

1 **Climate warming changes synchrony of plants and pollinators**

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3 **Running head:** Plant pollinator shifts

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5 Jonas Freimuth¹, Oliver Bossdorf¹, J. F. Scheepens², Franziska M. Willems^{1,3}

6

7 ¹Plant Evolutionary Ecology, Institute of Evolution and Ecology, University of Tübingen, Auf
8 der Morgenstelle 5, 72076 Tübingen, Germany

9 ²Plant Evolutionary Ecology, Faculty of Biological Sciences, Goethe University Frankfurt,
10 Max-von-Laue-Str. 13, 60438 Frankfurt am Main, Germany

11 ³ Conservation Biology, University of Marburg, Marburg, Germany (present address)

12

13 **Correspondence:** Franziska Willems, email: franziska.willems@biologie.uni-marburg.de,
14 phone: +49 6421 2823374

15

16 **OrcIDs:**

17 Jonas Freimuth: <https://orcid.org/0000-0002-9259-3696>

18 Oliver Bossdorf: <https://orcid.org/0000-0001-7504-6511>

19 J.F. Scheepens: <https://orcid.org/0000-0003-1650-2008>

20 Franziska Willems: <https://orcid.org/0000-0002-5481-3686>

21

22 **Abstract**

23 Climate warming changes the phenology of many species. When interacting organisms
24 respond differently, climate change may disrupt their interactions and affect the stability of
25 ecosystems. Here, we used GBIF occurrence records to examine phenology trends in plants
26 and their associated insect pollinators in Germany since the 1980s. We found strong
27 phenological advances in plants, but differences in the extent of shifts among pollinator
28 groups. The temporal trends in plant and insect phenologies were generally associated with
29 interannual temperature variation, and thus likely driven by climate change. When examining
30 the synchrony of species-level plant-pollinator interactions, their temporal trends differed
31 among pollinator groups. Overall, plant-pollinator interactions become more synchronized,
32 mainly since the phenology of plants responded more strongly to climate change than that of
33 the pollinators. However, if the observed trends continue, many interactions may become
34 more asynchronous again in the future. Our study suggests that climate change affects the
35 phenologies of both plants and insects, and that it also influences the synchrony of plant-
36 pollinator interactions.

37

38 **Keywords:** asynchrony, GBIF, mismatch, phenology, pollination mode

39 **Introduction**

40 Phenological events are periodically occurring events in the life cycle of organisms. The
41 timing of these events often depends on environmental factors such as temperature or
42 photoperiod, and it is well known that climate change affects some of these and thus changes
43 the phenologies of many organisms [1,2]. With such phenology shifts, there is increasing
44 concern about possible phenological mismatches between interacting organisms, which could
45 exceed the natural resilience of ecosystems [3,4]. Climate change-induced phenological shifts
46 have been documented extensively for individual species [5], but we still know much less
47 about how these shifts affect ecological interactions. Kharouba et al. [6] recently reviewed 54
48 published interaction studies across ecosystems and interaction types and found no clear
49 general trend, with about half of the studied interactions becoming more asynchronous but the
50 other half becoming even more synchronized through climate change.

51 Plant-pollinator systems are among the biotic interactions expected to suffer most from
52 a mismatch of phenological events [7]. Several previous studies have observed mismatches
53 [8,9], but in others plants and pollinators seemed to be able to keep up with each other [10].
54 An interesting question in this context is also which of the two partners is advancing faster if
55 there is an increasing mismatch. So far, the evidence here is also mixed. For instance Gordo
56 & Sanz [11] found pollinators to advance faster than trees, and Parmesan [5] that butterflies
57 advanced faster than herbaceous plants, but in a study by Kudo & Ida [12] it was the plants –
58 spring ephemerals – that advanced faster than their bee pollinators.

59 Mismatches of plant-pollinator interactions can have negative consequences for both
60 partners. For the pollinators, this can include lower survival rates, a decreased overall fitness
61 and higher parasite loads [13]. Moreover, mismatches might also impact pollinator
62 demography, the body sizes [8] and frequencies of sexes, and thus population viability [13].
63 On the plant side, desynchronized pollinator interactions are mainly expected to impact plant

64 fitness and thus long-term population growth and survival. For instance, Kudo & Ida [12]
65 found that seed counts were reduced in early-flowering spring ephemerals after
66 desynchronization with their bee pollinators. However, in another study fly-pollinated plants
67 did not show similar responses [14].

68 Testing hypotheses about plant-pollinator responses to climate change is not trivial.
69 Since changes in phenology take place on the scale of decades [15], we need long-term data.
70 A possible source of long-term data on plant phenology are herbarium specimens [16,17],
71 which can indicate the day of year that a specific species was flowering in a given location
72 and year. Herbarium data provide unique historical depth, but they need to be treated with
73 caution because of the sampling biases associated with them [18,19]. In recent years the
74 digitization of herbaria as well as other collections and observation data, including on other
75 taxa such as pollinating insects, e.g. from long-term monitoring networks, is creating an
76 increasing number of public data bases that contain vast amounts of natural history data of
77 large spatial and temporal scales [20]. These data are increasingly used for analyses of broad
78 ecological trends and global changes [19,21]. One of the largest and most important hubs of
79 large-scale and long-term ecological data sets is the Global Biodiversity Facility (GBIF), an
80 intergovernmental initiative and public data base that provides access to biodiversity data
81 compiled from various individual sources like academic institutions, government agencies or
82 independent collections [22].

83 Another matter is finding a measure for changes in phenology. Primack et al. [23]
84 demonstrated that the collection dates of herbarium specimens of a plant species in a year can
85 be used as a proxy for flowering time in that year. The same approach of using occurrence
86 records in natural history collections or other data bases can in principle be used to estimate
87 the activity times of other groups of organisms such as insects ([6] and references therein).
88 For instance, analyses of natural history collections in the UK have demonstrated phenology

89 changes in bees [9] and butterflies [24]. Thus, the occurrences of plants and insects in GBIF
90 may be used to estimate activity shifts of different taxonomic groups, as well as their
91 synchrony. When we use the term ‘activity’ in this paper, we refer to the period in an
92 organism’s life when it can interact with its ecological partner. For plants this is the period of
93 flowering, for insect pollinators, such as bees, flies, beetles, and butterflies, it is the period of
94 flight.

95 We used data from GBIF to study phenological shifts of plants and insect pollinators in
96 Germany, at the level of taxonomic groups as well as individual species’ interactions. We
97 asked the following questions: (i) Are there long-term trends in the phenology of plants and
98 pollinators? (ii) If yes, are phenology trends related to climate change? (iii) How does climate
99 change affect the synchrony of interactions between individual plant and pollinator species?

