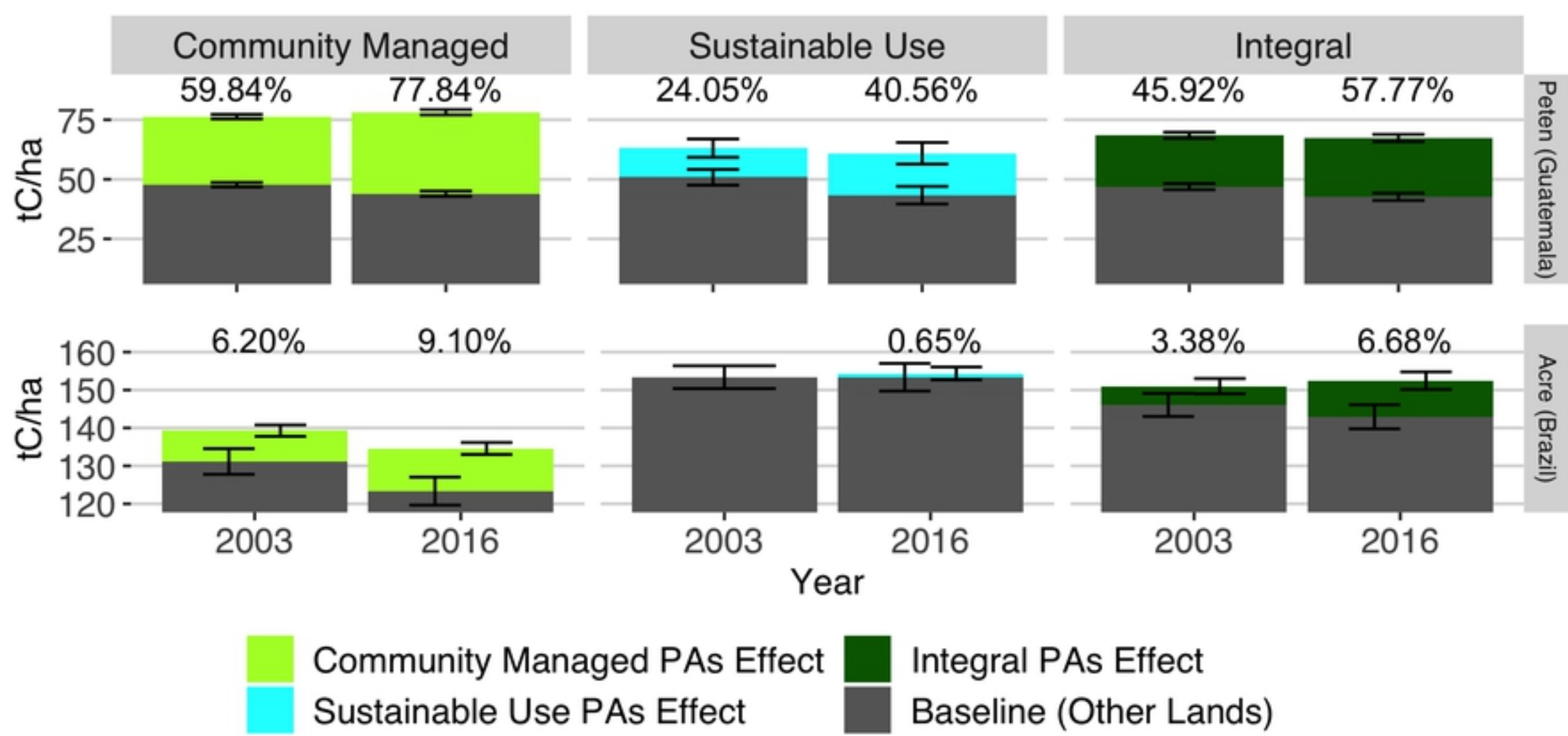


Fig 5

