Invasive grass dominance over native forbs is linked to shifts in the bacterial rhizosphere

microbiome

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Abstract

Rhizosphere microbiomes have received growing attention in recent years for their role in plant

health, stress tolerance, soil nutrition, and invasion. Still, relatively little is known about how

these microbial communities are altered under plant competition, and even less about whether

these shifts are tied to competitive outcomes between native and invasive plants. We investigated

the structure and diversity of rhizosphere bacterial and fungal microbiomes of native annual

forbs and invasive annual grasses individually and in competition using high-throughput

amplicon sequencing of the bacterial 16S rRNA gene and the fungal ITS region. We assessed

how differentially abundant microbial families correlate to plant biomass under competition We

find that bacterial diversity and structure differ between native forbs and invasive grasses, but

fungal diversity and structure do not. Further, bacterial community structures under competition

are distinct from individual bacterial community structures. We also identified five bacterial

families that varied in normalized abundance between treatments and that were correlated with

plant biomass under competition. We speculate that invasive grass dominance over these natives

may be partially due to effects on the rhizosphere community, with changes in specific bacterial

families potentially benefiting invaders at the expense of natives.

Keywords: rhizosphere microbiome, competition, invasive species, plant-microbe interactions,

annual grassland

Declarations

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Competing interests

The authors declare that the research was conducted in the absence of any commercial or

financial relationships that could be construed as a potential conflict of interest.

Availability of data and material

The datasets and scripts supporting the conclusions of this article are available in the GitHub

repository, DOI:10.5281/zenodo.4422181. The sequence reads generated for this 16S rRNA gene

and ITS region amplicon project were deposited at Genbank under accession no. PRJNA666893.

Code availability

No custom software was used. All R scripts used to process the data are available in the GitHub

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Authors' contributions

MLL, CLE, and HK conceived of, planned, and carried out the experiment. MLL and CLE

processed and analyzed the data and wrote the paper. All authors contributed to the final version

of the manuscript.

Ethics approval

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Consent to participate

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Not applicable.

Introduction

Plants and their associated soil microbial communities have coevolved complex dynamic

relationships through time. The soil directly surrounding plant roots, known as the rhizosphere, is

home to a diversity of microbes that play vital roles in plant health [1], nutrition [2], and stress

tolerance [3,4] and is distinct from the surrounding bulk soil [1,5–7]. Though rhizosphere

composition is driven predominantly by abiotic factors [8–10], these communities can be highly

host plant-specific [11,12], with plants exhibiting strong local adaptations to their home

microbial community [13–15]. Further, there is growing evidence that microbial communities

vary in the presence of certain neighbors [16], and can even affect competitive outcomes

between plants [17–19] with the potential to drive either coexistence or exclusion [20,21].

Microbes can affect plant-host competition both directly via resource availability and

indirectly through community interactions. For example, microbes can make nitrogen more

readily available to plants [22,23], which could preferentially benefit hosts with faster uptake

rates [24,25] thereby increasing their competitive success [26]. Plants may also indirectly

compete for beneficial plant-growth-promoting microbes [7,27,28] or come into contact with

novel pathogens through competition with a novel plant, such as an invader [29]. These

interactions can lead to the loss of specialized microbes in the inferior competitor [30] and drive

the resulting microbial community to resemble that of the dominant competitor [18,19].

Shifts in the microbiome are a potentially important contributor to the dominance of

invasive plants. Although there is strong evidence for local adaptation between plants and their

home microbiome, invasive plants are relatively novel players, causing disruptions in important

plant-microbe interactions by increasing soil microbial activity [31], reducing microbial biomass

[32] and diversity [33], increasing nitrification rates [24], and changing microbiome composition

[34,35]. These plant-soil feedbacks can benefit the invader [35,36], selectively harm natives [37],

and/or benefit other invaders [38]. While it is known that invaders and natives can harbor distinct

microbial communities, and that competition between host plants can affect microbiome

composition (and vice versa), little is known of how changes in the abundance of specific

microbial taxa may be tied to the competitive dominance of invaders and/or the competitive

inferiority of natives.

Serpentine soils in California annual grasslands are characterized by high heavy metal

content and low levels of essential plant nutrients, and as a result, are home to a unique

community of native forbs. While some of these soils are rocky and shallow, other serpentine

areas are deeper and more finely textured ("lush" serpentine) and therefore hold more water and

nitrogen [39], making them more favorable to invasive annual grass establishment and growth.

These fast-growing invasive grasses outcompete the less abundant and competitively inferior

native forbs, contributing to their declines across California grasslands [40–42]. In native-

dominated serpentine areas, microbes have been shown to increase seedling survival [43] and

facilitate heavy metal tolerance [44]. With the invasion of grasses, microbiome changes may

strengthen the advantage grasses have over natives if they recruit beneficial microbes away from

natives or harbor novel microbes harmful to natives. Alternatively, locally adapted microbes may

be harming invasive grasses, helping native forbs to persist in these areas.

We sought to understand how competition between native forbs and invasive grasses

affects the bacterial and fungal rhizosphere microbiomes, and how microbes may be shaping

competitive outcomes in this community. Our main questions were:

(Q1) Do microbiomes differ between invasive grasses and native forbs?

(Q2) Do microbiomes of grass-forb pairs (i.e. under competition) differ from the

microbiomes of (a) invasive grasses and (b) native forbs?

(Q3) If so, are microbiomes of grass-forb pairs sourced more from invasive grasses,

native forbs, or equally from both?

(Q4) Given compositional changes in the microbiome in grass-forb pairs, which specific

microbial families are driving these changes?

(Q5) Are abundances of these microbial families correlated with plant performance in

grass-forb pairs?

To answer these questions, we conducted a manipulative shade-house experiment in

which forbs and grasses were grown in field-collected soil both individually and in grass-forb

pairs. For each question, we looked at effects in both the bacterial and the fungal rhizosphere

measured through high throughput amplicon sequencing, focusing our analysis on group-level

results (i.e. forb, grass, and grass-forb pairs).

Methods

Experimental set-up

In spring and summer 2017, we collected soil and seed from a lush serpentine annual

grassland at the University of California McLaughlin Natural Reserve [45] (N 38°52', W

122°26'). Seeds were collected from >10 individuals each of six common native annual forb

species (Fig. 1a; *Lasthenia californica, Clarkia purpurea, Agoseris heterophylla, Calycadenia*

pauciflora, Hemizonia congesta, and *Plantago erecta)* and three dominant invasive annual

grasses (Fig. 1b; *Avena fatua, Taeniatherum caput-medusae,* and *Bromus hordeaceus*). In

December 2017, seeds were sown into physan-washed pots with a mixture of 65% field-collected

soil and 35% sterile autoclaved sand to improve drainage. Each species was grown alone (one

individual per pot, five replicate pots per species, 45 total pots: 30 forbs, 15 grasses) to assess the

microbial communities of individual plant species. To assess the effects of competition between

natives and invasives, each forb species was grown with each grass species in a pairwise factorial

design for a total of 18 different species pairs (one individual of each grass and forb per pot, five

replicate pots per combination, 90 total pots). We chose to focus on grass-forb pairs for our

competition treatment as this treatment is the most relevant to understanding how microbial

changes in the rhizosphere may contribute to invasive grass dominance. Due to plant mortality

and a sample mix-up, final numbers were as follows: grass-forb pairs = 84, forbs = 27, grasses =

15. Pots were placed in a shade-house open to natural temperature variation in the UC Davis

Orchard Park Greenhouse and soil moisture was maintained with an automatic drip irrigation

system (Fig. 1c).