100

101 **Methods**

102 *Phenology data*

103 We worked with occurrence records of plants and insects available from the GBIF database
104 [25–28]. For the plants, we restricted ourselves to species covered by the BioFlor database of
105 plant traits [29], because we originally intended to classify plants by their level of pollinator
106 dependence – an idea we later abandoned. For the insects we restricted ourselves to beetles
107 (Coleoptera), flies (Diptera), bees (Hymenoptera) as well as butterflies and moths
108 (Lepidoptera), as these groups contain most insect pollinators [30]. We used the R package
109 *rgbif* [31] to download all available records of the above taxa from GBIF. Our basic criteria
110 for including records were that they originated from Germany, and that their basis of record
111 (as defined in GBIF) was either a living specimen (e.g., a captured insect), a human
112 observation (i.e., an observation of a species made without collecting it), just an observation
113 (i.e., when the exact type of observation was not clear), or a preserved specimen (e.g., an

114 herbarium record or a collected specimen). If names of plant species were not accepted
115 names, we used the R package *taxsize* [32] to check the names against the GBIF backbone
116 taxonomy and determine the actual accepted name.

117 Prior to the data analyses, we subjected the data to several steps of quality control. First,
118 we removed all records from before 1980 as these turned out to be too inconsistent, with few
119 records per year and large gaps due to consecutive years without records. We also removed
120 the records from 2021 as the year had not been complete at the time of our analysis. Second,
121 we removed all records from the days of year (DOY) 1, 95, 121, 163, 164, 166 and 181, and
122 in particular DOY 365 and 366 from the National Museum of Natural History in Luxembourg
123 because the high numbers of records on these days indicated that either records without
124 collecting date had been assigned these by default, or the dates were used by BioBlitz events
125 where very large numbers of records are taken on a specific day of the year. Including these
126 data would have strongly biased the intra-annual distributions of our records. Finally, we
127 removed some records for which no elevation or temperature data could be obtained (see
128 below).

129 To ensure reasonable coverage of the studied time interval, we then restricted the
130 records to species which had at least 10 occurrence records in every decade covered (with the
131 year 2020 included in the last decade).

132 After these data curation steps, we maintained just above 12 million occurrence records
133 that covered altogether 1,764 species, with 11.4 million records of 1,438 plant species, around
134 590,000 records of 207 species of butterflies and moths, some 76,000 records of 20 bee
135 species and 30,000 records of 22 fly species, and almost 25,000 records of 77 species of
136 beetles (**Table 1**). There were large differences between plants and insects not only in the
137 numbers of records but also in their temporal distribution across the studied period (**Figure S**
138 **1**). While plants had relatively even record numbers across years, the insect groups, in

139 particular flies and bees, were strongly underrepresented in the earlier decades, and record
140 numbers increased rapidly in the last 20 years, probably due to the advent of platforms like
141 iNaturalist.org and naturgucker.de, which allow recording of species occurrences by citizen
142 naturalists, and which made up most of the insect occurrence data for Germany in GBIF.
143

144 *Temperature and elevation data, and individual interactions*

145 Besides the main phenology data from GBIF records, we obtained several other data sets
146 required for our analyses. To test for associations with climate, we used temperature data
147 from the Climate Research Unit (CRU, <https://crudata.uea.ac.uk>) at the University of East
148 Anglia, specifically the Time-Series dataset version 4.05 [33], which contains gridded
149 temperature data from 1901-2020 at a resolution of 0.5° longitude by 0.5° latitude. From this
150 dataset we extracted the monthly mean temperatures and averaged them to obtain the annual
151 mean temperatures at the sites of occurrence records. To be able to control for elevation at the
152 locations of occurrence records, we used elevation data at a 90 m resolution from the NASA
153 Shuttle Radar Topography Mission (SRTM), obtained from the SRTM 90m DEM Digital
154 Elevation Database [34] and accessed through the *raster* package [35] in R.

155 Finally, we obtained data on individual plant-pollinator interactions from a United
156 Kingdom (UK) database on plant-pollinator interactions [36] hosted by the Centre for
157 Ecology and Hydrology (CEH). This database included all known interactions between plants
158 and flower-visiting bees, butterflies, and hoverflies (but unfortunately neither beetles nor
159 moths nor flies other than hoverflies) in the UK, a country similar to Germany in terms of
160 climate and species composition. While these interaction data are unlikely to represent all
161 possible species interactions in Germany, we could not find any similar data for our study
162 area, and so we used this dataset as a prediction for interactions taking place in Germany. To
163 ensure that interactions reflected pollination, we excluded plants that were known not to

164 depend on insects for pollination. This was the case when a plant was pollinated abiotically or
165 only through selfing, or when it reproduced exclusively vegetatively.

166

167 *Data analysis*

168 All data wrangling and analysis was done in R [37]. Before analysing phenology data, we
169 examined patterns of climate change in Germany through a linear model that regressed the
170 annual mean temperature values at the collection sites (= the corresponding $0.5^\circ \times 0.5^\circ$ grid
171 cells) over time.

172 To understand phenology changes in plants and insects, we first estimated the
173 phenological shifts in each taxonomic group (i.e., plants, beetles, flies, bees, and
174 butterflies/moths) as the slopes of linear regressions linking activity, i.e., the DOY of a record
175 to its year of observation. We estimated taxonomic group-specific phenological shifts in two
176 linear mixed-effect models: one that estimated shifts over time and one that related phenology
177 variation to temperature. Both models included the DOY of a record as the response variable,
178 and the latitude, longitude, and elevation of a record as fixed effects. The temporal-change
179 model additionally included the year of a record as a fixed effect and as a random effect
180 across species; the temperature-change model instead included the annual mean temperature
181 at the site of a record as a fixed effect and as a random effect across species. We used the
182 *lme4* package [38] in R to fit these models, and assessed model fits by visually inspecting the
183 relationships between residuals and fitted values, and between residuals and covariates
184 (**Supplementary diagnostic plot documents**). As the random effects from the models agreed
185 well enough with more complex generalized additive models, we considered our linear model
186 reasonably robust.

187 After estimating the average taxonomic group-level phenological shifts, we compared
188 these between different insect orders and plants through ANOVAs, using Tukey post-hoc

189 tests for pairwise tests of differences. We assessed the normal distribution of the slopes
190 through boxplots and checked potential outliers for their plausibility by examining the data
191 from which these slopes were estimated. The same procedure was used in all subsequent
192 ANOVAs. We tested for an association between temporal shifts and temperature change by
193 calculating the Pearson correlations between the slopes of time- and temperature-relationships
194 within each taxonomic group. To test whether a taxonomic group had an average temporal
195 shift or average temperature sensitivity different from zero, we performed a one-sample *t*-test
196 on the slopes in each taxonomic group for both time and temperature slopes, adjusting the *P*-
197 values for multiple testing using the false discovery rate [39].