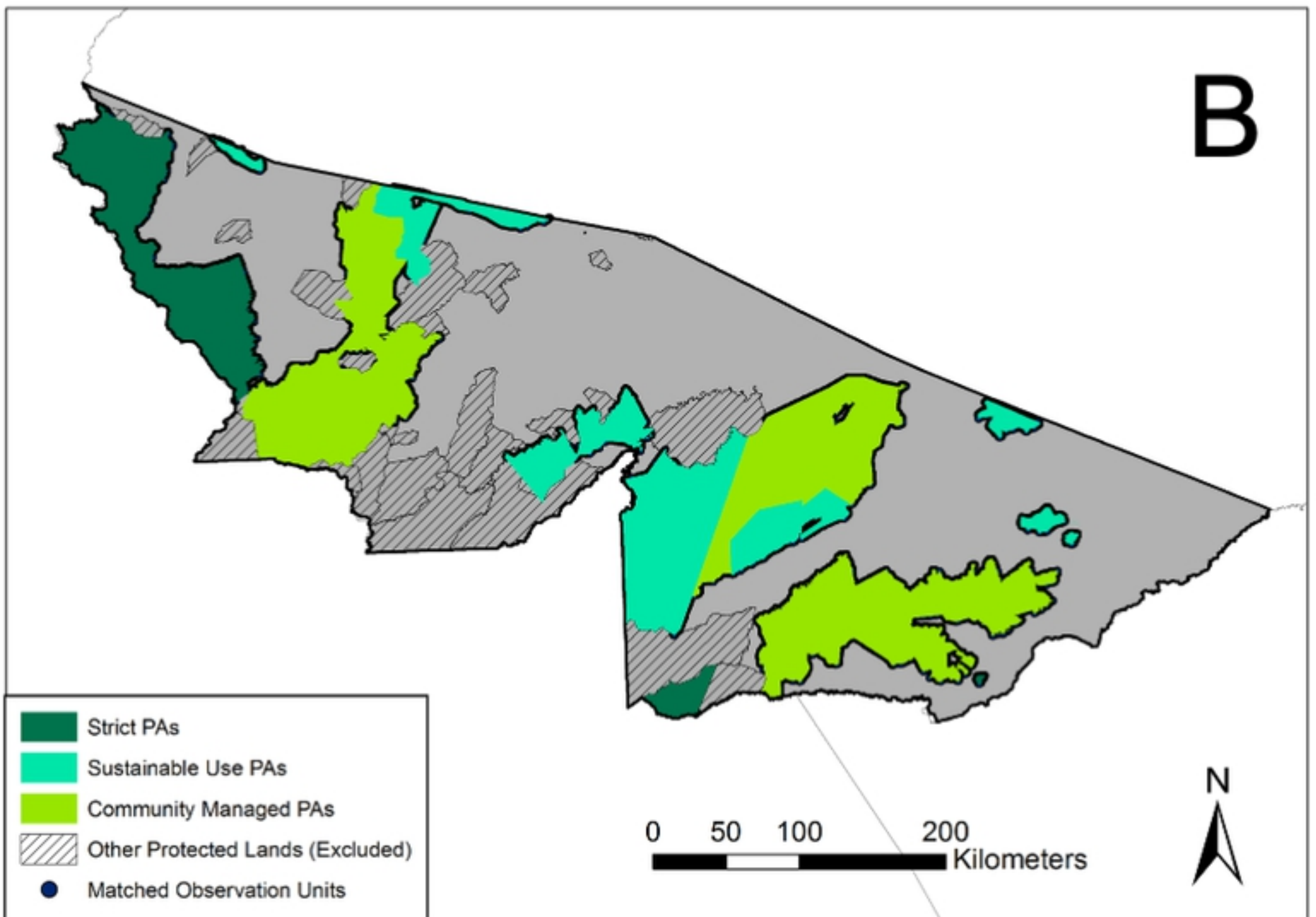