Sample collection

In April 2018, we sampled rhizosphere microbial communities by gently shaking off soil

from plant roots, then submerging the roots in autoclaved nanopure water in a 50 mL conical

tube and vortexing to obtain the rhizosphere soil similar to Edwards et al. [1]. Conical tubes were

then centrifuged at 4000 g for 1 minute to obtain rhizosphere soil pellets. In grass-forb pairs, we

sampled the joint rhizosphere microbiome, similar to Sun et al. [46], as roots were completely

interwoven.

To assess how the microbial community may aid in invasive grass dominance over native

forbs, we harvested, dried, and weighed aboveground plant biomass. Roots were not included

due to the difficulty of identifying roots to species in the competition treatment. Previous work in

this system has shown that competition with invasive grasses lowers fitness of native forbs

through decreased seed production[41,47]. While we were not able to measure seed production

of plants in this study due to rhizosphere sampling prior to seed set, aboveground biomass in

annual plants is often used as a surrogate for competitive ability [48,49] and is well correlated

with seed production [50,51] and is therefore a reasonable estimate of plant performance.

Molecular methods and sequence generation

Briefly, we extracted DNA from the 135 rhizosphere soil pellets using the MoBio

PowerSoil DNA Extraction kit (MO BIO Laboratories, Inc., Carlsbad, CA, USA). Samples were

then sent to the Integrated Microbiome Resource (IMR) at Dalhousie University to amplify and

sequence the 16S rRNA gene using the 515FB-806RB primer set [52,53] and ITS2 region using

the ITS86F-ITS4R primer set [54]. The sequence reads generated for this 16S rRNA gene and

ITS region amplicon project were deposited at Genbank under accession no. PRJNA666893.

Sequence processing

Sequence data was processed using the DADA2 workflow [55] in R [56] to create

Amplicon Sequence Variant tables (ASV). Chimeras, contaminants, mitochondria and

chloroplasts were removed from these tables prior to further analysis. For questions 1-3, the 16S

rRNA gene dataset was rarefied to 9434 reads per sample and the ITS region dataset was rarefied

to 7557 reads per sample. These rarefaction levels were chosen, after examining rarefaction

curves for saturation, based on the size of the sample with the smallest number of reads in order

to retain the maximum number of samples for downstream analysis.

Further methodological details related to experimental set-up, sample collection and

processing, sequence generation, and data processing prior to downstream analyses can be found

in the Supplemental Materials.

Data analysis (Q1-Q2)

To evaluate how rhizosphere communities varied in structure (beta diversity) between

grasses, forbs, and grass-forb pairs, we conducted a Principal Coordinates Analysis (PCoA) on

weighted UniFrac of each sample for rarefied 16S rRNA gene data and on Bray-Curtis

dissimilarities for rarefied ITS region data. We tested for significant differences in mean

centroids between groups (native forbs, invasive grasses, grass-forb pairs) using a permutational

multivariate analysis of variance (PERMANOVA) and between mean dispersions using the

betadisper function in vegan [57]. We then conducted post-hoc analyses with Benjamini-

Hochberg p-value corrections [58]. We also investigated whether these groups varied compared

to background soil (i.e. the community in the soil-sand mix at the start of the experiment) by

performing PERMANOVA tests. Finally, we investigated the degree of species-level variation

within both grasses and forbs by running PERMANOVA tests on each group separately to

evaluate whether there was more variation within or between groups.

To assess how rhizosphere communities varied in alpha diversity between grasses, forbs

and grass-forb pairs, we calculated Shannon diversity of each sample and conducted a Kruskal-

Wallis test to test whether alpha diversity varied between these groups. We then conducted post-

hoc analyses using a Dunn Test with Benjamini-Hochberg p-value corrections.

Data analysis (Q3)

To evaluate whether a larger proportion of the joint microbiome community was

predicted to be sourced from forbs or grasses, we used SourceTracker, a Bayesian source

tracking classifier [59]. SourceTracker uses Gibbs sampling and Dirichlet distributions to

estimate the proportional abundance of an ASV originating from provided source populations.

First, we trained SourceTracker on the rarified ASV tables of forbs, grasses, and background soil

and then tested the trained Bayesian model on the grass-forb pair joint microbiomes to estimate

the proportion of ASVs originating from each of these sources. The model assumes that joint

microbiomes contain a combination of colonists from known and unknown sources (i.e., any

ASVs that are absent from the provided source datasets, and thus any sources not sampled here

such as air, water, etc.) and estimates the fraction of ASVs detected in the joint microbiome that

originated from grasses, forbs, background soil or unknown habitats. We used Wilcoxon signed-

rank tests to assess differences in the SourceTracker predicted proportions between different

sources for the bacterial and fungal communities, and then also tested for differences in the

proportions of each individual source between the two communities.

Data Analysis (Q4)

We used differential abundance analysis to identify microbial families whose abundance

varied between groups in order to investigate which families were driving observed differences

in community structure and composition. To do this, we first summed the raw read counts of

each full data set (i.e. not rarefied), then used the DESeq2 R package [60] on counts to examine

the $\log_2$fold change (i.e. differential abundance) of families between pairwise contrasts of groups

(forbs, grasses, and grass-forb pairs). We chose to investigate differences at the family-level to

maximize the potential for understanding the functional role of microbes while minimizing zero-

inflation observed at higher resolutions. Only 1% of bacterial ASVs were unable to be

taxonomically classified at the family level, while 49% of fungal ASVs could not be assigned to

a family. Families with Bonferroni corrected p-values < 0.01 were classified as differentially

abundant. In subsequent analyses, we focus on families that varied between both grass and forb

individual microbiomes and between individual microbiomes and joint grass-forb pair

microbiomes to highlight how baseline differences between groups are related to differences

observed during competition.

Data analysis (Q5)

To understand whether the abundances of these microbial families may be tied to the

dominance of grasses over forbs, we first assessed how biomass varied between forbs and

grasses both alone and in pairs, then we investigated the relationship between plant biomass in

pairs and microbial family abundance. To compare biomass of forbs and grasses grown alone,

we conducted a linear mixed effects model with biomass of plants grown alone as the response

variable and functional group (grass or forb) as the predictor variable. Biomass was log-

transformed to fit model assumptions and we included a plant species random intercept to

account for species-level differences. To compare biomass of grasses and forbs in grass-forb

pairs we ran a similar model on log biomass of plants in paired pots, again with a random

intercept for each species. We explored other random effect structures that included a term

accounting for non-independence of pots, however ultimately did not include this term as it did

not include this term as it did

not improve the model based on the log-likelihood.