198 Finally, we analysed the synchrony of plants and pollinators using the data on
199 individual plant-pollinator interactions. For each of the plant-insect pollinator pairs predicted
200 from the UK dataset we calculated the differences of the slopes and intercepts (= value for
201 plant – value for pollinator) for the taxonomic groups estimated in the temporal-change
202 model. From this we obtained a linear equation describing the temporal change in their
203 asynchrony, with positive values in a given year describing asynchrony with pollinators active
204 earlier than plants. For each group, we then tested whether the average shift of asynchrony
205 differed from zero by using a one-sample *t*-test on the slopes for time slopes, and *P*-values
206 adjusted for multiple testing using false discovery rates. Last, we tested for differences in the
207 average shifts between interacting groups (i.e., hoverfly-plant, bee-plant, and butterfly-plant)
208 through ANOVAs of the temporal asynchrony shift data and Tukey post-hoc tests that
209 examined pairwise differences between interaction groups. Together with the information
210 about the shifts of taxonomic groups and individual species, these analyses of plant-pollinator
211 synchrony also allowed to infer which groups were the main agent of change in interactions.

212

213 **Results**

214 Across all collection locations, there was a strong overall trend of climate warming. The
215 annual mean temperature increased by 0.49 °C per decade ($F_{1, 8477} = 2038$, $R_{adj}^2 = 0.19$, $P <$
216 0.001 for the linear model), with a total increase of ~2 °C across the study period 1980-2020
217 (**Figure S2**).

218

219 *Temporal trends in plant and insect phenology*

220 Out of the five studied taxonomic groups, three showed on average significant phenological
221 shifts towards earlier activity: plants, flies, and butterflies/moths. For beetles and bees, the
222 trends were less consistent (**Figure 1A**; **Table 1**). The phenology of plants advanced more
223 strongly than all of the insect pollinators except flies (**Figure 1A**). Among the insects, flies
224 and butterflies/moths advanced most strongly (**Table 1**), with both groups showing distinct
225 temporal shifts from the beetles (among-group differences: ANOVA, $F_{4, 1759} = 33.44$, $P <$
226 0.001; pairwise differences: Tukey post-hoc with $\alpha = 0.05$). There was generally substantial
227 variation of temporal trends within taxonomic groups: Even in plants, flies, and
228 butterflies/moths where >85% of species advanced their phenology (**Table 1**), there were also
229 some for which the opposite was true (**Figure 1B**). In bees and beetles, the numbers of
230 species with advanced versus delayed phenology were more even.

231

232 *Temperature sensitivity of plant and insect phenology*

233 When we analysed phenology variation in relation to climate instead of temporal trends, (i.e.
234 testing for climate sensitivities), the results for the different taxonomic groups were similar,
235 but not entirely (**Figure 1**): In addition to the plants, flies and butterflies/moths, also the
236 beetles significantly shifted their phenology with increasing temperatures (**Table 1**), and the
237 only groups for which the average temperature shifts differed significantly were plants and

238 beetles, and plants and butterflies/moths, respectively (**Figure 1B**; among-group differences:
239 ANOVA $F_{4, 1759} = 22.19$, $P < 0.001$; pairwise differences: Tukey post-hoc with $\alpha = 0.05$). In
240 these comparisons, the plants were generally the group with the larger temperature
241 sensitivities. At the level of individual species, the directions and magnitudes of temperature
242 sensitivities were generally strongly correlated with that of the temporal shifts in phenology
243 (**Figure 1C**), i.e., species which strong climate sensitivities were also the ones that displayed
244 large changes over time. In all taxonomic groups the majority of species showed negative
245 temperature sensitivity, i.e., accelerating phenology at higher temperature (**Table 1**), but there
246 was substantial variation in all groups, with a fraction of species showing the opposite
247 responses.

248

249 *Synchrony of plant-pollinator interactions*

250 When examining the synchrony of individual plant-pollinator interactions (predicted by the
251 UK dataset), we found that all groups showed negative average shifts, i.e., decreasing the
252 phenological lead of pollinators relative to the plants (**Table 2**). The three examined pollinator
253 groups differed in their magnitudes of synchrony changes (**Figure 2A**), but, except in
254 interactions of plants with flies, they did not become more asynchronous but rather more
255 synchronized during the last decades (**Figure 2B**). The magnitudes of temporal changes were
256 generally greatest for plant-bee interactions, with similar shifts for plant-butterfly and plant-
257 hoverfly interactions (**Table 2**; ANVOA: $F_{2, 4399} = 363.33985 = 326.8$, $P < 0.001$; pairwise
258 differences: Tukey post-hoc with $\alpha = 0.05$). In all three interaction groups, the insects tended
259 to be the earlier partner at the start of the study period. However, since plants advanced their
260 phenology faster than most insect groups, they tended to ‘catch up’ over time, relative to the
261 pollinators (**Figure 2 B**). This is observable both in plant-butterfly and plant-bee interactions.
262 and in the case of plant-hoverfly interactions the plants even ‘overtook’ the insects, with the

263 average asynchrony values for this group becoming more negative during the 1980s,
264 indicating that plants were gradually becoming the earlier partner in these interactions.

265

266 **Discussion**

267 In this study, we took advantage of large collections of occurrence records to examine
268 phenological trends of flowering plants and insect pollinators in Germany. We asked whether
269 phenology changes affected the synchrony of predicted interactions of plants and insects, and
270 whether observed changes in phenology, and variation therein, were related to the different
271 taxonomic groups' responses to climate warming. Our results showed that the phenological
272 shifts of plants and insects differed, with plants generally shifting most strongly, and
273 substantial variation among the insect groups. These changes took place across a period in
274 which the mean annual temperature increased in Germany. As pollinators had historically
275 often been active before the plants, the observed faster phenology changes of the plants
276 resulted in increased plant-pollinator synchrony in predicted plant interactions with bees and
277 butterflies during the last decades. For interactions between plants and hoverflies, however,
278 we find a trend towards greater asynchrony. At the level of individual species, there was
279 generally a strong correlation between temporal shifts and temperature sensitivities of their
280 phenology, which suggests that the two are causally related. Our results therefore indicate that
281 both the phenologies of different taxonomic groups and the synchrony of plant-pollinator
282 interactions are changing because of ongoing anthropogenic climate change.