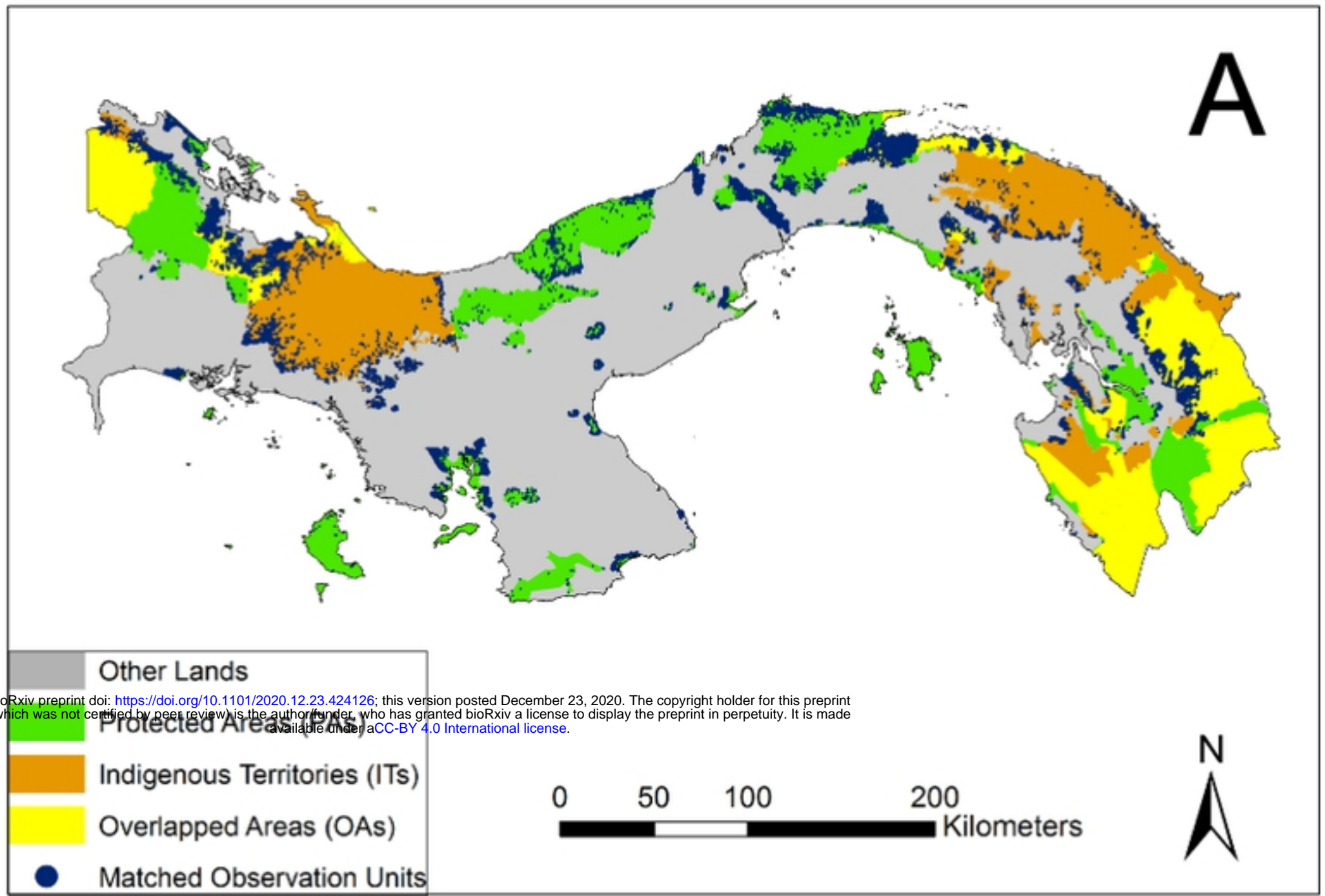


Fig 6