If grass dominance over forbs was controlled wholly by density dependent effects and

competitive hierarchies rather than microbial interactions, we would expect an inverse

relationship between grass biomass and forb biomass. To investigate the degree to which forb

biomass was linked to grass biomass in pairs we used similar models with log-transformed forb

biomass as the response variable and log-transformed grass biomass as the predictor variable and

a random intercept for forb species.

To investigate the relationship between performance in grass-forb pairs and microbial

abundance, we ran linear mixed effects models for each host group (grass and forb) and each

differentially abundance microbial family with plant biomass from grass-forb pairs as the

response variable and regularized log abundance of the focal family (calculated using the rld

function in the DESeq2 package) as the predictor variable. Regularized log abundance (hereafter

referred to as normalized abundance) is on the log2 scale and is normalized to library size to

allow for direct comparison of microbial abundance across samples. We also included plant

species as a random intercept.

Finally, we calculated the proportion of each differentially abundant family in the joint

microbiome predicted to originate from either the individual forb or grass microbiomes by

grouping ASVs into taxonomic families and then summarizing the fraction of ASVs within each

family attributed to each possible source using SourceTracker.

Results

Bacterial microbiomes differ between invasive grasses and native forbs

Overall, the rhizosphere bacterial microbiome varied between invasive grasses and native

forbs, but the fungal microbiome did not. The structure of forb bacterial microbiomes was

significantly different from the structure of grass bacterial microbiomes ($p < 0.001$; Fig. 2).

Further, bacterial Shannon diversity (Fig. 3) was significantly higher in forb microbiomes than in

grass microbiomes (p = 0.035). Conversely, fungal communities showed marginally

nonsignificant differences in structure (p = 0.07; Fig. S1) and no significant differences in

Shannon diversity of fungal communities among treatments (p = 0.85; Fig. S2). Community

structure was significantly different between background soil and all treatment groups for both

the bacterial (p < 0.001, Fig. S3a) and fungal communities (p < 0.001, Fig. S3b). There were few

species-level differences, with more variation between groups than between species within

groups (Figs. S4, S5).

Bacterial microbiomes of grass-forb pairs differ from the microbiomes of grasses and forbs

Grass-forb pairs displayed significantly different joint bacterial microbiomes from

microbiomes of both groups grown alone (forb: p < 0.001; grass: p < 0.001; Fig. 2). Grass-forb

pair microbiomes also had marginally higher Shannon diversity than grass microbiomes (p =

0.059), but did not significantly differ from forb microbiomes (p = 0.292). As reported above,

there were no significant differences found between treatments in the fungal microbiome, so we

did not conduct post-hoc analyses to assess pairwise contrasts between treatment groups.

Microbiomes of grass-forb pairs are sourced equally from invasive grasses and native forbs

For bacteria, SourceTracker identified similar fractions of ASVs in grass-forb pairs

predicted to have originated from grasses and from forbs (p = 0.341, grass mean ± se: 40.34 ±

1.33%, forb: 42.13 ± 1.14%). Only relatively low proportions of the community came from

background soil or unknown sources (background: 8.57 ± 0.48%, unknown: 8.96 ± 0.60%). For

fungi, the fraction of the community predicted to originate from grasses was marginally higher

than that from forbs (p = 0.066, grass: 38.88 ± 2.26%, forb: 32.65 ± 2.01%). Additionally, the

fraction of the community originating from forbs was significantly lower than the fraction of the

bacterial community originating from forbs ($p < 0.001$). A small proportion of the fungal

community was predicted to derive from background soil ($5.16 \pm 0.67\%$). A much larger

proportion of the fungal community was estimated to be from unknown sources ($23.31 \pm 2.11\%$),

which was significantly higher than the contribution of unknown sources to the bacterial

community ($p < 0.001$).

Six bacterial and one fungal family displayed differential abundances between forbs,

grasses, and pairs

Differential abundance analysis (DESeq2) revealed 9 bacterial families (out of a total of

302 bacterial families) (Fig. S6a) that differed in abundance between individual grass and forb

microbiomes, with six of these families remaining differentially between joint grass-forb pair

microbiomes and individual microbiomes: Methylophilaceae, Fibrobacteraceae,

Clostridiaceae_1, Burkholderiaceae, Rhodocyclaceae, and Veillonellaceae (Fig. 4). Of these six

bacterial families, only Burkholderiaceae had an untransformed mean relative abundance of

greater than one percent ($11.69 \pm 0.39\%$; Table S1). There were an additional 10 bacterial

families that varied between pairs and individuals, but were not found to be differentially

abundant between grasses and forbs grown alone (Fig. S7).

Methylophilaceae was the only family with higher normalized abundance in forbs than in

grasses and their normalized abundance declined in pairs relative to forbs. Fibrobacteraceae,

Clostridiaceae_1 and Veillonellaceae were all higher in normalized abundance in grasses than in

forbs and increased in grass-forb pairs relative to forbs. Both Burkholderiaceae and

Rhodocyclaceae were higher in grasses than in forbs, but decreased in grass-forb pairs relative to

grasses.

310    Using DESeq2, we found only one fungal family (out of 231) that was differentially

311 abundant between grasses and forbs, and between pairs and individuals (Tubeufiaceae) (Fig.

312 S6b, Fig. S8). This family had higher normalized abundance in forbs than in grasses and

313 subsequently lower normalized abundance in grass-forb pairs than in forbs. Two other families

314 that did not vary between grasses and forbs showed variation between pairs and either grasses or

315 forbs.

316 *Normalized abundances of key microbial families are correlated with grass dominance over*

317 *forbs*

318    When forbs and grasses were grown alone, grass and forb biomass did not significantly

319 differ ($p = 0.435$), however in grass-forb pairs, grass biomass was significantly higher than forb

320 biomass ($p < 0.001$), indicating that despite no inherent differences in biomass between groups at

321 this stage, grasses outperformed forbs when grown together. We found no significant

322 relationship between forb biomass and grass biomass in grass-forb pairs (est = -0.153, se =

323 0.102, $p = 0.138$), suggesting that decreases in forb biomass were not solely due to increases in

324 grass biomass. Instead, individual plant performance in pairs was closely related to microbial

325 abundance. Of the six families that differentially varied between grasses and forbs grown alone,

326 and remained differentially abundant in pairs, all but Rhodocyclaceae displayed relationships

327 with either forb or grass biomass in paired pots (Fig. 5, Table S2). Moreover, the four families

328 that displayed negative correlations with forb biomass and/or positive correlations with grass

329 biomass (Burkholderiaceae, Fibrobacteraceae, Clostridiaceae_1, and Veillonellaceae) had higher

330 predicted proportions originating from grass microbiomes according to SourceTracker, while the

331 one family that displayed a positive correlation with forb biomass and a negative correlation with

grass biomass (Methylophilaceae) had a higher predicted proportion originating from forb

microbiomes (Fig. 6).