283

284 *Caveats*

285 When interpreting the results of our study, it is important to consider some caveats of the
286 collections data and occurrence records we used. For instance, the spatial distribution of
287 collections data is usually quite heterogeneous, and this was also true for our data (See

288 diagnostic plots 7 & 8 in both **supplementary diagnostic plot documents**). We attempted to
289 correct for spatial heterogeneity in our analyses by including latitude, longitude, and altitude
290 in the statistical models. Moreover, our data are not only spatially but also temporally
291 heterogenous. While the plant data are well-distributed across decades, the pollinator data are
292 sparser during earlier decades, making slope estimation more sensitive to outliers, particularly
293 for individual species. For entire taxonomic groups, however, these effects should cancel out
294 each other.

295 Another problem is that not all of the recorded plants were in bloom at the time of
296 recording or not all insects were adults. Particularly in the GBIF category “human
297 observation” a fraction of records is of plants in a vegetative state. In a preliminary inspection
298 of a subset of 23 early-flowering species we found only about 75% of the records with images
299 to be of flowering species. Because of the large number of plant records, and the lack of
300 images for many, it is impossible to evaluate this problem for our entire data set. However,
301 we don’t expect any systematic trends in such ‘inactive’ specimens over time, so they should
302 not have biased our analyses but rather increased the overall noise in our data.

303 The data we used allowed us to estimate average phenology/activity, but it did not allow
304 to disentangle different aspects of phenology such as changes in first activity versus peak
305 activity, or the duration of phenology. Some of these aspects might change even if peak
306 activity remains unchanged, and it is possible that we have missed some dimensions of
307 phenology variation, and their temporal changes.

308

309 *Phenological shifts over time*

310 We found that both plants and some insect pollinator groups significantly advanced their
311 phenology. The advances of plants in our data are stronger than those reported in most
312 previous publications. For example, Fitter & Fitter [40] compared the first flowering dates of

313 hundreds of plant species in England between 1954-1990 and 1991-2000, and they found an
314 average advancement of -4.5 ± 0.8 (mean $\pm 1.96 * SE / 95\% CI$, same on all errors given
315 below) days. A more recent long-term analysis by CaraDonna et al. [41] of phenology
316 changes in subalpine meadow plants in the Rocky Mountains found an average advancement
317 of peak flowering of 2.5 ± 0.2 days per decade from 1974 to 2012 for individual plant species,
318 and 5.3 ± 1.7 days and 3.3 ± 1.6 days advancement of the spring and summer peak floral
319 abundances of the entire plant community. All of these estimates are lower than the -7.2 ± 0.4
320 days per decade we found over the period from 1980 to 2020. Thus, our data seem to indicate
321 that plants in Germany are shifting their phenology more strongly than in many other regions.

322 For insects, previous studies are less consistent, with widespread but not universal
323 advances in spring phenology (mostly associated with warming) over the last decades [42].
324 For instance, a long-term study of butterflies in California showed that their peak flight times
325 advanced on average by -1.7 ± 1 days per decade [43], a weaker trend than what we estimated
326 in our data from Germany (-3.9 ± 0.6 days per decade). For bees, our data suggest an average
327 shift of 2.6 ± 4 days per decade; however, the large confidence interval makes comparisons
328 difficult. In New York, Bartomeus et al. [44] found that bee phenology advanced by -1.9 ± 0.1
329 days per decade during 1965-2011. This less than what we found in our data, but the
330 discrepancy could be explained by the overall later time period of our study. For flies, there is
331 little previous data on phenology shifts [45], except for Olsen et al. [46] who found an
332 advancement of -6.2 days of the 10th percentile of collection day (their measure of first flight)
333 during 2000-2018. Although this appears similar to the -5.0 days per decade we found, a
334 direct comparison is difficult here because of the different phenology measures. For the last of
335 our studied insect groups, the beetles, we found no significant temporal shift of phenology at
336 all, but also no previous studies except for such on individual pest species like bark beetles or

337 potato beetles. Thus, we cannot judge how common the lack of temporal shifts is that we
338 observed in our beetle data.

339

340 *Temperature sensitivities of plants and insects*

341 At the level of individual species, there were strong correlations between the temporal shifts
342 of species phenologies and their temperature sensitivities, i.e., how phenology was associated
343 with interannual temperature variation. This was true for all taxonomic groups, and it strongly
344 suggests a causal link between phenology shifts and climate change, supporting previous
345 studies such as CaraDonna et al. [41] and Song et al. [47]. The taxonomic groups differed in
346 their average temperature sensitivities, and these differences did not completely match the
347 ones observed in temporal shifts. While plants generally tended to be the most temperature-
348 sensitive group, there were also significant, albeit more moderate, temperature sensitivities in
349 all pollinator groups except for the bees. Previous studies on insect phenology in the
350 temperate zone (reviewed in [42]) have shown that increased spring temperatures are indeed
351 often associated with earlier insect emergence, but that this pattern cannot be generalized as
352 easily as for plants, as temperature–phenology relationships of insects are more complex.
353 While many insects are able to plastically respond to warmer temperatures by speeding up
354 their rates of development (and thus potentially emerge earlier), others have been found to not
355 respond, or even to delay their phenology. This can be because insect development depends
356 on other cues such as rainfall [48], because insects require a cold period during their diapause
357 (if climate warming reduces this chilling period, this may even increase the amount of
358 warming required for subsequent emergence; [42]), or because hibernation states are not
359 temperature-sensitive. Fründ et al. [49] showed that bees overwintering in larval stages
360 responded to higher winter temperatures with delayed emergence, while bees overwintering
361 as adults showed advanced emergence (but had greater weight losses during overwintering).

362 There were some species with negative temperature sensitivities, i.e., delayed phenologies, in
363 our data. This also connects well to some of the findings reviewed by Forrest [50], for
364 instance that during the winter aboveground-nesting bees experience different temperatures
365 than the plants they feed on during the summer. Such microclimate differences between
366 overwintering insects could sometimes explain contrasting climate responses. Furthermore,
367 warming can change the number of generations per year (voltinism; [42]). All the above-
368 mentioned mechanisms can increase variation in the phenological responses of insects to
369 climate warming and may explain why climate change is not always accompanied by
370 phenological advances but might also cause delays – as we observed e.g., in some beetles.