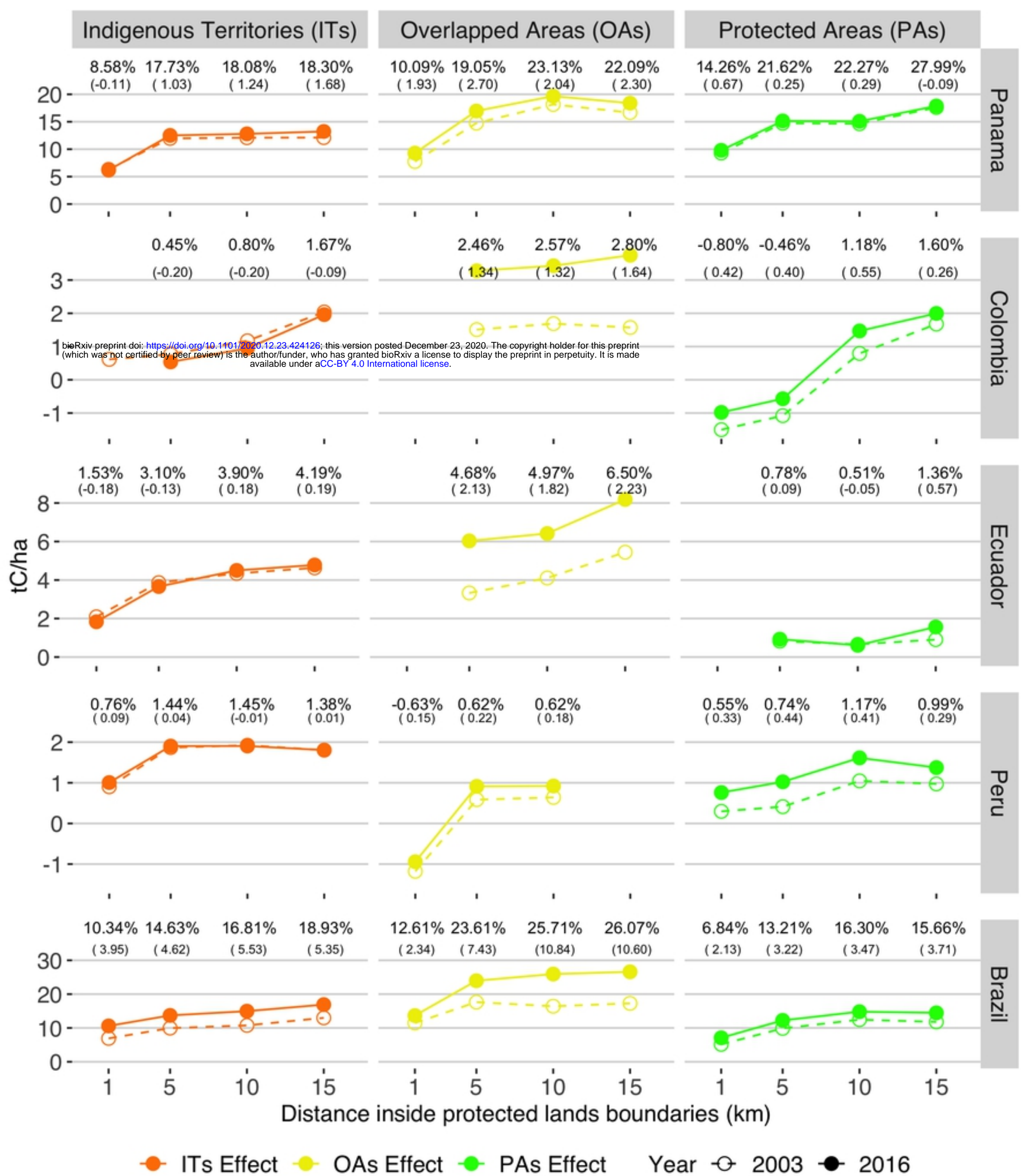


Fig 7