The normalized abundance of Methylophilaceae was positively correlated with forb

biomass in pairs, and negatively correlated with grass biomass in pairs, indicating that as the

normalized abundance of this family decreased, grass performance increased and forb

performance decreased. The normalized abundance of Fibrobacteraceae was negatively

correlated with forb biomass in pairs, but was not correlated with grass biomass, indicating that

as the normalized abundance of this family increased, forb performance decreased despite no

change in grass performance. The normalized abundance of Clostridiaceae_1 was positively

correlated with grass biomass, and marginally negatively correlated with grass biomass. The

normalized abundance of Burkholderiaceae was positively correlated with grass biomass and

marginally negatively correlated with forb biomass. Finally, the normalized abundance of

Veillonellaceae was marginally positively correlated with grass biomass.

In the remaining 13 differentially abundant bacterial families, only Weeksellaceae, which

did not vary between grasses and forbs, showed a significant relationship with biomass (Fig. S9,

S10; Table S2). Similarly, only the fungal family Ceratobasidiaceae , which again did not vary

between forbs and grasses, showed a significant relationship with biomass (Fig. S11, S12; Table

S2).

Discussion

The rhizosphere bacterial community of native forbs is structurally distinct from those of

invasive grasses and has higher alpha diversity. Interactions between grasses and forbs here

correlated with shifts in the bacterial rhizosphere towards communities dissimilar to both

individual forb and grass microbiomes, with marginally higher diversity than grasses alone.

Studies in similar annual grasslands have found little change in bacterial structure with increased

invasive grass abundance [61–63] but invader dominance has been linked to decreased bacterial

diversity in other systems [33,64]. While we expected the microbiomes of grass-forb pairs to be

a mixed community sourced from both forbs and grasses, we also expected the joint rhizosphere

to be dominated by the microbes of the dominant competitor as in Hortal et al. [18] and Lozano

et al. [19]. In partial support of this, we found that four out of the six families that varied both

between forbs and grasses and between individuals and pairs came predominantly from grasses

and increased in abundance relative to forbs, while the one family that was majority sourced

from forbs saw a decrease in abundance. At the community level, however, we found that the

novel assemblage of bacteria in pairs was sourced nearly equally from both groups. These results

suggest that the invasions of grasses into native forb habitats may be associated with microbiome

shifts in both groups; however, these changes may have more negative consequences for native

forbs than invasive grasses.

Consistent with invasive annual grasses aiding in native forb declines across California

annual grasslands, we found that grasses outperformed forbs in paired pots despite no differences

when grown alone. Plants were well-watered and harvested before shading could become

problematic, therefore differences in performance likely resulted from below-ground

interactions. Given the large role played by microbes in below-ground interactions, the lack of a

relationship between grass and forb performance in pairs, and the relationship between microbial

abundance and plant performance in pairs, we speculate that invasive grass dominance over

native forbs here is partially mediated by microbes. Invaded grassland soils tend to be depleted in

plant available nitrogen due to the nutrient-demanding nature of invasive grasses [32,62,63], but

with increased microbial activity and faster nitrogen cycling [24,31,63]. These microbe-driven

changes may selectively benefit fast-growing grasses over natives, especially in nutrient-poor

serpentine soils where added nitrogen has been found to preferentially aid invasives [51,65].

Invasive grasses may further benefit from decreased microbial diversity, which has been linked

to both decreased forb nitrogen uptake and increased grass nitrogen uptake [66]. Beyond

competition for nutrients, grasses and forbs may also indirectly compete through recruitment of

microbes harmful to natives [67] and/or helpful to invasives [36]. Other studies in serpentine

grasslands have shown that soil primed with invasive grass negatively affected native forb

growth due to changes in the microbial community [68,69]. Although future work should be

conducted to tease apart the role of plant density on microbial shifts and to understand the degree

to which microbial shifts drive competitive outcomes in this system, our results add to a body of

evidence that the invasion of grasses alters the microbiome and suggest that their dominance may

be linked to these shifts.

Plant performance in pairs could not be explained by the performance of their neighbor.

Instead, plant performance in pairs was tied to the abundance of key bacterial families. We found

links between bacterial normalized abundance and plant biomass in grass-forb pairs for five of

the six main differentially abundant families: three that were correlated with decreased forb

performance: Burkholderiacae (marginal), Clostridacaeae_1 (marginal), and Fibrobacteraceae,

three that were correlated with increased grass performance: Burkholderiaceae, Clostridiaceae_1

(marginal), and Veillonellaceae (marginal), and one that was correlated with both decreased

grass performance and increased forb performance: Methylophilaceae.

Methylophilaceae have been found in soils with high heavy metal content and include

microbes critical for heavy metal attenuation and immobilization [70,71]. Members of

Methylophilaceae also form beneficial symbioses with plants [72–74]. This family may include

important taxa for serpentine-adapted species with roles in heavy metal attenuation and plant

growth [73]. Forbs contributed a higher percentage of ASVs from this family to the joint grass-

forb rhizosphere microbiome, further supporting close associations between native forbs and this

group. If interactions with invasive grasses are driving declines in this family as our results

suggest, grass dominance in this system may be partially due to a decrease in locally adapted

microbes similar to Cavalieri et al. [30].

Grasses were likely driving higher normalized abundances of Fibrobacteraceae,

Veillonellaceae, and Clostridiaceae_1, as they contributed a higher percentage of ASVs from

these families to the joint rhizosphere microbiome in pairs. Although Fibrobacteraceae is

typically lower in abundance in soils with high heavy metal content [71], this family is also

positively associated with invasive grass dominance [75]. Further, all three families are known

for their cellulose-degrading properties [76–78], and therefore may contain important taxa for

organic matter degradation [79]. Both Fibrobacteraceae and Clostridiaceae_1 contain known

nitrogen-fixers [80–82]. If taxa in these families are increasing nutrient availability, they could

be disproportionately helping fast-growing invasive annual grasses [51,65].

Burkholderiaceae are considered keystone members in grasslands as endophytes or

pathogens [83–86] and are known for their antimicrobial properties [87–89], nitrogen-fixing

abilities [80,90], and competitive dominance, especially in N-limited environments [91]. Some

members are even known to suppress fungal pathogens [92,93]. It is possible that the

Burkholderiacae observed in pairs here are either providing positive functional benefits to

grasses, or are less harmful to grasses than forbs as pathogens.