371

372 *Changes in plant-pollinator synchrony*

373 When we analysed the synchrony of predicted plant-pollinator interactions, we found clear
374 trends in shifting synchrony, with different magnitudes in the pollinator groups. As the
375 phenology of plants generally advanced faster than that of insects during the last decades, but
376 plants had historically been the later partner in most plant-pollinator interactions, these shifts
377 overall led to greater plant-pollinator synchrony. However, if the observed trends continue,
378 then many of the studied interactions will soon reach absolute synchrony, and after that the
379 interactions may become more asynchronous again, albeit in the opposite direction. For plant-
380 hoverfly interactions this point has already been reached. If interactions will become more
381 asynchronous again in the future, then the resilience of pollinator networks, in particular
382 through pollinator generalism, could buffer some of the impacts of phenological mismatches
383 [8]. However, while generalist pollinators make up the larger part of the interactions in most
384 pollination networks, some plant-pollinator interactions are highly specialized, and these
385 might be the ones suffering most from future mismatches [51]. One idea for future research

386 could therefore be to focus specifically on specialist pollinators, or to compare long-term
387 trends and climate change effects on generalist versus specialist plant-pollinator interactions.

388

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393

394 **Authors contributions**

395 JF and FMW conceived the study, JF and FMW collected the data, JF analysed the data and
396 wrote the first draft of the manuscript, with guidance from FMW. JFS and OB provided input
397 to data analysis and manuscript writing. All authors read and approved the final manuscript.

398

399 **Data availability**

400 The R code used to conduct the analysis is available at
401 <https://github.com/jonasfreimuth/Phenological-shifts-germany>.

402

403 **References**

404 [1] Cleland EE., Chuine I, Menzel A, Mooney HA. & Schwartz MD. 2007 Shifting plant
405 phenology in response to global change. *Trends in Ecology & Evolution* **22**, 357–365.

406 DOI: 10.1016/j.tree.2007.04.003.

407 [2] Thackeray SJ., Henrys PA., Hemming D, Bell JR., Botham MS., Burthe S, Helaouet P,
408 Johns DG., Jones ID. & Leech DI. *et al.* 2016 Phenological sensitivity to climate across

409 taxa and trophic levels. *Nature* **535**, 241–245. DOI: 10.1038/nature18608.

- 410 [3] Memmott J, Craze PG., Waser NM. & Price MV. 2007 Global warming and the
411 disruption of plant-pollinator interactions. *Ecology letters* **10**, 710–717. DOI:
412 10.1111/j.1461-0248.2007.01061.x.
- 413 [4] Duchenne F, Thébault E, D. Michez, Elias M, Drake M, Persson M, Rousseau-Piot J-S,
414 Pollet M, Vanormelingen P & Fontaine C. 2020 Phenological shifts alter the seasonal
415 structure of pollinator assemblages in Europe. *Nat Ecol Evol* **4**, 115–121. DOI:
416 10.1038/s41559-019-1062-4.
- 417 [5] Parmesan C. 2007 Influences of species, latitudes and methodologies on estimates of
418 phenological response to global warming. *Global Change Biol* **13**, 1860–1872. DOI:
419 10.1111/j.1365-2486.2007.01404.x.
- 420 [6] Kharouba HM., Ehrlén J, Gelman A, Bolmgren K, Allen JM., Travers SE. & Wolkovich
421 EM. 2018 Global shifts in the phenological synchrony of species interactions over recent
422 decades. *Proceedings of the National Academy of Sciences of the United States of*
423 *America* **115**, 5211–5216. DOI: 10.1073/pnas.1714511115.
- 424 [7] Scheffers BR., Meester L de, Bridge TC. L., Hoffmann AA., Pandolfi JM., Corlett RT.,
425 Butchart SH. M., Pearce-Kelly P, Kovacs KM. & Dudgeon D *et al.* 2016 The broad
426 footprint of climate change from genes to biomes to people. *Science (New York, N.Y.)*
427 **354**. DOI: 10.1126/science.aaf7671.
- 428 [8] Miller-Rushing AJ., Høye TT., Inouye DW. & Post E. 2010 The effects of phenological
429 mismatches on demography. *Philosophical transactions of the Royal Society of London.*
430 *Series B, Biological sciences* **365**, 3177–3186. DOI: 10.1098/rstb.2010.0148.
- 431 [9] Robbirt KM., Roberts DL., Hutchings MJ. & Davy AJ. 2014 Potential disruption of
432 pollination in a sexually deceptive orchid by climatic change. *Current biology : CB* **24**,
433 2845–2849. DOI: 10.1016/j.cub.2014.10.033.

- 434 [10] Bartomeus I, Ascher JS., Wagner DL., Danforth BN., Colla S, Kornbluth S & Winfree R.
435 2011 Climate-associated phenological advances in bee pollinators and bee-pollinated
436 plants. *Proceedings of the National Academy of Sciences of the United States of America*
437 **108**, 20645–20649. DOI: 10.1073/pnas.1115559108.
- 438 [11] Gordo O; Sanz JJ. 2006 Temporal trends in phenology of the honey bee *Apis mellifera*
439 (L.) and the small white *Pieris rapae* (L.) in the Iberian Peninsula (1952-2004).
440 *Ecological Entomology* **31**, 261–268. DOI: 10.1111/j.1365-2311.2006.00787.x.
- 441 [12] Kudo G; Ida TY. 2013 Early onset of spring increases the phenological mismatch
442 between plants and pollinators. *Ecology* **94**, 2311–2320. DOI: 10.1890/12-2003.1.
- 443 [13] Schenk M, Krauss J & Holzschuh A. 2018 Desynchronizations in bee-plant interactions
444 cause severe fitness losses in solitary bees. *The Journal of animal ecology* **87**, 139–149.
445 DOI: 10.1111/1365-2656.12694.
- 446 [14] Kudo G, Nishikawa Y, Kasagi T & Kosuge S. 2004 Does seed production of spring
447 ephemerals decrease when spring comes early? *Ecological Research* **19**, 255–259. DOI:
448 10.1111/j.1440-1703.2003.00630.x.
- 449 [15] Parmesan C. 2006 Ecological and Evolutionary Responses to Recent Climate Change.
450 *Annu. Rev. Ecol. Evol. Syst.* **37**, 637–669. DOI:
451 10.1146/annurev.ecolsys.37.091305.110100.
- 452 [16] Jones CA; Daehler CC. 2018 Herbarium specimens can reveal impacts of climate change
453 on plant phenology; a review of methods and applications. *PeerJ* **6**, e4576. DOI:
454 10.7717/peerj.4576.
- 455 [17] Lang PL. M., Willems FM., Scheepens JF., Burbano HA. & Bossdorf O. 2019 Using
456 herbaria to study global environmental change. *The New phytologist* **221**, 110–122. DOI:
457 10.1111/nph.15401.