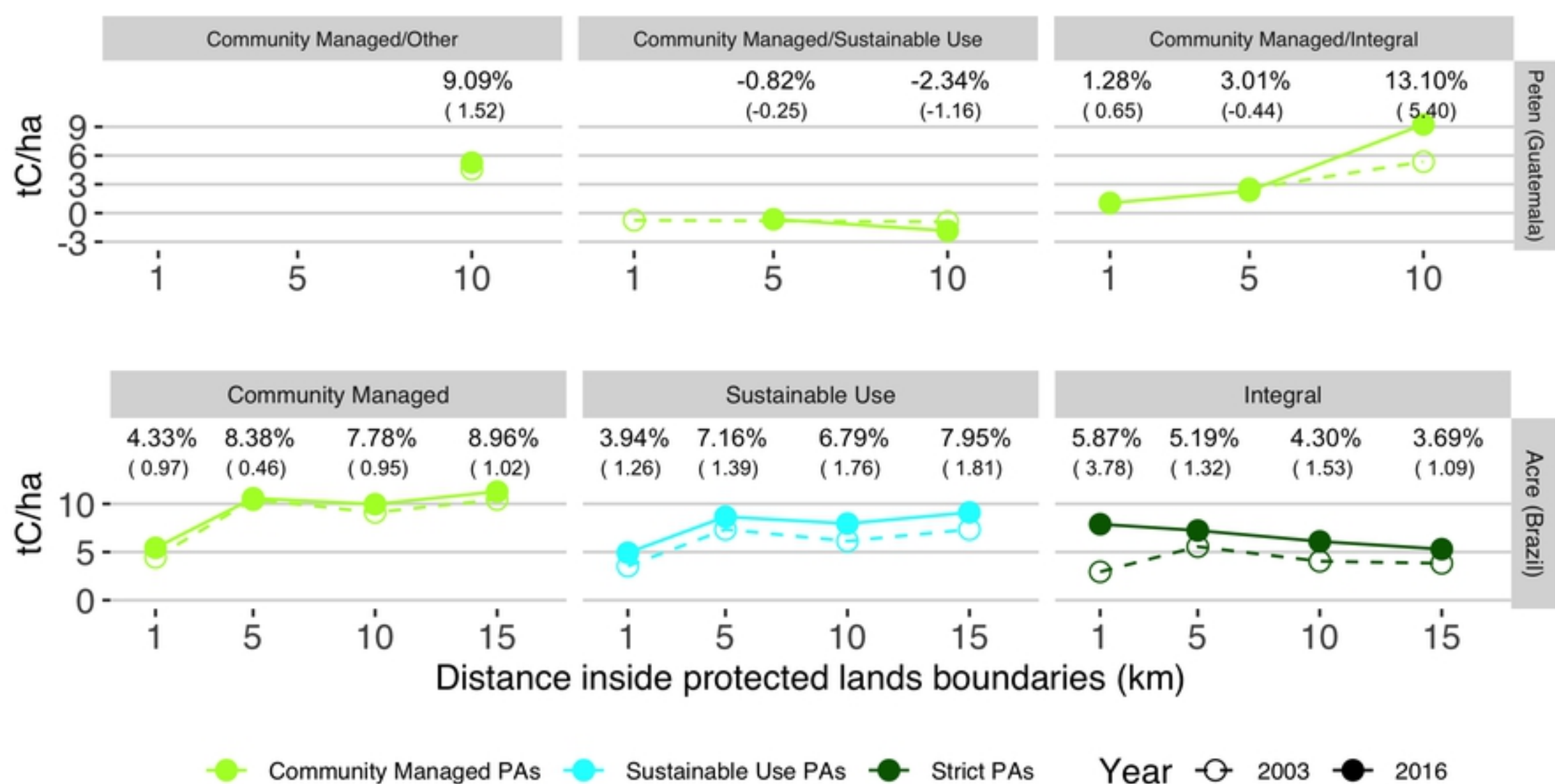


Fig 8