In contrast to our bacterial findings, we found no significant differences in fungal

communities across treatments, either in terms of structure or diversity. This suggests loose

fungal associations with plant hosts, possibly due to functional redundancy of fungi across large

geographic scales [94]. Within annual grasslands, however, there is evidence of invader-driven

fungal shifts [61] and increases in fungal relative abundance [95]. Our limited fungal findings

may indicate that fungi are generally less important than bacteria for competitive outcomes

between plants [96]. Alternatively, fungi may be important, but complex fungal-fungal

interactions at the community level may cancel out fitness effects on plants by beneficial and

detrimental fungi [97]. In addition, experimental constraints such as unaccounted fungal regional

source pools (such as airborne local spores or local watersheds [94]), plant age and development

at time of sampling [16], and methodology (e.g. sampling rhizosphere here instead of root

endophytes (e.g. Emam et al. [98]) or fungi in the rhizoplane (e.g. Edwards et al. [1])), may have

prevented the observation of interactions between the fungal community and plant competition.

Conclusions

Native annual forbs are host to a community of microbes distinct from, and more diverse

than, those of their invasive competitors, but these close associations may be disrupted by

invasive grasses. The joint bacterial rhizosphere of invasive grasses and native forbs differed

from those of plants grown alone, with grasses contributing more to the abundance of ASVs

from families that were linked to decreases in forb performance and/or increased in grass

performance. Moreover, three of these families were higher in abundance in joint microbiomes

relative to those of forbs alone. Forbs on the other hand contributed more to the abundance of

ASVs from only one family that was associated with increased forb performance and decreased

grass performance, but this family declined in abundance in the joint rhizosphere relative to

forbs. Given the correlative nature of our study, more research is needed to understand the role of

the microbiome in invasive grass dominance over native forbs and to clarify whether the

observed changes in plant performance were indeed driven by these candidate bacterial families.

Regardless, our study highlights the importance of considering the microbiome in ecosystems

facing dominance by invasive species.

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References

1. Edwards J, Johnson C, Santos-Medellin C, Lurie E, Podishetty NK, Bhatnagar S, et al.

Structure, variation, and assembly of the root-associated microbiomes of rice. Proceedings of the

National Academy of Sciences of the United States of America. 2015;112:E911–20.

2. Richardson AE, Barea J-M, McNeill AM, Prigent-Combaret C. Acquisition of phosphorus and

nitrogen in the rhizosphere and plant growth promotion by microorganisms. Plant Soil.

2009;321:305–39.

3. Yang J, Kloepper JW, Ryu C-M. Rhizosphere bacteria help plants tolerate abiotic stress.

Trends in Plant Science. 2009;14:1–4.

4. Doubková P, Suda J, Sudová R. The symbiosis with arbuscular mycorrhizal fungi contributes

to plant tolerance to serpentine edaphic stress. Soil Biology and Biochemistry. 2012;44:56–64.

5. Berendsen RL, Pieterse CMJ, Bakker PAHM. The rhizosphere microbiome and plant health.

Trends in Plant Science. 2012;17:478–86.

6. Philippot L, Raaijmakers JM, Lemanceau P, van der Putten WH. Going back to the roots: the

microbial ecology of the rhizosphere. Nature Reviews Microbiology. Nature Publishing Group;

2013;11:789–99.

7. Bever JD, Platt TG, Morton ER. Microbial Population and Community Dynamics on Plant

Roots and Their Feedbacks on Plant Communities. Annu Rev Microbiol. Annual Reviews;

2012;66:265–83.

8. Xue P-P, Carrillo Y, Pino V, Minasny B, McBratney AB. Soil Properties Drive Microbial

Community Structure in a Large Scale Transect in South Eastern Australia. Scientific Reports.

Nature Publishing Group; 2018;8:11725.

9. Singh BK, Munro S, Potts JM, Millard P. Influence of grass species and soil type on

rhizosphere microbial community structure in grassland soils. Applied Soil Ecology.

2007;36:147–55.

10. de Ridder-Duine AS, Kowalchuk GA, Klein Gunnewiek PJA, Smant W, van Veen JA, de

Boer W. Rhizosphere bacterial community composition in natural stands of Carex arenaria (sand

sedge) is determined by bulk soil community composition. Soil Biology and Biochemistry.

2005;37:349–57.

11. Aleklett K, Leff JW, Fierer N, Hart M. Wild plant species growing closely connected in a

subalpine meadow host distinct root-associated bacterial communities. PeerJ. PeerJ Inc.;

2015;3:e804.

12. Lakshmanan V. Chapter Three - Root Microbiome Assemblage is Modulated by Plant Host

Factors. In: Bais H, Sherrier J, editors. Advances in Botanical Research. Academic Press; 2015.

p. 57–79.

13. Rúa MA, Antoninka A, Antunes PM, Chaudhary VB, Gehring C, Lamit LJ, et al. Home-field

advantage? evidence of local adaptation among plants, soil, and arbuscular mycorrhizal fungi

through meta-analysis. BMC Evol Biol. 2016;16:122.

14. Schultz PA, Michael Miller R, Jastrow JD, Rivetta CV, Bever JD. Evidence of a mycorrhizal

mechanism for the adaptation of Andropogon gerardii (Poaceae) to high- and low-nutrient

prairies. Am J Bot. 2001;88:1650–6.

15. Johnson NC, Wilson GWT, Bowker MA, Wilson JA, Miller RM. Resource limitation is a

driver of local adaptation in mycorrhizal symbioses. PNAS. National Academy of Sciences;

2010;107:2093–8.

16. Chaparro JM, Badri DV, Vivanco JM. Rhizosphere microbiome assemblage is affected by

plant development. The ISME Journal. Nature Publishing Group; 2014;8:790–803.

17. Siefert A, Zillig KW, Friesen ML, Strauss SY. Soil microbial communities alter conspecific

and congeneric competition consistent with patterns of field coexistence in three Trifolium

congeners. Journal of Ecology. 2018;106:1876–91.

18. Hortal S, Lozano YM, Bastida F, Armas C, Moreno JL, Garcia C, et al. Plant-plant

competition outcomes are modulated by plant effects on the soil bacterial community. Sci Rep.

2017;7:17756.

19. Lozano YM, Hortal S, Armas C, Pugnaire FI. Soil micro-organisms and competitive ability

of a tussock grass species in a dry ecosystem. Journal of Ecology. 2019;107:1215–25.

20. Kandlikar GS, Johnson CA, Yan X, Kraft NJB, Levine JM. Winning and losing with

microbes: how microbially mediated fitness differences influence plant diversity. Ecology

Letters. 2019;22:1178–91.

21. Kandlikar GS, Yan X, Levine JM, Kraft NJB. Soil Microbes Generate Stronger Fitness

Differences than Stabilization among California Annual Plants. The American Naturalist. The

University of Chicago Press; 2021;197:E30–9.

22. Franche C, Lindström K, Elmerich C. Nitrogen-fixing bacteria associated with leguminous

and non-leguminous plants. Plant Soil. 2009;321:35–59.

23. Heijden MGAVD, Bardgett RD, Straalen NMV. The unseen majority: soil microbes as

drivers of plant diversity and productivity in terrestrial ecosystems. Ecology Letters.