- 458 [18] Daru BH., Park DS., Primack RB., Willis CG., Barrington DS., Whitfeld T.J. S., Seidler
459 TG., Sweeney PW., Foster DR. & Ellison AM. *et al.* 2018 Widespread sampling biases
460 in herbaria revealed from large-scale digitization. *The New phytologist* **217**, 939–955.
461 DOI: 10.1111/nph.14855.
- 462 [19] Maldonado C, Molina CI., Zizka A, Persson C, Taylor CM., Albán J, Chilquillo E,
463 Rønsted N & Antonelli A. 2015 Estimating species diversity and distribution in the era of
464 Big Data: to what extent can we trust public databases? *Global ecology and*
465 *biogeography: a journal of macroecology* **24**, 973–984. DOI: 10.1111/geb.12326.
- 466 [20] Newbold T. 2010 Applications and limitations of museum data for conservation and
467 ecology, with particular attention to species distribution models. *Progress in Physical*
468 *Geography: Earth and Environment* **34**, 3–22. DOI: 10.1177/0309133309355630.
- 469 [21] Chapman, A. D. 2005 *Uses of primary species-occurrence data, version 1.0.*
470 Copenhagen.
- 471 [22] GBIF: The Global Biodiversity Information Facility. 2022 *What is GBIF?*
472 <https://www.gbif.org/what-is-gbif>. Accessed 11 February 2022.
- 473 [23] Primack D, Imbres C, Primack RB., Miller-Rushing AJ. & Del Tredici P. 2004
474 Herbarium specimens demonstrate earlier flowering times in response to warming in
475 Boston. *American journal of botany* **91**, 1260–1264. DOI: 10.3732/ajb.91.8.1260.
- 476 [24] Brooks SJ., Self A, Toloni F & Sparks T. 2014 Natural history museum collections
477 provide information on phenological change in British butterflies since the late-
478 nineteenth century. *International Journal of Biometeorology* **58**, 1749–1758.
- 479 [25] GBIF.org. 2021 *Occurrence Download*: The Global Biodiversity Information Facility.
480 DOI: 10.15468/dl.t2x9cj.
- 481 [26] GBIF.org. 2021 *Occurrence Download*: The Global Biodiversity Information Facility.
482 DOI: 10.15468/dl.z9wz76.

- 483 [27] GBIF.org. 2021 *Occurrence Download*: The Global Biodiversity Information Facility.
484 DOI: 10.15468/dl.3anj3a.
- 485 [28] GBIF.org. 2021 *Occurrence Download*: The Global Biodiversity Information Facility.
486 DOI: 10.15468/dl.hb2fgs.
- 487 [29] Klotz S, Kühn I & Durka W. 2002 BIOLFLOR – Eine Datenbank zu Biologisch-
488 Ökologischen Merkmalen der Gefäßpflanzen in Deutschland. In *Schriftenreihe für*
489 *Vegetationskunde*, vol 38, pp. 1–333.
- 490 [30] Kevan PG; Baker HG. 1983 Insects as Flower Visitors and Pollinators. *Annu. Rev.*
491 *Entomol.* **28**, 407–453. DOI: 10.1146/annurev.en.28.010183.002203.
- 492 [31] Chamberlain, S. A. & Boettiger, C. 2017 *R Python, and Ruby clients for GBIF species*
493 *occurrence data*. DOI: 10.7287/peerj.preprints.3304v1.
- 494 [32] Chamberlain SA; Szöcs E. 2013 taxize: taxonomic search and retrieval in R.
495 *F1000Research* **2**, 191. DOI: 10.12688/f1000research.2-191.v2.
- 496 [33] University of East Anglia Climatic Research Unit, Harris, I. C., Jones, P. D. & Osborn,
497 T. 2021 *CRU TS4.05. Climatic Research Unit (CRU) Time-Series (TS) version 4.05 of*
498 *high-resolution gridded data of month-by-month variation in climate (Jan. 1901- Dec.*
499 *2020)*. <https://catalogue.ceda.ac.uk/uuid/c26a65020a5e4b80b20018f148556681>.
500 Accessed 1 September 2021.
- 501 [34] Jarvis, A., Reuter, H. I., Nelson, A. & Guevara, E. 2021 *Hole-filled seamless SRTM data*
502 *V4*. <https://srtm.csi.cgiar.org>.
- 503 [35] Hijmans, R. J. 2020 raster: Geographic Data Analysis and Modeling. [https://CRAN.R-](https://CRAN.R-project.org/package=raster)
504 [project.org/package=raster](https://CRAN.R-project.org/package=raster).
- 505 [36] Redhead, J. W., Coombes, C. F., Dean, H. J., Dyer, R., Oliver, T. H., Pocock, M., Rorke,
506 S. L., Vanbergen, A. J., Woodcock, B. A. & Pywell, R. F. 2018 Plant-pollinator

- 507 interactions database for construction of potential networks: NERC Environmental
508 Information Data Centre. DOI: 10.5285/6d8d5cb5-bd54-4da7-903a-15bd4bbd531b.
- 509 [37] R Core Team. 2008 R: A language and environment for statistical computing. Vienna,
510 Austria. <http://www.r-project.org/>.
- 511 [38] Bates D, Mächler M, Bolker B & Walker S. 2015 Fitting Linear Mixed-Effects Models
512 Using lme4. *J. Stat. Soft.* **67**. DOI: 10.18637/jss.v067.i01.
- 513 [39] Benjamini Y; Hochberg Y. 1995 Controlling the False Discovery Rate: A Practical and
514 Powerful Approach to Multiple Testing. *Journal of the Royal Statistical Society: Series B*
515 *(Methodological)* **57**, 289–300. DOI: 10.1111/j.2517-6161.1995.tb02031.x.
- 516 [40] Fitter AH; Fitter RS. R. 2002 Rapid changes in flowering time in British plants. *Science*
517 *(New York, N.Y.)* **296**, 1689–1691. DOI: 10.1126/science.1071617.
- 518 [41] CaraDonna PJ., Iler AM. & Inouye DW. 2014 Shifts in flowering phenology reshape a
519 subalpine plant community. *Proceedings of the National Academy of Sciences of the*
520 *United States of America* **111**, 4916–4921. DOI: 10.1073/pnas.1323073111.
- 521 [42] Forrest JR. K. 2016 Complex responses of insect phenology to climate change. *Current*
522 *opinion in insect science* **17**, 49–54. DOI: 10.1016/j.cois.2016.07.002.
- 523 [43] Forister ML; Shapiro AM. 2003 Climatic trends and advancing spring flight of butterflies
524 in lowland California. *Global Change Biol* **9**, 1130–1135. DOI: 10.1046/j.1365-
525 2486.2003.00643.x.
- 526 [44] Bartomeus I, Park MG., Gibbs J, Danforth BN., Lakso AN. & Winfree R. 2013
527 Biodiversity ensures plant-pollinator phenological synchrony against climate change.
528 *Ecol Letters* **16**, 1331–1338. DOI: 10.1111/ele.12170.
- 529 [45] Hassall C, Owen J & Gilbert F. 2017 Phenological shifts in hoverflies (Diptera:
530 Syrphidae): linking measurement and mechanism. *Ecography* **40**, 853–863. DOI:
531 10.1111/ecog.02623.

- 532 [46] Olsen K, Holm TE., Pape T & Simonsen TJ. 2020 Natural history museum collection and
533 citizen science data show advancing phenology of Danish hoverflies (Insecta: Diptera,
534 Syrphidae) with increasing annual temperature. *PloS one* **15**, e0232980. DOI:
535 10.1371/journal.pone.0232980.
- 536 [47] Song Z, Fu YH., Du Y, Li L, Ouyang X, Ye W & Huang Z. 2020 Flowering phenology
537 of a widespread perennial herb shows contrasting responses to global warming between
538 humid and non-humid regions. *Funct Ecol* **34**, 1870–1881. DOI: 10.1111/1365-
539 2435.13634.
- 540 [48] Bonal R, Hernández M, Espelta JM., Muñoz A & Aparicio JM. 2015 Unexpected
541 consequences of a drier world: evidence that delay in late summer rains biases the
542 population sex ratio of an insect. *Royal Society open science* **2**, 150198. DOI:
543 10.1098/rsos.150198.
- 544 [49] Fründ J, Zieger SL. & Tschardt T. 2013 Response diversity of wild bees to
545 overwintering temperatures. *Oecologia* **173**, 1639–1648. DOI: 10.1007/s00442-013-
546 2729-1.
- 547 [50] Forrest JR. K. 2015 Plant-pollinator interactions and phenological change: what can we
548 learn about climate impacts from experiments and observations? *Oikos* **124**, 4–13. DOI:
549 10.1111/oik.01386.
- 550 [51] Bascompte J; Jordano P. 2007 Plant-Animal Mutualistic Networks: The Architecture of
551 Biodiversity. *Annu. Rev. Ecol. Evol. Syst.* **38**, 567–593. DOI:
552 10.1146/annurev.ecolsys.38.091206.095818.
553

554 **Table 1** Overview of the studied taxonomic groups, their numbers of species and total records, ranges of records per species, and the average
 555 temporal shifts and temperature sensitivities (± 1 S.E.) of their phenology, the significance level of the slope estimate, and the respective
 556 fractions of species that showed a negative slope in their shifts and temperature sensitivities.

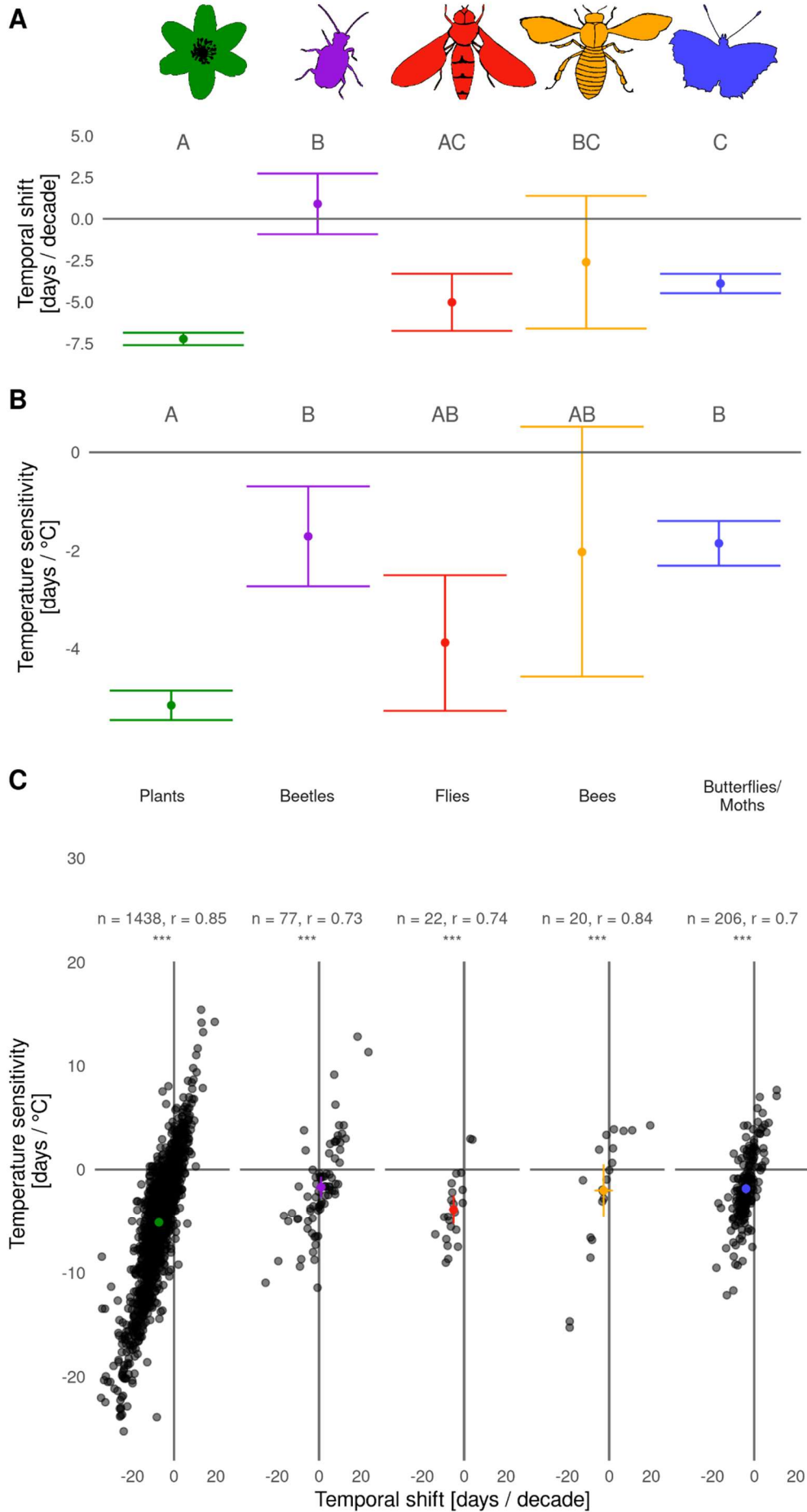
Tax. group	# Species	# Records	# Records/Species			Temporal shifts		Temperature sensitivity			
			Min	Median	Max	Days / decade	<i>P</i> -value	% Species Advancing	Days / 1°C	<i>P</i> -value	% Species Advancing
Beetles	77	24,758	55	196	1,879	0.9 ± 0.9	0.340	44%	-1.7 ± 0.5	< 0.001	74%
Flies	22	30,927	166	675	9,194	-5.0 ± 0.9	< 0.001	91%	-3.9 ± 0.7	< 0.001	91%
Bees	20	76,438	91	599	15,059	-2.6 ± 2.0	0.270	65%	-2.0 ± 1.3	0.140	60%
Butterflies	206	589,970	69	799	34,065	-3.9 ± 0.3	< 0.001	86%	-1.9 ± 0.2	< 0.001	74%
Plants	1,438	11,357,981	62	2,280	121,612	-7.2 ± 0.2	< 0.001	86%	-5.2 ± 0.2	< 0.001	84%