2008;11:296–310.

24. Hawkes CV, Wren IF, Herman DJ, Firestone MK. Plant invasion alters nitrogen cycling by

modifying the soil nitrifying community. Ecology Letters. 2005;8:976–85.

25. Lee MR, Flory SL, Phillips RP. Positive feedbacks to growth of an invasive grass through

alteration of nitrogen cycling. Oecologia. 2012;170:457–65.

26. Funk JL, Vitousek PM. Resource-use efficiency and plant invasion in low-resource systems.

Nature. 2007;446:1079–81.

27. Hodge A, Fitter AH. Microbial mediation of plant competition and community structure.

Functional Ecology. 2013;27:865–75.

28. Fitzpatrick CR, Copeland J, Wang PW, Guttman DS, Kotanen PM, Johnson MTJ. Assembly

and ecological function of the root microbiome across angiosperm plant species. PNAS.

2018;115:E1157–65.

29. Eppinga MB, Rietkerk Max, Dekker Stefan C., De Ruiter Peter C., Van der Putten Wim H.,

Van der Putten Wim H. Accumulation of local pathogens: a new hypothesis to explain exotic

plant invasions. Oikos. 2006;114:168–76.

30. Cavalieri A, Bak F, Garcia-Lemos AM, Weiner J, Nicolaisen MH, Nybroe O. Effects of

Intra- and Interspecific Plant Density on Rhizosphere Bacterial Communities. Front Microbiol

[Internet]. Frontiers; 2020 [cited 2020 Oct 26];11. Available from:

https://www.frontiersin.org/articles/10.3389/fmicb.2020.01045/full

31. Esch EH, Hernández DL, Pasari JR, Kantor RSG, Selmants PC. Response of soil microbial

activity to grazing, nitrogen deposition, and exotic cover in a serpentine grassland. Plant Soil.

2013;366:671–82.

32. Carey CJ, Blankinship JC, Eviner VT, Malmstrom CM, Hart SC. Invasive plants decrease

microbial capacity to nitrify and denitrify compared to native California grassland communities.

Biol Invasions. 2017;19:2941–57.

33. Shi S, Nuccio E, Herman DJ, Rijkers R, Estera K, Li J, et al. Successional Trajectories of

Rhizosphere Bacterial Communities over Consecutive Seasons. mBio [Internet]. American

Society for Microbiology; 2015 [cited 2020 Oct 7];6. Available from:

https://mbio.asm.org/content/6/4/e00746-15

34. Komatsu KJ, Simms EL. Invasive legume management strategies differentially impact

mutualist abundance and benefit to native and invasive hosts. Restoration Ecology.

2020;28:378–86.

35. Rodríguez-Caballero G, Caravaca F, Díaz G, Torres P, Roldán A. The invader Carpobrotus

edulis promotes a specific rhizosphere microbiome across globally distributed coastal

ecosystems. Science of The Total Environment. 2020;719:137347.

36. Rout ME, Chrzanowski TH, Westlie TK, DeLuca TH, Callaway RM, Holben WE. Bacterial

endophytes enhance competition by invasive plants. American Journal of Botany.

2013;100:1726–37.

37. Callaway RM, Cipollini D, Barto K, Thelen GC, Hallett SG, Prati D, et al. Novel Weapons:

Invasive Plant Suppresses Fungal Mutualists in America but Not in Its Native Europe. Ecology.

2008;89:1043–55.

38. Blank RR, Mackey B, Morgan T. Do a native and two exotic grasses respond differently in

soils conditioned by native vegetation versus an exotic grass? Rhizosphere. 2016;2:38–47.

39. Eskelinen A, Harrison S. Exotic plant invasions under enhanced rainfall are constrained by

soil nutrients and competition. Ecology. 2014;95:682–92.

40. Davies KW. Plant community diversity and native plant abundance decline with increasing

abundance of an exotic annual grass. Oecologia. 2011;167:481–91.

41. LaForgia ML, Harrison SP, Latimer AM. Invasive species interact with climatic variability to

reduce success of natives. Ecology. 2020;n/a:e03022.

42. Bennett AE, Thomsen M, Strauss SY. Multiple mechanisms enable invasive species to

suppress native species. American Journal of Botany. 2011;98:1086–94.

43. Igwe AN, Vannette RL. Bacterial communities differ between plant species and soil type,

and differentially influence seedling establishment on serpentine soils. Plant Soil. 2019;441:423–

37.

44. Porter SS, Chang PL, Conow CA, Dunham JP, Friesen ML. Association mapping reveals

novel serpentine adaptation gene clusters in a population of symbiotic Mesorhizobium. The

ISME Journal. Nature Publishing Group; 2017;11:248–62.

45. Clary J. McLaughlin Reserve [Internet]. naturalreserves. 2017 [cited 2020 Dec 3]. Available

from: https://naturalreserves.ucdavis.edu/mclaughlin-reserve

46. Sun C, Liu G, Xue S. Interaction Between Plant Competition and Rhizospheric Bacterial

Community Influence Secondary Succession of Abandoned Farmland on the Loess Plateau of

China. Front Plant Sci [Internet]. Frontiers; 2018 [cited 2020 Oct 29];9. Available from:

https://www.frontiersin.org/articles/10.3389/fpls.2018.00898/full

47. LaForgia ML. Impacts of invasive annual grasses and their litter vary by native functional

strategy. Biol Invasions [Internet]. 2021 [cited 2021 Apr 28]; Available from:

https://doi.org/10.1007/s10530-021-02527-2

48. Carmona CP, Bello F de, Azcárate FM, Mason NWH, Peco B. Trait hierarchies and

intraspecific variability drive competitive interactions in Mediterranean annual plants. Journal of

Ecology. 2019;107:2078–89.

49. Goldberg DE, Landa K. Competitive Effect and Response: Hierarchies and Correlated Traits

in the Early Stages of Competition. Journal of Ecology. [Wiley, British Ecological Society];

1991;79:1013–30.

50. Miller TE. Effects of emergence time on survival and growth in an early old-field plant

community. Oecologia. 1987;72:272–8.

51. Going BM, Hillerislambers J, Levine JM. Abiotic and biotic resistance to grass invasion in

serpentine annual plant communities. Oecologia. 2009;159:839–47.

52. Apprill A, McNally S, Parsons R, Weber L. Minor revision to V4 region SSU rRNA 806R

gene primer greatly increases detection of SAR11 bacterioplankton. Aquatic Microbial Ecology.

2015;75:129–37.

53. Parada AE, Needham DM, Fuhrman JA. Every base matters: assessing small subunit rRNA

primers for marine microbiomes with mock communities, time series and global field samples.

Environmental Microbiology. 2016;18:1403–14.

54. Beeck MOD, Lievens B, Busschaert P, Declerck S, Vangronsveld J, Colpaert JV.

Comparison and Validation of Some ITS Primer Pairs Useful for Fungal Metabarcoding Studies.