557

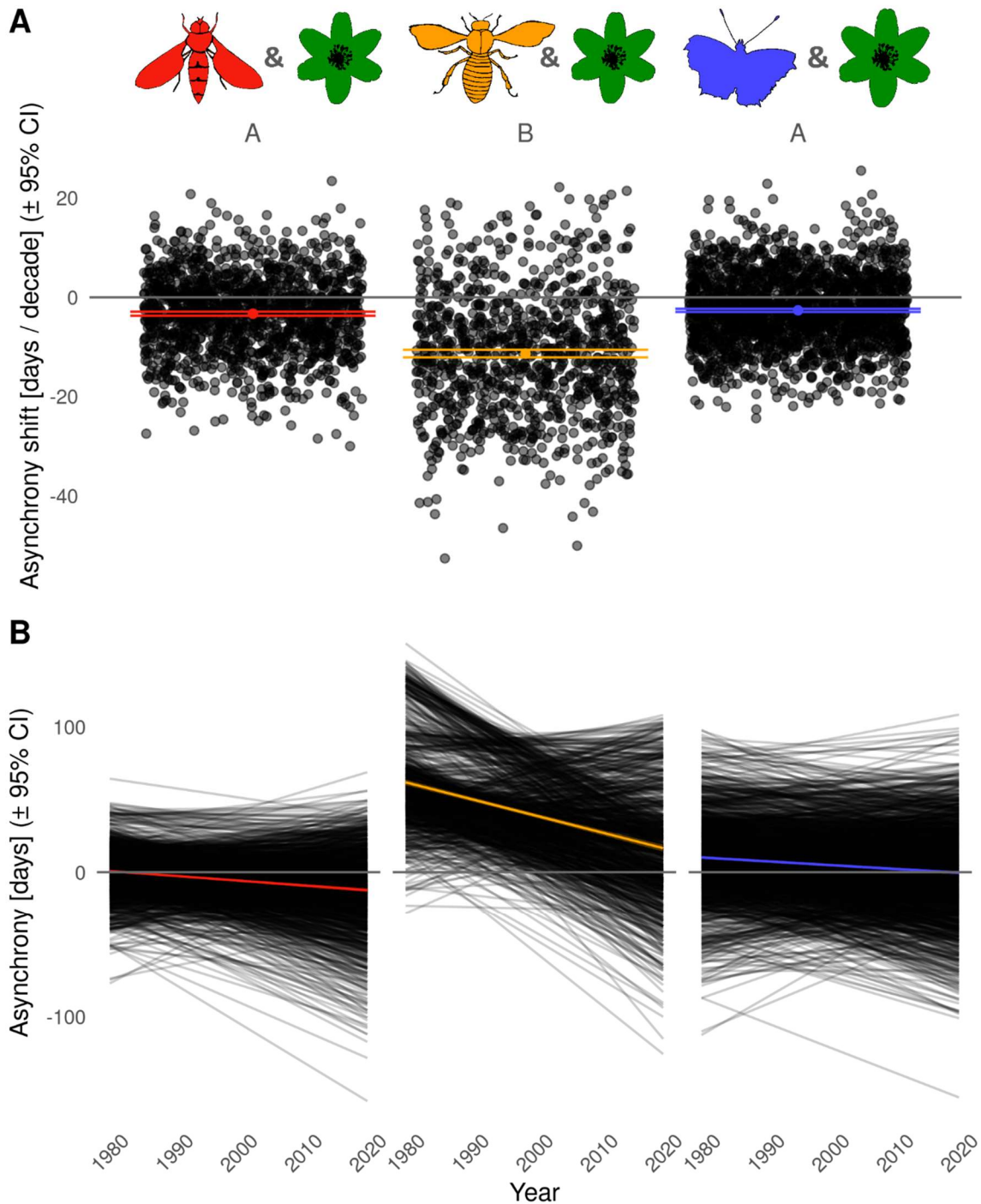
558 **Table 2** Overview of the plant-pollinator interaction groups, the numbers of individual interactions, as well as plant and pollinator species
 559 studied in each, and the observed average temporal changes and temperature sensitivities (± 1 S.E.) of their asynchrony. All estimates in the
 560 last two columns are significantly different from zero at $P < 0.001$.

Interaction group	# Interactions	# Species		Temporal shifts	Temperature sensitivity
		Plants	Pollinators	Days / decade	Days / 1 °C
Hoverfly - Plant	1,407	219	17	-3.3 ± 0.21	-1.81 ± 0.17
Bee - Plant	1,008	256	12	-11.34 ± 0.39	-5.68 ± 0.24
Butterfly - Plant	1,987	260	40	-2.63 ± 0.16	-3.09 ± 0.13

561



563 **Figure 1** Temporal trends (A), temperature sensitivities (B), and their correlation (C) of the
564 phenology of plants and pollinator groups in Germany from 1980 to 2020. In (A) and (B) the
565 values are groups averages \pm 95% confidence intervals, and groups with different letters
566 above are significantly different (Tukey post-hoc tests with $\alpha = 0.05$). In (C) the coloured dots
567 are the averages of taxonomic groups, with 95% confidence intervals, and the black dots are
568 individual species. The Pearson correlation coefficient (r) of each relationship is given above
569 the plot; all correlations are significant at $P < 0.001$.



570

571 **Figure 2** Estimated shifts in asynchrony of individual plant-pollinator interactions (predicted
572 from a UK dataset), separately for different interaction groups. Asynchrony is the difference
573 in the estimated yearly mean DOY of activity between the plant species and the pollinator
574 species in an interaction. (A) Average decadal asynchrony changes of individual interactions
575 (grey dots), and averaged for each group (coloured dots and 95% CI whiskers). Interaction
576 groups with different letters above are significant different (Tukey post-hoc tests with $\alpha =$

577 0.05). (B) Asynchrony changes over time (= slopes and intercepts of pollinators subtracted
578 from plant slopes and intercepts), with black lines for individual interactions and the coloured
579 lines for means across interaction groups.

580