PLOS ONE. Public Library of Science; 2014;9:e97629.

55. Callahan BJ, McMurdie PJ, Rosen MJ, Han AW, Johnson AJA, Holmes SP. DADA2: High-

resolution sample inference from Illumina amplicon data. Nature Methods. Nature Publishing

Group; 2016;13:581–3.

56. R Core Team. R: A language and environment for statistical computing. Vienna, Austria: R

Foundation for Statistical Computing; 2018.

57. Dixon P. VEGAN, a package of R functions for community ecology. Journal of Vegetation

Science. 2003;14:927–30.

58. Salazar G. EcolUtils: Utilities for community ecology analysis [Internet]. 2020 [cited 2020

Dec 7]. Available from: https://github.com/GuillemSalazar/EcolUtils

59. Knights D, Kuczynski J, Charlson ES, Zaneveld J, Mozer MC, Collman RG, et al. Bayesian

community-wide culture-independent microbial source tracking. Nature Methods. Nature

Publishing Group; 2011;8:761–3.

60. Love MI, Huber W, Anders S. Moderated estimation of fold change and dispersion for RNA-

seq data with DESeq2. Genome Biology. 2014;15:550.

61. Gravuer K, Eskelinen A, Winbourne JB, Harrison SP. Vulnerability and resistance in the

spatial heterogeneity of soil microbial communities under resource additions. PNAS. National

Academy of Sciences; 2020;117:7263–70.

62. Carey CJ, Beman JM, Eviner VT, Malmstrom CM, Hart SC. Soil microbial community

structure is unaltered by plant invasion, vegetation clipping, and nitrogen fertilization in

experimental semi-arid grasslands. Front Microbiol [Internet]. Frontiers; 2015 [cited 2020 Jun

1];6. Available from: https://www.frontiersin.org/articles/10.3389/fmicb.2015.00466/full

63. Dickens SJM, Allen EB, Santiago LS, Crowley D. Exotic annuals reduce soil heterogeneity

in coastal sage scrub soil chemical and biological characteristics. Soil Biology and Biochemistry.

2013;58:70–81.

64. Kowalchuk GeorgeA, Buma DS, de Boer W, Klinkhamer PGL, van Veen JA. Effects of

above-ground plant species composition and diversity on the diversity of soil-borne

microorganisms. Antonie Van Leeuwenhoek. 2002;81:509.

65. Eskelinen A, Harrison S. Erosion of beta diversity under interacting global change impacts in

a semi-arid grassland. Journal of Ecology. 2015;103:397–407.

66. Wagg C, Schlaeppi K, Banerjee S, Kuramae EE, van der Heijden MGA. Fungal-bacterial

diversity and microbiome complexity predict ecosystem functioning. Nat Commun.

2019;10:4841.

67. Stinson KA, Campbell SA, Powell JR, Wolfe BE, Callaway RM, Thelen GC, et al. Invasive

Plant Suppresses the Growth of Native Tree Seedlings by Disrupting Belowground Mutualisms.

PLOS Biology. Public Library of Science; 2006;4:e140.

68. Batten KM, Scow KM, Espeland EK. Soil Microbial Community Associated with an

Invasive Grass Differentially Impacts Native Plant Performance. Microb Ecol. 2008;55:220–8.

69. Ba L, Facelli E, Facelli JM. Plant-mycorrhizal fungi feedbacks: potential accomplices of

Avena barbata's high invasiveness. Plant Ecol. 2018;219:1045–52.

70. Drewniak L, Krawczyk PS, Mielnicki S, Adamska D, Sobczak A, Lipinski L, et al.

Physiological and Metagenomic Analyses of Microbial Mats Involved in Self-Purification of

Mine Waters Contaminated with Heavy Metals. Front Microbiol [Internet]. Frontiers; 2016

[cited 2020 Jul 2];7. Available from:

https://www.frontiersin.org/articles/10.3389/fmicb.2016.01252/full

71. Syranidou E, Thijs S, Avramidou M, Weyens N, Venieri D, Pintelon I, et al. Responses of

the Endophytic Bacterial Communities of Juncus acutus to Pollution With Metals, Emerging

Organic Pollutants and to Bioaugmentation With Indigenous Strains. Front Plant Sci [Internet].

Frontiers; 2018 [cited 2020 Jul 2];9. Available from:

https://www.frontiersin.org/articles/10.3389/fpls.2018.01526/full

72. Trotsenko YuA, Ivanova EG, Doronina NV. Aerobic Methylotrophic Bacteria as

Phytosymbionts. Microbiology. 2001;70:623–32.

73. Fedorov DN, Doronina NV, Trotsenko YuA. Phytosymbiosis of aerobic methylobacteria:

New facts and views. Microbiology. 2011;80:443.

74. Macey MC, Pratscher J, Crombie AT, Murrell JC. Impact of plants on the diversity and

activity of methylotrophs in soil. Microbiome. 2020;8:31.

75. Fahey C, Koyama A, Antunes PM, Dunfield K, Flory SL. Plant communities mediate the

interactive effects of invasion and drought on soil microbial communities. The ISME Journal.

Nature Publishing Group; 2020;1–14.

76. Rubino F, Carberry C, M Waters S, Kenny D, McCabe MS, Creevey CJ. Divergent

functional isoforms drive niche specialisation for nutrient acquisition and use in rumen

microbiome. ISME J. 2017;11:932–44.

77. Ozbayram EG, Ince O, Ince B, Harms H, Kleinsteuber S. Comparison of Rumen and Manure

Microbiomes and Implications for the Inoculation of Anaerobic Digesters. Microorganisms.

Multidisciplinary Digital Publishing Institute; 2018;6:15.

78. Vahidi MF, Gharechahi J, Behmanesh M, Ding X-Z, Han J-L, Salekdeh GH. Diversity of

microbes colonizing forages of varying lignocellulose properties in the sheep rumen. PeerJ. PeerJ

Inc.; 2021;9:e10463.

79. Nuccio EE, Starr E, Karaoz U, Brodie EL, Zhou J, Tringe SG, et al. Niche differentiation is

spatially and temporally regulated in the rhizosphere. ISME J. 2020;1–16.

80. Dos Santos PC, Fang Z, Mason SW, Setubal JC, Dixon R. Distribution of nitrogen fixation

and nitrogenase-like sequences amongst microbial genomes. BMC Genomics. 2012;13:162.

81. Abdul Rahman N, Parks DH, Vanwonterghem I, Morrison M, Tyson GW, Hugenholtz P. A

Phylogenomic Analysis of the Bacterial Phylum Fibrobacteres. Front Microbiol [Internet]. 2016

[cited 2020 Oct 7];6. Available from: https://www.ncbi.nlm.nih.gov/pmc/articles/PMC4704652/

82. Pepe-Ranney C, Koechli C, Potrafka R, Andam C, Eggleston E, Garcia-Pichel F, et al. Non-

cyanobacterial diazotrophs mediate dinitrogen fixation in biological soil crusts during early crust

formation. ISME J. 2016;10:287–98.

83. Eberl L, Vandamme P. Members of the genus Burkholderia: good and bad guys. F1000Res.

2016;5:1007.

84. Banerjee S, Schlaeppi K, Heijden MGA van der. Keystone taxa as drivers of microbiome

structure and functioning. Nature Reviews Microbiology. 2018;16:567.

85. Genin S, Boucher C. Lessons learned from the genome analysis of ralstonia solanacearum.

Annu Rev Phytopathol. Annual Reviews; 2004;42:107–34.

86. Mansfield J, Genin S, Magori S, Citovsky V, Sriariyanum M, Ronald P, et al. Top 10 plant

pathogenic bacteria in molecular plant pathology. Molecular Plant Pathology. 2012;13:614–29.

87. Becker MH, Brucker RM, Schwantes CR, Harris RN, Minbiole KPC. The Bacterially

Produced Metabolite Violacein Is Associated with Survival of Amphibians Infected with a

Lethal Fungus. Appl Environ Microbiol. American Society for Microbiology; 2009;75:6635–8.

88. Depoorter E, Bull MJ, Peeters C, Coenye T, Vandamme P, Mahenthiralingam E.

Burkholderia: an update on taxonomy and biotechnological potential as antibiotic producers.

Appl Microbiol Biotechnol. 2016;100:5215–29.

89. Shehata HR, Ettinger CL, Eisen JA, Raizada MN. Genes Required for the Anti-fungal

Activity of a Bacterial Endophyte Isolated from a Corn Landrace Grown Continuously by

Subsistence Farmers Since 1000 BC. Front Microbiol [Internet]. Frontiers; 2016 [cited 2020 Oct

8];7. Available from: https://www.frontiersin.org/articles/10.3389/fmicb.2016.01548/full

90. Estrada-de los Santos P, Palmer M, Chávez-Ramírez B, Beukes C, Steenkamp ET, Briscoe L,

et al. Whole Genome Analyses Suggests that Burkholderia sensu lato Contains Two Additional

Novel Genera (Mycetohabitans gen. nov., and Trinickia gen. nov.): Implications for the

Evolution of Diazotrophy and Nodulation in the Burkholderiaceae. Genes (Basel) [Internet].

2018 [cited 2020 Oct 7];9. Available from:

https://www.ncbi.nlm.nih.gov/pmc/articles/PMC6116057/

91. Elliott GN, Chou J-H, Chen W-M, Bloemberg GV, Bontemps C, Martínez-Romero E, et al.

Burkholderia spp. are the most competitive symbionts of Mimosa, particularly under N-limited

conditions. Environmental Microbiology. 2009;11:762–78.

92. Kang Y, Carlson R, Tharpe W, Schell MA. Characterization of Genes Involved in

Biosynthesis of a Novel Antibiotic from Burkholderia cepacia BC11 and Their Role in

Biological Control of Rhizoctonia solani. Appl Environ Microbiol. 1998;64:3939–47.

93. Chapelle E, Mendes R, Bakker PAH, Raaijmakers JM. Fungal invasion of the rhizosphere

microbiome. The ISME Journal. Nature Publishing Group; 2016;10:265–8.

94. Talbot JM, Bruns TD, Taylor JW, Smith DP, Branco S, Glassman SI, et al. Endemism and

functional convergence across the North American soil mycobiome. PNAS. National Academy

of Sciences; 2014;111:6341–6.

95. Wolkovich EM, Lipson DA, Virginia RA, Cottingham KL, Bolger DT. Grass invasion

causes rapid increases in ecosystem carbon and nitrogen storage in a semiarid shrubland. Glob

Change Biol. 2010;16:1351–65.

96. Levine JM, Adler PB, Yelenik SG. A meta-analysis of biotic resistance to exotic plant

invasions. Ecology letters. 2004;7:975–89.

97. Facelli E, Smith SE, Facelli JM, Christophersen HM, Smith FA. Underground friends or

enemies: model plants help to unravel direct and indirect effects of arbuscular mycorrhizal fungi

on plant competition. New Phytologist. 2010;185:1050–61.

98. Emam T. Local soil, but not commercial AMF inoculum, increases native and non-native

grass growth at a mine restoration site. Restoration Ecology. 2016;24:35–44.

Figure 1. (a) Native annual forb species (from left to right, top to bottom): *Clarkia purpurea,*

Calycadenia pauciflora, Lasthenia californica, Agoseris heterophylla, Plantago erecta, and

Hemizonia congesta. (b) Invasive annual grasses (from left to right): *Bromus hordeaceus,*

Taeniatherum caput-medusae, and *Avena fatua.* (c) Shade-house set-up. Photos of *C. purpurea*,

P. erecta, H. congesta, *C. pauciflora*, *B. hordeaceus*, *T. caput-medusae*, and *A. fatua* by Paul

Aigner. Remaining photos by authors.

Figure 2. Bacterial microbiome structure differs between grasses and forbs, and shifts during

competition. Principal coordinates analysis (PCoA) visualization of weighted UniFrac distances

of bacterial communities associated with the rhizosphere. Points in the ordination are colored and

represented by shapes based on the rhizosphere of forbs grown alone (purple circles), grasses

grown alone (green triangles) and grass-forb pairs, i.e. competition (yellow squares). Ellipses

represent the 95% confidence interval around the centroid of each group.

Figure 3. Shannon diversity of bacterial microbiomes is highest in forbs. Shannon diversity was

used to assess alpha diversity for bacterial microbiomes for each treatment (forbs grown alone,

grasses grown alone, competition). Comparisons that are significantly different from each other

are notated by different letters (e.g. a vs. b).

Figure 4. Bacterial families that are differentially abundant between grass and forb microbiomes,

and between grass-forb pairs and either grass or forb microbiomes. Using DESeq2, bacterial

families were identified whose abundance differed significantly between treatments (forbs grown

alone, grasses grown alone, grown together in competition). Each plot shows the normalized

abundance of families for each treatment and comparisons that are significant from each other

are notated by different letters (e.g. a vs. b).

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762   **Figure 5.** Bacterial family abundances correlated with forb (purple) and grass (green) biomass in

763   pairs. Solid line indicates significance ($p < 0.05$), dashed line indicates marginal significance (p

764   < 0.10). For regression results see Table S2.

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Figure 6. Proportion of ASVs predicted to colonize from each source for each of the six

differentially abundant families. SourceTracker was used to predict whether ASVs in

competition treatments originated from grasses (dark green), forbs (purple), the background soil

mix used in the experiment (blue), or unknown habitats (light green) and then the results were

summarized by grouping ASVs by taxonomic family. Families that were negatively correlated

with forb performance and/or positively correlated with grass performance were predominantly

sourced from grass. Methylophilaceae, the one family that showed a positive relationship with

forb biomass under competition and a negative relationship with grass biomass under

competition, was predominantly sourced from forbs.

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